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1 **Does harvesting amplify environmentally induced population fluctuations**
2 **over time in marine and terrestrial species?**

3

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11

12 **Abstract**

- 13 1. In marine and terrestrial ecosystems, organisms are affected by environmental
14 variations that cause fluctuations in population size. The harvest-interaction
15 hypothesis predicts that environmentally-induced fluctuations in population size are
16 magnified by harvesting. Empirical evidence is urgently needed in the context of
17 global change because greater fluctuations will increase extinction risk.
- 18 2. Here, we review theoretical and empirical work that has addressed the harvest-
19 interaction hypothesis in fish, birds and mammals. We identify the mechanisms by
20 which harvesting might make population size more variable over time and thereby
21 increase the risk of extinction.
- 22 3. Theoretical models show that harvest can modify population structure in time and
23 space, and that changes in the amplitude and synchrony of population dynamics both
24 increase extinction risk. Empirical evidence indicates that fishing amplifies the effects
25 of environmental changes on the population variability, but no empirical study of

26 terrestrial species has tested for amplified environmentally-induced fluctuations due to
27 hunting.

28 4. *Synthesis and applications.* In terrestrial species, theoretical studies have evaluated
29 how environmentally-induced fluctuations in population size are magnified by
30 different harvest strategies, but there is now an urgent need for an empirical evaluation
31 of this hypothesis. Future research is needed to explore how hunting and climate
32 interact and to test whether hunting enhances environmentally induced fluctuations in
33 population numbers of terrestrial species.

34

35 **KEYWORDS**

36 climate change, environmental factors, fishing, hunting, population dynamics, harvesting,
37 population size, environmentally induced fluctuations

38

39 1 | INTRODUCTION

40 Free-ranging populations live in a global environment that is rapidly changing. All
41 ecosystems, from marine to terrestrial, are impacted by global drivers of environmental
42 change. Climate variation is one of the most important drivers of annual fluctuations in
43 population size (Sæther 1997; Gaillard, Festa-Bianchet & Yoccoz 1998; Sæther *et al.* 2016).
44 Variability in climate may influence population size through changes in phenotypic traits such
45 as body condition or key demographic rates including fecundity or survival (Jenouvrier 2013;
46 Jenouvrier *et al.* 2018). For instance, climate variation impacts age of maturity, recruitment
47 rates and hence the overall population dynamics of large herbivores (Sæther 1997; Post *et al.*
48 1999). Climate effects can have strong consequences on population persistence, because
49 greater environmentally-induced fluctuations in population size will lead to higher extinction
50 risk (Lande, Engen & Saether 2003). Exploited natural populations are also affected by
51 commercial, recreational or subsistence harvest, with removal of individuals through fishing
52 or hunting. The magnitude of the fluctuations in population size can be large if declining
53 populations are overharvested or populations that start to rebound have low harvest rates
54 (Fryxell *et al.* 2010). To maintain sustainable harvest and to avoid extinction due to
55 overexploitation in the current context of global change, it is crucial to include the effects of
56 environmental trends and stochastic variation in the environment in the development of
57 sustainable harvest strategies (May *et al.* 1978; Lande, Engen & Saether 1995; Brander 2007).
58 The question of whether harvesting amplifies the effects of environmental changes on the
59 variability in population size over time and increases extinction risk (hereafter called
60 “*harvest-interaction hypothesis*”) or alternatively, stabilizes population dynamics has been an
61 unsolved problem for some years. During the last fifteen years, the question of potential
62 interactions between climate and harvest has received growing interest in marine ecology,
63 mainly due to the collapse of commercially important fisheries (Hutchings & Myers 1994;

64 Myers, Hutchings & Barrowman 1997). Here, we assess whether harvesting can amplify
65 environmentally-induced population fluctuations and affect dynamics of exploited
66 populations in the changing environments in both marine and terrestrial ecosystems.

67 Our objective was to assess the ecological consequences of harvesting in a variable
68 environment by identifying the mechanisms causing fishing/hunting to increase
69 environmentally-induced population fluctuations in terms of changes in population size over
70 time. Here, we review theoretical and empirical studies based on observational or
71 experimental approaches that have evaluated the harvest-interaction hypothesis of higher
72 variability in exploited population sizes over time. Based on our findings, we propose new
73 avenues of research for understanding the effects of harvest on exploited populations of
74 marine and terrestrial organisms.

75

76 **2 | HARVEST EFFECTS ON AGE/SIZE STRUCTURE**

77 Harvesting can be proportional (or balanced) on all age/size classes in a population.
78 However, removal of individuals through harvesting is rarely random with respect to size or
79 demographic classes (Kuparinen & Festa-Bianchet 2017). If the oldest and largest individuals
80 are preferentially removed, or if harvesting pressure is high, the proportion of old individuals
81 in the populations decreases. In exploited populations of northern cod *Gadus morhua*, the age
82 structure contained 18 age classes in the 1960s, but the older fish disappeared and the age
83 structure was truncated to 9 age classes more recently (Drinkwater 2002). The northern cod is
84 not an exception: age truncation is pervasive in exploited populations where the proportion of
85 individuals in the oldest age classes drastically decreases at a severe rate (reviewed by Barnett
86 *et al.* 2017 for 63 exploited populations of fish). A growing body of empirical evidence shows
87 that fishing plays a key role in changes in size structure, and that climate changes such as
88 increasing water temperature contribute to a less extent to the observed changes in size

89 distribution (Tu, Chen & Hsieh 2018). Size-selective harvest is problematic because the oldest
90 individuals may have different fecundity and survival rates that buffer the population against
91 environmental changes. In fish, large/old females are generally more fertile than small/young
92 ones and spawn during different temporal and spatial windows (Berkeley, Chapman & Sogard
93 2004), hence spreading the risk of losing all the offspring when conditions are unfavorable. In
94 ungulates, selective harvest can also lead to age truncation with a larger proportion of
95 individuals in younger age-classes than populations not subject to hunting. Many studies have
96 documented that in long-lived species, young individuals exhibit more variable survival rates
97 over years than adults, adult survival being canalized and thus buffered against environmental
98 variation (Gaillard & Yoccoz 2003), a process called “demographic buffering” (Morris &
99 Doak 2004). Consequently, harvest-induced changes in age structure might increase
100 sensitivity of the population dynamics to climate variation if the remaining age classes are
101 more sensitive to climate variation. Conversely, one can expect dampened environmentally-
102 induced population fluctuations if harvest targets young individuals that are the most sensitive
103 to climate variation. Harvest strategies are often designed to target specific demographic
104 classes. In many European ungulates, harvest quotas are designed to take more calves and
105 yearlings with low reproductive value (Apollonio, Andersen & Putman 2010). In fisheries
106 management, evidence is also accumulating that regulations stipulating a maximum size of
107 fish that can be harvested and thus favoring the removal of the youngest/smallest individuals
108 can reduce the negative effects of harvest (Birkeland & Dayton 2005). Selective harvest
109 directed towards some specific age classes implies that hunters can identify the age of the
110 individuals. Group composition, phenotypic traits or coat color can be used to identify age
111 classes. For example, in wild boar (*Sus scrofa*), a social group is led by an old female,
112 followed by juveniles striped until 4 months of age and wearing a reddish coat until 6 months

113 of age (Gamelon *et al.* 2012), making age determination straightforward. In fish, gill nets are
114 commonly used tools to select and remove individuals of specific size-classes.

115 A growing body of empirical evidence shows that age truncation might enhance climate
116 sensitivity and generate fluctuations in population size over years (Anderson *et al.* 2008). In
117 2012, from long-term data of 6 fish species, Rouyer *et al.* (2012) showed that high mortality
118 due to fishing is associated with a change in age structure that enhances the population
119 susceptibility to extrinsic environmental forcing (table 1). Five years later, Shelton and
120 Mangel (2011) confirmed through modeling and demographic data collected from the
121 literature including 45 fish species (table 1) that age-selective fishing increases the relative
122 contribution of recruits and thus amplifies the destabilizing effect of climate variation. This
123 finding has been further discussed by Sugihara *et al.* (2011). Recently, Cameron *et al.* (2016)
124 experimentally tested the hypothesis of enhanced fluctuations in population size when
125 adult/old individuals were removed using a terrestrial species, the soil mite (*Sancassania*
126 *berlese*). The experiment consisted of different harvesting treatments under different regimes
127 of environmental fluctuations. The authors showed that harvesting adults increased population
128 variability but only in variable environments. In a constant environment, harvest had no effect
129 on the coefficient of variation (CV) of population sizes over time. The laboratory study
130 provided experimental evidence that harvesting directed towards adults can increase climatic
131 sensitivity of exploited populations (table 1). Harvest may also select for accelerated life
132 histories with higher fecundity, early maturity and short lifespan (Anderson *et al.* 2008;
133 Allendorf & Hard 2009; Kuparinen & Festa-Bianchet 2017). When the rate of turnover of
134 individuals is accelerated, a population is moved towards the fast end of the continuum of life
135 history variation (see Gaillard *et al.* 2016) and becomes more prone to greater fluctuations in
136 population sizes in response to variation in environmental conditions (Gamelon *et al.* 2014).
137 In particular, environmental fluctuations generate temporal autocorrelations among fecundity,

138 survival and other demographic rates that in turn influence both the mean and the variance of
139 population size (Engen et al. 2013). The effects of temporal autocorrelation on the
140 demography depend on life history variation: species with fast life histories are more sensitive
141 to autocorrelation than slow species with opposite characteristics (Paniw, Ozgul & Salguero-
142 Gómez 2018). A shift towards younger age classes due to harvest that accelerates the pace of
143 life may thus also favour autocorrelation in population dynamics. Thus, it may be difficult to
144 disentangle whether changes in population size are caused by harvest or changes in climate
145 conditions, which in turn may increase the risk for overexploitation and population collapse.

146

147 **3 | HARVEST EFFECTS ON SPATIAL DISTRIBUTION**

148 The distributional range of some exploited species may shift in response to climate
149 change (Mysterud & Sæther 2010). For example, cod and pollock (*Gadus chalcogrammus*)
150 distributions have changed over the last years in the northern oceans (Hollowed & Sundby
151 2014). Northbound shifts in spatial distribution allow individuals to better cope with changing
152 environmental conditions. Harvesting per se may affect the spatial distribution of exploited
153 species (Takashina & Mougi 2015). For instance, in a network of patches with and without
154 harvest, individuals can move from patches without harvest to harvested patches, a process
155 called compensatory immigration (see Gervasi *et al.* 2015 for an example on wolverines *Gulo*
156 *Gulo*). Alternatively, when exposed to harvest, individuals might move to new habitats and
157 exploit any available niche to avoid mortality risk (see Lefebvre *et al.* 2017 for an example on
158 greater snow goose *Anser caerulescens*). Displacement can result in high spatial aggregation,
159 or enhanced spatial heterogeneity of animal distribution (see Ciannelli *et al.* 2013 for a review
160 on fish). Similarly, different age classes can inhabit contrasting habitats according to their
161 age-specific needs, mobile abilities and competitive advantages (Rijnsdorp & Pastoors 1995;
162 Nøttestad *et al.* 1999; Fauchald, Mauritzen & Gjøsæter 2006; Breau, Cunjak & Bremset

163 2007), resulting in homogeneous populations in terms of spatial structure capable of
164 occupying various habitats. Harvest-induced changes in (st)age structure can also alter spatial
165 structure and favour heterogeneity (Kuo *et al.* 2016). Empirical work has shown that such a
166 constriction of geographic distributions associated with harvest pressure can weaken the
167 populations' ability to buffer the effects of unfavorable environmental conditions (Planque *et*
168 *al.* 2010). Populations become more vulnerable to climate variability, and fluctuations in
169 population numbers and risk of local extinction increase (Hsieh *et al.* 2008; Kuo *et al.* 2016).
170 If the goal of managers is to decrease population sizes, for instance to control overabundant
171 populations of geese, increasing vulnerability to climate variability and local extinction risks
172 might be a desirable outcome. Similarly, migration is a simple way to buffer climate-driven
173 variations in terrestrial and marine ecosystems. Seasonal migration is a widespread
174 phenomenon in vertebrates, being observed in fish, birds and mammals (Milner-Gulland,
175 Fryxell & Sinclair 2011). Migration patterns can be culturally transmitted through social
176 learning in both marine (Corten 2002) and terrestrial environments (Jesmer *et al.* 2018).
177 Harvest of mature individuals may lead to a loss of transmission of migration knowledge to
178 younger individuals, alter migration patterns and lead to greater sensitivity to climate change
179 (Fryxell *et al.* 2010). Hence, harvesting can affect the spatial distribution of exploited species
180 and shifts in distribution can be unfavourable when facing a changing climate (see Hsieh *et al.*
181 2010 for a review in fisheries).

182 Moving from the effects of harvesting on the spatial structure of a single population to
183 populations structured in space, populations separated in space that exhibit correlated
184 fluctuations in population size over years are theoretically more prone to extinction because
185 local patches can go extinct without being rescued by immigration from other populations
186 (Heino *et al.* 1997; Earn, Levin & Rohani 2000; Engen 2007). One of the key insights that has
187 been derived from spatial population models is that increased synchrony in population

188 fluctuations over larger areas tends to reduce population size and increase the risk of
189 extinction (Foley 1994; Engen, Lande & Saether 2002). Recent theoretical analyses have
190 shown that proportional harvest itself, where the annual yield is proportional to the population
191 size, tends to induce spatial synchrony (Engen 2017; Engen, Cao & Sæther 2018). Greater
192 synchrony implies that spatially variation in harvest strategies including protected areas could
193 be a way of avoiding this effect. In fact, theoretical results have shown that adjusting the
194 harvesting rate according to local population fluctuations can result in reduced spatial scaling
195 of the population synchrony. Spatio-temporal changes in the covariation of key environmental
196 variables are expected due to changes in climate, which may provide a mechanism to
197 counteract climate-induced changes in the dynamics of exploited species. For example, if the
198 spatial scaling of the environmental noise is altered because of alteration of the climate
199 regime, the proportion of juveniles included in the harvest will influence the spatial scaling of
200 the distribution of abundance (Engen, Lee & Sæther 2018). Growing empirical evidence in
201 fisheries indicates that synchronized harvest at a large spatial scale plays a key role in
202 affecting the degree of spatial synchrony in population fluctuations, providing support for
203 theoretical expectations. For instance, Frank et al. (2016) showed that the size of 22
204 populations of cod fluctuate synchronously in North Atlantic not solely due to climate but
205 also due to harvesting.

206

207 **4 | HARVEST EFFECTS ON DENSITY REGULATION**

208 Fluctuations in size of natural populations are driven by climate effects (Sæther *et al.*
209 2000; Coulson *et al.* 2001; Stenseth *et al.* 2003; Berryman & Lima 2006) and also by density
210 dependence, involving a negative feedback of the population density on the population
211 growth rate (Royama 1992; Turchin 1995). Mechanisms can vary but include intraspecific
212 competition, social behavior and other factors. Importantly, climate effects can interact with

213 density-dependent regulation of population dynamics (Turchin 1995; Stenseth *et al.* 2003;
214 Barbraud & Weimerskirch 2003; Gamelon *et al.* 2017; Hansen *et al.* 2019). For example,
215 effects of climate on demographic rates and population size can be stronger at high density
216 than at low density because when climate conditions are bad, food resources are more limiting
217 if the population is at or above carrying capacity (see Gamelon *et al.* 2017 for a case study on
218 dippers *Cinclus cinclus*; see Hansen *et al.* 2019 for a case study on Svalbard reindeer *Rangifer*
219 *tarandus*). If the population is above carrying capacity, the remaining individuals in the
220 population after harvest might exhibit improved fecundity and survival via *compensatory*
221 *density dependence* (Bonenfant *et al.* 2009), being for instance more likely to colonize
222 habitats of better quality instead of occupying less optimal places (MacCall 1990; Marshall &
223 Frank 1995). In contrast, reduced population sizes can be associated with lower demographic
224 performance for the remaining individuals via Allee effects such as higher risk of starvation
225 due to stress (reviewed by Courchamp, Berec & Gascoigne 2008; Péron 2013). The interplay
226 between climate and density dependence (Gamelon *et al.* 2017; Hansen *et al.* 2019), and
227 between harvest and density dependence (Boyce, Sinclair & White 1999; Sandercock *et al.*
228 2011) have been well studied, but an integrative approach to the combined effects of
229 harvesting and climate on density-regulated populations remains lacking. An integrated
230 approach is strongly needed because complex interactions between harvest, density
231 dependence and climate variation might arise. For instance, harvest might influence the way
232 in which a population responds to density dependence, make it more or less vulnerable to
233 climate variation, and thus might ultimately amplify or dampen fluctuations in population
234 sizes over time.

235

236 **5 | EVIDENCE FOR HARVEST EFFECTS ON POPULATION DYNAMICS**

237 Empirical evidence is accumulating in marine species of fish that harvesting might
238 magnify environmentally-induced fluctuations in population sizes. In a pioneering study,
239 Hsieh et al. (2006) compared the temporal variability of 13 exploited vs. 16 unexploited fish
240 populations living in a common environment and subject to the same environmental
241 fluctuations. From long-term data, they found that the CV of annual larval abundance was
242 lower in unexploited than in exploited populations and concluded that the addition of
243 mortality from fishing resulted in increased temporal variability of all species subject to
244 variable environments (table 1). In terrestrial species, some studies have shown that specific
245 harvesting strategies may increase fluctuations in population size and favor cyclicity,
246 accentuating the risk of population collapse when environmental conditions are harsh (Fryxell
247 et al. 2010). In a case study of black grouse (*Tetrao tetrix*), Jonzen et al. (2003) demonstrated
248 that harvesting can cause quasi-cyclic fluctuations when the harvest fraction was time-
249 dependent, thus reducing the ability of the population to recover when environmental
250 conditions are unsuitable. Similarly, Bunnefeld et al. (2011) used simulations to explore the
251 effects of selective harvest by sex and age on population cycles in red grouse (*Lagopus*
252 *lagopus*). In general, a *threshold harvest strategy* that involves immediate removal of all
253 individuals above a critical population size generates a large variance in annual yield (Sæther,
254 Engen & Solberg 2001; Sæther et al. 2005) and large fluctuations in age distributions. Engen
255 et al. (1997) therefore proposed a *proportional threshold strategy*, in which only a fraction of
256 the excess individuals above a lower critical threshold is removed. Proportional threshold
257 involves a reduction of the threshold when harvest is allowed compared to a pure threshold
258 strategy. The modified strategy allows for an increase in the number of years harvesting is
259 permitted, which in turn generates less variability in the annual yield (Sæther, Engen &
260 Solberg 2001; Aanes et al. 2002; Sæther et al. 2005). Thus, the strategy should make it easier

261 to separate out climate-induced influences on the population dynamics and hence adjust the
262 harvest tactics accordingly.

263

264 **6 | EVIDENCE FOR HARVEST EFFECTS ON COMMUNITY DYNAMICS IN** 265 **SPACE**

266 Species live in communities with other species and can compete for limiting resources.

267 The harvest of one species may induce cascading effects on other species because it may
268 affect interspecific interactions and hence ecosystem processes leading to alternative stable
269 states (Barkai & McQuaid 1988; Kirby, Beaugrand & Lindley 2009; Selkoe *et al.* 2015;
270 Gårdmark *et al.* 2015). Surprisingly, a topic that to a large extent has been ignored is how
271 harvesting affects the fluctuations of population size of interacting species. Jarillo *et al.*
272 (2018) showed that proportional harvesting, in which the catches at each location are
273 proportional to the population size, tends to synchronize fluctuations in population size of
274 competing harvested species in space, and thus ultimately increases the probability of
275 extinction. However, if other harvesting strategies are applied, such as *fixed quota* strategy for
276 catches more concentrated in areas with high population sizes, the outcome can be opposite:
277 harvesting can reduce the synchrony of the species (Engen, Cao & Sæther 2018) and thus the
278 risk of local extinction. However, fixed quota harvest strategies are generally not sustainable
279 on the long-run (Lande, Engen & Saether 1995; Lande, Sæther & Engen 1997). Thus,
280 different harvesting strategies can have contrasting effects on vulnerability of interacting
281 species over large geographical areas to changes in climate and cause spatial decoupling of
282 ecosystem structures. Interactions imply that harvest of a given species may affect the
283 dynamics of entire food webs, also affecting non-harvested species. The removal of species at
284 low-trophic levels affects the dynamics of upper-trophic levels and thus the entire community
285 (Smith *et al.* 2011). Similarly, when apex predators are removed from the ecosystem or when

286 their dynamics are modified by trophic down-grading, the entire ecosystem can be
287 destabilized through trophic cascades, involving changes in the abundance of multiple
288 interacting species subsequently to the change in abundance of a single species (McCann
289 2012; Terborgh 2015). Cascades can modify the response of the entire community to
290 changing climate conditions (Estes *et al.* 2011). For instance, the removal of carnivores might
291 lead to high abundance in herbivores as ungulates, higher herbivory and lower frequency and
292 intensity of wildfires (Estes *et al.* 2011). In marine systems, overfishing of large-bodied fish
293 might reduce herbivory, increase levels of disease and prevent the recovery of corals after
294 cyclones (Hughes *et al.* 2003). In this later example, interactions among species are essential
295 to the resilience of the community (Loreau 2010), and when some species decline or are
296 removed from the system, interactions are broken and the ability of the entire community to
297 recover from disturbances such as changing climate conditions decreases. Understanding such
298 interactions has received an increasing interest in the recent years and multi-species
299 interactions are now included in models for sustainable harvest of exploited populations
300 (Leslie & McLeod 2007; Persson, Van Leeuwen & De Roos 2014). Ecosystem-based
301 approaches are crucial to gain a comprehensive understanding of harvest-induced effects at
302 the community level.

303

304 **7 | CONCLUSIONS**

305 Does harvesting increase environmentally-induced fluctuations in population size?
306 Among empiricists, this question has mainly been addressed by marine ecologists and
307 evidence for the “harvest-interaction hypothesis” is accumulating (table 1). The combined
308 effect of fishing and climate on population fluctuations in fish stocks has received increasing
309 interest in the recent years (Lindegren *et al.* 2013; Frank *et al.* 2016) and both theoretical and
310 empirical works have provided evidence that specific types of harvesting strategies might

311 increase environmentally-induced fluctuations in population size and thus extinction risk. In
312 terrestrial environments, studies evaluating the “harvest-interaction hypothesis” have mainly
313 developed theoretical approaches to evaluate how contrasting harvesting strategies might
314 enhance or reduce environmentally-induced fluctuations in population size. Surprisingly,
315 empirical evidence is lacking. Few field experiments have been conducted and little is known
316 on the combined effect of hunting and climate on population fluctuations in terrestrial species.
317 Available studies have measured the relative contributions of climate, density dependence and
318 harvesting on the population growth rate fluctuations (Koons *et al.* 2015 for a study on
319 American bison *Bison bison*; López-Montoya, Moro & Azorit 2017 for a study on red *Cervus*
320 *elaphus hispanicus* and fallow deer *Dama dama*) whereas other studies have evaluated the
321 effect of different harvesting strategies on population dynamics (McGowan 1975; Mentis &
322 Bigalke 1985; Connelly *et al.* 2003; Gervasi *et al.* 2015). Other studies have compared the
323 dynamics of harvested vs. natural populations in similar environments (Palmer & Bennett Jr
324 1963; Pedersen *et al.* 2004). Tests of how harvesting enhances or dampens environmentally-
325 induced fluctuations in population sizes by comparing the dynamics of harvested vs. natural
326 populations in similar environments have not yet been conducted. We strongly encourage
327 analyses of this question on a large range of species with contrasting life histories (slow
328 versus fast species), with contrasting abilities to face environmental fluctuations (mobile
329 versus resident species, ectotherms versus endotherms), at different trophic levels (carnivores
330 versus herbivores) by using some of the existing long-term field studies of exploited
331 populations. Long-term individual-based studies with capture-mark-recapture data and
332 population counts are available for many exploited bird and mammalian populations.
333 Analyses of time series of classified counts and harvest data can be complicated by hunting
334 regulations and hunter preferences and also by covariation between abundance and
335 regulations. As stated by Festa-Bianchet *et al.* (2017) “*It remains unclear, therefore, whether*

336 *the response of heavily hunted populations to changes in density, weather, and resource*
337 *availability differs from that documented by long-term studies of un hunted populations.” A*
338 promising avenue of research is to explore how hunting and climate interact and evaluate
339 whether, like observed in marine species, hunting enhances environmentally-induced
340 fluctuations in population sizes for terrestrial exploited species. Both the theory and the data
341 are now available to address this question of major importance in ecology, management and
342 conservation in exploited avian and mammalian populations.

343

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349

350 **AUTHORS' CONTRIBUTIONS**

351 M.G. led the writing of the manuscript. B.K.S. and B.-E.S contributed to the drafts and gave
352 final approval for publication.

353

354 **DATA ACCESSIBILITY**

355 Data have not been archived because this article does not contain data.

356 Table 1 – Case studies with evidence that harvesting enhances environmentally-induced fluctuations in population sizes.

Species	Environment	Approach	Evidence for effects of harvest on greater environmentally-induced fluctuations in population sizes	Source
29 species	Marine	Observational	“This well-documented mechanism [age truncation] suggests how fishing can make populations more susceptible to extrinsic environmental forcing.”	Hsieh et al. 2006
6 species	Marine	Observational	“Our results demonstrate that populations exhibiting a truncation of their age structure were associated with shorter-term fluctuations in time series, which suggests an increased sensitivity to the short-term changes in the environment.”	Rouyer et al. 2012
45 species	Marine	Observational	“In a variable environment, the addition of mortality from fishing leads to increased temporal variability for all species examined.”	Shelton and Mangel 2011
1 species of acarid mite	Terrestrial	Experimental	“Only when environmental variation and harvesting were combined did we see strong evidence of magnified fluctuations in abundance.”	Cameron et al. 2016

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