

Bisphenol A interferes with first shell formation and development of the serotoninergic system in early larval stages of Mytilus galloprovincialis

Angelica Miglioli, Teresa Balbi, Lydia Besnardeau, Remi Dumollard, Laura

Canesi

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altered spatial pattern of expression of TYR and HOX 1 depending on the concentration and developmental stage





Highlights

- Bisphenol A-BPA affects early larval development in *Mytilus*
- Concentration dependent effects on deposition of organic matrix and CaCO₃
- Impact on transcription of genes involved in shell biogenesis Tyrosinase and HOX1
- Impaired development of serotonin(5-HT) neurons and changes in 5-HTR expression
- BPA interferes with shell formation and development of the serotoninergic system

1	Bisphenol A interferes with first shell formation and development of the serotoninergic		
2	system in early larval stages of Mytilus galloprovincialis		
3 4 5 6	A. Miglioli ^{1,2} , T. Balbi ¹ , L. Besnardeau ² , R. Dumollard ^{2#} , and L. Canesi ^{1#*} ¹ Dipartimento di Scienze della Terra, dell'Ambiente e della Vita, DISTAV, Università di Genova, Corso Europa 26, 16132 Genova, Italy		
7 8	² Laboratoire de Biologie du Developpement de Villefranche-sur-mer, Institut de la mer, Sorbonne Université, CNRS, 181 Chemin du Lazaret, 06230 Villefranche-sur-mer, France		
9 10	# These Authors equally contributed to the ms.		
11			
12	*Corresponding Author		
13	Laura Canesi		
14	DISTAV, Università di Genova		
15	Laura.Canesi@unige.it		
16			

17 Abstract

Bisphenol A-BPA, a widespread plastic additive, is an emerging contaminant of high concern and a 18 potential endocrine disruptor in mammals. BPA also represents a potential threat for aquatic 19 species, especially for larval stages. In the marine bivalve *Mytilus galloprovincialis*, BPA has been 20 previously shown to affect early larval development and gene transcription. In this work, the effects 21 22 of BPA (0.05-0.5-5 µM) were further investigated at different times post fertilization (24-28-32-48 hpf). BPA induced concentration-dependent alterations in deposition of the organic matrix and 23 calcified shell at different larval stages, as shown by double calcofluor/calcein staining, resulting in 24 altered phenotypes at 48hpf. Transcription of tTyrosinase-TYR, that plays a key role in 25 remodelling of the shell organic matrix, and of HOX1, a member of homeobox genes involved in 26 27 larval shell formation and neurogenesis, were evaluated by In Situ Hybrydization-ISH. BPA altered the spatial pattern of expression of both genes, with distinct effects depending on the concentration 28 and developmental stage. Moreover, BPA affected the time course of mRNA levels for TYR from 29 30 24 to 48hpf. BPA impaired development of serotonin-5-HT-immunoreactive neurons at different times pf; at 48hpf, the reduction in the number of serotoninergic neurons was associated with 31 developmental delay and downregulation of the 5-HT receptor-5-HTR. All the effects were 32 observed from the lowest concentration tested, corresponding to detectable BPA levels in 33 contaminated coastal waters. These data demonstrate that BPA interferes with key processes 34 occurring during the first developmental stages of mussels, thus representing a potential threat for 35 natural populations. 36

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Keywords: Bisphenol A, mussel, early development, shell formation, tyrosinase, hox1, serotonin

39 **1. Introduction**

Bisphenol A (BPA), a monomer used in the production of epoxy resins and polycarbonate plastics 40 is continuously produced and released into the environment in large quantities globally through 41 sewage discharges, landfill leachate, surface runoff, and release from plastic debris. Although it 42 degrades quickly, due to its continual inputs, it is pseudo-persistent and ubiquitous in different 43 aquatic compartments, with reported concentrations between ng and µg/L (Flint et al., 2012; Canesi 44 45 and Fabbri, 2015; Corrales et al., 2015; ECHA, 2017; Sousa et al., 2018; Staples et al., 2018; Wu and Seebacher, 2020). For freshwater, measurements are on average in the range between 10 and 46 some 100 ng/l with mean maximum concentrations of 2000 ng/L for freshwater and above 8000 47 ng/L in marine waters (ECHA, 2017 and refs. therein). In Europe, concentrations of BPA were 48 49 above detection limits in 70% and 61% of freshwater and marine water samples (Staples et al., 2018). BPA is weakly estrogenic and a potential endocrine disrupting chemical-EDC in mammals, 50 and it also represents a potential concern for aquatic organisms, as shown by its reproductive and 51 52 developmental effects on aquatic vertebrates and invertebrates, in particular on early life stages of molluscs and amphibians (Canesi and Fabbri, 2015; ECHA, 2017; Cuvillier-Hot and Lenoir, 2020; 53 Wu and Seebacher, 2020). In July 2019, the General Court of the European Union confirmed that 54 BPA must be listed as a substance of 'very high concern' (https://www.ehn.org/bisphenol-a-55 2639174350.html). 56

57 Marine invertebrates are sensitive to environmental stressors, especially during early larval stages. 58 In addition, taxa characterized by calcified structures (e.g. molluscs, echinoderms) are markedly 59 vulnerable to global changes, including ocean acidification and pollution (Hamdoun and Epel, 50 2007; Przeslawski et al., 2015). In this context, bivalve molluscs represent an important ecological 51 group that, by living in complex environments such as coastal ecosystems, experience both natural 52 environmental fluctuations (pH, temperature, salinity) and exposure to anthropogenic chemicals, 53 and also show rapid adaptation capabilities to environmental fluctuations (Bitter et al., 2019). In the

mussel *Mytilus galloprovincialis*, BPA has been shown to affect early development in the 48h 64 embryotoxicity assay in a wide concentration range (Fabbri et al., 2014). Further studies 65 demonstrated that BPA, from nanoMolar concentrations, significantly affected transcription of 66 several genes, including those involved in biomineralization and the serotonin receptor 5-HTR, both 67 at the trocophora and D-veliger stage (24 and 48h pf post fertilization-pf, respectively) (Balbi et al., 68 2016). However, it is not known whether the observed changes in gene transcription induced by 69 70 BPA are associated with impaired shell growth and formation of serotoninergic neurons, two key processes occurring during the first stages of larval development. 71

In M. galloprovincialis, the processes leading to the formation of the first larval shell have been 72 73 recently investigated. The time course of both organic matrix and $CaCO_3$ deposition was evaluated, from the initial shell only made of organic matrix and characterized by a saddle shape (the shell 74 field), to development of the first calcified larva, the D-veliger (Miglioli et al., 2019). These 75 76 processes were associated with upregulation of genes involved in biomineralization (Tyrosinase, Chitin synthase, Carbonic anhydrase, Extrapallial Protein). In particular, the results underlined the 77 78 pivotal role of tyrosinase in matrix deposition and remodelling and in driving first shell calcification 79 (Miglioli et al., 2019).

In the present work, the effects of BPA (0.05, 0.5 and 5 μ M) on the mechanisms involved in shell 80 formation, and on the development of the serotoninergic system, were investigated in mussel larvae 81 82 at close times pf (24, 28, 32, and 48h). Deposition of the organic matrix and of the calcified shell were evaluated by confocal imaging of Calcofluor/Calcein staining. The effect of BPA on larval 83 phenotypes at 48hpf was also evaluated. Transcription of tyrosinase (TYR) was quantified by qPCR 84 85 and its spatial pattern of expression by In Situ Hybridization-ISH. The effects of BPA on localization of transcripts for *M. galloprovincialis* Homeobox 1 (HOX1), a conserved transcription 86 87 factor involved in ectodermal patterning and differentiation (Barucca et al., 2003; Perez-Parallè et al, 2005) and in molluscan larval shell formation and neurogenesis (Huan, et al. 2020) was also 88 evaluated by ISH. Development of the serotoninergic system was investigated by evaluating 89

90 transcription of the serotonin receptors 5-HTR by qPCR and the presence of serotoninergic neurons
91 by 5HT immunoreactivity.

92

93 2 Materials and Methods

94 2.1 Animal handling, gamete collection and fertilization

All procedures were carried out as previously described (Miglioli et al., 2019). and details are reported in SI. Fertilized eggs were exposed to different concentrations of BPA, depending on the endpoint measured (final concentrations 0.05, 0.5 and 5 μ M, corresponding to 11.4, 114 and 1140 μ g/L). These concentrations fall within the range previously utilized for evaluating the effects of BPA on mussel larval development at 48 hpf (Fabbri et al., 2014).

100 2.2 Effect of Bisphenol A on formation of the first larval shell: morphometric analysis

Fertilized eggs were exposed to BPA in 24-well plates at 30 min pf. Stock solutions of BPA (0.5 M 101 in Dimethyl-sulfoxide/DMSO) were diluted in MFSW at 100 µM and tenfold serially diluted to 102 103 obtain the desired concentrations (100 nM, 1 µM and 10 µM - 2x with respect to final concentration). The time-course of shell formation was investigated at different times pf, between 104 the trochophore and the D-veliger stage (24, 28, 32 and 48 hpf). For details on early larval 105 development see SI. Control samples in MFSW added with the vehicle at the final concentration 106 utilized (0.01% DMSO) were run in parallel. Fluorescent dyes were employed to visualize the 107 components of the larval shell as previously described (Miglioli et al., 2019): Calcofluor White 108 Fluorescent Brightener 28 (Sigma Aldrich, Lyon, France) for organic matrix and Calcein (Sigma 109 Aldrich, Lyon, France) for CaCO₃ deposition, respectively, as previously described (Miglioli et al., 110 111 2019) and details are reported in SI.

Larvae were imaged at each experimental time point with a Leica SP8 Confocal Laser Scanning
Microscope (CLSM - Leica, France) scanning sequentially Calcofluor, Calcein and Brightfield
signals with a 0.5 µm Z-stack interval. Channels were merged, 3D rendered and rotated to measure

the area (in μ m²) of one organic/calcified valve in each larva by manual drawing using IMAGEJ software (Kapsenberg et al., 2018). Area measurements were performed on larvae obtained from 5 independent parental pairs (N=5) and on least 15 individuals for each experimental condition. Data were normalized with respect to controls for each parental pair in different experimental conditions (times pf and exposure).

At 48 hpf, the effect of BPA was also evaluated by scoring shell malformations and measuring shell
lengths in larvae from 5 independent parental pairs and in at least 50 larvae for each parental pair
(see SI). 2.3 In situ hybridization – ISH

The ISH protocol previously described for *M. galloprovincialis* larvae (Miglioli et al., 2019) was 123 used to investigate the expression patterns of Mg-tyrosinase (Mg-Tyr, GenBank: KV583276.1) and 124 Mg-Homeobox1A (Mg-HOX1A, GenBank: FN597580.1). Primer pairs were designed to amplify 125 approximately 1-1.5 Kb cDNA fragment of Mg-Tyr (5'-ATGCGATTCTTTATACATGAAA-3';3'-126 127 TTGGTGGTTTTTGGTACATGT-5') and Mg-HOX1A (5'-ATGAATACGGAGAGTGATTACACA-3'; 3'-TCGACTGTCCTGACATAAACCCAT-5'). 128 The 129 amplicons were then used to synthesize sense and antisense digoxigenin-labelled RNA probes as 130 previously described (Miglioli et al., 2019). Larvae (control and BPA-exposed) were grown in 50 mL culture flasks as described above. At different times pf (24, 28, 32 and 48 h), larvae were 131 sampled and concentrated with a 60 µm nylon filter, fixed overnight in 4% Paraformaldehyde in 132 Phosphate Buffer Saline-PBS (137 mM NaCl, KCl 2.7 mM, 10 mM Na₂HPO₄, KH₂PO₄ 1.8 mM, 133 pH 7.4), washed in the same buffer and kept in 100% methanol at -20°C. Analyses were performed 134 on larvae obtained from at least 4 independent parental pairs (100 individuals for each parental pair 135 and experimental condition). 136

137 *2.4 qPCR*

Eggs were fertilized in polystyrene 6-well plates and treated with 2x BPA solution in order to obtain
the final nominal concentration of 0.05 μM in a 8 mL volume. Control samples were run in parallel.
Larvae were collected at 24, 28, 32 and 48 hpf by a nylon mesh (20 μm pore-filter) and washed

with artificial seawater (ASW) (Balbi et al., 2016). Three wells for each stage were pooled in order
to obtain approximately 7000 embryos/replicate. RNA extraction, retro-transcription and qPCR
were carried out as previously described (Balbi et al., 206; Miglioli et al., 2019) and are described
in SI.

145 2.5 5-HT immunocytochemistry

Antibodies against the neurotransmitter serotonin (5-HT) were used to visualize the development of 146 the serotonergic system in control, 0.05 and 0.5 µM BPA- exposed larvae at different times pf, 147 using an adaptation of the protocols already available for larvae of other bivalves (Voronezhskaya 148 et al., 2008, Yurchenko et al., 2018). Subsamples of larvae employed for ISH were used for 5-HT 149 150 immunostaining. Larvae were rehydrated from 100% Methanol with 3 x 15 min washes in PBS, decalcified by treatment with 0.1 M Ethylenediaminetetraacetic acid (EDTA) in PBS for 1 h at 151 room temperature-RT, and rinsed with 3 x 15 min washes in PBS. In order to eliminate non-specific 152 153 binding sites, larvae were incubated overnight in Blocking Solution [10% normal goat serum, 0.25% bovine serum albumin (BSA), 0.03% sodium azide (NaN₃), PBS-Tx 1% (Triton X-100)] at 154 155 4° C. After 3 x 15 min washes in PBS-Tx 0.1%, larvae were soaked for 10 min at RT in the primary 156 Antibody Buffer (1:10 Blocking Buffer in PBS), then added with a polyclonal primary antibody against Rabbit 5-HT (Immunostar, Hudson, WI, USA) at a final dilution 1:10.000 and incubated for 157 5 days at 4° C. After 3 x 1 h rinsing in PBS-Tx 0.1%, larvae were incubated for 3 days at 4°C with 158 Goat Anti-Rabbit Rhodamine Red[™]-X (RRX) IgG secondary antibody (TRICT, Ex/Em: 570/590 159 nm - Jackson laboratory, Bar Harbor, Maine, USA) at a final dilution 1:500 in PBS-Tx 0.1%. Non-160 specific binding was evaluated in subsamples incubated only in the presence of the secondary 161 antibody (not shown). Samples were rinsed with 3 x 1 h washes in PBS, stained with 1 µg/ml 162 Hoechst in PBS (UV, Ex/Em: 352/461 nm- Hoechst 33342, Invitrogen) for 30 min at room 163 164 temperature to visualize the nuclei, finally rinsed 3 x 15 min in PBS and mounted with glycerol phosphate buffered solution containing antifading agent (CitiFluorTM AF1, pH 10, Agar scientific). 165 Embryos were imaged with a Leica Sp8 CLSM; Hoechst, 5-HT (ir) and Brightfield signals were 166

167 sequentially scanned with a 0.3 μ m Z-stack interval. Using ImageJ software, channels were merged 168 and subsequent Z-stacks were assembled in order to assess the number of immunoreactive (*ir*) cells 169 by counting Hoechst stained nuclei in the 5-HT *ir* region as shown in Fig. S1.

170 2.6 Statistical analysis

All data, obtained from 5 or 4 independent parental pairs, depending on the experiment (mean ± SD) were analysed by one-way nonparametric ANOVA (Kruskal-Wallis test) followed by the Tukey's test (p<0.05). All statistical calculations were performed using GraphPad Prism 5 software (GraphPad Inc.).

175

176 **3. Results**

177 *3.1 Effect of Bisphenol-A on shell biogenesis and larval development*

Fertilized eggs were exposed to different concentrations of BPA (0.05, 0.5 and 5 µM) and the 178 179 growth of the organic matrix and calcified shell was evaluated by morphometric analysis of stained larvae at different times pf, as previously described (Miglioli et al., 2019). In Fig. 1 representative 180 images are reported of Calcofluor/Calcein staining in control larvae and larvae exposed to different 181 182 concentrations of BPA (from left to right) at 24, 28, and 32hpf, showing the organic matrix (blue) and calcified component (green) of the growing shell. As previously described, in control samples 183 at 24hpf, the initial larval shell was characterized by a saddle shape (the shell field) completely 184 made of organic matrix, with no signs of calcification (Miglioli et al., 2019). BPA exposure induced 185 alterations of the organic matrix, that was reduced to a thin line at the highest concentration tested 186 (5 µM). At 28hpf, when first calcification was visible in control larvae, increasing concentrations of 187 BPA induced changes in both structures. At 32hpf, control samples showed increased calcification, 188 with evident accretion rings, while the hinge region was still not calcified. At this time point, all 189 190 BPA concentrations affected the organic matrix as well as the calcified shell. In particular, larvae exposed to both 0.5 and 5 µM BPA had the typical appearance of immature stages, and no accretion 191 rings were visible in the calcifying shell. At 48hpf, control larvae reached the normal D-veliger 192

193 stage, characterized by extensive calcification and a straight hinge. In BPA-exposed larvae, D-194 veligers were also observed; however, lower concentrations of BPA induced shell malformations, in 195 particular in the hinge region, and also altered the pattern of calcification. At the highest 196 concentration tested (5 µM), BPA led to disruption of both shell matrix and CaCO₃ deposition.

The effects of BPA on the growth of organic matrix and calcified shell were quantified by 197 measuring the areas of each shell component (μm^2) as previously described (Miglioli et al., 2019). 198 Data are reported in Fig. 2, as % values recorded in BPA-exposed larvae with respect to controls at 199 200 24 (A), 28 (B) and 32hpf (C). At 24hpf, BPA induced a dose-dependent decrease of the area occupied by the organic matrix (blue), that was significant at 0.5 and 5 µM BPA (-28% and 64%, 201 202 respectively; p<0.05 (Fig. 2A). A similar effect on the organic matrix was observed at 28 hpf (Fig. 2B). Moreover, at this stage BPA induced a concentration dependent decrease in the calcified areas 203 (green) that paralleled those observed for the organic matrix; such a decrease was significant at all 204 205 the concentrations tested (from -20% to -50% at 0.05 and 5 µM, respectively; p<0.05) (Fig. 2B). Similar decreases in both shell components were observed at 32hpf at all the concentrations tested 206 207 (Fig. 2C), and they were associated with evident phenotypic changes (Fig. 1). The results are 208 summarized in Fig. 2D, where data are reported as organic matrix:calcified shell ratios in different experimental conditions. In control larvae this ratio showed a significant decrease from 28 to 32 209 hpf, reflecting progressive calcification. BPA induced a concentration-dependent increase in the 210 area occupied by the organic matrix with respect to that of the calcified shell, indicating a delay in 211 CaCO₃ deposition. 212

In Fig. 3A main phenotypical alterations induced by BPA at 48 hpf are reported (malformed hinge, protruding mantle, arrested trocophora). BPA induced a concentration dependent decrease in the percentage of normal D-larvae (from 94% in controls to 18% at the highest BPA concentration) (Fig. 3B). A small increase in the percentage of malformed shells (convex or indented hinge) was observed with respect to the number of total larvae at all the concentrations tested (from a minimum of 2.7% to a maximum of 9.4% in control and larvae exposed to 5 μ M BPA. Moreover, a

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progressive delay in development was observed, indicated by the increase in the proportion of immature D-veligers, characterized by protruding mantle, at all the concentrations tested (from 3.4% in controls to 25.7% in larvae exposed to 5 μ M BPA). In addition, a clear concentration dependent increase in the proportion of larvae withheld at the trocophora stage was observed, reaching 50% of total larvae at the highest concentration tested. When the average shell length of all D-veligers (normal, malformed, immature) was measured, both 0.5 and 5 μ M BPA also induced a significant decrease in size with respect to controls (Fig. 3C) (p<0.05).

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3.2. Effect of BPA on gene expression in the shell field: tyrosinase and HOX1

After screening different genes involved in first shell formation (such as CS, CA, EP) (not shown), 228 TYR and HOX1 were selected as possible targets for BPA on the basis of their localization in the 229 area of the shell field (Miglioli et al., 2019; Pin Huan et al., 2020). Expression of tyrosinase (TYR) 230 231 was first quantified by qPCR in control larvae and larvae exposed to the lowest concentration of BPA (0.05 µM). In Fig. 4, data on the time course of the level of TYR mRNA transcripts in control 232 233 and BPA-exposed larvae at different times pf are reported with respect to eggs. In control larvae, a 234 large upregulation was observed from 24hpf, that peaked at 32hpf (with an increase of about 8000folds with respect to eggs), followed by a decrease at 48 hpf. Such trend was affected by exposure 235 to BPA, that significantly increased TYR transcription at 24 and 32 hpf (+67% and +34 % 236 compared to controls respectively, p<0.05). The effects of BPA on the TYR expression pattern were 237 then evaluated by ISH (Fig. 5). As previously described (Miglioli et al., 2019) in larvae grown in 238 physiological conditions, TYR was highly expressed in the shell field area from 24 hpf, and its 239 240 localization subsequently paralleled the growth of the organic matrix as the trochophore developed into the first D-Veliger at 48 hpf, when the signal was weaker and restricted to the margins of the 241 242 growing shell (Fig. 5, first column). At 24 and 28 hpf, lower concentrations of BPA (0.05 and 0.5 µM) did not significantly affect TYR expression pattern. However, from 32hpf, an altered spatial 243 pattern was observed at all the concentrations tested, that was more evident at 48 hpf, when the 244

245 TYR signal was detected in the whole larval body. At the highest concentration (5 μ M) BPA 246 affected localization of TYR transcripts at all times pf.

ISH was also utilized to investigate the expression pattern of HOX1 mRNA and the results are 247 reported in Fig. 6. In control larvae, HOX1 expression was highest at 24 hpf, when it was localized 248 at the external margins of the shell field, but not in its centre, as clearly observed in the dorsal view 249 250 of the larva. From 28 hpf, the signal was concentrated at two distinct, small areas along the hinge 251 region (lateral view, black arrows), and it progressively weakened until becoming hardly detectable at 48hpf. BPA clearly affected HOX1 expression pattern at all times pf and at increasing 252 concentrations. At 24 hpf, HOX1 transcripts progressively filled the centre of the shell field; at 253 254 28hpf the two distinct sites of expression (black arrows) were closer to each other until almost in contact in correspondence to the indented hinge at the highest concentration tested. At 32 hpf, BPA-255 256 exposed larvae showed a stronger signal (similar to that observed in control larvae at 28 hpf) 257 indicating developmental delay, and the two sites of expression were closer with respect to controls. Finally, at 48 hpf, BPA apparently increased the HOX1 signal, but it strongly altered its pattern of 258 259 expression in the whole larvae at all concentrations tested.

260 3.3 Effect of BPA on development of the serotonergic system

Levels of mRNA transcripts for the serotonin receptor 5-HTR were quantified by qPCR in control larvae and larvae exposed to BPA (0.05 μ M) and data on the time course of 5-HTR transcription at different times pf are reported with respect to eggs in Fig. 7. In control samples, after a first upregulation at 24 hpf (about 150-folds with respect to eggs), a decrease in mRNA abundance for 5-HTR was observed at both 28 and 32 hpf, followed by strong upregulation at 48 hpf (about 700 folds vs eggs). BPA exposure did not affect transcription of 5-HTR between 24 and 32 hpf, whereas a significant decrease was observed at 48 hpf (-64% with respect to controls; p<0.05).

5-HT immunofluorescence labelling was utilized to detect serotonin positive cells by confocal microscopy, together with Hoechst staining to identify cell nuclei (Fig. S1), in control larvae and larvae exposed to 0.05 μ M BPA at (24, 28, 32 and 48hpf). The results are reported in Fig. 8 and

Fig. 9, respectively. In control larvae 5-HT immunoreactivity (ir) was detectable from 24 hpf and it 271 272 was concentrated in few cells (Fig. 8A, columns a and b) that were located in the pre-trochal area of the larva, close to the apical cilia (Fig. S2). As shown more detail in Fig 8A (column c), two 5-HT-273 274 ir cells closely adjacent to each other could be identified in trochophora at 24hpf. As the larva developed, the number of 5-HT-ir cells progressively increased to three at 28 hpf, and to four in the 275 late trocophora (32 hpf), where the emergence of tiny neurites could be also observed. In BPA-276 exposed larvae 5-HT immunoreactivity, was clearly affected with respect to control larvae; in 277 particular, no more than two 5-HT-ir cells could be observed from the early to the late trocophora 278 stage (from 24 to 32 hpf) (Fig. 8B). At 48 hpf (Fig. 9), control D-veligers showed up to seven 279 280 serotonin positive neurons, organized into a compact pyramidal structure (columns a and b), and emerging axons could be clearly identified expanding towards the hinge in the dorsal side of the 281 larva (Fig. 9A, column c). At this stage, in BPA-exposed larvae, different effects were observed 282 depending on the BPA-induced phenotype. In D-veligers showing protruding mantles, only five, 283 loosely connected 5-HT-ir cells were observed (column c). Moreover, ectopic 5-HT-ir cells were 284 285 occasionally observed in the protruding mantle (not shown). In arrested trocophorae, no more than 286 three 5-HT-ir neurons could be observed. Higher concentrations of BPA (0.5 µM) more severely affected 5-HT immunoreactivity (Fig. S3 and Fig. S4). From 24 to 32 hpf, only one and two weakly 287 288 5-HT-ir cells could be observed (Fig. S3). At 48hf, a stronger reduction in 5-HT-ir cell number was observed in malformed and immature D-veligers and in arrested trocophorae, where three, four and 289 two serotonin positive cells could be observed, respectively (Fig. S4). 290

291

292 4. Discussion

The results of this work obtained in the mussel *M. galloprovincialis* represent the first data on the effects of BPA, a contaminant of emerging concern, on first shell formation and development of serotoninergic system, two key processes of early larval development in bivalves. Different effects of BPA could be observed from the lowest concentration tested (0.05 μ M, corresponding to 11.4 μ g/L), that falls within the environmental concentrations of BPA detected in some coastal areas (Corrales et al., 2015).

Increasing concentrations of BPA affected the growth of both components of the shell (organic ma-299 300 trix and CaCO₃ deposition), this mainly resulting in developmental delay. When the overall impact of BPA on the growth of the two shell components was quantified by the organic matrix:calcified 301 302 shell ratio, a concentration and time dependent increase was observed at increasing times pf, indi-303 cating an unbalance in the processes of shell formation. In particular, BPA induced alterations in the pattern of deposition of both the organic matrix, that were progressively detectable from the early 304 trochophora stage, and of CaCO₃, with barely detectable shell accretion rings and non-305 306 homogeneous calcification at later stages. These effects were reflected at 48 hpf by a concentration dependent increase in the frequency and severity of altered shell phenotypes, resulting in develop-307 mental arrest at the highest concentration tested. The double Calcofluor/Calcein staining, previous-308 309 ly utilized to characterize the process of first shell formation in larvae grown in physiological conditions (Miglioli et al., 2019) and in response to acidification (Kapsenberg et al., 2018), thus appears a 310 311 suitable method to evaluate not only phenotypical changes induced by chemical exposure, but also 312 to identify early contaminant-induced alterations in organic matrix deposition versus calcification.

With regards to the mechanisms involved in the effects of BPA, we have previously shown that, 313 314 among different genes involved in shell formation, tyrosinase was the most upregulated gene in mussel early larval stages; moreover, TYR expression paralleled the formation of the shell field and 315 growth of the organic matrix from 24 hpf (Miglioli et al., 2019). These data indicated a key role of 316 this phenol oxidoreductase, in the early and progressive guiding of matrix deposition and shell 317 growth in M. galloprovincialis. The results obtained here show that in BPA exposed larvae the ef-318 fects on shell growth were associated with quantitative and qualitative changes in transcription of 319 320 tyrosinase from 24 to 48hpf: from the lowest concentration tested, BPA affected both the time course of transcription and the spatial pattern of expression of TYR, as shown by qPCR and ISH, 321 respectively. These data indicate expression of TYR, a key gene in matrix deposition, represents a 322

significant target for BPA in *Mytilus* early larval development. However, tyrosinases participate in 323 the cross-linking and maturation of shell matrix proteins by catalysing the quinoprotein oxidation 324 (Ren et al., 2020). Therefore, BPA may act on other process upstream of matrix deposition and 325 326 more directly responsible for the initial shell patterning. Homeobox-containing genes play essential roles in the regulation of development and differentiation of many organisms. In particular, HOX 327 genes, a subset of homeobox genes that encode transcription factors containing a highly conserved 328 329 homeodomain-binding motif, are involved in anterior-posterior axial patterning and morphogenetic processes (Gehring et al., 1994). Several HOX genes have been identified in Mollusca; however, 330 due to the vast diversity of body plans in each group (from bivalves to cephalopods), their function-331 332 al role in development in each group is still to be elucidated (Huan et al., 2020). In gastropods, expression of HOX genes is initially detected in the trochophora shell field. Later, during develop-333 ment, these genes are expressed in the mantle, demonstrating their continuous role in larval shell 334 335 formation and differentiation of mantle edge that secretes the adult shell (Samadi and Steiner, 2009, 2010). In bivalves, including M. galloprovincialis, different HOX genes have been identified (Ba-336 337 rucca et al., 2003; Perez-Parallè et al, 2005); however, their role in development is still unknown. It 338 has been recently proposed that HOX expression in the ventral ectoderm likely contributes to ventral patterning, such as neurogenesis, whereas HOX expression on the dorsal side is strongly corre-339 340 lated with shell formation and exhibits lineage-specific characteristics in each class of mollusc (Huan et al., 2020). Here we show that in *M. galloprovincialis* HOX1 was strongly expressed at 24 341 hpf in the dorsal side of the trocophora around the shell field, but not in its centre, and expression 342 progressively declined at longer times pf. Exposure to BPA, from environmental concentrations 343 344 clearly induced dose-dependent alterations in the transcription pattern of HOX1 at all times pf. Although other Mytilus HOX genes will participate in larval development, these results clearly show 345 that expression of HOX1, a transcription factor involved in ectodermal patterning, is concentrated 346 in the shell field, thus supporting its role in initial shell formation, and that it represents a significant 347

target for BPA during larval development. These results support the hypothesis that BPA can affectshell biogenesis from the very beginning of the morphogenetic process.

Previous data showed that both BPA and the natural estrogen 17β -estradiol (at 10 μ g/L, a 350 351 concentration close to the lowest utilized in the present work) could also affect neuroendocrine signalling in early larvae, including transcription of the serotonin receptor 5-HTR (Balbi et al., 352 2016). When transcription of 5-HTR was evaluated in control larvae, upregulation was observed at 353 354 all times pf with respect to eggs, with highest levels of mRNA transcripts in early trocophorae and D-veligers (at 24 and 48 hpf, respectively). BPA induced significant downregulation of 5-HTR only 355 at 48 hpf, confirming previous data (Balbi et al., 2016). In order to evaluate if BPA could affect 356 development of the serotoninergic system, we looked for the presence of 5-HT immunoreactive 357 neurons. In bivalves, neurogenesis begins with the emergence of the apical sensory cells 358 (Voroneskaya et al., 2008 and refs. quoted therein). In M. trossulus, the earliest 5-HT 359 360 immunoreactive (5-HT-ir) neuron was detected in the trochophora in the region of the developing apical organ (AO), the most conserved larval sensory structure, which later becomes a part of the 361 emerging cerebral ganglion (Voronezhskaya et al., 2008). In different bivalve species, the AO 362 363 develops from early trochophora where it consists of both 5-HTir and FMRFamide ir cells (Voronezhskaya et al., 2008; Yurchenko et al., 2018; Pavlicek et al., 2018). In M. trossulus and 364 365 *Crassostrea gigas*, the AO consists of three and five 5-HT*ir* cells in trochophorae and in D-veligers, respectively (Voronezhskaya et al., 2008; Yurchenko et al., 2018). Our data show that in control 366 larvae of *M. galloprovincialis*, from the appearance of two 5-HT-*ir* cells in early trocophora at 24 367 hpf, the number of serotonin positive neurons progressively increased up to seven at 48 hpf, with 368 369 the appearance of tiny neurites. These are the first results on the identification of 5-HTir neurons in the AO of *M. galloprovincialis* early larval stages. The higher number of 5-HT-ir neurons detected 370 371 in the present work with respect to previous studies in other bivalves is probably due to the utilization of the nuclear staining and 3D confocal imaging, which allows a more accurate cell 372 counting on sequential z-stacks of 5-HT-ir and Hoechst merged fluorescence signals (Fig. S1). 373

Exposure to, environmental concentrations of BPA affected the development of 5-HT-*ir* neurons. A decrease in the number of 5-HT-*ir* cells was detected from 28hpf. At 48 hpf, the extent of reduction in cell number was associated with different degrees of developmental delay (from seven in normal D-veligers, to five in early veligers and three in larvae withheld at the trocophora stage, respectively). Higher concentrations resulted in even stronger reductions in the number of serotonin positive neurons, that also in this case were associated with altered phenotypes at 48hpf (malformed, immature, arrested).

The results demonstrate that in mussels exposure to BPA interferes with the development of another ectodermal structure, i.e. neurogenesis of the serotoninergic system. Although the reduction in number 5-HT-*ir* cells was detectable from the trocophora stages, the most evident effects were detected at 48 hpf, when both the number of 5-HT-*ir* cells and the amount of 5-HTR transcripts were reduced. Overall, the results indicate that exposure to BPA can induce neurodevelopmental toxicity in *M. galloprovincialis*.

Current knowledge about the correlation between developmental exposure to BPA and levels of 5-387 388 HT is relatively scarce. In mammalian models, it has been proposed that early life exposure to BPA may alter the metabolism of 5-HT and other neurotransmitters, causing neurodevelopmental 389 disorders and behavioural defects (Castro et al., 2015; Xin et al., 2018). In male zebrafish, BPA-390 induced behavioural changes were associated with reduced levels of 5-HT in the hypothalamus and 391 down-regulation of tyrosine hydroxylase in brain tissues (Li et al., 2017). In molluscs, 5-HT 392 signalling is involved in a number of key physiological processes from reproduction, to 393 neurogenesis and muscle contraction (Joyce and Vogeler 2018; Kim et al., 2019; Yurchenko et al., 394 395 2019; Sharker et al., 2019); consequently, exposure to BPA during early life stages may interfere with the onset of the serotoninergic system with potentially dramatic outcomes in wildlife 396 397 populations. Moreover, BPA may affect other components involved in development of the neuroendocrine systems, such as FMRFamide or other neurotransmitters. 398

Overall the results indicate that in Mytilus early larval stages exposure to BPA can have a 399 significant impact not only on initial shell formation but also on development of the nervous 400 system, indicating that this compound can act through pleiotropic mechanisms. Further research is 401 needed to investigate the mechanisms of action of BPA on different target cells of developing 402 larvae and the possible relationship between the two processes. A recent study showed that, in early 403 oyster larvae, shell formation was modulated by monoamine neurotransmitters 5-HT and dopamine 404 (DA) through TGF- β smad pathway, triggering the expression of tyrosinase to form initial shell 405 (Liu et al., 2020). In this light, it is tempting to hypothesize that exposure to BPA may directly 406 affect the metabolism of 5-HT (or other monoamines like DA), leading to dysregulation of organic 407 408 matrix deposition and consequent calcification. Thus BPA has a strong neurodevelopmental impact in bivalves. However, it remains to be shown whether the effects of BPA exposure could be 409 reversed after subsequent recovery to clean conditions or by other environmental factors. Increasing 410 411 knowledge on the first steps of physiological larval development in different bivalve species may help identifying early signs of impact of different environmental stressors, from exposure to 412 emerging contaminants to ocean acidification, thus contributing to a better understanding of 413 potential perturbation of developmental processes of marine invertebrates in a global change 414 scenario. 415

416

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424

425

426 Figure Legends

Figure 1 - Effect of Bisphenol-A (0.05, 0.5 and 5 μ M) on the time course of early shell formation in *M. galloprovincialis* from the trochophora to the first D-veliger stage at different times postfertilization (from top to bottom: at 24, 28, 32, 48 hpf). Columns correspond to experimental conditions (from left to right: Control, 0.05, 0.5, 5 μ M BPA). Representative images (lateral view) show merged fluorescent signals of calcofluor (blue) and calcein (green) staining, respectively, the organic matrix and CaCO₃ deposition. (Scale bars: 10 μ m).

433

Figure 2 - Effect of BPA (0.05, 0.5 and 5 μ M) on the growth of organic matrix (blue) and calcified shell (green) in mussel larvae. Areas of each shell component (μ m²) measured in a single valva of BPA-exposed larvae are reported as % values of control (unexposed) larvae at 24 (A), 28 (B) and 32 hpf (C). In D) ratios of areas matrix/calcified shell in different experimental conditions are reported. Measurements were made on 15 larvae for each condition obtained from 5 parental pairs (N=5). * $p \le 0.05$, BPA-exposed *vs* Controls; # Control 28 hpf *vs* 32 hpf (one way ANOVA Kruskall-Wallis followed by Tukey's test).

441

442 Figure 3 - Effect of BPA (0.05, 0.5 and 5 μ M) on shell morphology at 48 hpf.

A) representative images of control larvae and of the main phenotypes observed in BPA-exposed larvae (scale bar: 20 μ m). B) percentage of different phenotypes: normally developed D-Veligers (white), with hinge malformations (light gray), with extruding mantle (dark grey) and arrested to the trochophora stage (black). C) Shell length of D-veligers, i.e. the anterior-posterior dimension of the shell parallel to the hinge line normalized as % of controls (Inset), * $p \le 0.05$. Measurements were made on 70 larvae for each experimental condition obtained from 5 parental pairs (N=5).

449

Figure 4 - Time course of tyrosinase (TYR) expression in control and BPA-exposed larvae of *M*. *galloprovincialis* evaluated by qPCR at different times pf. Data, reported as fold changes with respect to eggs, represent the mean \pm SD (N=4). Data were analysed by one way ANOVA Kruskall-Wallis followed by Tukey's test and statistical differences between control and BPA-exposed samples at each developmental stage are reported (* *p* < 0.01). Expression of all transcripts was significantly different at all times pf with respect to eggs (*p* < 0.01, not shown).

456

457

458 Figure 5 - Effect of Bisphenol A on TYR expression pattern evaluated by ISH in mussel larvae459 from 24 to 48 hpf.

Columns (from left to right: Control, 0.05, 0.5 and 5 μM BPA); rows (from top to bottom: 24, 28,
32, 48 hpf). Scale bar: 20 μm

462

Figure 6 - HOX1 expression pattern evaluated by ISH in mussel larvae.

A) Control larvae at 24, 28, 32, 48 hpf (lateral view); B) Effect of BPA 0.05, 0.5 and 5 μM at 24
hpf (dorsal view). The red line indicates the perimeter of the shell field. Scale bar: 20 μm.

466

Figure 7 - Transcriptional profile of the serotonin receptor (5-HTR) during early larval development in *M. galloprovincialis* evaluated by qPCR at different times pf. Data, reported as fold changes with respect to eggs, represent the mean \pm SD (N=4). Data were analysed by 1-way nonparametric ANOVA plus Tukey's test and statistical differences between control and BPA-exposed samples at each developmental stage are reported (* *p*<0.01). Expression of all transcripts was significantly different at all times pf with respect to eggs (one way ANOVA followed by Tukey's test; *p* < 0.01, not shown).

474

Figure 8 - Confocal images of serotonin immunoreactive cells (5-HT-*ir*) in early larval stages of *M*.

476 galloprovincialis from 24 to 32 hpf. A) control; B) BPA-exposed (0.05 μ M).

477 5-HT-ir cells are shown in Red/Pink (Ex/Em: 590/617 nm), DAPI stained nuclei are shown in blue

478 (Ex/Em: 358/461 nm). Rows correspond to hours post fertilization: 24, 28, 32 hpf. Columns show:

479 a) brightfield and 5-HT merged channels, scale bar: $20 \ \mu m$;

480 b) 5-HT and DAPI merged channels, scale bar: $20 \ \mu m$.

481 c) 4-5 X magnified images of column B: white asterisks indicate the nuclei of 5-HT-*ir* cells (*) and

482 white arrowheads emerging neurites, respectively, scale bar: $5 \mu m$.

483 A) In control larvae, 5-HT immunoreactivity was concentrated in compact developing neurons from

- 484 24 hpf. In column c) the presence of two, three, and four-five 5-HT-*ir* cells can be identified,
 485 respectively, from 24 to 32 hpf (*).
- 486 B) In larvae exposed to BPA (0.05 μ M), 5-HT immunoreactivity was apparently more scattered.
- 487 Moreover, only two 5-HT-*ir* cells could be observed from 24 to 32 hpf.
- 488
- 489 **Figure 9**

490 Confocal images of serotonin immunoreactive cells (5-HT-ir) in control and BPA-exposed early

491 larval stages of *M. galloprovincialis* at 48 hpf. All details as in Fig. 8.

492 In A) and B) images of control and BPA-exposed larvae are reported, showing representative BPA-

- 493 induced phenotypes and 5-HT immunoreactivity, respectively, scale bars: $20 \ \mu m$.
- 494 In C) details are reported showing that BPA-induced phenotypic alterations are associated with a
- reduction in the number of 5-HT-*ir* cells (*); arrowheads indicate emerging neurites. Scale bar: 5
- 496 μm.
- 497

498 References

- ASTM Standard E724-98, 2012. Guide for Conducting Static Acute Toxicity Tests Starting
 with Embryos of Four Species of Saltwater Bivalve Molluscs. ASTM International, West
 Conshohocken, PA.
- Balbi, T., Franzellitti, S., Fabbri, R., Montagna, M., Fabbri, E., Canesi, L., 2016. Impact of
 bisphenol A (BPA) on early embryo development in the marine mussel Mytilus
 galloprovincialis: Effects on gene transcription. Environmental Pollution 218, 996–1004.
 https://doi.org/10.1016/j.envpol.2016.08.050
- Barucca, M., Olmo, E., Canapa, A., 2003. Hox and paraHox genes in bivalve molluscs.
 Gene 317, 97–102. https://doi.org/10.1016/s0378-1119(03)00657
- 508 4. Bitter, M.C., Kapsenberg, L., Gattuso, J.-P., Pfister, C.A., 2019. Standing genetic variation
 509 fuels rapid adaptation to ocean acidification. Nature Communications 10.
 510 <u>https://doi.org/10.1038/s41467-019-13767-1</u>
- 5. Canesi, L., Fabbri, E., 2015. Environmental Effects of BPA. Dose-Response 13,
 155932581559830. https://doi.org/10.1177/1559325815598304
- 6. Castro, B., Sánchez, P., Miranda, M. T., Torres, J. M., & Ortega, E. (2015). Identification of
 dopamine- and serotonin-related genes modulated by bisphenol A in the prefrontal cortex of
 male rats. Chemosphere, 139, 235–239. https://doi.org/10.1016/j.chemosphere.2015.06.061
- Corrales, J., Kristofco, L.A., Steele, W.B., Yates, B.S., Breed, C.S., Williams, E.S., Brooks,
 B.W., 2015. Global Assessment of Bisphenol A in the Environment. Dose-Response 13,
 155932581559830. https://doi.org/10.1177/1559325815598308
- Su Cuvillier-Hot, V., Lenoir, A., 2020. Invertebrates facing environmental contamination by
 endocrine disruptors: Novel evidences and recent insights. Molecular and Cellular
 Endocrinology 504, 110712. https://doi.org/10.1016/j.mce.2020.110712
- 522 9. ECHA (2017). Agreement Of The Member State Committee On The Identification Of 4,4'523 Isopropylidenediphenol (Bisphenol A) As A Substance Of Very High Concern According
 524 To Articles 57 And 59 Of Regulation (Ec) 1907/20061. Adopted on 14 June 2017.
 525 Available at

- https://www.echa.europa.eu/documents/10162/13638/svhc_msc_agreement_bpa_ii_en.pdf/a
 c9efb97-c06b-d1a7-2823-5dc69208a238.
- Fabbri, R., Montagna, M., Balbi, T., Raffo, E., Palumbo, F., Canesi, L., 2014. Adaptation of
 the bivalve embryotoxicity assay for the high throughput screening of emerging
 contaminants in Mytilus galloprovincialis. Marine Environmental Research 99, 1–8.
 https://doi.org/10.1016/j.marenvres.2014.05.007
- Flint, S., Markle, T., Thompson, S., Wallace, E., 2012. Bisphenol A exposure, effects, and
 policy: A wildlife perspective. Journal of Environmental Management 104, 19–34.
 https://doi.org/10.1016/j.jenvman.2012.03.021
- Gehring, W.J., Qian, Y.Q., Billeter, M., Furukubo-Tokunaga, K., Schier, A.F., ResendezPerez, D., Affolter, M., Otting, G., Wüthrich, K., 1994. Homeodomain-DNA recognition.
 Cell 78, 211–223. https://doi.org/10.1016/0092-8674(94)90292-5
- Hamdoun, A., Epel, D., 2007. Embryo stability and vulnerability in an always changing
 world. Proceedings of the National Academy of Sciences 104, 1745–1750.
 https://doi.org/10.1073/pnas.0610108104
- Huan, P., Liu, G., Wang, H., Liu, B., 2013. Identification of a tyrosinase gene potentially
 involved in early larval shell biogenesis of the Pacific oyster Crassostrea gigas.
 Development Genes and Evolution 223, 389–394. https://doi.org/10.1007/s00427-013-0450z
- Huan, P., Wang, Q., Tan, S., Liu, B., 2019. Dorsoventral decoupling of Hox gene expression
 underpins the diversification of molluscs. Proceedings of the National Academy of Sciences
 117, 503–512. https://doi.org/10.1073/pnas.1907328117
- 548 16. Joyce, A., Vogeler, S., 2018. Molluscan bivalve settlement and metamorphosis:
 549 Neuroendocrine inducers and morphogenetic responses. Aquaculture 487, 64–82.
 550 https://doi.org/10.1016/j.aquaculture.2018.01.002
- 17. Kim, K.S., Kim, M.A., Sohn, Y.C., 2019. Molecular characterization, expression analysis, and functional properties of multiple 5-hydroxytryptamine receptors in Pacific abalone (Haliotis discus hannai). General and Comparative Endocrinology 276, 52–59.
 https://doi.org/10.1016/j.ygcen.2019.03.001

- Li, X., Sun, M.-Z., Li, X., Zhang, S.-H., Dai, L.-T., Liu, X.-Y., Zhao, X., Chen, D.-Y., & 18. 555 Feng, X.-Z. (2017). Impact of low-dose chronic exposure to Bisphenol A (BPA) on adult 556 557 male zebrafish adaption to the environmental complexity: Disturbing the color preference 186. 558 patterns and reliving the anxiety behavior. Chemosphere, 295-304. https://doi.org/10.1016/j.chemosphere.2017.07.164 559
- Liu, Z., Zhou, Z., Zhang, Y., Wang, L., Song, X., Wang, W., Zheng, Y., Zong, Y., Lv, Z., &
 Song, L. (2020). Ocean acidification inhibits initial shell formation of oyster larvae by
 suppressing the biosynthesis of serotonin and dopamine. Science of The Total Environment,
 735, 139469. https://doi.org/10.1016/j.scitotenv.2020.139469
- Miglioli, A., Dumollard, R., Balbi, T., Besnardeau, L., Canesi, L., 2019. Characterization of
 the main steps in first shell formation in Mytilus galloprovincialis : possible role of
 tyrosinase. Proceedings of the Royal Society B: Biological Sciences 286, 20192043.
 https://doi.org/10.1098/rspb.2019.2043
- Page, L. R., & Kempf, S. C. (2009). Larval apical sensory organ in a neritimorph gastropod,
 an ancient gastropod lineage with feeding larvae. Zoomorphology, 128(4), 327–338.
 https://doi.org/10.1007/s00435-009-0093-9
- Pavlicek, A., Schwaha, T., & Wanninger, A. (2018). Towards a ground pattern
 reconstruction of bivalve nervous systems: neurogenesis in the zebra mussel Dreissena
 polymorpha. Organisms Diversity & Evolution, 18(1), 101–114.
 https://doi.org/10.1007/s13127-017-0356-0
- 575 23. Pérez-Parallé, M.L., Carpintero, P., Pazos, A.J., Abad, M., Sánchez, J.L., 2005. The HOX
 576 Gene Cluster in the Bivalve Mollusc Mytilus galloprovincialis. Biochemical Genetics 43,
 577 417–424. https://doi.org/10.1007/s10528-005-6780-4
- Przesławski, R., Byrne, M., Mellin, C., 2015. A review and meta-analysis of the effects of
 multiple abiotic stressors on marine embryos and larvae. Global Change Biology 21, 2122–
 2140. https://doi.org/10.1111/gcb.12833
- 25. Ren, G., Chen, C., Jin, Y., Zhang, G., Hu, Y., Shen, W., 2020. A Novel Tyrosinase Gene
 Plays a Potential Role in Modification the Shell Organic Matrix of the Triangle Mussel
 Hyriopsis cumingii. Frontiers in Physiology 11. https://doi.org/10.3389/fphys.2020.00100

- Samadi, L., Steiner, G., 2009. Involvement of Hox genes in shell morphogenesis in the
 encapsulated development of a top shell gastropod (Gibbula varia L.). Development Genes
 and Evolution 219, 523–530. https://doi.org/10.1007/s00427-009-0308-6
- Samadi, L., Steiner, G., 2010. Expression of Hox genes during the larval development of the
 snail, Gibbula varia (L.)—further evidence of non-colinearity in molluscs. Development
 Genes and Evolution 220, 161–172. https://doi.org/10.1007/s00427-010-0338-0
- Schmittgen, T.D., Livak, K.J., 2008. Analyzing real-time PCR data by the comparative CT
 method. Nature Protocols 3, 1101–1108. https://doi.org/10.1038/nprot.2008.73
- 592 29. Sharker, M.R., Sukhan, Z.P., Kim, S.C., Lee, W.K., Kho, K.H., 2019. Identification,
- characterization, and expression analysis of a serotonin receptor involved in the reproductive
 process of the Pacific abalone, Haliotis discus hannai. Molecular Biology Reports 47, 555–
 567. https://doi.org/10.1007/s11033-019-05162-2
- Sousa, A.C., 2018. Environmental contaminants and endocrine disruption: the story of
 obesogens. Endocrine Abstracts. https://doi.org/10.1530/endoabs.56.s4.3
- Staples, C., van der Hoeven, N., Clark, K., Mihaich, E., Woelz, J., Hentges, S., 2018.
 Distributions of concentrations of bisphenol A in North American and European surface
 waters and sediments determined from 19 years of monitoring data. Chemosphere 201, 448–
 458. https://doi.org/10.1016/j.chemosphere.2018.02.175
- Voronezhskaya, E.E., Nezlin, L.P., Odintsova, N.A., Plummer, J.T., Croll, R.P., 2008.
 Neuronal development in larval mussel Mytilus trossulus (Mollusca: Bivalvia).
 Zoomorphology 127, 97–110. https://doi.org/10.1007/s00435-007-0055-z
- Wu, N.C., Seebacher, F., 2020. Effect of the plastic pollutant bisphenol A on the biology of
 aquatic organisms: A meta- analysis. Global Change Biology.
 https://doi.org/10.1111/gcb.15127
- 34. Xin, F., Fischer, E., Krapp, C., Krizman, E. N., Lan, Y., Mesaros, C., ... Bartolomei, M. S.
 (2018). Mice exposed to bisphenol A exhibit depressive-like behavior with neurotransmitter
 and neuroactive steroid dysfunction. Hormones and Behavior, 102, 93–104.
 https://doi.org/10.1016/j.yhbeh.2018.05.010

612	35.	Yang, B., Pu, F., Li, L., You, W., Ke, C., Feng, D., 2017. Functional analysis of a tyrosinase
613		gene involved in early larval shell biogenesis in Crassostrea angulata and its response to
614		ocean acidification. Comparative Biochemistry and Physiology Part B: Biochemistry and
615		Molecular Biology 206, 8–15. https://doi.org/10.1016/j.cbpb.2017.01.006

- 36. Yurchenko, O. V., Skiteva, O. I., Voronezhskaya, E. E., & Dyachuk, V. A. (2018). Nervous
 system development in the Pacific oyster, Crassostrea gigas (Mollusca: Bivalvia). Frontiers
 in Zoology, 15(1). https://doi.org/10.1186/s12983-018-0259-8
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Credit Author Statement

LC, AM, RD, TB conceived the experiments. AM, TB, LB, performed experiments. AM, RD, TB, LC writing and editing. LC supervision. All authors have read and agreed to the published version of the manuscript.

Supplementary material to Miglioli et al.,

Bisphenol A interferes with first shell formation and development of the serotoninergic

system in early larval stages of Mytilus galloprovincialis

Early development of mussel larvae

In bivalve early development, the onset of shell formation, occurs at the end of the gastrulation, when a group of epithelial cells thickens, forming the shell field. These cells transiently invaginate, forming the shell gland during early trochophore stage (generally from 20 hpf), a swimming larva characterized by two rows of cilia. The trochophora larval body is divided by the prototroch, the equatorial band of ciliated cells, into a pre-trochal area, characterized by the apical tuft, and a posttrochal area distinguished by the presence on the dorsal of the shell forming tissue, the shell field, on the dorsal side of the larva. The shell gland subsequently evaginates and the shell field spreads by flattening of the cells and by mitotic divisions, thus becoming the calcifying mantle. The peripheral cells of the shell gland, that are not internalized during invagination, produce an extracellular lamella – the future periostracum, whose function is to provide the organic matrix support for initial mineral deposition. The evagination is accompanied by the extension in size of the periostracum. Between the periostracum and the cells of the shell field, the primary mineralization takes place. From 28 hpf, shell growth proceeds until 32 hpf: he early shell, the prodissoconch I, exhibits a granular aspect and develops during the late trochophore stage. By this time, the shell field has completed its evagination defining the hinge axis and the body organization undergoes a dramatic change, the larva assuming a triangular shape characteristic of the early veliger stage, characterized by the velum (ciliated swimming and feeding organ). It is followed by the prodissoconch II stage, formed during the D-veliger stage at 48 hpf. The prodissoconch II shell is characterized by concentric growth lines, which mark a change in the calcifying regime.

References

Marin F. 2012 The formation and mineralization of mollusk shell. Frontiers in Bioscience S4, 1099–1125. (doi:10.2741/s321)

Methods

Animal handling, gamete collection and fertilization

Sexually mature specimens of *M. galloprovincialis* were collected from a natural population in the Bay of Villefranche-sur-mer (43.682°N, 7.319°E - France) during the spawning season of 2018/19 (December-March). Animals were kept and maintained at the Institut de la Mer de Villefranche

(IMEV) by the Centre de Ressources Biologiques Marines of the institute (CRBM), transferred to the laboratory and acclimatized in flow-through vessels containing filtered natural seawater MFSW (Millipore filtered seawater) (pH 7.9-8.0, 38 ppt salinity, 15°C). Gametes were obtained by spontaneous spawning; egg quality (shape, size) and sperm motility were checked using an inverted microscope. Sperm and eggs were kept at 16°C until fertilization, that was carried out with an egg:sperm ratio of 1:10. After 30 min, fertilization success (n. fertilized eggs/n. total eggs x 100) was verified by microscopic observation (usually > 90%) and fertilized eggs were transferred to culture vessels at a density of 200 larvae mL⁻¹.

Calcofluor/Calcein staining

Stock solutions of Calcein (4 M in Dimethyl-sulfoxide/DMSO) were diluted in MFSW at 4 mM (4x final concentration). Calcofluor (UV channel, Exc: 408 nm/Em: 450–490 nm) was added to each single well at the time of sampling and prior to fixation (final concentration 0.02 mM in 0.01% DMSO), as previously described (Miglioli et al., 2019). In each well, aliquots of 0.5 mL of 4x Calcein (FITC channel, Exc: 488 nm/Em: 520–560 nm) solution and 1 mL of 2x the BPA solution were added to 0.5 mL of fertilized egg suspension (0.01% DMSO in MFSW) to reach the final nominal concentrations (0.05 0.5, 5 μ M BPA and 1 mM calcein – 0.01% DMSO) in a 2 mL volume. Concentrations of BPA in test solutions were checked by solid-phase extraction SPE/LC/MS as previously described (Fabbri et al., 2014, Balbi et al., 2016).

Larval growth at 48 hpf

At 48 hpf, the effect of BPA was also evaluated by scoring shell malformations and measuring shell lengths in larvae from 5 independent parental pairs and in at least 50 larvae for each parental pair. The frequency of shell malformations (% values) and mean shell lengths (μ m) were measured. A larva was considered normal when the shell was D-shaped (straight hinge) and the mantle did not protrude out of the shell, and malformed if the larva had not reached the stage typical for 48 h (trocophore or earlier stages) or when some developmental defects were observed (concave, malformed or damaged shell, protruding mantle) (Fabbri et al., 2014). The acceptability of test results was based on control samples for a percentage of normal D-shell stage larvae >75% (ASTM, 2012).

qPCR

The larval suspension was centrifuged at 800x g for 10 min at 4°C and the obtained pellets were lysed in 1 mL of TRI Reagent (Sigma Aldrich, Milan, Italy). Total RNA was extracted following the manufacturer's instructions (SigmaAldrich, Milan, Italy). RNA concentration and quality were verified using the Qubit RNA assay (Thermo Fisher, Milan, Italy) and electrophoresis using a 1.5% agarose gel under denaturing conditions. First strand cDNA for each sample was synthesized from 1 µg total RNA (Balbi et al., 2016). Primers pairs employed for qPCR analysis and their accession numbers in GenBank tyrosinase-TYR [KC878467.1], were: forward CGATTCTTTATACATGAAATCTGTG-reverse AAACCGTTATAACAACGTGCTAA: forward CAGCTGCAAGATCGAGGATT-reverse serotonin receptor-5-HTR [AB526218], TGAAGCCATCTTGACTGACG. qPCR reactions were performed in triplicate in a final volume of 15 µL containing 7.5 µL iTaq universal master mix with ROX (BioRad Laboratories, Milan, Italy), 5 µL diluted cDNA and 0.3 µM specific primers. A control lacking cDNA template (no-template) was included in the qPCR analysis to determine the specificity of target cDNA amplification. Amplifications were performed in a StepOne real time PCR system apparatus using a standard 'fast mode' thermal protocol (Thermo Fisher, Milan, Italy). Melting curves were utilized for each target mRNA in order to verify the specificity of the amplified products and the absence of artefacts. Dilution series of cDNA were used to calculate the amplification efficiency of each primer pair. HEL and EF- α 1 were used for data normalization as the best performing combination of reference gene products (EF1/HEL) (Balbi et al., 2016). Analyses were performed on at least four independent mRNA samples. Calculations of relative expression of target mRNAs was performed by a comparative CT method (Schmittgen and Livak, 2008) using the StepOne software tool (Thermo Fisher). Data are reported as fold changes observed within each life stage with respect to unfertilized eggs in control and BPA-exposed samples.

Figures

Figure S1 - Counting of immunoreactive cells stained with anti-serotonin antibody (5-HT-*ir* cells) in a 48 hpf D-veliger from control conditions. The figure shows a montage of sequential z-stacks of 5-HT-*ir* staining (in red) and Hoechst (in blue) merged on the image. Numbers indicate the nuclei and white arrowheads indicate the emerging neurites. Z-distance between 2 images is 0.9 μ m. Scale bars: 5 μ m. The nucleus of cell 1 is visible from z1 to z6, cell 2 from z3 to z12, cell 3 from z6 to z13, cell 4 from z12 to z15, cell 5 from z13 to z19, cell 6 from z15 to z16, cell 7 from z20 to z24.



Figure S2 - Immunolocalization of 5-HT-*ir* cells in a mussel larva at 24 hpf (trocophora, lateral view). 5-HT-*ir* cells are shown as red fluorescence, the perimeter of the shell field is indicated in blue. The trochophora larval body is divided into a pre-trochal (on top) and post-trochal region by the prototroch, the equatorial band of ciliated cells. The pre-trochal area is characterized by the

apical tuft (black arrowhead), while the post-trochal area by the presence of the shell forming tissue, the shell field, on the dorsal side of the larva (in blue). 5-HT-*ir* cells (in red) are located in the pre-trochal area, close to the apical tuft. The perimeter of the growing shell and 5-HT-*ir* cells are placed on the opposite sides of the larva: as the shell always grows on the dorsal side, 5-HT-*ir* cells are located ventrally. Scale bar : $20 \,\mu$ m.



Figure S3 - Effect of BPA (0.5 μ M) on serotonin immunoreactivity (5-HT-*ir*) in early larval stages of *M. galloprovincialis* from 24 to 32 hpf. For control images, refer to Fig. 8. All details as in Fig. 8. In BPA-exposed larvae only one 5-HT-*ir* cells could be observed at 24hpf and two cells at 28 and 32 hpf. Scale bars: 20 μ m.



Figure S4 - Effect of BPA (0.5 μ M) on serotonin immunoreactivity (5-HT-*ir*) in early larval stages of *M. galloprovincialis* at 48 hpf. All details as in Fig. 9. In BPA-exposed larvae phenotypic alterations were associated with a reduction in the number of 5-HT-*ir* neurons (three in indented hinge, four in protruding mantle and two in arrested trocophora, respectively). Scale bars: 20 μ m.

