



HAL
open science

Siloxane removal for biogas purification by low cost mineral adsorbent

S. Pioquinto García, L.Á. Garza Rodríguez, D. Bustos Martínez, F.D.J. Cerino Córdova, E. Soto Regalado, Sylvain Giraudet, N.E. Dávila Guzmán

► To cite this version:

S. Pioquinto García, L.Á. Garza Rodríguez, D. Bustos Martínez, F.D.J. Cerino Córdova, E. Soto Regalado, et al.. Siloxane removal for biogas purification by low cost mineral adsorbent. *Journal of Cleaner Production*, 2021, 286, pp.124940. 10.1016/j.jclepro.2020.124940 . hal-03130413

HAL Id: hal-03130413

<https://hal.science/hal-03130413>

Submitted on 18 Feb 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

Siloxane removal for biogas purification by low cost mineral adsorbent

Sandra Pioquinto García^a, Luis Ángel Garza Rodríguez^c, Diana Bustos Martínez^a, Felipe de Jesús Cerino Córdova^{a,1}, Eduardo Soto Regalado^a, Sylvain Giraudet^{b*}, Nancy Elizabeth Dávila Guzmán^{a*}

^a*Universidad Autónoma de Nuevo León, UANL, Facultad de Ciencias Químicas.*

Av. Universidad S/N, Cd. Universitaria, San Nicolás de los Garza, Nuevo León, CP 66455, México.

^b*École Nationale Supérieure de Chimie de Rennes, UMR CNRS 6226, 11 Allée de Beaulieu, 35708 Rennes, Francia*

^c*Instituto Tecnológico de Estudios Superiores de Monterrey, Departamento de Química y Nanotecnología, Ave. Eugenio Garza Sada 2501 Sur, Monterrey, Nuevo León, CP 64849, México*

^{a*} **Corresponding author:** Nancy Elizabeth Dávila Guzmán

Phone number: 52-8183294000, ext. 3475

Fax number: 52-8183294000, ext. 6282

E-mail: nancy.davilagz@uanl.edu.mx

^{b*} **Corresponding author:** Sylvain Giraudet

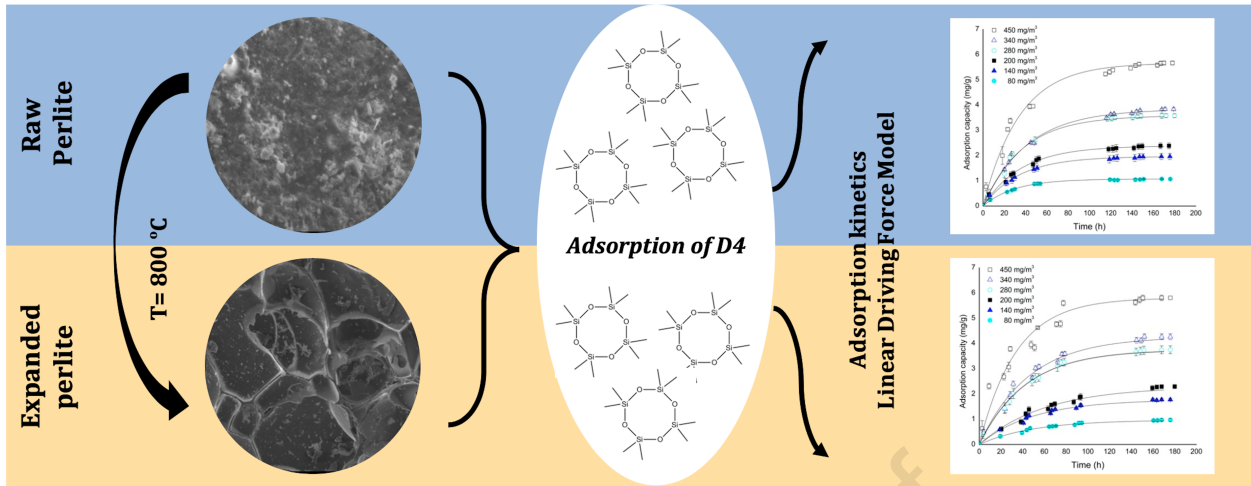
Phone number: +33 (0)2 23 23 80 15

E-mail: sylvain.giraudet@ensc-rennes.fr

¹ Current address for Felipe Cerino: Universidad Autónoma de Nuevo León, UANL, Facultad de Ingeniería Mecánica y Eléctrica. Av. Universidad S/N, Cd. Universitaria, San Nicolás de los Garza, Nuevo León, CP 66455, México.

Sandra Pioquinto-Garcia: Investigation, Writing- Original draft, Formal analysis; **Nancy Davila-Guzman:** Conceptualization, Methodology, Writing- Review & Editing, Project administration, Funding acquisition; **Luis Garza-Rodriguez:** Conceptualization, Funding acquisition; **Sylvain Giraudet:** Methodology, Formal Analysis, Supervision; **Diana Bustos-Martinez:** Validation, Resources; **Felipe Cerino-Cordova:** Writing- Review & Editing, **Eduardo Soto-Regalado:** Funding acquisition, Methodology.

Journal Pre-proof



Journal Pre-proof

ABSTRACT

The potential use of raw and expanded perlite as low-cost adsorbents for biogas purification has been investigated. The thermal expansion of perlite causes a reduction in the density of silanol groups from 2515.43 to 653.75 OH/nm²; in contrast, the specific surface area of perlite increased two-fold due to the thermal expansion. To determine the equilibrium adsorption capacity and the adsorption kinetics batch experiments were conducted. The adsorption capacities are in the following order: activated carbon (6.8 mg/g) > silica gel (6.6 mg/g) > expanded perlite (5.81 mg/g) > raw perlite (5.6 mg/g) when compared at the same experimental conditions. The equilibrium adsorption data showed that perlite can be used to reduce the octamethylcyclotetrasiloxane concentration below 28 mg/m³, as recommended by leading manufacturers. The adsorption kinetics of octamethylcyclotetrasiloxane onto raw and expanded perlite followed the Linear-Driving Force model suggesting that the mass transfer is the rate-controlling step. In addition to its low cost, expanded perlite has the advantage of requiring lower desorption temperature (200 °C) for regeneration in comparison to the reported values for activated carbon (> 400 °C) and fast desorption kinetics (20 min), which could contribute to a cleaner production of biogas.

Keywords: Octamethylcyclotetrasiloxane; biogas; perlite; adsorption; isotherm; kinetics

Abbreviations	
a	Elovich parameter

b	Elovich parameter
AC	activated carbon
Co	initial concentration
C _e	concentration at equilibrium
D4	octamethylcyclotetrasiloxane
d	diameter of adsorbent particle
D _s	effective diffusion coefficient
EP	expanded perlite
K _L	effective mass transfer coefficient
K _R	Ritchie kinetic rate constant
LDF	linear driving force
\dot{m}_{ads}	adsorbent mass per hour
p _{ads}	price of the adsorbent
q	adsorption capacity at time t
q _e	adsorption capacity at equilibrium
Q _{bg}	biogas flow rate
R ²	coefficient of determination
RP	raw perlite
SE	standard error
SG	silica gel
t ₀	Elovich constant
WWTP	wastewater treatment plant
α	effective volume ratio

ϵ_{bg}	amount of electricity generated from biogas
-----------------	---

1. Introduction

The United Nations, in 2015, proposed promoting the use of renewable energy as a priority of the 2030 Agenda for Sustainable Development (United Nations, 2015), since there is great potential to use clean and renewable energy in the world, especially in the use of biomass for the production of biofuels such as bioethanol, biodiesel, and biogas (Tabatabaei et al., 2020). Biogas is generated from organic matter in an anaerobic environment (Santos et al., 2018). In the world, the United States (US) and Europe are the main producers of biogas (Scarlat et al., 2018). According to the American Biogas Council, the US has a potential biogas production from wastewater treatment plants (WWTP) and landfills. WWTP could generate 2.97 billion Nm^3/year of biogas to make 5.6 billion kWh of electricity/year. While the biogas from landfills could amount to 7.47 billion Nm^3/year to produce 22.5 billion kWh of electricity/year (American Biogas Council, 2014). In Europe, Germany is positioned as the leader in the production of biogas, with 94 biogas plants and 120 biogas upgrading plants (including landfill plants) and upgrading capacity of 1.79 billion Nm^3/year (Ullah Khan et al., 2017). The Mexican biogas production from WWTP is estimated to be 0.14 billion Nm^3/year , which could produce 0.332 billion kWh of electricity/year (IMTA, 2016). The landfill gas potential is estimated to be 4.97 billion Nm^3/year , considering 24.91 million tons of waste in landfills (REMBIO, 2012).

The biogas composition is highly dependent on the origin and the production conditions such as microorganisms, temperature, and impurities (Chaemchuen et al., 2016, Tabatabaei, et al., 2020). Biogas is generally a mixture of 50-70 v% methane, 30-50 v% carbon dioxide, and <6 v% residual compounds such as water vapor, hydrogen sulfide, and siloxanes (Poloncarzova et al., 2011). Hydrogen sulfide and siloxanes must be removed from the biogas to increase the calorific value and avoid damage to the combustion system (Álvarez-Flórez and Egusquiza, 2015, Kapoor et al., 2020). Removal of hydrogen sulfide can be achieved by biological treatments that are inexpensive and environmentally friendly (more extensive review on removal of hydrogen sulfide by biological treatment can be found in Khoshnevisan et al., 2017). On the other hand, the siloxanes are undesirable in a gas combustion system due to the formation of silicon dioxide (Álvarez-Flórez and Egusquiza, 2015). The silicon dioxide solid deposits damage parts of the combustion system such as pistons, cylinder heads, and valves (Matsui and Imamura, 2010) causing low efficiency and high repair costs (Ajhar et al., 2010). In WWTP and landfills, typical concentrations of siloxanes in biogas are between 1-400 mg/Nm³ (Dewil et al., 2006), where octamethylcyclotetrasiloxane (D4) is the common siloxane compound found (Tran et al., 2019a). Although, there is no legislation to regulate the amount of siloxanes to avoid damages in the combustion systems, some recommendations had been made for limiting the concentration of siloxanes in biogas from 0.3 to 28.0 mg/m³ (Gaj, 2017).

There are two types of methods for eliminating siloxanes in biogas: pretreatment methods and post-purification methods. The first might include direct sludge purging processes or heat treatment before anaerobic digestion. The second type, which is the

primary method for commercial applications, includes technologies such as adsorption, absorption, cryogenics, filtration by membrane, catalysis, and biological processes (Wang et al., 2019). Several publications detail the techniques for eliminating siloxanes by post-purification methods in which the efficiencies, advantages, and disadvantages are indicated (Gong et al., 2015; Ruiling et al., 2017; Shen et al., 2018). In general, the use of some post-purification techniques could reach removal efficiencies greater than 90 %, to achieve such high efficiency implies that the cost of investment and operation would also be high.

The most common biogas purification technology is adsorption, the use of alumina, silica gel, activated carbon, zeolites, and polymeric adsorbents have been reported for removal of siloxanes. Note that the activated carbon is included in all comparisons because this adsorbent is the most used in the biogas purification process (Cabrera-Codony et al., 2018). The adsorption process with activated carbon still has some drawbacks associated with the high regeneration temperatures (400–1000 °C), incomplete desorption (Giraudet et al., 2014) and, consequently, a high cost (Santos-Clotas et al., 2019). This study aims to evaluate a low-cost mineral as a new material for siloxane removal, which has a silicon dioxide composition similar to silica gel, approximately 80 % (Alkan et al., 2005), and it is three times cheaper than activated carbon.

2. Materials and Methods

2.1 Adsorbent and characterization

Raw Perlite (RP) and Expanded Perlite (EP) were provided by TERMOLITA (Mexico), PICTACTIF NC 60 activated carbon was supplied by PICA Corp. (France), and silica gel was supplied by PROLABO (France). The materials were sieved using a mesh No. 20, and the moisture was removed at 108 °C for 24 hours in a Memmert® oven.

The analysis of Fourier transform infrared (FTIR), X-ray diffraction (XRD), and thermogravimetry, required RP and EP as powders. Thus, a Rock Lab Benchmill 50 vibratory mill was used to grind them. For the adsorption tests, nitrogen with purity > 99.995 %, hydrogen with purity > 99.999 %, and synthetic air (O₂ 20 % ± 2%) were supplied by the company Air Liquide (France).

The crystalline structure of the RP and EP was evaluated via an X-ray powder diffractometer Siemens brand model E04-0012. The functional groups on the surface of RP and EP were identified using an FTIR analysis performed in an Agilent Technologies Cary 630 FTIR device, from 4000 to 400 cm⁻¹ at a resolution of 16 cm⁻¹ triangular apodization type, Mertz phase-type, with a KBr disk being used as a background. The morphology of RP and EP was studied by using the images obtained by scanning electron microscope (SEM) with a JSM 6301F (JEOL®) equipment for classical images, and a JSM 6400 (JEOL®) coupled to a power dispersion spectrometer (EDS) for the elemental analysis of the surface. The specific surface area of RP, EP, and silica gel were obtained by analyzing the N₂ adsorption-desorption at 77 K carried out in an Autosorb-1-MP Quantachrome Instruments® (from ENSCR, France) and by using the BET equation.

The thermogravimetric analysis (TGA) of RP and EP was carried out to determine the thermal stability of the adsorbents; a TA instruments Q500 thermogravimetric analyzer

was used with a temperature range from 50-850 °C, a heating rate of 10 °C/min, and N₂ atmosphere. The density of the silanol groups was estimated from the thermal analysis following the procedure reported elsewhere (de Farias and Airoidi, 1998). Briefly, the mass of silanol groups was obtained from the mass loss detected from the thermogravimetric curve. This mass loss is due to the condensation of two silanol groups. Then, the density of silanol groups is calculated by dividing the amount of silanol groups by the perlite surface area.

2.2 Adsorption experiments.

The adsorption tests were carried out at six siloxane D4 initial concentrations. First, 20 µL of D4 (analytical grade, liquid, 98% purity, Sigma-Aldrich) were allowed to volatilize in a 2 L glass flask. Then, different amounts of D4 vapor (18-104 mL) were taken to be injected into a second glass flask (2 L), previously flushed with dry air, to obtain the desired concentration of D4, ranging from 80 to 450 mg/m³. The adsorption tests were carried out in 2 L glass flasks, placing 0.154 g of adsorbent material into a steel basket inside the flask, the particle size used in all experiments was 1 mm, constant temperature (25 °C), pressure of 1 atm, and a D4 initial concentration from 80 to 450 mg/m³ (Figure S1). Each flask was stirred at 700 min⁻¹ for 170 h to ensure the equilibrium was reached. During adsorption experiments, samples were taken for further analysis of concentration by using gas chromatography. Each sample was injected into an HP 6890 Series II gas chromatography system with a flame ionization detector (FID) to measure the concentration of siloxane D4. The column used was an HP-624 Special Analysis capillary column (30.0 m x 0.25 mm inner diameter x 1.40 µm

film thickness). The inlet temperature was set at 150 °C, oven at 145 °C, and detector at 250 °C. The N₂ was used as carrier gas at a flow rate of 2.3 mL/min and 17.71 psi. The injected volume was 0.5 mL with splitless, and the duration of each analysis was 3.4 min. The adsorption capacity was obtained by the mass balance, using Eq. (1):

$$q = \frac{(C_0 - C)V}{m} \quad \text{Eq. (1)}$$

Where q is the adsorption capacity (mg/g), C_0 is the D4 initial concentration (mg/m³), C is the D4 concentration at time t (mg/m³), V is the volume (m³), and m is the adsorbent mass (g). Batch adsorption experiments were carried out by triplicate (coefficient of variation < 7%) and the adsorption capacity was obtained by the average value of the three tests.

2.3 Adsorbent Regeneration Studies

The regeneration of EP was studied by carrying out an adsorption experiment with 7.0 g of EP at an initial concentration of 450 mg/m³ and 25 °C, as described in Sec. 2.2. When the equilibrium was reached, the flask was insulated and placed on a heating plate at 200 °C for the desorption of D4 for 10 h. Several samples were taken for further concentration analysis on GC to obtain the desorption kinetics.

2.4 Mathematical models

For dilute systems, the Henry isotherm model is used (Eq. 2), where there is a linear relationship between the equilibrium concentration and the equilibrium adsorption capacity (Cooney, 1998).

$$q_e = KC_e \quad \text{Eq. (2)}$$

where K is the Henry constant (m^3/g), C_e and q_e are the D4 concentration (mg/m^3) and adsorption capacity (mg/g) at equilibrium, respectively.

For the adsorption kinetics, four models were used to predict the evolution of the adsorption capacity through time. The Elovich kinetic model (Eq. 3) was developed for gas-solid adsorption systems (Russo et al., 2017) and establishes that the active sites of the adsorbents are heterogeneous and the mass transfer resistance to be negligible.

$$q = \frac{1}{b} \ln(t + t_o) - \frac{1}{b} \ln(t_o) \quad \text{Eq. (3)}$$

Where q is the adsorbed amount at t (mg/g), b is the Elovich constant (g/mg), $t_o = 1/ab$, a is the initial adsorption rate ($\text{mg}/\text{g} \cdot \text{h}$) and t is the time (h).

The Ritchie kinetic model (Eq. 4) considers adsorption at specific sites on the surface of a solid (Ritchie, 1977). The rate of adsorption depends on the fraction of sites that are unoccupied at time t .

$$\frac{q_e^{n-1}}{(q_e - q)^{n-1}} = (n - 1)k_R t + 1 \quad \text{Eq. (4)}$$

where q_e and q are the adsorption capacity at equilibrium (mg/g) and the adsorption capacity (mg/g) at time t , respectively. n is the number of sites on the surface occupied by an adsorbed gas and is equal to two for second-order reaction kinetics (Altindal et al., 2014), k_R is the kinetic rate constant ($1/\text{h}$), and t is the time (h).

The mass transfer models consider the external mass transfer resistance or the intraparticle diffusion as the rate-controlling steps. Despite this physical mechanisms, among the mass transfer models most reported in the literature to describe the adsorption kinetics of gases and liquids in solid materials (such as carbon dioxide, nitrous oxide, and acetylene), are the linear driving force (LDF) model (Yao and Tien,

1992) and the Crank model (Crank, 1975), which consider a single adsorption rate. This rate of adsorption lumps together the external and internal mass transfers.

The LDF model (Eq. 5) assumes that the adsorption rate is proportional to the difference between the average concentration of adsorbate at the gas-solid interface $\bar{C}^*(t)$ and the average adsorbate concentration in the adsorbent particle $\bar{C}(t)$ (Sircar and Hufton, 2000):

$$\frac{d\bar{C}(t)}{dt} = k_L[\bar{C}^*(t) - \bar{C}(t)] \quad \text{Eq. (5)}$$

The analytical solution of the LDF model is shown in Eq. 6:

$$\frac{q}{q_e} = 1 - e^{-k_L t} \quad \text{Eq. (6)}$$

Where q is the adsorption capacity at time t (mg/g), q_e is the adsorption capacity at equilibrium (mg/g), k_L is the effective mass transfer coefficient (1/h), and t is the time (h). Additionally, the Crank model (Eq. 7) represents the analytical solution of the equation of the Fick's second law, when the external mass transfer resistance is negligible. This model considers that the rate-limiting step is the intraparticle mass transfer, free-solute adsorbent particle, well-stirred solution of limited volume, and uniform solute concentration. In Eq. (7), the total amount of solute adsorbed after time t is related to the total amount of solute adsorbed after infinite time (i.e. at equilibrium) as follows:

$$\frac{q}{q_e} = 1 - \sum_{n=1}^{\infty} \frac{6\alpha(\alpha+1)\exp(-D_s q_n^2 t/d^2)}{9+9\alpha+q_n^2 \alpha^2} \quad \text{Eq. (7)}$$

where q is the adsorption capacity at time t (mg/g), q_e is the adsorption capacity at equilibrium (mg/g), d is the diameter of the particle (m), D_s is the effective diffusion coefficient (m²/h), and t is the time (h). α is the effective volume ratio and is expressed

in terms of the final uptake of solute by Eq. (8), q_n represent the non-zero solutions of Eq. (9):

$$\tan q_n = \frac{3q_n}{3 + \alpha q_n^2} \quad \text{Eq. (8)}$$

$$\frac{q_e}{VC_o} = \frac{1}{1 + \alpha} \quad \text{Eq. (9)}$$

The parameters of the isotherm and kinetics models were obtained by non-linear regression, the Solver® tool of Microsoft Excel was used to minimize the standard error (Eq. 10), and the coefficient of determination (R^2) was also calculated by Eq. 11.

$$SE = \sqrt{\frac{(q_{e,exp} - q_{e,cal})^2}{n - (k + 1)}} \quad \text{Eq. (10)}$$

$$R^2 = \frac{\Sigma(q_{e,cal} - \bar{q}_{e,exp})^2}{\Sigma(q_{e,cal} - \bar{q}_{e,exp})^2 + \Sigma(q_{e,cal} - q_{e,exp})^2} \quad \text{Eq. (11)}$$

Where $q_{e,cal}$ is the adsorption capacity at equilibrium calculated by the kinetic model (mg/g), $q_{e,exp}$ is the adsorption capacity at equilibrium from experimental data (mg/g), $\bar{q}_{e,exp}$ is the average adsorption capacity at equilibrium obtained from the experimental data (mg/g), n is the number of experimental data, and k is the number of the model parameters.

2.5 Removal cost analysis

A simple calculation of the cost associated with the EP and activated carbon usage for the removal of siloxane was performed. First, the adsorbent mass per hour (\dot{m}_{ads}) was calculated by considering the adsorption capacity (q) of each adsorbent, the siloxane concentration (C), and the biogas flow rate of 1 m³/h (Q_{bg}), as shown in Eq. (12). Then, the cost of the usage of EP and activated carbon was obtained from Eq. (13).

$$\dot{m}_{ads} = (Q_{bg})(C)/q \quad \text{Eq. (12)}$$

$$cost = (\dot{m}_{ads})(p_{ads})/\varepsilon_{bg} \quad \text{Eq. (13)}$$

where p_{ads} is the price of the adsorbent, and ε_{bg} is the amount of electricity generated from biogas.

3. Results and discussion

3.1 Perlite characterization

The examination of the crystalline structure of RP and EP were determined by X-ray diffraction analysis (Figure 1). A broad peak was observed, centered at 25 degrees (2θ), related to the absence of any ordered crystalline structure. Thus, this material could be considered mainly amorphous (Celik et al., 2013). The surface functional groups were determined by FTIR analysis (Figure 2). A broad band around 3500 cm^{-1} (A) was identified, and it is attributed to the stretching mode of O-H groups. In comparison, the band at 1630 cm^{-1} (B) is characteristic of the bending mode of OH groups of Si-OH and water molecules adsorbed on the perlite surface. The band at 1386 cm^{-1} (C) corresponds to the C-H deformation of CH_2 and CH_3 groups (Liu and Huang, 2002). On the other hand, lower intensity bands were observed at $1200\text{-}1000 \text{ cm}^{-1}$ (D) and 800 cm^{-1} (E) related to the Si-O vibrations of asymmetric Si-O-Si stretching and the Si-O vibrations of symmetric stretching of Si-O-Al.

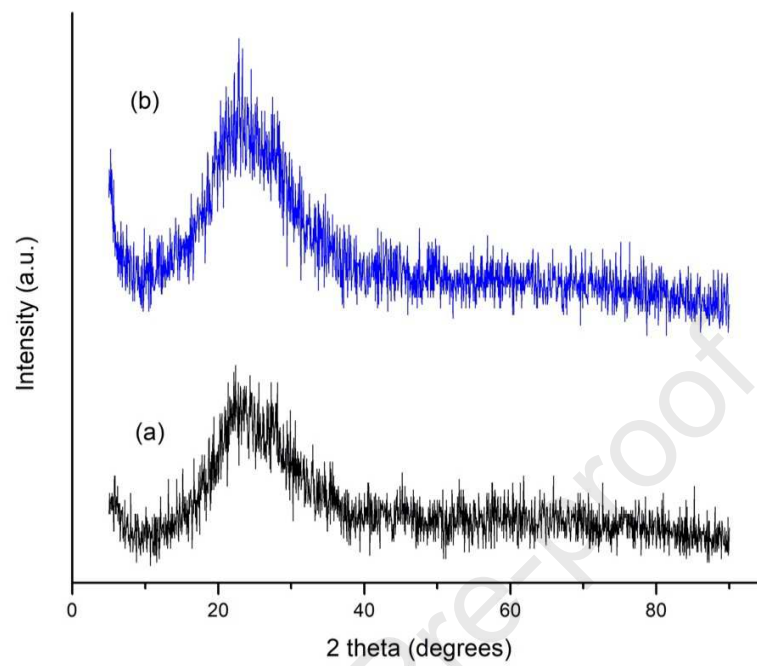


Figure 1. X-ray diffraction patterns of a) raw perlite and b) expanded perlite.

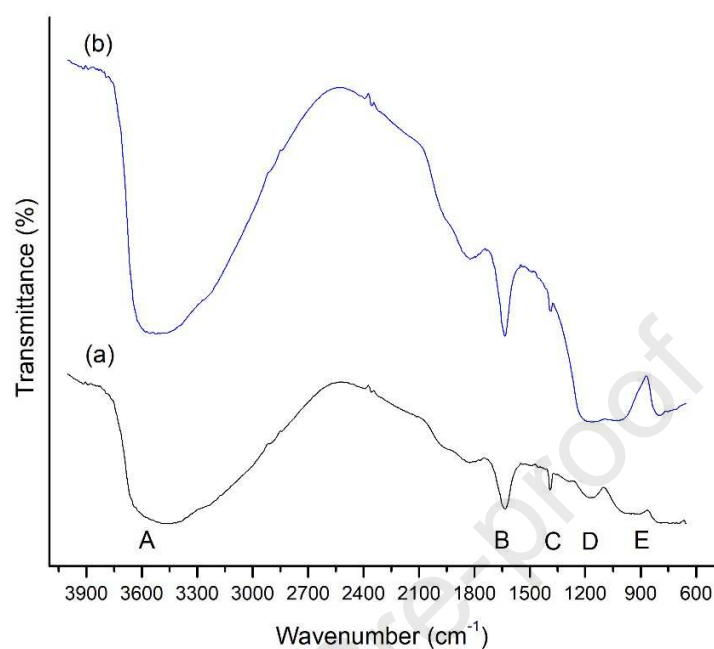


Figure 2. Fourier transform infrared spectra of the (a) raw perlite and (b) expanded perlite.

According to Figure 2, the intensity of the absorption bands A, B, C, and E on the EP surface was more significant than that shown by RP. This behavior could be attributed to the increase in the exposure of the silanol functional groups after the thermal expansion process of RP. Silanol groups have been described as adsorption sites for polar molecules, such as D4, which suggests that both adsorbent materials may exhibit a potential adsorption capacity of D4 (Sigot et al., 2014).

Also, SEM was used to analyze the morphology and texture of RP and EP (Figure 3). In general, RP and EP have a rough outer surface and a white vitreous luster due to its amorphous silica content. Moreover, the RP micrograph (Figure 3 left) reveals a rough surface with slits of a few micrometers in diameter ($< 10\mu\text{m}$). On the contrary, EP

exhibits shallow grooves about 20-100 μm formed by smooth leaves (Figure 3 right) due to the high temperatures exposition (800-1000 $^{\circ}\text{C}$) during the thermal expansion process.

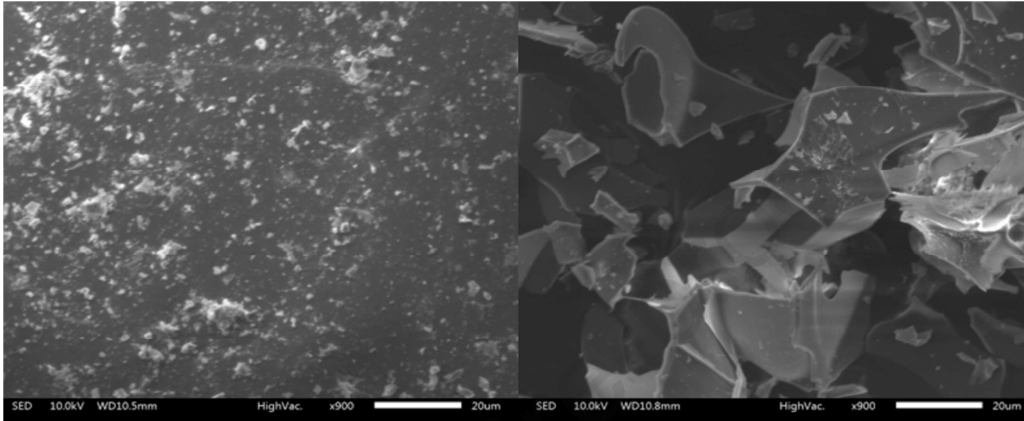


Figure 3. Scanning electron microscopy images of raw perlite (left) and expanded perlite (right).

SEM/EDS was used to investigate the chemical elements on the surface of RP and EP. The most abundant chemical elements in both materials are oxygen (O), silicon (Si) and aluminum (Al), in the form of silicon dioxide and aluminum dioxide (see Table 1). Additionally, TGA was used to determine the thermal stability of RP and EP (Figure 4). The TGA curve of RP showed a continuous decrease in weight in the range of (50-800) $^{\circ}\text{C}$. The weight loss over the range temperature from 50 to 120 $^{\circ}\text{C}$ is due to the elimination of the water absorbed on the surface (loss of 0.11 %) and the weight loss observed up to 800 $^{\circ}\text{C}$ (approximately 3.43 %) can be attributed to the dehydroxylation of the silanol groups (Rouliat et al., 2006).

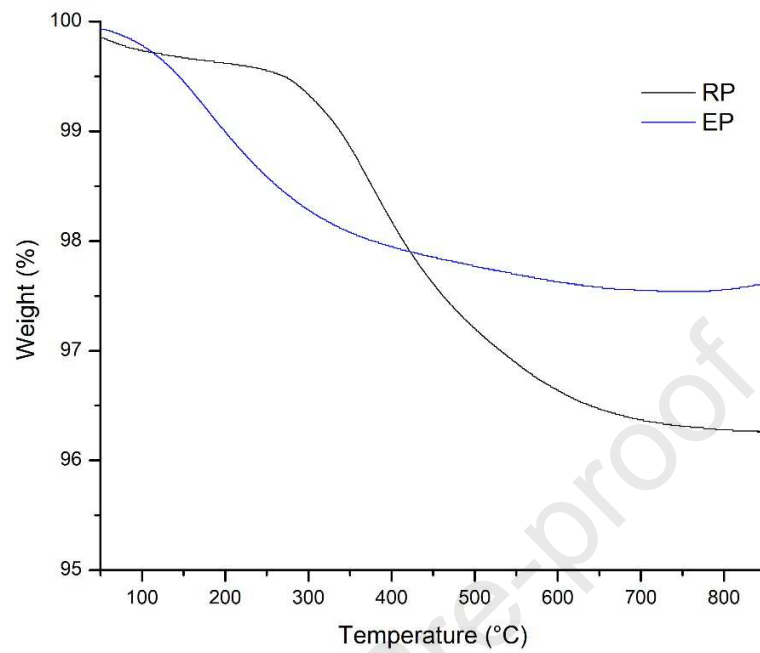


Figure 4. Thermogravimetric analysis curve for (a) raw perlite and (b) expanded perlite.

Table 1. Relative elementary composition of raw perlite and expanded perlite.

Element	Raw perlite (% weight)	Expanded perlite (% weight)
O	56.84	55.98
Mg	0.56	0.00
Al	5.18	5.64
Si	23.17	31.81
K	3.02	4.16
Ca	0.22	0.00
Ti	0.83	0.00
Fe	13.91	0.00
Na	0.87	2.41

The density of the silanol groups was calculated to quantify the dehydroxylation of EP after been subjected to thermal expansion. The weight loss of RP and EP was associated to the condensation of the silanol groups distributed on the surface of the adsorbents (de Farias and Airoidi, 1998). The density of the silanol group was 2515.43 OH/nm² and 653.75 OH/nm² for RP and EP, respectively, showing that the dehydroxylation process decreases 3.85 folds the number of silanol groups. Due to the higher amount of silanol groups on the surface of the RP, a greater siloxane adsorption capacity in this material would be expected when compared with the adsorption capacity of the EP.

The specific surface area was determined by N₂ adsorption-desorption on RP and EP (Figure S2). The adsorption isotherms of RP and EP are type IV (IUPAC classification), which are characteristic of mesoporous materials. The specific surface obtained by the BET method for EP was 2.12 m²/g, which is twice the value of RP (1.2 m²/g). Again, the increase in the surface area is due to the expansion process of RP when subjected to high temperatures (above 700 °C).

The mean pore size of RP and EP is 47.13 Å and 43.42 Å (Figure S3), respectively, whereas the D4 molecule size is approximately 10 Å. Thus, no resistance to intraparticle diffusion is expected (Jiang et al., 2016).

3.2 Adsorption Equilibrium

The adsorption capacity at equilibrium was determined as a function of the D4 equilibrium concentration, the adsorption isotherms of RP and EP were carried out with

a fixed amount of 0.154 g of adsorbent, temperature of 25 °C, 1 atm, and an initial concentration of D4 from 80 to 450 mg/m³.

This study was carried out at low D4 concentration, similar to the values found in biogas streams at WWTP and landfills (Bak et al., 2019; Tran et al., 2019b). At low concentration, the adsorption isotherms of D4 exhibited a linear behavior, which was described by Henry's Law model (Figure 5). The values of the Henry constant for D4 adsorption on RP and EP were 0.1023 and 0.1359 m³ D4/g, respectively. The R² values of Henry model for RP and EP were 0.980 and 0.972, respectively. These results could be an indication of a stronger affinity between EP and D4 than to RP since the Henry constant is a measure of the interaction between a molecule and the functional groups in the adsorbent solid.

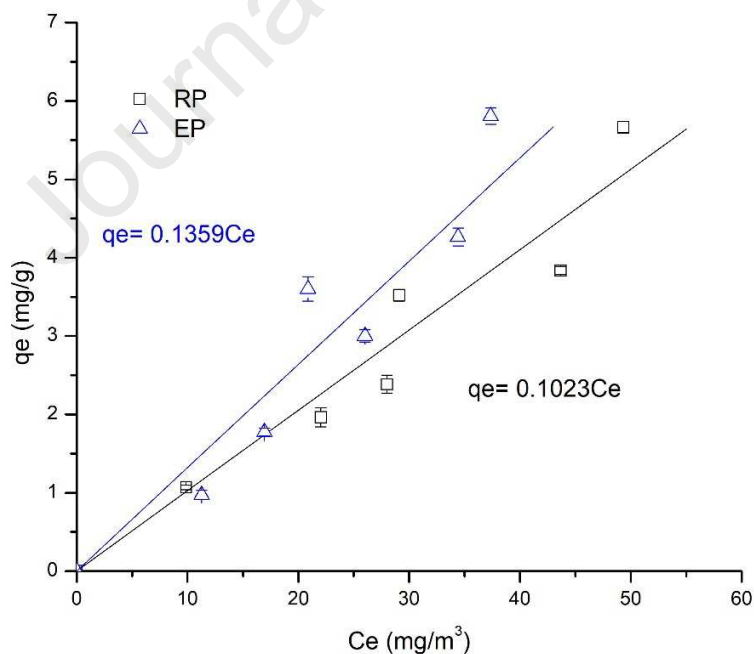


Figure 5. Adsorption isotherm of D4 on raw perlite (\square) and expanded perlite (\circ). The solid lines represent the Henry Isotherm model. Error bars represent standard deviation of triplicate measurements.

In contrast, most studies of the siloxanes dynamic adsorption experiments have been carried out at higher siloxanes concentrations than those commonly found in real biogas samples. This is to shorten the contact time in the adsorption experiments (Cabrera-Codony et al., 2018; Tran et al., 2019a). The same applies to the few experiments performed in batch systems (Nam et al., 2013), which prevents comparison between the adsorption capacity of perlite with other materials already reported (see Table 2).

Table 2. Equilibrium adsorption capacities of several adsorbents reported for the removal of siloxane D4.

System Configuration	Adsorbent	Siloxane concentration (mg/m ³)	Equilibrium adsorption capacity (mg/g)	Reference
Dynamic	Activated carbon	559*	404	(Oshita et al., 2010)
Batch	Activated carbon	7140	90	(Nam et al., 2013)
	Silica gel		56	
	Alumina oxide		34	
Dynamic	Activated carbon	400	52	(Sigot et al., 2014)
	13X Zeolite		113	
	Silica gel		216	
Dynamic	Activated carbon	1540	319.9	(Cabrera-Codony et al., 2018)
Batch	Activated carbon	164000*	577	(Santos-Clotas et

				al., 2019)
Bench-scale	Iron oxide	1822*	1.86	(Bak et al., 2019)
	Iron oxide hydroxide		1.33	
	Activated carbon		5.35	
	Silica gel		1055.32	
	Zeolite		1.73	
Dynamic	Activated carbon	12145	273	(Tran et al., 2019a)
Dynamic	Activated carbon	12000	269	(Tran et al., 2019b)
Dynamic	Silica gel	83820	236.2	(Meng et al., 2020)
	Modified silica gel		367.1	
Batch	Raw perlite	450	5.6	This work
	Expanded perlite		5.8	
	Activated carbon		6.8	
	Silica gel		6.6	

*Calculated at 1atm and 25 °C.

Accordingly, the removal of D4 by silica gel and activated carbon were performed under the same conditions previously described in Section 2.2. The adsorption capacities of siloxane D4 followed this order: activated carbon (6.8 mg/g) > silica gel (6.6 mg/g) > expanded perlite (5.8 mg/g) > raw perlite (5.6 mg/g). The results showed that the EP adsorption capacity is on average about 0.88 and 0.86 times the adsorption capacities of silica gel and activated carbon, respectively (Figure 6). These differences are mainly attributed to the specific surface area of the adsorbent materials (Oshita et al., 2010). The specific surface area of silica gel is 749.62 m²/g, and the specific surface area of activated carbon is 1240.0 m²/g (Delage et al., 2000), which are much larger values than those corresponding to RP (1.2 m²/g) and EP (2.12 m²/g). The specific surface

area of EP could explain why it had a slightly higher siloxane adsorption capacity than the RP. Moreover, it is assumed that the density of silanol groups on the RP and EP surface is not directly proportional to the adsorption capacity of siloxanes at low region D4 concentration.

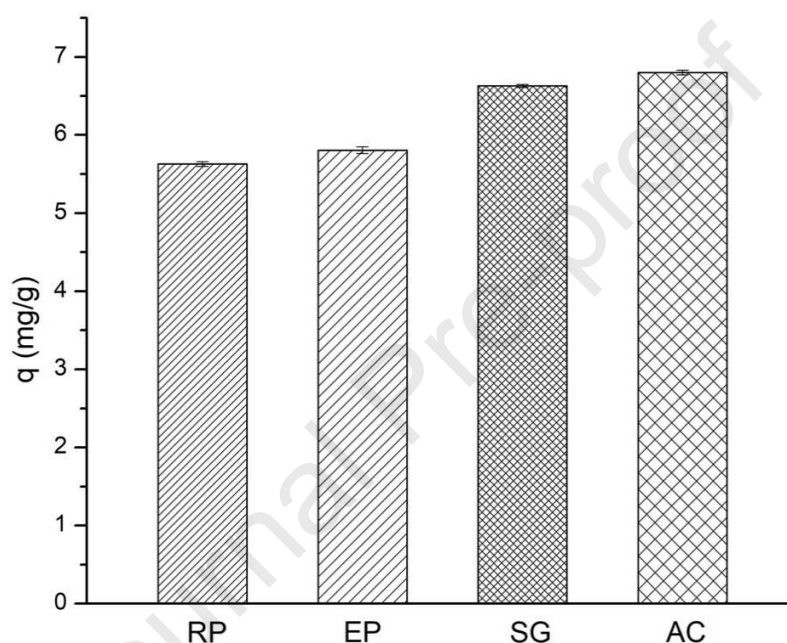


Figure 6. Equilibrium adsorption capacity of raw perlite (RP), expanded perlite (EP), silica gel (SG), and activated carbon (AC). Experimental conditions: 0.154 g adsorbent mass, 25 °C, 450 mg/m³ D4 initial concentration. Error bars represent standard deviation of triplicate measurements.

3.3 Adsorption Kinetics

The adsorption kinetics of D4 on RP and EP were performed using 0.15 g of each adsorbent in a batch system at 25 °C, 1 atm, and initial concentration of D4 from (80-

450) mg/m^3 . As can be seen in Figure 7, there is a rapid increase in the adsorption capacity in the first 10 h for both adsorbents, followed by a slower increase until reaching a plateau in approximately 170 h. Additionally, an increase in the adsorption capacity is observed as the initial concentration of D4 rises, while the equilibrium time does not depend on the initial concentration. Therefore, the equilibrium time is the same for all studied experiments at different initial concentrations. For instance, at $80 \text{ mg}/\text{m}^3$, the equilibrium adsorption capacities were $1.06 \text{ mg D4}/\text{g}$ and $0.97 \text{ mg D4}/\text{g}$ for RP and EP, respectively; while at $450 \text{ mg}/\text{m}^3$, an increase in the equilibrium adsorption capacities of 81% and 83.3% were obtained for RP and EP, respectively. As the initial D4 concentration increases, the number of active sites occupied on the perlite surface increments as well, causing an increase of the adsorption capacity. It is important to mention that long time is required for RP and EP to achieve equilibrium (170 h), which can be due to the low specific surface area ($< 2 \text{ m}^2/\text{g}$). Nevertheless, these low-cost materials can reduce the siloxane D4 concentration to values lower than the recommended by leading manufacturers ($< 28 \text{ mg}/\text{m}^3$) (Gaj, 2017).

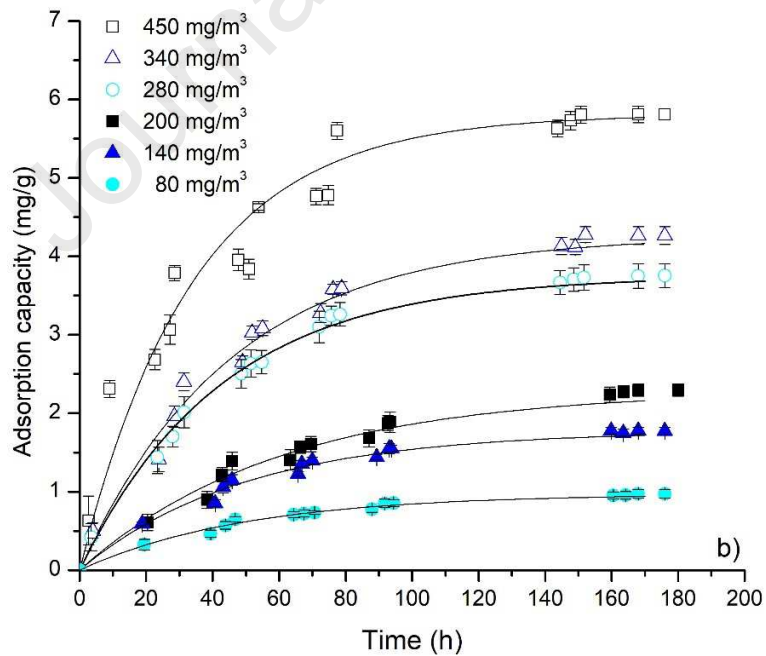
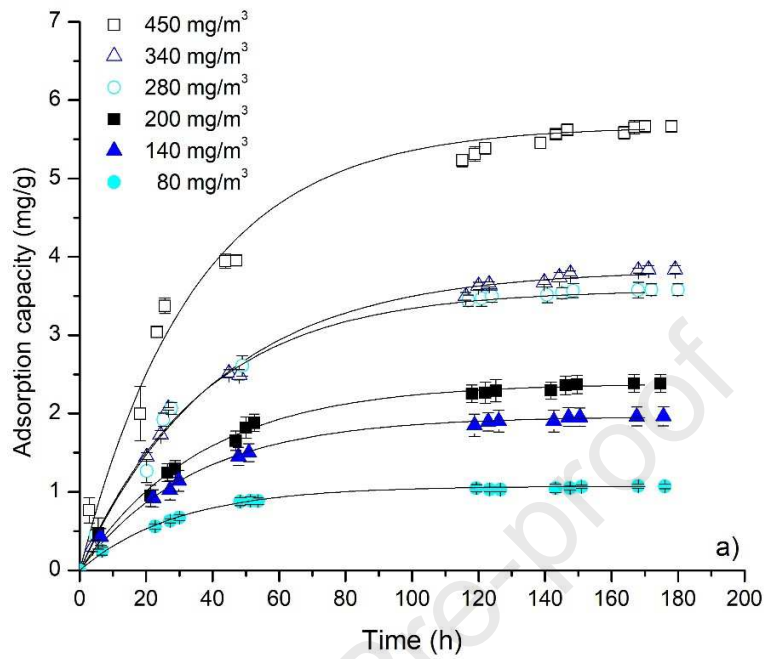


Figure 7. Adsorption kinetics of siloxane D4 onto (a) raw perlite and (b) expanded perlite at several initial concentrations of D4: (●) 80, (▲) 140, (■) 200, (○) 280, (△) 340, and (□) 450 mg/m³. Solid lines represent the fitting results of the LDF model.

On the other hand, four mathematical models, including the Elovich, Ritchie, LDF, and Crank models were tested to predict the adsorption kinetics of D4 onto RP and EP. The standard error, coefficient of determination, and the model parameters are listed in Tables 3 and 4 for the adsorption of D4 onto RP and EP, respectively. The Elovich model is a well-known kinetics model that describes the adsorption kinetics of gas molecules on solids. Although it has been applied to describe chemisorption kinetics, the phenomena behind the equation are related to the linear increase in the activation energy for adsorption with the surface coverage of the adsorbent. The standard errors and coefficients of determination obtained by the Elovich model are in the range of 0.059-0.336 and 0.970-0.990, respectively. These high values can be due to the fact that this model have two adjustable parameters. In contrast, when the Crank model was used, the adsorption kinetics data exhibited the highest SE values, suggesting that the adsorption of D4 onto RP and EP are not controlled by internal mass transfer. Similarly, the Ritchie model does not predict well the adsorption kinetics data according to the low R² obtained values. On the other hand, the LDF model present the highest R² (0.974-0.993) and the lowest SE values (0.016-0.404) indicating that the mass transfer is the rate-controlling step.

Table 3. Parameters of adsorption kinetic models of siloxane D4 onto raw perlite.

Elovich Model					
Co (mg/m ³)	to (h)	a (mg/g·h)	b (g/mg)	SE	R ²
80	3.742 (±1.290)	0.078 (±0.020)	3.411 (±0.246)	0.059	0.974
140	5.236 (±1.891)	0.111 (±0.032)	1.722 (±0.060)	0.067	0.991
200	6.034 (±3.414)	0.122 (±0.044)	1.360 (±0.143)	0.105	0.985
280	8.331 (±1.522)	0.147 (±0.019)	0.819 (±0.032)	0.194	0.982
340	10.519 (±1.758)	0.133 (±0.014)	0.716 (±0.035)	0.149	0.990
450	4.802 (±1.086)	0.335 (±0.051)	0.621 (±0.046)	0.225	0.988
Ritchie Model					
Co (mg/m ³)		k _R (1/h)		SE	R ²
80		0.061 (±0.008)		0.090	0.923
140		0.052 (±0.006)		0.178	0.920
200		0.049 (±0.007)		0.226	0.908
280		0.045 (±0.003)		0.416	0.888
340		0.040 (±0.002)		0.413	0.895
450		0.053 (±0.004)		0.504	0.925
LDF Model					
Co (mg/m ³)		k _L (1/h)		SE	R ²
80		0.034 (±0.003)		0.016	0.998
140		0.029 (±0.002)		0.038	0.997
200		0.027 (±0.001)		0.059	0.995
280		0.027 (±0.001)		0.101	0.994
340		0.024 (±0.001)		0.092	0.996
450		0.029 (±0.002)		0.216	0.989
Crank Model					
Co (mg/m ³)		D _s (m ² /h)		SE	R ²
80		1.29×10 ⁻⁹		0.323	0.624
140		1.29×10 ⁻⁹		0.679	0.771
200		1.29×10 ⁻⁹		0.803	0.827
280		1.29×10 ⁻⁹		1.285	0.843
340		1.29×10 ⁻⁹		1.379	0.889
450		1.29×10 ⁻⁹		1.883	0.913

a, to, and b: Elovich constants, K_R: Ritchie kinetic rate constant, K_L: effective mass transfer coefficient, D_s: effective diffusion coefficient, Co: D4 initial concentration, SE: standard error, R²: coefficient of determination. The standard deviations are reported in parentheses.

Table 4. Parameters of adsorption kinetic models of siloxane D4 onto expanded perlite.

Elovich Model					
Co (mg/m ³)	to (h)	a (mg/g·h)	b (g/mg)	SE	R ²

80	11.824 (± 3.608)	0.031 (± 0.006)	2.736 (± 0.159)	0.044	0.980
140	12.870 (± 1.117)	0.052 (± 0.005)	1.493 (± 0.035)	0.075	0.982
200	24.314 (± 6.559)	0.046 (± 0.010)	0.887 (± 0.074)	0.110	0.979
280	9.350 (± 3.985)	0.143 (± 0.041)	0.750 (± 0.072)	0.186	0.981
340	10.830 (± 2.357)	0.146 (± 0.023)	0.634 (± 0.040)	0.229	0.978
450	4.209 (± 1.582)	0.384 (± 0.123)	0.618 (± 0.064)	0.356	0.970

Ritchie Model

Co (mg/m ³)	k_R (1/h)	SE	R ²
80	0.038 (± 0.003)	0.097	0.868
140	0.038 (± 0.001)	0.186	0.859
200	0.028 (± 0.003)	0.278	0.814
280	0.041 (± 0.005)	1.787	0.828
340	0.038 (± 0.003)	0.451	0.884
450	0.055 (± 0.005)	0.531	0.929

LDF Model

Co (mg/m ³)	k_L (1/h)	SE	R ²
80	0.020 ($\pm 7.39 \times 10^{-4}$)	0.034	0.987
140	0.020 ($\pm 4.86 \times 10^{-4}$)	0.072	0.983
200	0.017 ($\pm 1.22 \times 10^{-3}$)	0.118	0.972
280	0.023 ($\pm 1.64 \times 10^{-3}$)	0.091	0.995
340	0.022 ($\pm 8.67 \times 10^{-4}$)	0.161	0.987
450	0.029 ($\pm 1.27 \times 10^{-3}$)	0.404	0.962

Crank Model

Co (mg/m ³)	D_s (m ² /h)	SE	R ²
80	1.37×10^{-9}	0.314	0.686
140	1.37×10^{-9}	0.574	0.795
200	1.37×10^{-9}	0.838	0.869
280	1.37×10^{-9}	1.459	0.918
340	1.37×10^{-9}	1.653	0.921
450	1.37×10^{-9}	1.885	0.914

a, to, and b: Elovich constants, K_R : Ritchie kinetic rate constant, K_L : effective mass transfer coefficient, D_s : effective diffusion coefficient, Co: D4 initial concentration, SE: standard error, R²: coefficient of determination. The standard deviations are reported in parentheses.

3.4 Adsorbent Regeneration Studies

The regeneration process occurs when a spent adsorbent material is subjected to the release of the adsorbed species to recover the adsorption sites, but with the minimum possible alteration of its physicochemical or textural properties. In this study, the

desorption of D4 from EP by thermal regeneration was evaluated at 200 °C. In Figure 8, it is observed that the mass of D4 desorbed is 0.08 mg, which is equivalent to 40 % of the total amount of D4 adsorbed on EP.

It is reported that during the desorption of D4 on activated carbon and silica gel, a chemical transformation of D4 occurs in cyclic oligomers such as hexamethylcyclotrisiloxane limiting the thermal regeneration due to the closing of pores and therefore the decrease in the adsorption capacity of siloxanes of the regenerated material, consequently, silica gel (Sigot et al., 2015) and activated carbon (Tran et al., 2019a) cannot be used in further adsorption process. In this study, low regeneration time and temperature were needed for the desorption of D4 from EP, suggesting weak interactions between the silanol groups at expanded perlite surface and the siloxane D4.

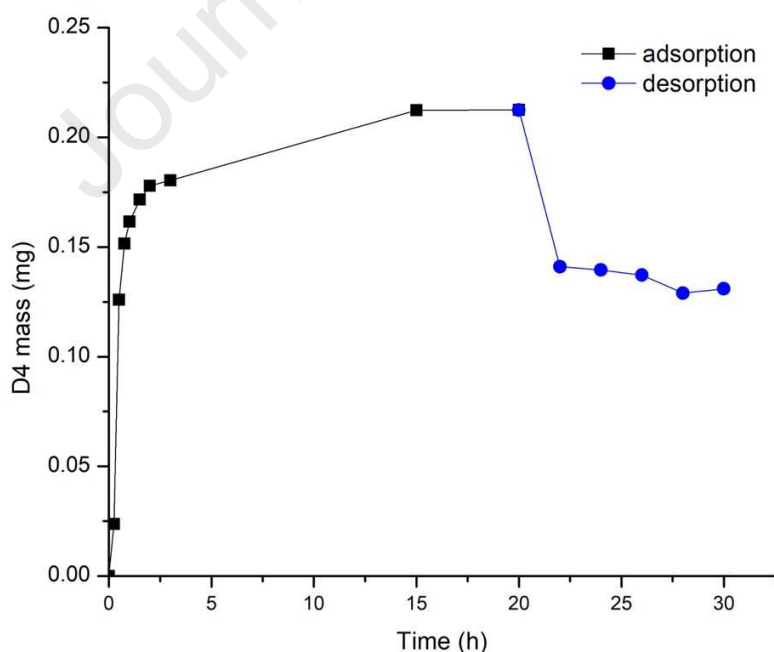


Figure 8. Adsorption-desorption kinetics of siloxane D4 onto expanded perlite.

Experimental conditions: 450 mg/m³ D4 initial concentration, desorption temperature of 200 °C.

3.5 Removal cost analysis

A simple estimation of the costs associated with the usage of EP and activated carbon for the removal of D4 was made taking into consideration 1 m³/h biogas, 130 mg/Nm³ siloxane concentration (a value commonly found in real samples of biogas), the adsorption capacity (5.8 and 6.8 mg/g for EP and activated carbon (Gupta et al., 2009), respectively), and the price of each material (US\$ 2.18/kg and US\$ 3.3/kg for EP and activated carbon, respectively). When the D4 initial concentration is 130 mg/Nm³, the adsorbent mass per hour (\dot{m}_{ads}) of activated carbon required for the purification of 1 m³/h of biogas is 0.0191 Kg activated carbon/h (obtained by Eq. 13). Since, 1 m³/h can be converted to 1.4 kWh of electricity (Mitiku Teferra and Wubu, 2019). Thus, the cost associated to the usage of activated carbon for the removal of 130 mg/Nm³ siloxane is US\$ 0.045/kWh (by using Eq. 12). In the case of EP, the \dot{m}_{ads} required is 0.0224 kg EP/h with a cost of US\$ 0.0349/kWh. Although the adsorption capacity of EP is slightly lower than the activated carbon, its low cost contributes to the economy of the purification process. This cost could be reduced by using EP in several adsorption/desorption cycles. It is important to mention that the cost of the thermal regeneration is not included. Hence, further studies need to be conducted to estimate the total cost associated to the siloxane removal from biogas that includes thermal regeneration cost and the reusability of EP.

4. Conclusions and future prospects

In this study, the use of perlite as a low cost and efficient adsorbent for removing siloxane has been demonstrated. The adsorbent characterization analysis revealed the morphological nature of RP and EP as amorphous materials with low specific surface area and mesoporous pore size distribution, but good thermal stability. In addition, silanol functional groups were identified by FTIR analysis, which are described as adsorption sites for siloxanes. Furthermore, it was evidenced that the thermal expansion of EP causes a reduction in the density of silanol groups from 2515.43 to 653.75 OH/nm²; in contrast, the specific surface area of EP increased two-fold due to the thermal expansion. Also, the isotherm adsorption data were adjusted to Henry's law model at low D4 concentrations. Similar adsorption capacities were obtained for EP, RP, silica gel, and activated carbon in the range of 5.6-6.8 mg D4/g. The adsorption kinetics followed the LDF model suggesting that the mass transfer is the rate-controlling step. EP could be regenerated at 200°C showing fast desorption kinetics, that could represent an advantage over activated carbon which requires temperatures higher than 400 °C.

As future work, the selectivity in multicomponent gas samples will be assessed on dynamic operation. Also, the increase of the specific surface area of perlite will be investigated by applying physicochemical treatments to enhance the adsorption kinetics rate. Finally, the comparative techno-economical evaluation of the perlite and activated carbon for biogas purification will be carried out by life cycle assessment methodology.

Data availability

Datasets related to this article can be found at <http://dx.doi.org/10.17632/4gbmztd92j.1>, an open-source online data repository hosted at Mendeley Data (Pioquinto-Garcia et al., 2019).

Acknowledgments

We would like to gratefully acknowledge to Facultad de Ciencias Químicas, UANL and Consejo Nacional de Ciencia y Tecnología de México (Scholarship number 604928). Also, we would like to thank Ignacio López-Barro, Edgar Valmaña, and Gerardo Magaña from the company Termolita SAPI de CV for the support provided during the realization of this project.

References

- Ajhar, M., Travasset, M., Yüce, S., Melin, T., 2010. Siloxane removal from landfill and digester gas - A technology overview. *Bioresour. Technol.* 101, 2913–2923. <https://doi.org/10.1016/j.biortech.2009.12.018>
- Alkan, M., Karadaş, M., Doğan, M., Demirbaş, Ö., 2005. Adsorption of CTAB onto perlite samples from aqueous solutions. *J. Colloid Interface Sci.* 291, 309–318. <https://doi.org/10.1016/J.JCIS.2005.05.027>
- Altındal, A., Kurt, Ö., Şengül, A., Bekaroğlu, Ö., 2014. Kinetics of CO₂ adsorption on ball-type dicopper phthalocyanine thin film. *Sens. Actuators B Chem.* 202, 373–381. <https://doi.org/10.1016/j.snb.2014.05.107>
- Álvarez-Flórez, J., Egusquiza, E., 2015. Analysis of damage caused by siloxanes in stationary reciprocating internal combustion engines operating with landfill gas. *Eng. Fail. Anal.* 50, 29–38. <https://doi.org/10.1016/j.engfailanal.2015.01.010>
- American Biogas Council, 2014. Current and Potential Biogas Production. <https://americanbiogascouncil.org/> (accessed 30 January 2019).
- Bak, C. u., Lim, C.J., Lee, J.G., Kim, Y.D., Kim, W.S., 2019. Removal of sulfur compounds and siloxanes by physical and chemical sorption. *Sep. Purif. Technol.* 209,

- 542–549. <https://doi.org/10.1016/j.seppur.2018.07.080>
- Cabrera-Codony, A., Santos-Clotas, E., Ania, C.O., Martín, M.J., 2018. Competitive siloxane adsorption in multicomponent gas streams for biogas upgrading. *Chem. Eng. J.* 344, 565–573. <https://doi.org/10.1016/J.CEJ.2018.03.131>
- Celik, A.G., Kilik, A.M., Cakal, G.O., 2013. Expanded perlite aggregate characterization for use as a lightweight construction raw material. *Physicochem. Probl. Miner. Process.* 49 685-700 ISSN 2084-4735. <https://doi.org/10.5277/ppmp130227>
- Chaemchuen, S., Zhou, K., Verpoort, F., 2016. From Biogas to Biofuel: Materials Used for Biogas Cleaning to Biomethane. *ChemBioEng Rev.* 3, 250–265. <https://doi.org/10.1002/cben.201600016>
- Cooney, D.O., 1998. *Adsorption Design for Wastewater Treatment*, CRC Press, Florida
- Crank, John., 1975. *The mathematics of diffusion*, second ed. Clarendon Press, London
- De Farias, R.F., Airoidi, C., 1998. Thermogravimetry as a Reliable tool to Estimate the Density of Silanols on a Silica Gel Surface. *J. Therm. Anal. Calorim.* 53, 751–756. <https://doi.org/10.1023/A:1010174224567>
- Delage, F., Pre, P., Le Cloirec, P., 2000. Mass transfer and warming during adsorption of high concentrations of VOCs on an activated carbon bed: Experimental and theoretical analysis. *Environ. Sci. Technol.* 34, 4816–4821. <https://doi.org/10.1021/es001187x>
- Dewil, R., Appels, L., Baeyens, J., 2006. Energy use of biogas hampered by the presence of siloxanes. *Energy Convers. Manag.* 47, 1711–1722. <https://doi.org/10.1016/j.enconman.2005.10.016>
- Gaj, K., 2017. Applicability of selected methods and sorbents to simultaneous removal of siloxanes and other impurities from biogas. *Clean Technol. Environ. Policy* 19, 2181–2189. <https://doi.org/10.1007/s10098-017-1422-1>
- Giraudet, S., Boulinguez, B., Le Cloirec, P., 2014. Adsorption and electrothermal desorption of volatile organic compounds and siloxanes onto an activated carbon fiber cloth for biogas purification. *Energy Fuels* 28, 3924–3932. <https://doi.org/10.1021/ef500600b>
- Gong, H., Chen, Z., Fan, Y., Zhang, M., Wu, W., Wang, W., 2015. Surface modification of activated carbon for siloxane adsorption. *Renew. Energy* 83, 144–150. <https://doi.org/10.1016/J.RENENE.2015.04.004>
- Gupta, V.K., Carrott, P.J.M., Ribeiro Carrott, M.M.L., Suhas, 2009. Low-Cost adsorbents: Growing approach to wastewater treatment a review. *Crit. Rev. Environ. Sci. Technol.* 39, 783–842. <https://doi.org/10.1080/10643380801977610>
- IMTA, 2016. Revisión y actualización del potencial de biomasa para generación de energía eléctrica a partir de plantas de tratamiento de aguas residuales presentado en el Inventario Nacional de Energías Renovables (INERE). <http://repositorio.imta.mx/handle/20.500.12013/1775> (accessed 30 January 2019)
- Jiang, T., Zhong, W., Jafari, T., Du, S., He, J., Fu, Y.-J., Singh, P., Suib, S.L., 2016. Siloxane D4 adsorption by mesoporous aluminosilicates. *Chem. Eng. J.* 289, 356–364. <https://doi.org/10.1016/J.CEJ.2015.12.094>
- Kapoor, R., Ghosh, P., Tyagi, B., Vijay, V. K., Vijay, V., Thakur, I. S., Kamyab, H., Nguyen, D. D., & Kumar, A., 2020. Advances in biogas valorization and utilization systems: A comprehensive review. *Journal of Cleaner Production*, 273, 123052. <https://doi.org/10.1016/j.jclepro.2020.123052>

- Khoshnevisan, B., Tsapekos, P., Alfaro, N., Díaz, I., Fdz-Polanco, M., Rafiee, S., Angelidaki, I., 2017. A review on prospects and challenges of biological H₂S removal from biogas with focus on biotrickling filtration and microaerobic desulfurization. *Biofuel Res. J.* 4, 741–750. <https://doi.org/10.18331/BRJ2017.4.4.6>
- Liu, C., Huang, P.M., 2002. Role of hydroxy-aluminosilicate ions (proto-imogolite sol) in the formation of humic substances. *Org. Geochem.* 33, 295–305. [https://doi.org/10.1016/S0146-6380\(01\)00161-9](https://doi.org/10.1016/S0146-6380(01)00161-9)
- Matsui, T., Imamura, S., 2010. Removal of siloxane from digestion gas of sewage sludge. *Bioresour. Technol.* 101, S29–S32. <https://doi.org/10.1016/j.biortech.2009.05.037>
- Meng, Z.Y., Liu, Y.H., Ma, Z.C., Hou, X.F., 2020. The regulation of micro/mesoporous silica gel by polyethylene imine for enhancing the siloxane removal. *Inorg. Chem. Commun.* 112, 107754. <https://doi.org/10.1016/j.inoche.2019.107754>
- Mitiku Teferra, D., Wubu, W., 2019. Biogas for Clean Energy, in: *Anaerobic Digestion*. IntechOpen. <https://doi.org/10.5772/intechopen.79534>
- Nam, S., Namkoong, W., Kang, J.-H., Park, J.-K., Lee, N., 2013. Adsorption characteristics of siloxanes in landfill gas by the adsorption equilibrium test. *Waste Manag.* 33, 2091–8. <https://doi.org/10.1016/j.wasman.2013.03.024>
- Oshita, K., Ishihara, Y., Takaoka, M., Takeda, N., Matsumoto, T., Morisawa, S., Kitayama, A., 2010. Behaviour and adsorptive removal of siloxanes in sewage sludge biogas. *Water Sci. Technol.* 61, 2003–2012. <https://doi.org/10.2166/wst.2010.101>
- Poloncarzova, M., Vejrazka, J., Vesely, V., Izak, P., 2011. Effective purification of biogas by a condensing-liquid membrane. *Angew. Chem. - Int. Ed.* 50, 669–671. <https://doi.org/10.1002/anie.201004821>
- REMBIO, 2012. Producción de biogás en México estado actual y perspectivas, Cuaderno Temático No. 5 Red Mexicana de Bioenergía. <http://rembio.org.mx> (accessed 30 January 2019).
- Ritchie, A.G., 1977. Alternative to the Elovich equation for the kinetics of adsorption of gases on solids. *J. Chem. Soc. Faraday Trans. 1 Phys. Chem. Condens. Phases* 73, 1650–1653. <https://doi.org/10.1039/F19777301650>
- Roulia, M., Chassapis, K., Kapoutsis, J.A., Kamitsos, E.I., Savvidis, T., 2006. Influence of thermal treatment on the water release and the glassy structure of perlite. *J. Mater. Sci.* 41, 5870–5881. <https://doi.org/10.1007/s10853-006-0325-z>
- Ruiling, G., Shikun, C., Zifu, L., 2017. Research progress of siloxane removal from biogas. *Int. J. Agric. Biol. Eng.* 10, 30–39. <https://doi.org/10.25165/IJABE.V10I1.3043>
- Russo, V., Trifuoggi, M., Di Serio, M., Tesser, R., 2017. Fluid-Solid Adsorption in Batch and Continuous Processing: A Review and Insights into Modeling. *Chem. Eng. Technol.* 40, 799–820. <https://doi.org/10.1002/ceat.201600582>
- Santos, I.F.S. dos, Gonçalves, A.T.T., Borges, P.B., Barros, R.M., da Silva Lima, R., 2018. Combined use of biogas from sanitary landfill and wastewater treatment plants for distributed energy generation in Brazil. *Resour. Conserv. Recycl.* 136, 376–388. <https://doi.org/10.1016/J.RESCONREC.2018.05.011>
- Santos-Clotas, E., Cabrera-Codony, A., Ruiz, B., Fuente, E., Martín, M.J., 2019. Sewage biogas efficient purification by means of lignocellulosic waste-based activated carbons. *Bioresour. Technol.* 275, 207–215. <https://doi.org/10.1016/j.biortech.2018.12.060>

- Scarlat, N., Dallemand, J.-F., Fahl, F., 2018. Biogas: Developments and perspectives in Europe. *Renew. Energy* 129, 457–472. <https://doi.org/10.1016/J.RENENE.2018.03.006>
- Shen, M., Zhang, Y., Hu, D., Fan, J., Zeng, G., 2018. A review on removal of siloxanes from biogas: with a special focus on volatile methylsiloxanes. *Environ. Sci. Pollut. Res.* <https://doi.org/10.1007/s11356-018-3000-4>
- Sigot, L., Ducom, G., Benadda, B., Labouré, C., 2014. Adsorption of octamethylcyclotetrasiloxane on silica gel for biogas purification. *Fuel* 135, 205–209. <https://doi.org/10.1016/j.fuel.2014.06.058>
- Sigot, L., Ducom, G., Germain, P., 2015. Adsorption of octamethylcyclotetrasiloxane (D4) on silica gel (SG): Retention mechanism. *Microporous Mesoporous Mater.* 213, 118–124. <https://doi.org/10.1016/j.micromeso.2015.04.016>
- Sircar, S., Hufton, J.R., 2000. Why does the linear driving force model for adsorption kinetics work? *Adsorption* 6, 137–147. <https://doi.org/10.1023/A:1008965317983>
- Tabatabaei, M., Aghbashlo, M., Valijanian, E., Kazemi Shariat Panahi, H., Nizami, A.-S., Ghanavati, H., Sulaiman, A., Mirmohamadsadeghi, S., Karimi, K., 2020. A comprehensive review on recent biological innovations to improve biogas production, Part 1: Upstream strategies. *Renew. Energy* 146, 1204–1220. <https://doi.org/10.1016/j.renene.2019.07.037>
- Tran, V.T.L., Gélín, P., Ferronato, C., Chovelon, J.M., Fine, L., Postole, G., 2019a. Adsorption of linear and cyclic siloxanes on activated carbons for biogas purification: Sorbents regenerability. *Chem. Eng. J.* 378, 122152. <https://doi.org/10.1016/j.cej.2019.122152>
- Tran, V.T.L., Gélín, P., Ferronato, C., Mascunan, P., Rac, V., Chovelon, J.-M., Postole, G., 2019b. Siloxane adsorption on activated carbons: Role of the surface chemistry on sorption properties in humid atmosphere and regenerability issues. *Chem. Eng. J.* 371, 821–832. <https://doi.org/10.1016/J.CEJ.2019.04.087>
- Ullah Khan, I., Hafiz Dzarfan Othman, M., Hashim, H., Matsuura, T., Ismail, A.F., Rezaei-DashtArzhandi, M., Wan Azelee, I., 2017. Biogas as a renewable energy fuel – A review of biogas upgrading, utilisation and storage. *Energy Convers. Manag.* 150, 277–294. <https://doi.org/10.1016/J.ENCONMAN.2017.08.035>
- United Nations, 2015. Transforming Our World: the 2030 Agenda for Sustainable Development. <https://www.unfpa.org/resources/transforming-our-world-2030-agenda-sustainable-development> (accessed 30 January 2019).
- Wang, G., Zhang, Z., Hao, Z., 2019. Recent advances in technologies for the removal of volatile methylsiloxanes: A case in biogas purification process. *Crit. Rev. Environ. Sci. Technol.* 49, 2257–2313. <https://doi.org/10.1080/10643389.2019.1607443>
- Yao, C., Tien, C., 1992. Approximation of intraparticle mass transfer in adsorption processes—I. Linear systems. *Chem. Eng. Sci.* 47, 457–464. [https://doi.org/10.1016/0009-2509\(92\)80033-9](https://doi.org/10.1016/0009-2509(92)80033-9)

Highlights:

1. The density of silanol groups in perlite decrease 74% after thermal expansion.
2. Perlite has similar siloxane adsorption capacity than activated carbon and silica gel.
3. A long time is required for perlite to achieved adsorption equilibrium (170 h).
4. Perlite can reduce siloxane concentration to 28 mg/m^3 as recommended by manufacturers.
5. Perlite showed fast desorption (20 min) and a low regeneration temperature ($200 \text{ }^\circ\text{C}$).

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Journal Pre-proof