



HAL
open science

Sex-specific costs of reproduction on survival in a long-lived seabird

Marta Cruz-Flores, Roger Pradel, Joël Bried, Jacob González-Solís, Raül Ramos

► **To cite this version:**

Marta Cruz-Flores, Roger Pradel, Joël Bried, Jacob González-Solís, Raül Ramos. Sex-specific costs of reproduction on survival in a long-lived seabird. *Biology Letters*, 2021, 17, pp.20200804. 10.1098/rsbl.2020.0804 . hal-03159171

HAL Id: hal-03159171

<https://hal.science/hal-03159171>

Submitted on 4 Mar 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 **Sex-specific costs of reproduction on survival in a long-lived seabird**

2

3 Marta Cruz-Flores^{1*}, Roger Pradel², Joël Bried³, Jacob González-Solís¹ & Raül Ramos¹

4 ¹ Departament de Biologia Evolutiva, Ecologia i Ciències Ambientals, Facultat de Biologia, Institut
5 de Recerca de la Biodiversitat (IRBio), Universitat de Barcelona, Barcelona, 08028, Spain.

6 ² CEFE UMR 5175, CNRS – Université de Montpellier – Université Paul-Valéry Montpellier – EPHE,
7 Montpellier, France.

8 ³ Centro Okeanos, MARE (Marine and Environmental Sciences Centre), IMAR and LARSyS
9 Associated Lab, Universidade dos Açores, Departamento de Oceanografia e Pescas, 9901-862
10 Horta, Açores, Portugal. Present address: 8 Avenue de la Reine Nathalie, 64200 Biarritz, France.

11

12

13 * Correspondence: Marta Cruz Flores, Departament de Biologia Evolutiva, Ecologia i Ciències
14 Ambientals, Facultat de Biologia, Universitat de Barcelona, Av. Diagonal 643, Barcelona 08028,
15 Spain. Email: martacruzflores@gmail.com, Phone: +34666776064

16

17 **Running Head (40 characters max.):** Sex-specific breeding costs on survival

18

19 **Word count:** 2,767

20

21 Abstract (2,000 characters max.)

22 Costs of reproduction on survival have captured the attention of researchers since Life History
23 Theory (LHT) was formulated. Adults of long-lived species may increase survival by reducing their
24 breeding effort or even skipping reproduction. In this study, we aimed to evaluate the costs of
25 current reproduction on survival and whether skipping reproduction increases adult survival in
26 a long-lived seabird. We used capture-mark-recapture data (1,450 encounters) from two
27 populations of Bulwer's petrel (*Bulweria bulwerii*), breeding in the Azores and Canary Islands,
28 North Atlantic Ocean. Using a multievent model with two different breeding statuses (breeders
29 vs. non-breeders), we calculated probabilities of survival and of transitions between breeding
30 statuses, evaluating potential differences between sexes. Females had lower survival
31 probabilities than males, independent of their breeding status. When considering breeding
32 status, breeding females had lower survival probabilities than non-breeding females, suggesting
33 costs of reproduction on survival. Breeding males had higher survival probabilities than non-
34 breeding males, suggesting that males do not incur costs of reproduction on survival and that
35 only the highest quality males have access to breeding. The highest and the lowest probabilities
36 of skipping reproduction were found in breeding males from the Azores and in breeding
37 males from the Canary Islands, respectively. Intermediate values were observed in the females
38 from both. This result is probably due to differences in the external factors affecting both
39 populations, essentially predation pressure and competition. The existence of sex-specific costs
40 of reproduction on survival in several populations of this long-lived species, may have important
41 implications for species population dynamics.

42

43 Keywords: Breeding costs, Bulwer's petrel, capture-mark-recapture, Life-history, predation,
44 sabbatical

45 **Introduction**

46

47 Life History Theory (LHT) postulates that individuals need to carefully trade off investment in
48 survival, growth and reproduction to maximize lifetime reproductive success [1–3]. As a result,
49 greater investment in one of these traits may result in reduced investment in others. Under
50 different environmental conditions, species, populations or even groups of individuals within
51 the same population, such as a given sex, may invest in life history traits differently, and
52 therefore face different life history trade-offs [4].

53

54 For example, reproduction trades off with survivorship by reducing the allocation of resources
55 for growth or self-maintenance and, depending on the species, the feeding, rearing, and caring
56 of offspring have important consequences for the physical state of progenitors [5]. In addition,
57 reproduction reduces the survivorship of individuals by increasing their vulnerability and
58 predation risk (*e.g.*, when building or attending nests, or when carrying embryos or eggs [6–8]).
59 In turn, predation can modify individual traits such as reproduction in order to reduce predation
60 risk (*e.g.*, reducing the number of offspring when predation risk is high [9,10]). Other aspects of
61 social interactions between populations and individuals, such as competition, can also alter
62 reproduction (*e.g.*, via competition for food, nesting places or mates [11]) or even increase direct
63 mortality [12]. All these costs may differ between sexes since, for example, females typically
64 take on extra energetic costs with parturition or egg-laying, which may compromise their own
65 body condition [13]; and males are most affected by intra and inter-specific competition for nest
66 sites than females [14,15].

67

68 According to the LHT, the trade-off between reproduction and survival differs between short-
69 and long-lived species. To maximize lifetime reproductive success, short-lived species show

70 higher fecundity and tend to invest in current reproductive events to a greater extent than long-
71 lived species, even when this effort may compromise individual survival and future breeding
72 attempts [1,16]. Contrarily, long-lived species have low fecundity, deferred maturity and high
73 adult survival rates [17–19]. Furthermore, progenitors of long-lived species can reduce their
74 reproductive investment and even skip the current breeding attempt, favouring self-
75 maintenance and survival until the next possible breeding attempt (a phenomenon known as
76 reproductive skipping behaviour, intermittent breeding or sabbatical years [20,21]). However,
77 the extent to which the trade-off between survival and reproduction differs between sexes in
78 long-lived species, which experience very different costs associated with reproduction, has been
79 little studied [22–24].

80

81 In this study, we analysed sex-specific costs of reproduction on survival in a long-lived species,
82 the Bulwer's petrel (*Bulweria bulwerii*), at two breeding colonies from two different populations
83 in the Atlantic Ocean. First, as reproduction may reduce the survivorship of individuals [17], we
84 expected lower survival probabilities in breeders than in non-breeders. Second, *a priori*, we did
85 not expect sex-specific costs of reproduction, since in Procellariiformes, females and males share
86 incubation and chick-rearing duties [25]. Females certainly incur the costs of egg production and
87 laying, but this is believed to be counterbalanced by a higher parental investment of males
88 through longer periods of colony attendance during the pre-laying period, longer incubation
89 shifts and/or higher chick provisioning rates [26–28]. Third, in the event of sex-specific costs of
90 reproduction on survival, we would expect the sex with the highest survival to show the highest
91 probability of skipping reproduction, since (a) sabbatical years are considered to increase adult
92 survival [29] and (b) sex-specific survival could imply a limitation in the availability of potential
93 mates [30,31]. Finally, since the predator density and competition for nest sites differs between
94 our two colonies [32,33], we expected to find different effects on the trade-off between survival

95 and reproduction between both. In the colony with higher competition pressure, we expected
96 a lower survival rate for males due to direct mortality from competition, since males tend to
97 defend their nest more than females [34]; and in the colony with higher predation pressure, we
98 expected a high incidence of skipping behaviour for both sexes as a strategy to reduce predation
99 risk until the next possible breeding attempt.

100

101 **Materials and methods**

102 **Species and study colonies**

103 Bulwer's petrel is a small (*ca.* 95 g) Procellariiforme whose maximum recorded longevity is 23
104 years [35]. In the Atlantic Ocean, it breeds on small islets and islands throughout all
105 Macaronesian archipelagos [36]. Like all Procellariiform species, it is socially monogamous, lays
106 a single egg per breeding attempt, and incubation and chick-rearing duties are shared by both
107 sexes [25].

108

109 Fieldwork was conducted at two different North Atlantic colonies situated *ca.* 1370 km apart,
110 namely on Vila Islet (hereafter Vila, Azores Islands; 36°55' N, 25°10' W; *ca.* 50 breeding pairs; J.
111 Bried unpublished data; [37]), where inter-specific competition for nesting places with Cory's
112 shearwaters (*Calonectris borealis*) is the main cause of Bulwer's petrel mortality [12,33]; and on
113 Montaña Clara Islet (hereafter M.Clara, Canary Islands; 29°18' N, 13°32' W; <100 breeding pairs
114 [38]), where the main cause of Bulwer's petrel mortality is presumably predation by Eleonora's
115 falcon (*Falco eleonora*), which breeds on the islet at high densities (77.7 nests/km² [32,38]).

116

117 **Capture-Mark-Recapture (CMR) sampling strategy**

118 At both colonies, we captured birds in their burrows. Each individual was ringed with a unique
119 stainless-steel ring at its first capture, with every subsequent recapture registered. The dataset
120 considered 1,450 encounters from the two study colonies: Vila (n=261 adults, data from 2007-
121 2012 and 2016) and M.Clara (n=416, 2010-2018). The reproductive performance of the
122 individuals was recorded each year during the entire sampling period on Vila, and for 2015-2018
123 on M.Clara. Forty-five females and 46 males on Vila (17.2 and 17.6% of the total number of
124 adults from Vila, respectively), and 24 females and 46 males on M.Clara (5.8 and 11.1%,
125 respectively) were molecularly sexed following Fridolfsson & Ellegren (1999) [39]. The sex
126 remained unknown for the other individuals. Differences in the sex ratios of sexed individuals
127 are due to different sampling strategies and fieldwork schedules on the two Islets. Capture-
128 mark-recapture data are accessible through a public repository [40]. Total numbers of
129 individuals captured at each colony per breeding status, sex and year (and proportion of each
130 combination of breeding status and sex, per year); and M-array summaries by colony and sex
131 are shown in the ESM 1 in Electronic Supplementary Material.

132

133 **CMR models and statistical analyses**

134 We first evaluated whether the general model for single-state data with only time-dependent
135 parameters (Cormack-Jolly-Seber [CJS] model) was an acceptable starting point for our dataset.
136 To do so, we carried out standard Goodness-Of-Fit (GOF) tests using U-CARE 2.3.4 [41].

137

138 Second, to construct our model, we defined five states: breeders and non-breeders, both by sex,
139 and dead birds. After running preliminary models (not shown), we considered that the unsexed
140 individuals were a mixture of 50% females and 50% males when modelling Initial States (IS)
141 probabilities. We modelled survival (ϕ , probability that an individual alive at Year t survives until
142 Year $t+1$) as either constant, conditioned on colony, sex, breeding status (breeders vs. non-

143 breeders), or on the breeding status by sex. In addition, we modelled the probability of
144 Transitions between Breeding Statuses (TBS), since we were interested in the proportion of
145 breeding adults that became non-breeders (*i.e.*, that performed reproductive skipping
146 behaviour). We also checked if differences in TBS occurred between colonies.

147

148 Regarding the detection process, we considered the alternative events of being or not being
149 detected with capture probability (p , probability that a Bulwer's petrel alive and present at the
150 breeding colony at Year t is captured during Year t). We started modelling capture probabilities
151 depending on colony and time to control for the sampling effort and set the years with missing
152 data for each colony. For those individuals detected, we defined six possible events related to
153 their reproductive performance/status and we constructed a multievent model with uncertainty
154 to define how events relate to each of the five states [42]. For more details about the model
155 definition see the ESM 2 in Electronic Supplementary Material.

156

157 CMR models were run using E-SURGE 2.1.4 [43], and model selection was based on the Akaike
158 Information Criterion corrected for overdispersion and small sample sizes (QAICc [44]). The
159 model with the lowest QAICc had the best compromise between bias and variance [45]. In
160 addition, Akaike weights were calculated as an index of the relative plausibility of each model
161 [46]. Following Burnham & Anderson (2002), and once we had selected the best model for IS, Φ ,
162 TBS and p (table 1 section A-D), we explored neighbouring models of interest to re-evaluate
163 early-dropped effects (table 1 section E).

164

165 **Results**

166 The overall GOF test for the CJS model was not significant, making it an acceptable point of
167 reference for further models (see the ESM 3 in Electronic Supplementary Material).

168

169 According to the best-supported models (models 9 and 10, table 1), survival probabilities
170 differed for each combination of breeding status and sex, but not between colonies; and capture
171 probabilities varied among colonies, years, and between breeding statuses. ΔQAICc of these two
172 models did not allow ruling out either of them. The main difference between models 9 and 10
173 was in TBS, identical or different between colonies, respectively. Model 10 made more sense
174 from a biological point of view due to the differences between the two islets in habitat
175 availability for Bulwer's petrel and predation pressure (both lower on Vila than on M.Clara; see
176 Discussion). Model 10 considered that breeding females had lower survival probabilities than
177 non-breeding females (mean \pm SE: 0.64 ± 0.04 and 0.73 ± 0.07 , respectively), whereas breeding
178 males had higher survival probabilities than non-breeding males (0.93 ± 0.04 and 0.80 ± 0.04 ,
179 respectively; figure 1). Thus, breeding females had the lowest survival rate while breeding males
180 had the highest survival rate (figure 1). Concerning TBS, breeding individuals were more likely to
181 breed again the next year than to take a sabbatical year. The probability of skipping reproduction
182 the next year was highest in the breeding males from Vila and lowest in the breeding males from
183 M.Clara (figure 2; see ESM 4 in Electronic Supplementary Material for more details). Despite
184 differences in TBS and in predation pressures between Vila and M.Clara, the model considering
185 inter-colony differences in survival was not a competitive model (model 11, table 1).

186

187 **Discussion**

188 Contrary to our expectations, survival probabilities varied in an opposite manner between
189 breeders and non-breeders when considering sex, highlighting sex-biased costs of reproduction
190 on survival. The lower survival rate observed in breeding females compared to non-breeding

191 females suggests that the former incur costs of reproduction in terms of survival. These costs
192 may arise from egg production and laying, which implies a higher energetic expense than for
193 males. They have been observed in many other species including Procellariiformes [13,24]. The
194 higher survival of breeding males compared to non-breeding males suggests that high quality
195 males are more likely to obtain mates, and less affected by the trade-off between reproduction
196 and survival than low quality males. Similar results were found in related species, including the
197 Monteiro's storm petrel (*Hydrobates monteiroi*), in which breeders (successful or failed,
198 regardless of sex) had higher survival probabilities than non-breeders [47]. High-quality males
199 may be older or more experienced, since individual quality and breeding performance of males
200 in long-lived species are often related to age or experience [48,49]. However, we were unable
201 to address this question. As in most Procellariiform species, Bulwer's petrels cannot be aged
202 using plumage [25], and during our study very few birds ringed as chicks, and therefore of known
203 age, have returned to breed. Overall, the costs of reproduction that breeding males face, from
204 the pre-laying stage until the end of the chick-rearing period, seem to be lower than those faced
205 by breeding females during the same period. Nevertheless, males are supposed to spend more
206 time ashore than females during the pre-laying period, when they have to defend their burrows
207 against conspecifics [34], making them more vulnerable to injuries or predation. They also invest
208 more time and/or resources than females into incubation and chick-rearing [26–28]. Since
209 predators mainly prey on poor-condition individuals [50], a higher exposure of breeding males
210 to predation could imply a selection against low-quality males, which would ultimately explain
211 the high survival of breeding males.

212

213 The sex bias we found in survival rates can affect the adult sex ratio in favour of males [51],
214 resulting in a reduction of the effective population size with respect to census size that assumes
215 sex parity [30,31]. This male-biased sex ratio would imply a higher number of sabbatical males

216 than females due to a lack of potential mates. As expected, we found this result on Vila, where
217 the higher proportion of sabbatical individuals compared to M.Clara, particularly in males (figure
218 2), is probably favoured by a shortage of suitable nesting places and a negligible predation
219 pressure. On Vila, inter-specific competition with Cory's shearwaters for nesting places
220 represents the main cause of Bulwer's petrel mortality [12,33]. This competition might explain
221 the high proportion of breeders skipping breeding the next year and the low proportion of non-
222 breeding individuals resuming breeding on Vila. In contrast, on M.Clara, the proportion of
223 sabbatical males resuming breeding was higher than the proportion of breeding males becoming
224 sabbatical. This result is possibly related to the presumably higher predation pressure on
225 M.Clara [38], to which males would be more exposed, and which results in new vacancies year
226 after year.

227

228 Studies evaluating the effects of predation on reproductive strategies, both in short- and long-
229 lived species, found decreases in breeding propensity or in investment into rearing duties to
230 ultimately reduce predation risk [10,52,53]. However, the Transitions between Breeding
231 Statuses (TBS) we found on M.Clara (with lower probabilities of skipping reproduction than on
232 Vila) may be also influenced by the effect of predation, which could reduce the intra-specific
233 competition for mates and nesting places on M.Clara [54], allowing birds to breed year after
234 year. Consequently, predation would be playing a key role in shaping life-history traits at this
235 locality (*i.e.*, reproductive investment / skipping reproduction). In addition, under high predation
236 pressure, reducing investment in breeding or skipping reproduction might not necessarily result
237 in higher survival probabilities and future opportunities for reproduction, and the fact that the
238 breeding individuals on M.Clara repeat breeding more often than those individuals from Vila
239 may support this hypothesis. Therefore, predation could explain why life-history traits can vary

240 among conspecific populations, and sometimes in a manner that is not expected under Life-
241 History Theory.

242

243 **Funding statement**

244 MCF was supported by a predoctoral grant from the *Universitat de Barcelona*. JB was supported
245 by the *Fundação para a Ciência e a Tecnologia* (PTDC/BIABDE/67286/2006), the *Fundo Europeu*
246 *de Desenvolvimento Regional* (Programme COMPETE, FCOMP-01-0124-FEDER-007061) and the
247 Programme MoniAves (launched by the *Direção Regional do Ambiente* [DRA] of the Azores and
248 coordinated by R.S. Santos). RR was supported by a postdoctoral *Ramón y Cajal* contract of the
249 Spanish *Ministerio de Economía, Industria y Competitividad* (MINECO; RYC-2017-22055). This
250 study was funded by MINECO (CGL2006-01315/BOS, CGL2009-11278/BOS, CGL2013-42585-P,
251 and CGL2016-78530-R) and *Fondos FEDER*, and the Research Executive Agency (European
252 Commission, FP7-PEOPLE-2013-CIG, 618841).

253

254 **Acknowledgements**

255 Fieldwork was conducted with the authorization of the DRA of the Azores and *Dirección General*
256 *de Protección de la Naturaleza del Gobierno de Canarias*. We thank everyone involved in
257 fieldwork. We thank two anonymous referees for constructive feedback.

258

259 **References**

- 260 1. Stearns SC. 1992 *The evolution of life histories*. Oxford University Press.
- 261 2. Lindström J. 1999 Early development and fitness in birds and mammals. *Trends Ecol.*
262 *Evol.* **14**, 343–348.
- 263 3. Metcalfe NB, Monaghan P. 2001 Compensation for a bad start: grow now, pay later?
264 *Trends Ecol. Evol.* **16**, 254–260.
- 265 4. Braendle C, Heyland A, Flatt T. 2001 Integrating mechanistic and evolutionary analysis

- 266 of life history variation. In *Mechanisms of Life History Evolution* (eds T Flatt, A Heyland),
267 pp. 3–10. United States: Oxford University Press Inc., New York.
- 268 5. Owens IPF, Bennett P. 1994 Mortality costs of parental care and sexual dimorphism in
269 birds. *Proc. R. Soc. B.* **257**, 1–8.
- 270 6. Kaitala A, Espadaler X, Lehtonen R. 2000 Ant predation and the cost of egg carrying in
271 the golden egg bug: experiments in the field. *Oikos* **89**, 254–258. (doi:10.1034/j.1600-
272 0706.2000.890205.x)
- 273 7. Rompré G, Robinson WD. 2008 Predation, nest attendance, and long incubation
274 -periods of two Neotropical antbirds. *Ecotropica* **14**, 81–87.
- 275 8. Seigel RA, Huggins MM, Ford NB. 1987 Reduction in locomotor ability as a cost of
276 reproduction in gravid snakes. *Oecologia* **73**, 481–485. (doi:10.1007/BF00379404)
- 277 9. Zanette LY, White AF, Allen MC, Clinchy M. 2011 Perceived Predation Risk Reduces the
278 Number of Offspring Songbirds Produce per Year. *Science (80-.)*. **334**, 1398–1401.
279 (doi:10.1126/science.1210908)
- 280 10. Candolin U. 1998 Reproduction under predation risk and the trade-off between current
281 and future reproduction in the threespine stickleback. *Proc. R. Soc. London. Ser. B Biol.*
282 *Sci.* **265**, 1171–1175. (doi:10.1098/rspb.1998.0415)
- 283 11. Watts HE, Holekamp KE. 2008 Interspecific competition influences reproduction in
284 spotted hyenas. *J. Zool.* **276**, 402–410. (doi:10.1111/j.1469-7998.2008.00506.x)
- 285 12. Ramos JA, Monteiro LR, Sola E, Moniz Z. 1997 Characteristics and Competition for Nest
286 Cavities in Burrowing Procellariiformes. *Condor* **99**, 634–641. (doi:10.2307/1370475)
- 287 13. Williams TD. 2005 Mechanisms Underlying the Costs of Egg Production. *Bioscience* **55**,
288 39.48.
- 289 14. Wiggins DA. 1995 Interspecific Competition for Nest Holes Causes Adult Mortality in the
290 Collared Flycatcher. *Condor* **97**, 445–450. (doi:10.2307/1369030)
- 291 15. Gustafsson L. 2008 Inter- and intraspecific competition for nest holes in a population of
292 the Collared Flycatcher *Ficedula albicollis*. *Ibis (Lond. 1859)*. **130**, 11–16.
293 (doi:10.1111/j.1474-919X.1988.tb00951.x)
- 294 16. Linden M, Møller AP. 1989 Cost of reproduction and covariation of life history traits in
295 birds. *Trends Ecol. Evol.* **4**, 367–371.
- 296 17. Stearns SC, Roff DA. 1992 *The evolution of life histories*. New York: Chapman and Hall.
- 297 18. Roff DA. 1993 *The Evolution Of Life Histories : Theory and Analysis*. Chapman and Hall.
- 298 19. Pugesek BH, Wood P. 1992 Alternate reproductive strategies in the California gull. *Evol.*
299 *Ecol.* **6**, 279–295. (doi:10.1007/BF02270965)
- 300 20. Kendall WL, Converse SJ, Doherty PF, Naughton MB, Anders A, Hines JE, Flint E. 2009
301 Sampling design considerations for demographic studies: a case of colonial seabirds.
302 *Ecol. Appl.* **19**, 55–68. (doi:10.1890/07-1072.1)
- 303 21. Sanz-Aguilar A, Tavecchia G, Genovart M, Igual JM, Oro D, Rouan L, Pradel R. 2011
304 Studying the reproductive skipping behavior in long-lived birds by adding nest
305 inspection to individual-based data. *Ecol. Appl.* **21**, 555–564. (doi:10.1890/09-2339.1)
- 306 22. Pyle P, Nur N, Sydeman WJ, Emslie SD. 1997 Cost of reproduction and the evolution of

- 307 deferred breeding in the western gull. *Behav. Ecol.* **8**, 140–147.
- 308 23. Tavecchia G, Pradel R, Boy V, Johnson AR, Cezilly F, Jan N. 2001 Sex- and Age-Related
309 Variation in Survival and Cost of First Reproduction in Greater Flamingos. *Ecology* **82**,
310 165–174.
- 311 24. Sanz-Aguilar A, Mínguez E, Oro D. 2012 Is laying a large egg expensive? Female-biased
312 cost of first reproduction in a petrel. *Auk* **129**, 510–516. (doi:10.1525/auk.2012.12011)
- 313 25. Brooke M. 2004 *Albatrosses and petrels across the world*. Oxford: Oxford University
314 Press.
- 315 26. Gray CM, Hamer KC. 2001 Food- provisioning behaviour of male and females Manx
316 shearwaters, *Puffinus puffinus*. *Anim. Behav.* **62**, 117–121.
317 (doi:10.1006/anbe.2001.1717)
- 318 27. Peck DR, Congdon BC. 2006 Sex-specific chick provisioning and diving behaviour in the
319 wedge-tailed shearwater *Puffinus pacificus*. *J. Avian Biol.* **37**, 245–251.
320 (doi:10.1111/j.2006.0908-8857.03558.x)
- 321 28. Megyesi JL, O’Daniel DL. 2020 Bulwer’s Petrel (*Bulweria bulwerii*). *Birds of the World*
322 (doi:10.2173/bow.bulpet.01)
- 323 29. Bradshaw CJA, McMahon CR. 2008 Fecundity. In *Encyclopedia of Ecology*, pp. 1535–
324 1543. Elsevier. (doi:10.1016/B978-008045405-4.00645-5)
- 325 30. Bessa-Gomes C, Legendre S, Clobert J. 2004 Allee effects, mating systems and the
326 extinction risk in populations with two sexes. *Ecol. Lett.* **7**, 802–812.
327 (doi:10.1111/j.1461-0248.2004.00632.x)
- 328 31. Jenouvrier S, Caswell H, Barbraud C, Weimerskirch H. 2010 Mating Behavior, Population
329 Growth, and the Operational Sex Ratio: A Periodic Two - Sex Model Approach. *Am. Nat.*
330 **175**, 739–752. (doi:10.1086/652436)
- 331 32. López-Darias M, Rumeu B. 2010 Status and population trend of eleonora’s falcon *Falco*
332 *eleonorae* in the Canary Islands. *Ornis Fenn.* **87**, 35–40.
- 333 33. Bried J, Bourgeois K. 2005 Which future for Bulwer’s petrel in the Azores? *Airo* **15**, 51–
334 55.
- 335 34. Warham J. 1990 *The petrels: their ecology and breeding systems*. London, UK: Academic
336 Press.
- 337 35. Blumstein D., Møller A. 2008 Is sociality associated with high longevity in North
338 American birds? *Biol. Lett.* **4**, 146–148. (doi:10.1098/rsbl.2007.0606)
- 339 36. Flood B, Fisher A. 2011 *Multimedia identification guide to North Atlantic seabirds:*
340 *Storm-petrels & Bulwer’s Petrel*.
- 341 37. Ramirez I, Geraldés P, Meirinho A, Amorim P, Paiva VH. 2008 *Áreas Importantes para as*
342 *Aves Marinhas em Portugal [Important Areas for Seabirds in Portugal]*. Lisboa: Projeto
343 LIFE04NAT/PT/000213 - Sociedade Portuguesa Para o Estudo das Aves.
- 344 38. Cruz-Flores M, Ramos R, Sardá-Serra M, López-Souto S, Militão T, González-Solís J. 2019
345 *Migración y ecología espacial de la población española de petrel de Bulwer*. *Monografía*
346 *nº 4 del programa Migra*. Madrid. (doi:https://doi.org/10.31170/0056)
- 347 39. Fridolfsson A-K, Ellegren H. 1999 A simple and universal method for molecular sexing of

- 348 non-ratite birds. *J. Avian Biol.* **30**, 116–121.
- 349 40. Cruz-Flores M, Pradel R, Bried J, González-Solís J, Ramos R. 2021 Data from: Sex-specific
350 costs of reproduction on survival in a long-lived seabird. DataverseCAT Repository.
351 (doi:doi/10.34810/data27)
- 352 41. Choquet R, Lebreton JD, Gimenez O, Reboulet AM, Pradel R. 2009 U-CARE: Utilities for
353 performing goodness of fit tests and manipulating CAPture-REcapture data. *Ecography*
354 (*Cop.*) **32**, 1071–1074. (doi:10.1111/j.1600-0587.2009.05968.x)
- 355 42. Pradel R. 2005 Multievent: An extension of multistate capture-recapture models to
356 uncertain states. *Biometrics* **61**, 442–447. (doi:10.1111/j.1541-0420.2005.00318.x)
- 357 43. Choquet R, Rouan L, Pradel R. 2009 *Program E-SURGE: A Software Application for*
358 *Fitting Multievent Models*. (doi:10.1007/978-0-387-78151-8)
- 359 44. Burnham KP, Anderson DR. 2002 *Model selection and multimodel inference: a practical*
360 *information-theoretic approach*. 2nd edn. New York: Springer. See
361 <http://linkinghub.elsevier.com/retrieve/pii/S0304380003004526>.
- 362 45. Burnham KP, Anderson DR, Huyvaert KP. 2011 AIC model selection and multimodel
363 inference in behavioral ecology: Some background, observations, and comparisons.
364 *Behav. Ecol. Sociobiol.* **65**, 23–35. (doi:10.1007/s00265-010-1029-6)
- 365 46. Wagenmakers E-J, Farrell S. 2004 AIC model selection using Akaike weights. *Psychon.*
366 *Bull. Rev.* **11**, 192–196. (doi:10.1021/ef300604q)
- 367 47. Robert A, Paiva VH, Bolton M, Jiguet F, Bried J. 2012 The interaction between
368 reproductive cost and individual quality is mediated by oceanic conditions in a long-
369 lived bird. *Ecology* **93**, 1944–1952. (doi:10.1890/11-1840.1)
- 370 48. Kokko H. 1998 Good genes, old age and life-history trade-offs. *Evol. Ecol.* **12**, 739–750.
- 371 49. Catry P, Phillips RA, Phalan B, Croxall JP. 2006 Senescence effects in an extremely long-
372 lived bird: the grey-headed albatross *Thalassarche chrysostoma*. *Proc. Biol. Sci.* **273**,
373 1625–30. (doi:10.1098/rspb.2006.3482)
- 374 50. Genovart M, Negre N, Tavecchia G, Bistuer A, Parpal L, Oro D. 2010 The young, the
375 weak and the sick: Evidence of natural selection by predation. *PLoS One* **5**, 1–5.
376 (doi:10.1371/journal.pone.0009774)
- 377 51. Székely T, Liker A, Freckleton RP, Fichtel C, Kappeler PM. 2014 Sex-biased survival
378 predicts adult sex ratio variation in wild birds. *Proc. R. Soc. B Biol. Sci.* **281**, 20140342.
379 (doi:10.1098/rspb.2014.0342)
- 380 52. Lima SL. 2009 Predators and the breeding bird: behavioral and reproductive flexibility
381 under the risk of predation. *Biol. Rev.* **84**, 485–513. (doi:10.1111/j.1469-
382 185X.2009.00085.x)
- 383 53. Sanz-Aguilar A, Martínez-Abraín A, Tavecchia G, Mínguez E, Oro D. 2009 Evidence-based
384 culling of a facultative predator: efficacy and efficiency components. *Biol. Conserv.* **142**,
385 424–431. (doi:10.1016/j.biocon.2008.11.004)
- 386 54. Cresswell W. 2008 Non-lethal effects of predation in birds. *Ibis (Lond. 1859)*. **150**, 3–17.
387 (doi:10.1111/j.1474-919X.2007.00793.x)

389

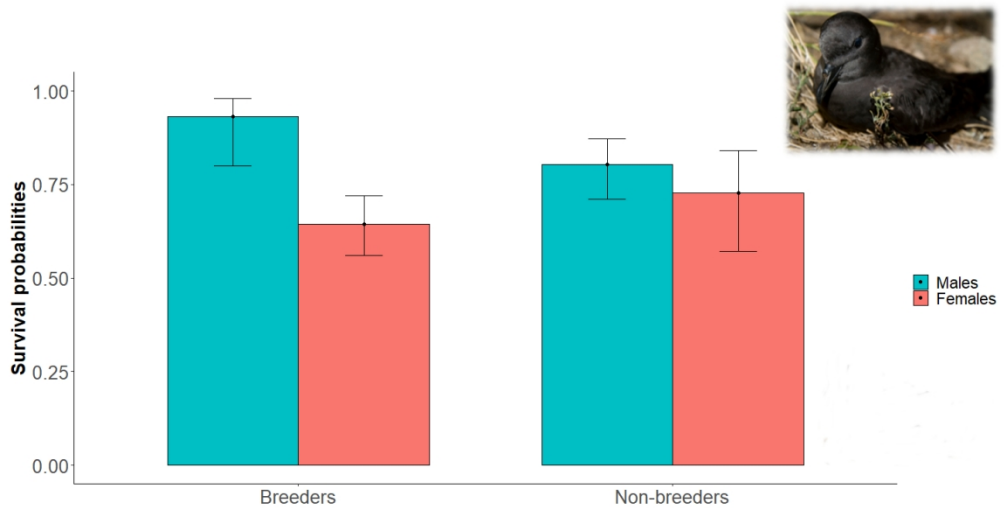
390 **Table 1.** Modelling Initial States (IS), survival (Φ), Transitions between Breeding Statuses (TBS) and capture (p) probabilities, for Bulwer's petrel on Vila and
 391 M.Clara, depending on time (years), sex (without differentiating breeding statuses), breeding status (breedStatus: breeders and non-breeders, without
 392 distinguishing sexes), each combination of breeding status and sex (breedStatus/sex) or colony (colony; while ~~colony~~ means no differences between colonies).
 393 Within each group of models (A-D), we provide QAICc values in ascending order. For groups B-E, we started to model from the best model obtained in the
 394 previous group.

					np	Dev	QAICc	Δ QAICc	Weight
A) Modelling Initial States probabilities (IS):									
1.	IS(breedStatus/sex*time)	Φ (c)	TBS(colony)	p (colony*time)	67	5,883.6	6,024.2	42.4	0.000
2.	IS(breedStatus/sex)	Φ (c)	TBS(colony)	p (colony*time)	45	5,945.5	6,038.4	56.5	0.000
B) Modelling survival probabilities (Φ):									
3.	IS(breedStatus/sex*time)	Φ (breedStatus/sex)	TBS(colony)	p (colony*time)	70	5,860.6	6,007.8	25.9	0.000
4.	IS(breedStatus/sex*time)	Φ (sex)	TBS(colony)	p (colony*time)	68	5,865.4	6,008.2	26.3	0.000
5.	IS(breedStatus/sex*time)	Φ (breedStatus/sex*colony)	TBS(colony)	p (colony*time)	74	5,855.7	6,011.8	29.9	0.000
6.	IS(breedStatus/sex*time)	Φ (breedStatus)	TBS(colony)	p (colony*time)	68	5,877.5	6,020.3	38.4	0.000
7.	IS(breedStatus/sex*time)	Φ (colony)	TBS(colony)	p (colony*time)	68	5,881.4	6,024.2	42.3	0.000
C) Modelling Transitions between Breeding Statuses probabilities (TBS):									
8.	IS(breedStatus/sex*time)	Φ (breedStatus/sex)	TBS(colony)	p (colony*time)	66	5,867.0	6,005.4	23.5	0.000
D) Modelling capture probabilities (p):									
9.	IS(breedStatus/sex*time)	Φ (breedStatus/sex)	TBS(colony)	p (colony*time) (breedStatus)	68	5,839.1	5,981.9	0.0	0.681
E) Neighboring models of model 9:									
10.	IS(breedStatus/sex*time)	Φ (breedStatus/sex)	TBS(colony)	p (colony*time) (breedStatus)	72	5,832.0	5,983.6	1.7	0.285
11.	IS(breedStatus/sex*time)	Φ (breedStatus/sex*colony)	TBS(colony)	p (colony*time) (breedStatus)	72	5,836.6	5,988.2	6.4	0.028
12.	IS(breedStatus/sex*time)	Φ (sex)	TBS(colony)	p (colony*time) (breedStatus)	66	5,853.1	5,991.5	9.7	0.005
13.	IS(breedStatus/sex*time)	Φ (breedStatus)	TBS(colony)	p (colony*time) (breedStatus)	66	5,860.6	5,999.0	17.1	0.000

395 (c) is constant; 'np' number of estimated parameters; 'Dev' deviance; 'QAICc' quasi-likelihood Akaike information criterion values for overdispersion and small samples; ' Δ QAICc' difference
396 between a specific model and the model with the lowest QAICc value (in bold). Weight is the probability that a model is the expected best model based on QAICc.

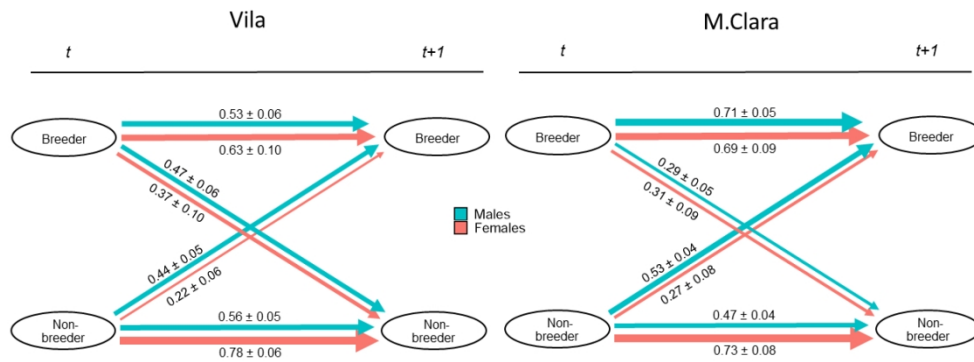
397 **Figure 1.** Adult survival probabilities (\pm 95% CI) of breeding and non-breeding Bulwer's petrels
398 obtained from the best-supported model (model 10, table 1). Photo credit: Raúl Ramos.

399 **Figure 2.** Transition probabilities among breeding statuses (\pm SE) for Vila and M.Clara obtained
400 from the best-supported model (model 10, table 1). Arrow width is proportional to probability
401 estimates.



Adult survival probabilities (\pm 95% CI) of breeding and non-breeding Bulwer's petrels obtained from the best-supported model (model 10, table 1). Photo credit: Raúl Ramos.

508x302mm (72 x 72 DPI)



Transition probabilities among breeding statuses (\pm SE) for Vila and M.Clara obtained from the best-supported model (model 10, table 1). Arrow width is proportional to probability estimates.

338x190mm (96 x 96 DPI)