



# Ultrafast reproducible synthesis of a Ag-nanocluster@MOF composite and its superior visible-photocatalytic activity in batch and in continuous flow

Ana Arenas-Vivo, Sara Rojas, Iván Ocaña, Ana Torres, Marta Liras, Fabrice Salles, Daniel Arenas-Esteban, Sara Bals, David Ávila, Patricia Horcajada

## ► To cite this version:

Ana Arenas-Vivo, Sara Rojas, Iván Ocaña, Ana Torres, Marta Liras, et al.. Ultrafast reproducible synthesis of a Ag-nanocluster@MOF composite and its superior visible-photocatalytic activity in batch and in continuous flow. *Journal of Materials Chemistry A*, 2021, 9 (28), pp.15704-15713. 10.1039/D1TA02251B . hal-03292627v2

**HAL Id: hal-03292627**

**<https://hal.science/hal-03292627v2>**

Submitted on 23 Jul 2021

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

# Ultrafast reproducible synthesis of a Ag-nanocluster@MOF composite and its superior visible-photocatalytic activity in batch and in continuous Flow

Ana Arenas-Vivo,<sup>a</sup> Sara Rojas,<sup>a</sup> Iván Ocaña,<sup>a</sup> Ana Torres,<sup>a</sup> Marta Liras,<sup>b,\*</sup> Fabrice Salles,<sup>c</sup> Daniel Arenas-Esteban,<sup>d</sup> Sara Bals,<sup>d</sup> David Ávila,<sup>e</sup> Patricia Horcajada<sup>a,\*</sup>

The (photo)catalytical properties of Metal-Organic Frameworks (MOFs) can be enhanced by postsynthetic inclusion of metallic species in their porosity. Due to their extraordinarily high surface area and well defined porous structure, MOFs can be used for the stabilization of metal nanoparticles with adjustable size within their porosity. Originally, we present here an optimized ultrafast photoreduction protocol for the *in situ* synthesis of tiny and monodispersed silver nanoclusters (AgNCs) homogeneously supported on the photoactive porous titanium carboxylate MIL-125-NH<sub>2</sub> MOF. The strong metal-framework interaction between -NH<sub>2</sub> and Ag clusters influences the AgNCs growth, leading to the surfactant free efficient catalyst AgNC@MIL-125-NH<sub>2</sub> with improved visible light absorption. The potential use of AgNC@MIL-125-NH<sub>2</sub> was further tested in challenging applications: *i)* the photodegradation of the emerging organic contaminants (EOCs) methylene blue (MB-dye) and sulfamethazine (SMT-antibiotic) in water treatment, and *ii)* the catalytic hydrogenation of *p*-nitroaniline (4-NA) to *p*-phenylenediamine (PPD) with industrial interest. It is noteworthy that compared with the pristine MIL-125-NH<sub>2</sub>, the composite presents an improved catalytic activity and stability, being able to photodegrade 92% of MB in 60 min, 96% of SMT in 30 min, and transform 100% of 4-NA to PPD in 30 min. Aside from these very good results, this study describes for the first time the use of a MOF in a visible light continuous flow reactor for wastewater treatment. With only 10 mg of AgNC@MIL-125-NH<sub>2</sub>, a high SMT removal efficiency over 70% is maintained after >2 h using water flow conditions found in real wastewater treatment plants, envisioning a future real application of MOFs in water remediation.

## Introduction

In the last years, water pollution is becoming a major concern due to novel and dangerous anthropogenic pollutants.<sup>1</sup> Particularly, emerging organic contaminants (EOCs) are found as hundreds contaminating water resources and include a wide different array of compounds: pharmaceuticals and personal care products (PPCPs), dyes, pesticides, veterinary products, industrial products, and engineered nanomaterials.<sup>2</sup> EOCs enter the environment from a large variety of sources and pathways (*e.g.*, wastewater treatment plants (WWTPs), hospital effluents, livestock activities, and industrial waste),<sup>3</sup> leading to concentrations in surface waters (even in tap drinking water) capable of causing detrimental effects to living organisms (from ng·L<sup>-1</sup> to µg·L<sup>-1</sup>).<sup>4,5</sup> It is therefore clear that currently used technologies are not equipped to remove them and more effective methods need to be developed.

Among the new processes to eliminate EOCs from water (*i.e.*, ozonization,<sup>6</sup> chlorination,<sup>7</sup> sonodegradation,<sup>8</sup> biodegradation,<sup>9</sup> catalysis,<sup>10</sup> activated carbon treatment<sup>11</sup>), photocatalysis shows great potential since it could be performed by sunlight illumination at ambient conditions. Metal oxide and chalcogenides semiconductors (*e.g.*, TiO<sub>2</sub>, ZnO, Fe<sub>2</sub>O<sub>3</sub>, CdS, GaP, ZnS) are promising heterogeneous photocatalysts, but their large scale applications are limited due to difficult post-separation and low energy conversion efficiency.<sup>12</sup>

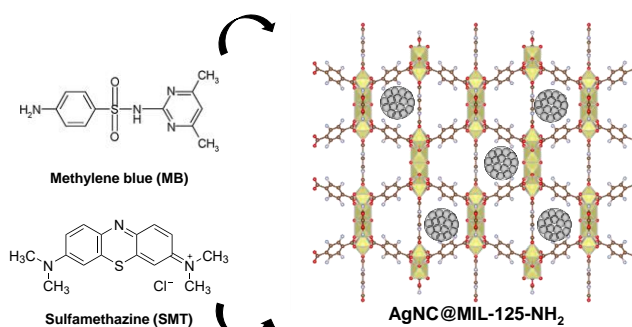
Metal-organic frameworks (MOFs), or crystalline porous coordination polymers,<sup>13</sup> outstand among other proposed materials for environmental remediation by photocatalysis, as they can act as both selective adsorbents of contaminants and catalysts, playing an active role in their degradation. They present many properties that commend them particularly in water treatment: *i)* large porosity (up to Brunauer–Emmett–Teller surface area ( $S_{\text{BET}}$ ) = 7800 m<sup>2</sup>·g<sup>-1</sup>),<sup>14,15</sup> associated with high sorption capacities; *ii)* active sites, where contaminants can be chemisorbed and/or degraded; *iii)* some are highly stable in water; *iv)* easily functionalizable cavities, where specific host-guest interactions may take place; *v)* opportunities for tailoring the MOF structure to allow shape selective adsorption and/or catalysis; *vi)* MOFs could be synthesized at large scale (few already commercialized);<sup>16</sup> and *vii)* they can be shaped into practical devices adapted for decontamination (*e.g.*, pellets, membranes, columns).<sup>17</sup> MOFs have already proven remarkable performances on the removal (adsorption and/or catalytic degradation) of EOCs from water. However, the majority of these studies concerns organic dyes (70% of the reports) being PPCPs elimination an emerging topic, only studied from end-2012 (bibliographic estimation by WOS February 2021).<sup>18</sup> Which is more, studies presented up to date are limited to only batch type reactors and investigation of MOFs photocatalysts under continuous flow is still a remaining subject of analysis.<sup>19,20</sup>

Despite the advantages of MOFs for the catalytic degradation of EOCs, as indicated above, the type of active sites on MOFs is mainly limited to unsaturated metal centers, the creation of structural defects around metals, or active sites grafted on the organic linker. However, MOFs can also incorporate catalytically active guest species into their pore spaces. Integration of MOFs and noble metal ions or nanoparticles (NPs) (*e.g.*, Pt, Pd, Ag, Au) to form new multifunctional composites has become a rapidly developed research area. Their incorporation can result in an enhancement on the photoconversion quantum yield and allows the extension of the absorption band gap to visible light, enabling water remediation with sunlight.<sup>21,22</sup> The growing number of research articles on this topic indicates that NPs@MOF nanocomposites will play an important role in heterogeneous photocatalysis.<sup>23</sup> Particularly in water decontamination, some outstanding results have been reported in the use of MIL-100(Fe)@TiO<sub>2</sub> in the methylene blue (MB) degradation, reaching 45 times highest rate of MB degradation compared to TiO<sub>2</sub>;<sup>24</sup> or MIL-125-NH<sub>2</sub>@Ag/AgCl in the degradation of rhodamine B, with an 80% of degradation in only 10 min.<sup>25</sup>

Among other metallic nanospecies of great interest are silver nanoclusters (AgNCs) due to their unique electronic properties,<sup>26</sup> as they present high catalytic efficiency and fluorescence resonance energy transfer;<sup>27</sup> tiny size (up to few nanometers);<sup>28</sup> low toxicity (e.g., at 250  $\mu\text{M}$ , cell viability is ca. 98% in human neonatal foreskin fibroblast cells),<sup>29,30</sup> and they are considered a less expensive alternative to AgNPs. Few reports have analyzed the use of MOFs as scaffolds for the growth and stabilization of AgNCs without the need of surfactants. However, long chemical reduction procedures with several steps have been followed, as in the complex infiltration procedures in the synthesis of Ag clusters (AgCs) within MIL-68, HKUST-1, MIL-101 or MOF-5,<sup>31–34</sup> ruling out their potential in real applications, like water remediation. Even UV photoreduction was previously used in the synthesis of Ag/rGO/MIL-125 composite, it required 2 h of irradiation to produce large AgNPs (5–15 nm) that were adsorbed on both, the reduced graphene oxide layers and the MOF outer surface.<sup>35</sup>

The work presented here describes an optimized ultrafast photoreduction protocol innovatively used for the formation of small AgNCs supported in the porosity of the photoactive Ti-MOF MIL-125-NH<sub>2</sub>. MIL-125-NH<sub>2</sub> or [Ti<sub>8</sub>O<sub>8</sub>(OH)<sub>4</sub>(BDC-NH<sub>2</sub>)<sub>6</sub>] is built from Ti-oxo-clusters coordinated to the 2-aminoterephthalate ligand (BDC-NH<sub>2</sub>). MIL-125-NH<sub>2</sub> has been selected as scaffold for the synthesis of AgNCs considering different factors: *i*) it represents a prototype and extensively studied carboxylate MOF family, *ii*) in terms of cost, it can be considered affordable in a future (already commercialized by different companies),<sup>36,37</sup> *iii*) its large surface area and high pore volume ( $S_{\text{BET}} \sim 1500 \text{ m}^2 \cdot \text{g}^{-1}$ ,  $V_p \sim 0.65 \text{ cm}^3 \cdot \text{g}^{-1}$  with octahedral  $\sim 12.5 \text{ \AA}$  and tetrahedral cavities  $\sim 6 \text{ \AA}$ , accessible via windows  $\sim 5\text{--}7 \text{ \AA}$ ),<sup>38</sup> *iv*) the available amino groups, which may facilitate a strong metal-framework interaction between -NH<sub>2</sub> and Ag clusters, preventing AgNCs leaching and leading to the surfactant free efficient composite (AgNC@MIL-125-NH<sub>2</sub>), and *v*) its theoretical band gap energy ( $E_g$ ) of 2.6 eV, that enables semiconducting application under visible light, compatible with the degradation of EOCs using sunlight.<sup>39</sup> Aside from this innovative synthetic method, this study describes the pioneer use of a MOF in a visible light continuous flow reactor for wastewater treatment. If initially, we have considered the vis-photodegradation of the standard dye methylene blue (MB + MOF + water; February WOS  $\sim 600$  papers, see chemical structure in **Figure 1**) for comparison reasons, later the photodegradation of a challenging antibiotic (sulfamethazine; SMT, **Figure 1**) was originally studied using a continuous flow homemade photoreactor. SMT presents a wide occurrence and relatively high level in natural aquatic environment (from 0.12 to 19.2 ppm),<sup>3,40,41</sup> associated with severe environmental and health issues (human average daily intake  $<0.05 \text{ ppm}$  for preventing hormone unbalance, oral rat median lethal dose ( $\text{LD}_{50}$ ) =  $2 \text{ g} \cdot \text{Kg}^{-1}$ ).<sup>42</sup> In all described studies, the EOCs photodegradation efficiency of the AgNC@MIL-125-NH<sub>2</sub> was compared with the pristine MIL-125-NH<sub>2</sub> and other reported photocatalysts, resulting in both, an improved activity and stability. Finally, due to the interest of AgNCs stabilization for other organic catalytic reactions, the AgNC@MIL-125-NH<sub>2</sub> was tested in the catalytic hydrogenation of *p*-nitroaniline (4-NA) to *p*-phenylenediamine (PPD), an important intermediate for a variety of industrial chemicals, resulting in very high efficiencies.<sup>43</sup>

**Figure 1.** Schematic view of the structure of AgNP@MIL-125-NH<sub>2</sub> (titanium polyhedra, oxygen, nitrogen, and carbon are represented in yellow, red, blue, and grey, respectively; hydrogen atoms are omitted for clarity). Ag-NCs are represented by grey spheres. Structure of the dye MB and antibiotic SMT are also given.



## Results and discussion

### Synthesis and characterization of AgNC@MIL-125-NH<sub>2</sub>

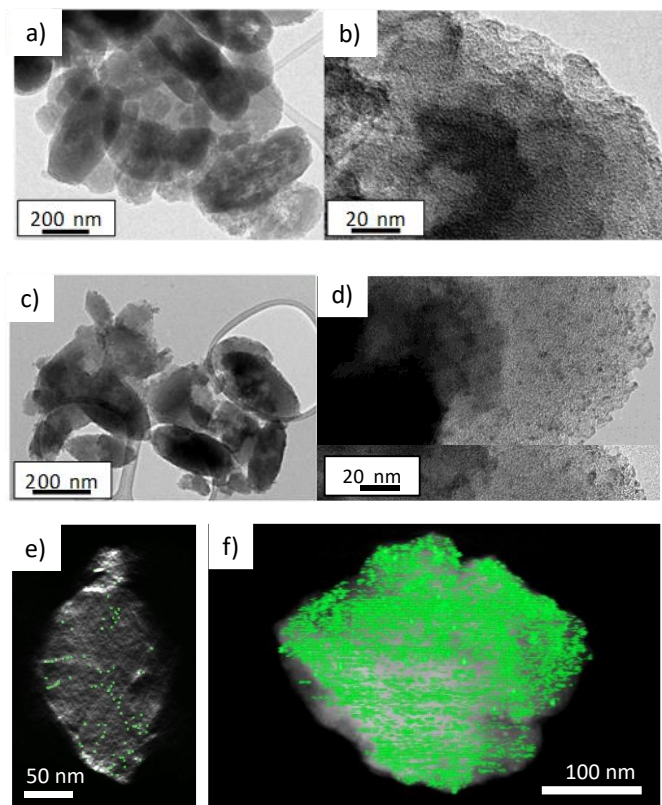
The synthesis of stable AgNCs of controlled size supported on MOFs porosity by means of photo-induced transformation was pioneeringly pursued. This photoreduction protocol might enhance overall composite (photo)catalytic activity while overcoming the long and intricate silver chemical reduction. Our inspiration come from Maretti *et al.*,<sup>44</sup> and Garcia-Bosch *et al.*,<sup>45</sup> works where fluorescent AgNCs were photosynthesized by means of Irgacure as photoinitiator, and stabilized by the presence of cyclohexane and a polymer with pendant -NH<sub>2</sub> groups, respectively. In our case, the visible-light-sensitive MIL-125-NH<sub>2</sub> was selected as the ideal scaffold as bearing -NH<sub>2</sub> functional groups potentially able to stabilize the AgNCs, aside from its porosity which could enable size precise control during the photoreduction reaction.

Briefly, AgNCs@MIL-125-NH<sub>2</sub> was prepared by an easy and fast two-step procedure (for details see **Supporting Information (SI) section 3**). First, the silver precursor (AgCF<sub>3</sub>COO) was impregnated within the MOF porosity prior diffusion (15 min). Then, the adsorbed silver was subsequently photoreduced in the presence of a photo-initiator (Irgacure-2959 or I-2959) in a controlled ultrafast way under stirring (15 s). To avoid the two main issues reported from bibliography (*i.e.*, preferential formation of AgNCs on the outer MOF surface, and polydispersity and heterogeneous distribution of AgNCs),<sup>46</sup> the design of the best synthetic strategy was optimized considering several parameters (irradiation time, reagents addition, stirring, MOF:Ag ratio; see transmission electron microscopy (TEM) images; **SI, section 3**). Discussing in brief their influence: *i*) *irradiation time*: excessive reaction time (15 s vs. 20 min or 1 h) leads to AgNCs aggregation through defect generation in the framework, procuring AgNPs larger than the pore

size, even coming out after 1 h of irradiation (**SI, Figure S5**); *ii) reagent addition*: a stepwise procedure (impregnation, then irradiation) leads to a better distribution of the AgNC through the MOF crystals than one-pot impregnation/irradiation as well as better control of particle size (**SI, Figure S6**); *iii) stirring* during the photoreduction step not only improve the sample homogeneity, associated with homogeneously distributed and monodispersed AgNCs, but also allows faster reaction kinetics (15 s vs. 1 min in presence and absence of stirring; **SI, Figure S7**); and *iv) MOF:Ag ratio*: enables the modulation of the amount of stabilized Ag and directly affects to reaction yield (see **SI, Figures S8 and Table S1**). In the end, a simple and efficient (yield >80% in only 20 min total process) two-step method was originally set up for obtaining tiny and monodispersed AgNCs ( $1.01 \pm 0.36$  nm, corresponding to  $\sim 28$  Ag atoms; **SI, Figure S9**) homogeneously distributed within the microporosity of MIL-125-NH<sub>2</sub> (**Figure 2**). Electron tomography studies were performed in order to ensure the location of the AgNC within the MOF particles. Based on a careful segmentation through the slices of the reconstruction (**Figure 2e and f, SI, Figure S10 and Movie 1**), the presence of the AgNCs within the MOF porous structure is confirmed. Moreover, many of the particles were found following same direction of the pores, thus supporting the porous structure during tilt series acquisition. Therefore, the synthetic procedure here presented ensured the formation of AgNC inside MIL-125-NH<sub>2</sub>, overcoming the limitations of previous works with MOFs with AgNPs exclusively adsorbed on their outer surface.<sup>31–35</sup> The ultrafast photoreduction reaction (15 s) prevented AgNCs uncontrolled growing and aggregation, further confined by MOF porosity (as AgNCs growth was limited within the size of MIL-125-NH<sub>2</sub> cavities of *ca.* 0.61 and 1.25 nm).<sup>38</sup> The synthetic procedure was followed with the non-functionalized MIL-125 material, obtaining bigger Ag NPs (*ca.* 7 nm), mainly located on the outer surface of the MOF particles (**SI, Figure S11**). These results support the effect of the -NH<sub>2</sub> group in the stabilization of the AgNCs. To the best of our knowledge, this is the first time tiny AgNCs have been synthesized by photoreduction using MOFs as support, and outstands from UV assisted silver photoreduction methods previously reported by means of reaction time, particle size and distribution within the porosity.<sup>35</sup> Further, images showed that the impregnation & photoreduction process was successful and barely aggressive with the MIL-125-NH<sub>2</sub> structure, which preserved its characteristic oval shape, with a relatively roughen surface. As the available -NH<sub>2</sub> seems to be the key, this ultrafast stabilization might be extrapolated to other photoactive microporous MOFs with available amino groups, such as the zirconium based UiO-66-NH<sub>2</sub>.

Further characterization by inductively coupled plasma-optical emission spectrometry (ICP-OES) evidenced that a significant amount of Ag was incorporated into the porous matrix ( $7.5 \pm 0.4$  wt.%), as also confirmed by thermogravimetric analyses (TGA; 7.7 wt.%, **SI, Figure S12**) and semi-quantitative energy dispersive X-ray spectroscopy (EDX; Ti:Ag molar ratio =  $7.0 \pm 0.5 : 3.0 \pm 0.5$ , or 21:1). Finally, the UV-vis solid reflectance spectroscopy revealed a significant change in AgNC@MIL-125-NH<sub>2</sub> light absorption, as reflectance increases beyond  $\sim 430$  nm compared to the pristine MIL-125-NH<sub>2</sub> (**SI, Figure S13**). This is associated with a reduction of the direct allowed transition energy band gap, being 2.40 and 2.53 eV for AgNC@MIL-125-NH<sub>2</sub> and MIL-125-NH<sub>2</sub>, respectively, due to the overlapping of the AgNC plasmon resonance band at  $\sim 450$  nm (values obtained by the Tauc Plot).<sup>47</sup> This last result suggests a strong association between the AgNCs and MIL-125-NH<sub>2</sub>, which might modify the photon absorption of the resulting composite.

The structural integrity of MIL-125-NH<sub>2</sub> after the AgNCs incorporation process was confirmed using X-ray powder diffraction (XRPD), without a noticeable decrease in the MOF crystallinity (**SI, Figure S14**). In addition, the main reflections of the face-centered cubic (fcc) structure of Ag<sup>0</sup> could not be seen, supporting the small size of the AgNCs. XRPD patterns further confirmed the formation of pure AgNC@MIL-125-NH<sub>2</sub> as no characteristic reflections of AgCF<sub>3</sub>COO neither the photoinitiator I-2959 are observed. Fourier transform infrared spectroscopy (FTIR) also corroborated that the structure remains stable after the impregnation & photoreduction process, and the proper removal of I-2959 (**SI, Figure S15 and S16**). The AgNC@MIL-125-NH<sub>2</sub> composite preserves the main vibrational modes of the hosting MOF such as coordination of carboxylates to the titanium ( $1550 - 1450$  cm<sup>-1</sup>).



**Figure 2.** TEM micrographs of the pristine MIL-125-NH<sub>2</sub> (a, b), and the AgNC@MIL-125-NH<sub>2</sub> composite (c, d). Orthoslice of the HAADF tomography reconstruction of one particle from the AgNC@MIL-125-NH<sub>2</sub> composite (e), and the 3D reconstruction where the AgNCs have been carefully segmented in green (f). Scale bar: (a, c) 200 nm, (b, d) 20 nm, (e) 50 nm, and (f) 100 nm.

Furthermore, a red shift in the bands associated to the symmetric and asymmetric -NH<sub>2</sub> stretching (from 3459 and 3356 cm<sup>-1</sup> in the pristine MOF, to 3442 and 3344 cm<sup>-1</sup> in the AgNC@MIL-125-NH<sub>2</sub>) suggests the stabilization of the AgNCs in the framework, and their interaction with the -NH<sub>2</sub> groups of the BDC-NH<sub>2</sub> ligand. This result is in agreement with previous FTIR modification, observed in the encapsulation of bigger AgNPs into the MIL-125-NH<sub>2</sub> solid.<sup>48</sup> This observation is also corroborated by the snapshots obtained from Monte Carlo simulations where AgNCs are distributed in the entire MOF porosity and the main interactions between Ag and MIL-125-NH<sub>2</sub> framework are observed with both the -NH<sub>2</sub> and benzene groups of the MIL-125-NH<sub>2</sub> (calculated distances = 2.8-2.9 Å, see SI, Figure S27a). Indeed, these large distances are mainly due to the absence of electrostatic interactions and only van der Waals interactions are taken into account in these calculations. The simulation results will be discussed in detail in the next section. Finally, the textural properties of both AgNC@MIL-125-NH<sub>2</sub> and the pristine MOF were evaluated by means of N<sub>2</sub> sorption at 77

K (SI, Figure S17). The specific surface area and pore volume of the composite were slightly lower than the pristine MOF ( $S_{\text{BET}} \sim 1430$  vs. 1560 m<sup>2</sup>·g<sup>-1</sup>;  $V_p \sim 0.61$  vs. 0.65 cm<sup>3</sup>·g<sup>-1</sup>, respectively), in agreement with the presence of AgNC in the composite. Hence the presence of the AgNCs within the cavities of MIL-125-NH<sub>2</sub> may explain these differences. Moreover, molecular simulation of AgNCs insertion gave reasonably similar textural properties for the composite ( $S_{\text{BET}} \sim 1390$  vs. 1925 m<sup>2</sup>·g<sup>-1</sup>;  $V_p \sim 0.63$  vs. 0.81 cm<sup>3</sup>·g<sup>-1</sup>, for AgNC@MIL-125-NH<sub>2</sub> and MIL-125-NH<sub>2</sub>, respectively). Differences between experimental and calculated textural properties of pristine MIL-125-NH<sub>2</sub> are due to the absence of defects and perfect porosity evacuation when simulating the crystalline structure of MIL-125-NH<sub>2</sub>.

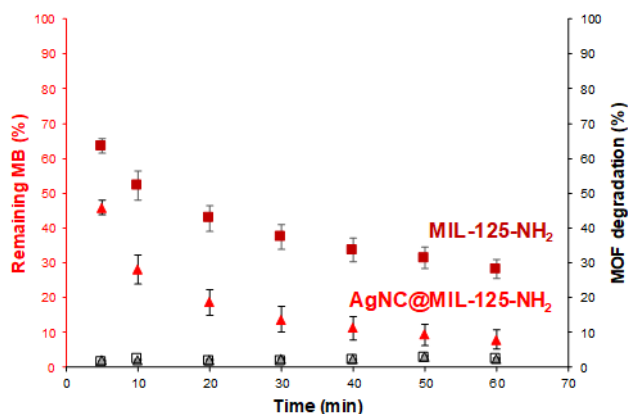
#### Photodegradation of water emerging pollutants

To evaluate the ability of AgNC@MIL-125-NH<sub>2</sub> in water purification, the powdered material was first suspended in a tap water solution of the selected contaminant, using a concentration range within that one commonly found in wastewater (200 ppm of MB and 10 ppm of SMT).<sup>41,49</sup> AgNC@MIL-125-NH<sub>2</sub> photocatalytic activity towards the degradation of EOCs was investigated under visible light irradiation, quantifying the contaminant elimination. In addition, AgNC@MIL-125-NH<sub>2</sub> stability was monitored during the EOCs photodegradation process (SI, Section 4).

#### Methylene blue (MB)

Remarkably, AgNC@MIL-125-NH<sub>2</sub> was able to photodegrade up to 72 ± 4 and 92 ± 3% of MB present in the solution within only 10 and 60 min, respectively (Figure 3, red triangles). The pristine MIL-125-NH<sub>2</sub> displays a slower kinetics, reaching 48 ± 1 and 72 ± 2% of MB photodegradation in 10 and 60 min, respectively (Figure 3, dark red squares). The rapid MB elimination kinetics is most likely attributed to: *i*) the lower energy band gap induced by the AgNC plasmon resonance (2.40 vs. 2.53 eV in AgNC@MIL-125-NH<sub>2</sub> and MIL-125-NH<sub>2</sub>, respectively); *ii*) the reduction of the electron hole recombination as AgNC work as electron trapping agent;<sup>50</sup> and *iii*) the chemical affinity of Ag to the -S and -N atoms of MB structure.<sup>51</sup> This chemical affinity of MB towards the AgNC@MIL-125-NH<sub>2</sub> is also seen in the experiments carried out in dark (SI, Figure S18), where MB physisorption trends are presented. After 60 min in dark, AgNC@MIL-125-NH<sub>2</sub> adsorbed 70% of the dye while the pristine MOF only removed 40% of the MB from water (comparatively, the same photodegraded amount in 10 min, Figure 3). This interaction between the dye and the catalyst framework likely affects to its photodegradation under visible light.

To shed some light on the MB degradation kinetics and to gain further understanding on the involved mechanism, the MB degradation kinetics using AgNC@MIL-125-NH<sub>2</sub> and MIL-125-NH<sub>2</sub> were fitted to a second order kinetics according to Eqn. (1) (SI, Section 4a), giving rise to  $k$  values of 0.0007 and 0.0001 g·mg<sup>-1</sup>·min<sup>-1</sup>, respectively. It can be seen that AgNC@MIL-125-NH<sub>2</sub> is a better photocatalyst as its kinetic constant is 7 times higher than the pristine MOF (SI, Figure S19).



**Figure 3.** MB photodegradation kinetics under visible light irradiation in the presence of AgNC@MIL-125-NH<sub>2</sub> (red triangles, left y axis) and MIL-125-NH<sub>2</sub> (dark red squares, left y axis). The MOFs degradation by means of BDC-NH<sub>2</sub> leaching was also represented for AgNC@MIL-125-NH<sub>2</sub> (white squares, right y axis), and MIL-125-NH<sub>2</sub> (grey triangles, right y axis).

These fittings also reveal that the rate-limiting step of the photodegradation is not the physisorption of the dye on the MOF porous surface, but the photocatalytic reaction on the AgNCs active site, after MOF photon absorption by the ligand-to-cluster transfer mechanism.<sup>52,53</sup>

Further, both solids showed an excellent stability since only ~2% of the total BDC-NH<sub>2</sub> ligand was leached after 1 h contact (as

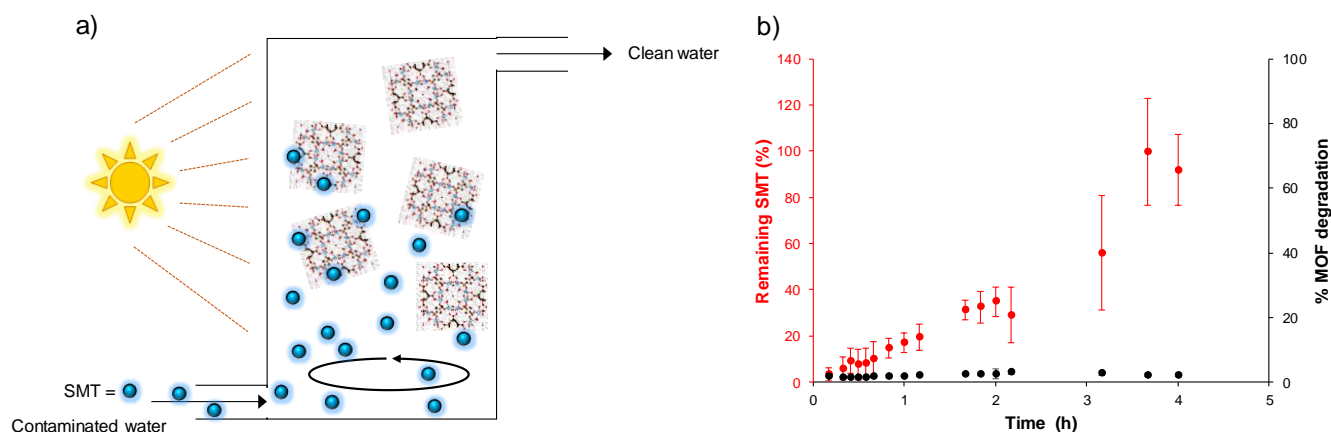
determined by high-performance liquid chromatography-HPLC). Besides, XRPD of both materials suggest the retainment of the crystalline character of the parenteral MIL-125-NH<sub>2</sub>, with a slight peak broadening associated with the MB adsorption (**SI, Figure S20**). The crystalline structure of AgNC@MIL-125-NH<sub>2</sub> was not almost disturbed after photodegradation of MB, suggesting the stabilization of the MIL-125-NH<sub>2</sub> scaffold by a potential “template” effect of the encapsulated AgNCs on its structure, reinforcing the composite. Note also here that upon TEM observations, we have observed a MIL-125-NH<sub>2</sub> stabilization effect under the beam by the presence of AgNCs. This increment in stability was also reported in the Pt@UiO-66-NH<sub>2</sub> composite after photocatalytic H<sub>2</sub> production.<sup>54</sup> These stability analyses are fundamental in water remediation. Despite, other MB photodegradation studies using MOFs do not often report these relevant data (**SI, Table S2**).

To deeply analyze the influence of AgNCs, the catalytic performances of AgNC@MIL-125-NH<sub>2</sub> were compared with its analogue AgNPs@MIL-125-NH<sub>2</sub>, based on an equivalent amount of AgNPs with a larger size (5 vs. 1 nm), recently reported by our group (see experimental detail in the **SI, Section S4**).<sup>48</sup> Although AgNPs@MIL-125-NH<sub>2</sub> also provided a fast MB degradation at short times (84 ± 4% after 1 h; **SI Figure S21**), the kinetics is slightly faster when using AgNC@MIL-125-NH<sub>2</sub>. TOF calculation (min<sup>-1</sup>) also revealed this faint increment in activity (**SI, Figure S22**) evidencing the relevance of the higher reactive surface area of the AgNCs vs. AgNPs (~1.25 times higher surface area with AgNCs) and the size dependent response of the AgNC surface plasmon resonance for the photocatalytic reaction, since both materials possess the same Ag content. Previous bibliography considers this beneficial catalytic effect as a consequence of the variation of the proportion of accessible active sites.<sup>55</sup> The number of atoms on the surface and in the edges differs with catalyst shape and size, and the number of low coordinated atoms increases when the catalyst size decreases, incrementing the reaction rate. We can tentatively estimate that almost all the 28 atoms of the AgNCs are on the surface (against the *ca.* 25% for the AgNPs).<sup>56</sup> However, there are some limitations about determining the particle size effects in catalysis, since the catalyst’ structure (shape) and even the MOF’ structure (crystallinity, density of defects), and the different accommodation and distribution of the catalyst on the host MOF support may be changed during the synthesis. Our results here agree with diverse studies using AgNPs as heterogeneous catalyst in the hydrogenation of nitroaromatics or CO oxidation, in which the catalyst activity increased with the decrease in Ag size.<sup>57,58</sup> Finally, to further contextualize these results, AgNC@MIL-125-NH<sub>2</sub> showed a MB photodegradation capacity much higher than the widely-known commercial TiO<sub>2</sub>, catalytically limited under visible light (18 ± 1%, 1 h, **SI, Figure S21**). Furthermore, the MB photodegradation capacity of AgNC@MIL-125-NH<sub>2</sub> is on the range, or even surpasses, the MB degradation capacity of other known MOFs and MOF composites (90% degradation in 60 min, **SI, Table S2**). Therefore, this work paves the way of the use of AgNC@MIL-125-NH<sub>2</sub> for the elimination of dyes as EOCs in wastewater due to its high photoactivity and aqueous stability.

### Sulfamethazine (SMT)

The excellent MB removal capacity and stability of AgNC@MIL-125-NH<sub>2</sub> prompted us to further investigate the elimination of a challenging EOC, the antibiotic SMT, under more real continuous flow conditions. First, we studied the kinetics of the SMT photodegradation process under static conditions, reaching *ca.* 96.6 ± 0.2% of SMT removal in only 30 min (**SI, Figure S23**) while keeping intact its crystalline structure and chemical integrity (**SI, Figures S23 and S24**; 5 ± 2% degradation). The integrity of AgNCs was also confirmed by EDX analysis (Ti:Ag molar ratio before and after catalysis = 7.0 ± 0.5 : 3.0 ± 0.5 vs. 7.0 ± 0.1 : 2.6 ± 0.2, respectively). As previously obtained in MB degradation, the pristine MIL-125-NH<sub>2</sub> displays a slower kinetics than its AgNC counterpart, with almost half of the SMT degradation. Again, the SMT photodegradation data were fitted to a second order kinetics according to Eqn. (1) (**SI, Section 4b**), giving rise to *k* values of 3.56 and 0.02 g·mg<sup>-1</sup>·min<sup>-1</sup> for AgNC@MIL-125-NH<sub>2</sub> and MIL-125-NH<sub>2</sub>, respectively. AgNC@MIL-125-NH<sub>2</sub> exhibits a *ca.* 200 times faster photodegradation kinetics than the pristine MOF (**SI, Figure S25**), being even faster than the three MOF-based materials reporting SMT degradation (**SI, Table S3**).





**Figure 4.** a) Schematic view of the AgNC@MIL-125-NH<sub>2</sub> based continuous flow system. b) Effect of time on the remaining SMT (red circles, left y axis) using AgNC@MIL-125-NH<sub>2</sub> reactor at a continuous flow of 30 mL·h<sup>-1</sup> of a 10 ppm SMT solution in water. The potential leaching of the ligand (black triangles, right x axis) was also monitored. All studies were carried out in triplicate.

Taking advantage of the efficient performance of AgNC@MIL-125-NH<sub>2</sub> in SMT photodegradation, for the first time a continuous flow photoreactor was developed as proof-of concept of the use of MOF-based materials for the efficient pollutant removal from water (**Figure 4a**; **SI, Section 4b**). Practically, a cylindrical glass photoreactor was designed and filled with AgNC@MIL-125-NH<sub>2</sub>, pumping a continuous flow of SMT-contaminated tap water. As shown in **Figure 4b**, a high removal efficiency over  $71 \pm 12\%$  was maintained after more than 2 h when using only 10 mg of catalyst under a flux of 30 mL·h<sup>-1</sup> or 80 L·m<sup>-2</sup>·h<sup>-1</sup>. It has to be pointed that the water flow applied during these experiments is within the range of the flow rate used in wastewater treatment plants (*i.e.*, 27 – 530000 L·m<sup>-2</sup>·h<sup>-1</sup>),<sup>59</sup> supporting the interest of AgNC@MIL-125-NH<sub>2</sub> as efficient photocatalyst in the practical removal of SMT. This flow is also comparable to other pure MOF-based systems, proposed for the EOCs removal by exclusively adsorption (100  $\mu\text{L}\cdot\text{min}^{-1}$  equivalent to 0.034 L·m<sup>-2</sup>·h<sup>-1</sup> for KOH@Ni<sub>8</sub>BDP<sub>6</sub>,<sup>20</sup> 0.05 mL·min<sup>-1</sup> equivalent to 58 L·m<sup>-2</sup>·h<sup>-1</sup> for ZIF-8,<sup>60</sup> and 2 mL·min<sup>-1</sup> equivalent to 764 L·m<sup>-2</sup>·h<sup>-1</sup> for MIL-101(Cr)).<sup>61</sup> However, the flow rate can be significantly increased when using different substrates as MOFs support (0.45 L·h<sup>-1</sup> for PES-SPES/MIL-53(Al)-NH<sub>2</sub>/PTFE without indicating the membrane surface,<sup>62</sup> 6.45 L·m<sup>-2</sup>·h<sup>-1</sup> for LacPAN-MIL-101-L,<sup>63</sup> 125.7 L·m<sup>-2</sup>·h<sup>-1</sup>·bar<sup>-1</sup> for UiO-66-NH<sub>2</sub>@MMM,<sup>64</sup> and 1000 L·m<sup>-2</sup>·h<sup>-1</sup> for UiO-66@wood).<sup>65</sup> This implies that our system, able to remarkably treat 6.50 L·h<sup>-1</sup>·g<sup>-1</sup> of contaminated water *via* visible light photocatalytic degradation, could be further improved by supporting the MOF on a specific substrate. The chemical and structural integrity of AgNC@MIL-125-NH<sub>2</sub> was confirmed ( $1.9 \pm 0.5\%$  of ligand leaching after 4 h; **Figure 4b** and **SI, Figure S26**).

In order to understand the impact of the presence of AgNCs in the MIL-125-NH<sub>2</sub> porosity on the adsorption of pollutants, Monte Carlo simulations were performed for pristine MIL-125-NH<sub>2</sub> and AgNC@MIL-125-NH<sub>2</sub> for the adsorption of SMT at low coverage (1 or 2 molecules *per* unit cell, **SI, Figure S27 and S28**). From these calculations, it was possible to extract the main adsorption sites and compare the behavior of the EOC guest hosted within the porosity. In the pristine MIL-125-NH<sub>2</sub>, SMT molecules strongly interacted with the MOF (theoretical adsorption enthalpy close to 127 kJ·mol<sup>-1</sup>). The main adsorption sites are evidenced by the extracted configurations from Monte Carlo: -NH<sub>2</sub> from SMT can interact by hydrogen bonds with the amino groups of the MIL-125-NH<sub>2</sub> ( $H_{\text{SMT}}\cdots H_{\text{MOF}} = 2.652 \text{ \AA}$ ;  $H_{\text{SMT}}\cdots N_{\text{MOF}} = 2.54 \text{ \AA}$ ), while secondary adsorption sites can also be observed between -NH<sub>2</sub> or -SO<sub>2</sub> groups from SMT and the carboxylate groups from the MOF ( $H_{\text{SMT}}\cdots O_{\text{carbMOF}} = 2.964 \text{ \AA}$ ;  $O_{\text{SMT}}\cdots C_{\text{carbMOF}} = 2.885 \text{ \AA}$ ) (**SI, Figure S27b**).

Regarding the AgNC@MIL-125-NH<sub>2</sub>, it could be seen that the SMT guest molecules were found in the large octahedral cavities ( $\sim 12.5 \text{ \AA}$ ) together with AgNCs (**SI, Figure S27c**). Furthermore, the main interaction sites are again associated to the formation of hydrogen bonds between the amino groups from the SMT and the MOF (with larger  $H_{\text{SMT}}\cdots N_{\text{MOF}}$  distances ranging between 2.6 and 2.7  $\text{\AA}$ ; as a typical illustration on **SI, Figure S27d**, we observe a 2.682  $\text{\AA}$  distance). In addition, AgNCs interact with the benzene rings of the SMT (distances 2.7-3.0  $\text{\AA}$ ). FTIR spectra of both SMT-loaded materials (MIL-125-NH<sub>2</sub>\_SMT and AgNC@MIL-125-NH<sub>2</sub>\_SMT, **SI, Figure S29**) confirm the configurations found by molecular simulation. A shift in the characteristic stretching  $\nu_{\text{asym}}(\text{N-H})$  and  $\nu_{\text{sym}}(\text{N-H})$  vibration bands of the amino group of the MOF (from 3360 to 3376 cm<sup>-1</sup> for the AgNC@MIL-125-NH<sub>2</sub>; and from 3349 and 3446 cm<sup>-1</sup> to 3386 and 3497 cm<sup>-1</sup> for the MIL-125-NH<sub>2</sub>) suggests the formation of interactions between these moieties. Finally, in the case of the pristine MIL-125-NH<sub>2</sub>, there is a shift of around 20 cm<sup>-1</sup> in the bands associated to the carboxylate groups found in the range of 1520-1650 cm<sup>-1</sup> and 1280 to 1400 cm<sup>-1</sup> for the  $\nu_{\text{asym}}$  and  $\nu_{\text{sym}}$ , respectively. This shift is not visible in the AgNC@MIL-125-NH<sub>2</sub>\_SMT solid, in total agreement with the simulation observations.

As a final point, a successful design of a water remediation continuous flow system requires prediction of the concentration-time profile or breakthrough curve for the effluent. Here, the efficiency of the system was assessed by the curve shape and the breakthrough time. The so-called Thomas model is commonly used in the modeling of fixed bed breakthrough curves in environmental sorption research (see **SI, Section 4** for further details).<sup>66</sup> From this equation, values describing the characteristic

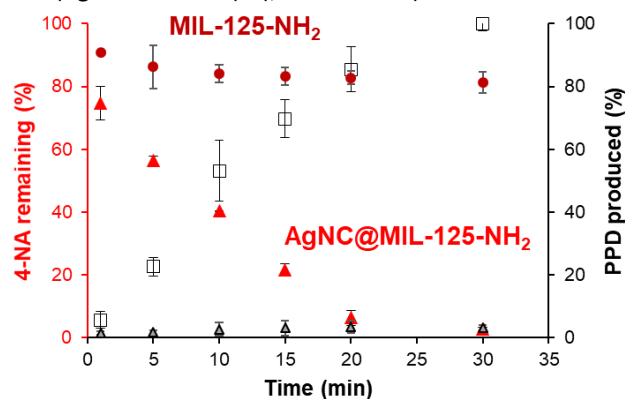
operational parameters of the column can be determined from a plot of  $C_t/C_0$  against  $t$  at a given bed height and flow rate using the non-linear regressive method. Thus, the Thomas model was found suitable for the description of whole breakthrough curve, with good correlation values ( $R^2 = 0.9624$ ; **SI, Figure S30**). This model, considering Langmuir kinetics of the degradation process and no axial dispersion, is derived with the assumption that the rate driving force obeys second-order reversible reaction kinetics,<sup>67</sup> giving a quite good kinetic constant ( $K_{Th}$ ) of  $3.49 \text{ mL mg}^{-1} \cdot \text{min}^{-1}$ . However, a direct comparison with other MOF-continuous flow systems is impossible since all the studies reported so far are exclusively based on sorption processes (mainly vapor phase) and not catalytic ones. In an attempt to position our results within the context of EOCs elimination from water, a comparison can be made with the kinetic constants obtained using MOFs or other adsorbent materials (*i.e.*, the adsorption of diazinon using MIL-100(Cr):  $K_{Th} = 5.11 - 8.33 \text{ mL} \cdot \text{min}^{-1} \cdot \text{mg}^{-1}$ ,<sup>61</sup> sulfamethazole with lignite activated coke:  $K_{Th} = 0.04 - 0.01 \text{ mL} \cdot \text{mg}^{-1} \cdot \text{min}^{-1}$ ,<sup>68</sup> or atenolol with activated carbon:  $K_{Th} = 2.28 \text{ mL} \cdot \text{mg}^{-1} \cdot \text{min}^{-1}$ ).<sup>69</sup> Finally, although this laboratory scale MOF photoreactor is still far from its application in wastewater treatments, our work presents the very first example of continuous-flow photodegradation of EOCs using MOFs with a considerably good breakthrough time and working under operation conditions on the range of those normally applied in wastewater treatment plants.

On the whole, the efficient removal of SMT and MB contaminants by AgNC@MIL-125-NH<sub>2</sub> together with its water robustness under real water continuous flow and vis-irradiation, make this composite a highly promising photocatalyst for the removal of emerging contaminants in water.

### Catalytic hydrogenation of 4-nitroaniline

AgNCs stabilization is not only of interest for the photodegradation of EOCs, but also for the catalysis of organic reactions at industrial scale. AgNPs have been previously reported for the hydrogenation of several nitroaromatic compounds.<sup>70,71</sup> In particular, *p*-nitroaniline (4-NA) is an important compound used as intermediate or precursor in the manufacture of organic synthesis, and it is used as target-model molecule in oxidation processes. In this context, heterogeneous catalysis using AgNCs supported on MOFs might be an interesting alternative for such processes, in terms of activity, lifetime and recyclability. In order to compare with other already described heterogeneous catalysts, the activity of AgNC@MIL-125-NH<sub>2</sub> towards catalytic hydrogenation of 4-NA to *p*-phenylenediamine (PPD) in the presence of NaBH<sub>4</sub> was studied, demonstrating its potential for further catalytic-related applications (**SI, Section 5**). As it can be seen from the **Figure 5**, complete hydrogenation of 4-NA is achieved at room temperature within 30 minutes when catalyzed by AgNC@MIL-125-NH<sub>2</sub>. AgNCs are the active catalytic sites, as the pristine MIL-125-NH<sub>2</sub> barely produces *ca.* 4% of the maximum theoretical PPD while consuming *ca.* 20% of 4-NA leading to the formation of undesired secondary products (which identification is out of the scope of this work).

Analysis of the reaction kinetics enabled the determination of the rate constant with a satisfactory fitting to a first order reaction kinetics following the Eqn (3) (**SI, Section 5**). The fitting of the data gave rise to a  $k$  value of  $0.119$  and  $0.002 \text{ min}^{-1}$  for AgNC@MIL-125-NH<sub>2</sub> and MIL-125-NH<sub>2</sub>, respectively (see **SI, Figure S33**). Results indicate that AgNC@MIL-125-NH<sub>2</sub> has a high catalytic activity towards 4-NA hydrogenation, within the same order of magnitude of previously reported works using other Ag functionalized MOFs (AgNC@MIL-101(Fe),  $k = 0.3 \text{ min}^{-1}$ ).<sup>72</sup>



**Figure 5.** 4-NA (2 mM) hydrogenation kinetics in the presence of AgNC@MIL-125-NH<sub>2</sub> (red triangles, left y axis) and MIL-125-NH<sub>2</sub> (b) (dark red squares, left y axis). The PPD formation in both experiments was also represented (white squares for AgNC@MIL-125-NH<sub>2</sub>, and MIL-125-NH<sub>2</sub> (grey triangles; right y axis).

### Conclusions

This work presents an ultrafast two-step photoreduction synthetic protocol for the stabilization of AgNCs within the MIL-125-NH<sub>2</sub> framework, with an average size of 1 nm (~28 Ag atoms) and homogeneously distributed along the MOF crystals. MIL-125-NH<sub>2</sub> framework prevents clusters aggregation and the strong metal-matrix interaction leads to a photoactive composite with a 2.4 eV energy band gap. This synthetic procedure can be in principle applied in other aminated MOFs in a way to achieve small metallic



Ag-clusters in their porosity. Further, the AgNC-MOF based photocatalyst here prepared demonstrates great potential for the visible light assisted degradation of challenging emerging organic contaminants in tap water, such as the dye MB and the antibiotic SMT. The photodegradation of contaminants under continuous flow using MOFs is performed here for the first time, enabling the purification of 6.50 L·h<sup>-1</sup>·g<sup>-1</sup> of SMT contaminated water *via* visible light photocatalytic degradation. Further, considering the results obtained in the 4-NA hydrogenation reaction, AgNC@MIL-125-NH<sub>2</sub> catalysts could be extended to many other industrially important aqueous heterogeneous reactions. By considering the final (real) working conditions (tap water, real contaminants concentration ranges found in natural waters, and continuous flow working conditions which are going to be use on the real scale), and the simultaneous study of potential leaching of the MOF building blocks (normally not considered), this work paves the way for the real future application of MOFs and MOF composites as photocatalysts for the elimination of challenging dyes, antibiotics and other widespread PPCPs in environmental remediation.

## Conflicts of interest

There are no conflicts to declare.

## Acknowledgements

The work has been supported by IMDEA Energy Foundation, BBVA Foundation (ref: IN[17]\_CBB\_QUI\_0197), Raphuel project (ENE2016-79608-C2-1-R, MINECOAEI/FEDER, UE) and MOFSEIDON project (PID2019-104228RB-I00, MCI/AEI/FEDER, UE). A.A.-V. and P.H. want to thank the financial support of Madrid Community (CAM PEJD-2016/IND-2828). S.R. and P.H. acknowledges the financial support of Madrid Community (Recruitment of Young doctors 2017, Talento Modality 2, 2017-T2/IND-5149). P.H. and M.L. acknowledges the Spanish Ramón y Cajal Programme (Grant Agreements 2014-15039 and 2015-18677, respectively). S.R. acknowledges the Spanish Juan de la Cierva Incorporación Fellowship (grant agreement no. IJC2019-038894-I). D. A. acknowledges the financial support through the Retos Project from MINECO with reference MAT2017-84385-R for funding.

## Notes and references

- 1 S. Zandaryaa, *A World Sci.*, 2011, **9**, 18–21.
- 2 I. Ali, M. Asim and T. A. Khan, *J. Environ. Manage.*, 2012, **113**, 170–183.
- 3 D. J. Lapworth, N. Baran, M. E. Stuart and R. S. Ward, *Environ. Pollut.*, 2012, **163**, 287–303.
- 4 H. B. Quesada, A. T. A. Baptista, L. F. Cusioli, D. Seibert, C. de Oliveira Bezerra and R. Bergamasco, *Chemosphere*, 2019, **222**, 766–780.
- 5 A. J. Ebele, M. Abou-Elwafa Abdallah and S. Harrad, *Emerg. Contam.*, 2017, **3**, 1–16.
- 6 M. Ibáñez, E. Gracia-Lor, L. Bijlsma, E. Morales, L. Pastor and F. Hernández, *J. Hazard. Mater.*, 2013, **260**, 389–398.
- 7 J. L. Acero, F. J. Benitez, F. J. Real, G. Roldan and E. Rodriguez, *Chem. Eng. J.*, 2013, **219**, 43–50.
- 8 E. A. Serna-Galvis, A. M. Botero-Coy, D. Martínez-Pachón, A. Moncayo-Lasso, M. Ibáñez, F. Hernández and R. A. Torres-Palma, *Water Res.*, 2019, **154**, 349–360.
- 9 N. H. Tran, T. Urase, H. H. Ngo, J. Hu and S. L. Ong, *Bioresour. Technol.*, 2013, **146**, 721–731.
- 10 N. Klammerth, N. Miranda, S. Malato, A. Agüera, A. R. Fernández-Alba, M. I. Maldonado and J. M. Coronado, *Catal. Today*, 2009, **144**, 124–130.
- 11 S. Wilhelm, A. Henneberg, H. R. Köhler, M. Rault, D. Richter, M. Scheurer, S. Suchail and R. Triebskorn, *Aquat. Toxicol.*, 2017, **192**, 184–197.
- 12 C. C. Wang, J. R. Li, X. L. Lv, Y. Q. Zhang and G. Guo, *Energy Environ. Sci.*, 2014, **7**, 2831–2867.
- 13 H. Li, M. Eddaoudi, M. O’Keeffe and O. M. Yaghi, *Nature*, 1999, **402**, 276–279.
- 14 O. K. Farha, I. Eryazici, N. C. Jeong, B. G. Hauser, C. E. Wilmer, A. A. Sarjeant, R. Q. Snurr, S. T. Nguyen, A. Ö. Yazaydin and J. T. Hupp, *J. Am. Chem. Soc.*, 2012, **134**, 15016–15021.
- 15 I. M. Hönicke, I. Senkovska, V. Bon, I. A. Baburin, N. Bönisch, S. Raschke, J. D. Evans and S. Kaskel, *Angew. Chemie Int. Ed.*, 2018, **57**, 13780–13783.
- 16 P. Silva, S. M. F. Vilela, J. P. C. Tomé and F. A. Almeida Paz, *Chem. Soc. Rev.*, 2015, **44**, 6774–6803.
- 17 U.-H. Lee, A. H. Valekar, Y. K. Hwang and J.-S. Chang, in *The Chemistry of Metal-Organic Frameworks: Synthesis, Characterization, and Applications*, 2016, pp. 551–572.
- 18 Z. Hasan, J. Jeon and S. H. Jhung, *J. Hazard. Mater.*, 2012, **209–210**, 151–157.
- 19 X. Zhang, J. Wang, X. X. Dong and Y. K. Lv, *Chemosphere*, 2020, **242**, 125144–125159.
- 20 S. Rojas, J. A. R. Navarro and P. Horcajada, *Dalt. Trans.*, 2021, **50**, 2493–2500.
- 21 E. Grabowska, A. Zaleska, S. Sorgues, M. Kunst, A. Etcheberry, C. Colbeau-Justin and H. Remita, *J. Phys. Chem. C*, 2013, **117**, 1955–1962.

- 22 X. Yu, L. Wang and S. M. Cohen, *CrystEngComm*, 2017, **19**, 4126–4136.
- 23 Y. Liu, Z. Liu, D. Huang, M. Cheng, G. Zeng, C. Lai, C. Zhang, C. Zhou, W. Wang, D. Jiang, H. Wang and B. Shao, *Coord. Chem. Rev.*, 2019, **388**, 63–78.
- 24 X. Liu, R. Dang, W. Dong, X. Huang, J. Tang, H. Gao and G. Wang, *Appl. Catal. B Environ.*, 2017, **209**, 506–513.
- 25 N. M. Mahmoodi, A. Taghizadeh, M. Taghizadeh and J. Abdi, *J. Hazard. Mater.*, 2019, **378**, 120741.
- 26 C. P. Joshi, M. S. Bootharaju and O. M. Bakr, *J. Phys. Chem. Lett.*, 2015, **6**, 3023–3035.
- 27 D. A. Islam, A. Chakraborty and H. Acharya, *New J. Chem.*, 2016, **40**, 6745–6751.
- 28 Q. M. Wang, Y. M. Lin and K. G. Liu, *Acc. Chem. Res.*, 2015, **48**, 1570–1579.
- 29 M. I. Setyawati, X. Yuan, J. Xie and D. T. Leong, *Biomaterials*, 2014, **35**, 6707–6715.
- 30 T. Soares, D. Ribeiro, C. Proença, R. C. Chisté, E. Fernandes and M. Freitas, *Life Sci.*, 2016, **145**, 247–254.
- 31 D. A. Islam, A. Chakraborty and H. Acharya, *New J. Chem.*, 2016, **40**, 6745–6751.
- 32 R. J. T. Houk, B. W. Jacobs, F. El Gabaly, N. N. Chang, A. A. Talin, D. D. Graham, S. D. House, I. M. Robertson and M. D. Allendorf, *Nano Lett.*, 2009, **9**, 3413–3418.
- 33 F. Su, Q. Jia, Z. Li, M. Wang, L. He, D. Peng, Y. Song, Z. Zhang and S. Fang, *Microporous Mesoporous Mater.*, 2019, **275**, 152–162.
- 34 B. Han, X. Hu, M. Yu, T. Peng, Y. Li and G. He, *RSC Adv.*, 2018, **8**, 22748–22754.
- 35 X. Yuan, H. Wang, Y. Wu, G. Zeng, X. Chen, L. Leng, Z. Wu and H. Li, *Appl. Organomet. Chem.*, 2016, **30**, 289–296.
- 36 AZO, No Title, <https://www.azom.com/article.aspx?ArticleID=16650>.
- 37 INSCX, Global Sales of Nanomaterials, <https://inscx.com/shop/product/titanium-metal-organic-framework-ti-nh2-mil-125-99-30-40um/>.
- 38 M. Dan-Hardi, C. Serre, T. Frot, L. Rozes, G. Maurin, C. Sanchez and G. Férey, *J. Am. Chem. Soc.*, 2009, **131**, 10857–10859.
- 39 M. De Miguel, F. Ragon, T. Devic, C. Serre, P. Horcajada and H. García, *ChemPhysChem*, 2012, **13**, 3651–3654.
- 40 S. Managaki, A. Murata, H. Takada, C. T. Bui and N. H. Chiem, *Environ. Sci. Technol.*, 2007, **41**, 8004–8010.
- 41 M. Petrovic and P. Verlicchi, *Contrib. to Sci.*, 2014, **10**, 135–150.
- 42 L. A. Poirier, D. R. Doerge, D. W. Gaylor, M. A. Miller, R. J. Lorentzen, D. A. Casciano, F. F. Kadlubar and B. A. Schweiz, *Regul. Toxicol. Pharmacol.*, 1999, **30**, 217–222.
- 43 H. Fiege, H.-W. Voges, T. Hamamoto, S. Umemura, T. Iwata, H. Miki, Y. Fujita, H.-J. Buysch, D. Garbe and W. Paulus, *Ullman's Encycl. Ind. Chem.*, 2012, **26**, 617–622.
- 44 L. Maretti, P. S. Billone, Y. Liu and J. C. Scaiano, *J. Am. Chem. Soc.*, 2009, **131**, 13972–13980.
- 45 N. García-Bosch, M. Liras, I. Quijada-Garrido and O. García, *RSC Adv.*, 2016, **6**, 67643–67650.
- 46 M. Meilikhov, K. Yusenko, D. Esken, S. Turner, G. Van Tendeloo and R. A. Fischer, *Eur. J. Inorg. Chem.*, 2010, 3701–3714.
- 47 B. D. Viezicke, S. Patel, B. E. Davis and D. P. Birnie, *Phys. status solidi*, 2015, **252**, 1700–1710.
- 48 A. Arenas-Vivo, G. Amariei, S. Aguado, R. Rosal and P. Horcajada, *Acta Biomater.*, 2019, **97**, 490–500.
- 49 G. Nibret, S. Ahmad, D. G. Rao, I. Ahmad, M. A. M. U. Shaikh and Z. U. Rehman, in *SSRN Electronic Journal*, Elsevier BV, 2019, pp. 1959–1969.
- 50 L. Collado, A. Reynal, F. Fresno, M. Barawi, C. Escudero, V. Perez-Dieste, J. M. Coronado, D. P. Serrano, J. R. Durrant and V. A. de la Peña O'Shea, *Nat. Commun.*, 2018, **9**, 1–10.
- 51 S. Nair and B. J. Tatarchuk, *Fuel*, 2010, **89**, 3218–3225.
- 52 D. Marcinkowski, M. Walesa-Chorab, V. Patroniak, M. Kubicki, G. Kadziolka and B. Michalkiewicz, *New J. Chem.*, 2014, **38**, 604.
- 53 C. Te Hsieh, W. S. Fan, W. Y. Chen and J. Y. Lin, *Sep. Purif. Technol.*, 2009, **67**, 312–318.
- 54 J. Xiao, Q. Shang, Y. Xiong, Q. Zhang, Y. Luo, S. Yu and H. Jiang, *Angew. Chemie*, 2016, **128**, 9535–9539.
- 55 C. R. Henry, C. Chapon, S. Giorgio and C. Goyhenex, in *Chemisorption and Reactivity on Supported Clusters and Thin Films*, Springer Netherlands, 1997, pp. 117–152.
- 56 D. J. Lewis, T. M. Day, J. V. MacPherson and Z. Pikramenou, *Chem. Commun.*, 2006, 1433–1435.
- 57 K. ichi Shimizu, Y. Miyamoto and A. Satsuma, *J. Catal.*, 2010, **270**, 86–94.
- 58 M. Lamothe, M. Plodinec, L. Scharfenberg, S. Wrabetz, F. Girgsdies, T. Jones, F. Rosowski, R. Horn, R. Schlögl and E. Frei, *ACS Appl. Nano Mater.*, 2019, **2**, 2909–2920.
- 59 F. Çeçen and O. Aktas, *Activated carbon for water and wastewater treatment: integration of adsorption and biological treatment*, 2011.
- 60 J. D. S. F. Da Silva, D. L. Malo, G. A. Bataglion, M. N. Eberlin, C. M. Ronconi, S. Alves and G. F. De Sá, *PLoS One*, 2015, **10**, 1–20.
- 61 S. M. Mirsoleimani-Azizi, P. Setoodeh, F. Samimi, J. Shadmehr, N. Hamed and M. R. Rahimpour, *J. Environ. Chem. Eng.*, 2018, **6**, 4653–4664.
- 62 Z. Jia, M. Jiang and G. Wu, *Chem. Eng. J.*, 2017, **307**, 283–290.
- 63 Z. Ren, J. Luo and Y. Wan, *Chem. Eng. J.*, 2018, **348**, 389–398.
- 64 H. Wang, S. Zhao, Y. Liu, R. Yao, X. Wang, Y. Cao, D. Ma, M. Zou, A. Cao, X. Feng and B. Wang, *Nat. Commun.*, 2019, 1–9.
- 65 R. Guo, X. Cai, H. Liu, Z. Yang, Y. Meng, F. Chen, Y. Li and B. Wang, *Environ. Sci. Technol.*, 2019, **53**, 2705–2712.
- 66 M. Patel, R. Kumar, K. Kishor, T. Mlsna, C. U. Pittman and D. Mohan, *Chem. Rev.*, 2019, **119**, 3510–3673.
- 67 H. C. Thomas, *J. Am. Chem. Soc.*, 1944, **66**, 1664–1666.
- 68 H. Li, J. He, K. Chen, Z. Shi, M. Li, P. Guo and L. Wu, *Materials (Basel)*, 2020, **13**, 16–20.

- 69 J. L. Sotelo, G. Ovejero, A. Rodríguez, S. Álvarez and J. García, *Ind. Eng. Chem. Res.*, 2012, **51**, 5045–5055.
- 70 P. Dauthal and M. Mukhopadhyay, *Korean J. Chem. Eng.*, 2015, **32**, 837–844.
- 71 K. Zhang, J. M. Suh, J. W. Choi, H. W. Jang, M. Shokouhimehr and R. S. Varma, *ACS Omega*, 2019, **4**, 483–495.
- 72 D. A. Islam, A. Chakraborty and H. Acharya, *New J. Chem.*, 2016, **40**, 6745–6751.