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Early Diagenesis in the Hypoxic and Acidified Zone of the Northern Gulf of Mexico: Is Organic Matter Recycling in Sediments Disconnected From the Water Column?

Christophe Rabouille^{1*}, Bruno Lansard¹, Shannon M. Owings², Nancy N. Rabalais³, Bruno Bombled¹, Edouard Metzger⁴, Julien Richirt⁴, Eryn M. Eitel², Anthony D. Boever², Jordon S. Beckler⁵ and Martial Taillefert²

¹ Laboratoire des Sciences du Climat et de l'Environnement, UMR 8212 CEA-CNRS-UVSQ, CEA-Orme des Merisiers, IPSL and Université Paris-Saclay, Gif-sur-Yvette, France, ² School of Earth and Atmospheric Sciences, Georgia Institute of Technology, Atlanta, GA, United States, ³ Department of Oceanography and Coastal Sciences, Louisiana State University, Baton Rouge, LA, United States, ⁴ LPG-BIAF, UMR 6112 CNRS, Université d'Angers, Angers, France, ⁵ Harbor Branch Oceanographic Institute, Florida Atlantic University, Boca Raton, FL, United States

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*Correspondence:

Christophe Rabouille
rabouill@lsce.ipsl.fr
orcid.org/0000-0003-1211-717X

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Hypoxia and associated acidification are growing concerns for ecosystems and biogeochemical cycles in the coastal zone. The northern Gulf of Mexico (nGoM) has experienced large seasonal hypoxia for decades linked to the eutrophication of the continental shelf fueled by the Mississippi River nutrient discharge. Sediments play a key role in maintaining hypoxic and acidified bottom waters, but this role is still not completely understood. In the summer 2017, when the surface area of the hypoxic zone in the nGoM was the largest ever recorded, we investigated four stations on the continental shelf differentially influenced by river inputs of the Mississippi-Atchafalaya River System and seasonal hypoxia. We investigated diagenetic processes under normoxic, hypoxic, and nearly anoxic bottom waters by coupling amperometric, potentiometric, and voltammetric microprofiling with high-resolution diffusive equilibrium in thin-films (DET) profiles and porewater analyses. In addition, we used a time-series of bottom-water dissolved oxygen from May to November 2017, which indicated intense O₂ consumption in bottom waters related to organic carbon recycling. At the sediment-water interface (SWI), we found that oxygen consumption linked to organic matter recycling was large with diffusive oxygen uptake (DOU) of 8 and 14 mmol m⁻² d⁻¹, except when the oxygen concentration was near anoxia (5 mmol m⁻² d⁻¹). Except at the station located near the Mississippi river outlet, the downcore pore water sulfate concentration decrease was limited, with little increase in alkalinity, dissolved inorganic carbon (DIC), ammonium, and phosphate suggesting that low oxygen conditions did not promote anoxic diagenesis as anticipated. We attributed the low anoxic diagenesis intensity to a limitation in organic substrate supply, possibly linked to the reduction of bioturbation during the hypoxic spring and summer.

Keywords: diagenesis, hypoxia, acidification, organic matter, continental shelf, oxygen, sediments, RiOMar

INTRODUCTION

Ocean de-oxygenation is a major concern due to warming and eutrophication, especially in the coastal zone (Pörtner et al., 2007; Diaz and Rosenberg, 2008; Gilbert et al., 2010; Zhang et al., 2010; Breitburg et al., 2018; Capet et al., 2020). The occurrence, frequency and extent of hypoxic events (dissolved oxygen concentration $< 63 \mu\text{mol L}^{-1}$) are increasing in many coastal zones due to global warming and local eutrophication. The oxygenation of sea water is fundamental to biological and biogeochemical processes. Its decline can cause major changes in ocean productivity, biodiversity, and biogeochemical cycles. Despite the recognized impact of hypoxia on many coastal ecosystems, the numerous benthic processes which promote or limit its negative effects are largely under-investigated. Oxygen availability affects remineralization processes and associated sources and sinks of important nutrient elements, such as nitrogen, phosphorus, and iron (Breitburg et al., 2018). In the general context of ocean acidification caused by anthropogenic CO_2 invasion (Orr et al., 2005), coastal acidification shows some specific features. In addition to surface invasion of CO_2 , coastal acidification is amplified by the microbial respiration of terrigenous and marine organic matter in bottom waters, and the accumulation of DIC in stratified waters (Cai et al., 2011). In coastal regions, seasonal hypoxia is generally associated with significant bottom water acidification because of concomitant DIC production and oxygen consumption during aerobic mineralization of organic matter thus imposing a double stressor to coastal ecosystems (Cai et al., 2011). Sediments play a key role in maintaining hypoxic and acidified bottom waters, but this role is still not completely understood (Middelburg and Levin, 2009).

The northern Gulf of Mexico (nGoM) and the Louisiana shelf near the Mississippi River delta includes one of the largest seasonal hypoxic zones of the coastal ocean (Rabalais et al., 2002). Coastal eutrophication associated with nutrient input from the Mississippi River watershed is responsible for the onset of hypoxia through increased primary production, organic matter sedimentation, and mineralization in the bottom waters of the stratified water column (Rabalais et al., 2002; Turner et al., 2006, 2008; Bianchi et al., 2010; Fennel and Testa, 2019). The aerobic microbial degradation of organic matter also generates *in situ* ocean acidification by the formation of metabolic CO_2 without significant total alkalinity (TA) production (Cai et al., 2011). The consumption of bottom water oxygen in the nGoM is intense and correlated to the vertical flux of organic matter originating from primary production in the Mississippi River plume and adjacent continental shelf waters (Redalje et al., 1994; Rabalais et al., 2014; Wang et al., 2018). This organic matter flux promotes a rise in bacterial oxygen consumption with depletion time of several days to a few weeks (Chin-Leo and Benner, 1992; Amon and Benner, 1998; Rabalais et al., 2007; Rabouille et al., 2008). Mobile muds contribute to sediment redistribution and export across the continental shelf (Corbett et al., 2004, 2007; McKee et al., 2004) and to oxygen consumption in the bottom water during resuspension events (Moriarty et al., 2018). In addition to water column respiration, sediment oxygen demand was identified as an important contributor to hypoxia (Turner et al., 2008) with

benthic oxygen consumption rates obtained by *in situ* benthic chambers ranging from 5 to 50 $\text{mmol O}_2 \text{ m}^{-2} \text{ d}^{-1}$ (Morse and Rowe, 1999; Rowe et al., 2002; Berelson et al., 2019) and by *ex situ* sediment core incubations ranging from 10 to 20 $\text{mmol O}_2 \text{ m}^{-2} \text{ d}^{-1}$ (Murrell and Lehrter, 2011). It was also proposed that dissolved inorganic carbon (DIC) benthic fluxes contribute to the acidification of bottom waters (Hu et al., 2017). As anoxic mineralization in sediments produces alkalinity (Krumins et al., 2013; Rassmann et al., 2020), benthic alkalinity fluxes may also mitigate bottom water acidification if the TA/DIC flux ratio is above one (Hu and Cai, 2011). Several 3D models dealing with hypoxia in the nGoM utilized sediment oxygen consumption (SOC) representations and DIC fluxes to explain the onset and maintenance of hypoxia and the related acidification throughout the summer season (Hetland and DiMarco, 2008; Fennel et al., 2013; Laurent et al., 2017). Specific sediment models were also developed with the aim of understanding oxygen fluxes and their relationship with organic matter deposition and oxygen concentration in the water column (Morse and Eldridge, 2007; Laurent et al., 2016). Benthic-pelagic exchange processes were recognized as key features of mathematical models for accurate representation of hypoxia occurrence and persistence in the nGoM (Fennel et al., 2013; Feist et al., 2016).

Yet, benthic processes involved in bottom water oxygen consumption and acidification remain poorly characterized. After the pioneering work of Lin and Morse (1991) on pyrite formation in the Gulf of Mexico, which included pore water depth profiles of sulfur species (SO_4^{2-} , $\Sigma\text{H}_2\text{S}$), few pore water data have been published in the following 20 years. Morse and Rowe (1999) investigated the benthic biogeochemistry at three stations beneath the Mississippi river plume and some other data have been published in a later paper (Morse and Eldridge, 2007). Pore water data obtained at five stations across the Louisiana shelf for 5 years (2002–2007) were published in a report (Murrell et al., 2013) without a thorough discussion of the underlying biogeochemical processes. Later, these processes were partially discussed for two of the five stations (Devereux et al., 2015; Laurent et al., 2016), although a larger collection of pore water profiles was published more recently (Devereux et al., 2019) together with sulfate reduction rate measurements. In contrast to the large sediment oxygen demands reported in the above studies, all data sets indicated poorly active subsurface sediments with limited increase of metabolite concentrations (NH_4^+ , DIC) in shelf sediments under the river plume and adjacent continental shelf waters.

In most of these studies, the contribution of benthic processes, in particular the role of dissolved iron, manganese and sulfide oxidation at the oxic-anoxic interface in sediments, to the onset and persistence of hypoxic bottom waters throughout the summer was not investigated thoroughly. Furthermore, the role of benthic processes on DIC and TA fluxes to bottom waters and its link to acidification during both normoxic and hypoxic periods was not considered, despite the potential release of alkalinity produced by increased anoxic diagenesis (Krumins et al., 2013; Rassmann et al., 2020).

In this paper, the role of early diagenesis in Louisiana shelf sediments on bottom water hypoxia was investigated in the

summer of 2017, when the surface area of the hypoxic zone was the largest ever recorded (Rabalais and Turner, 2019). Surface sediments and pore waters within and beyond the hypoxic zone were examined using multiple techniques, including *in situ* and *ex situ* micro-electrodes, diffusive equilibrium in thin films (DET) probes and classical pore water extractions. Time series data (7 months) of bottom water dissolved oxygen from a fixed station in 20-m water depth collected simultaneously were also included to examine the oxygen conditions before, during, and after hypoxia and assess the dynamics of oxygen consumption in the water column. The main objectives of this paper are to: (i) examine the contribution of water column, interface and subsurface sediments to the overall oxygen demand of bottom waters; (ii) investigate whether anoxic early diagenesis in subsurface sediments (metal oxides and sulfate reduction) is enhanced by hypoxic conditions; and (iii) examine the contribution of early diagenesis to bottom water acidification through benthic DIC and TA production and their potential release to the overlying waters.

MATERIALS AND METHODS

Site Characteristics and Sampling

The Louisiana shelf near the Mississippi River delta in the nGoM is one of the largest seasonal hypoxic zones of the coastal ocean (Rabalais et al., 2002; Rabouille et al., 2008). These sediments contain high concentrations of particulate organic carbon (POC) which decrease slightly from the River mouth (1.4% OC dry weight) to the shelf (1.1% OC dw) (Goni et al., 1997; Gordon and Goni, 2004; Bianchi et al., 2014). Although terrestrial organic carbon from C3 and C4 plants is present, a decreasing proportion of terrestrial organic carbon and a progressive replacement by marine organic matter is observed westward (Waterson and Canuel, 2008) with a proportion of 60–80% of marine OC in shelf sediments. Most of the recycled carbon is derived from marine organic carbon as attested by the $\delta^{13}\text{C}$ signature of bottom water DIC (Wang et al., 2018). The export of POC to the Mississippi canyon is also significant (Bianchi et al., 2006) and probably related to mobile mud resuspension and transport (Corbett et al., 2004, 2007). Sedimentation rates indicate large sedimentation near the river outlet ($>2.5\text{ cm y}^{-1}$) with values ranging between 0.1 and 0.7 cm y^{-1} on the Louisiana shelf (Gordon and Goni, 2004; Corbett et al., 2006).

The MissRhoDia 1 cruise was conducted in the nGoM in the plume of the Mississippi River on the R/V *Savannah*, operated by the Skidaway Institute of Oceanography at the University of Georgia (United States), from July 19th to August 13th 2017. Four stations were investigated on the Louisiana continental shelf along a 170 km east-west transect from the river mouth to the open shelf (Figure 1). Station 2b was located near the Mississippi River delta at 65 m depth, whereas the three other stations (4, MK, and 5B) were located on the Louisiana shelf around the 20 m isobath in and around the reported hypoxic zone. Indeed, the reported hypoxia frequency for these three stations are 50–100% during the hypoxic season (Rabalais and Turner, 2019). Four other stations were investigated during the same cruise

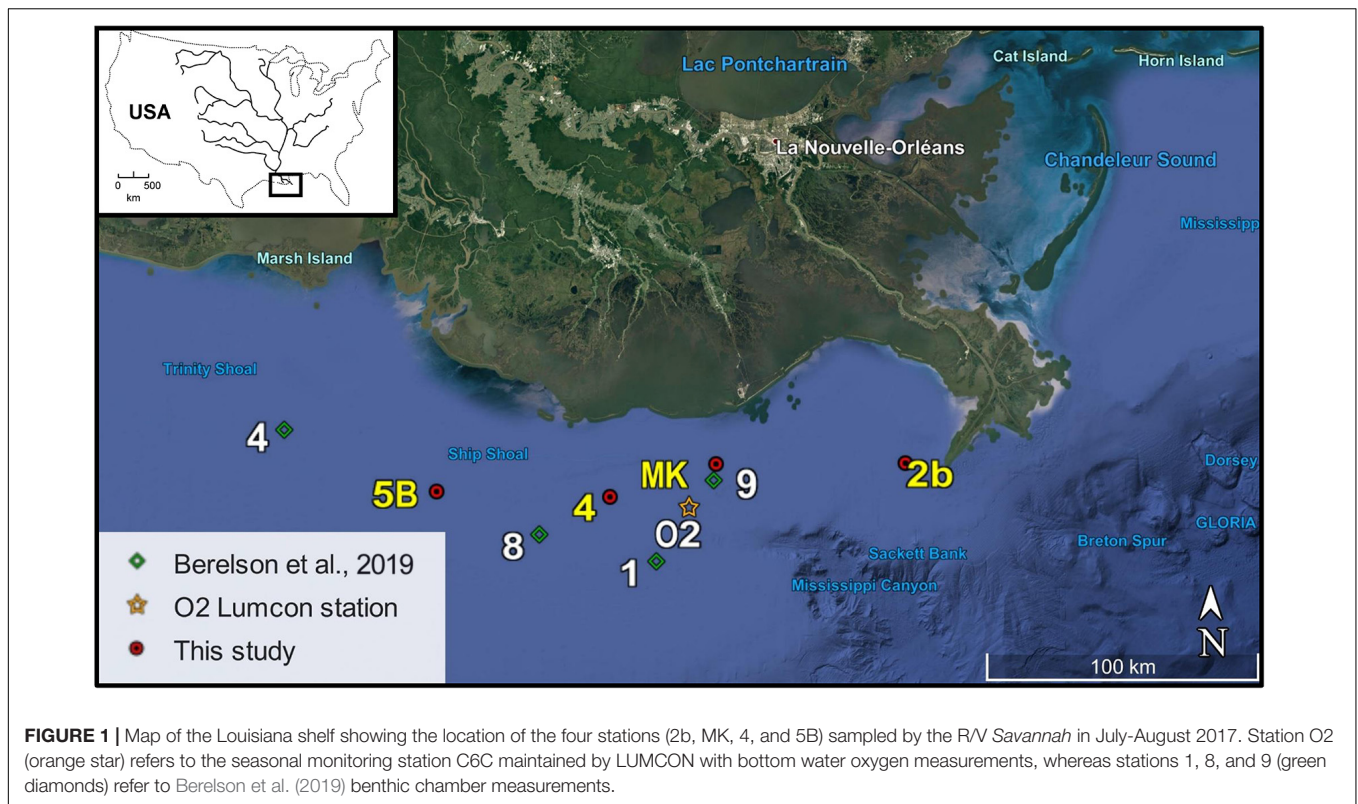
on the slope and are presented in a companion paper (Owings et al., 2021). At each station, CTD profiles were measured, bottom water was sampled with Niskin bottles, one microprofiler deployment was conducted and three sediment cores were collected using a MC-800 multicorer (Ocean Instruments) and immediately treated on board: one for porosity, one for voltammetry followed by porewater extraction, and one for DET measurements. In addition, dissolved oxygen time series in bottom waters obtained in 2017 at station C6C are included, as a subset of the time series obtained over the last 20 years by LUMCON (Rabalais and Turner, 2019).

CTD and Bottom Water Measurements

Seawater samples were collected using a rosette system (10 × 6-L Niskin bottles) equipped with a conductivity-temperature-depth sensor (CTD, Seabird® SBE 19), a dissolved O₂ sensor (Seabird® SBE 43), a turbidity sensor (ECO-BB, Wet Lab®) and a fluorometer (Seapoint®). The temperature, conductivity, turbidity and fluorometer probes were calibrated by the manufacturer. Bottom water samples were collected using the CTD/Rosette as close as possible to the seafloor approximately 1.5–2 m above the seabed. Temperature and salinity were immediately measured using a manual thermometer (precision 0.1°C) and a thermo-salinometer (WTW 331, precision 0.1). DIC and TA samples were collected in 500 mL borosilicate glass bottles, poisoned with 100 μL of saturated HgCl₂ and measured by the SNAPO-CO₂ at the LOCEAN French laboratory (Paris, France). These analyses were performed by a potentiometric closed-cell titration (DOE, 1994), on 120 mL sample volumes, at 25°C and with a precision of 0.15%. Bottom water pH was measured on board in triplicates by the unpurified m-cresol spectrophotometric method of Clayton and Byrne (1993) at 25°C with uncertainties < 0.01 pH units. The pH at *in situ* conditions was then calculated using the CO₂sys program (Pierrot et al., 2006) with *ex situ* pH, TA, salinity, temperature, and pressure as inputs parameters. Following the recommendations of Orr et al. (2015), all calculations with CO₂sys were performed using the acid dissociation constants of carbonate Ka1 and Ka2 reported by Lueker et al. (2000) and total dissolved boron (B_T) reported by Uppström (1974). The *in situ* pH is reported on the total proton scale (pH_T). In order to validate our pH values, we compared the obtained *in situ* pH_T to that calculated using measured DIC and TA (SNAPO-CO₂) at *in situ* pressure, salinity and temperature by the CO₂sys program. These two independent values were in good agreement within 0.015 pH units or better.

Continuous O₂ Records in Bottom Waters

Data from the C6C station (O₂ on Figure 1) were obtained every 15 min for bottom water dissolved oxygen from May 2017 to Nov 2017. The data were collected at 1 m above the seabed using dissolved oxygen meters YSI6600 calibrated before and after deployment with Winkler titrations. Water quality probes were replaced optimally at 4- to 6-week intervals to ensure adequate battery life and to avoid fouling, which was minimal because of low oxygen concentrations and high turbidity at these depths.



Independent measurements of oxygen were also obtained with a profiling YSI6820 on dive trips when changing the water quality probes to provide an independent quality control on the dissolved oxygen concentration of the deployed YSI6600. The YSI6820 was also pre- and post-calibrated and compared to Winkler titrations in the laboratory.

***In situ* Micro-Electrode Measurements in Sediment**

The *in situ* distributions of O₂ and pH at the sediment-water interface (SWI) were obtained using a benthic micro-profiler (Rassmann et al., 2016, 2020) equipped with five Clark Au/Ag oxygen microelectrodes, two pH glass microelectrodes, two pH references (Ag/AgCl) and one resistivity electrode. After landing on the sediment, the sensing head of the lander carrying the microelectrodes moved vertically over 4 cm across the SWI with a resolution of 200 μm. The O₂ microelectrodes were calibrated with a two-point calibration technique using the bottom water O₂ concentration determined by Winkler titration and the anoxic pore waters. The pH microelectrodes were calibrated using NBS buffers (pH 4.00, 7.00, and 9.00 at 20°C), and electrode slopes were used to calculate pH variations (ΔpH) in the pore waters at *in situ* temperature. The ΔpH values were then added to the spectrophotometrically determined pH_T of the bottom waters to obtain pore water pH_T. Signal drift of O₂ and pH microelectrodes during profiling was checked to be less than 5%. Diffusive Oxygen Uptake rates (DOU) were calculated for each micro-profile using the observed O₂ concentration gradient calculated over

400 μm in the sediment, the porosity value and the O₂ diffusion coefficient at *in situ* temperature (Broecker and Peng, 1974) corrected for tortuosity. The oxygen penetration depth (OPD) was operationally defined as the depth where O₂ concentrations dropped below 1 μmol L⁻¹ (Rabouille and Gaillard, 1991).

Sediment Porosity Measurements

For each station, a sediment core was dedicated to porosity measurements. The core was sub-cored using a 50-ml plastic syringe with end cut off and then cut into 2 mm slices on the first centimeter, every 5 mm on the next 9 cm and finally on 2 cm slices for the rest of the core (Cathalot et al., 2010). The sediments from each section were stored in petri dishes and frozen. Porosity was measured in the laboratory based on the salt-corrected mass difference between wet and dry samples (1 week at 60°C) and a sediment density of 2.65 g cm⁻³.

Voltammetric *ex situ* Profiling

A replicate core collected from the same MC800 deployment was profiled in millimeter increments for dissolved oxygen [O_{2(aq)}], thiosulfate (S₂O₃²⁻), total dissolved sulfide (ΣH₂S = H₂S + HS⁻ + S⁰ + S_x²⁻), dissolved iron(II) (Fe²⁺), and dissolved manganese(II) (Mn²⁺) with 100 μm gold/mercury (Au/Hg) voltammetric microelectrodes (Owings et al., 2021). The electrodes were prepared daily and deployed on a computer-controlled micromanipulator with a DLK 70 electrochemical analyzer (Analytical Instrument Systems, Inc.) in a three-electrode system (Ag/AgCl reference, Pt counter, and Au/Hg working electrodes) as reported previously

(Brendel and Luther, 1995; Luther et al., 2008). All measurements were performed according to Owings et al. (2021). Except for $O_{2(aq)}$ which was calibrated using the Winkler-determined overlying water measurements, five-point external calibrations in 0.54 M NaCl were performed with Mn^{2+} as pilot ion, and $S_2O_3^{2-}$, Fe^{2+} , and ΣH_2S were quantified from the pilot ion method (Brendel and Luther, 1995) with the following minimum detection limits (MDLs): $\sim 15 \mu M$ for Mn^{2+} , $\sim 25 \mu M$ for Fe^{2+} , and $\sim 0.2 \mu M$ for ΣH_2S . The electrochemical system also allows detection of organic complexes of Fe(III) [org-Fe(III)] (Taillefert et al., 2000) and aqueous iron sulfide clusters [$FeS_{(aq)}$] as intermediates in the precipitation of $FeS_{(s)}$ (Theberge and Luther, 1997). As the chemical composition of these species is variable and not well defined, they cannot be quantified and are thus reported in normalized current intensities (Meiggs and Taillefert, 2011).

Pore Water Extraction and Analyses

After voltammetric profiles were completed, typically 3 h after core recovery, the same core was transferred to a glove bag (Sigma Aldrich) with a N_2 atmosphere for sectioning in 7–10 mm increments down to 12.5 cm and then in 20 mm increments down to the bottom of the core (Owings et al., 2021). For station 2b, the increment was increased to 50 mm after 20 cm depth. Sediment sections were partitioned into 50 mL Falcon tubes, centrifuged for 10 min at 3,000 rpm under N_2 atmosphere to extract pore waters, which were then immediately filtered through 0.2-micron Whatman® nitrocellulose Puradisc syringe filters into 15 mL Falcon tubes under N_2 atmosphere. The pore waters were then split for onboard analyses of dissolved iron [Fe(II) and Fe(III)] speciation, dissolved manganese (Mn_d), dissolved inorganic phosphate (DIP), DIC, and TA. The remaining aliquots were acidified (HCl 0.1 N) and preserved at 4°C or at $-20^\circ C$ until analysis at Georgia Tech (NH_4^+ , SO_4^{2-} , NO_3^-).

Pore water TA measurements were conducted according to Dickson et al. (2007), with a 3 mL open-cell potentiometric titration (Rassmann et al., 2016) using 0.01 M HCl, which was calibrated every day using Seawater Reference Material (standard batch #150) provided by the Scripps Institution of Oceanography (A. G. Dickson laboratory). Error in this analysis represents the standard deviation from the mean of duplicate samples and averages 0.5%. DIC was determined by flow injection analysis, if necessary after 5 mM $ZnCl_2$ amendment to avoid dissolved sulfide interference (Hall and Aller, 1992) with an uncertainty of 2–3%. Total dissolved Mn (Mn_d) was determined in duplicate using the porphyrin kinetic spectrophotometric method (Madison et al., 2011) modified to account for dissolved Fe^{2+} interferences (Owings et al., 2021). Mn_d was quantified by converting the final absorbance to concentration using the molar absorptivity ($\epsilon = 91.2 \pm 0.4 \times 10^{-3} \mu M cm^{-1}$) determined from external calibration curves with Mn^{2+} (MDL: 0.11 μM , 3σ of blanks, $n = 24$). Dissolved orthophosphates (ΣPO_4^{3-}) were quantified with the methylene blue method after correction for silica interference (Murphy and Riley, 1962). Finally, dissolved Fe^{2+} was quantified using the ferrozine method without hydroxylamine whereas total dissolved Fe (Fe_d) was determined by reacting a separate aliquot with

hydroxylamine hydrochloride (0.2 M) in the dark for 24 h. Dissolved Fe^{3+} was obtained by subtracting the dissolved Fe^{2+} concentration from total dissolved Fe (Stookey, 1970; Viollier et al., 2000). Dissolved ammonium (NH_4^+) was quantified by the indophenol blue method (Strickland and Parsons, 1972) and anions (SO_4^{2-} , NO_2^- , and NO_3^-) via HPLC with UV detection using a 3.2 mM $NaCO_3/1.0$ mM $NaHCO_3$ buffer (SO_4^{2-} ; Minimum detection limit = 114 μM respectively) and 54 mM NaCl (NO_2^- NO_3^- , MDL = 0.12, and 0.11 μM) eluents (Beckler et al., 2014). All standard deviations reported for pore water measurements represent analytical error propagated from calibration curves.

“Diffusive Equilibrium in Thin Films” Profiling

Two-dimensional (2D) distributions of total dissolved iron (Fe_d) were obtained by insertion of a 2D-DET probe into a separate sediment core collected simultaneously at each station. The 2D-DET probe was composed of a polycarbonate plate as support, a $100 \times 40 \times 0.5$ mm hydrated polyacrylamide gel (details in Jézéquel et al., 2007) and a 0.1 mm-thick PVDF membrane (Durapore®). Before deployment, the probe was deoxygenated by N_2 bubbling for at least 3 h. Probes were inserted in the sediment cores for 3 h for chemical equilibrium with the pore waters. During this time, oxygen concentration in the core overlying water was maintained by bubbling an N_2 /air mix. Temperature of the core was also maintained at *in situ* value using a Thermo-Fisher cryostat. At the same time, a 0.5 mm-thick gel was also equilibrated with a solution of ascorbic acid (30 mM) and ferrozine (12.2 mM) to form a reagent gel. After retrieval, the probe gel was laid onto the reagent gel. After 15 min, a magenta color whose intensity is proportional to total iron content was read using an office flatbed scanner. Standardization was realized according to Cesbron et al. (2014) with $(NH_4)_2Fe(SO_4)_2 \cdot 6H_2O$ and ascorbic acid solutions. Scanned images were processed using ImageJ software and the green band was used for calibration and quantification (Jézéquel et al., 2007). For comparison to extracted porewaters, average profiles were calculated using the mean and standard deviations of all pixels within the same depth interval as core slices.

Calculation of the Carbonate Saturation State

Calcium carbonate ($CaCO_3$) saturation states with respect to calcite (Ω_{Ca}) and aragonite (Ω_{Ar}), reported classically as the ion activity product divided by the apparent solubility equilibrium constant of calcite or aragonite, were calculated using the CO2SYS program (Pierrot et al., 2006) following the recommendations of Orr et al. (2015), using measured values of pH, TA, temperature, salinity, DIP, and silicate from Berelson et al. (2019). Calcium carbonate saturation states at a millimeter scale near the SWI were calculated from *in situ* pH profiles and linear interpolation of the centimeter-scale TA profiles. The uncertainties of Ω_{Ca} and Ω_{Ar} were estimated using the standard deviation of the mean *in situ* pH profiles and the pore water TA ($\pm 15 \mu mol kg^{-1}$).

Integrated Reaction Rate Calculations Using PROFILE

To quantify the mineralization of organic carbon in shelf sediments, integrated reaction rates of DIC production were calculated using the PROFILE software (Berg et al., 1998) using different bioirrigation rate constants (α in y^{-1}). Alpha coefficients were constant with depth until 5–6 cm depth below which they rapidly decreased to 0. Four to five values were used for each profile fit in order to bracket the best fit for each station. Measured porosity profiles were used with diffusion coefficients adjusted to the observed water temperature and salinities. Fits to the measured DIC concentration profiles were optimized by the software and computed integrated reaction rates were then converted to $\text{mmol m}^{-2} \text{d}^{-1}$.

RESULTS

Water Column Measurements

Vertical distributions of temperature, salinity, fluorescence, turbidity, and dissolved O_2 concentration in the water column are shown in **Figure 2**. Surface water temperatures ranged from 28 to 30°C, with the lowest temperature observed near the Mississippi River delta (St. 2b), and water temperatures remained consistently high to 20 m in depth ($\geq 28^\circ\text{C}$). The water column showed a haline stratification at all stations, with the lowest surface salinity measured at St. 4 ($S = 12$), which was probably most directly in the path of the freshwater discharge, and the highest surface salinity observed at St. 5B ($S = 24$). In contrast, bottom water salinities were close to marine waters (**Table 1**). This haline stratification (Wiseman et al., 1997) was correlated to a strong gradient in dissolved O_2 concentration from the surface to the bottom. Dissolved O_2 concentrations were slightly supersaturated around $240 \mu\text{mol L}^{-1}$ in surface waters at St. MK and 4, reached close to $190 \mu\text{mol L}^{-1}$ at St. 2b, and were slightly lower at Station 5B ($180 \mu\text{mol L}^{-1}$). Dissolved O_2 concentrations decreased with depth at all stations (**Figure 2**). Bottom water O_2 concentration reached 109 and $106 \mu\text{mol L}^{-1}$ at St. 2b and 4, respectively (**Table 1**). Near-hypoxic values were observed at St. MK ($64 \mu\text{mol L}^{-1}$) while hypoxic values were observed at St. 5B ($18 \mu\text{mol L}^{-1}$). Bottom water pH_T followed the decrease in oxygen concentrations, with larger values at stations 2b and 4, slightly lower pH_T at St. MK, and a minimum was measured at St. 5B (**Table 1**). The Chl-*a* concentration was relatively low ($2\text{--}5 \text{ mg m}^{-3}$) within the uppermost 10 m of the water column at all stations and decreased below the pycnocline (**Figure 2**) to reach a constant value of 2 mg m^{-3} except at St. 2b. The turbidity signal was relatively high above the pycnocline at stations 2b, MK and 4, which are located in the coastal waters around the Mississippi river mouth, and decreased deeper. A nepheloid layer was mostly detected close to the seafloor at St. MK, and to a lesser extent at St. 2b and 5B.

Bottom Water Oxygen Time Series

Dissolved oxygen time series in bottom waters are reported for 7 months at site C6C in the center of the seasonal

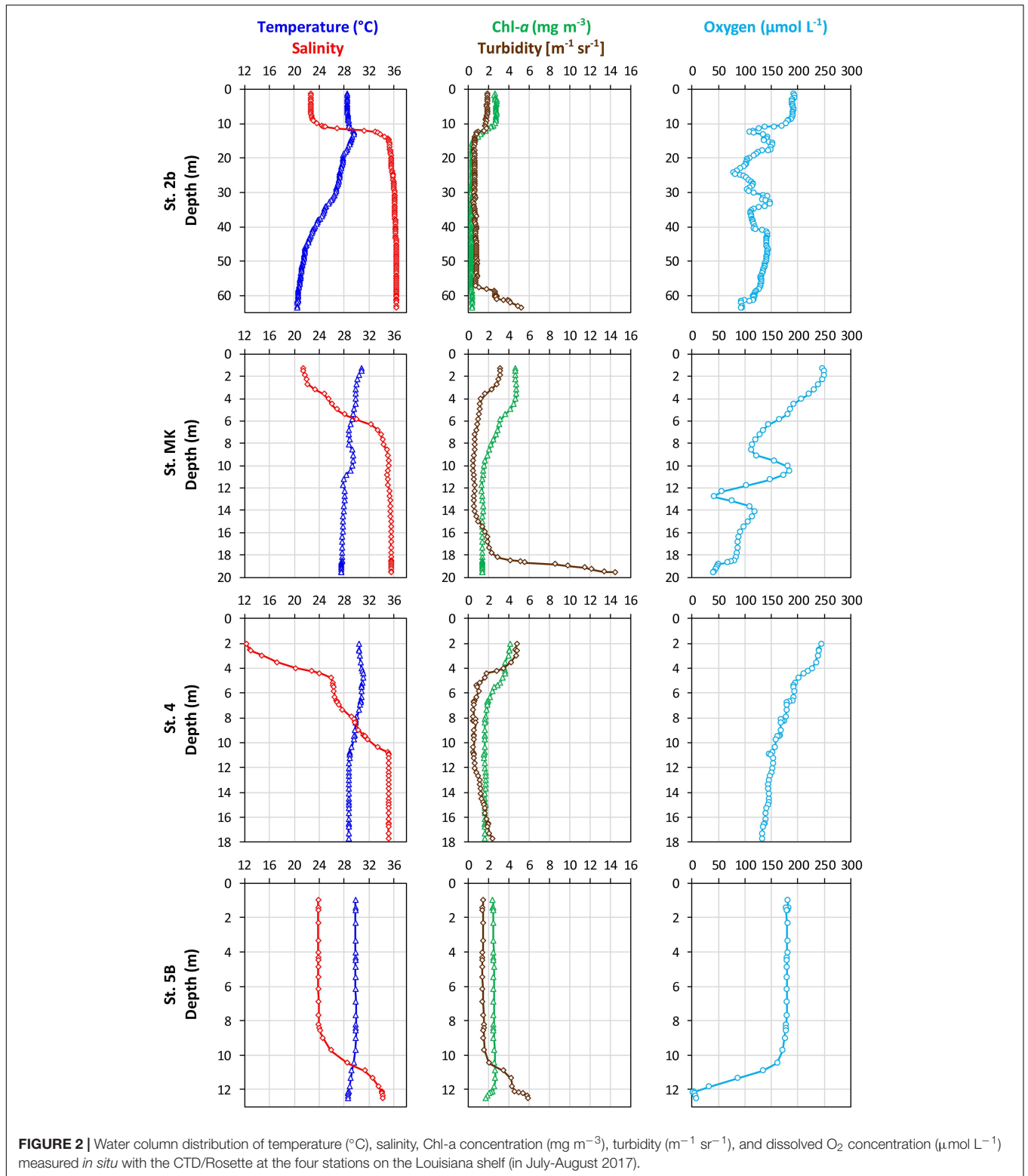
hypoxic zone. The bottom water oxygen concentrations were consistent with independent measurements conducted during maintenance visits (red diamonds). The time series indicated extended periods of hypoxia in May–June, July, and August–September with re-oxygenation of the bottom waters in June after the tropical storm Cindy on June 21st and in September–October (**Figure 3**). Periods of rapid consumption of oxygen occurred during the summer: at the beginning of July (1st July), the oxygen concentration in the bottom waters dropped from $150 \mu\text{mol L}^{-1}$ to hypoxia ($63 \mu\text{mol L}^{-1}$) in 30 h. Similarly, on July 16th, the oxygen concentration switched from hypoxia to near anoxia in only 2 days.

Oxygen and pH Micro-Profiles

The *in situ* distribution of oxygen at the SWI revealed a large decrease of oxygen concentration in the first millimeters of the sediment (**Figure 4**). Stations MK and 5B, which showed the lowest oxygen concentration in the bottom water, nearly hypoxic and hypoxic respectively, showed highly reproducible profiles and were characterized by average OPD of 0.4–0.6 mm (**Table 2**). In contrast, St. 4 and 2b, in oxic conditions, displayed more heterogeneous sediment profiles with average OPD of 1.6–2.0 mm. DOU rates ranged between 5 and $15 \text{ mmol m}^{-2} \text{d}^{-1}$ and decreased with decreasing bottom water O_2 from St. 2b to St. 5B (**Table 2**).

The *in situ* pH_T microprofiles acquired using the benthic micro-profiler were characterized by large pH decreases below the SWI (**Figure 5A**). Two types of profiles were observed. At the shelf stations (4, MK, and 5B), pH showed a large decrease in the oxic zone (a few millimeters) and stabilized around 6.8–7.0 at depth. On the contrary, pH profiles at the river mouth station (St. 2b) showed a less intense decrease across the oxic-anoxic interface, a rebound below the OPD, followed by a secondary decrease to reach a minimum value around 7.4 ± 0.1 at depth (**Figure 5**). The pH stabilized within 2 cm below the SWI in all profiles.

In bottom waters, Ω_{Ca} and Ω_{Ar} showed oversaturation with respect to calcite and aragonite. They ranged between 2.5 and 4 for calcite and between 1.7 and 2.6 for aragonite, with maximal values at St. 4 and minimal values at St. 5B (**Figure 5B**). At St. 2b, the pore water saturation state decreased with depth across the oxycline. Although pore waters remained oversaturated with respect to calcite ($\Omega_{Ca} > 1$), pore waters reached aragonite saturation ($\Omega_{Ar} = 1$) at the OPD at this station. As the pH rebounded deeper in the sediment, the pore waters returned to oversaturation with respect to both calcite and aragonite and decreased again during anoxic diagenesis. Similarly, at St. 4, MK, and 5B, Ω_{Ca} and Ω_{Ar} both decreased across the SWI from above saturation in the overlying waters to near saturation or undersaturation at the OPD. However, whereas Ω_{Ca} remained close to one at all stations by 3 cm depth, Ω_{Ar} showed clear undersaturation below the SWI reaching values as low as 0.50, 0.64, and 0.45 at depth at St. 4, MK, and 5B. Thus, dissolution of aragonite likely occurs in these sediments bathed with hypoxic and acidified waters, whereas calcite dissolution appears to be limited in surface sediments.



Pore Water Profiles and DET Distributions

Pore water profiles showed diagenetic footprints with generally increasing DIC, TA, NH_4^+ , and DIP concentrations downcore

(Figure 6). However, pore water profiles were characterized by contrasting signatures depending on their location: St. 2b displayed large DIC and TA production in pore waters with deeper concentrations greater than 10 mmol kg^{-1} , accompanied

TABLE 1 | Position, depth, and sampling date of the four stations on the Louisiana shelf (nGoM) along with parameters recorded in the bottom waters (T, S, O₂, pH, TA, and DIC).

Station	Date	Lat (°N)	Long (°W)	Depth (m)	T _{BW} (°C)	S _{BW}	O _{2-BW} (μmol L ⁻¹)	pH _{T-BW}	TA _{BW} (μmol kg ⁻¹)	DIC _{BW} (μmol kg ⁻¹)
2b	Aug-6-2017	28°54.0	89°35.4	64.5	20.9	36.5	108.8 ± 0.2	7.88 ± 0.01	2,420 ± 4	2,218 ± 3
MK	Jul-26-2017	28°54.1	90°17.8	20.3	27.8	35.9	64.0 ± 0.5	7.78 ± 0.01	2,383 ± 4	2,184 ± 3
4	Jul-29-2017	28°47.7	90°41.5	19.2	35.8	106.2 ± 0.4	7.82 ± 0.01	2,389 ± 4	2,164 ± 3	
5B	Aug-1-2017	28°48.6	91°20.3	15.4	28.8	34.2	17.6 ± 0.5	7.62 ± 0.01	2,366 ± 4	2,244 ± 3

pH is reported at in situ condition. The error represents the standard deviation between replicate measurements (n = 3).

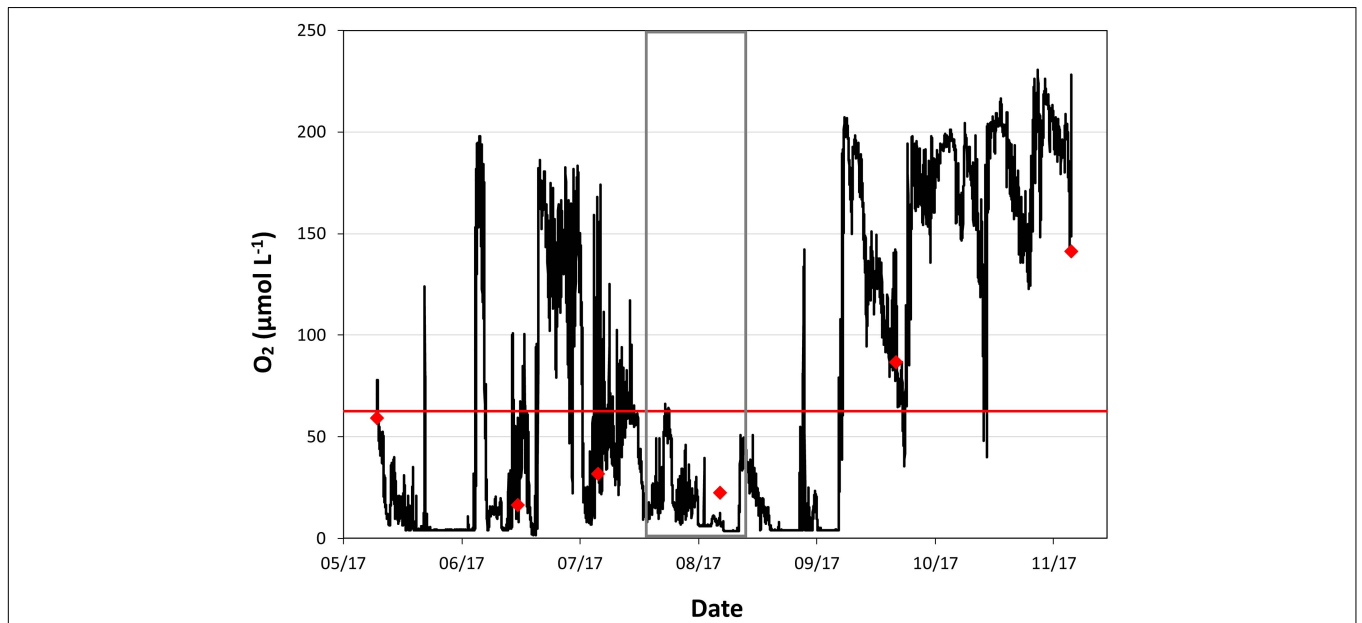


FIGURE 3 | Time series of dissolved oxygen concentration in bottom water (μmol L⁻¹) at Station C6C (O₂ on **Figure 1**) showing substantial variability linked to tropical storms (Cindy on June 21st 2017) and subsequent water column mixing followed by rapid de-oxygenation at the beginning of July 2017 and later. The red line indicates the level of hypoxia (O₂ = 63 μmol L⁻¹). Red diamonds represent dates of sensor replacements and independent measurements of oxygen (see section “Materials and Methods”). Gray rectangle represents the sampling period during MissRhoDia 1 cruise.

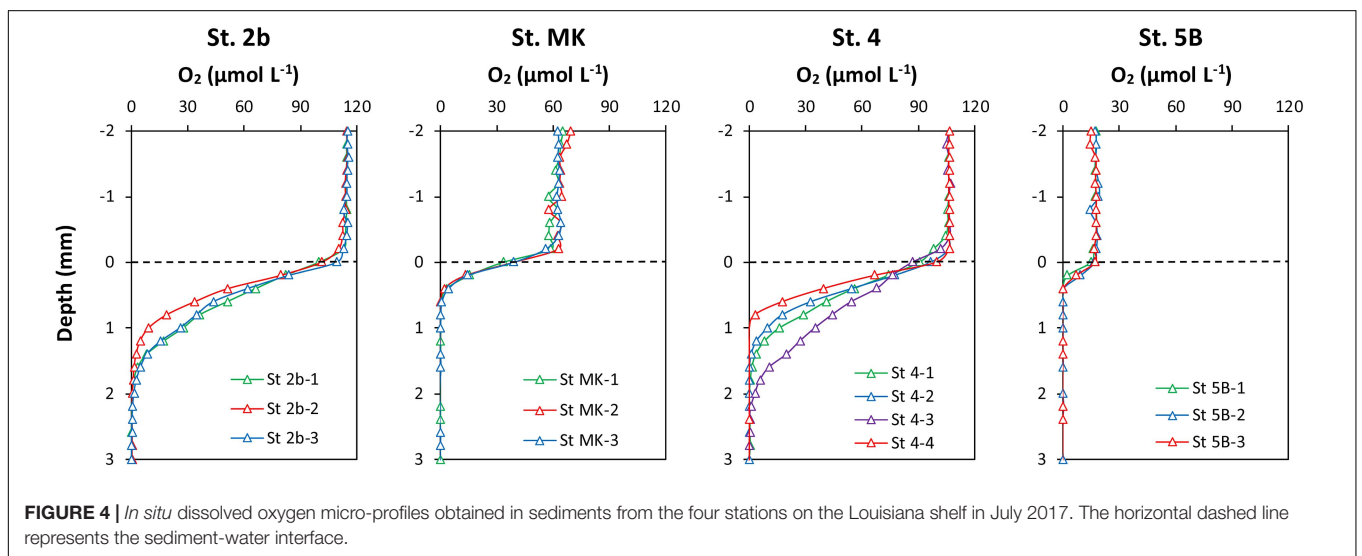


FIGURE 4 | *In situ* dissolved oxygen micro-profiles obtained in sediments from the four stations on the Louisiana shelf in July 2017. The horizontal dashed line represents the sediment-water interface.

TABLE 2 | Oxygen concentration in bottom water (O_2^{BW}), diffusive oxygen uptake (DOU), and oxygen penetration depth (OPD) at the Louisiana shelf in July-August 2017.

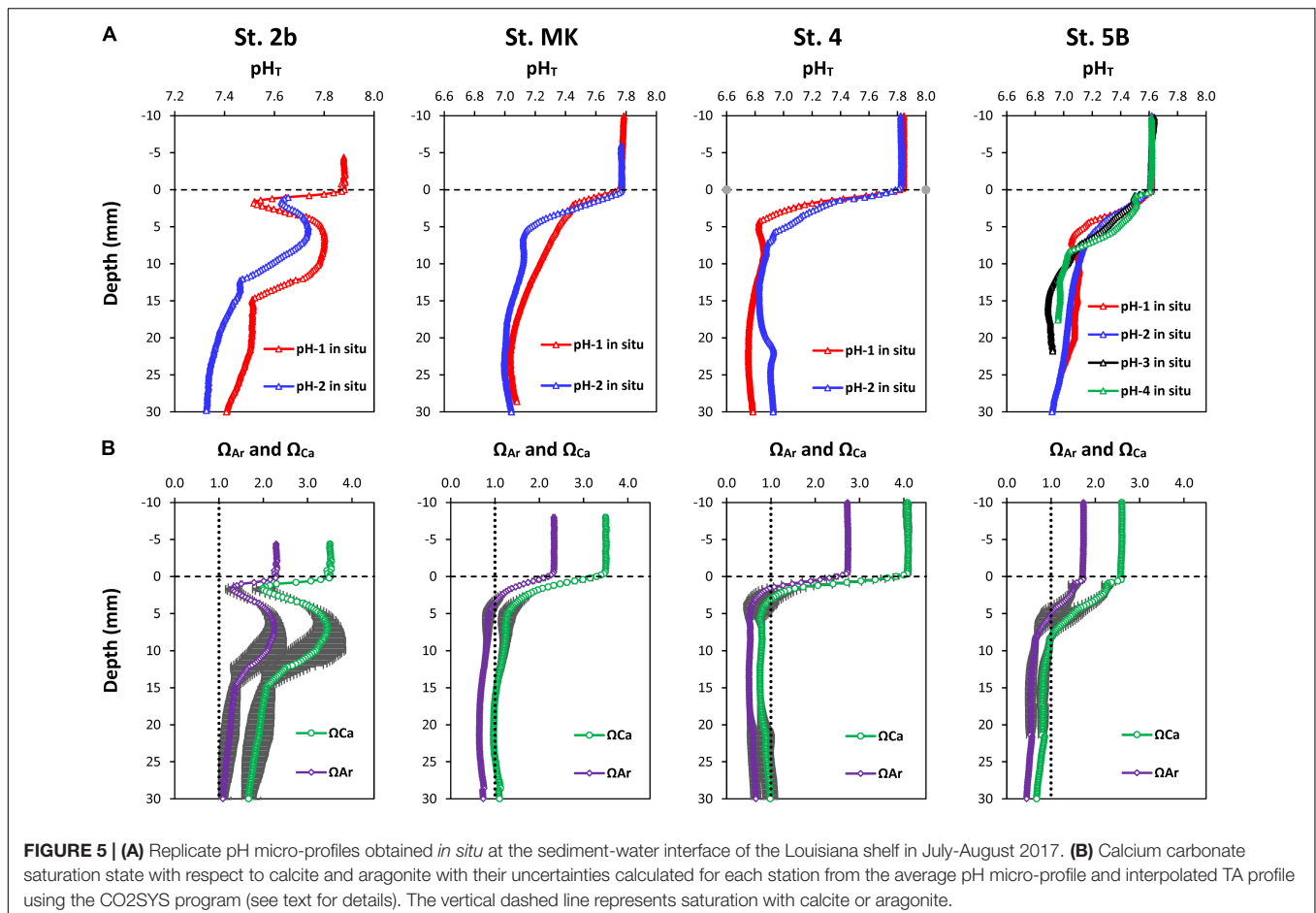
Stations	O_2^{BW} ($\mu\text{mol L}^{-1}$)	DOU ($\text{mmol m}^{-2} \text{d}^{-1}$)	OPD (mm)
2b	108.8	15.1 ± 3.0	2.0 ± 0.2
MK	64	8.0 ± 0.8	0.6 ± 0.0
4	106.2	14.1 ± 6.2	1.6 ± 0.5
5B	17.6	4.8 ± 0.3	0.4 ± 0.0

The DOU and OPD errors represent the standard deviation from the mean values obtained from individual profiles ($n = 3-4$).

by a large SO_4^{2-} concentration decrease of around 10 mM. The production of DIC was correlated with an increase of up to 1 mM NH_4^+ and 300 μM DIP. Nitrate concentration showed a rapid decrease from the bottom water and stabilization at depth to low values (2–3 μM). Dissolved Mn and Fe showed large and similar increases in the first centimeters below the SWI (up to 300 μM) and a slow decrease with depth. It is noteworthy that Fe^{3+} dominated dissolved iron speciation (Figure 6). On the contrary, the diagenetic signatures at all other stations (St. 4, MK, and 5B) were different: they were characterized by slight increases in DIC and TA (2–3 mmol kg^{-1}), with occasional decrease at depth such as observed at St. MK or 5B (Figure 6). Limited decreases in SO_4^{2-} (<2 mM) were visible on profiles,

and small increases of NH_4^+ and DIP concentration (<250 μM and <100 μM respectively) were also observed. Dissolved Fe and Mn showed subsurface peaks of smaller amplitude than St. 2b (around 50 μM Mn and 100 μM Fe) except at St. 5B where the dissolved Mn subsurface peak reached 300 μM .

The 2D distribution of dissolved iron obtained with the DET probe revealed little lateral variability except at the onset of dissolved iron production (Figure 7). In particular, no sign of large disturbance by burrows or other biogenic structures were observed on 2D images except at St. 2b. This is confirmed by the small standard deviation of the pixel lines on the gel image as calculated in the 1-D profiles (black lines). Dissolved iron concentrations from extracted pore waters (red dots) and gel bands averaged over the same depth interval (black dots) were in a good agreement except for St. 2b where the dissolved iron peak in extracted pore waters were about 300 μM and located at 1 cm depth. Such variability between cores suggests heterogeneity at this station, which is the closest to the river delta. Dissolved iron was not detected above 0.5 cm depth at all stations except for St. 4, which showed a single enriched zone at the SWI. At St. 2b, dissolved iron appeared at 2 cm depth, reached 150 μM at 3 cm, and remained stable further down in contrast to the extracted pore water profile. At St. 4 and MK, dissolved iron was present just below the SWI and reached a



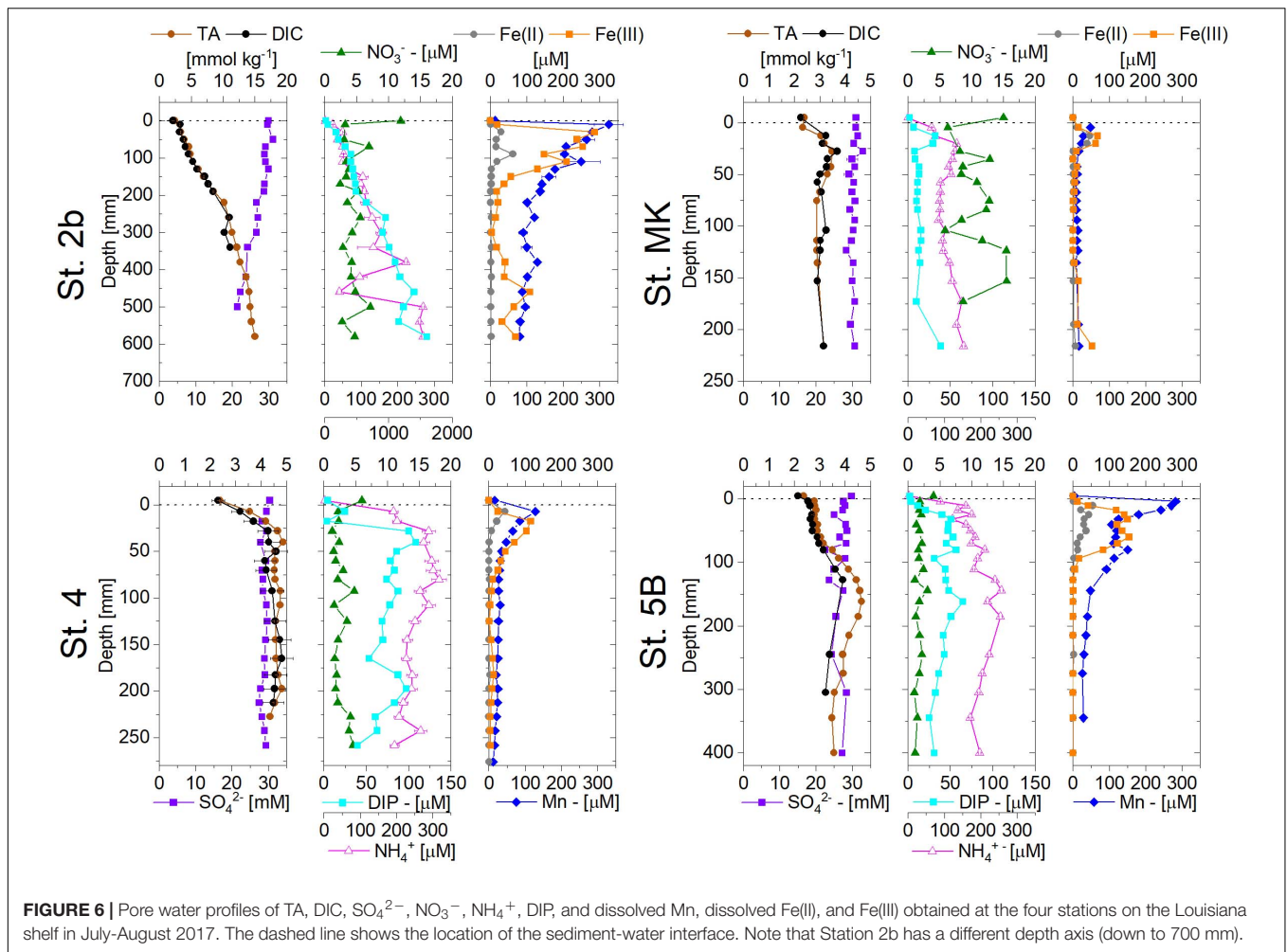


FIGURE 6 | Pore water profiles of TA, DIC, SO_4^{2-} , NO_3^- , NH_4^+ , DIP, and dissolved Mn, dissolved Fe(II), and Fe(III) obtained at the four stations on the Louisiana shelf in July-August 2017. The dashed line shows the location of the sediment-water interface. Note that Station 2b has a different depth axis (down to 700 mm).

maximum concentration of about 100 and 70 μM , respectively, near 2 cm depth. Below, dissolved iron concentration slowly decreased to 50 and 30 μM at 8 cm depth, respectively. Station 5B showed the highest Fe_d concentration in the DET, with 200 μM at 2.5 cm depth.

Voltammetric Profiles

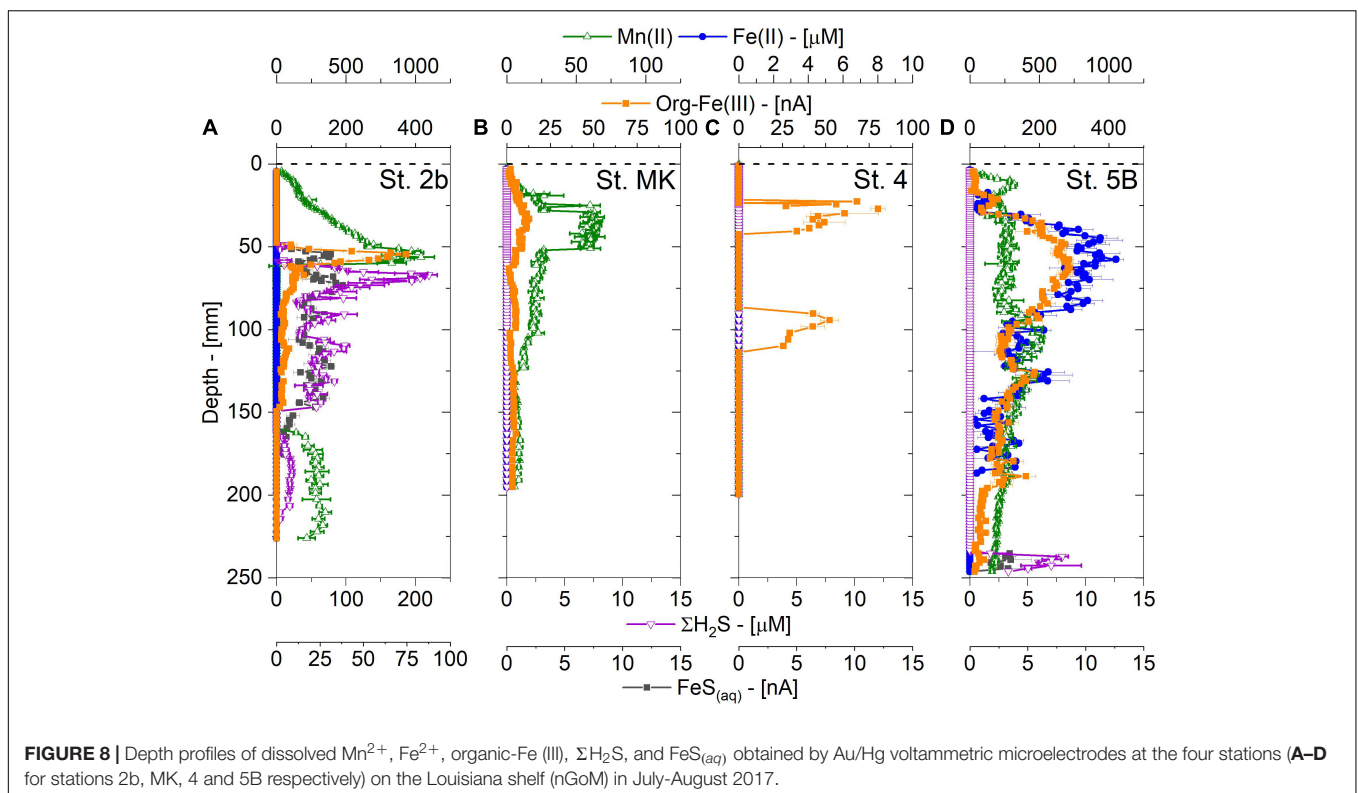
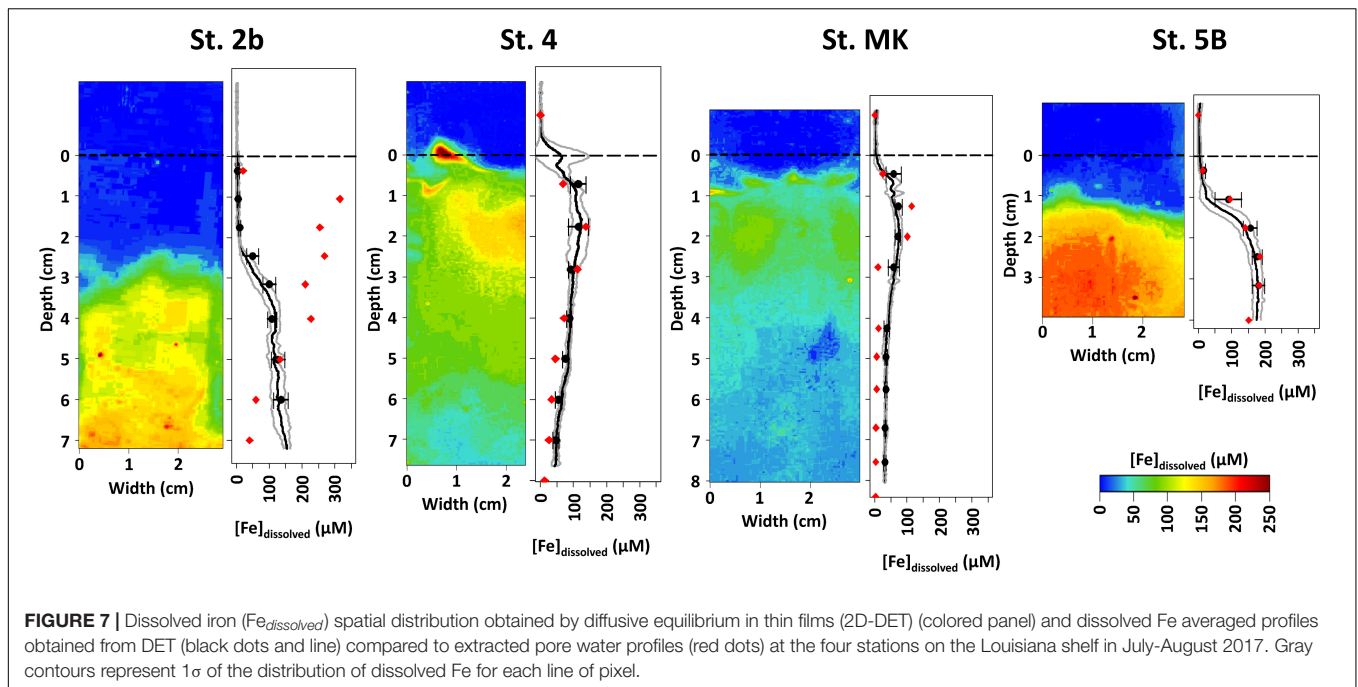
Voltammetric microprofiles measured immediately after sediment core collection revealed that metal reduction occurred at all stations albeit with different intensities (Figure 8). At St. 2b, dissolved Mn^{2+} showed a pronounced peak up to 1,000 μM at 5 cm with lower values in deeper layers, while dissolved iron showed a coinciding peak in org-Fe(III) rather than Fe^{2+} concentrations at 5 cm depth. At the other stations the dissolved metal signals were much lower except at St. 5B, which showed Mn^{2+} and Fe^{2+} peaks of 250 and 750 μM and pervasive occurrence of org-Fe(III) complexes in much larger current intensities. The large concentrations of dissolved Fe at St. 2b and 5B were also visible in the pore water extraction (Figure 6) and 2-D DET profiles (Figure 7).

The most striking difference between St. 2b and the other shelf stations was revealed by the total dissolved sulfide ($\Sigma\text{H}_2\text{S}$)

and $\text{FeS}_{(aq)}$ depth profiles. A $\Sigma\text{H}_2\text{S}$ peak reaching 200 μM was detected between 5 and 15 cm with simultaneous production of $\text{FeS}_{(aq)}$ clusters over the same depth range, suggesting particulate FeS formation at St. 2b which was corroborated by solid phase data (Owings et al., 2021). In contrast, the other stations did not display any detectable dissolved sulfide (except < 10 μM at depth at St. 5B), nor $\text{FeS}_{(aq)}$ clusters throughout these cores.

DISCUSSION

The Louisiana shelf undergoes seasonal hypoxia and acidification, as a result of the biogeochemical mineralization of organic matter and subsequent oxygen consumption and DIC production in the stratified water column. The shelf waters and biogeochemical condition are influenced by the discharge of the Mississippi Atchafalaya River system, increased primary production and thermal warming in spring and summer (Rabalais et al., 2002; Cai et al., 2011; Laurent et al., 2017; Fennel and Testa, 2019). In this discussion, the oxygen depletion in the water column and its dynamics are explored, the effect of diagenetic processes



in generating or maintaining hypoxia and contributing to acidification on the shelf is discussed, including the diagenetic processes underlying this oxygen consumption, and a conceptual framework is presented that may explain both the present results and biogeochemical data obtained from previous studies.

High O_2 Consumption in the Water Column and at the Sediment-Water Interface

Oxygen depletion was particularly developed in the summer 2017 (Rabalais and Turner, 2019). Hypoxia was present all over the

Louisiana shelf, in a continuous band from the Mississippi River delta to the Texas shelf at depths between 10 and 20 m. In fact, the hypoxic zone reached its maximum area in 2017, based on a record dating back to 1985 (Rabalais and Turner, 2019).

The shelf water column measurements from the present study showed a gradient of oxygen depletion from oxic at St. 4, to nearly hypoxic at St. MK and hypoxic at St. 5B (**Figure 3**). The water column of all stations revealed both thermal and haline stratification, although the vertical salinity gradient due to seawater dilution in the Mississippi River plume (6–10 m thick at our stations) contributed mainly to this stratification. As the plume was turbid, primary production was limited and Chl-*a* never exceeded 5 mg m^{-3} at the surface in contrast to previous studies that demonstrated larger chlorophyll content in the distal plume in spring and summer (Dagg et al., 2008; Lohrenz et al., 2008). The presence of a significant concentration of Chl*a* in bottom waters can probably be related to freshly settled detritus as previously suggested (Dortch et al., 1994). The bottom boundary layer was characterized by an increase in turbidity (St. 2b, MK, and 5B), which can be related to the nepheloid layer (Corbett et al., 2004, 2007). Oxygen concentrations were at or below the hypoxic level at stations MK and 5B, showing a clear sign of oxygen consumption in the shelf waters (Turner et al., 2008). It has been proposed that the “time to reach hypoxia” from oxic conditions could be around 1 month on the Louisiana shelf (ranging from a few weeks to 2 months; Chin-Leo and Benner, 1992; Dortch et al., 1994; Amon and Benner, 1998) as a result of the aerobic microbial degradation of organic matter and water temperature. Time series in bottom waters at St. C6C indicated that oxygen consumption leading to the hypoxic threshold can be even more rapid in summer, reaching a few days (**Figure 2**). As the oxygen consumption in the nGoM bottom waters is attributed to the excess organic matter linked to high primary production during spring algal bloom (Rabalais et al., 2002; Turner et al., 2008) and microbial activity is high in the water column (Dortch et al., 1994; Amon and Benner, 1998), it can be inferred that the observed rapid drop in bottom water oxygen concentration is related to the mineralization of a large quantity of labile organic matter.

In addition to the rapid consumption of oxygen by aerobic microbial respiration in the stratified bottom waters, the surface sediment also plays a prominent role in consuming dissolved oxygen. Large SOC, which represents the sum of DOU and the fauna-mediated O_2 flux, have been already documented in the area (Rowe et al.; 2002; Murrell and Lehrter, 2011; Lehrter et al., 2012), indicating that SOC could significantly contribute to the oxygen drawdown in bottom waters (20–30%; Dortch et al., 1994; Murrell and Lehrter, 2011; Feist et al., 2016). The data presented in **Figure 4**, displaying low OPD in sediments, suggest large oxygen consumption possibly linked to the degradation by aerobic respiration of organic matter deposited at the SWI. Indeed, the DOU rates (**Table 2**) were high at St. 2b, 4, and MK (respectively, 15, 14, and $8 \text{ mmol m}^{-2} \text{ d}^{-1}$) where O_2 concentrations were above hypoxic level. The DOU rate was lower at St. 5B ($4.8 \text{ mmol m}^{-2} \text{ d}^{-1}$) limited by O_2 concentration below the hypoxic threshold. DOU is most often lower than SOC because bioirrigation is not considered in DOU (Glud et al., 1998;

Glud, 2008). Despite this difference, the DOU rates obtained in this study were comparable to the SOC values obtained by Lehrter et al. (2012); $5.2 \pm 2.5 \text{ mmol m}^{-2} \text{ d}^{-1}$ for mid shelf stations), Murrell and Lehrter (2011); shelf average $11.6 \pm 2.2 \text{ mmol m}^{-2} \text{ d}^{-1}$, Rowe et al. (2002); $13 \pm 3 \text{ mmol m}^{-2} \text{ d}^{-1}$ at St. C6 close to C6C and St. MK), and Berelson et al. (2019); 4 and $12 \text{ mmol m}^{-2} \text{ d}^{-1}$ at St. 1 and 9; **Figure 1**). These high oxygen consumptions indicate that the sediment plays a significant role in controlling bottom water oxygen concentrations. As dissolved O_2 was confined to a narrow layer at the top of the sediment column, where labile OM from the water column accumulates (OPD = 0.5–2 mm, **Figure 4**), the consumption of oxygen could largely be due to the mineralization of labile organic matter recently deposited at the SWI. Dissolved metal re-oxidation may also play a role in DOU rates, as diffusive flux calculations indicate that a maximum of 20% of oxygen uptake could be used for Mn re-oxidation at St. 5B. SOC would therefore participate in the overall oxic mineralization and O_2 consumption in the bottom waters. It is noteworthy that this large oxic mineralization at the SWI would also contribute to the acidification of bottom waters, as aerobic respiration produces metabolic CO_2 with negligible TA, therefore decreasing the bottom water pH. This is apparent on the gradient from St. 4 to St. 5B where bottom water pH drops by 0.2 units and O_2 from 106 to $18 \mu\text{mol L}^{-1}$ (**Table 1**). The pH and O_2 values measured in the bottom water are in good agreement with previous data reported for the Louisiana Shelf (see Annex 1; Cai et al., 2011).

Low Intensity of Early Diagenesis in Shelf Sediments

A close examination of the pore water profiles obtained from the Mississippi River delta (St. 2b) to the shelf reveals a contrasted situation. A large increase in metabolite concentrations was observed at St. 2b near the Mississippi River delta where previous papers have indicated high accumulation of terrestrial organic matter (Bianchi et al., 2002). DIC, NH_4^+ , and DIP concentrations reach values of 10 mmol kg^{-1} , 1.5 mM , and $200 \mu\text{M}$, respectively (**Figure 6**), which together indicate a high mineralization activity in these sediments. At the same time dissolved iron and manganese peaked in surface pore waters with concentrations above $300 \mu\text{M}$ in the first few centimeters (**Figures 6–8**) indicating significant Fe and Mn reductions. Accompanied by the large decrease in SO_4^{2-} (**Figure 6**) and the presence of dissolved sulfide and $\text{FeS}_{(aq)}$ (**Figure 8**), pore water data indicates active sulfate reduction at this station (St. 2b).

In contrast, metabolite concentration increases observed at the other shelf stations (4, MK, and 5B) were limited with maximum DIC concentrations of 5 mmol kg^{-1} ($\Delta\text{DIC} < 3 \text{ mmol kg}^{-1}$), and low NH_4^+ production ($< 100 \mu\text{M}$). At the same time, low SO_4^{2-} decrease in pore waters ($\Delta\text{SO}_4 < 2 \text{ mM}$) was observed at these three stations. The shelf pore water data are similar to other profiles obtained in the same area [St. C06 – Devereux et al. (2019); St. Z02 – Devereux et al. (2015) and Laurent et al. (2016)]. In shelf sediments other than St. 2b, sulfate reduction seemed to be minimal as shown by the absence of sulfide or $\text{FeS}_{(aq)}$ clusters in pore waters (**Figure 8**). Only metal diagenesis

(Mn and Fe reduction) appeared to be active in shelf sediments. Both pore water extractions (Figure 6) and 2-D DET profiles (Figure 7) displayed remarkably similar dissolved Fe(II) profiles, which showed distinct and coherent peaks (80–300 μM) in subsurface sediments indicative of substantial iron reduction in these sediments. Dissolved Mn and Fe peaks were weaker at St. 4 and MK than at St. 5B. This latter station is located westward under the direct influence of the Atchafalaya River, which delivers Mn- and Fe-rich detrital particles to fuel the sedimentary Mn and Fe cycles. It is noteworthy that St. 2b shows similar enhanced Fe and Mn diagenesis probably due to similar inputs of detrital material by the Mississippi River (Owings et al., 2021). Sulfate reduction generally constitutes a major diagenetic pathway in coastal marine sediments which receive significant organic matter inputs (Canfield et al., 1993). The low decrease of SO_4^{2-} in pore water profiles at St. 4, MK, and 5B is clearly indicative that early diagenesis in Louisiana shelf sediments is not intense compared to Station 2b.

In order to quantify the organic carbon mineralization in these shelf sediments, the PROFILE software (Berg et al., 1998) was used and fitted to the DIC profiles of the four investigated stations. As bioirrigation is a major control on benthic fluxes in coastal areas (Glud et al., 1998) and some profiles show a double-bend pattern typical of irrigated profiles (e.g., the DIC profile at St. 2b), different bioirrigation rate constants (α in y^{-1}) were used to calculate integrated DIC production rates in these sediments. In each case, the best fit of the PROFILE model to the data is displayed in Table 3.

Integrated DIC production calculated using the PROFILE model at the River delta station 2b matched the carbon remineralization (16.5 $\text{mmol m}^{-2} \text{d}^{-1}$) rate calculated for the same station from a complete diagenetic model run at steady-state (Owings et al., 2021). Integrated DIC production rates revealed that the mineralization rate at this station (St. 2b) is greater than the average mineralization rate at the other stations on the Louisiana shelf (St. 4, MK, and 5B), even though mineralization rates are highly variable with a strong positive link to bottom water oxygen levels (Table 3). The large DIC production rate at St. 2b is accompanied by a large SO_4^{2-} decrease which indicates prevalent sulfate reduction in sediments near the Mississippi River delta. In contrast, lower DIC productions at St. 4, MK, and 5B (average $6.8 \pm 5 \text{ mmol m}^{-2} \text{d}^{-1}$) were related to limited SO_4^{2-} decreases indicating low sulfate reduction rates. These findings indicate that anoxic

diagenesis was not favored in the nearly-hypoxic (MK) and hypoxic (5B) zones of the Louisiana shelf. These results rather suggest that manganese and iron reduction were the most active pathways and produced the majority of DIC at these two stations, while sulfate reduction was negligible.

Integrated DIC production rates from the present study on the Louisiana shelf ($0.7\text{--}12 \text{ mmol m}^{-2} \text{d}^{-1}$) can be compared to benthic DIC fluxes at steady-state. They fall in the lower range of core incubations measurements (averages $15.2 \pm 2.1 \text{ mmol m}^{-2} \text{d}^{-1}$ for shelf stations; Lehrter et al., 2012) and are lower than *in situ* benthic DIC fluxes measured in August 2011 ($18\text{--}59 \text{ mmol m}^{-2} \text{d}^{-1}$; Berelson et al., 2019). In the case of the Lehrter et al. (2012) study, interannual and seasonal variability may explain the observed difference as half of the measurements were conducted during spring, before deposition of the spring bloom to the sediment. Nevertheless, the benthic DIC fluxes measured at station Z02-Z03 (equivalent to station MK and 5B) were approximately double (or triple) the integrated DIC production rates calculated in the present study. The largest benthic DIC flux ($59 \text{ mmol m}^{-2} \text{d}^{-1}$; Station 9) in Berelson's study could be attributed to a diagenetic hot spot, as the other fluxes (stations 1 and 8) hovered around $18\text{--}19 \text{ mmol m}^{-2} \text{d}^{-1}$, in the upper range of Lehrter's values. Thus, benthic DIC fluxes, either obtained by *in situ* measurements or *ex situ* sediment incubations, showed values that were approximately two to three times larger than the internal production of DIC calculated in the present study using DIC depth profiles and bioirrigation. One possible bias of this comparison is the fact that fluxes and integrated DIC production rates were obtained from different years and seasons. Indeed, although previously published DIC fluxes cover a large range, their average shows a clear difference by a factor of two to three with the integrated DIC production rate calculated in this study.

We can thus hypothesize that such differences could be related to the efficient mineralization of organic matter at the SWI, which is not captured by standard centimeter-scale pore water profiles used for PROFILE calculations. In this hypothesis, oxic mineralization and metabolic CO_2 production at the SWI could be a major component of DIC fluxes measured by whole core incubations despite not being detected in pore water profiles due to the low vertical resolution. Calcium carbonate dissolution near the SWI is another potential process that may increase DIC fluxes (Rao et al., 2014). The surface sediment (Figure 5B) showed near-saturation with respect to calcite ($\Omega_{Ca} = 0.8\text{--}1$) and undersaturation with respect to aragonite ($\Omega_{Ar} = 0.5$). Although carbonate minerals in these sediments probably consist of detritic calcite and foraminifera shells (calcite), to the best of our knowledge, the type of calcium carbonate present in the Louisiana shelf sediments is not known. More importantly, the CaCO_3 content of these sediments is apparently low, with an average 0.5% inorganic carbon by weight (Gordon and Goni, 2004). As the calcium carbonate content is low, with an even lower proportion of aragonite, and porewaters were saturated with calcite in surface sediments, it is reasonable to assume that carbonate dissolution is limited and does not contribute significantly to the DIC flux.

TABLE 3 | Integrated DIC production (in $\text{mmol m}^{-2} \text{d}^{-1}$) for the station studied in 2017.

	Integrated DIC production ($\text{mmol m}^{-2} \text{d}^{-1}$)	O_2 ($\mu\text{mol L}^{-1}$)	α (y^{-1})
St. 2b	16.5	108.8	40
St. MK	7.7	64	40
St. 4	12	106.2	20
St. 5B	0.7	17.6	0

Oxygen concentration in the bottom water is provided together with the irrigation rate constant (α expressed in y^{-1}) found for the best fit of the model to the data.

and sediment interior (**Figure 9**). Bottom waters and surface sediments probably show large mineralization activity, as evidenced by the rapid (**Figures 3, 4**) O_2 consumption related to the large quantities of labile organic matter raining from surface waters during the spring/summer. Conversely, labile substrate availability is probably restricted in the deeper sediments as a result of the low bioturbation year-round because of the dominance of surface deposit feeders (**Figure 9**). In this scenario, oxygen consumption and DIC production at the SWI would be strengthened during the summer period, as long as oxygen remains near the hypoxic threshold, due to the deposition of labile organic matter at the sediment surface (**Figure 9**). In this case, oxygen consumption at the SWI may increase acidification of bottom waters as a result of the DIC produced during oxic degradation (**Supplementary Figure 1**). A powerful counteracting mechanism would rely on alkalinity production during anoxic mineralization. Indeed, in these sediments, metal reduction (Mn and Fe) is active, which may produce significant alkalinity concentration. However, reoxidation of these dissolved reduced metals with oxygen at the SWI generally consumes most of the alkalinity produced by these reduction reactions (Hu and Cai, 2011). This produces limited net TA fluxes and limits acidification mitigation by benthic alkalinity production. Compounded by the lack of sulfate reduction as a result of the decrease in labile organic inputs in the sediment interior, net alkalinity-generating processes such as FeS production and burial following sulfate reduction (Hu and Cai, 2011; Rassmann et al., 2020), which mitigate acidification in the coastal ocean, are most probably limited at the investigated stations.

CONCLUSION

This study investigated early diagenesis in Louisiana shelf sediments using high resolution techniques near the SWI (amperometric, voltammetric, and potentiometric microelectrodes; DET) and state-of-the-art pore water extractions. A spring to fall time series of bottom water oxygen concentration was also used to investigate hypoxia generation and the biogeochemical dynamics of oxygen consumption in the water column.

From these data, it was possible to conclude that organic matter recycling in the water column near the SWI is disconnected from diagenesis in the sediment interior. Indeed, oxygen consumption was particularly active in the water column during summer as a result of the warm temperature and availability of labile organic matter. Dissolved oxygen was rapidly depleted in the water column, and prolonged periods of hypoxia existed during the summer stratification, therefore attesting the intensity of organic matter mineralization and the lability of the organic substrates. The onset of hypoxia was accompanied by large recycling of organic matter at the SWI as evidenced by dissolved oxygen microprofiles and DOU rates. In contrast, early diagenesis in the shelf sediment interior (St. 4, MK, and 5B) was not intense as revealed by the limited increase in pore water metabolites (DIC, NH_4^+ , and DIP), little decrease in SO_4^{2-} concentration

and small DIC production rates in the sediment interior. This leads to the conclusion that hypoxia does not favor anoxic diagenesis on the Louisiana shelf, at least at the investigated stations.

The lack of connectivity between surface and subsurface sediments was attributed to the lack of bioturbation year-round and especially in summer, when fresh labile organic matter is deposited at the sediment surface. Hypoxic conditions, starting at the end of the spring prevent benthic macrofauna from burying fresh organic matter within the sediment, therefore limiting substrate availability in the deeper sediment layers (2–20 cm). This lack of organic matter entrainment to the sediment interior may have a profound effect on the onset and maintenance of hypoxia and acidification in bottom waters. Indeed, aerobic processes at the SWI consume oxygen and produce DIC without alkalinity, which favor hypoxia and acidifies the bottom waters without any “temporary storage” typically associated with mixing by bioturbation in the deeper sediment layers and slow diffusion in the pore waters. In addition, alkalinity-producing processes linked to anoxic diagenesis (production and burial of FeS) are absent in these hypoxic sediments, thus preventing mitigation of overlying water acidification.

DATA AVAILABILITY STATEMENT

The datasets presented in this study can be found in the SEANOE online repository: <https://www.seanoe.org/data/00630/74215/>.

AUTHOR CONTRIBUTIONS

CR and BL wrote the first draft of the manuscript. SO, BB, EM, JR, EE, AB, JB, MT, and CR participated to the cruise, collected the samples, and, with the addition of BL, performed the analysis. NR provided oxygen time series data. All authors contributed to the manuscript by text writing, editions and comments.

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REFERENCES

- Aller, R. C., Aller, J. Y., Zhu, Q., Heilbrun, C., Klingensmith, I., and Kaushik, A. (2019). Worm tubes as conduits for the electrogenic microbial grid in marine sediments. *Sci. Adv.* 5:eaaw3651. doi: 10.1126/sciadv.aaw3651
- Amon, R. M. W., and Benner, R. (1998). Seasonal patterns of bacterial abundance and production in the Mississippi River plume and their importance for the fate of enhanced primary production. *Microb. Ecol.* 35, 289–300. doi: 10.1007/s002489900084
- Baustian, M. M., and Rabalais, N. N. (2009). Seasonal composition of benthic macroinfauna exposed to hypoxia in the northern Gulf of Mexico. *Est. Coast* 32, 975–983. doi: 10.1007/s12237-009-9187-3
- Beckler, J. S., Nuzzio, D. B., and Taillefert, M. (2014). Development of single-step liquid chromatography methods with ultraviolet detection for the measurement of inorganic anions in marine waters. *Limnol. Oceanogr. Meth.* 12, 563–576. doi: 10.4319/lom.2014.12.563
- Berelson, W., McManus, J., Sevmann, S., and Rollin, N. (2019). Benthic fluxes from hypoxia-influenced Gulf of Mexico sediments: Impact on bottom water acidification. *Mar. Chem.* 209, 94–106. doi: 10.1016/j.marchem.2019.01.004
- Berg, P., Risgaard-Petersen, N., and Rysgaard, S. (1998). Interpretation of measured concentration profiles in sediment pore water. *Limnol. Oceanogr.* 437, 1500–1510. doi: 10.4319/lo.1998.43.7.1500
- Bianchi, S. T., DiMarco, S., Cowan, J. H. Jr., Hetland, R. D., Chapman, P., Day, J. W., et al. (2010). The science of hypoxia in the Northern Gulf of Mexico: A review. *Sci. Tot. Environ.* 408, 1471–1484. doi: 10.1016/j.scitotenv.2009.11.047
- Bianchi, S. T., Goni, M. A., Allison, M. A., Chen, M. A., and McKee, B. A. (2014). “Sedimentary carbon dynamics of the Atchafalaya and Mississippi River delta system and associated margin,” in *Biogeochemical dynamics at major river-coastal interfaces: linkages with global change*, eds S. T. Bianchi, M. A. Allison, and W. J. Cai (New York: Cambridge University Press), 473–502. doi: 10.1017/CBO9781139136853.024
- Bianchi, T. S., Allison, M. A., Canuel, E. A., Corbett, D. R., McKee, B. A., Sampere, T. P., et al. (2006). Rapid Export of Organic Matter to the Mississippi Canyon. *Eos* 87, 565–573. doi: 10.1029/2006EO500002
- Bianchi, T. S., Mitra, S., and McKee, B. A. (2002). Sources of terrestrially-derived organic carbon in lower Mississippi River and Louisiana shelf sediments: implications for differential sedimentation and transport at the coastal margin. *Mar. Chem.* 77, 211–223. doi: 10.1016/S0304-4203(01)00088-3
- Breitburg, D., Levin, L. A., Oschlies, A., Grégoire, M., Chavez, F. P., Conley, D. J., et al. (2018). Declining oxygen in the global ocean and coastal waters. *Science* 359:aam7240. doi: 10.1126/science.aam7240
- Brendel, P. J., and Luther, G. W. III (1995). Development of a gold amalgam voltammetric microelectrode for the determination of dissolved Fe, Mn, O₂, and S(-II) in porewaters of marine and freshwater sediments. *Environ. Sci. Tech.* 29, 751–761. doi: 10.1021/es00003a024
- Briggs, K. B., Hartmann, V. A., Yeager, K. M., Shivarudrappa, S., Diaz, R. J., Osterman, L. E., et al. (2015). Influence of hypoxia on biogenic structure in sediments on the Louisiana continental shelf. *Estuar. Coast. Shelf Sci.* 164:147e160. doi: 10.1016/j.ecss.2015.07.019
- Broecker, W. S., and Peng, T. H. (1974). Gas exchange rates between air and sea. *Tellus* 26, 21–35. doi: 10.3402/tellusa.v26i1-2.9733
- Cai, W.-J., Hu, X., Huang, W.-J., Murrell, M. C., Lehrter, J. C., Lohrenz, S. E., et al. (2011). Acidification of subsurface coastal waters enhanced by eutrophication. *Nat. Geosci.* 4, 766–770. doi: 10.1038/ngeo1297
- Canfield, D. E., Jorgensen, B. B., Fossing, H., Glud, R., Gundersen, J., Ramsing, N. B., et al. (1993). Pathways of organic carbon oxidation in three continental margin sediments. *Mar. Geol.* 113, 27–40. doi: 10.1016/0025-3227(93)90147-N
- Capet, A., Cook, P., Garcia-Robledo, E., Hoogakker, B., Paulmier, A., Rabouille, C., et al. (2020). Facing Marine Deoxygenation. *Front. Mar. Sci.* 7:46. doi: 10.3389/fmars.2020.00046
- Cathalot, C., Rabouille, C., Pastor, L., Deflandre, B., Viollier, E., Buscail, R., et al. (2010). Temporal variability of carbon recycling in coastal sediments influenced by rivers: assessing the impact of flood inputs in the Rhone River prodelta. *Biogeosciences* 7, 1187–1205. doi: 10.5194/bg-7-1187-2010
- Cesbron, F., Metzger, E., Launeau, P., Deflandre, B., Delgard, M.-L., Thibault, et al. (2014). Simultaneous 2D Imaging of Dissolved Iron and Reactive Phosphorus in Sediment Porewaters by Thin-Film and Hyperspectral Methods. *Environ. Sci. Technol.* 48, 2816–2826. doi: 10.1021/es404724r
- Chin-Leo, G., and Benner, R. (1992). Enhanced bacterioplankton production and respiration at intermediate salinities in the Mississippi River plume. *Mar. Ecol. Prog. Ser.* 82, 87–103. doi: 10.3354/meps087087
- Clayton, T. D., and Byrne, R. H. (1993). Spectrophotometric seawater pH measurements: total hydrogen ion concentration scale calibration of m-cresol purple and at-sea results. *Deep Sea Res.* 40, 2115–2129. doi: 10.1016/0967-0637(93)90048-8
- Corbett, D. R., Dail, M., and McKee, B. A. (2007). High-frequency time-series of the dynamic sedimentation processes on the western shelf of the Mississippi River Delta. *Contin. Shelf Res.* 27, 1600–1615. doi: 10.1016/j.csr.2007.01.025
- Corbett, D. R., McKee, B. A., and Allison, M. (2006). Nature of decadal-scale sediment accumulation on the western shelf of the Mississippi River delta. *Continental Shelf Res.* 26, 2125–2140. doi: 10.1016/j.csr.2006.07.012
- Corbett, D. R., McKee, B. A., and Duncan, D. (2004). An evaluation of mobile mud dynamics in the Mississippi River deltaic region. *Mar. Geol.* 209, 91–112. doi: 10.1016/j.margeo.2004.05.028
- Dagg, M. J., Bianchi, T. S., McKee, B. A., and Powell, R. (2008). Fates of dissolved and particulate materials from the Mississippi river immediately after discharge into the northern Gulf of Mexico, USA, during a period of low wind stress. *Contin. Shelf Res.* 28, 1443–1450. doi: 10.1016/j.csr.2006.12.009
- Devereux, R., Lehrter, J. C., Beddick, D. L. J., Yates, D. F., and Jarvis, B. M. (2015). Manganese, iron, and sulfur cycling in Louisiana continental shelf sediments. *Contin. Shelf Res.* 99, 46–56. doi: 10.1016/j.csr.2015.03.008
- Devereux, R., Lehrter, J., Cicchetti, G., Beddick, D. L. J., Yates, D. F., Jarvis, B. M., et al. (2019). Spatially variable bioturbation and physical mixing drive the sedimentary biogeochemical seascape in the Louisiana continental shelf hypoxic zone. *Biogeochem* 143, 151–169. doi: 10.1007/s10533-019-00539-8
- Diaz, R. J., and Rosenberg, R. (2008). Spreading Dead Zones and Consequences for Marine Ecosystems. *Science* 321, 926–929. doi: 10.1126/science.1156401
- Dickson, A. G., Sabine, C. L., and Christian, J. R. (Eds.) (2007). *Guide to Best Practices for Ocean CO₂ Measurements*. Sidney, BC: PICES Special Publication.
- DOE (1994). “Handbook of methods for the analysis of the various parameters of the carbon dioxide system in sea water. Version 2,” in *ORNL/CDIAC-74*, eds A. G. Dickson and C. Goyet (Washington, D.C: DOE).
- Dortch, Q., Rabalais, N. N., Turner, R. E., and Rowe, G. T. (1994). Respiration rates and hypoxia on the Louisiana shelf. *Estuaries* 17, 862–872. doi: 10.2307/1352754
- Feist, T. J., Pauer, J. J., Melendez, W., Lehrter, J., DePetro, P. A., Rygwelski, K. R., et al. (2016). Modeling the relative importance of nutrient and carbon loads, boundary fluxes, and sediment fluxes on Gulf of Mexico hypoxia. *Environ. Sci. Technol.* 50, 8713–8721. doi: 10.1021/acs.est.6b01684
- Fennel, K., and Testa, J. M. (2019). Biogeochemical controls on coastal hypoxia. *Annu. Rev. Mar. Sci.* 11, 4.1–4.26. doi: 10.1146/annurev-marine-010318-095138
- Fennel, K., Hu, J. T., Laurent, A., Marta-Almeida, M., and Hetland, R. D. (2013). Sensitivity of hypoxia predictions for the northern Gulf of Mexico to sediment oxygen consumption and model nesting. *J. Geophys. Res.* 118, 990–1002. doi: 10.1002/jgrc.20077
- Gilbert, D., Rabalais, N. N., Diaz, R. J., and Zhang, J. (2010). Evidence for greater oxygen decline rates in the coastal ocean than in the open ocean. *Biogeosciences* 7, 2283–2296. doi: 10.5194/bg-7-2283-2010

SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2021.604330/full#supplementary-material>

- Glud, R. N. (2008). Oxygen dynamics of marine sediments. *Mar. Biol. Res.* 4, 243–289. doi: 10.1080/17451000801888726
- Glud, R. N., Holby, O., Hoffmann, F., and Canfield, D. E. (1998). Benthic mineralization and exchange in Arctic sediments (Svalbard, Norway). *Mar. Ecol. Prog. Ser.* 173, 237–251. doi: 10.3354/meps173237
- Goñi, M. A., Ruttenberg, K. C., and Eglinton T. I. (1997). Sources and contribution of terrigenous organic carbon to surface sediments in the Gulf of Mexico. *Nature* 389, 275–278. doi: 10.1038/38477
- Gordon, E. S., and Goni, M. A. (2004). Controls on the distribution and accumulation of terrigenous organic matter in sediments from the Mississippi and Atchafalaya River margin. *Mar. Chem.* 92, 331–352. doi: 10.1016/j.marchem.2004.06.035
- Hall, P. J., and Aller, R. C. (1992). Rapid, small-volume, flow injection analysis for SCO₂ and NH₄⁺ in marine and freshwaters. *Limnol. Oceanogr.* 37, 1113–1119. doi: 10.4319/lo.1992.37.5.1113
- Hetland, R. D., and DiMarco, S. F. (2008). How does the character of oxygen demand control the structure of hypoxia on the Texas-Louisiana continental shelf? *J. Mar. Syst.* 70, 49–62. doi: 10.1016/j.jmarsys.2007.1003.1002
- Hu, X. P., and Cai, W. J. (2011). An assessment of ocean margin anaerobic processes on oceanic alkalinity budget. *Glob. Biogeochem. Cycles* 21:GB3003. doi: 10.1029/2010GB003859
- Hu, X. P., Li, Q., Huang, W. J., Chen, B. S., Cai, W. J., Rabalais, N. N., et al. (2017). Effects of eutrophication and benthic respiration on water column carbonate chemistry in a traditional hypoxic zone in the northern Gulf of Mexico. *Mar. Chem.* 194, 33–42. doi: 10.1016/j.marchem.2017.04.004
- Jézéquel, D., Brayner, R., Metzger, E., Viollier, E., Prévot, F., and Fiévet, F. (2007). Two-dimensional determination of dissolved iron and sulfur species in marine sediment pore-waters by thin-film based imaging. Thau lagoon (France). *Estuar. Coast. Shelf Sci.* 72, 420–431. doi: 10.1016/j.ecss.2006.11.031
- Krumins, V., Gehlen, M., Arndt, S., Van Cappellen, P., and Regnier, P. A. G. (2013). Dissolved inorganic carbon and alkalinity fluxes from coastal marine sediments: model estimates for different shelf environments and sensitivity to global change. *Biogeochem* 10, 371–398. doi: 10.5194/bg-10-371-2013
- Laurent, A., Fennel, K., Cai, W. J., Huang, W. J., Barbero, L., and Wanninkhof, R. (2017). Eutrophication-induced acidification of coastal waters in the northern Gulf of Mexico: Insights into origin and processes from a coupled physical biogeochemical model. *Geophys. Res. Lett.* 44:2016GL071881. doi: 10.1002/2016GL071881
- Laurent, A., Fennel, K., Wilson, R., Lehrter, J., and Devereux, R. (2016). Parameterization of biogeochemical sediment–water fluxes using in situ measurements and a diagenetic model. *Depart. Oceanogr.* 13, 77–94. doi: 10.5194/bg-13-77-2016
- Lehrter, J. C., Beddick, D. L. J., Devereux, R., Yates, D. F., and Murrell, M. C. (2012). Sediment-water fluxes of dissolved inorganic carbon, O₂, nutrients, and N₂ from the hypoxic region of the Louisiana continental shelf. *Biogeochem* 109, 233–252. doi: 10.1007/s10533-011-9623-x
- Levin, L. A., Ekau, W., Gooday, A. J., Jorissen, F. J., Middelburg, J. J., Naqvi, W. A., et al. (2009). Effects of natural and human-induced hypoxia on coastal benthos. *Biogeochem* 6, 2063–2098. doi: 10.5194/bg-6-2063-2009
- Lin, S., and Morse, J. W. (1991). Sulfate reduction and iron sulfide mineral formation in Gulf of Mexico anoxic sediments. *Amer. J. Sci.* 201, 55–89. doi: 10.2475/ajs.291.1.55
- Lohrenz, S. E., Redalje, D. G., Cai, W. J., Acker, J., and Dagg, M. J. (2008). A retrospective analysis of nutrients and phytoplankton productivity in the Mississippi River plume. *Contin. Shelf Res.* 28, 1466–1475. doi: 10.1016/j.csr.2007.06.019
- Lueker, T. J., Dickson, A. G., and Keeling, C. D. (2000). Ocean pCO₂ calculated from dissolved inorganic carbon, alkalinity, and equations for K₁ and K₂: validation based on laboratory measurements of CO₂ in gas and seawater at equilibrium. *Mar. Chem.* 70, 105–119. doi: 10.1016/S0304-4203(00)00022-0
- Luther, G. W., Glazer, B., Ma, S., Trouwborst, R., Moore, T., Metzger, E., et al. (2008). Use of voltammetric solid-state (micro)electrodes for studying biogeochemical processes: Laboratory measurements to real time measurements with an in situ electrochemical analyzer (ISEA). *Mar. Chem.* 108, 221–235. doi: 10.1016/j.marchem.2007.03.002
- Madison, A. S., Tebo, B. M., and Luther, G. W. III (2011). Simultaneous determination of soluble manganese(III), manganese(II) and total manganese in natural (pore)waters. *Talanta* 84, 374–381. doi: 10.1016/j.talanta.2011.01.025
- McKee, B. A., Aller, R. C., Allison, M. A., Bianchi, T. S., and Kineke, G. C. (2004). Transport and transformation of dissolved and particulate materials on continental margins influenced by major rivers: benthic boundary layer and seabed processes. *Contin. Shelf Res.* 24, 899–926. doi: 10.1016/j.csr.2004.02.009
- Meiggs, D., and Tallefert, M. (2011). The effect of riverine discharge on biogeochemical processes in estuarine sediments. *Limnol. Oceanogr.* 56, 1797–1810. doi: 10.4319/lo.2011.56.5.1797
- Metzger, E., Langlet, D., Viollier, E., Koron, N., Riedel, B., Stachowitsch, M., et al. (2014). Artificially induced migration of redox layers in a coastal sediment from the Northern Adriatic. *Biogeochem* 11, 2211–2224. doi: 10.5194/bg-11-2211-2014
- Middelburg, J. J., and Levin, L. A. (2009). Coastal hypoxia and sediment biogeochemistry. *Biogeochem* 6, 1273–1293. doi: 10.5194/bg-6-1273-2009
- Moriarty, J. M., Harris, C. K., Friedrichs, M. A. M., Fennel, K., and Xu, K. H. (2018). Impact of seabed resuspension on oxygen and nitrogen dynamics in the northern Gulf of Mexico: A numerical modeling study. *J. Geophys. Res.* 123, 7237–7263. doi: 10.1029/2018JC013950
- Morse, J. W., and Eldridge, P. M. (2007). A non-steady state diagenetic model for changes in sediment biogeochemistry in response to seasonally hypoxic/anoxic conditions in the “dead zone” of the Louisiana shelf. *Mar. Chem.* 106, 239–255. doi: 10.1016/j.marchem.2006.02.003
- Morse, J. W., and Rowe, G. T. (1999). Benthic biogeochemistry beneath the Mississippi river plume. *Estuaries* 22, 206–214. doi: 10.2307/1352977
- Murphy, J., and Riley, J. P. (1962). A modified single solution method for the determination of phosphate in natural waters. *Analyt. Chim. Acta* 27, 31–36. doi: 10.1016/S0003-2670(00)88444-5
- Murrell, M. C., Aukamp, J. R., Beddick, D. L. J., Devereux, R., Greene, R. M., Hagy, J. D. I., et al. (2013). *Gulf of Mexico hypoxia research program data report: 2002–2007*. Washington, DC: U. S. Environmental Protection Agency.
- Murrell, M., and Lehrter, J. (2011). Sediment and lower water column oxygen consumption in the seasonally hypoxic region of the Louisiana continental shelf. *Estuar. Coasts* 34, 912–924. doi: 10.1007/s12237-010-9351-9
- Pörtner, H.-O., Roberts, D. C., Masson-Delmotte, V., Zhai, P., Tignor, M., Poloczanska, E., et al., (eds.) (In press). *IPCC Special Report on the Ocean and Cryosphere in a Changing Climate*.
- Orr, J. C., Epitalon, J. M., and Gattuso, J. P. (2015). Comparison of ten packages that compute ocean carbonate chemistry. *Biogeosciences* 12, 1483–1510. doi: 10.5194/bg-12-1483-2015
- Orr, J. C., Fabry, V. J., Aumont, O., Bopp, L., Doney, S. C., Feely, R. A. et al. (2005). Anthropogenic ocean acidification over the twenty-first century and its impact on calcifying organisms. *Nature* 437, 681–686. doi: 10.1038/nature04095
- Owings, S., Brethous, L., Eitel, E. M., Fields, B. P., Boever, A., Beckler, J., et al. (2021). Differential manganese and iron recycling and transport in continental margin sediments of the Northern Gulf of Mexico. *Mar. Chem.* 229:103908. doi: 10.1016/j.marchem.2020.103908
- Pierrot, D., Lewis, E., and Wallace, D. W. R. (2006). *MS Excel Program Developed for CO₂ System Calculations*. ORNL/CDIAC-105a. Tennessee: Carbon Dioxide Information Analysis Center.
- R Core Team (2017). *R: A Language and Environment for Statistical Computing*. Vienna: R Foundation for Statistical Computing. Available online at: <http://www.R-project.org/>
- Rabalais, N. N., and Baustian, M. M. (2020). Historical shifts in benthic infaunal diversity in the northern Gulf of Mexico since the appearance of seasonally severe hypoxia. *Divers* 12:49. doi: 10.3390/d12020049
- Rabalais, N. N., and Turner, R. E. (2019). Gulf of Mexico Hypoxia: Past, Present, and Future. *Limnol. Oceanogr. Bull.* 28, 117–124. doi: 10.1002/lob.10351
- Rabalais, N. N., Cai, W. J., Cartensen, J., Conley, D. J., Fry, B., Hu, X., et al. (2014). Eutrophication-driven deoxygenation in the coastal ocean. *Oceanography* 27, 172–183. doi: 10.5670/oceanog.2014.21
- Rabalais, N. N., Harper, D. E. J., and Turner, R. E. (2001a). “Responses of nekton and demersal and benthic fauna to decreasing oxygen concentrations,” in *Coastal hypoxia: consequences for living resources and ecosystems. Coastal and estuarine studies*, eds N. N. Rabalais and R. E. Turner (Washington, D.C.: American Geophysical Union), 115–128. doi: 10.1029/CE058p0115
- Rabalais, N. N., Smith, L. E., Harper, D. E. J., and Justic, D. (2001b). “Effects of seasonal hypoxia on continental shelf benthos,” in *Coastal hypoxia:*

- consequences for living resources and ecosystems. *Coastal and estuarine studies*, eds N. N. Rabalais and R. E. Turner (Washington, D.C: American Geophysical Union), 211–240. doi: 10.1029/CE058p0211
- Rabalais, N. N., Turner, R. E., and Wiseman, W. J. (2002). Gulf of Mexico hypoxia, a.k.a. “the dead zone”. *Ann. Rev. Ecol. Syst.* 33, 235–263. doi: 10.1146/annurev.ecolsys.33.010802.150513
- Rabalais, N. N., Turner, R. E., Sen Gupta, B. K., Boesch, D. F., Chapman, P., and Murrell, M. C. (2007). Hypoxia in the northern Gulf of Mexico: does the science support the plan to reduce, mitigate, and control hypoxia? *Estuar. Coast.* 30, 753–772. doi: 10.1007/BF02841332
- Rabouille, C., and Gaillard, J. F. (1991). A coupled model representing the deep-sea organic carbon mineralization and oxygen-consumption in surficial sediments. *J. Geophys. Res. Oceans* 96, 2761–2776. doi: 10.1029/90jc02332
- Rabouille, C., Conley, D. J., Dai, M. H., Cai, W. J., Chen, C. T. A., Lansard, B., et al. (2008). Comparison of hypoxia among four river-dominated ocean margins: The Changjiang (Yangtze), Mississippi, Pearl, and Rhone rivers. *Contin. Shelf Res.* 28, 1527–1537. doi: 10.1016/j.csr.2008.01.020
- Rao, A. M. F., Malkin, S. Y., Montserrat, F., and Meysman, F. J. R. (2014). Alkalinity production in intertidal sands intensified by lugworm bioirrigation. *Estuar. Coast. Shelf Sci.* 148, 36–47. doi: 10.1016/j.ecss.2014.06.006
- Rassmann, J., Eitel, E., Cathalot, C., Brandily, C., Lansard, B., Taillefert, M., et al. (2020). Benthic alkalinity and DIC fluxes in the Rhône River prodelta generated by decoupled aerobic and anaerobic processes. *Biogeosciences* 17, 13–33. doi: 10.5194/bg-17-13-2020
- Rassmann, J., Lansard, B., Pozzato, L., and Rabouille, C. (2016). Carbonate chemistry in sediment porewaters of the Rhône River delta driven by early diagenesis (northwestern Mediterranean). *Biogeosciences* 13, 5379–5394. doi: 10.5194/bg-13-5379-2016
- Redalje, D. G., Lohrenz, S. E., and Fahnenstiel, G. L. (1994). The relationship between primary production and the vertical export of particulate organic matter in a river-impacted coastal ecosystem. *Estuar* 17, 829–838. doi: 10.2307/1352751
- Riedel, B., Pados, T., Pretterebner, K., Schiemer, L., Steckbauer, A., Haselmair, A., et al. (2014). Effect of hypoxia and anoxia on invertebrate behaviour: ecological perspectives from species to community level. *Biogeosc* 11, 1491–1518. doi: 10.5194/bg-11-1491-2014
- Robertson, D., Welsh, D. T., and Teasdale, P. R. (2009). Investigating biogenic heterogeneity in coastal sediments with two-dimensional measurements of iron(II) and sulfide. *Environ. Chem.* 6, 60–69. doi: 10.1071/EN08059
- Rowe, G. T., Kaegi, M. E. C., Morse, J. W., Boland, G. S., and Briones, E. G. E. (2002). Sediment community metabolism associated with continental shelf hypoxia, Northern Gulf of Mexico. *Estuaries* 25, 1097–1106. doi: 10.1007/BF02692207
- Stookey, L. L. (1970). Ferrozine- A New Spectrophotometric Reagent for Iron. *Anal. Chem.* 42, 779–781. doi: 10.1021/ac60289a016
- Strickland, J. D. H., and Parsons, T. R. (1972). *A practical handbook of sea-water analysis*. Ottawa: Fisheries Research Board of Canada, 167.
- Taillefert, M., Bono, A. B., and Luther, G. W. (2000). Reactivity of Freshly Formed Fe(III) in Synthetic Solutions and (Pore)Waters: Voltammetric Evidence of an Aging Process. *Environ. Sci. Technol.* 34, 2169–2177. doi: 10.1021/es990120a
- Theberge, S. M., and Luther, G. W. (1997). Determination of the Electrochemical Properties of a Soluble Aqueous FeS Species Present in Sulfidic Solutions. *Aquat. Geochem.* 3, 191–211. doi: 10.1023/A:1009648026806
- Thibault de Chanvalon, A., Metzger, E., Mouret, A., Knoery, J., Geslin, E., et al. (2017). Two-dimensional mapping of iron release in marine sediments at submillimetre scale. *Mar. Chem.* 191, 34–49. doi: 10.1016/j.marchem.2016.04.003
- Turner, R. E., Rabalais, N. N., and Justic, D. (2006). Predicting summer hypoxia in the northern Gulf of Mexico: Riverine N, P, and Si loading. *Mar. Poll. Bull.* 52, 139–148. doi: 10.1016/j.marpolbul.2005.08.012
- Turner, R. E., Rabalais, N. N., and Justic, D. (2008). Gulf of Mexico hypoxia: Alternate states and a legacy. *Environ. Sci. Technol.* 42, 2323–2327. doi: 10.1021/es071617k
- Uppström, L. R. (1974). The Boron/Chlorinity Ratio of Deep-Sea Water from the Pacific Ocean. *Deep Sea Res. Oceanogr.* 21, 161–162. doi: 10.1016/0011-7471(74)90074-6
- Viollier, E., Inglett, P. W., Hunter, K., Roychoudhury, A. N., and Van Cappellen, P. (2000). The ferrozine method revisited: Fe(II)/Fe(III) determination in natural waters. *Appl. Geochem.* 15, 785–790. doi: 10.1016/S0883-2927(99)00097-9
- Wang, H. J., Hu, X. P., Rabalais, N. N., and Brandes, J. (2018). Drivers of oxygen consumption in the northern Gulf of Mexico hypoxic waters—A stable carbon isotope perspective. *Geophys. Res. Lett.* 45, 10528–10538. doi: 10.1029/2018GL078571
- Waterson, E. J., and Canuel, E. A. (2008). Sources of sedimentary organic matter in the Mississippi River and adjacent Gulf of Mexico as revealed by lipid biomarker and 13CTOC analyses. *Org. Geochem.* 39, 422–439. doi: 10.1016/j.orggeochem.2008.01.011
- Wiseman, W. J., Rabalais, N. N., Turner, R. E., Dinnel, S. P., and MacNaughton, A. (1997). Seasonal and interannual variability within the Louisiana coastal current: stratification and hypoxia. *J. Mar. Syst.* 12, 237–248. doi: 10.1016/S0924-7963(96)00100-5
- Zhang, J., Gilbert, D., Gooday, A., Levin, L., Naqvi, S. W. A., Middelburg, J. J. et al. (2010). Natural and human-induced hypoxia and consequences for coastal areas: synthesis and future developmen. *Biogeosc* 7, 1443–1467. doi: 10.5194/bg-7-1443-2010

Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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