

Multistability of model and real dryland ecosystems through spatial self-organization

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Robbin Bastiaansen^{a,1}, Olfa Jaïbi^a, Vincent Deblauwe^{b,c}, Maarten B. Eppinga^d, Koen Siteur^{e,f}, Eric Siero^g, Stéphane Mermoz^h, Alexandre Bouvet^h, Arjen Doelman^a, and Max Rietkerk^d

^aMathematical Institute, Leiden University, 2300 RA Leiden, The Netherlands; ^bInternational Institute of Tropical Agriculture, BP 2008 (Messa), Yaounde, Cameroon; ^cCenter for Tropical Research, Institute of the Environment and Sustainability, University of California, Los Angeles, Los Angeles, CA 90095, USA; ^d Department of Environmental Sciences, Copernicus Institute, Utrecht University, 3508 TC Utrecht, The Netherlands; ^eDepartment of Estuarine and Delta Systems, Royal Netherlands Institute for Sea Research and Utrecht University, 4401 NT Yerseke, the Netherlands; ^fShanghai Key Laboratory for Urban Ecological Processes and Eco-Restoration & Center for Global Change and Ecological Forecasting, School of Ecological and Environmental Science, East China Normal University, 200241 Shanghai, China; ^gInstitute for Mathematics, Carl von Ossietzky University Oldenburg, 26111 Oldenburg, Germany; ^hCESBIO, Université de Toulouse, CNES, CNRS, IRD, UPS, France

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Spatial self-organization of dryland vegetation constitutes one of the most promising indicators for an ecosystem's proximity to desertification. This insight is based on studies of reaction-diffusion models that reproduce visual characteristics of vegetation patterns observed on aerial photographs. However, until now the development of reliable early warning systems has been hampered by the lack of more in-depth comparisons between model predictions and real ecosystem patterns. In this paper, we combined topographical data, (remotely sensed) optical data and in-situ biomass measurements from two sites in Somalia to generate a multi-level description of dryland vegetation patterns. We performed an in-depth comparison between these observed vegetation pattern characteristics and predictions made by the extended-Klausmeier model for dryland vegetation patterning. Consistent with model predictions, we found that for a given topography, there is multi-stability of ecosystem states with different pattern wavenumbers. Furthermore, observations corroborated model predictions regarding the relationships between pattern wavenumber, total biomass and maximum biomass. In contrast, model predictions regarding the role of slope angles were not corroborated by the empirical data, suggesting that inclusion of small-scale topographical heterogeneity is a promising avenue for future model development. Our findings suggest that patterned dryland ecosystems may be more resilient to environmental change than previously anticipated, but this enhanced resilience crucially depends on the adaptive capacity of vegetation patterns.

vegetation patterns | spatial self-organization | Busse balloon | arid ecosystems | ecosystem resilience

key aim of ecological modeling is to generate an understanding of the mechanisms driving observed patterns (1). A significant challenge in this pursuit, however, is that multiple alternative processes may generate the same emergent outcome (1–4), a phenomenon also referred to as equifinality (5, 6). As a result, modeling efforts may reveal that a particular ecological pattern can be explained by a suite of alternative driver mechanisms. Therefore, a match between a pattern simulated with a mechanistic model and a pattern observed in a real ecosystem may only constitute limited support for the modeled mechanism being its true driver (2, 5, 6).

Pattern-oriented modeling (2, 7) aims to address the challenge of equifinality of alternative model formulations. In this approach, model assessment is based on the degree to which the output corresponds to observed patterns. A distinction is made between strong and weak patterns. Strong patterns are the dominant emergent features a model should reproduce, such as the cycles within predator and prey population sizes, or a spatial distribution of vegetation patches (6, 7). Weak patterns are typically qualitative relationships, such as the existence of a population over a specific timespan, or a positive association between one state variable and another (6, 7). Rather than comparing model output to a single strong pattern, additional comparisons to multiple weak patterns, at different scales or levels of organization, provide more power to model validation and selection procedures (2, 6, 7).

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A specific type of ecological patterns that has received considerable attention is regular spatial patterning of sessile biota (8). On flat terrain, the reported patterns are gaps, labyrinths, and spots (9, 10). On sloping grounds banded patterns form, their regular spacing enabling a description of the characteristic band-inter-band period and wavenumber. Evidence is accumulating that these patterns are self-organized, meaning that the larger-scale patterning is driven by internal ecosystem processes operating at smaller scales (8, 11). The crucial component in this self-organization process is a long-range negative effect of biota on itself, either directly or through modulation of resource availability. In cases where this long-range negative feedback is coupled to a locally positive feedback, the mechanism creating pattern formation may be linked to the existence of alternative stable states, as well

Significance Statement

Today, vast areas of drylands in semi-arid climates face the dangers of desertification. To understand the driving mechanisms behind this effect, many theoretical models have been created. These models provide insight in the resilience of dryland ecosystems. However, until now, comparisons with reality were merely visual. In this article, a systematic comparison has been performed using data on wavenumber, biomass and migration speed of vegetation patterns in Somalia. In agreement with reaction-diffusion models, a wide distribution of regular pattern wavenumbers has been found in the data. This highlights the potential for extrapolating predictions of those models to real ecosystems, including those that elucidate how spatial self-organization of vegetation enhances ecosystem resilience.

K.S., E.S., A.D., and M.R. initiated the research, R.B., O.J., V.D., M.B.E., K.S., E.S., A.D. and M.R. designed research, S.M. and A.B. provided biomass data; V.D. provided other data and processed data; E.S. and R.B. created the model figures, R.B and O.J. analyzed data; R.B., V.D., M.B.E. and K.S. wrote the paper. O.J., E.S., A.D., and M.R. provided feedback on draft versions of the manuscript.

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¹To whom correspondence should be addressed. E-mail: r.bastiaansen@math.leidenuniv.nl

as the possibility of so-called catastrophic shifts between these states (11). This phenomenon has been most prominently studied in (semi-)arid ecosystems, where decreases in resource availability or increases in grazing pressure may trigger catastrophic shifts from vegetated states to desert states without vegetation (12–14). In this context, the formation of regular spatial vegetation patterns may indicate proximity to a threshold of catastrophic change (11).

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There is a long tradition in the scientific literature of explaining regular spatial patterning with reaction-diffusion models (15-17). In line with this work, a variety of reaction-diffusion models has been applied to investigate selforganization in (semi-)arid ecosystems (9, 10, 18, 19). Despite the broad support for the findings obtained with these models and their implications for (semi-)arid ecosystem functioning, comparisons of model results with empirical data have mainly been limited to comparison of a single strong pattern, namely the spatial distribution of vegetation patches. Until now, the few studies considering additional weak patterns have shown that reaction-diffusion model simulations successfully reproduce associations between pattern shape and aridity, and associations between pattern shape and slope of the terrain (20). In addition, models that account for sloped terrain also seem to capture the observed migration of the location of banded patterns in uphill direction (21). Despite these promising agreements between model results and empirical data, a more systematic comparison between model results and data, based on multiple patterns at different levels of organization (2, 7), was still lacking.

Advanced model analyses that have recently been applied to ecological models have yielded a number of findings which, when confronted with high quality remote sensing products, makes a more systematic comparison possible. More specifically, recent theoretical studies have shown that for a given environmental condition (i.e. a given parameter combination), not a single ecosystem state, but multiple ecosystem states with patterns spanning a range of wavenumbers may be stable, hence observable (22–24). The range of observable patterns, across a range of environmental conditions forms a bounded region in (parameter, wavenumber)-space. This region is referred to as the Busse balloon, after F.H. Busse, who studied similar phenomena in the field of fluid dynamics (25). Although the patterned ecosystem states in the Busse balloon are defined by their wavenumber, other properties, like migration speed and spatially averaged biomass, have also been studied (26) and are suggested to depend on the position of a system within the Busse balloon. These theoretical findings provide multiple additional weak patterns that can be compared to empirical data, providing opportunities for more powerful tests of the validity of the developed reaction-diffusion models to describe dryland ecosystems.

The aim of this study was to confront theoretical findings regarding pattern wavenumber, biomass and migration speed with the same pattern properties derived from aerial imagery and remote sensing products of banded vegetation patterns in the Horn of Africa, a location with prominent undisturbed presence of vegetation pattern formation. Hence, a multilevel comparison between theory and data in line with the pattern-oriented modeling approach was conducted (2, 6, 7).

1. Theory

A. Model description. Multiple reaction-diffusion models of dryland vegetation dynamics include a mechanism in which vegetation acts as an ecosystem engineer, locally increasing the influx of available water (9, 10, 18, 19). Despite the different nuances between these models, a number of predictions can be robustly derived from these frameworks. One of the simplest of these ecosystem models - and the archetype considered in this article - is an extended version of the dryland ecosystem model by Klausmeier (18, 22), which we will refer to as the extended-Klausmeier model. This model describes the interaction between water, w, and plant biomass, n. A non-dimensional version of this model is used for the purposes of this article. A dimensional version of the model and the physical meaning of its parameters can be found in SI 1. The model is given by the following equations

$$\begin{cases} \frac{\partial w}{\partial t} &= e \frac{\partial^2 w}{\partial x^2} + \frac{\partial (vw)}{\partial x} + a - w - wn^2, \\ \frac{\partial n}{\partial t} &= \frac{\partial^2 n}{\partial x^2} - mn + wn^2. \end{cases}$$
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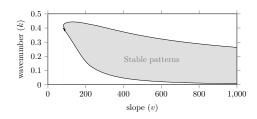
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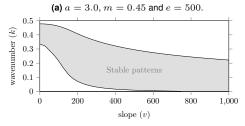
The reaction terms model the change in water as a combined effect of rainfall (+a), evaporation (-w) and uptake by plants $(-wn^2)$. The change of plant biomass comes from mortality (-mn) and plant growth $(+wn^2)$. Dispersion by plants is modeled as diffusion and the movement of water as a combined effect of diffusion and advection. The latter is due to gradients in the terrain, which are proportional to the slope parameter

B. Theoretical outcomes.

B.1. Multi-stability of patterned states. Reaction-advectiondiffusion equations in general – and the extended-Klausmeier model in particular – exhibit a vast variety of spatial patterns (27, 28). However, not all feasible patterns are stable solutions of these models. Which patterned states are stable (hence, observable) depends on the combination of the model parameters. For regular patterns, the concept of the Busse balloon can help to illustrate this dependency (25). A Busse balloon is a model dependent shape in the (parameter, wavenumber)-space that indicates all combinations of parameter and wavenumber that represent stable solutions of the model. If, for a given set of model parameters, a wavenumber k lies within the Busse balloon, then regular patterns with wavenumber k are observable. So, in measurements, all (non-transient) patterns are expected to be present in the Busse balloon.

Typically, the Busse balloon is a high-dimensional structure due to the number of parameters in a system. Therefore, usually, only one parameter is varied when a Busse balloon is visualized. This produces a 2D-slice of the full Busse balloon. 237 In the context of desertification research, the straightforward choice would be to vary the rainfall (23). However, mean annual rainfall was relatively constant in our study sites during the observation period considered. Instead, topography (i.e. the slope gradient) comprised the main source of environmental variation within our study areas. Thus, relevant theoretical predictions for our study sites can be generated by varying the slope parameter v (while keeping rainfall constant). Here, we present two of such 2D-Busse balloon slices for the extended-Klausmeier model (Figure 1), which were constructed by tracking the boundary of the Busse balloon





(b) a = 2.5, m = 0.45 and e = 500.

Fig. 1. (*slope,wavenumber*)-Busse balloon slices for the extended-Klausmeier model for two different values of the rainfall parameter a. A banded pattern solution to the extended-Klausmeier model with slope v and wavenumber k is stable if the (v,k)-combination lies inside the Busse balloon. This indicates that a wide spread of (v,k)-combinations yields stable banded patterns. The latter are therefore expected for a broad range of wavenumbers — and not for specific (v,k)-choices only. The shape of a Busse balloon can change between models and between parameter values. This is illustrated in the figures which were computed for different a-values.

using numerical continuation methods (23, 24, 29, 30). The shaded region in these figures indicates the combinations of pattern wavenumber k and slope v for which stable solutions exist. Thus, a specific slope v does not lead to one specific pattern. In fact, the model shows multi-stability; a given slope v can sustain a continuous range of wavenumbers k. A similar spread in wavenumbers is expected in the real system.

Though the Busse balloon indicates which patterns might be observed, it does not specify the likelihood of finding a certain pattern with specific wavenumber k within this range. Recent numerical studies suggest that the (entire) history of environmental changes is relevant in the selection process (26, 31). To understand these hysteretic dynamics, it is vital to acknowledge that model patterns do not change their wavenumber unless they have to (23, 30): if an environmental change forces the system outside of the Busse balloon, the current pattern has become unstable, and will need to adapt into a new pattern that is again stable – thus part of the Busse balloon. During this (fast) adaption, only part of the vegetation bands are lost, while the remaining bands increases in size; these adaptions thus have limited effect on the total biomass in the system (23). Hence multiple wavenumber adaptations are expected to occur after each other that will, gradually, lead to a complete desertification of the system (23). Precisely which wavenumber k gets selected at each of these destabilizations is difficult to predict, though, for low noise levels, a preference for an approximate period doubling is expected, i.e. the wavenumber gets halved (23).

Numerical simulations help to get an insight in the kind of wavenumber distribution one ought to expect in observations. To illustrate the typical spread in wavenumber, a total of 200 simulations on a flat terrain (v=0) were run, where the rainfall parameter was slowly decreased from a=3 to a=0.5. The initial configurations for these runs were chosen randomly, but close to the equilibrium state of uniform biomass before

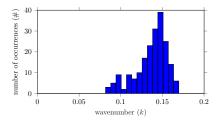


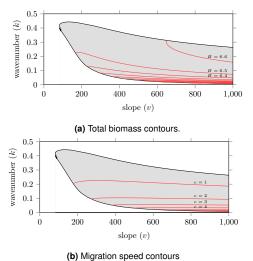
Fig. 2. Histogram demonstrating a spread in wavenumber (k) at the end of 200 simulations of the extended-Klausmeier model on a flat terrain (v=0) with model parameters e=500 and m=0.45. These simulations had a random initial configuration close to a stable fully vegetated state. A climate change was simulated by decreasing the rainfall parameter a linearly from a=0.5 to over the course of a=0.5 time unit, causing several pattern selections and corresponding changes in wavenumber.

the onset of patterns (between 90% and 110% of the uniform vegetated equilibrium state). At the end of each simulation – after several pattern selections – the wavenumber of the remaining pattern was measured. This gives a snapshot of the wavenumber distribution, similar to the snapshots acquired from observations. Note that a similar experiment was done before, albeit on a much smaller scale (30). The histogram of the resulting wavenumbers is shown in Figure 2. It shows a substantial spread, which goes from a wavenumber of 0.08 to 0.16 (a difference of 100%).

B.2. Biomass & migration speed. Besides a wavenumber, each ecosystem state also has a specific biomass and a specific pattern migration speed. The biomass of regular patterned states has been studied using numerical simulations (23) and more general formulas have been derived for small-wavenumber patterns (32). Both indicate that the biomass (per unit area) is positively correlated with both the wavenumber k of the pattern and the slope parameter v (23); see also Figure 3a. This has a physical interpretation: both steeper slopes and higher wavenumbers (lower wavelengths) reduce the time it takes for water to reach vegetation bands, and thereby reduce water losses during the transportation process. As a result, the vegetation will be able to harvest water from the uphill interbands more effectively. The biomass per wavelength is also of interest. The same studies indicate that the band biomass (per wavelength) is increased when the wavenumber k is decreased and when the slope v is increased. Hence, vegetation bands are expected to have more biomass when other vegetation is farther away, because of the larger (upslope) inter-band area water can be collected from.

The theoretical predictions for migration speed (of a pattern's location) are a bit more subtle. For terrains with a constant slope, numerical simulations have been done (33, 34) and general formulas have been determined for small-wavenumber patterns (32, 35). In these idealized limit cases, migration speed is negatively correlated with wavenumber k and positively correlated with slope v. However, beyond these idealizations, numerical computations show the contour lines are slightly humped, see Figure 3b. This indicates a (slightly) negative correlation between speed and slope v for large slopes.

C. Testable predictions. The theoretical findings in this section lead to predictions that can be confronted with the field data. First of all, the model possesses a Busse balloon, which



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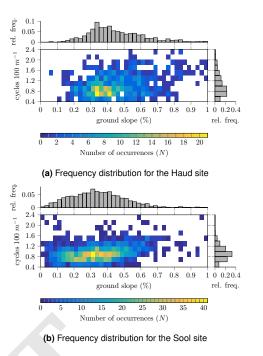
Fig. 3. (slope, wavenumber)-Busse balloon slices for the extended Klausmeier model that include contours for the total biomass (per area) B (a) and the migration speed c (b). Biomass (per area) is positively correlated with both wavenumber k and slope v: the migration speed is negatively correlated with the wavenumber k. Model parameters used: a = 3, m = 0.45, e = 500.

should lead to a wide spread in observable pattern wavenumbers (Figures 1 and 2). Moreover, biomass and migration speed are affected by pattern wavenumber. The biomass (per unit area) is expected to be positively correlated with both the wavenumber and the slope of the terrain (Figure 3a). Migration speed is expected to decrease as a function of pattern wavenumber; the effect of slope on the migration speed is context-specific, as it can be either positive or negative depending on the specific topographical and environmental conditions (Figure 3b).

2. Data acquisition & processing

For this comparison study, two sites were selected in Somalia. The first one $(8^{\circ}0'14'' \text{ to } 8^{\circ}15'11''\text{N}; 47^{\circ}11'54'' \text{ to } 47^{\circ}31'4''\text{E})$ is located in the Haud pastoral region, which will be referred to as the 'Haud' site. The other site (9°18′49″ to 9°34′34″N; 48°8′15" to 48°43′15"E) is located in the Sool-Plateau pastoral area and will be called the 'Sool' site. Both sites mainly exhibit banded vegetation and have ground slopes ranging from zero to one percent. Vegetation mainly constitutes of perennial grasses, which typically have an average life time of one to seven years (36–38). A more detailed description of these sites can be found in SI 2; a map with the location of these sites along with the mean annual rainfall in these areas is shown in Figure S1.

To study the pattern properties in these study areas, each site was divided into square windows (of size $750m \times 750m$ for the Haud site and of size $1010m \times 1010m$ for the Sool site). As has been done in previous studies, the type of pattern (e.g. bare soil, banded vegetation), along with its wavenumber, was determined using spectral analysis (20, 39–41). Only those windows were kept that exhibited banded vegetation with a wavenumber that could be determined with enough certainty (i.e. between 0.4 and 2.5 cycles per 100m). Moreover, windows with a too large curvature were ignored, because the theoretical predictions only apply to terrains with a constant slope. To obtain data on the migration speed of the banded vegetation,



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Fig. 4. Frequency distribution of banded patterns as function of ground slope and wavenumber (number of cycles per 100m) for the Haud site (a) and Sool site (b). The distribution on the right indicates the relative frequency of banded vegetation with corresponding wavenumber. The color gradient indicates the amount of windows (N).

a cross-spectral analysis was performed, along the lines of previous studies (21, 42, 43). A more in-depth explanation of these processing steps can be found in SI 4.

The topographical data used in this article were derived from the ALOS World 3D (AW3D) digital raster elevation model; biomass data for the Haud site have been retrieved from a recently made map on (above-ground) biomass of African savannahs and woodlands (44) (no reliable data for the Sool site was available). Finally, optical data were acquired from various sources: three multispectral WorldView-2 images were mosaicked and used as reference for the Haud site; a panchromatic Ikonos 'Geo' Imagery was acquired for the same site. For the Sool site, six WorldView-2 images were used and a panchromatic SPOT4 image preprocessed to level 2A was used as reference layer (©Cnes 2004 – Spot Image distribution). Moreover, two $7\mu m$ digitized panchromatic declassified Corona spy satellite image, national intelligence reconnaissance system, available from the USGS were obtained for the Haud and the Sool sites. More information about these data sets can be found in SI 3.

3. Results

Empirical Busse balloon. The most prominent pattern prop- 486 erty studied in this article is the pattern wavenumber, which was derived from aerial imagery. The resulting distribution of wavenumbers is reported in Figure 4. These figures show the number of windows that have a particular slope-wavenumber combination. Also given is the relative frequency that indicates the spread of wavenumbers across all windows. This data displays banded vegetation with wavenumbers roughly between 0.4 and 2.0 cycles per 100m. Importantly, this large spread is present for all of the ground slope values which had a representative sample size and could not be explained by

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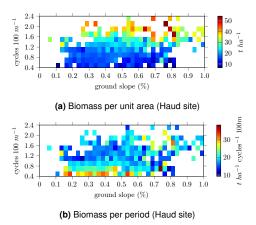


Fig. 5. Biomass distribution per area (a) and per period (b) as a function of ground slope and wavenumber (cycles per 100m) for the Haud site. The color gradient indicates the amount of biomass measured for a particular (slope, wavenumber)-combination.

present heterogeneities in elevation or rainfall. This shows that for a given environmental condition not a single wavenumber pattern, but rather multiple patterns spanning a sizable range of wavenumbers are observable. Additionally, measurements used to determine the migration speed show barely any changes in wavenumber over the scope of 39 years (consistent with (43)), indicating that these patterns are in fact quite stable. Therefore, the observations are in agreement with the existence of a Busse balloon in the real ecosystem.

Biomass and migration speed. The processed biomass data for the Haud site is shown in Figure 5. In Figure 5a the relation between biomass per area (in $t ha^{-1}$) is plotted against the ground slope and the wavenumber. From the same data the biomass per period is computed – which is biomass per area divided by the window's wavenumber. The resulting plot is given in Figure 5b. The measurements of biomass show agreement with theoretical predictions of model studies; in both, the total biomass increases (all slopes: r = 0.80, n = 714, P < 0.001; linear regression) and the biomass per period decreases when the wavenumber increases (all slopes: r = -0.30, n = 714, P < 0.001; linear regression). However, a more in-depth inspection reveals disagreements. For one, the effect of ground slope is not strongly present in the data, though its effect is clear in the extended-Klausmeier model (Figure 3a). Additionally, the more refined details of wavenumber dependence also differ (it is concave in the theoretical model and convex in the real-life data).

The migration speed is plotted in Figure 6 for both the Haud and the Sool sites. These measurements show an increase in speed when the wavenumber decreases (Haud: r=-0.65, n=104, P<0.001; Sool: r=-0.67, n=79, P<0.001; linear regression), corroborating theoretical predictions (see Figure 3b).

4. Discussion

Leading ecological frameworks emphasize the potential role of regular spatial vegetation patterns as indicators for proximity to catastrophic ecosystem shifts (11, 45). In these frameworks, however, mono-stability of patterns is implicitly assumed, suggesting that for a given environmental condition there is only

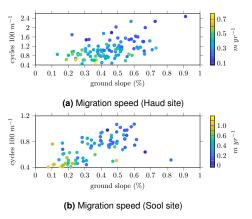


Fig. 6. Observed (average) migration speed of vegetation bands in the Haud (a) and the Sool (b) sites over the course of 39 years as a function of ground slope and wavenumber (cycles per 100m). The colour gradient indicates the migration speed for a particular (*slope*, *wavenumber*)-combination. The sign indicates the direction of migration relative to the slope, with positive and negative values indicating upslope and downslope migration respectively.

one stable vegetated state, i.e. a single pattern with a specific wavelength (11, 45). Subsequent theoretical insights have challenged this view, highlighting the possibility of multi-stability of patterns, bounded by the so-called Busse balloon. In this study, we provide the first empirical evidence corroborating the existence of a Busse balloon for stable vegetation patterns in dryland ecosystems. Specifically, our two study sites in Somalia revealed the sustained (i.e. over a 39-year period) co-occurrence of banded vegetation with wavenumbers varying over a substantial range. Our findings have major implications for the way in which vegetation patterns indicate ecosystem resilience and mediate ecosystem responses to environmental change.

Specifically, the existence of a Busse balloon implies that an ecosystem's resilience can no longer merely be defined by the magnitude of environmental change it can cope with (46). In these systems there is not one tipping point, but a cascade of destabilizations – indicated by the boundary of the Busse balloon. When environmental changes push a patterned ecosystem beyond the boundary of the Busse balloon, a wavelength adaptation occurs, and typically part of the vegetation patches are lost, while the remaining patches grow in size. The extent of these adaptations depends on the rate of environmental change (23, 26, 47, 48). Moreover, human activities or natural variations can cause local disturbances, diminishing the regularity of ecosystem patterns. The recovery process from such disturbances may involve a rearrangement of patches in the landscape (23, 32). Again, the extent to which such recovery is possible depends on the rate of environmental change that the ecosystem is exposed to (47). Hence, the existence of a Busse balloon of stable dryland vegetation patterns suggests that adaptability of patches to changing environmental conditions provides a more comprehensive indicator for the ecosystem's resilience than the shape of the pattern itself, as suggested in current leading frameworks (11, 45). Future studies in this direction should provide a more thorough understanding of what determines the spatial rearrangement of vegetation patches resulting from disturbances, environmental changes and spatial heterogeneities in the landscape.

The pattern-oriented modeling approach was mainly devel-

oped to aid model development and design, but the approach can also be used to evaluate the success of existing models to explain multiple strong and weak patterns observed (7). This so-called 'reverse pattern-oriented modeling' approach (7) was used in the current study. Such systematic comparisons between model predictions and empirical data can be part of an iterative process toward further model improvement (5, 6). In this context, it is interesting to note the discrepancy that we observed between model predictions and field measurements of the influence of the ground slope on pattern migration speeds. Because topography critically changes the distribution of water within ecosystems, it also alters the migration speed of patterns. Therefore, future model developments should relax the assumption of uniform slopes, and examine the effects of more complex topographies for dryland ecosystem dynamics.

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Since their appearance on aerial photographs in the 1950s (49), the origin of regular vegetation patterns in dryland ecosystems has been a topic of fascination within the scientific community. The study of these patterns through reaction-diffusion modeling subsequently highlighted the importance of these patterns for the functioning of dryland ecosystems, and

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their response to environmental change. The recent increase in the availability of optical and topographical data provides unprecedented opportunities to confront model predictions with empirical data (6, 31). In this study, we combined these data sources with in-situ measurements of biomass, enabling the comparison of multiple pattern characteristics of Somalia drylands with predictions derived from reaction-diffusion modeling. The empirical evidence corroborates theories of multi-stability of patterned vegetation states, improving our understanding of these systems' resilience to environmental change. In addition, our results call for more detailed investigations of the role of small-scale topographic variability in pattern formation and migration, to be undertaken in future studies.

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