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Y. Colin, M.-P. Turpault, L. Fauchery, Marc Buée, Stéphane Uroz. Forest plant cover and mineral type determine the diversity and composition of mineral-colonizing fungal communities. European Journal of Soil Biology, 2021, 105, pp.103334. 10.1016/j.ejsobi.2021.103334 . hal-03336939

HAL Id: hal-03336939 https://hal.science/hal-03336939v1

Submitted on 2 Aug 2023

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22 ABSTRACT

23 Soil fungi play a fundamental role in the cycling of nutrients and tree nutrition in forests. As mycorrhizal fungi are closely associated with their host trees and provide them essential 24 25 base cations and phosphorus, we hypothesize that they actively mine soil minerals and 26 transfer their nutritive content to their host. Therefore, mycorrhizal fungi are expected to be 27 important colonizers of soil mineral surfaces relative to other fungal trophic guilds. To test 28 this hypothesis, 50 µm nylon mesh bags were filled with three different mineral types (i.e. 29 calcite, apatite and obsidian) and buried in the same soil colonized by three forest stands (i.e. 30 beech, Corsican pine and coppice with standards). After an in situ incubation period of 29 31 months in the topsoil, mineral dissolution was detectable in the mesh bags, and the analysis of 32 the fungal communities was therefore undertaken. Minerals presented a lower fungal biomass 33 than the surrounding bulk soil and no difference between the three mineral types tested was 34 observed. The high-throughput sequencing (HTS) approach showed that the fungal 35 communities differed among the three forest stands, but in each case, mineral associated 36 fungal taxa differed significantly from those in the bulk soil. To a lesser extent, fungal 37 communities further varied between the three mineral types. Overall, trophic inference 38 analyses suggested that two groups of plant-associated fungi, namely the ectomycorrhizal and 39 endophytic fungi, were the major fungal mineral colonizers. Saprophytes also appeared 40 significantly enriched on minerals comparatively to the surrounding bulk soil. Together, our 41 data reveal at local scale that the mineral-associated fungal communities are first determined 42 by the forest plant cover and then by the type of mineral.

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44 **Key words**: forest; minerals; fungi; tree stand; ITS sequencing

45 **1. INTRODUCTION**

46 Soil minerals and rocks play a fundamental role in the long-lasting functioning of nutrientpoor ecosystems with low-human impact, where they represent the main source of inorganic 47 48 nutrients (i.e., K, Mg, Ca, P)[1]. This is especially the case in temperate forests, which are 49 usually developed on acidic and nutrient-poor soils, while the fertile and easily manageable 50 soils have been dedicated to agriculture [2]. The access to these inorganic nutrients is 51 especially important for the trees that require long periods of time to grow and significant 52 amount of nutrients. Therefore their full development relies on the inorganic nutrients 53 released from the soil minerals or coming from organic matter decomposition and 54 atmospheric deposition [3]. However, each mineral is characterized by its own physico-55 chemical properties (i.e., composition, weatherability, size, age) and its dissolution depends 56 on its properties and the local conditions. In acidic soils, minerals are mainly represented by phyllosilicates (e.g., biotite) or tectosilicates (e.g., K feldspar, quartz), which are the most 57 58 recalcitrant to dissolution [4]. Consequently, their nutritive content is not directly accessible 59 to tree roots, underlining the pivotal role of soil mineral weathering processes in tree nutrition 60 and ecosystem functioning. Mineral weathering occurs through a combination of abiotic and 61 biotic reactions [5-8]. Abiotic weathering operates mostly through the flow of water and 62 carbon dioxide occurring into the soil, the soil acidification and the freeze-thaw changes that progressively break down and dissolve minerals and rocks [9]. Plants, macro- and micro-63 64 organisms can also physically break down the mineral particles through the pressure exerted 65 (e.g., root and mycelium) or directly solubilise minerals through their activities. Plants can 66 also locally change the soil properties (i.e., pH) and promote mineral weathering [7, 10-16]. 67 However, we are far from knowing which microbial taxa are involved in the mineral 68 weathering process.

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69 Plants themselves can physically disrupt mineral particles through the pressure exerted by 70 their roots on mineral surfaces or chemically through the action of chelating and acidifying 71 molecules contained in root exudates [17,18]. Higher mineral weathering rates have been 72 reported for minerals collected in the rhizosphere than in the surrounding bulk soil [19-25]. 73 However, significant differences were observed between plant species growing in a same soil, 74 suggesting a direct species effect or an indirect effect linked to their associated microbiota [19,24]. In this sense, the presence of effective mineral weathering microorganisms (i.e., 75 76 bacteria and fungi) was reported in the rhizosphere of different plants [7,14], suggesting that 77 the ability of plants to weather minerals is not purely a function of their own activity, but a joint activity of the roots and their microbiota. 78

79 The ability to weather minerals has been reported for a broad range of microorganisms 80 including soil bacteria, lichen and fungi [6,7,14,26-28]. Fungi are major actors in forest 81 ecosystem due to their ability to decompose complex organic compounds [29,30] and for 82 some of them, to establish symbiotic associations with plants [31]. Fungal hyphae can explore 83 important volume of soil and colonize pores non-accessible to the tree roots. For symbiotic 84 fungi, this ability improves the access and transfer of nutrients and water to the host plants. 85 The discovery of micrometer-size pores in mineral particles was attributed to the mineral 86 weathering activity of the fungal hyphae [14,32,33]. Although their real implication in the 87 formation of these pores is not established, several mycorrhizal and non-mycorrhizal species 88 have been reported as capable of weathering minerals through bio-mechanical forcing and/or 89 the production of protons and oxalate [14,28,34-39], and consecutively are likely to promote 90 nutrient transfer to plants [21,40,41]. In addition, mycorrhizal fungi with low mineral 91 weathering activities exhibit high-affinity ion transporter systems resulting in a greater 92 capacity to absorb inorganic nutrients from the soil [42].

93 In the soil environment, minerals represent a particular interface due to their physical 94 properties, but also their nutrient content and reactivity to the local conditions. Different 95 factors such as mineral composition, particle size and age have been suggested to condition 96 biomass and mineral weathering activity of fungi [40-46]. Such selective process was also 97 reported on rock surfaces, where significant changes of the fungal community composition 98 were observed according to the mineral composition [47-49]. However in the soil, few studies 99 have applied high-throughput sequencing to provide a good understanding of the composition 100 and potential drivers structuring mineral-colonizing fungal communities [49-53]. In this 101 context, our study aimed at investigating how the soil fungal communities adapt to different 102 types of minerals incubated in situ below different tree stands and which ecological guilds of 103 fungi are mainly associated with these different types of minerals. We hypothesized that (1) 104 the fungal biomass, diversity and community composition on mineral surfaces differ from the 105 surrounding soil (BS), (2) minerals result in an enrichment of different fungal communities 106 according to their physico-chemical properties, and (3) mycorrhizal fungi are major 107 colonizers of minerals particles. To test these hypotheses, pure minerals (i.e., calcite, apatite, 108 and obsidian) enclosed in mesh bags were introduced in the topsoil of the long-term forest 109 observatory (LTO) of Breuil-Chenue (France) for an incubation period of 29 months. The 110 composition and structure of the fungal communities in the different mesh bags as well as of 111 the surrounding bulk soil were analyzed using high-throughput sequencing. The fungal 112 population sizes in mesh bags and in the soil were determined by quantitative PCR (qPCR).

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2. MATERIALS AND METHODS

115 2.1 Sampling site description and experimental design

116 The present study was performed in the Long Term Observatory (LTO) of Breuil-Chenue 117 (Morvan region; 47°18'N, 4°5'E, France). This experimental site is characterized by monospecific plots (0.1 ha) of different tree species planted in 1976 and growing under the 118 119 same climatic conditions. We obtained a permit from the LTO manager to work on the plots 120 planted with beech (Fagus sylvatica L.(Bc)), Corsican pine (Pinus nigra Arn. subsp. laricio 121 Poiret var. Corsicana, (Cp)) or colonized by the native forest (coppice with standards, 122 (CwS)). The term 'coppice with standards' refers to the forestry practice that consists in 123 regularly harvesting the coppice (Quercus sessiliflora Smith, Betula verrucosa Ehrh, and 124 Corylus avellana L.) of the native forest and the conservation of oldest trees (Fagus 125 sylvatica). The soil is classified as a Typic Dystrochrept [54] developed on the "Pierre qui 126 Vire" granite. The bulk soil is sandy-loam textured (60% sands and <20% clays) and acidic 127 (pH_{H2O} ca. 4)[55]. A thick layer of humus is present as moder on the site. The soil 128 characteristics are presented in Table S1.

To investigate the impact of mineral substrates on soil fungal community structure, three mineral types (calcite, apatite, and obsidian) were selected. The minerals used in this work are the same as those used in [11] to analyse the mineral-associated bacterial communities. The chemical composition (Table S2) and the way of preparation of the minerals are described in [11]. These minerals have been selected because of their differences of chemical composition and weatherability. Apatite (P and Ca rich) is encountered in the Breuil-Chenue granite as an accessory mineral, while obsidian (Si, Al and Fe rich) and calcite (Ca rich) are not naturally

136 present in this soil. For each mineral type, particles were subjected to a mechanical grinding 137 step followed by washing / sieving to conserve only particles of 100 to 200 µm in size. The 138 minerals were washed several times to remove any soluble fraction adsorbed on their surface, 139 according to [19]. This size was selected to allow an observable dissolution in few years in 140 our experimental conditions. Nylon bags (mesh size, 50 µm; 4 by 10 cm) were filled with 3 141 grams of pure minerals (i.e., calcite or apatite or obsidian) calibrated in size following the 142 procedure described in references [56,57]. The mesh size $(50\mu m)$ was selected to prevent root 143 colonization and to allow microbial colonization and water circulation [58]. On June 13, 144 2012, mesh bags (n=4 per mineral type) were installed horizontally within the organo-mineral layer (5-15 cm) below the three tree stands (Coppice, Beech and Corsican pine) as described 145 146 in [11]. Soil layers (i.e. litter, organic layer (0-5 cm) and organo-mineral layer (5-15 cm)) 147 were removed one by one to install the mesh bags. The sampling date was determined on the 148 basis of preliminary experiments in order to obtain conditions where mineral dissolution was 149 observable [19,59] and to have established fungal mineral colonization.

150 On November 4, 2014 (29 months later), all the mesh bags (n=4 for each mineral) and the 151 surrounding bulk soil (BS; n=3 below each tree stand) were sampled. For each tree stand, soil 152 chemical properties (pH, total C and N, C/N ratio, organic matter content, phosphorus, 153 cationic exchange capacity and all the cations) were determined and are presented in [11]. 154 After 29 months of incubation in the soil conditions of the Breuil-Chenue LTO, the relative 155 mass loss (final weight/initial weight x 100) was observable and quantifiable for calcite 156 minerals (~15%) and to a lesser extent for apatite and obsidian (~3%) (Figure 1). A 157 subsample of each mineral type was directly imaged by stereomicroscopy and scanning 158 electron microscopy in order to examine the potential presence of fungal hyphae on the 159 mineral particles (Figure 2).

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2.2 DNA extraction, quantitative PCR and ITS metabarcoding inventory

162 The genomic DNA used in this study are the same than those analyzed in [11]. Total DNA was used to quantify the total fungi 18S rRNA gene copies using the 18S rRNA gene-specific 163 164 primers [10 µM each; FR1/FF390 (total fungi; [60]). The quantification performed by qPCR 165 gives a good approximation of the population size (i.e., 18S rRNA gene copies), a quantity 166 related to the fungal biomass. Absolute quantifications were performed using serial dilutions of standard plasmids containing a total fungal 18S rDNA insert (from 10⁹ to 10² gene 167 168 copies/µl) and the SsoAdvanced Universal SYBR Probes Supermix (classical qPCR) from 169 BioRad. Total fungal quantification was performed using the following cycle parameters: 40 170 cycles, with 95°C for 15 s, 50°C for 30 s and 70°C for 60 s. For each qPCR run, a melting 171 curve was performed at the end. The population sizes of total fungi are expressed as 172 log[number of 18S rRNA gene copies per gram dry weight soil]. Subsequently, amplicon 173 libraries were created using barcode-tagged primers for the primer pair ITS1F/ITS2. Each 174 reaction contained 1X PCR Mastermix (5 PRIME®), 500 nM ITS1F primer, 500nM ITS2 175 primer and 7 ng of DNA (as measured by a Nanodrop spectrophotometer). Pyrosequencing 176 was performed by Beckman Coulter Genomics (Danvers, MA, USA) on a Genome Sequencer 177 (GS) FLX 454 Titanium System (Roche).

178

179 2.3 Bioinformatic analyses

The sequencing effort resulted in 516,120 unidirectional ITS sequences across 9 bulk soil samples and 36 mineral samples. The ITS sequences were sorted according to the samples and adapters were deleted. Subsequently, ITS sequences shorter than 150 bp and with

183 expected error > 0.5 were removed. All chimeras and singletons were discarded before further 184 analysis using UPARSE [61]. Two mineral samples collected below the Coppice stand (1 185 mesh bag containing calcite and another containing apatite) resulted in few ITS sequences and 186 were thus discarded from downstream analyses. Overall, a total of 269,739 high-quality ITS 187 sequences were obtained. We rarefied the library size to 2,936 randomly selected reads per 188 sample to correct for differences in sequencing depth and to obtain unbiased non-parametric 189 richness estimates of microbial community diversity [62]. A total of 128,398 high quality ITS 190 sequences were subsequently analyzed and resulted in the identification of 1,295 fungal 191 OTUs. Using UPARSE algorithm in USEARCH, an OTU-based analysis was performed to 192 calculate alpha-diversity estimates in our samples with a cutoff of 3% dissimilarity. Then, ITS 193 sequences were assigned to taxonomic groups based on pairwise similarity searches against 194 the curated UNITE fungal ITS sequence database. Finally, the trophic mode was determined 195 for each OTU using the FUNGuild software [63]. Only the FUNGuild assignments with a 196 confidence score of "Probable" or "Highly Probable" were considered. The taxa with 197 assignment scores below these confidence scores were considered as 'unassigned'. For our 198 analyses, we assigned taxa to one of the six guilds: symbiotroph, saprotroph, pathotroph, 199 pathotroph-saprotroph, unassigned and others. The category "others" grouped other modes 200 that represented a small percentage of sequence reads and were not analysed further.

201

202 2.4 Statistical analyses

Based on the fungal OTUs table, alpha-diversity estimates were calculated using the function "rarefy" from the package 'Vegan' [64]. Richness (S_{obs}) was calculated as the number of observed OTUs in each sample. The diversity within each individual sample was estimated using the non-parametric Shannon index. Correlation analyses were done to 207 determine whether the soil chemical properties were linked to the different indices. We also assessed whether fungal population size and alpha diversity varied using a Student's t-test. 208 209 One-factor analysis of variance (ANOVA) was done independently below each tree stand. 210 The factor considered was the compartment (i.e., bulk soil vs mineral) or the sample type (*i.e.*, 211 bulk soil, calcite, apatite and obsidian). Similar analyses were performed to test the effect of 212 minerals and mineral type on the alpha-diversity within the two main fungal phyla (i.e., 213 Basidiomycota and Ascomycota). Concerning the beta-diversity, we applied a Hellinger 214 transformation of the OTUs table and generated a pairwise distance matrix using Bray-Curtis 215 dissimilarity using the Vegdist function [64]. Samples were then compared to determine 216 whether there were significant differences in the variance and means between tree stands (i.e., 217 Beech, Coppice and Corsican pine), compartments (i.e., bulk soil vs mineral) and mineral 218 types (*i.e.*, calcite, apatite and obsidian) using Nonmetric multidimensional scaling (NMDS) 219 ordinations and permutational multivariate analysis of variance (PERMANOVA). Variance 220 partitioning analysis was performed using the 'varpart' function in the 'vegan' package. The 221 main contributing taxa for dissimilarity between bulk soils and minerals were identified at 222 different taxonomical levels through a similarity percentages (SIMPER) analysis using the 223 function 'simper' from the 'vegan' package [64]. In addition, changes in fungal taxonomical 224 composition and functional guilds (i.e., FUNGuild) between bulk soils and minerals and 225 between the three mineral types were tested using the Student's t-test or ANOVA analysis. 226 When required, a p adjustment was done using the false-Discovery-Procedure (FDR) [65] to 227 determine the p-value threshold. All statistical analyses were carried out in R (v.3.5.1) (R 228 Core Team 2018).

229

230 2.5 Accession numbers

- 231 The ITS pyrosequences generated for this study were submitted to the Sequence Read
- 232 Archive (SRA) under the accession number PRJNA507076.

233 **3. RESULTS**

3.1 Quantification of the fungal population sizes

235 For each tree stand, lower fungal populations (i.e., biomass based on qPCR) were observed 236 on minerals compared to the surrounding bulk soil (BS; Student's t-test, p<0.05; Figure 2A). 237 The stronger difference was observed between calcite samples and their related BS samples 238 below beech and Corsican pine stands (Figure 2A). No difference in population size was 239 observed for each tree stand between the different mineral types (i.e. calcite, apatite and 240 obsidian) (ANOVA; p>0.05), excepted between calcite and apatite below beech stand (Figure 241 2A). Finally, a subsample of minerals was imaged using stereomicroscopy and scanning electron microscopy (SEM), which confirmed fungal colonization of mineral surfaces (Figure 242 243 3).

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245 *3.2 Fungal assemblage from metabarcoding inventory*

Almost 97% of the ITS sequences were classified into five different phyla, predominantly,
Basidiomycota (51.2%), Ascomycota (37.8%) and Zygomycota (7.5%). At the order level, it
was possible to classify about 64% of sequences, the most common orders being *Russulales*, *Boletales* and *Agaricales* among Basidiomycota; *Hypocreales* and *Helotiales* among
Ascomycota; and *Mortierellales* among *Zygomyota*. A total of 1,295 OTUs was generated
from this sequence dataset.

252

254 The fungal diversity was similar in the bulk soil samples of the different tree stands 255 (H'_{Beech}: 2.3 ± 0.1 ; H'_{Coppice}: 2.6 ± 0.5 ; H'_{Corsican pine}: 2.7 ± 0.5) (Figure 2B and Table S3). 256 However, below the beech and Coppice with standards stands significantly higher diversity 257 values were observed for calcite samples than in the related BS samples (Student's t-test; 258 p < 0.05; H'_{Beech}: 3.6 ± 0.9; H'_{Coppice}: 4.1 ± 0.2), but not below Corsican pine (H'_{Corsican pine}: 4.0 259 \pm 0.3, p>0.05; Figure 2B). This increase was mainly attributed to higher Shannon index 260 values in the Basidiomycota (Student's t-test; all p<0.05; Figure S1). No significant 261 relationship was observed between the Shannon index or the observed number of OTUs and 262 the soil chemical properties (p>0.05).

263

3.4 Fungal community structure

265 According to NMDS and PERMANOVA analyses, the tree stand was the main factor 266 explaining the fungal community structure (PERMANOVA : F=9.0; p<0.001) (Figure 4). 267 Noticeably, for each tree stand, a significant compartment effect (*i.e.*, mineral vs BS) was 268 observed (PERMANOVA; Beech: F=4.4; p<0.01; Coppice: F=3.0; p<0.01 and Corsican pine: 269 F=2.8; p<0.01) (Figures 4 and S4). In a lesser extent, the mineral type (*i.e.*, calcite, apatite and 270 obsidian) significantly impacts the structure of the mineral-associated fungal communities in 271 each tree stand (PERMANOVA ; Beech: F=2.1; p<0.01; Coppice: F=2.0; p<0.01 and 272 Corsican pine: F=2.1; p<0.01) (Figures 4 and S4).

Together the tree stand and minerals explained 30.9% of the observed variation at the OTU level as stated by a variance partitioning analysis. Tree stand effect was able to independently explain 28.5% of the variation observed, while the compartment (mineral vs bulk soil) alone explained only 3.2%. When the BS samples were excluded from the analysis to concentrate on the potential effect of the mineral type (i.e., apatite vs obsidian vs calcite), the tree stand and minerals explained 24% of the observed variation. Again, the tree stand effect was the main factor explaining the variance (23%), while the type of mineral alone explained 2.3% of the variation.

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282

2. 3.5 Fungal assemblage changes between BS and mineral samples

283 A similarity percentage analysis (SIMPER) determined that the changes in proportion of 284 the Basidiomycota and Ascomycota contributed respectively to 44.1 - 49.4% and 21.3 -285 28.8% of the observed dissimilarity between BS and mineral samples (Table S5). 286 Basidiomycota were by far the most abundant phylum in the BS samples (relative abundance 287 in the BS samples below each tree stand: Beech: $87.4\% \pm 2.4$; Coppice: $72.4\% \pm 10.0$; 288 Corsican pine: $77.6\% \pm 11.8$) and their relative abundance strongly declined on mineral 289 surfaces (relative abundance in the mineral samples below each tree stand: Beech: $36.3\% \pm$ 290 16.8; Coppice: $51.6\% \pm 31.6$; Corsican pine: $44.9\% \pm 21.1$) (Figure 5A). In contrast, 291 Ascomycota, whose importance was limited in BS (relative abundance: Beech = $8.9\% \pm 2.3$; 292 Coppice = $17.9\% \pm 6.8$; Corsican pine = $16.2\% \pm 8.6$), increased in relative abundance on 293 mineral surfaces (Beech = $56.5\% \pm 15.1$; Coppice = $40.4\% \pm 24.8$; Corsican pine = $34.4\% \pm$ 294 16.3) (Figure 5A). Depending on the tree stand analysed, these compositional changes were assessed as statistically significant for Basidiomycota (ANOVA; Beech: F=26.2, p<0.001; 295 296 Coppice: F=1.2, P=0.29; Corsican pine: F=6.4, p<0.05) and Ascomycota (ANOVA; Beech: 297 F=28.0, p<0.001; Coppice: F=2.3, P=0.16; Corsican pine: F=3.4, P=0.09) (Table S4).

At the order level, *Russulales* and *Agaricales* were responsible for respectively 15.6 - 31.9% and 5.1 - 7.0% of the observed dissimilarity (Table S5). These two orders were more abundant in BS samples than on minerals (ANOVA; all p<0.05, except for *Agaricales* below Corsican pine) (Table S3). On the contrary, the *Hypocreales* and *Helotiales* orders each responsible for nearly 3% of the observed dissimilarity (Table S5) were more abundant on minerals.

When regarding the 25 most abundant species identified in our dataset, 13 were related to ECM fungi, one was related to ericoid fungi and the other appeared related to saprophytes/pathogens and potential endophytes. Notably, ITS sequences related to *Russula ochroleuca*, *Russula puellaris*, *Lactifluus glaucescens* and *Amanita rubescens* were significantly enriched in BS compared to minerals (Table S7).

309

310 *3.6 Fungal assemblage changes between mineral types*

311 When mineral samples were examined without the related BS samples in each tree stand, 312 the composition of fungal communities was shown to vary according to the mineral type. This 313 effect was observable at the phylum level, but was only statistically significant below the 314 Coppice with standards stand (i.e., the native forest). In this tree stand, Basidiomycota 315 appeared more abundant on obsidian (70.6% \pm 16.2) than on calcite (15.5% \pm 5.3), while the 316 opposite trend was observed for Ascomycota (Obs = $26.6\% \pm 16.6$ and Cal = $67.5\% \pm 2.5$) 317 and Zygomycota (Obs = $1.9\% \pm 0.6$ and Cal = $11.3\% \pm 1.6$) (ANOVA, p<0.05) (Figure 5A 318 and Table S7). At the order level, Hypocreales and Tremellales were more abundant on 319 calcite (9.2-20.8% and 1.4-7.1%, respectively) than on apatite (0.8-1.4% and 0.4-1.1%, 320 respectively) and obsidian (1.1-6.0% and 0.5-4.1%, respectively) (ANOVA, all p<0.05, except for Corsican pine _{Hypocreales}) (Table S8). Noticeably, several fungal species belonging to
the *Mortierellales* (*Mortierella horticola*, *M. alpine*, *M gamsii* and *M. macrocystis*) and *Hypocreales* (*Ilyonectria radicicola*) orders appeared more abundant on calcite than on other
minerals below the Beech and Coppice stands, some of them showing significant differences
(ANOVA, p<0.05, Table S9).

326

327

3.7 Trophic mode prediction of fungal communities

328 A total of 419 fungal OTUs was successfully assigned to a trophic mode. These OTUs 329 represented 78.6-90.9% of the total ITS sequences in BS samples and 49.9-65.9% in the 330 mineral samples. Symbiotrophic fungi presented higher relative abundance in BS (Beech = 331 $84.5\% \pm 1.7$; Coppice = $64.6\% \pm 12.9$ and Corsican pine = $73.8\% \pm 11.4$) than on minerals (Beech = 40.0 % \pm 18.2; Coppice = 46.9% \pm 32.8 and Corsican pine = 31.5% \pm 24.2) 332 333 (ANOVA; Beech: F=16.9, p=0.001; Coppice: F=0.8, p=0.39 and Corsican pine: F=8.3, 334 p=0.01) (Figure 5B and Table 1). On the contrary, saprotrophic and pathogenic fungi 335 presented a lower relative abundance in the BS samples (saprotrophs: Beech = $3.5\% \pm 0.8$; 336 Coppice = $9.5\% \pm 5.0$; Corsican pine = $8.1\% \pm 3.8$ and pathotrophs: Beech = $0.9\% \pm 0.2$; 337 Coppice = $1.3\% \pm 0.5$; Corsican pine = $0.9\% \pm 0.3$) than on mineral surfaces (saprotrophs: 338 Beech = 8.8 $\% \pm 6.6$; Coppice = 8.6 $\% \pm 6.9$; Corsican pine = 23.8 $\% \pm 7.8$ and pathotrophs: 339 Beech = $2.35\% \pm 2.3$; Coppice = $7.12\% \pm 9.2$; Corsican pine = $4.7\% \pm 6.5$)(Figure 5B and 340 Table 1AB).

When mineral samples were examined without the related BS samples, both obsidian and apatite were characterized by higher relative abundance of symbiotrophic fungi (36.3-64.8% and 47.4-60.5%, respectively) than calcite (9.4-25.1%)(Figure 5B and Table 1B). Pathotrophic fungi were significantly enriched on calcite (Beech: $4.9\% \pm 2.4$ and Coppice: 20.3% ± 2.1) compared to obsidian (Beech: $0.9\% \pm 0.6$ and Coppice: $1.1\% \pm 0.1$) and apatite particles (Beech: $1.2\% \pm 1.0$ and Coppice: $2.0\% \pm 1.3$) (ANOVA, all p<0.05). Higher saprotrophic relative abundance was observed on calcite (Beech: $14.9\% \pm 7.3$; Coppice: $16.4\% \pm 1.7$) than on obsidian particles (Beech: $4.5\% \pm 1.4$; Coppice: $3.9\% \pm 0.9$) (ANOVA, all p<0.05) (Figure 5B and Table 1B).

350

351 **4. DISCUSSION**

352 *4.1 Mineral chemistry does not determine fungal biomass*

353 The potential effect of the mineral chemistry on fungal colonization of minerals has been 354 reported previously, mostly in boreal forests [45,46,66-68]. Previous studies revealed an 355 increase of the fungal biomass in mesh bags amended with minerals carrying nutritive 356 elements (i.e., P-carrying minerals (i.e., apatite) or K- or Mg-carrying mineral (i.e., hornblende and biotite)). The preferential colonization of minerals carrying limiting nutrients 357 358 would represent an adaptation of fungal communities, or an adaptive trait of trees extended to 359 their symbiotic fungi, to the nutrient limitations occurring in forest soils. However, this 360 increase of biomass on the minerals was not observed in all the experimental sites studied and 361 was shown to vary from one year to the next one on a same site, suggesting combined effects 362 of mineral chemistry, nutrient availability into the soil and nutrient status of the considered 363 forests [43,45,66]. In our study, quantitative PCR analyses indicated that fungal population 364 sizes were lower on mineral surfaces than in surrounding bulk soil. Although the physical 365 differences between the two matrixes (i.e. mineral particles vs the bulk soil) may explain 366 these lower abundances on minerals, similar conclusions were obtained in previous studies

[51,53]. Apatite did not exhibit an increase in fungal biomass in comparison with obsidian 367 and calcite, and the lower fungal biomass was observed for calcite (Ca-carrying mineral). 368 369 Nevertheless, our results are hardly comparable with those of previous studies [43,45,66] due 370 to several differences in terms of fungal quantification methods (*i.e.*, ergosterol vs qPCR), 371 mesh bag strategy (*i.e.*, bags mainly filled with quartz and amended with limited quantity of 372 minerals vs bags fully filled with mineral), forest type (*i.e.*, boreal vs temperate), soil depth 373 (i.e., different soil horizons) and incubation times of mineral in soils (*i.e.*, few months vs 374 years).

375

376 4.2 Land cover primarily determines the structuration of mineral-associated fungal
377 communities

378 The richness, diversity and fungal community composition are known to be driven by the 379 pedo-climatic parameters and land cover [69-71]. At local scale (i.e., plot), fungal 380 communities are also vertically structured (i.e., soil horizons)[72,73]. The the strong effect of 381 the land cover (*i.e.*, tree stand effect) on the composition of soil fungal communities observed 382 in our study was expected and confirmed the important relationship existing between the tree 383 species and their associated fungi [69,74]. Russulales, Agaricales, Boletales, Atheliales, 384 Thelephorales, Mortierellales, Pezizales represented the main orders identified in bulk soil 385 samples and differences in taxonomic composition according to tree species were consistent 386 with previous studies carried out on the same experimental site [69,75]. Notably, our results 387 highlighted that the soil fungal community structure was significantly impacted by minerals 388 and that this compartment effect (*i.e.*, minerals vs bulk soil) was first conditioned by the land 389 cover. These results are relatively new as most of our knowledge on the mineral-colonizing 390 fungal communities comes from mesh bag studies done below Picea abies stands and in boreal forests [44-46,66]. The differences observed may be due to the sampling strategy used in our study. The small-scale approach we used permitted to limit the effects of different confounding factors, especially the soil heterogeneity. On the contrary, the previous studies used a large-scale sampling strategy where the soil heterogeneity was higher, hiding probably the mineral type effect. Interestingly, other recent studies [51,53] reported a mineral type effect using a small-scale approach.

397

398 4.3 Taxonomic modification of the fungal communities colonizing minerals compared to 399 the surrounding bulk soil

400 HTS results revealed that minerals, and especially calcite, tended to harbour more diverse 401 fungal communities than the surrounding bulk soil. A net increase of the relative abundance 402 of ITS sequences assigned to Ascomycota and Zygomycota was observed on minerals, while 403 the Basidiomycota orders Russulales and Agaricales, abundant in both compartments, were 404 significantly more abundant in the bulk soil. One may explain these fungal compositional 405 changes between BS and mineral samples by technical biases. The mesh size would favour 406 some taxa more than others (see the review of [76]). The use of fresh mineral surfaces (i.e., 407 mineral particles generated by mechanical grinding; see Material and methods section) may 408 also favour opportunistic taxa (r-strategy) as it was suggested for bacteria [16]. On the other 409 hand, the minerals contained in the mesh bags contrast considerably with the surrounding 410 bulk soil in that they contain low levels of directly available organic and inorganic nutrients. 411 Therefore, these nutrient-poor conditions may limit the development of fungal biomass and 412 select for specific microorganisms capable of developing under such oligotrophic conditions or of recovering the nutrients entrapped into the minerals. Although the mesh size and fresh 413 414 mineral surface biases cannot be completely excluded, previous studies validated the mesh415 bag use, including several studies on organic matter decomposition [77,78] or mineral 416 colonization [51,58]. Notably, Wallander et al. [58] evidenced that after a trenching 417 experiment, the dominant fungi colonizing mineral-amended mesh bags were mycorrhizal 418 ones as in our study. Interestingly, in our study, while all the mesh bags contained ECM 419 related taxa (i.e., Russulales, Agaricales, Boletales), some were highly colonized and others 420 poorly colonized, confirming a patchy distribution of the fungi in the soil [79-81]. As an 421 example, Boletales related sequences appeared increased in abundance on minerals compared 422 to the bulk soil, but only below the coppice with standards stand (Table S6 and S10). In 423 addition, several OTUs affiliated to the Archaeorhizomyces genus were also detected in our 424 dataset, with some of them enriched on mineral surfaces, but only specifically below certain 425 tree species. As an example, the OTU47 was more abundant on apatite and obsidian than in 426 the bulk soil or on calcite, but only in the beech stand. Notably, Pinto-Figueroa et al. [82] recently reported an increase of the abundance of Archaeorhizomycetes in mineral rich 427 428 environment. These authors suggested that some representatives of this fungal class may be 429 adapted to mineral environment with potential capacities of mineral weathering. Together, our 430 results highlight that no specific fungal species is selected by the minerals whatever the tree 431 stand, but at the same time that the distribution of certain classes (Agaricales, Heliotiales, 432 Hypocreales) is relatively well conserved.

433

434 4.4 Mineral chemistry drives the structuration of the mineral-associated fungal
435 communities

A separate analysis, comparing only the three different minerals types and excluding the
BS samples, revealed significant differences between mineral-associated fungal communities.
Overall, Ascomycota and Zygomycota tended to be more enriched on calcite, while
20

439 Basidiomycota were more abundant on obsidian and apatite. Similar structuring effects of the 440 mineral type have also been reported for the fungal communities colonizing the surface of a 441 granite outcrop, where certain fungal ribotypes were specifically associated with K-felspar, 442 plagioclase or quartz [48,49]. Noticeably, quartz was shown to select particular taxa, though it 443 is a nutrient-poor mineral mainly composed of silica [48,49]. By contrast, other studies 444 reported that the mineral type has no effect on the structuration of the fungal communities [40, 43, 45]. These differences could be explained by the fact that previous studies used mesh 445 446 bags mainly filled with quartz particles and amended with only 1% of the mineral tested (*i.e.*, 447 apatite), while we filled mesh bags with 100% of the mineral tested. This is an important 448 difference as our conditions corresponded to an important hotspot of nutrients, while in the 449 previous conditions those nutritive elements were highly diluted in quartz. As quartz was 450 shown to have specific selective effect [48,49], the absence of difference observed in these 451 studies may be due to a major effect of quartz and a minor effect of the 1% amendment. The 452 differences in fungal assemblages observed on the surface of the three mineral types tested 453 here may be attributed to different intrinsic (i.e. chemical composition of the mineral 454 considered, their nutrient content, their weatherability) and extrinsic (i.e., physico-chemical 455 conditions of the surrounding soil) factors. In this way, the strong balance differences 456 between Ascomycota and Basidiomycota measured on calcite, in comparison to those 457 observed on apatite and obsidian, could be attributed to the nutritive content of calcite (*i.e.*, 458 Ca) and / or by the potential increase of the local pH due to carbonates released during the 459 dissolution of this mineral [11]. At the order level, calcite particles were enriched in specific 460 taxa belonging to the Hypocreales, Filobasidiales, Tremellales or Mortierellales. Notably, the 461 Hypocreales order encompass taxa frequently identified as endophytes in plant roots [83] and 462 which has been shown to facilitate the uptake of micronutrients by host plants and were

suggested to weather mineral substrates through the production of siderophores [84-86]. 463 464 Concerning the *Mortierellales* order, several species related to the genera *Mortierella* (*i.e.*, *M*. 465 horticola, M. alpine, M. longigemmata, M. gamsii and M. macrocystis) and Ilyonectria were 466 also enriched on calcite or apatite in our study. Notably, Mortierella possess functional 467 abilities related to organic matter decomposition [87] and mineral weathering [88,89], and 468 their presences in mesh bag amended with apatite has also been reported [40,44]. Together, 469 these results suggest that *Mortierella* species are well adapted to several environments, 470 including minerals, and are potentially capable of weathering minerals. As the representative 471 of this genus are culturable, it opens new perspectives in the understanding of their potential 472 role in nutrient cycling and mineral weathering.

473

474 *4.5 Contrasted responses of the different trophic guilds of fungi to the mineral habitat*

475 Fungal assemblage differences between the compartments (i.e., mineral vs BS) or the 476 mineral type (*i.e.*, calcite, obsidian and apatite) were also illustrated by trophic mode analysis 477 (FUNguild). Symbiotrophic fungi represented the most abundant group among the assigned 478 trophic modes detected in the soil horizon considered and on the mineral surfaces. However, 479 symbiotrophs represented a lower fraction on minerals than in the surrounding bulk soil, 480 especially on calcite (Figure 5b). Previous study demonstrated in vitro the selective 481 colonization of ECM fungi associated to Pinus sylvestris in patches of nutrient-rich potassium 482 feldspar in comparison with patches of nutrient-poor quartz [68]. Recently, Sun et al. [52] 483 reported that after 6-month incubation in the E soil horizon of a mixed forest, mesh bags filled 484 with apatite, serpentine, quartz or feldspar were dominantly colonized by symbiotrophic 485 fungi. Together these data evidence both *in vitro* and *in situ* the ability of symbiotrophic fungi 486 to explore their environment and their potential key role in nutrient cycling and mineral 487 weathering [90]. Additionally, saprotophic and pathotrophic fungi strongly increase in relative 488 abundance of ITS sequences in the mesh bags compared to the surrounding bulk soil, 489 suggesting an adaptation to the nutrient-poor conditions occurring in this mineral 490 environment. We cannot exclude that the incubation period (*i.e.*, 29 months) can be long 491 enough for accumulation of microbial necromass in the mesh bags due to the turnover of 492 ectomycorrhizal mycelia, resulting in a gradual increase in abundance of decomposer fungi 493 over time. It could also be suggested that the flux of dissolved organic matter produced when 494 percolating water passes through the forest floor could enter the mesh bag and thus represent 495 a source of energy and carbon for the microorganisms. In this sense, the presence of 496 heterotrophic bacteria isolated from the same mineral particles supports the presence of 497 available carbon substrates in the mesh bags [11]. However, the accumulation of organic 498 matter in the mesh bags was not assessed in the current study, precluding any conclusion on this point. Alternatively, the presence of saprotophic fungi on minerals may be due to their 499 500 mineral weathering ability. Indeed, some genera have been reported for their ability to 501 solubilize phosphate (Aspergillus, Penicillium, Trichoderma, Mucor, Candida, Discosia, 502 Eupenicillium, Gliocladium) [91-93]. Recently, Kirtzel et al. [94] revealed that the wood-503 rotting basidiomycota Schizophyllum commune was capable of increasing mineral weathering 504 through a combined effect of organic acids, siderophores, enzymes and mechanical pressure. 505 The patterns observed and the potential role of these different guilds in the mineral 506 weathering process may differ according to the soil horizons considered. Indeed, fungal 507 communities are known to strongly differ across the soil profile [37,72,95]. Notably, topsoil 508 horizons are considered to be inhabited by communities more strongly involved in organic 509 matter recycling, while deeper soil horizons due to their low content of carbon and organic 510 matter are inhabited by oligotrophic communities more effective at mobilizing nutrients from

other sources (i.e., minerals). Complementary studies considering mesh bags introduced in
different soil horizons are required to test this functional distribution.

513

514 **5. CONCLUSIONS**

515 Our study highlights that minerals represent a specific habitat for fungi (i.e., the 516 mineralosphere; [8]), where particular taxa are enriched. While the physico-chemical 517 properties of the minerals impact the community composition of the fungi, no effect was 518 observed for the fungal biomass development. Our analyses also evidenced that 519 environmental conditions (i.e., pH, land cover) strongly condition the colonization of minerals 520 by fungi. Indeed, although the work was done on the same soil in our study, the land cover 521 remains the first driver of fungal community. The mineral type effect observed in our study 522 was probably made possible by the soil horizon considered (i.e. topsoil) and the small-scale 523 sampling strategy used in our study, which was determined to correspond to the scale of 524 mineral/microorganisms interaction (i.e., µm or mm) and to limit the soil heterogeneity at 525 both abiotic (i.e., pH, nutrient content) and biotic levels. Interestingly, a mineral type effect 526 was also demonstrated for bacteria, but with a stronger effect, suggesting that minerals are 527 active interfaces to consider in soil biology [11,16,51]. The next step will be to increase the 528 sampling scale to integrate the soil heterogeneity and to determine whether the mineral type 529 effect is conserved whatever the soil depth. Our data revealed that symbiotrophic fungi are the 530 main colonizers of most of the mineral types tested, confirming the key role played by 531 mycorrhizal fungi in nutrient cycling, tree nutrition and potentially in mineral weathering. 532 However, specific patterns of colonisation of the minerals by fungi have been observed below 533 each tree stand, suggesting a strong impact of the soil chemistry and/or of the nutritive requirements of each tree species. Together the new data obtained in this study suggest a 534

535 contrasted response of the different guilds of fungi (*i.e.*, symbiotrophic, endophytic, 536 saprotrophic and pathotrophic fungi) according to the mineral type. A major objective will be 537 to investigate the role of the different guilds of fungi in mineral weathering. The culturable 538 fungi isolated from the Breuil-Chenue LTO will help to fill this gap. The impact of the soil 539 horizon considered on the distribution and functional role of these fungal communities will be 540 also a challenging question for future researches.

541

542 **5. ACKNOWLEDGMENTS**

543 This work was funded by the Laboratory of Excellence Arbre (ANR-11-LABX-0002-01) 544 in the frame of the multidisciplinary project INABACT and the ANR JC BACTOWEATHER 545 (ANR-11-JSV7-0001). Y. Colin is a postdoctoral scientist supported by grants from ANR JC, 546 ANDRA, and Lorraine Region. We thank A. Legout and J. Ranger for helpful comments and 547 for giving us access to the site of Breuil-Chenue. We thank C. Rose from the certified facility 548 in Functional Ecology (PTEF OC 081) from UMR 1137 EEF and UR 1138 BEF for imaging 549 mineral particles. The UMR1136 and UR1138 as well as the PTEF platform are supported by 550 the French Agency through the Laboratory of Excellence Arbre (ANR-11-LABX-0002-01). 551 We thank C. Calvaruso and C. Cochet, C. Pantigny and M. Beuret for helpful discussion and 552 for technical assistance.

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Figure legends

Figure 1. Relative of mass loss of the mineral particles recovered in the mesh bags, after 29 months of incubation in the different tree stands of the Breuil-Chenue forest site. Abbreviations were used as follows: B, Beech; P, Pine and C, Coppice with standards. For each mineral type (apatite, obsidian, calcite), the percentage of mass loss was determined and statistically compared between the different tree stands through an ANOVA test. Each value presented is the mean \pm the standard error of the mean of four replicates. Different letters indicate significant differences (P < 0.05).

Figure 2. Changes in fungal population size (A) and Shannon diversity (B) between bulk soil and mineral samples below the different tree stands. Fungal population size was estimated based on 18SrRNA qPCR and data are presented as log-transformed [18S rRNA gene copies per gram of soil or mineral]. Shannon index was calculated from the sequencing data for the total fungal community. For both estimates, letters above the boxplots denote significantly different groups from Tukey's post-hoc tests (P<0.05). Boxplots indicate the median, interquartile range, reasonable range of the data and outliers.

906 Figure 3: Photographic and microscopy (SEM) images of mineral particles recovered 907 in the mesh bags, after 29 months of incubation in the different tree stands of the Breuil-908 **Chenue forest site.** A) White fungal hyphae colonizing apatite particles under Coppice stand; 909 B) Black fungal hyphae colonizing obsidian particles under Coppice stand; C) Direct 910 observation of an obsidian particle incubated in the Beech stand; D) Direct observation of an 911 apatite particle incubated in the Beech stand E) Direct observation of an obsidian particle 912 incubated in the Coppice stand; F) Direct observation of an obsidian particle incubated in the 913 Corsican pine stand. The red arrows indicate fungal hyphae. The mineral particles were 914 directly observed without any preparation.

Figure 4: Non-metric multidimensional scaling (nMDS) plot of fungal community structure from bulk soil and mineral samples. Ordination plot is based on based on *Bray-Curtis (Hellinger*-transformed) dissimilarities. Symbols represent the different sample types (bulk soil, calcite, apatite and obsidian) and colors reflect the different tree stands (beech: red, coppice with standards: green and Corsican pine: blue). Ellipses represent 95% confidence intervals of centroids. Figure 5: Relative distribution of the major phyla (A) and fungal trophic modes (B) detected in bulk soil and mineral samples below the different tree stands. The values presented are average values from the replicates. The category 'other' groups minor phyla or minor trophic modes, while the 'unassigned' refer to the sequences only affiliated as fungi or without known trophic mode.





















Table 1: Relative distribution of the main fungal trophic modes detected A) in the bulk soil and mineral compartments below each tree stand or B) in the different mineral types below each tree stand. The data presented are relative abundances (mean \pm standard deviation) based on ITS sequence distribution across the samples. Significant differences between bulk soil and mineral samples are indicated are presented by asterisks ('*' means the p-value < 0.05, '**' means the p-value < 0.01, '***' means the p-value < 0.001). The term 'others' refers to the other minor trophic modes proposed by Funguild. Comparison symbols (< and >) were used to describe significant differences between calcite (c), apatite (a) and obsidian (o) samples.

A)

				beech					C	Coppice	with st	tand	dards		Corsican Pine									
	bul	k se	oil	minerals				bu	lk s	oil	minerals				bu	oil	mi							
Trophic guilds	mean ± sd			mean	±	sd	pval	mean	±	sd	mean	±	sd	pval	mean	±	sd	mean	±	sd	pval			
Symbiotroph	84.46	±	1.68	40.01	±	18.22	**	64.57	±	12.88	46.86	±	32.77	0.391	73.78	±	11.39	31.50	±	24.23	*			
Pathotroph	0.94	±	0.24	2.33	±	2.34	0.335	1.28	±	0.49	7.12	±	9.20	0.309	0.94	±	0.31	4.75	±	6.47	0.339			
Saprotroph	3.46	±	0.79	8.79	±	6.06	0.163	9.51	±	5.02	8.59	±	6.86	0.835	8.07	±	3.78	23.83	±	7.82	**			
Pathotroph/Saprotroph	0.59	±	0.12	1.24	±	1.13	0.351	1.43	±	0.70	0.93	±	1.03	0.454	0.67	±	0.24	4.53	±	3.40	0.078			
Others	1.80	±	0.98	1.25	±	1.23	0.49	1.88	±	0.77	2.24	±	2.35	0.801	1.76	±	0.57	1.47	±	0.67	0.497			
Unassigned	8.75	±	1.83	46.38	±	12.73	***	21.33	±	10.49	34.26	±	18.82	0.288	14.78	±	7.08	33.93	±	16.84	0.081			

B)

	Beech											Coppice with standards												Corsican pine									
	C	alcit	te	apatite			obsidian				calcite			apatite			obsidian				calcite			apatite			obsidian						
	mean ± sd			mean	±	sd	mean	±	sd	pval	mean	±	sd	mean	±	sd	mean	±	sd	pval	mean	±	sd	mean	±	sd	mean	±	sd	pval			
Symbiotroph	25.10	±	23.44	51.35	±	10.00	43.58	±	9.09	0.102	9.37	±	5.99	60.49	±	34.72	64.76	±	19.36	0>C***	10.81	±	8.70	47.37	±	27.03	36.32	±	20.19	0.076			
Pathotroph	4.89	±	2.37	1.21	±	0.97	0.90	±	0.60	c>a=o*	20.33	±	2.12	1.96	±	1.33	1.09	±	0.10	c>a=o**	9.68	±	9.90	0.92	±	0.40	3.65	±	1.38	0.145			
Saprotroph	14.94	±	7.29	6.92	±	0.53	4.51	±	1.36	C>0**	16.38	±	1.68	7.08	±	8.31	3.88	±	0.95	c>0*	21.12	±	2.10	27.50	±	11.50	22.85	±	7.66	0.536			
Pathotroph.Saprotroph	2.37	±	1.41	0.73	±	0.27	0.62	±	0.11	c>0*	2.36	±	0.57	0.30	±	0.19	0.33	±	0.14	c>a=o*	7.48	±	1.21	1.38	±	1.36	4.71	±	3.76	c>a***			
Unassigned	50.53	±	17.57	38.87	±	10.34	49.75	±	8.09	0.386	49.23	±	3.25	26.39	±	22.01	28.93	±	20.20	0.280	48.87	±	7.43	21.82	±	18.63	31.10	±	11.55	0.051			
Others	2.18	±	1.92	0.92	±	0.09	0.64	±	0.32	0.179	2.34	±	0.86	3.77	±	4.12	1.01	±	0.64	0.342	2.03	±	0.78	1.00	±	0.12	1.36	±	0.54	0.071			