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► **To cite this version:**

Rémi Prudhomme, Adriana De Palma, Patrice Dumas, Ricardo Gonzalez, Paul Leadley, et al.. Combining mitigation strategies to increase co-benefits for biodiversity and food security. *Environmental Research Letters*, 2020, 15 (11), pp.114005. 10.1088/1748-9326/abb10a . hal-03206590

HAL Id: hal-03206590

<https://hal.science/hal-03206590>

Submitted on 21 Feb 2024

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To cite this article before publication: Rémi Prudhomme *et al* 2020 *Environ. Res. Lett.* in press <https://doi.org/10.1088/1748-9326/abb10a>

Manuscript version: Accepted Manuscript

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Combining mitigation strategies to increase co-benefits for biodiversity and food security

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Received xxxxxx

Accepted for publication xxxxxx

Published xxxxxx

Length:4143

Abstract

World agriculture needs to find the right balance to cope with the trilemma between feeding a growing population, reducing its impact on biodiversity and minimizing greenhouse gas (GHG) emissions. In this paper, we evaluate a broad range of scenarios that achieve 4.3 GtCO_{2,eq}/year GHG mitigation in the Agriculture, Forestry and Other Land-Use (AFOLU) sector by 2100. Scenarios include varying mixes of three GHG mitigation policies: second-generation biofuel production, dietary change and reforestation of pasture. We find that focusing mitigation on a single policy can lead to positive results for a single indicator of food security or biodiversity conservation, but with significant negative side effects on others. A balanced portfolio of all three mitigation policies, while not optimal for any single criterion, minimizes trade-offs by avoiding large negative effects on food security and biodiversity conservation. At the regional scale, the trade-off seen globally between biodiversity and food security is nuanced by different regional contexts.

Keywords: Mitigation, Global scale, Land-use, Trade-off, Dietary change, Reforestation, Bioenergy, Food security, Biodiversity

1. Introduction

Land is a multi-purpose asset that may involve conflicts in its use. Formerly restricted to the local level, conflicts have become increasingly global over the last few decades because of the rapid intensification of international exchanges (Liu *et al* 2013). Currently, the joint challenges of global food security, climate change mitigation and conservation of biodiversity give a new

dimension to this issue, involving new types of trade-offs and synergies while strengthening the global dimension.

Assessments based on global land-use models have shown that mitigation policies relying on large-scale second-generation biofuel production have important environmental implications and, especially if forest protection measures are implemented, adversely impact food prices (Popp *et al* 2011, Humpenöder *et al*

2018, Heck *et al* 2018a). Afforestation is also associated with significant increase in food prices (Kreidenweis *et al* 2016) whereas dietary change policies may have the opposite effect (Stevanović *et al* 2017). Combining measures appears to be an appropriate solution to minimize negative effects, but the nature of the combinations promotes either biodiversity or food security (Humpenöder *et al* 2018, Obersteiner *et al* 2016, Visconti *et al* 2016).

Trade-offs between biodiversity and climate mitigation needed to be considered. While some mitigation policies such as carbon storage in forests can maintain biodiversity (Watson *et al* 2018), other options could increase pressure on biodiversity. Strong climate change mitigation scenarios relying on bioenergy are typically harmful to biodiversity either due to land-use change related to second-generation biofuel production (Hill *et al* 2018, Newbold 2018, Shukla *et al* 2019) or due to increased wood harvest for fuel in biodiversity hotspots (Jantz *et al* 2015). However, scenarios without strong climate mitigation are also associated with high impacts of climate change on biodiversity, especially in the second part of the century (Newbold 2018).

Combining on a global scale a model of agricultural intensification with a statistical model of biodiversity provides a unique framework for understanding (i) the impact of different GHG mitigation policies (second-generation biofuel production, dietary change and reforestation of pastures) on both biodiversity and food security and (ii) the degree of conflict or synergy between such policies.

2. Method

2.1 Overview of the modelling framework

The food system is represented by the Nexus Land-Use (NLU) model (Souty *et al* 2012). This global model of agricultural intensification describes the worldwide land-use system, computes cost-optimal food security indicators (average cost of production per calorie produced and food price per calorie produced), calculates associated agricultural and land-use change with respect to GHG emission goals and generates land-use maps (Souty *et al* 2012).

These land-use maps are converted into impacts on biodiversity through global estimation of two indicators of local biodiversity - Biodiversity Intactness Index (BII) (De Palma *et al* 2019) and within-sample Species Richness (SR) - applying a mixed-effect modelling structure based on the PREDICTS database (Hudson *et al* 2017). BII indicates the average abundance of a large, diverse set of native species in a given area, relative to their abundance in a pristine reference condition (Scholes and Biggs 2005) (See the Method section). The SR reports the number of species, relative to the number expected in a natural system (Section 6 in supporting information). These two indicators can provide complementary insights as they address both the number of species and differences in the composition of ecological communities present in ecosystems (Section 6 in supporting information). To clarify the impacts of GHG mitigation policies on these indicators, we made some changes to the framework used in Hill *et al* 2018: represented grassy and woody second-generation biofuel as highly intensified perennials.

Using this framework, we assess the impact on biodiversity and food security of land-use-based mitigation scenarios that provide mitigation of 4.3 GtCO_{2,eq}/year in 2100 (which is the AFOLU sector's share of the mitigation needed to limit global warming to 2° in 2100: Wollenberg *et al.* 2016). This mitigation target is calculated based on a methodology detailed in supporting information Table 2. We set a common target for every scenario to make their impacts on biodiversity and food security comparable.

We infer from these scenarios whether the relationship between biodiversity and food security in the presence of mitigation policies is synergistic or antagonistic and how the policy mix influences this relationship.

To mitigate the 4.3 GtCO_{2,eq}/year in 2100, we built scenarios that are combinations of second-generation biofuel production (between 0 and 112 EJ/year in 2100), dietary change (a convergence towards the consumption of 432 kcal/capita/day of animal products which is a reduction except in Africa for nutritional reasons) and reforesting pastures (between 0 and 31% of global pasture reforested). Each mitigation scenario

is detailed in supporting information. The mitigation effort of each of these policies (second-generation biofuel production, dietary change and reforestation of pastures) is then defined as the percentage of each policy in total mitigated emissions (Section 7 in supporting information). To cover a broad range of scenarios and represent a uniform distribution of mitigation policies (second-generation biofuel, dietary change and reforestation), the scenarios are constructed according to a full factorial design (Section 8 and 9 in supporting information). The experimental design involves taking mitigation efforts ranging from 0 to 100% for each policy in 10% steps while keeping the sum of efforts equal to 100% (Section 8 in supporting information).

Finally, we detail the distribution of these impacts across 12 large regions of the world. In this study, the mitigation effort is unequally distributed among the regions and depends on the amount of pasture available to reforest, the current diet and the regional cost of second-generation biofuel production. To compare the impacts of these heterogeneous mitigation efforts between regions and with the global figures, we calculate the relative change in biodiversity and food security divided by the relative change in regional emissions (Section 6 in supporting information for details of these indicators).

This downscaling highlights the influence of the regional context on the sensitivity of responses of both regional biodiversity and food security to mitigation policies.

2.2 Description of the NLU and PREDICTS models

2.2.1 Estimating agricultural production

The global Nexus Land-Use model (NLU) is used to represent the agricultural sector (see Souty *et al* 2012 for more details). It allows us to represent agricultural intensification and the distribution of cropland, pastures and forest at the global scale. Crop intensification is explicitly represented in NLU with a concave production function and fertilizer prices are computed from energy prices (Brunelle *et al* 2015). Two livestock systems are considered: a grass-based system and a mixed crop-livestock system.

Regional production cost is minimized under a supply-use equilibrium with a simplified representation of international trade. Based on an interpretation of the Ricardian theory, the boundary between the mixed crop-livestock system and the grass-fed livestock system changes according to the equalization of rent. In the mixed crop-livestock system, cropland distribution is based on potential yield, with rent increasing with land quality. In this model forest area is exogenously defined by scenarios. A detail description of these elements is provided in section 1 in supporting information.

2.2.2 Estimating agricultural emissions.

Agricultural emissions are calculated by NLU using the IPCC Tier 1 method for production in the plant food sector and the IPCC Tier 2 method for the livestock sector (IPCC 2006). In the livestock sector, emissions from manure management (CH₄ and N₂O) and enteric fermentation (CH₄) are computed. In the plant food sector, emissions from fertilization (N₂O) and rice cultivation (CH₄) are computed. Carbon dioxide (CO₂) emissions are also computed for land-use changes (Le Quéré *et al* 2009) and for fossil fuel substituted by second-generation biofuel (detailed in the description of biofuel scenarios in supporting information).

2.2.3 Estimating biodiversity impacts.

Biodiversity impacts are estimated by the PREDICTS modelling framework (Purvis *et al* 2018) which considers land-use to be the main driver of biodiversity losses (Díaz *et al* 2019).

The statistical models linking biodiversity to drivers are underpinned by a large, global and taxonomically broad database of terrestrial ecological communities facing land-use pressures. Among the biodiversity models provided by the PREDICTS framework, we chose BII because of its use in the Planetary Boundaries framework and SR because of its wide use despite its known limitations. The species richness model (SR) is a mixed-effect model computing the number of species present in a given area. The total abundance model computes the sum of all individuals of all species present in the ecosystem. The compositional similarity model computes the percentage of individuals common to the studied ecosystem and the reference ecosystem

for each grid of a 0.5° map. The abundance map was then multiplied by the compositional similarity map to produce the map of abundance-based BII (De Palma *et al* 2019) (Section 6 in supporting information). These three PREDICTS models include different levels of management (intensive, light or minimal) and different types of land cover (forest, pasture, rangeland, annual cropland, perennial cropland and urban zones).

2.2.4 Linking PREDICTS and NLU. In NLU, 60 land classes are defined in the reference year according to their potential yield. Different crop types are defined for each land-class: “Dynamic” crops and “other” crops (See supporting information).

In PREDICTS, three levels of intensification break down perennial crops, annual crops and nitrogen-fixing crops into a “minimal”, “light” and “intense” use category.

NLU crop types are aggregated into a single category and then split into PREDICTS crops categories (perennial, annual and nitrogen-fixing crops) based on their relative proportion of the crop mix in the reference year.

For the reference year, a Generalized Additive Model (GAM) is computed to match the relative proportion of “minimal”, “light” and “intense” cropland with the 60 NLU land classes (See supporting information, Fig. 7). A GAM is used to avoid making assumptions about the form of the relationship between the intensification in PREDICTS and the land classes of NLU and to avoid giving too much weight in the relationship to uncertain extreme values (Section 4 in supporting information).

Pastures in NLU mixed crop-livestock and pastoral production systems are aggregated into a single pasture category. In PREDICTS, pastures include rangeland, “light” and “intense” pastures. Among the aggregated pasture category of NLU, rangeland areas are defined on the basis of reference rangeland map (Hurtt *et al* 2011). For the remaining pastures, livestock density is defined on the basis of livestock density maps (Robinson *et al* 2014). In the reference year, a GAM is computed to match the relative proportion of “light” and “intense” pasture with livestock density maps (See supporting information, Fig. 8).

2.2.5 Baseline scenario. The population follows changes in the Shared Socio-economic Pathway (SSP2) (Riahi *et al* 2017). Food demand follows FAO projections (Alexandratos and Bruinsma 2012a) with a global mean food supply at the household level in 2100 of 2585 kcal/capita/day of vegetable products and 615 kcal/capita/day of animal products. International trade parameters are kept constant. Forest, which is exogenous to the model, follows current trends (Hurtt *et al* 2011) until 2050 and then stabilizes. Fertilizer prices are computed based on energy prices (Brunelle *et al* 2015) taken from the baseline of IMACLIM-R (Waisman *et al* 2012). This leads to a global average calorie price of 79\$/Mkcal, a global average production cost of 43\$/Mkcal, a 14% reduction of BII by land-use change occurring between 2001 and 2100 and a 15% reduction of species richness by land-use change occurring between 2001 and 2100 compared to the biodiversity levels in 2001.

2.2.6. Mitigation scenarios to achieve 2°C of global warming in 2100. We combine three mitigation policies in mitigation scenarios to achieve 4.3 GtCO_{2,eq}/year of mitigated emissions in 2100. 4.3 GtCO_{2,eq} /year is the target for the AFOLU sector to achieve 2°C of global warming. We deduced this target by applying the share of mitigated emissions by the AFOLU sector in overall mitigated emissions between the RCP2.6 and the baseline 2030 (Wollenberg *et al* 2016) to mitigated emissions between the RCP2.6 and the baseline of the marker model IMAGE (Gidden *et al* 2019) in 2100 (See Table 2 in supporting information) .

To obtain a broad representation of the possible combinations between second-generation biofuel production, dietary change and reforestation, we use a full factorial design (see Fig.11 in supporting information), which covers second-generation biofuel production ranging from 0 to 112 EJ, animal product consumption ranging from FAO trends (Alexandratos and Bruinsma 2012b) to a convergence towards 432 kcal/capita/year (see supporting information, Table 2), and pasture reforestation ranging from 0% to 31% (see supporting information, Table 3).

To achieve 4.3 GtCO_{2,eq}/year of mitigated emissions by means of dietary change, we replace the consumption

of animal products by plant products in the Agrimonde scenarios called AG1 (Paillard *et al* 2014). This leads to a convergence of the overall animal consumption towards 432 kcal/capita/day in all regions. The consumption of ruminant products obtained is 183 kcal/capita/year in 2050 for Brazil, Canada, Europe, USA, FSU, OECD Pacific and Rest of LAM, 91 kcal/capita/year in 2050 for India, Rest of Asia and China, 154 kcal/capita/year for Middle-East and 65 kcal/capita/year for Africa (Section 10.1 in supporting information). The rest of animal product consumption (in the 432 kcal/capita/day) is composed of monogastric and aquatic products (See supporting information, Table 2).

The reforestation scenario follows the same philosophy as the natural climate solutions reforestation scenario presented in (Griscom *et al* 2017) by reforesting pastures (Section 10.3 in supporting information). The 31 % of pastures reforested in the world corresponds to the reforestation of 186 Mha with a carbon sequestration of 23 tCO₂/ha in global average (See Table 5 in supporting information). This reforestation area is in the lower range of afforestation potential in baseline (Intergovernmental Panel on Climate Change 2014) (See supporting information for more details about the regional distribution, Table 3).

The second-generation biofuel production of 112 EJ is coherent with the literature that agrees on a technical bioenergy potential of at least 100EJ (Edenhofer *et al* 2011, Creutzig *et al* 2015). To limit the competition of the biofuel production with food, second-generation biofuels in the scenarios are deployed in the form of grassy crops in Europe and the USA, and in the form of woody crops in the rest of the world (Section 10.2 in supporting information).

2.2.7. Indicators of food security and biodiversity. We use four indicators to represent impacts of mitigation policies on biodiversity and food security:

- Global food price (\$/Mkcal) (Section 6 in supporting information).
- Crop production cost per unit of food energy produced (\$/Mkcal) (Section 6 in supporting information)

- Species richness (Section 2.2.3 in the method section)
- Biodiversity intactness index (BII) (Section 2.2.3 in the method section)

3. Results

The scatter of points representing the impacts of land-based mitigation scenarios is widely spread over the output space and has concave boundaries, indicating a moderate trade-off between biodiversity and food security for a greenhouse gas reduction objective compatible with 2 degrees of global warming (Fig. 1 and see supporting information for other indicators). A table presenting the values of the four indicators (calorie price, cost, BII and SR) for each scenario is provided in section 13 in the supporting information.

Scenarios with high second-generation biofuel production are located well inside the cloud of points. This shows that second-generation biofuel production is a less effective mitigation option for reconciling biodiversity and food security objectives than scenarios containing more reforestation or dietary change (Fig. 1 and see supporting information for other indicators). Moreover, scenarios with low levels of biodiversity (especially low SR) tend to be those including high levels of second-generation biofuel production (supporting information Fig. 17).

Mitigation scenarios focusing almost exclusively on dietary change or reforestation are either at the upper right hand or at the lower left hand of the cloud of points. This indicates that they perform well in relation to one indicator but have negative side-effects on at least one of the others (See Fig. 14 in supplementary information). The reforestation of large proportions of the world's pastures is positive for both biodiversity indicators but causes a sharp increase in food prices and food cost, thus threatening food security (supporting information Fig. 14). By contrast, scenarios with significant dietary changes are less positive in terms of biodiversity but have lower impacts on food prices and food production costs (Fig. 1 and supporting information Fig. 17). The cost of food production and the price of food vary in the same direction across mitigation strategies but with larger change in price (Section 11.1 of the SI for an explanation of the

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3 difference between response of food price and cost).
4 The impacts of a dietary change on biodiversity vary
5 according to the indicator under consideration with a
6 reduction in BII (-0.7 % per Gt CO_{2,eq}) and an increase in
7 SR (+0.3% per Gt CO_{2,eq}) compared to the baseline. A
8 major dietary change leads to less intensification of
9 livestock farming, which in NLU model results in an
10 increase in the pasture area without any influence on
11 forest area (Section 12.3 in supporting information).
12 Grazing land has a high species richness but its
13 ecological communities have little similarity with those
14 found in natural environments, which explains why
15 there is a low BII and high SR in a dietary change
16 scenario.

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22 [Insert Figure 1 here (See Figure section after the
23 reference section)]
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27
28 Figure 1. Impacts of mitigation scenarios achieving 4.3
29 GtCO_{2,eq}/year of mitigated emissions in 2100 based on
30 combinations of second-generation biofuel production, dietary
31 change and reforestation. Outputs are presented as the relative
32 change in biodiversity intactness index (BII) and food price in the
33 climate mitigation scenarios with respect to the baseline (no
34 mitigation policy). At the global scale, all climate mitigation
35 scenarios reduce emissions by construction of baseline emissions
36 (13.9 GtCO_{2,eq} in 2100) by around 4.3 GtCO_{2,eq}/year in 2100. Each
37 climate mitigation scenario is coloured according to the dominant
38 mitigation policy. The dominant emission reduction policy in each
39 scenario is the one that contributes the most to emissions
40 reductions. "Others" in the legend represents scenarios where no
41 option accounts for more than 50% of the mitigation effort. The
42 percent of mitigation policy is the share of mitigated emissions of
43 the dominant mitigation policy (second-generation biofuel
44 production, dietary change and reforestation). The food price and
45 the BII in baseline in 2100 are also indicated.

46
47 [Insert Figure 2 here (See Figure section after the
48 reference section)]
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50
51 Figure 2. Impacts of mitigation scenarios reaching 4.3
52 GtCO_{2,eq}/year in 2100 of mitigated emissions based on
53 combinations of second-generation biofuel production, dietary
54 change and reforestation. Outputs are presented as the relative
55 change in Species Richness (SR) and cost of food production in the
56 climate mitigation scenarios with respect to the baseline (no
57 mitigation policy). Each climate mitigation scenario is coloured
58 according to the dominant mitigation policy. The dominant emission
59

reduction policy in each scenario is the one that contributes the
most of emissions reductions. "Others" in the legend represents
scenarios where no option accounts for more than 50% of the
mitigation effort. For additional details, see Fig. 1. The food
production cost and the SR in baseline in 2100 are also indicated.
Impacts of mitigation on other indicators are provided in section 12
in supporting information.

3.1 Portfolios of land-use-based mitigation scenarios reduce the trade-off between biodiversity and food security

On a global scale, mitigation scenarios that spread mitigation efforts between several policies (reforestation, second-generation biofuel production and dietary change) avoid extreme negative side effects. Among these scenarios involving different emission mitigation policies, some mitigation scenarios can improve both the protection of biodiversity and food security in 2100 compared to the baseline without mitigation policies (scenarios in the upper left-hand quadrant of the Fig. 1).

These mitigation scenarios are mainly mixes of reforestation and dietary change associated with low second-generation biofuel production. For example second-generation biofuel production of 10 EJ/year in 2100 (10% of the mitigation effort) associated with reforestation of 11 % of pasture (40% of the mitigation effort) and a low consumption of ruminant product of 150 kcal/capita/day (50% of the mitigation effort) decrease the food price by 13% compared to the baseline and increase BII by 1.2% for a decrease of 30% of emissions in the AFOLU sector compared to the baseline (Fig. 1).

3.2 Trade-off and synergies between food security and biodiversity conservation in mitigation policies at the regional scale

The trade-off between BII and food prices seen at a global scale is also found within some regions such as the region covering Japan, South Korea, Australia and New Zealand (OECD Pacific in Fig. 3), Europe, the USA and China, but differs in other regions because of the regional context. For example, food prices and BII are relatively insensitive to mitigation strategies in the former Soviet Union. In Canada and the Middle East, food prices, but not BII are very sensitive to mitigation

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3 strategies. Due to the small fraction of agricultural land
4 in these regions (Hurtt *et al* 2011), their average
5 regional levels of biodiversity are mainly influenced by
6 the state of their natural areas and not by agricultural
7 land-use changes (Fig. 3).

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10 The choice of the optimal regional mitigation mix for BII
11 and the food price varies from one region to another.
12 On the one hand, large-scale bioenergy production
13 systematically increases food prices and reduces
14 biodiversity compared to a baseline without a
15 mitigation strategy. On the other hand, the influence of
16 reforestation and diet change in the mitigation strategy
17 depends on the regional context. For example, a lower
18 ruminant consumption decreases in the African dietary
19 change (-46 kcal/capita/day) compared to the other
20 regions (for example, -199 kcal/capita/day in Brazil in
21 Table 2) leads to high levels of BII and a low food price
22 (Fig. 3). For other regions like Brazil and the rest of Latin
23 America, both dietary change and reforestation
24 mitigation scenarios lead to higher price than the
25 baseline without a mitigation strategy.
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31 [Insert Figure 3 here (See Figure section after the
32 reference section)]
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38 Figure 3. Regional impacts of mitigation scenarios
39 reaching 4.3 GtCO₂,eq/year in 2100 of mitigated
40 emissions globally based on combinations of second-
41 generation biofuel production, dietary change and
42 reforestation. Outputs are presented as the relative
43 change in biodiversity intactness index (BII) and food
44 price in the climate mitigation scenarios with respect
45 to the baseline (no mitigation policy). These indicators
46 compare changes in relative BII and relative food
47 prices by considering different amounts of regionally
48 mitigated emissions indicated for each region by an
49 average of mitigated emission in the different
50 mitigation scenarios. These indicators are also
51 provided at the global scale in Fig. 1. Similar graphs for
52 other biodiversity and food security indicators are
53 provided in supporting information
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4. Discussion and conclusion

The major contribution of this study is that it is the first
to explore the full range of combinations of key land-
based climate mitigation options – bioenergy,
reforestation and dietary change – on biodiversity and
food systems. The model projections are not
predictions of future outcomes but do provide insight
into the synergies and trade-offs among land-based
mitigation measures, as well as their respective
advantages and drawbacks. Another important novelty
of this study is that it provides a global perspective of
the impact of agricultural intensification and land use
changes within the agricultural sector on biodiversity of
both agricultural intensification and land use changes
within the agricultural sector (conversion of the
pastoral system into a mixed pasture-and-crop system),
while previous studies have focused on the impact of
mitigation scenarios on habitats of high ecological value
such as "biodiversity hotspots" (Obersteiner *et al* 2016)
or forests (Humpenöder *et al* 2018). Considering the
impact of agricultural intensification on biodiversity
provides several new insights. For example, the
substantial agricultural intensification induced by a
reforestation scenario mitigates the initial BII increase
inside the reforested area. In addition, the reduction in
extent of the crop-pasture mix system in favour of the
pastoral system in scenarios of significant dietary
change may lead to strong change in ecological
community composition, as evidenced by the reduction
in BII (Fig. 1). However, as the reforestation rates are
exogenously set in NLU, the reduction in extent of the
crop-pasture mix system is probably overestimated in
this study, leading to an underestimation of the BII and
the food price (Section 11 of the supporting information
for a description of main mechanisms).

Another major contribution of this study is an
understanding of the impacts of different land-use-
based mitigation scenarios on different biodiversity
indicators: (i) the "naturalness" of ecosystems through
the BII and (ii) the local "extirpation risk" through the BII
and SR (Karp *et al* 2015). Although we do not estimate
extinction rates in this study, the biodiversity indicators
computed in our mitigation scenarios provide
additional and consistent information to the extinction
risk in global biodiversity hotspots already studied by,

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2
3 e.g., Obersteiner *et al* 2016. Reforestation scenarios are
4 beneficial to these three indicators, second-generation
5 biofuel is detrimental to these three indicators and
6 decreasing pressure on land through dietary change has
7 a beneficial effect on SR and biodiversity hotspot
8 preservation but decreases BII due to an increase in the
9 area of pasture.

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12 The inclusion of the impacts of these policies on
13 biodiversity is a first step towards a deeper integration
14 of biodiversity into the socio-ecological system used in
15 environmental assessment of mitigation options. The
16 crucial role of biodiversity in food production is well
17 established and its integration into land-use models can
18 significantly change the relationship between
19 biodiversity protection and food security (Commission
20 on Genetic Resources for Food and Agriculture and
21 Food and Agriculture Organization of the United
22 Nations 2019).

23
24
25 A portfolio of mitigation strategies reduces side-effects
26 on biodiversity and food security compared to siloed
27 strategies and allows several SDGs to be achieved
28 simultaneously (Humpenöder *et al* 2018, Obersteiner
29 *et al* 2016, Bertram *et al* 2018, Minx *et al* 2018). For
30 example, reforestation of 22% of pasture (70% of the
31 mitigation effort) and a dietary change of 90
32 kcal/capita/day from ruminant toward plant
33 consumption (30% of the mitigation effort) is the best
34 scenario to minimize negative impacts on our measures
35 of biodiversity, food security and mitigation in the
36 agricultural sector at the global scale. The portfolio
37 effect is explained in this scenario by the
38 complementarity of mitigation policies. The synergy is
39 particularly strong between dietary change and
40 reforestation strategies, as this combination allows for
41 land to be spared through a reduction in overall food
42 production, meaning it can both store carbon and
43 preserve biodiversity (Stevanović *et al* 2017, Herrero
44 *et al* 2016, Ewers *et al* 2009). On the other hand, the
45 increase in second-generation biofuel production
46 reduces the positive synergies between food security
47 and biodiversity conservation even with an optimistic
48 assumption about the quantity of emissions reduced
49 per unit of second-generation biofuel produced
50 (Searchinger *et al* 2018).

In this study, mitigation effort is allocated between
regions according to reforestation potential, biofuel
prices and the difference between local diet and a
reference diet, but without considering the equitability
or mitigation cost of this distribution of the effort. The
relationships between biodiversity and food security we
report could change when these allocation criteria are
considered. Moreover, the potential for mitigation of
emissions, food insecurity and biodiversity loss in the
AFOLU sector, although very high (Heck *et al* 2018b,
Tubiello *et al* 2015, Tilman *et al* 2017), may not be
exploited due to equitability of the allocation of effort
or high mitigation costs (Tilman *et al* 2017, van den Berg
et al 2019, Markel *et al* 2018).

In this study, we show the importance of considering
the regional context, which strongly nuances the global
trade-offs between biodiversity protection and food
security protection. This study should therefore be
complemented by future work that take into account
the regional context. More specifically, soil carbon
sequestration (Lal 2004) in regions with degraded soils
such as southern Europe, some parts of Asia and Africa,
or increased Nitrogen Use Efficiency (NUE) (Zhang *et al*
2015, Bodirsky *et al* 2014) in regions with low NUE such
as China or India. Other important dimensions could be
added to our analysis. For instance, the nutritional
qualities and health benefits of food diets could be also
considered in relation to recent work on the topic; and
more location-sensitive measures of biodiversity, such
as extinction rates or extinction risk, could usefully be
added.

Our study shows the importance of combining
exploratory with target-seeking scenarios to include
new objectives such as we have done here with
biodiversity. This approach differs from many others
used in climate scenarios that select the scenario with
the lowest implementation cost regardless of
preferences toward other objectives, as biodiversity
conservation or food security. For example, the RCP2.6
scenario (Vuuren *et al* 2011), implying an important
second-generation biofuel production (equivalent to
181 E_j) leads to relatively low food prices at the expense
of low SR levels (See Table 3). In this scenario, the
negative effect on biodiversity is mainly due to the
significant production of second-generation biofuel (Hill

et al 2018, Jantz *et al* 2015). In contrast, our approach allows the assessment of a wide variety of combinations of mitigation policies and does not make implicit assumptions about preferences between biodiversity and food security.

Acknowledgements

This work was supported by a doctoral school ABIES grant (provided by AgroParistech). This article has also benefited from the support of the LabEx BASC (ANR-11-LABX-0034), the Long-term modelling chair for sustainable development (Ponts Paristech-Mines Paristech) funded by ADEME, Grt-Gaz, Schneider Electric, EDF, French Environment Ministry, the aid from the French state managed by the ANR under the "Investissements d'avenir" programme (ANR-16-CONV-0003) and the funding by the NERC (award NE/M014533/1).

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Figure list:

Fig. 1:

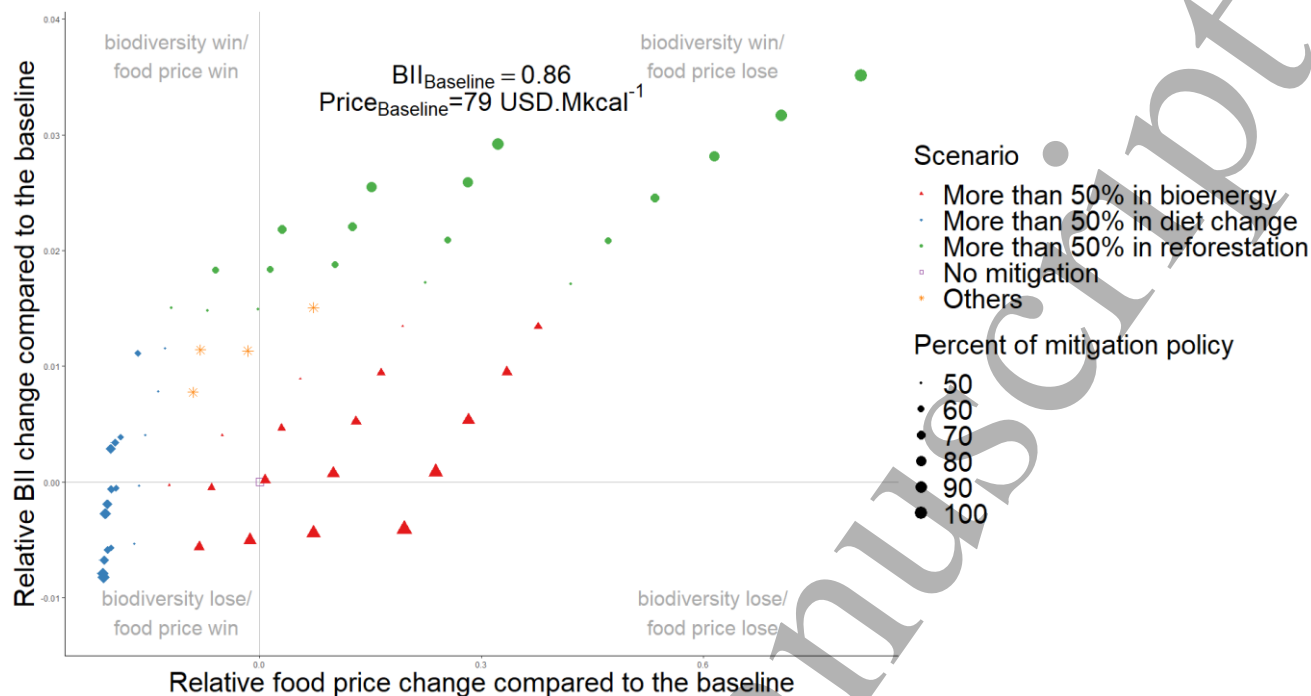


Fig. 2:

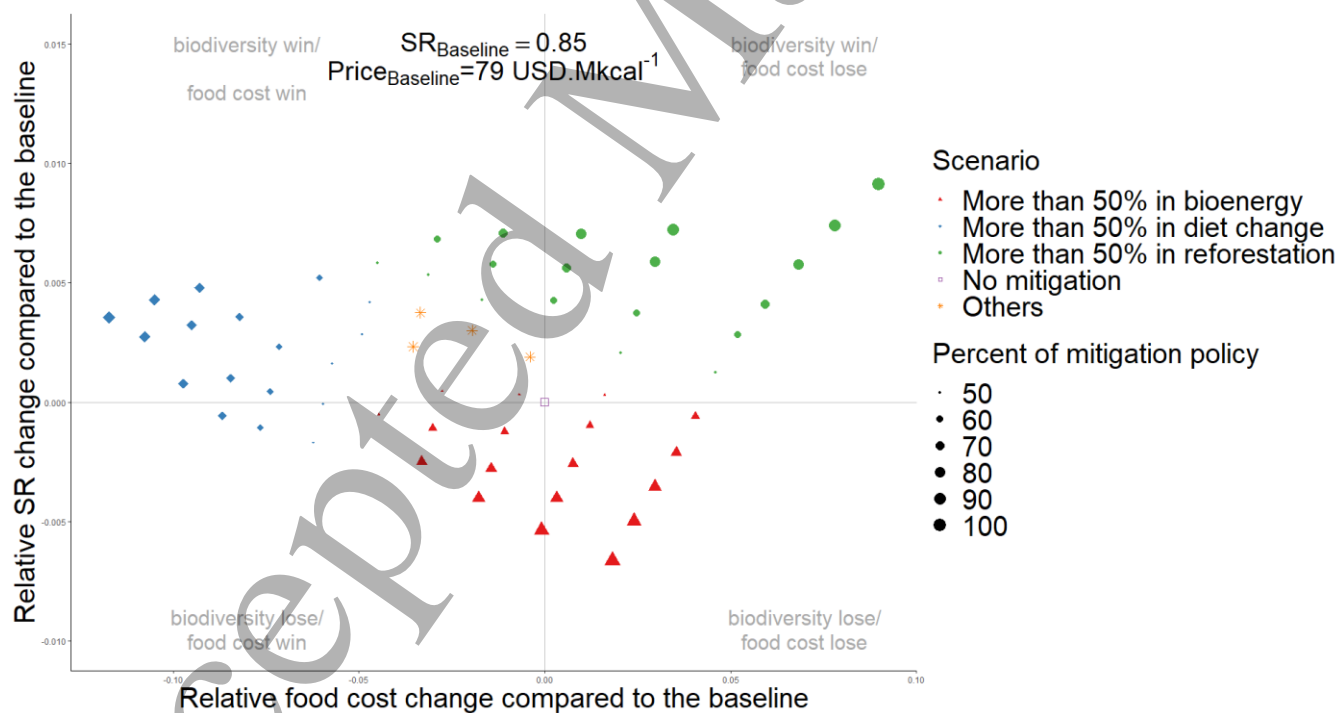


Fig. 3:

