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Asma Grira, M. Antiñolo, André Canosa, Alexandre Tomas, E. Jiménez, et al.. An experimental study of the gas-phase reaction between Cl atoms and trans-2-pentenal: Kinetics, products and SOA formation. Chemosphere, 2021, 276, pp.130193. 10.1016/j.chemosphere.2021.130193. hal-03195966

HAL Id: hal-03195966 https://hal.science/hal-03195966v1

Submitted on 21 Apr 2021

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1 2	An experimental study of the gas-phase reaction between Cl atoms and <i>trans-2</i> -pentenal: kinetics, products and SOA formation
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23 Abstract

The gas-phase reaction of trans-2-pentenal (T2P) with Cl atoms was studied at atmospheric 24 pressure and room temperature. A rate coefficient of $(2.56\pm0.83)\times10^{-10}$ cm³ molecule⁻¹ s⁻¹ was 25 obtained using the relative rate method and isoprene, cyclohexane and ethanol as reference 26 compounds. The kinetic study was carried out using a 300-L Teflon bag simulation chamber 27 (IMT Lille Douai-France) and a 16-L Pyrex cell (UCLM-Ciudad Real-Spain), both coupled to 28 the Fourier transform infrared (FTIR) technique. Gas-phase products and secondary organic 29 aerosol (SOA) formation were studied at UCLM using a 16-L Pyrex cell and a 264-L quartz 30 31 simulation chamber coupled to the FTIR and gas-chromatography-mass spectrometry (GC-MS) techniques. HCl, CO, and propanal were identified as products formed from the studied 32 reaction and quantified by FTIR, being the molar yield of the latter (5.2±0.2)%. Formic acid 33 34 was identified as a secondary product and was quantified by FTIR with a yield of (6.2 ± 0.4) %. In addition, 2-chlorobutanal and 2-pentenoic acid were identified, but not quantified, by GC-35 36 MS as products. The SOA formation was investigated using a fast mobility particle sizer 37 spectrometer. The observed SOA yields reached maximum values of around 7% at high particle mass concentrations. This work provides the first study of the formation of gaseous 38 39 and particulate products for the reaction of Cl with T2P. A reaction mechanism is suggested to explain the formation of the observed gaseous products. The results are discussed in terms 40 of the structure-reactivity relationship, and the atmospheric implications derived from this 41 42 study are commented as well.

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44 Keywords: Unsaturated aldehyde, kinetics, mechanism, organic aerosols, atmospheric
45 lifetimes

46

47 1. Introduction

48 It is well known that non-methane biogenic volatile organic compounds (BVOCs) account for approximately 90% of VOCs present in the atmosphere. BVOCs play an important role in the 49 50 chemistry of the troposphere since their oxidation provides major reaction pathways for the formation of Secondary Organic Aerosols (SOAs) (Hoffmann et al., 1997, Griffin et al., 51 52 1999). In addition, the reactivity of BVOCs has an impact on the HO_x (OH + HO₂), NO_x (NO 53 + NO₂) and the tropospheric ozone budgets. Therefore, it is important to understand the 54 oxidation of BVOCs in the atmosphere to assess their role in the troposphere's chemistry. This is crucial for atmospheric models to predict the formation of atmospheric photo-oxidants, 55 56 SOA and the HO_x cycle (Kanakidou et al., 2005, Rudich et al., 2007).

Unsaturated aldehydes are BVOCs that may be emitted into the atmosphere from the
enzymatic oxidation of linolenic and linoleic acids present in vegetation in response to stress
(Winer et al. 1992, Hatanaka, 1993, Heiden et al., 2003, Seco et al., 2007).

They may also be formed in situ as a result of the oxidation of dienes in the atmosphere (Orlando and Tyndall, 2002). Besides their emission from biogenic sources, unsaturated aldehydes are also emitted into the atmosphere from anthropogenic sources such as the combustion of wood, polymers, tobacco and gasoline (Magneron et al., 2002) and from chemical industries (Graedel and Crutzen, 1986).

Atmospheric mixing ratios of unsaturated aldehydes are in the order of ppb. For instance, the mixing ratio of acrolein, the smallest unsaturated aldehyde, in several cities was reported to be up to 12 ppb (Seaman et al., 2006, Faroon et al., 2008). Larger unsaturated aldehydes, such as trans-2-pentenal (T2P), have also been detected in the atmosphere with mixing ratios of a few ppb in soybean seeds after anaerobic incubation (Gardner et al., 1996). T2P is also emitted by vegetation such as orange (Moshonas et al., 1973), tomato (Tandon et al., 2000) and black tea (Yamanishi et al., 1972). Besides, T2P is emitted directly from food products such as olive oil

(Angerosa et al., 2000, Masella et al., 2019), milk powder (Cheng et al., 2019), wine (Saison et al., 2009, Martins et al., 2013), dry sausages, cheese and bread (Bianchi et al., 2007) or during food preparation as clam frying (Cheng et al., 2019). T2P has also been detected in human breath (Ruzsanyi et al., 2002) and cigarette smoke (Hatai et al., 2019). Further, this species is used to protect crop plants from pests and weeds (Rodríguez-Kabana et al., 2008).

Like other unsaturated aldehydes, once emitted into the atmosphere, T2P is expected to be 78 removed mainly by reaction with OH and NO₃ radicals and by photolysis during the daytime 79 (Houk, 1976, Calvert et al., 2011). However, reactions with chlorine (Cl) atoms may play an 80 important role in the removal of these species in the atmosphere and maybe competitive with 81 those due to the OH radicals. Heterogeneous CINO₂ formation in polluted environments has 82 83 been shown as a major route for the production of reactive gas-phase chlorine due to photolysis (Wang et al., 2019, Faxon et al., 2015). In coastal areas and in polluted northern 84 hemisphere areas, the concentration of Cl atoms reaches 1×10^5 atoms cm⁻³ (Pszenny et al., 85 1993, Spicer et al., 1998, Ezell et al., 2002, Tackett et al., 2007) and 5×10^4 atoms cm⁻³ 86 (Hossaini et al., 2016), respectively, making the reactions of Cl with VOCs a potentially 87 88 significant elimination route in the atmosphere for these species. Therefore, the study of these 89 processes is necessary to assess the fate of these species in the atmosphere and to better 90 evaluate the atmospheric implications of their oxidation processes.

91 To our knowledge, the rate coefficient for the T2P+Cl reaction, k_{T2P} , was measured by 92 Rodríguez et al., 2005 by a relative method at room temperature and an atmospheric pressure 93 of air and N₂ in a Teflon bag using the GC-FID (Flame ionization detector).

94 The value provided by Rodríguez et al., $(1.31\pm0.19)\times10^{-10}$ cm³ molecule⁻¹ s⁻¹, is the only 95 experimental data for k_{T2P} . However, an estimation of k_{T2P} , calculated using the Structure-

Activity Relationships (SARs) method, 3.47×10⁻¹⁰ cm³ molecule⁻¹ s⁻¹, is not in agreement 96 97 with the experimental value (Teruel et al., 2009). This value is close to the gas kinetic limit for such a bimolecular elementary reaction. To elucidate this discrepancy, the present paper 98 99 describes the determination of the rate coefficient k_{T2P} and aims at evaluating the impact of the 100 Cl + T2P reaction in the removal of T2P in the atmosphere. In addition to the kinetic investigation, the identification and quantification of the gas-phase products and the SOA 101 102 formation induced by reaction R1 are provided for the first time in the present work.

103 2. Materials and methods

104 2.1. Kinetic study

The kinetic study was carried out at 296±2 K and 730±20 Torr of air by a relative method 105 using two set-ups described below. 106

2.1.1. D-ASC chamber (IMT Lille Douai, France) 107

108 A detailed presentation of this experimental set-up can be found elsewhere (Turpin et al., 2006), and only a brief description is given here. The Douai Atmospheric Simulation 109 Chamber (hereafter D-ASC) is a 300-L Teflon bag enclosed inside a wooden box and 110 111 equipped with a 2-L multiple-reflection optical cell with 10 m of optical path coupled to an 112 FTIR spectrometer (Nicolet 550). The IR spectra of the gaseous samples (T2P and the reference compound and/or reaction products) were recorded with a resolution of 0.5 cm⁻¹ in 113 the spectral range 4000-650 cm⁻¹ by accumulating 32 interferograms. In our experiments, Cl 114 atoms were generated from the photolysis of molecular chlorine using 1 or 2 actinic lamps 115 116 (Philips TMX 200 LS, emitting in the range 300-460 nm, $\lambda_{max} = 365$ nm), which were fixed 117 on the internal walls of the box. All reactions were studied in purified zero air (relative 118 humidity < 2 ppm, CO and CO₂ < 80 ppb). T2P and the reference compounds were injected

into the Teflon bag using syringes. The sampling was made automatically each 5 min for
analysis by FTIR. At the end of each experiment, the Teflon bag and the optical cell were
cleaned several times by filling with zero air followed by pumping.

122 2.1.2. 16-L cell (University of Castilla La Mancha UCLM - Ciudad Real, Spain)

This experimental set-up was described in detail elsewhere (Ballesteros et al., 2009, Sleiman 123 124 et al., 2014), and only a brief outline is provided here. It consists of a 16-L Pyrex White-type 125 cell that has 96 m of the optical path. The cell is equipped with a Nexus 870 Thermo Nicolet FTIR spectrometer. IR spectra were recorded with a resolution of 2 cm^{-1} in the IR spectral 126 range of 4000 to 650 cm⁻¹. The cell is surrounded by 4 actinic lamps (Philips TL-K 40 W, 127 λ_{max} = 365 nm). All experiments were carried out in synthetic air (purity 99.999%, Air 128 129 Liquide). The introduction of reactants in the gas-phase was done by expansion from a glass 130 manifold system. At the end of each experiment, the cell was cleaned by filling with air and pumping down to 10^{-2} mbar several times. 131

132 2.1.3. Rate coefficient determination

The relative rate coefficient of the gas-phase reaction of T2P with Cl was measured in the presence of a reference compound (isoprene, cyclohexane or ethanol) using either the D-ASC chamber or the 16-L cell. The experimental conditions used in this study are summarized in Table 1. The reactions of interest are the simultaneous reactions of T2P and the reference compound with chlorine atoms, with rate coefficients as k_{T2P} and k_{Ref} , respectively:

$$T2P + Cl \xrightarrow{k_{T2P}} Products \tag{R1}$$

$$Reference + Cl \xrightarrow{\kappa_{Ref}} Products \tag{R2}$$

138

Before starting the kinetic study, tests were carried out to check the stability of T2P and the 139 140 reference compounds by quantifying dark reactions (loss onto the walls and potential reaction with Cl₂) and photolysis processes. The loss rate coefficients due to dark reactions and to 141 142 photolysis are defined as k_{WLX} and k_{PLX}, where X compound could be either T2P or reference compound. The values are summarized in Table 1. The rate coefficients, $k_{\rm L}$, accounting for all 143 these secondary loss processes are included in Eq. 1. They typically represent from 6 to 15% 144 145 of the logarithm components according to the reference compound employed in the 146 experiments. The time evolution of the concentration of T2P and the reference compound is, then, described by the following expression: 147

$$ln\left(\frac{[T2P]_{0}}{[T2P]_{t}}\right) - k_{L,T2P}t = \frac{k_{T2P}}{k_{Ref}} \left[ln\left(\frac{[Ref]_{0}}{[Ref]_{t}}\right) - k_{L,Ref}t\right]$$
(Eq. 1)

148 where $[T2P]_0$, $[Ref]_0$ and $[T2P]_t$, $[Ref]_t$ are the concentrations of T2P and the reference 149 compound at times t=0 and t, respectively, and $k_{L,T2P}$ and $k_{L,Ref}$ are the loss rate coefficients 150 due to secondary processes for T2P and the reference compound, respectively. The linear 151 least-square analysis at zero intercept of the $ln([T2P]_0/[T2P]_t) - k_{L,T2P}t$ against 152 $ln([Ref]_0/[Ref]_t) - k_{L,Ref}t$ plot yields k_{T2P}/k_{Ref} . The rate coefficient of T2P toward chlorine 153 atoms k_{T2P} was then calculated using the known rate coefficient k_{Ref} .

During the experiments performed in the D-ASC chamber, isoprene was used as the reference 154 compound considering a Cl-rate coefficient k_{Isoprene} of $(4.80 \pm 1.25) \times 10^{-10} \text{ cm}^3$ molecule⁻¹ s⁻¹. 155 156 This value is the average of those provided by (Ragains and Finlayson-Pitts, 1997, Fantechi et al., 1998 and Orlando et al., 2003). In the 16-L cell, two reference compounds were 157 employed: ethanol $(k_{\text{Ethanol}} = (1.0 \pm 0.2) \times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$ (IUPAC, 2006) and 158 cyclohexane $(k_{Cyclohexane} = (2.90 \pm 1.12) \times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1})$. This latter rate coefficient is 159 the average of six values given by (Atkinson and Aschmann, 1985, Wallington et al., 1988, 160 Rowley et al., 1992, Aschmann and Atkinson, 1995, Li and Pirasteh, 2006, Anderson et al., 161

162 2007). The error in k_{Ref} represents twice the standard deviation calculated from the averages
163 of individual k_{Ref}.

164 The reagents (T2P and the reference compound) consumption in the D-ASC chamber was monitored as a function of time, using the following IR spectral features: 1175-1115 cm⁻¹ and 165 942-855 cm⁻¹ for T2P and isoprene, respectively. In the 16-L cell, the IR spectral features 166 used were as follows: 1750-1700 cm⁻¹ for T2P, 2890-2830 cm⁻¹ for cyclohexane and 930-840 167 cm⁻¹ for ethanol. Once in the reactor, the gas mixture was allowed to stabilize for 40 min in 168 the D-ASC chamber and 10 min in the 16-L cell. Indeed, the stabilization time of reagents is 169 dependent on the volume of the reactor. The initial concentrations of T2P, reference 170 compounds and Cl₂ were obtained based on the injected amounts (D-ASC) or the pressure 171 172 measurements (16-L cell) and are summarized in Table 1.

173 The overall error Δk_{T2P} in the individual rate coefficient comes from statistical errors Δk_{T2P} 174 (*stat*) and systematic errors Δk_{T2P} (*syst*) as follows:

$$\Delta k_{T2P} = \sqrt{\Delta k_{T2P} (stat)^2 + \Delta k_{T2P} (syst)^2}$$
(Eq. 2)

175 Statistical errors $\Delta k_{T2P}(stat)$ are calculated according to the following equation:

$$\frac{\Delta k_{T2P} (stat)}{k_{T2P}} = \frac{\Delta (Slope)}{Slope} + \frac{\Delta k_{Ref}}{k_{Ref}}$$
(Eq. 3)

176 $\Delta(Slope)$ comes from the determination of the slope in Eq. 1 and results from the least-squares 177 analysis of Figure 1. Δk_{Ref} is the uncertainty on the rate coefficient of the used reference 178 compounds. In this work, this value represents two standard deviations obtained from the 179 averaging of several values of k_{Ref} found in the literature, as explained above. Systematic 180 errors $\Delta k_{T2P}(syst)$ are estimated to 10%. They come mainly from the uncertainty in measuring 181 the concentrations of the reagents (measurements of the areas of the spectroscopic peaks,

possible interference, and temperature instability). To minimize this error, 32 IR spectra wererecorded for each experiment.

184 2.2. Mechanistic study

In independent experiments without adding a reference compound, the products formed during the gas-phase reaction of T2P with chlorine atoms were identified and quantified using two different detection methods: FTIR spectroscopy, coupled to the 16-L cell described above, and GC-MS technique, coupled to the 264-L Pyrex chamber (Ciudad Real-Atmospheric Simulation Chamber, CR-ASC), described below.

In the 16-L Pyrex cell, the concentrations of T2P and the formed products were followed in situ by monitoring the evolution of their IR absorption bands. The IR spectral features used for this analysis were as follows: 1180-1120 cm⁻¹ for T2P, 875-817 cm⁻¹ for propanal, 1840-1780 cm⁻¹ for formic acid, 2230-2060 cm⁻¹ for CO and 3100-2600 cm⁻¹ for HCl. Examples of measured spectra are shown in Figure S1. The identification of the products was done using standard reference libraries of IR spectra.

The CR-ASC set-up was presented in detail elsewhere (Ballesteros et al., 2009), and only a 196 197 brief description is provided here. It consists of a 264-L Pyrex cylindrical reactor with four 198 quartz windows on the sides. The chamber is surrounded by a set of 8 actinic lamps (Philips 199 BL TL 40W/10 1SL/25, emitting in the range 340-400 nm, $\lambda_{max} = 365$ nm) used to generate Cl atoms from the photolysis of Cl₂. A GC-MS apparatus (Thermo Electron, Trace GC Ultra 200 and DSQ II) was employed to analyze the gas mixture from the chamber. The Solid-Phase 201 Micro-Extraction (SPME) technique was used for sampling the mixture using a 50/30 µm 202 203 divinylbenzene/carboxen/polydimethylsiloxane fiber as an absorber. The coated fiber was exposed to the gas mixture during an optimized time of 15 min. The fiber samples were then 204

thermally desorbed in the heated (at 523 K) GC injection port, and the products were analyzed 205 206 by GC-MS by using temperature ramps that ranged between 313 and 573 K. The 207 identification of the reaction products was carried out by using the electronic impact 208 ionization mass spectrum of each retention time. The consumption of T2P and the formation 209 of the products were examined by monitoring the areas of corresponding chromatographic 210 peaks observed at a certain retention time. The initial concentration of T2P and Cl₂ used for the mechanistic study are summarized in Table 2. Examples of a chromatogram and obtained 211 212 mass spectra of the detected products are shown in Figure S2.

213 2.3. Secondary Organic Aerosol (SOA) study

214 **2.3.1.** Experimental apparatus

The set-up used has already been presented elsewhere (Antiñolo et al., 2019). Hence, only a 215 brief description will be given here with a special emphasis on the specificities of the present 216 work. The study of the SOA formation in reaction R1 was carried out at 296±2 K and 730±20 217 218 Torr of synthetic air using both the CR-ASC and the 16-L cell-described above. Ranges of initial concentrations of T2P and Cl₂ were $(1.56-9.23)\times10^{14}$ molecule cm⁻³ and $(1.55-1.5)\times10^{14}$ 219 15.8)×10¹⁴ molecule cm⁻³, respectively. The loss of T2P, necessary to quantify the SOA yield, 220 was measured using the FTIR spectrometer described in section 2.1.2. Cl atoms were 221 222 generated in the CR-ASC as described in section 2.2.

A Fast Mobility Particle Sizer (FMPS) spectrometer (TSI 3091) was used to measure the number of particles formed with diameters ranging between 5.6 and 560 nm. Although it allows a time resolution of 1 s, averaged values for 1 min were considered in the present investigation. Mass concentrations were determined using an SOA density of 1.4 g.cm⁻³, which corresponds to the recommended value of Hallquist et al., 2009.

The gas mixture from the CR-ASC chamber entered in the FMPS through a 1 µm cut cyclone at a 10 L min⁻¹ flow rate, and a 10 L min⁻¹ filtered exhaust flow rate from the FMPS was then directed to the 16-L cell for FTIR measurements and then again to the CR-ASC, closing the circuit and allowing mixing of the gases in the system. Data were recorded with both FTIR and FMPS typically during 80 min as follows: 10 min before starting the reaction (dark measurements), 50 min under irradiation, followed by 20 min in dark conditions.

234 2.3.2. Determination of the SOA production

235 The SOA yield Y_{SOA} is defined by the following equation:

$$Y_{SOA} = \frac{M_{SOA}}{\Delta[T2P]} \tag{Eq. 4}$$

236 From the time evolutions of the SOA mass production M_{SOA} and the reactant consumption $\Delta[T2P]$, it is possible to plot M_{SOA} as a function of $\Delta[T2P]$ and determine Y_{SOA} from the slope. 237 Both M_{SOA} and $\Delta[T2P]$ were corrected to account for their loss in the dark when no Cl 238 239 reaction is taking place, as explained in a previous study (Antiñolo et al. 2019). The loss of M_{SOA} was characterized during the last 20 min of each experiment under the dark conditions 240 from the FMPS measurements, and its rate coefficient was in the $(5.22-6.97)\times10^{-4}$ s⁻¹ range. 241 $\Delta / T2P$ was corrected considering the FTIR measurements of T2P in the first 10 min of the 242 243 experiment under dark conditions from which it was possible to detect a loss rate coefficient within the $(1.16-20.8) \times 10^{-4}$ s⁻¹ range. Tests were also done to confirm that no SOA was 244 formed during the Cl₂ dark reaction or exposure of T2P under the light. 245

In each experiment, M_{SOA} reaches a maximum value $M_{SOA,max}$, which is directly correlated to the SOA yield Y_{SOA} . Indeed, SOA formation can be described by the one-product model proposed by (Pankow, 1994) for which the following equation was recommended by (Odum et al., 1996):

$$Y_{SOA} = M_{SOA,max} \left(\frac{\alpha K_p}{1 + M_{SOA,max} K_p} \right)$$
(Eq. 5)

where α is the mass-based gas-phase stoichiometric coefficient of a model product, and K_p represents its gas-particle partitioning equilibrium constant. The present study allows us to obtain these coefficients as explained in the discussion section 3.3.

253 **2.4.** Chemicals

The reagents employed were obtained from the following sources: T2P (≥95%, SigmaAldrich), isoprene (99%, Sigma-Aldrich), cyclohexane (99.9%, Sigma-Aldrich), ethanol
(≥99.5%, Sigma-Aldrich), Cl₂ (10% in N₂, Air Products) at IMT Lille Douai and Cl₂ (99.8%,
Sigma-Aldrich) at UCLM.

258 3. Results and discussion

259 3.1. Kinetic study

The plots of $ln([T2P]_0/[T2P]_t) \cdot k_{L,T2Pt}$ vs. $ln([Ref]_0/[Ref]_t) \cdot k_{L,Ref}t$ are presented in Figure 1. Relative rate coefficients were obtained for each reference compound. Individual and average k_{2TP} are summarized in Table 3. The good linearity with a correlation coefficient greater than 98% and an intercept close to zero for all the reference compounds indicates that no other secondary reactions occur. The error reported on the average rate coefficient shown in Table 3 is two standard deviations. The overall error Δk_{T2P} in the individual rate coefficient values ranges from 24 to 42%.

Two determinations for the rate coefficient of the studied reaction exist in the literature: an experimental value, $(1.31\pm0.19)\times10^{-10}$ cm³ molecule⁻¹ s⁻¹ (Rodríguez et al., 2005), and a rate coefficient estimated using the SAR method, 3.47×10^{-10} cm³ molecule⁻¹ s⁻¹ (Teruel et al., 270 2009). The present value of $(2.56\pm0.83)\times10^{-10}$ cm³ molecule⁻¹ s⁻¹ lays between these two 271 values.

272 k_{T2P} obtained in the present work is found to be higher than that provided by Rodríguez et al., also obtained using a relative method, by about a factor of 2. In Rodríguez et al.'s work, the 273 274 losses of T2P and the reference compounds (ethane, propene and 1-butene) were monitored 275 by GC-FID detection. The source of discrepancy with Rodríguez et al.'s work is not clear. First, the Cl-rate coefficients for the reference compounds used by Rodríguez et al. cannot be 276 the reason for this discrepancy. These rate coefficients (in cm^3 molecule⁻¹ s⁻¹) were 277 $(5.85\pm0.55)\times10^{-11}$ (Hitsuda et al., 2001), $(2.31\pm0.29)\times10^{-10}$ (Stutz et al., 1998) and 278 $(3.00\pm0.40)\times10^{-10}$ (Orlando et al., 2003) for ethane, propene and 1-butene, respectively. 279 Indeed, a re-evaluation of the value obtained by Rodríguez et al., for k_{T2P} using the rate 280 coefficients recommended by IUPAC, 2006 for ethane and propene $(5.9 \times 10^{-11} \text{ and } 2.70 \times 10^{-10})$ 281 cm³ molecule⁻¹ s⁻¹, respectively), gives $k_{T2P} = 1.39 \times 10^{-10}$ cm³ molecule⁻¹s⁻¹, which remains 282 283 within the error bars given by these authors. The only difference that we can find between the two works, apart of the used technique, is that in Rodríguez et al.'s work the losses were not 284 corrected although that wouldn't explain the discrepancy. It is important to note that in the 285 present work, the rate coefficient of the reaction of T2P with Cl was measured in two 286 287 different set-ups under different conditions. Very close rate coefficients values were obtained making our reported value reliable. The rate coefficient of the reaction of T2P with Cl atoms 288 estimated by Teruel et al., 2009 using the SAR method is 25% higher than that determined in 289 the present work. The source of this discrepancy is probably the uncertainties of the SAR 290 291 method that aims to estimate rate coefficients based on the reactivity of other compounds considering only their structure. 292

As generally observed for VOCs, Cl atoms are more reactive than OH radicals toward these species because they are stronger electrophiles than OH radicals. As expected, then, the rate coefficient for reaction R1 is almost 1 order of magnitude greater than the rate coefficient for the OH + T2P reaction, 3.33×10^{-11} cm³ molecule⁻¹ s⁻¹ (average from (Davis et al., 2007) and (Albaladejo et al., 2002)).

In Table 4, the rate coefficients of the reactions of a series of C_3 - C_7 unsaturated aldehydes with Cl atoms are displayed. Our result confirms that the structure of unsaturated aldehydes has a very small impact on their reactivity towards Cl atoms, as was previously observed (Ullerstam et al., 2001, Colmenar et al., 2015)) opposite to their reactivity towards OH and NO₃ radicals (Tables S1 and S2). The reactivity towards Cl seems to be fairly independent of the substitution on the olefinic carbons for unsaturated aldehydes, as stated by Blanco et al., 2010.

To evaluate the effect of the -CHO group on the reactivity, a comparison between the 305 306 reactions of Cl with a series of unsaturated aldehydes to those with the corresponding alkenes 307 would be useful. Unfortunately, no data are available in the literature for the reaction of Cl with *trans*-2-pentene, the corresponding alkene of T2P. However, data for other unsaturated 308 309 aldehydes and their corresponding alkenes are available. The rate coefficients of the reactions of Cl with crotonaldehyde and 3-methyl-2-butenal (see Table 4) are 34% and 36% lower than 310 those with *trans*-2-butene (3.58×10^{-10} cm³ molecule⁻¹ s⁻¹ provided by Kaiser et al., 2007) and 311 2-methyl-propene $(3.38 \times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1} \text{ provided by Blanco et al., 2010}),$ 312 313 respectively. This may be due to the negative inductive effect of the carbonyl-containing 314 group, which may deactivate the double bond toward the Cl atom addition. However, this decrease in reactivity due to the electron-withdrawing effect may be compensated by the fact 315 that the -CHO group is a reactive site. This was discussed in the study carried out by 316

317 (Thévenet et al., 2000), where the rate coefficient of the reaction of Cl with acrolein 318 $(2.2 \times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1})$ was compared to that for propene $(2 \times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1})$ 319 and was shown to be slightly higher within uncertainties showing that the –CHO group is a 320 reactive site, although H abstraction from the –CHO group remains a minor channel 321 compared to the Cl-addition to the double bond, which is likely to be the major one as stated 322 by these authors.

323 **3.2.** Mechanistic study

324 **3.2.1.** Molar yields of reaction products

325 This work represents the first mechanistic study on the T2P+Cl reaction. It has been carried 326 out at 296±2 K and 730±20 Torr of total pressure. By FTIR, HCl, CO, propanal and formic 327 acid were identified and quantified as products formed from the studied reaction (Figure S1). 328 Figure 2 shows a typical concentration vs. time profile for the consumption of T2P and the formation of products. The plots of the *Product*_t vs. $\Delta [T2P]$ for propanal and formic acid are 329 330 shown in Figure 3. [*Product*]_t is the product concentration at a time t, Δ [T2P] is the consumed 331 concentration of T2P at a time t. For propanal, a correction of its concentration was done according to the method described by Ceacero-Vega et al. (2012) and considering the IUPAC 332 recommendation for the rate coefficient of the Cl-reaction of propanal: 1.3×10^{-10} cm³ 333 molecule⁻¹ s⁻¹ (Atkinson et al. 2006). 334

The molar yields obtained from the slopes of the plot depicted in Figure 3 are: (5.2 ± 0.2) % and (6.2±0.4)%, for propanal and formic acid, respectively. The overall errors in the formation yields are due to systematic errors σ_{syst} mentioned above and statistical errors resulting from a least-squares analysis of the [*Product*]_t vs. Δ [*T2P*] plots. CO is expected to be formed as a primary product and also as a final product from the studied reaction and may be generated in the oxidation of other species in the mechanism. For that reason, a molar yield for this

341 compound is not reported in this work. HCl yield is not reported either because it may be a 342 primary product, or it can be generated in the H-abstraction of other reaction products. 343 Propanal is expected to be formed as a primary product from the reaction of T2P with Cl. Formic acid may be produced through the very rapid oxidation of other species generated in 344 345 the oxidation of T2P or as a consequence of heterogeneous reactions on the chamber walls though no evidence of this phenomenon was observed in our experiments. The residual IR 346 spectrum presented in Figure S1 shows peaks that could correspond to ketene in the 2200-347 2100 cm⁻¹ region, according to Wallington et al. (1996), and 2-chlorobutanal considering the 348 peak typical for a carbonyl group (1750 cm⁻¹), the bands due to C-H bending and stretching 349 $(1500-1000 \text{ cm}^{-1} \text{ and } 3000-2800 \text{ cm}^{-1}, \text{ respectively})$ and the band that can correspond to the 350 C-Cl stretching (1000-650 cm⁻¹). However, none of these two compounds could be confirmed 351 by FTIR or quantified given the lack of standards of these products. 352

Three of the peaks observed in the chromatogram obtained by GC-MS could be assigned to 353 354 reaction products (Figure S2). The peak detected at 4.03 min, for which the mass spectrum 355 (MS) shows peaks at m/z = 106 (molecular ion, M), 78 and 80 (fragment M-CO or M-C₂H₄) 356 for the two Cl isotopes), 77 (M-C(O)H), and 41 (C₃H₅), could be assigned to 2-chlorobutanal, 357 and that would confirm the observations made by FTIR. A chlorinated product that could not be identified was observed at 5.09 min retention time: its MS shows the isotopic pattern for Cl 358 359 in the 77 and 105 peaks, although it was not possible to clearly identify the molecular ion. 360 The last peak in the chromatogram, at 5.82 min, was attributed to 2-pentenoic acid, given that in the MS there are peaks at m/z = 100 (M), 82 (M-H₂O) and 55 (M-C(O)OH). This 361 362 identification was confirmed by comparing the reference MS library installed in the CG-MS 363 software used in this study. The quantification of these products was not possible due to the 364 lack of standards. However, based on the area of the observed peaks in the chromatogram and 365 assuming the same response for all of them, it seems that the yield of 2-chlorobutanal and the

unidentified chlorinated product is higher than that for 2-pentenoic acid (Figure S2). The carbon balance was calculated considering only propanal and CO. The obtained value was shown to not exceed 15% at the end of the reaction showing that additional formed oxidation products could not be detected or quantified.

370 3.2.2. Proposed reaction mechanism

371 As other unsaturated aldehydes (Canosa-Mas et al., 2001, Orlando and Tyndall, 2002, 372 Magneron et al., 2002), the degradation of T2P initiated by Cl atoms is expected to proceed 373 via two basic mechanisms: abstraction of the hydrogen atom from the -CHO group and/or addition to the C=C double bond. In Scheme 1, the reaction pathways that may justify the 374 375 observed products are shown. The addition of a Cl atom may occur on two sites of the C=C double bond: adjacent to the -C(O)H group (α) or on the double-bonded carbon furthest from 376 377 the aldehydic group (β). As seen in Scheme 1, the α -addition forms the CH₃CH₂CHCHClC(O)H radical, which will react with O₂ followed by a reaction with RO₂ to 378 the corresponding chloroalkoxy radical CH₃CH₂C(O)HCHClC(O)H. 379 lead to The 380 decomposition of this chloroalkoxy radical may proceed via two channels. The first one leads 381 to the formation of propanal (CH₃CH₂C(O)H), 2-oxoacetylchloride (ClC(O)C(O)H), formyl 382 chloride HC(O)Cl and CO. Propanal may further react with Cl atoms to form HCl and 383 CH₃CH₂C(O) radicals (Le Crâne et al., 2004). The second channel leads to the formation of 384 chloromalonaldehyde (C(O)HCHClC(O)H) and acetaldehyde CH₃C(O)H. Among these 385 products, only propanal and CO were observed and quantified in this work.

386 The β -addition mechanism proceeds through the CH₃CH₂CHClCHC(O)H radical, which will 387 react with O₂ followed by a reaction with RO₂ to lead to the CH₃CH₂CHClC(O)HC(O)H 388 radical. This radical may decompose to lead to the formation of 2-chlorobutanal (CH₃CH₂CHClC(O)H) identified by GC-MS, but not quantified and further to the formation 389 identified FTIR. 390 of CO and quantified by The decomposition of the

391 CH₃CH₂CHClCOHC(O)H radical may also lead to the formation of glyoxal (C(O)HC(O)H) 392 and 1-chloropropanal (CH₃CH₂C(O)Cl) after the conversion of the reaction of the 393 chloroperoxy radical CH₃CH₂C(O₂)HCl with RO₂. Glyoxal (C(O)HC(O)H) and 1-394 chloropropanal were not observed in this work, suggesting that this channel either does not 395 occur in the timescale of the experiments or is a minor route in the mechanism. Note also that 396 glyoxal may be readily photolyzed under UV irradiation. The CH₃CH₂CHClC(O)HC(O)H 397 radical may react with O_2 to form chloropropylglyoxal (CH₃CH₂CHClC(O)C(O)H). This 398 product may be the unidentified one observed by GC-MS but could not be confirmed.

399 The H-abstraction from the -C(O)H group may form the $CH_3CH_2CHCHC(O)$ radical and 400 HCl. In the presence of O_2 , the CH₃CH₂CHCHC(O) radical may be converted to the peroxy radical CH₃CH₂CHCHC(O)O₂, which generates 2-pentenoic acid (CH₃CH₂CHCHC(O)OH) 401 402 and ozone in the presence of HO₂. 2-pentenoic acid was detected in SPMS/GC-MS analysis. 403 On the other hand, the CH₃CH₂CHCHC(O)O₂ radical may also react with RO₂ and generates 404 the CH₃CH₂CHCHC(O)O radical. This radical can decompose to carbon dioxide (CO₂) and 405 ethylvinyl radical (CH₃CH₂CHCH), which is converted to the CH₃CH₂CHCH(O₂) radical in 406 the presence of O_2 . The reaction of this later with RO_2 may lead to the formation of butanal 407 (CH₃CH₂CH₂C(O)H) not observed in this work. The reaction of CH₃CH₂CHCH(O₂) with 408 RO₂ may also lead to propanal and CO formation, both quantified by FTIR in this work.

In the light of the discussion above, the relative contribution of addition and abstraction channels in the suggested mechanism could not be determined. In fact, the product formed only through the β -addition channel and observed in this work (2-chlorobutanal) could not be quantified. Among the products expected to occur from the α -addition channel, only propanal and CO were detected in this work. These products may also be formed through the other channels, as mentioned before.

415 **3.3. SOA study**

Figure S3 shows the relationship between the initial concentration of T2P and Cl₂ with $M_{SOA,max}$. A clear dependency of $M_{SOA,max}$ on the amount of Cl₂ introduced in the system was observed, whereas no obvious correlation can be identified with the initial concentration of T2P.

420 Typical aerosol size distributions at several reaction times for given initial concentrations of T2P and Cl₂ are presented in Figure S4. In Figure 4, the SOA mass concentration M_{SOA} is 421 plotted as a function of the consumed T2P. The SOA yield Y_{SOA} is obtained from linear least-422 423 squares analysis. The values of Y_{SOA} are summarized together with the experimental 424 conditions in Table 5. Despite the data scattering (Figure 5), the calculated Y_{SOA} fits 425 reasonably well to a one-product model (Odum et al., 1996), given by Equation 5. A leastsquare regression on the data gives the following parameters: $\alpha = (9.95\pm2.52)\%$ and $K_p =$ 426 $(6.00\pm2.68) \times 10^{-4} \text{ m}^3 \text{ }\mu\text{g}^{-1}$. Uncertainties are statistics and represent two standard deviations 427 428 (2**σ**).

Many products result from the oxidation of a VOC, so this fitting allows for easy comparison
of yield curves, but, as reported by Cai and Griffin, 2006, the fitted parameters provide no
specific information about the real oxidation products and ascribing meaning to them should
be avoided.

The SOA formation yields for the reaction of other unsaturated compounds with Cl have been previously determined. For *trans*-2-methyl-2-butenal, Y_{SOA} were recently determined to be between 0.26 and 1.65% (Antiñolo et al., 2020). These SOA yields are a little lower than the values determined in the present work for the reaction of Cl with T2P.

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437 4. Atmospheric implications and conclusion

The atmospheric lifetime τ_{Cl} of T2P due to its reaction with Cl is estimated from k_{T2P} obtained 438 in this work according to the following equation: $\tau_{Cl} = 1/(k_{T2P} \times [Cl])$. [Cl] represents the 439 global tropospheric concentration of Cl of 1×10^3 atoms cm⁻³ averaged over 24 hours 440 (Wingenter et al., 1999). With this value, a τ_{Cl} of 44 days is calculated. The reaction of T2P 441 with Cl may become significant in coastal areas and in polluted northern hemisphere areas 442 where the concentration of Cl atoms can reach 1×10^5 molecule cm⁻³ (Singh et al., 1996) and 443 5×10^4 atoms cm⁻³ (Hossaini et al., 2016), respectively, yielding to a lifetime shorter than 1 444 445 day which makes the reaction of T2P with Cl a potentially significant elimination route in the atmosphere. The calculated lifetime of T2P due to its reaction with OH is 7 hours using a 24-h 446 average concentration of 1×10^6 molecules cm⁻³ for OH (Atkinson et al., 1997) and the average 447 of the rate coefficient values reported in the literature by (Davis et al., 2007) and (Albaladejo 448 et al., 2002) $(3.33 \times 10^{-11} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1})$. A short lifetime (5 hours) of T2P is calculated for 449 450 its reaction with NO₃ radicals showing that this process is an important removal pathway for T2P during nighttime (Kalalian et al., 2019). These authors provided a lifetime of 9 days 451 towards reaction with ozone. The tropospheric lifetime considering the four chemical removal 452 453 pathways, is shorter than 3 hours. Further, the UV cross-sections of T2P have been recently 454 determined by Kalalian et al., 2019. A lower limit of 44 min was provided by these authors 455 for the T2P lifetime due to photolysis in the UV solar actinic region. This value was 456 calculated using a quantum yield of unity and a photolysis rate corresponding to June 30 at 457 midday for Paris (zenith angle = 26°). It should be noted however that the overall quantum 458 yield in the actinic region for T2P are not known and that for smaller unsaturated aldehydes such as *trans*-crotonaldehyde it has been found to be quite small since Magneron et al., 2002 459 460 obtained an upper value of 0.03 for this species.

The degradation products (and molar yields) of T2P in the presence of air were HCl, CO, and propanal ($5.2\pm0.2\%$). Besides, chlorinated products, such as 2-chlorobutanal, were identified as reaction products of the T2P + Cl reaction but could not be quantified. Highly oxidized compounds, such as 2-pentenoic acid and formic acid ($6.2\pm0.4\%$), were also detected in this work. Formic acid can be produced as a consequence of heterogeneous reactions on the chamber walls.

Given the formation of HCl, the T2P + Cl reaction may contribute to the acidity of marine 467 468 atmospheres. On the other hand, in polluted areas (high-NO_x conditions), the formation of CO may contribute to photochemical pollution via its OH-reaction, which produces CO₂ and H 469 470 atoms that quickly combine with O_2 to form HO_2 radicals regenerating OH and producing O_3 471 in the troposphere. It is important to note that in polluted areas, the reaction of T2P with Cl 472 may be a source of other toxic molecules such as PAN type species. In addition, the chemical mechanism displayed in Scheme 1 for NO_x -free conditions will evolve by including RO_2 + 473 NO and $RO_2 + NO_2$ reactions. 474

As observed in this work, the high particle mass concentration was due to the large amount of 475 476 T2P present in the system (22.5 ppm). In a real atmosphere, the mixing ratio of T2P is at ppb levels. Considering the small SOA yield formed in the T2P + Cl reaction ($Y_{SOA} < 7\%$), it is 477 478 not expected that T2P could contribute to the formation of ultrafine particles in clean 479 atmospheres. The dependence of Y_{SOA} with the maximum SOA mass is well-described by a one-product absorptive model that yielded the Odum's parameters α and K_p , the latter being 480 two orders of magnitude lower than the one reported for α -pinene, β -pinene and d-limonene 481 482 (Cai and Griffin, 2006). Even though the uncertainties in Y_{SOA} are high due to the lack of 483 knowledge in the SOA density, among other assumptions that were made, the determined

484 SOA yields are in the order of magnitude of those reported for the Cl + toluene reaction (Cai
485 et al., 2008) and other similar biogenic compounds as isoprene (Wang and Ruiz, 2017).

486 Additionally, insights in the reaction mechanism for the degradation of T2P initiated by Cl atoms were reported here. From the obtained results, the first step of the reaction mechanism 487 488 seems to proceed mainly by H-abstraction of the aldehydic group and by β -addition of Cl atoms to the double bond. The carbon balance calculated in this study did not exceed 15% at 489 the end of the reaction showing that additional oxidation products could not be detected or 490 quantified. Further studies, with complementary detection techniques, are clearly needed to 491 492 better understand the mechanism and accurately evaluate the carbon balance and the relative contribution of addition and abstraction channels in this mechanism. 493

494 Acknowledgement

495 The authors are grateful to the INSU-LEFE French program, the IEA program funded by 496 CNRS France and the Brittany Council. IMT Lille Douai acknowledges funding by the 497 French ANR agency under contract No. ANR-11-LabX-0005-01 CaPPA (Chemical and 498 Physical Properties of the Atmosphere), the Région Hauts-de-France, the Ministère de 499 l'Enseignement Supérieur et de la Recherche (CPER Climibio), and the European Fund for 500 Regional Economic Development. This research was also funded by the Consejería de Educación, Cultura y Deportes de la Junta de Comunidades de Castilla-La Mancha through 501 502 the project SBPLY/19/180501/000052 and the University of Castilla-La Mancha (UCLM) 503 through the project 2019-GRIN-27175 (Ayudas para la financiación de actividades de 504 investigación dirigidas a grupos). M. Antiñolo acknowledges UCLM Plan Propio de Investigación for funding her research contract. 505

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Tables

Reactor	Lamps	Compound	Initial concentration	kwl,x± σ_1^{a}	$\mathbf{k}_{\mathrm{PL},\mathrm{X}} \pm \boldsymbol{\sigma}_{2}^{\mathrm{a}}$	
			(× 10 ¹³ molecule cm ⁻³)	(×10 ⁻⁵ s ⁻¹)	(×10 ⁻⁵ s ⁻¹)	
D-ASC	1-2	T2P	5.04 - 15.10	1.73 ± 0.06 –	1.85 ± 0.10 -	
				6.55 ± 0.39	6.43 ± 0.41	
		Isoprene	5.90 - 14.80	$0.50 \pm 0.12 -$	$0.54 \pm 0.16 -$	
				2.42 ± 0.07	2.68 ± 0.12	
		Cl ₂	25.00 – 37.26			
16-L cell	4	T2P	46.20 - 93.00	$0.48 \pm 0.15 -$	2.95 ± 0.13 -	
			0	1.02 ± 0.08	3.57 ± 0.07	
		Cyclohexane	57.60 - 66.00	0.11 ± 0.09 –		
				1.07 ± 0.21		
		Ethanol	74.40 - 110	$1.72 \pm 0.07 -$		
				2.08 ± 0.40		
		Cl2	66.00 – 132			

Table 1. Experimental conditions used during our kinetic study.

^a: uncertainty is $\pm 1\sigma$.

Reactor	Lamps	Compound	Initial concentration range
			(× 10 ¹⁴ molecule cm ⁻³)
16-L cell (4 runs)	3	T2P	2.78-5.91
		Cl_2	8.34-9.97
CR-ASC (2 runs)	8	T2P	9.05-9.64
		Cl ₂	8.34-8.76

Table 2. Experimental conditions used during the mechanistic study.

Table 3. Rate coefficients for the reaction of T2P with Cl atoms at 296 ± 2 K and 730 ± 20

Torr.

Reactor	Reference compound	k _{T2P} /k _{Ref}	$k_{T2P} \pm \Delta k_{T2P}^{a}$ (× 10 ⁻¹⁰ cm ³ molecule ⁻¹ s ⁻¹)
D-ASC	Isoprene (6 runs)	0.50 ± 0.04	2.40 ± 0.82
16-L Pyrex cell	Cyclohexane (2 runs) Ethanol (3 runs)	0.79 ± 0.01 2.99 ± 0.08	2.29 ± 0.92 2.99 ± 0.69
	Average	2.56 ± 0.83^{b}	

 $k_{\text{Isoprene}} = (4.8 \pm 1.25) \times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1} \text{ (see section 2.1.3)}$ $k_{\text{Ethanol}} = (1.0 \pm 0.2) \times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1} \text{ (see section 2.1.3)}$ $k_{\text{Cyclohexane}} = (2.90 \pm 1.12) \times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1} \text{ (see section 2.1.3)}$ $a^{\text{uncertainty}} \text{ calculated according to equations (4-5)}$ $b^{\text{two}} \text{ standard deviations}$

Unsaturated aldehyde	k _{Cl}	Reference
	$(\times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1})$	
Acrolein	2.2 ± 0.3	(Thévenet et al. 2000)
(CH ₂ =CH-CHO)	2.2 ± 0.3	(Canosa-Mas et al. 2001)
	1.8 ± 0.3	(Ullerstam et al. 2001)
	2.5 ± 0.7	(Wang et al. 2002)
Crotonaldehyde	2.6 ± 0.3	(Thévenet et al. 2000)
(CH ₃ -CH=CH-CHO)	2.2 ± 0.4	(Ullerstam et al. 2001)
	3.2 ± 0.9	(Wang et al. 2002)
3-methyl-2-butenal	2.48 ± 0.71	(Blanco et al. 2010)
(CH ₃) ₂ C=CH-C(O)H		
Trans-2-pentenal	1.31 ± 0.19	(Rodríguez et al. 2005)
(CH ₃ -CH ₂ -CH=CH-CHO)	3.47	(Teruel et al. 2009) (SAR Estimation)
	2.53 ± 0.83	
		This work
Trans-2-hexenal	1.92 ± 0.22	(Rodríguez et al. 2005)
(CH ₃ -CH ₂ -CH ₂ -CH=CH-CHO)		
Trans-2-heptenal	2.4 ± 0.29	(Rodríguez et al. 2005)
(CH ₃ -CH ₂ -CH ₂ -CH ₂ -CH=CH-CHO)		

Table 4. Summary of the rate coefficients for the reactions of a series of C_3 - C_7 unsaturated aldehydes with Cl atoms.

[T2P] ₀ (× 10 ⁴ μg m ⁻³)	[Cl ₂] ₀ (× 10 ⁴ µg m ⁻³)	Δ[T2P] (× 10 ⁴ μg m ⁻³)	$\frac{M_{SOA,max}}{(\times 10^2 \mu g m^{-3})}$	Y _{SOA} ^a (%)
8.03	1.86	1.80	1.20	0.82 ± 0.02
9.39	2.43	1.75	1.75	1.15 ± 0.02
13.14	4.28	3.08	3.42	1.34 ± 0.02
3.30	3.16	1.56	3.39	2.07 ± 0.09
12.62	8.45	4.75	9.11	2.15 ± 0.02
2.23	2.43	0.88	2.56	2.47 ± 0.08
5.75	8.07	3.36	8.18	2.59 ± 0.05
3.05	8.07	1.67	8.20	2.79 ± 0.08
2.99	8.33	1.62	9.72	3.79 ± 0.10
5.75	7.89	2.19	8.97	3.79 ± 0.10
9.32	13.74	2.79	13.48	4.26 ± 0.11
6.69	12.41	4.31	15.89	4.62 ± 0.27
2.23	7.00	1.35	9.88	4.77 ± 0.28
7.89	18.97	4.89	21.56	5.17 ± 0.39
6.59	16.10	4.43	16.62	5.24 ± 0.54
9.32	17.34	2.46	19.00	6.16 ± 0.10

Table 5. Experimental conditions and results obtained for the SOA study.

^a Y_{SOA} is determined from the slope of $M_{SOA,max}$ vs. Y_{SOA} . Errors are 2σ .

Figure and Scheme Legends

Figure 1. Plot of $ln([T2P]_0/[T2P]_t)-k_{L,T2P}t$ vs. $ln([Ref]_0/[Ref]_t)-k_{L,Ref}t$ according to Eq. 1. Data obtained using the D-ASC (isoprene) and the 16L-cell (Cyclohexane and Ethanol).

Figure 2. Concentration-time profiles for T2P and the major products observed by FTIR using the 16-L cell.. Initial concentrations: 5.91×10^{14} molecule.cm⁻³ for T2P and 9.09×10^{14} molecule.cm⁻³ for Cl₂.

Figure 3. Product yields for HC(O)OH and propanal.

Figure 4. SOA mass concentration M_{SOA} produced from the T2P+Cl reaction as a function of the consumed T2P with an initial concentration of 4.03×10¹⁴ molecule.cm⁻³ for T2P and 6.72×10¹⁴ molecule.cm⁻³ for Cl₂. The SOA yield Y_{SOA} is obtained from a linear least-squares analysis.

Figure 5. Plot of SOA yield Y_{SOA} as a function of $M_{SOA,max}$ for the reaction between T2P and Cl.

Scheme 1. Proposed mechanism for the reaction of T2P with Cl atom: α -addition (a), β -addition (b) and H-abstraction (c).

Figures



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- A rate coefficient of (2.56±0.83)×10⁻¹⁰ cm³ molecule⁻¹ s⁻¹ was obtained for the reaction of T2P with Cl atoms under atmospheric conditions.
- This work is the first study of the gas-phase products and the SOA formation from the reaction of T2P with Cl.
- The T2P + Cl reaction may contribute to the acidity of marine atmospheres.
- The T2P + Cl reaction proceeds mainly by H-abstraction of the aldehydic group and by β-addition of Cl atoms to the double bond.
- Atmospheric lifetimes of T2P toward chemical removals are of few hours.

Journal Prevention

Declaration of interests

 \boxtimes The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: