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► **To cite this version:**

Kathrin Fuchs, Lutz Merbold, Nina Buchmann, Gianni Bellocchi, Marco Bindi, et al.. Evaluating the Potential of Legumes to Mitigate N<sub>2</sub>O Emissions From Permanent Grassland Using Process-Based Models. *Global Biogeochemical Cycles*, 2020, 34 (12), 10.1029/2020GB006561 . hal-03082769

**HAL Id: hal-03082769**

**<https://hal.science/hal-03082769v1>**

Submitted on 17 Oct 2023

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# Global Biogeochemical Cycles

## RESEARCH ARTICLE

10.1029/2020GB006561

### Key Points:

- A partial substitution of fertilizer nitrogen with symbiotically fixed nitrogen could mitigate nitrous oxide (N<sub>2</sub>O) emissions in grasslands by around 130 Gg yr<sup>-1</sup>
- Experimentally testing this mitigation option is challenging so modeling offers means to identify the optimum legume/fertilizer combination
- The models showed that net benefits to N<sub>2</sub>O mitigation and yield can be achieved across a wide range of legume/fertilizer combinations

### Supporting Information:

- Supporting Information S1

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### Citation:

Fuchs, K., Merbold, L., Buchmann, N., Bellocchi, G., Bindi, M., Brill, L., et al. (2020). Evaluating the potential of legumes to mitigate N<sub>2</sub>O emissions from permanent grassland using process-based models. *Global Biogeochemical Cycles*, 34, e2020GB006561. <https://doi.org/10.1029/2020GB006561>

Received 29 JAN 2020

Accepted 16 OCT 2020

Accepted article online 21 OCT 2020

## Evaluating the Potential of Legumes to Mitigate N<sub>2</sub>O Emissions From Permanent Grassland Using Process-Based Models

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**Abstract** A potential strategy for mitigating nitrous oxide (N<sub>2</sub>O) emissions from permanent grasslands is the partial substitution of fertilizer nitrogen (N<sub>fert</sub>) with symbiotically fixed nitrogen (N<sub>symb</sub>) from legumes. The input of N<sub>symb</sub> reduces the energy costs of producing fertilizer and provides a supply of nitrogen (N) for plants that is more synchronous to plant demand than occasional fertilizer applications. Legumes have been promoted as a potential N<sub>2</sub>O mitigation strategy for grasslands, but evidence to support their efficacy is limited, partly due to the difficulty in conducting experiments across the large range of potential combinations of legume proportions and fertilizer N inputs. These experimental constraints can be overcome by biogeochemical models that can vary legume-fertilizer combinations and subsequently aid the design of targeted experiments. Using two variants each of two biogeochemical models (APSIM and DayCent), we tested the N<sub>2</sub>O mitigation potential and productivity of full factorial combinations of legume proportions and fertilizer rates for five temperate grassland sites across the globe. Both models showed that replacing fertilizer with legumes reduced N<sub>2</sub>O emissions without reducing productivity across a broad range of legume-fertilizer combinations. Although the models were consistent with the relative changes of N<sub>2</sub>O emissions compared to the baseline scenario (200 kg N ha<sup>-1</sup> yr<sup>-1</sup>; no legumes), they predicted different levels of absolute N<sub>2</sub>O emissions and thus also of absolute N<sub>2</sub>O emission reductions; both were greater in DayCent than in APSIM. We recommend confirming these results with experimental studies assessing the effect of clover proportions in the range 30–50% and ≤150 kg N ha<sup>-1</sup> yr<sup>-1</sup> input as these were identified as best-bet climate smart agricultural practices.

## 1. Introduction

Nitrous oxide (N<sub>2</sub>O) is a powerful greenhouse gas (GHG) and plays a dominant role in stratospheric ozone depletion (Ravishankara et al., 2009; Smith et al., 2014). Agriculture contributes 58–84% to global anthropogenic N<sub>2</sub>O emissions (Mosier et al., 2004; P. Smith et al., 2008) and is thus important for N<sub>2</sub>O mitigation (Ciais et al., 2013; K. A. Smith, 2017). Most agricultural N<sub>2</sub>O emissions arise from inputs of nitrogen (N) by being too high or out of synchrony with crop or pasture demand, resulting in the applied N being susceptible to losses through leaching or gaseous emissions including N<sub>2</sub>O (Davidson & Kanter, 2014).

Grassland systems cover about 40% of the Earth's land surface and a quarter of these are located in temperate zones (White et al., 2000). Globally estimated N<sub>2</sub>O emissions from managed grasslands contributed 54% to

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the total agricultural emissions in 2006 (Dangal et al., 2019). These N<sub>2</sub>O emissions significantly increased during recent decades due to increased mineral and organic fertilizer applications and excreta deposition from 1.74 Tg N<sub>2</sub>O-N (1961) to 3.11 Tg N<sub>2</sub>O-N (2014) (Dangal et al., 2019).

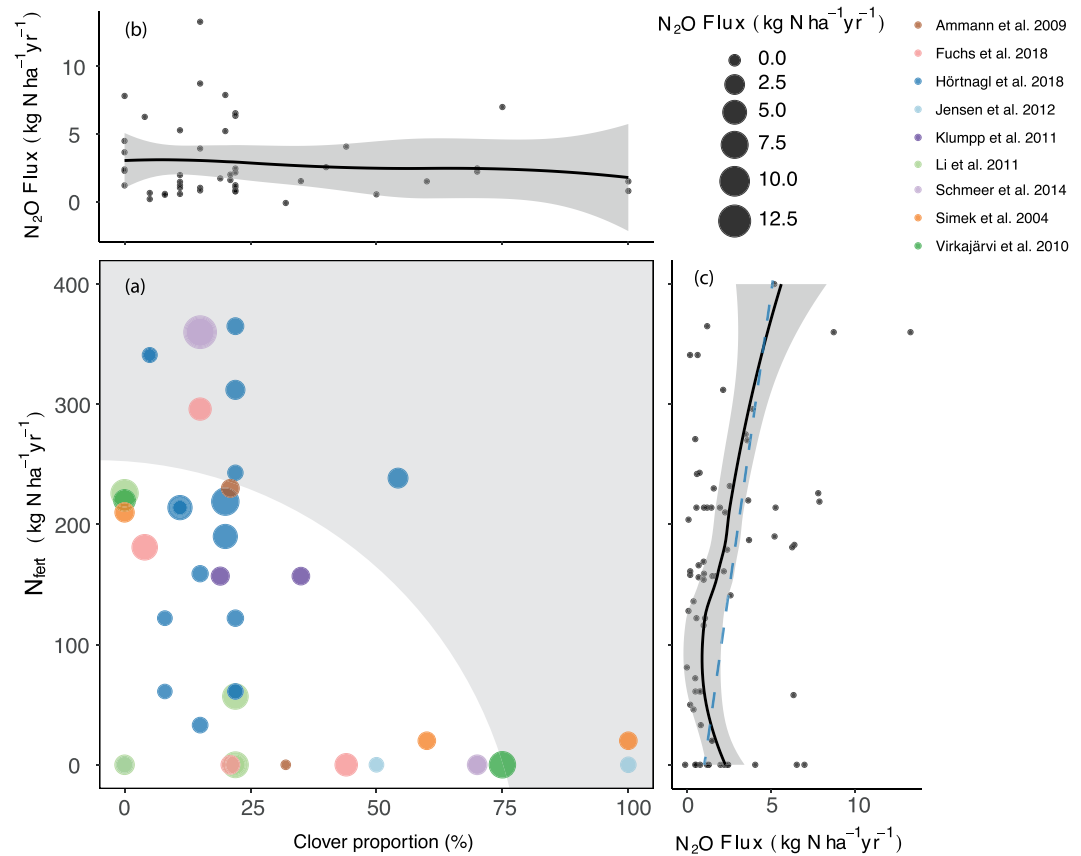
Besides fertilizer amendments (Flechard et al., 2007; K. A. Smith et al., 2012; Hörtnagl et al., 2018), recycling of N through the excreta of grazing animals (Saggar et al., 2004), aerial N deposition (Well & Butterbach-Bahl, 2010) and biological nitrogen fixation (BNF) via legume symbiosis with *Rhizobia* (Jensen et al., 2012) deliver important N inputs to grasslands while having the potential to increase N<sub>2</sub>O emissions. Environmental variables, nitrogen rate, type of N input, timing, and placement of organic and mineral N fertilizers are important factors that influence the magnitude and duration of N<sub>2</sub>O emission pulses (Baggs, 2011; Flessa, 2012; Shcherbak et al., 2014; K. A. Smith et al., 1997, 2012).

Nitrous oxide mitigation strategies for grasslands mainly focus on an increased N use efficiency (NUE) (K. A. Smith et al., 1997) by adjusting the amount, type, and timing of fertilizer to better match plant N demand (Bolan et al., 2004; Eckard et al., 2010; de Klein & Eckard, 2008; D. Li et al., 2013; Luo et al., 2010; Saggar et al., 2013). Other mitigation strategies aim to alter soil conditions affecting the microbial production of N<sub>2</sub>O. Approaches here include liming of acid soils and the application of nitrification inhibitors (Cahalan et al., 2015; Galbally et al., 2010; Lam et al., 2017). The use of legumes has also been proposed as a mitigation strategy (Jensen et al., 2012; Rochette & Janzen, 2005). Legumes can (a) use symbiotically fixed N (N<sub>symb</sub>) from the atmosphere as substitute for mineral fertilizer, which has high GHG costs associated with its manufacture (1.6–6.4 kg CO<sub>2</sub>-eq per kg fertilizer N; Brentrup & Pallière, 2008) and (b) provide a N source for plants that is better synchronized to plant demand than the alternative, which is episodic applications of N fertilizer. In addition, there is evidence that suggest mixtures of legumes and grasses can produce higher yields than monocultures of grasses or legumes (“overyielding”) (Finn et al., 2013; Kirwan et al., 2007; Lüscher et al., 2014). The potential to sustain or even increase yields would be a highly desirable characteristic of any mitigation option both in terms of adoption of the technology by farmers (Vellinga et al., 2011) and its benefit in reducing GHG emissions per area of land and per unit of product, referred to as “N<sub>2</sub>O emission intensity” (N<sub>2</sub>O-Int) (van Groenigen et al., 2010).

Despite its theoretical advantages, the potential to mitigate N<sub>2</sub>O emissions with legumes remains largely untested (Flessa et al., 2014; Osterburg et al., 2013). Only few field experiments have measured N<sub>2</sub>O emissions for different legume proportions in permanent grassland systems, and these have shown mixed results (Ammann et al., 2009; Fuchs et al., 2018; Hörtnagl et al., 2018; Ingram et al., 2015; Jensen et al., 2012; Klumpp et al., 2011; D. Li et al., 2011; Schmeer et al., 2014; Simek, 2004; Virkajärvi et al., 2010) (Figure 1; see Supporting Information Text S1). Although N<sub>2</sub>O emissions across these studies were highly variable, a smoothed fit suggests constant emissions across the full range of clover proportions used in these studies (Figure 1b). In contrast, N<sub>2</sub>O emissions increase with increased fertilizer rates (Figure 1c). In order to test the robustness of legumes as a N<sub>2</sub>O mitigation option, it would be valuable to extend the experimental evidence to better cover a wider range of legume proportions, for example, by examining the potentially interesting combination of a low N fertilizer level and relatively high (30–80%) proportion of clover (Figure 1).

Currently, the potential of grass-legume mixtures is poorly exploited (Phelan et al., 2015). While New Zealand and Switzerland have promoted grass-legume mixtures and have grasslands with up to 30% legume content as the norm, the use of legumes is not the standard in many other countries. The area of grasslands containing legumes is furthermore not routinely captured in country statistics nor are the legume proportions. Similarly, there are large spatial and temporal variations in legume proportions caused by varying seed mixes and difficulties in legume establishment, management, and persistence. Due to this lack of data, a quantitative estimate of the *status quo* is difficult. This motivated the assessment of several simplified scenarios to approximate the magnitude of the global mitigation potential.

Costs and workload constrain experimental measurements to a few locations and seasons, with restrictions on the number of treatments and replicates. Overall, experiments investigating N<sub>2</sub>O emissions in swards of different clover proportions are limited by the temporal coverage and/or by the number of treatments (clover-fertilizer combinations). Studies applying the eddy covariance (EC) technique to acquire data of high temporal coverage and resolution can clearly resolve N<sub>2</sub>O emission peaks at the field-scale (Fuchs et al., 2018; Hörtnagl et al., 2018), while the number of simultaneously observed treatments in EC studies is very limited.

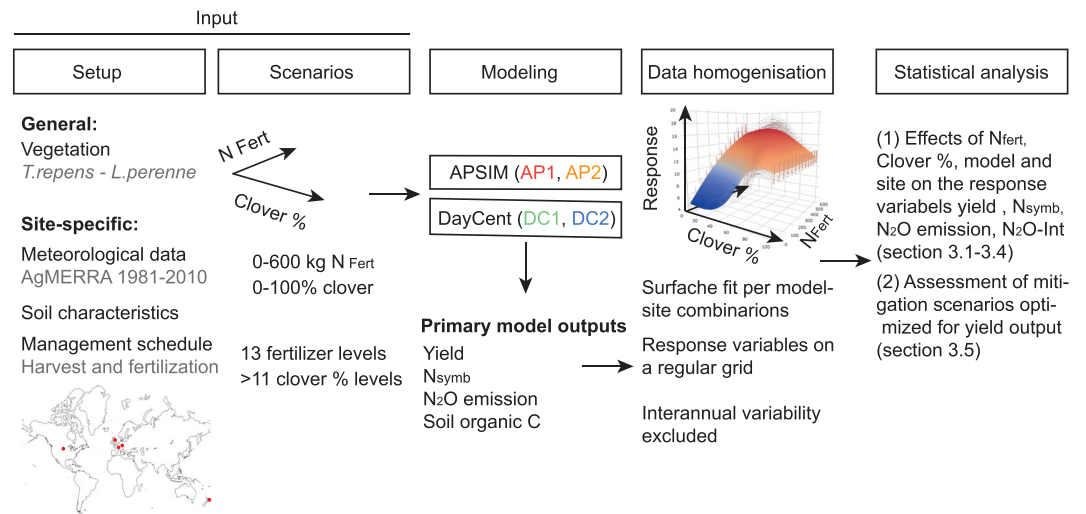


**Figure 1.** Observed data from literature (source = color) with documented N<sub>2</sub>O emissions, clover proportions and fertilizer amounts. Panel (a) relates observed clover proportions (horizontal axis) and fertilizer levels (vertical axis) at different sites to N<sub>2</sub>O emissions (bubble area). The gray-shaded area in (a) indicates that these combinations are unrealistic or of little or no practical interest, but some are important for testing model performance. The plots at the margins relate (b) clover proportion to N<sub>2</sub>O emissions across all fertilizer levels and (c) fertilizer amendments to N<sub>2</sub>O emissions across all clover proportions. The dashed blue line indicates N<sub>2</sub>O emissions according to the IPCC emission factor (1% of N<sub>fert</sub>), black lines indicate locally weighted regression (loess) fits of the displayed data and the shaded area their 95% confidence intervals (Cleveland & Devlin, 1988). All studies are from the temperate zone except Virkajärvi et al. (2010), which was in boreal grasslands characterized by freeze-thaw cycles.

Studies with more treatments tend to use manual GHG chambers to investigate N<sub>2</sub>O emissions, but these are generally limited by their coarse temporal resolution (e.g., Niklaus et al., 2016) and are therefore less capable of resolving temporal dynamics. However, if sampling frequency is too low, data are less suitable for annual N<sub>2</sub>O flux budgets (Barton et al., 2015). On the other hand, manual and automatic GHG chamber studies with higher sampling frequency (Klumpp et al., 2011; D. Li et al., 2011) covering a full growing season have often been constrained to two to five clover-fertilizer combinations only. In addition, chambers cover relatively small sampling areas (usually less than a few m<sup>2</sup>). While any in situ measurement delivers valuable data, a systematic assessment of the effect of different clover proportions on N<sub>2</sub>O emissions at different sites and scales (plot to ecosystem to catchment) is clearly lacking and unlikely to be resolved experimentally.

Biogeochemical process models allow for the ex-ante and ex-post estimation of GHG emissions (including N<sub>2</sub>O emissions) under many different management regimes. In particular, they can be used for a systematic assessment of mitigation strategies across several sites, investigating different rates of N fertilizer and levels of legume proportion simultaneously, which has so far not been achieved in field N<sub>2</sub>O measurements. Models can predict these different scenarios over a long time-span and reveal important long-term effects (e.g., Lugato et al., 2018).

In this study, we use the biogeochemical models APSIM (Holzworth et al., 2014) and DayCent (Del Grosso et al., 2001; Parton et al., 1998, 2001). These models have the ability to simulate key processes of C and N



**Figure 2.** Overview of the modeled scenarios: 13 levels of varying fertilizer amount (0–600 kg N ha<sup>-1</sup> yr<sup>-1</sup>) and >11 levels of clover proportion (0–100%) were simulated. Fertilizer types included both organic fertilizer (OF) and mineral fertilizer (MF). All simulations were run at five sites G1 (USA), G2 (New Zealand), G3 (France), G4 (United Kingdom), and G5 (Switzerland) by the biogeochemical process models APSIM and DayCent, both in two variants (see text for details). In order to obtain the response variables on a homogeneous grid and exclude inter-annual variability, surfaces were fitted to model outputs (yield, N<sub>2</sub>O emissions) per model-site combination on the surface expanded by actual clover percentage and annual fertilizer amount.

cyclung, including key processes of N<sub>2</sub>O production, nitrification, and denitrification and, importantly, have been validated with field data at the study sites used with respect to N<sub>symb</sub> (Fitton et al., 2019) and combinations of biomass yields and N<sub>2</sub>O emissions (Ehrhardt et al., 2018; Fuchs et al., 2020). The number of deployed models in this current study was lower than those in the publication cited above because most of the biogeochemical models were not able to simulate grass-legume interaction explicitly, as required for this study.

We conducted a simulation study at five temperate grassland sites using the two biogeochemical models, each in two variants. Our specific objectives were (1) to systematically assess the effects of a wide range of legume proportions and N fertilizer amounts on yields, N<sub>2</sub>O emissions, and N<sub>2</sub>O emission intensities (N<sub>2</sub>O-Int) and (2) to identify the optimal N<sub>2</sub>O mitigation strategy.

## 2. Materials and Methods

A set of theoretical scenarios, that is, combinations of legume proportions (0–100%) and N fertilizer amounts (N<sub>fert</sub>; 13 levels 0–600 kg N ha<sup>-1</sup> yr<sup>-1</sup>) in a full factorial design, was run at five temperate grassland sites across the globe for a simulation period of 30 years (1981–2010) (Figure 2). Although the higher fertilizer rates are of little practical relevance, they are important for testing model responses and constraining statistically derived surfaces. Grass-clover mixtures of two species *Lolium perenne* and *Trifolium repens* were modeled as examples of species that are widespread across the temperate climatic zone. Mineral fertilizer in form of ammonium nitrate was simulated. The starting point for this study was the final stage of an international modeling intercomparison exercise initiated by the Global Research Alliance (Integrative Research Group) as described by Ehrhardt et al. (2018). For simplicity, and in order to exclude the confounding effects of grazing by further enhancing variability, we performed the study on non-grazed grassland systems. Mitigation of N<sub>2</sub>O emissions via changes in grazing practices (i.e., livestock density and N fertilizer inputs) have previously been investigated by Sándor et al. (2018).

### 2.1. Models Used in This Study

This exercise was performed by a consortium of modelers with the biogeochemical models APSIM (Holzworth et al., 2014; in two variants) and DayCent (Del Grosso et al., 2001; Parton et al., 1998; in two variants) according to an agreed protocol. Models were previously calibrated (corresponding to Stage 5 in

Ehrhardt et al., 2018) with respect to plant productivity, soil moisture and temperature, soil mineral N content, and model uncertainties of the respective models were evaluated in an international model intercomparison (Ehrhardt et al., 2018; Table S1) with further testing of yield and N<sub>2</sub>O emissions at two levels of pasture clover content in Fuchs et al. (2020). Biological nitrogen fixation and yields in grass-clover mixtures were validated at one site described by Fitton et al. (2019). The scenarios required modelers to manipulate the legume proportion of the grasslands. In a physical experiment, this would be achieved by sowing or over-sowing the desired species, with the ultimate legume proportion resulting from interactions between management and environmental conditions (see Fuchs et al., 2018). APSIM simulates grasslands in a way that is comparable with a physical experiment where the modeler can influence, but not directly control, the proportion of legumes (dynamic legume proportion), while in DayCent the legume proportion is explicitly specified by the modeler (Table S2; Fitton et al., 2019). Biological nitrogen fixation from grasslands with varying legume proportion can be simulated by APSIM directly as the dependency of BNF on species composition is represented in the model (Text S2). In contrast, DayCent requires the user to specify a potential maximum N<sub>symb</sub> that should, sensibly, vary with legume proportion. In Fitton et al. (2019), we developed a method to derive this DayCent parameter (potential maximum N<sub>symb</sub>) based on the legume percentage of the pasture. Note that DayCent internally reduces the actual N<sub>symb</sub> as fertilization increases. This method is described in Text S3 and was validated in Fitton et al. (2019).

#### 2.1.1. APSIM

APSIM (v7.10 r4162), short for Agricultural Production Systems sIMulator (Holzworth et al., 2014), is a process-oriented simulation framework comprising several modules. The model primarily operates on a daily time-step, but several individual modules work with shorter internal time discretization. APSIM was used for this study in two variants differing in their soil water modules, the SWIM module based on Richards' equation (here named AP1; Huth et al., 2012; SWIM for Soil and Water integrated Model), and the capacitance-based Soil Water module (here named AP2; Probert et al., 1998; further details can be seen in Text S2). Sward growth was simulated with the AgPasture module (F. Li et al., 2011) with recent additions to model reallocation of N reserves (Vogeler & Cichota, 2016) and an implementation of the Penman-Monteith equation designed for intermingled canopies (Snow & Huth, 2004). Soil organic matter and nitrogen transformations were simulated with the SoilN module (Probert et al., 1998) with a derivation of the N<sub>2</sub>:N<sub>2</sub>O ratio adapted from DayCent (Thorburn et al., 2010) and a simple routine to calculate volatilization losses (Vogeler et al., 2010). APSIM simulates N<sub>2</sub>O production from nitrification and denitrification. Nitrification is implemented with a Michaelis-Menten approach, with a fixed fraction of NH<sub>4</sub><sup>+</sup> for nitrification. N<sub>2</sub>O production via nitrification is down-regulated at suboptimal temperature, moisture, and pH in the respective layer. Denitrification depends on NO<sub>3</sub><sup>-</sup> and active C concentrations in the respective layer. The maximum denitrification rate is regulated downwards and depends on temperature and soil moisture in each layer (for further details on the simulation and parameterization of BNF and N<sub>2</sub>O emissions in APSIM, see Text S2). Simulated pasture composition in APSIM is dynamic and adapts to prevailing conditions, such as fertilizer inputs, much as one might expect a physical pasture to behave. To create variation in legume content for any given fertilizer regime, the simulated pastures were resown every third year, much as might be done in a physical experiment.

#### 2.1.2. DayCent

DayCent, the daily time-step model that has evolved from the Century model (Del Grosso et al., 2001; Parton et al., 1998), was used in this study in the variants DayCent v4.5 2010 (here named DC1) and DayCent v4.5 2013 (here named DC2). These variants differ in their calculations of solar radiation and in their calculations of maintenance and growth respiration, as well as the implementation of N<sub>2</sub>O emissions due to freeze-thaw cycles in the latter (please find further details in Text S3). DayCent includes four major sub-models: (1) the plant growth sub-model calculates biomass production and allocates net primary production (NPP) to various plant pools; (2) the soil organic matter (SOM) sub-model simulates decomposition of dead plant material (litter) and SOM and allocates soil carbon to three SOC pools and the litter pool; (3) the soil water sub-model simulates water flow between different layers; and (4) the trace gas flux sub-model simulates gaseous emissions. DayCent simulates N<sub>2</sub>O losses based on the leaky pipe metaphor where nitrification and denitrification are conceptualized as leaky pipes through which N is cycling but losing some gaseous N through a hole (Parton et al., 2001). Nitrification and denitrification rates depend on NH<sub>4</sub><sup>+</sup> (for nitrification) and NO<sub>3</sub><sup>-</sup> concentrations and labile carbon availability (for denitrification), soil water content and

temperature. While the proportion of the nitrification and denitrification escaping as gaseous N as either ( $\text{NO}_x$ ,  $\text{N}_2\text{O}$ , or  $\text{N}_2$ ) is fixed, the ratio of  $\text{N}_2\text{O}$  with respect to  $\text{NO}_x$  and  $\text{N}_2$  depends on soil water content and soil properties regulating soil gas diffusivity (Parton et al., 2001; see Text S3).

### 2.1.3. Empirical Findings for the Effect of $\text{N}_{\text{fert}}$ and Clover on Yields

In order to present model results in the context of empirical findings, we displayed example data from two studies for the dependency of yields on fertilizer N (Reid, 1983; here E1) and clover proportion (Nyfeler et al., 2009; here E2) in Figure 4. These can be regarded as case studies with the purpose of showing the reader what patterns we would expect. We used the regression equation from Reid (1983) as displayed in their Table 2. In contrast to our model outputs, grass-clover mixtures in Reid (1983) did not contain exactly 50% clover in dry biomass. Besides Reid (1983), see also further studies of Sparrow (1979), Prins (1983), and Whitehead (1995) for comparable data. To provide empirical evidence for the effect of clover percentage on yields, we used the equation from Nyfeler et al. (2009), equation 1 with the coefficients according to their Table S3 for mixtures of *T. repens* and *L. perenne*, excluding the other species. The Nyfeler study had  $50 \text{ kg N ha}^{-1} \text{ yr}^{-1}$  as the minimum fertilizer rate instead of  $0 \text{ kg N ha}^{-1} \text{ yr}^{-1}$  as in our study.

## 2.2. Site Information

The sites used for the simulations were five grassland sites in the temperate zone (Table S3). Daily weather data for 1980–2010 was supplied by A. Ruane from the AgMERRA data set (Ruane et al., 2015). Each modeling group used their plant growth model parameterizations, with the exceptions noted below, as developed in the fully informed or calibrated (Stage 5) modeling reported by Ehrhardt et al. (2018). The exception to the parameterization was in relation to the grass-legume ratio and symbiotic N fixation as described in the specific model sections in the supporting information.

## 2.3. Validation Results

For net primary production, the applied models in the validation exercise showed relative biases of  $-0.04$ – $0.74$  and RRMSE of  $0.70$ – $0.89$  (Ehrhardt et al., 2018). For  $\text{N}_2\text{O}$  emissions, relative biases were  $-0.80$ – $0.03$  and RRMSE ranged between  $0.13$  and  $0.81$  (Ehrhardt et al., 2018; see Table S1a). A detailed validation of  $\text{N}_2\text{O}$  emissions and its drivers at the Swiss site revealed that both models used here generally performed better than the Swiss IPCC calculations for that site (Fuchs et al., 2020).

The relationship between yields and legume percentage was validated in Fitton et al. (2019), showing that APSIM most accurately predicted yields at 0% and 100% clover (bias below 2%), but underestimated yields at 25–50% clover by 23–27% and overestimated yields at 75% clover by 10%. DayCent estimated yields with a bias below 10% across all clover percentages, except at 75% clover percentage where it overestimated yields by 33% (Fitton et al., 2019; see Table S1b).

## 2.4. Modeling Protocol

Management events included cutting and fertilization. Cutting events were simulated monthly during the growing season (May–October for the European sites G3–5, May–September for G1 in North Dakota, and September–May for G2 in New Zealand; each on the first day of the month) and followed by fertilization on the fifth day of the month (Table S4). In line with agricultural practices, the exception deviating from this scheme was that there was no fertilizer applied after the last harvest of the season and fertilization in spring started in the month before the first cut in each year (Table S4). The residual biomass after all cuts was  $1,200 \text{ kg DM ha}^{-1}$ , and all cut biomass was removed from the simulation. The total amount of N to be applied annually was split equally between the individual applications, which were surface applied. The mineral N was applied as ammonium-nitrate ( $\text{NH}_4\text{-NO}_3$ ). Aerial deposition of N was set to zero. The sites were presumed to be flat; hence, there were no N losses via run-off. Model outputs included yields, harvested C and N, plant N uptake,  $\text{N}_2\text{O}$  emissions, soil organic carbon (SOC), soil organic nitrogen (SON), nitrate and ammonium concentrations, and nitrate leaching (Table S5).

## 2.5. Statistical Analysis

The dynamic simulation of clover proportions in APSIM resulted in clover proportions that were irregularly distributed between 0% and 100% and therefore not suitable for direct analysis. From visual inspection, we observed no trends in the output data. In order to (1) obtain a response value on a regular grid for any clover-fertilizer combination per site, per model and per fertilizer type and to (2) exclude the inter-annual

variability and analyze an aggregated outcome across years, we fitted response surfaces on the area spanned by clover proportion and fertilizer amount (Figure 2; see also Figure S1 for an example surface fit). These smooth surfaces can be considered as an interpolation, aggregating the 30 years of model outputs. We used the tensor product smoother to include two dimensions (clover proportion and fertilizer amount), suitable due to its scale-invariant properties (Wood, 2006). We prescribed a fixed number of knots for both dimensions ( $k = 4$ , resulting in  $4^2 = 16$  knots) in order to hold the number of parameters consistent across all surface fits. We had also investigated surface responses using a linear model with second order polynomials for both predictors and an interaction. However, residual analysis of the results from that approach showed that the polynomial fits violated the assumptions of the linear model (i.e., residual distribution showed no normality and homogeneous variance) in most site-model combinations. Thus, from visual inspection and residual analysis, the chosen method using the tensor product smoother better represented the characteristics of the data and was therefore chosen for this study. Daily model outputs were aggregated to annual values before the fitting. The surface fitting and statistical analyses were carried out using the open source software R (R Core Team, 2016) and specifically the “mgcv” package (Wood, 2011) for fitting smoothing splines, building the tensor product of the predictors ( $N_{\text{fert}}$  and clover %) with the function `te()`, and fitting the smoothing splines on this tensor product with `gam()` (Wood, 2011). Further, the “data.table” package (Dowle & Srinivasan, 2017) was used for efficiently handling the large data sets. ANOVA was used in order to systematically assess the effects of clover proportion, fertilizer amount, model and site on yields,  $N_{\text{symp}}$ ,  $N_2O$  emissions, and  $N_2O$  emission intensities ( $N_2O\text{-Int}$  calculated as mass unit of  $N_2O$  emissions per mass unit biomass yield). For the statistical analysis, the post-processed data set (after surface fitting) was used as this had the advantage of a balanced design. The ANOVA was performed separately for APSIM and DayCent, since the patterns varied between models such that averaging across model types would make it impossible to address model-specific behaviors and their implications on the results. To compare the effect of clover proportion and fertilizer amount on a response variable (e.g., yield,  $N_2O$  emissions, and  $N_2O\text{-Int}$ ) across several sites, the response variable was presented as standardized response. Here, the standardization was consistently defined as the response variable divided by the maximum of the respective response variable across all scenarios (combinations of clover proportion and fertilizer amount) per site and model.

$$Y'_{i,m,s} = \frac{Y_{i,m,s}}{\max \{Y_{i,m,s}\}}$$

$Y'_{i,m,s}$  is the standardized response variable for scenario  $i$ , fixed model  $m$ , and fixed site  $s$ , and  $Y_{i,m,s}$  is the original response variable. The results are presented per model (DayCent, APSIM) across model variants in order to assess the general pattern rather than the individual model variants' differences. Mean values and ranges given in the text and figures refer to the mean and range across variants and not the inter-annual which was excluded by the surface fitting procedure.

A baseline scenario of no clover with  $200 \text{ kg N ha}^{-1} \text{ yr}^{-1}$  fertilizer applied was used for comparing mitigation scenarios. This baseline reasonably represents intensively used cut-and-carry systems, corresponding to N recommendations for fertilized grassland present in several countries. To test if a small clover percentage would significantly change the results under the baseline, we regarded clover percentages of 5% as a second baseline scenario. Here, we were not trying to mimic the actual management at the sites nor intending to address site-specific optima (species, N rates, grazing). Instead, we conceptually prescribed fixed management schedules in order to look at the response surfaces of fixed comparable setups across sites. While corresponding approximately to actual fertilizer amounts at the G5 site, which was the highest input system of any of the sites in this study, this baseline deviated from the actual fertilizer amendments on the other sites (Table S3). Nevertheless, we used this as baseline because we were interested in the general patterns using the sites as “generalizable” examples.

## 2.6. Estimating the Global Technical Mitigation Potential of Replacing $N_{\text{fert}}$ With BNF

Our work focuses on managed pasturelands with C3 grasses and is not applicable to extensive rangeland and C4 pastureland. Dangal et al. (2019) estimated that pastures based on C3 grasses cover 570 million ha of the terrestrial surface, based on the HYDE3.2 data set by Klein Goldewijk et al. (2017). We used the estimated global amount of  $2.9 \text{ Tg N}_2\text{O-N yr}^{-1}$  for the recent period 2001–2014 together with their estimate of 11% of these emissions being attributed to synthetic fertilizer, resulting in emissions of  $319 \text{ Gg Tg N}_2\text{O-N yr}^{-1}$



**Table 1**  
Effects of Site, Model, Fertilizer Amount, and Clover Percentage on Yield, N<sub>2</sub>O Emission, and N<sub>2</sub>O-Int

		Yield t DM ha <sup>-1</sup> yr <sup>-1</sup>		N <sub>2</sub> O emission kg N <sub>2</sub> O-N ha <sup>-1</sup> yr <sup>-1</sup>		N <sub>2</sub> O-Int g N <sub>2</sub> O-N kg <sup>-1</sup> DM <sup>-1</sup>	
		APSIM	DayCent	APSIM	DayCent	APSIM	DayCent
Site	Grand mean	10.3	6.8	0.52	5.31	0.04	0.83
		***	***	***	***	***	***
	G1	8.3 <sup>a</sup>	5.0 <sup>a</sup>	0.38 <sup>b</sup>	2.53 <sup>a</sup>	0.034 <sup>a</sup>	0.595 <sup>a</sup>
	G2	11.9 <sup>b</sup>	7.3 <sup>b</sup>	0.53 <sup>cd</sup>	3.57 <sup>b</sup>	0.034 <sup>a</sup>	0.494 <sup>a</sup>
	G3	10.7 <sup>c</sup>	6.7 <sup>bc</sup>	0.43 <sup>a</sup>	7.11 <sup>c</sup>	0.028 <sup>a</sup>	1.071 <sup>b</sup>
	G4	9.2 <sup>b</sup>	6.3 <sup>c</sup>	0.61 <sup>de</sup>	6.32 <sup>c</sup>	0.054 <sup>b</sup>	1.015 <sup>b</sup>
Model	G5	11.3 <sup>bc</sup>	8.8 <sup>d</sup>	0.78 <sup>e</sup>	10.4 <sup>d</sup>	0.059 <sup>b</sup>	1.231 <sup>b</sup>
		***	***	ns	***	***	***
	AP1	10.8 <sup>a</sup>	-	0.55 <sup>a</sup>	-	0.037 <sup>a</sup>	-
	AP2	9.8 <sup>b</sup>	-	0.51 <sup>a</sup>	-	0.043 <sup>b</sup>	-
	DC1	-	5.8 <sup>a</sup>	-	4.91 <sup>a</sup>	-	0.917 <sup>a</sup>
N <sub>fert</sub>	DC2	-	7.8 <sup>b</sup>	-	5.75 <sup>b</sup>	-	0.751 <sup>b</sup>
		***	***	***	***	***	***
	0	5.3 <sup>a</sup>	5.8 <sup>a</sup>	0.01 <sup>a</sup>	1.83 <sup>a</sup>	0.006 <sup>a</sup>	0.341 <sup>a</sup>
	150	9.1 <sup>b</sup>	6.7 <sup>b</sup>	0.31 <sup>b</sup>	4.35 <sup>b</sup>	0.025 <sup>b</sup>	0.681 <sup>a</sup>
	300	11.3 <sup>c</sup>	7.0 <sup>b</sup>	0.66 <sup>c</sup>	6.59 <sup>c</sup>	0.054 <sup>c</sup>	0.989 <sup>b</sup>
	450	12.4 <sup>cd</sup>	7.1 <sup>b</sup>	1.21 <sup>d</sup>	8.37 <sup>cd</sup>	0.096 <sup>d</sup>	1.232 <sup>bc</sup>
Clover %	600	13.3 <sup>d</sup>	7.3 <sup>b</sup>	1.90 <sup>e</sup>	9.62 <sup>d</sup>	0.145 <sup>e</sup>	1.392 <sup>c</sup>
		***	***	***	ns	***	ns
	0	8.7 <sup>a</sup>	6.1 <sup>a</sup>	0.50 <sup>a</sup>	4.80 <sup>a</sup>	0.044 <sup>a</sup>	0.886 <sup>a</sup>
	25	9.5 <sup>a</sup>	7.1 <sup>bc</sup>	0.48 <sup>a</sup>	5.31 <sup>a</sup>	0.040 <sup>ab</sup>	0.785 <sup>a</sup>
	50	10.7 <sup>b</sup>	7.3 <sup>c</sup>	0.46 <sup>a</sup>	5.51 <sup>a</sup>	0.034 <sup>b</sup>	0.769 <sup>a</sup>
	75	11.8 <sup>b</sup>	6.9 <sup>bc</sup>	0.54 <sup>ab</sup>	5.51 <sup>a</sup>	0.034 <sup>b</sup>	0.823 <sup>a</sup>
	10.8 <sup>b</sup>	6.5 <sup>ab</sup>	0.68 <sup>b</sup>	5.46 <sup>a</sup>	0.050 <sup>ab</sup>	0.884 <sup>a</sup>	

Note. Predicted values per model (APSIM, DayCent) are shown for mineral fertilizer scenarios (see Table S7 including organic fertilization). Different small letters indicate significant differences among levels (Tukey's HSD) per factor (Site, Model, N<sub>fert</sub>, Clover %). ns indicates no significant differences for the respective factor.

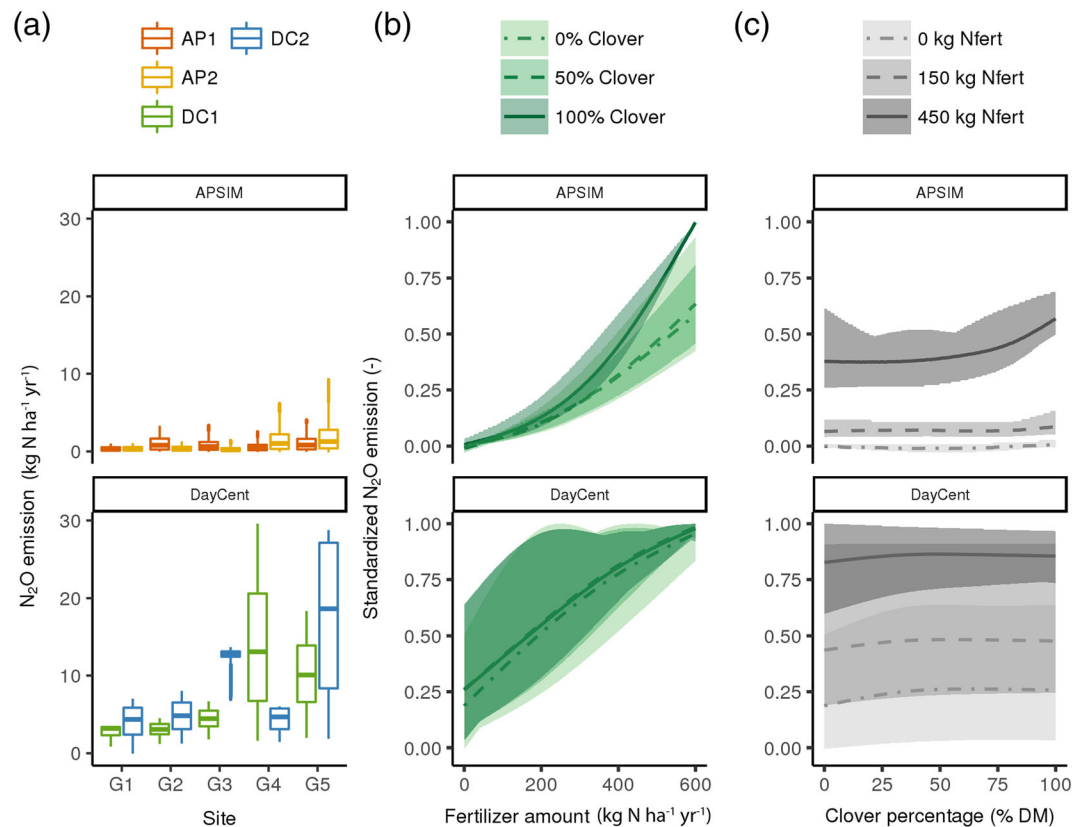
\*\*\*Significant effect at a significance level of  $p < 0.01$ .

attributed to fertilizer application during this period. Based on this estimate of N<sub>2</sub>O emissions from global fertilizer application on grasslands, we used our estimated relative values for N<sub>2</sub>O mitigation averaged across sites to calculate the global technical mitigation potential of replacing N<sub>fert</sub> with BNF under the baseline and different mitigation scenarios. For the global estimate, the mitigation effect was considered relative to the N<sub>2</sub>O emissions attributed to fertilizer use, which are the site emissions at the fertilizer scenario minus the background N<sub>2</sub>O emissions, that is, N<sub>2</sub>O emissions under the scenario of no fertilizer nor symbiotically fixed N input.

### 3. Results

#### 3.1. Nitrous Oxide Emissions

Nitrous oxide emissions ranged between 0.01 and 1.9 kg N<sub>2</sub>O-N ha<sup>-1</sup> yr<sup>-1</sup> (0.1–0.4% of N<sub>fert</sub>) across scenarios for APSIM variants and between 1.8 and 9.6 kg N<sub>2</sub>O-N ha<sup>-1</sup> yr<sup>-1</sup> (1.4–5.5% of N<sub>fert</sub>) for DayCent model variants averaged across sites (Table 1, Figure 3). The emissions varied significantly among models, sites, and fertilizer levels (Tables 1 and S4). Clover proportion did not affect N<sub>2</sub>O emissions in DayCent (Table 1) and did not affect N<sub>2</sub>O emissions in APSIM at 0–75% clover but significantly increased N<sub>2</sub>O emissions at the 75–100% clover level (Table 1) due to increased N<sub>2</sub>O emissions at high fertilizer levels of 450–600 kg N ha<sup>-1</sup> yr<sup>-1</sup> in clover-rich grasslands (Table 2, Figure 3). The N<sub>2</sub>O emissions depended not only on the amount of N input but also on the N source. For the same total N input (here defined as N<sub>fert</sub> + N<sub>symb</sub>), lower N<sub>2</sub>O emissions were found when a larger fraction of the N input came from symbiotic N fixation (as N<sub>symb</sub>) rather than from fertilization (N<sub>fert</sub>). The mitigation option would be ineffective if N<sub>2</sub>O emissions were independent of the source of N; thus, this finding shows that an important precondition for effectiveness of the mitigation strategy was met. The N<sub>2</sub>O emissions were highest at maximal fertilizer levels and



**Figure 3.** (a) Boxplots of  $\text{N}_2\text{O}$  emissions across all mineral N scenarios (clover-fertilization combinations) per site (G1–G5) for APSIM (top) and DayCent variants (bottom); (b) dependency of standardized  $\text{N}_2\text{O}$  emissions (y-axis) on fertilizer amount (x-axis) at 0% clover percentage (dot-dashed) 50% clover percentage (dashed) and 100% clover percentage (solid line); and (c) dependency of standardized  $\text{N}_2\text{O}$  emissions (y-axis) on clover percentage (x-axis) for the fertilizer levels 0 kg N (dot-dashed), 150 kg N (dashed), and 450 kg N (solid line) (APSIM: top, DayCent: bottom).  $\text{N}_2\text{O}$  emissions in b and c were standardized with the maximum yields per model-site combination. Lines display the mean across sites, shaded areas indicate the range across sites per model family.

were generally smallest in the absence of fertilizer. While  $\text{N}_2\text{O}$  emissions were higher for the DayCent variants, APSIM variants predicted higher losses via nitrate leaching (Figure S4).

### 3.1.1. Effect of Fertilization on Nitrous Oxide Emissions

In the absence of fertilization, models simulated average  $\text{N}_2\text{O}$  emissions of  $0.01 \text{ kg N}_2\text{O-N ha}^{-1} \text{ yr}^{-1}$  (APSIM) and  $1.83 \text{ kg N}_2\text{O-N ha}^{-1} \text{ yr}^{-1}$  (DayCent) across clover proportions (Table 1). At  $150 \text{ kg N}$  fertilization levels,  $\text{N}_2\text{O}$  emissions increased to  $0.31 \text{ kg N}_2\text{O-N ha}^{-1} \text{ yr}^{-1}$  (APSIM) and  $4.35 \text{ kg N}_2\text{O-N ha}^{-1} \text{ yr}^{-1}$  (DayCent) (Table 2). At maximal fertilization levels ( $600 \text{ kg N}$ ),  $\text{N}_2\text{O}$  emissions reached  $1.9 \text{ kg N}_2\text{O-N ha}^{-1} \text{ yr}^{-1}$  (APSIM) and  $9.6 \text{ kg N}_2\text{O-N ha}^{-1} \text{ yr}^{-1}$  (DayCent), averaged across clover scenarios, model variants and sites (Table 1). Fertilization increased  $\text{N}_2\text{O}$  emissions for all model-site-combinations. However, different types of relationships were observed. The APSIM models showed an almost exponential increase of  $\text{N}_2\text{O}$  with fertilization, (Figure 3), while the DayCent models simulated approximately linear relationships, and in some cases (e.g., G3) lower  $\text{N}_2\text{O}$  emission increments at higher fertilizer rates (leveling-off) (Figure 3).

### 3.1.2. Effect of Clover Proportion on Nitrous Oxide Emissions

In general, clover proportion did not affect  $\text{N}_2\text{O}$  emissions across models and sites (Figure 3, Tables 1 and 2) between 0% and 75% clover (Table 1). As an exception, APSIM simulated significantly increased  $\text{N}_2\text{O}$  emissions at a clover proportion of 100% in the higher fertilized scenarios ( $\geq 450 \text{ kg N}_{\text{fert}}$ ), depicting the effects of high N inputs in these scenarios. However, one has to keep in mind that those clover-fertilizer combinations are rather artificial and of no practical importance. A notable characteristic was that DayCent simulated background  $\text{N}_2\text{O}$  emissions of  $1.8 (0.4\text{--}4.3) \text{ kg N}_2\text{O-N ha}^{-1} \text{ yr}^{-1}$ , without N fertilization or clover in the field. APSIM showed no  $\text{N}_2\text{O}$  fluxes without N fertilization independent of clover level.

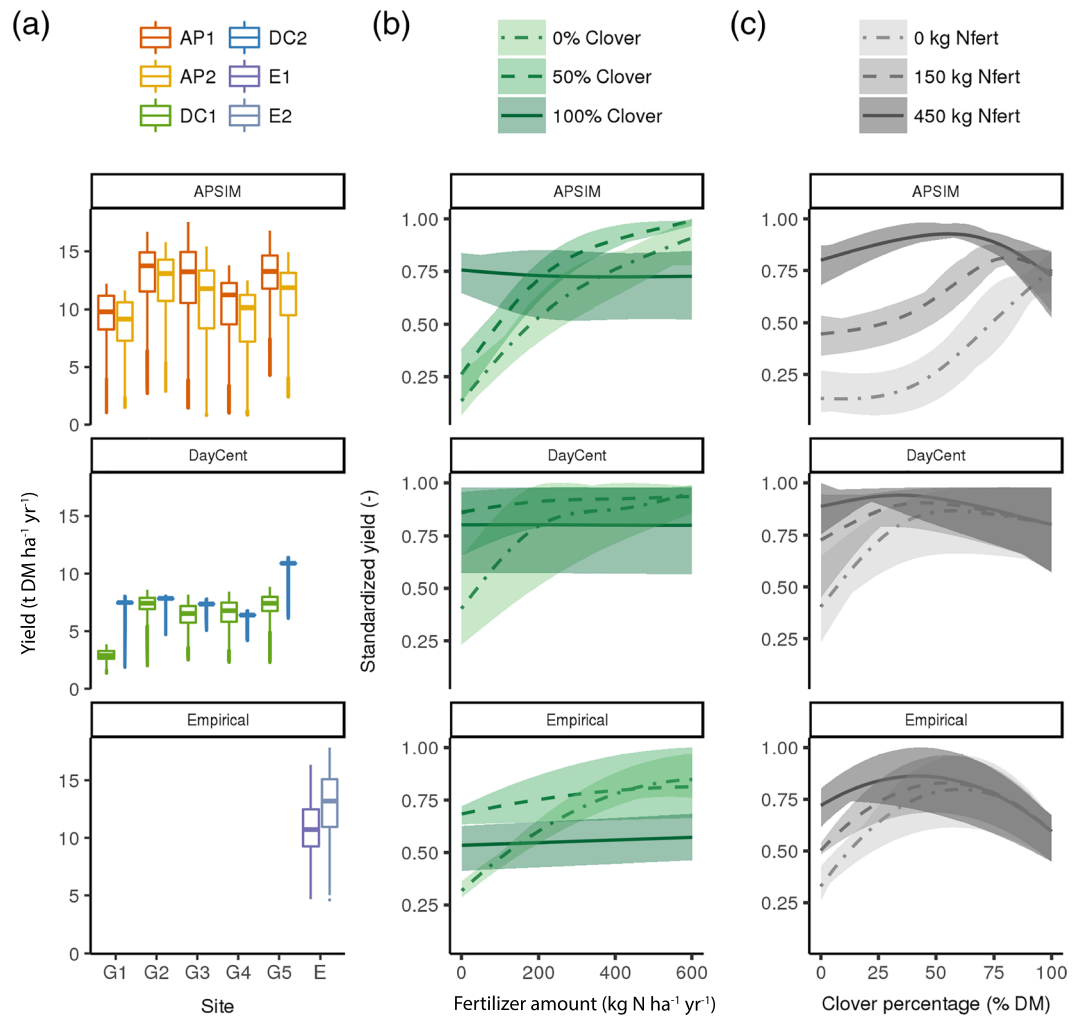
**Table 2**  
Average and Range of Yields, N<sub>2</sub>O Emission, and N<sub>2</sub>O-Int for Different Combinations of N<sub>fert</sub> and Clover Proportion Across Sites and Model Variants for APSIM and DayCent for Mineral Fertilizer N Scenarios

	N 0	N 150	N 300	N 450	N 600
APSIM Yield (t DM ha <sup>-1</sup> yr <sup>-1</sup> )					
0% clover	2.0 (0.9–3.6)	6.5 (5.4–7.6)	9.7 (8.7–11.2)	11.7 (9.7–13.8)	13.4 (9.7–16.0)
25% clover	2.2 (1.2–3.4)	7.2 (6.1–8.7)	10.8 (9.4–12.1)	12.9 (10.9–14.1)	14.2 (11.2–15.8)
50% clover	3.9 (2.6–5.3)	9.1 (7.5–10.9)	12.2 (10.6–13.7)	13.6 (11.6–15.0)	14.5 (16.4–15.7)
75% clover	7.5 (4.9–9.9)	11.8 (10.1–13.4)	13.1 (10.9–14.8)	13.1 (10.5–14.6)	13.6 (10.5–15.7)
100% clover	11.1 (8.5–13.3)	10.9 (7.3–13)	10.7 (7.0–12.9)	10.7 (7.0–12.9)	10.7 (7.0–12.8)
DayCent Yield (t DM ha <sup>-1</sup> yr <sup>-1</sup> )					
0% clover	3.2 (1.6–4.2)	5.8 (4.9–7.3)	6.8 (5.8–8.6)	7.0 (5.5–8.9)	7.6 (5.8–9.7)
25% clover	5.9 (3.7–7.6)	7.0 (5.2–8.9)	7.4 (5.6–9.4)	7.5 (5.4–9.5)	7.7 (5.5–9.9)
50% clover	7.0 (4.8–9.1)	7.3 (5.2–4.9)	7.4 (5.3–9.5)	7.5 (5.3–9.6)	7.5 (5.3–9.7)
75% clover	6.9 (5.0–9.0)	6.9 (5.1–9.0)	7.0 (5.1–0.1)	7.0 (5.1–9.1)	7.1 (5.1–9.2)
100% clover	6.5 (4.6–8.6)	6.5 (4.9–8.6)	6.5 (4.9–8.6)	6.5 (4.9–8.5)	6.5 (4.9–8.5)
APSIM N <sub>2</sub> O emission (kg N <sub>2</sub> O-N ha <sup>-1</sup> yr <sup>-1</sup> )					
0% clover	0 (0–0)	0.2 (0.1–0.3)	0.6 (0.3–1.0)	1.1 (0.6–1.9)	1.7 (0.9–3.1)
25% clover	0 (0–0)	0.2 (0.10.3)	0.6 (0.3–1.0)	1.1 (0.5–2.0)	1.8 (0.8–3.2)
50% clover	0 (0–0)	0.2 (0.1–0.3)	0.6 (0.2–1.0)	1.2 (0.5–2.1)	2.0 (0.8–3.5)
75% clover	0 (0–0)	0.2 (0.1–0.3)	0.6 (0.3–1.1)	1.3 (0.6–2.4)	2.3 (1.0–4.3)
100% clover	0 (0–0)	0.2 (0.1–0.4)	0.8 (0.4–1.5)	1.8 (0.7–3.6)	3.4 (1.0–6.7)
DayCent N <sub>2</sub> O emission (kg N <sub>2</sub> O-N ha <sup>-1</sup> yr <sup>-1</sup> )					
0% clover	1.9 (0.4–4.3)	4.6 (2.3–7.6)	7.5 (3.7–13.5)	10.0 (4.7–19.8)	11.9 (5.1–22.9)
25% clover	2.4 (0.5–5.3)	5.0 (2.3–7.9)	7.9 (3.8–14.4)	10.4 (4.6–20.6)	12.2 (5.1–23.4)
50% clover	2.6 (0.6–5.6)	5.2 (2.4–8.0)	8.1 (3.8–14.6)	10.6 (4.6–20.7)	12.4 (5.1–23.5)
75% clover	2.6 (0.6–5.6)	5.2 (2.4–8.0)	8.0 (3.7–14.4)	10.6 (4.6–20.5)	12.4 (5.1–23.3)
100% clover	2.6 (0.6–5.6)	5.1 (2.3–7.9)	8.0 (3.7–14.3)	10.6 (4.6–20.3)	12.4 (5.1–23.0)
APSIM N <sub>2</sub> O-Int (g N <sub>2</sub> O-N kg <sup>-1</sup> DM <sup>-1</sup> )					
0% clover	0.010 (0.003–0.034)	0.028 (0.019–0.043)	0.059 (0.040–0.096)	0.097 (0.061–0.159)	0.134 (0.079–0.220)
25% clover	0.006 (0.001–0.018)	0.027 (0.018–0.04)	0.057 (0.033–0.090)	0.093 (0.051–0.149)	0.131 (0.069–0.210)
50% clover	0.002 (0.000–0.006)	0.023 (0.013–0.034)	0.052 (0.025–0.082)	0.092 (0.045–0.149)	0.140 (0.073–0.228)
75% clover	0.001 (0.001–0.002)	0.017 (0.010–0.026)	0.049 (0.03–0.079)	0.105 (0.063–0.177)	0.179 (0.108–0.313)
100% clover	0.001 (0.001–0.006)	0.025 (0.016–0.037)	0.075 (0.048–0.128)	0.172 (0.108–0.315)	0.311 (0.166–0.585)
DayCent N <sub>2</sub> O-Int (g N <sub>2</sub> O-N kg <sup>-1</sup> DM <sup>-1</sup> )					
0% clover	0.63 (0.41–1.05)	0.82 (0.45–1.18)	1.09 (0.55–1.57)	1.34 (0.70–2.12)	1.49 (0.80–2.35)
25% clover	0.42 (0.26–0.84)	0.73 (0.38–1.11)	1.05 (0.52–1.51)	1.33 (0.66–2.09)	1.50 (0.78–2.34)
50% clover	0.37 (0.22–0.78)	0.72 (0.36–1.11)	1.07 (0.51–1.50)	1.36 (0.65–2.10)	1.57 (0.79–2.39)
75% clover	0.41 (0.25–0.82)	0.78 (0.38–1.15)	1.16 (0.52–1.56)	1.49 (0.67–2.19)	1.73 (0.81–2.77)
100% clover	0.41 (0.25–0.84)	0.84 (0.38–1.20)	1.29 (0.54–1.83)	1.68 (0.70–2.75)	1.99 (0.85–3.57)
Empirically estimated N <sub>2</sub> O emission <sup>a</sup> (kg N <sub>2</sub> O-N ha <sup>-1</sup> yr <sup>-1</sup> )					
IPCC	0.0 (0.0–0.0)	1.5 (0.5–4.5)	3.0 (0.9–9.0)	4.5 (1.4–13.5)	6.0 (1.8–18.0)
Shcherbak et al. (2014) <sup>b</sup>	0.0	2.0	5.5	10.6 <sup>c</sup>	17.4 <sup>c</sup>

Note. For comparison, we include the IPCC N<sub>2</sub>O emission estimate with its 95% confidence intervals as well as the empirical estimates by Shcherbak et al. (2014). <sup>a</sup>For models empirically estimating N<sub>2</sub>O emission, no distinction between different clover proportions is available. <sup>b</sup>Empirical model for “perennial grass/forage” in Table S3 of Shcherbak et al. (2014); emission of N<sub>2</sub>O-N = 1.067 a (7.76 a 0.0353\*N<sub>fert</sub>)\*N<sub>fert</sub>)\*10<sup>-3</sup>. <sup>c</sup>Even though Shcherbak et al. (2014) displayed N<sub>2</sub>O emission estimates for only N up to 300 kg N in their paper, the N fertilizer amounts in grasslands used for deriving their empirical relationship covered the higher levels (see supplementary information of Shcherbak et al., 2014), and we thus included levels up to 600 kg in our table.

### 3.2. Sward Productivity

Annual yields of the grass-clover mixtures ranged between 0.9 and 15.7 t DM ha<sup>-1</sup> yr<sup>-1</sup> across models and sites, clover, and fertilizer combinations (Figure 4a, Table 2). For APSIM, average yields across scenarios (i.e., variations in fertilizer inputs and clover proportion) ranged between 8.3 t DM ha<sup>-1</sup> yr<sup>-1</sup> at Site G1 and 11.9 t DM ha<sup>-1</sup> yr<sup>-1</sup> at Site G2 (Table 1). DayCent simulated lower average yields, ranging between 5.0 t DM ha<sup>-1</sup> yr<sup>-1</sup> (G1) and 8.8 t DM ha<sup>-1</sup> yr<sup>-1</sup> (G5) across scenarios (Table 1). Variability between scenarios (within model) was higher for APSIM (mean sum of squares MSQ<sub>Nfert</sub> = 890; MSQ<sub>Clover%</sub> = 238.4) compared to DayCent (MSQ<sub>Nfert</sub> = 24.6, MSQ<sub>Clover%</sub> = 24.6; Table S6). The models agreed that highest productivity levels were simulated at G2 and G5, medium productivity at G3, while lower productivity was simulated for G4 and G1 (Figure 4a, Table 1).



**Figure 4.** (a) Boxplots of yields across all mineral N scenarios (clover-fertilization combinations) per site (G1–G5) for APSIM (top) and DayCent variants (middle); (b) dependency of standardized yield (vertical axis) on fertilizer amount (horizontal axis) at 0% clover percentage (dot-dashed), 50% clover percentage (dashed), and 100% clover percentage (solid line); and (c) dependency of standardized yield (vertical axis) on clover percentage (horizontal axis) for the fertilizer levels 0 kg N (dot-dashed), 150 kg N (dashed), and 450 kg N (solid line) (APSIM: top, DayCent: middle). Yields in b and c were standardized with the maximum yields per model-site combination. Lines display the mean across sites, and shaded areas indicate the range across sites per model family. The bottom panels display conceptually our expectations from empirical findings. For the effect of fertilization (b), we displayed the empirical-derived regression equation from Reid (1983) (E1) and refer to both equation and coefficients in Table 2 in their publication. In contrast to our model outputs, grass-clover mixtures in Reid (1983) did not contain exactly 50% clover in dry biomass. For the effect of clover percentage (c), we used the empirical-derived equation from Nyfeler et al. (2009) (E2), equation 1 with the coefficients according to their Table S3 for mixtures of *T. repens* and *L. perenne*, excluding the other species. The Nyfeler study had 50 kg N as the minimum fertilizer rate, displayed dot-dashed, not 0 kg N. For the empirical studies in the bottom panel, the line indicates the mean and the shaded area the range across years.

### 3.2.1. Effects of Fertilizer Amount on Yield

Yields of grass monocultures significantly increased with the amount of fertilizer added. Average yields were lowest for scenarios without fertilizer (5.3 t DM ha<sup>-1</sup> yr<sup>-1</sup> in APSIM and 5.8 t DM ha<sup>-1</sup> yr<sup>-1</sup> in DayCent and up to 13.3 t DM ha<sup>-1</sup> yr<sup>-1</sup> in APSIM and 7.3 t DM ha<sup>-1</sup> yr<sup>-1</sup> in DayCent at 600 kg N input when averaged across sites and clover proportions; Table 1). While APSIM suggested that increased fertilization still increased yields at the highest fertilizer levels (i.e., significant yield differences between 450 and 600 kg N<sub>fert</sub>), DayCent simulated yields leveling off above ~200 kg N fertilization (i.e., no significant yield increase at the fertilizer levels ≥300 kg N compared to the 150 kg N; Table 1, Figure 4b). In unfertilized grass monocultures, APSIM simulated yields of 2.0 t DM ha<sup>-1</sup> yr<sup>-1</sup> (0.9–3.6 t DM ha<sup>-1</sup> yr<sup>-1</sup> across sites, Table 2)

**Table 3**  
Average and Standard Deviation Across Sites (sd) of  $N_{\text{symb}}$  for Different Fertilizer Rates and Clover Percentages as Simulated by APSIM and DayCent

$N_{\text{fert}}$ ( $\text{kg ha}^{-1} \text{ yr}^{-1}$ )	Clover percentage (%)	APSIM		DayCent	
		$N_{\text{symb}}$ ( $\text{kg ha}^{-1} \text{ yr}^{-1}$ )	sd $N_{\text{symb}}$ ( $\text{kg ha}^{-1} \text{ yr}^{-1}$ )	$N_{\text{symb}}$ ( $\text{kg ha}^{-1} \text{ yr}^{-1}$ )	sd $N_{\text{symb}}$ ( $\text{kg ha}^{-1} \text{ yr}^{-1}$ )
0	0	0	0	0	0
0	20	9	10	211	78
0	50	96	27	377	139
0	70	231	66	422	153
0	100	465	65	420	153
150	0	0	0	0	0
150	20	48	11	190	87
150	50	168	34	309	143
150	70	285	50	331	157
150	100	316	63	306	158
300	0	0	0	0	0
300	20	66	16	152	87
300	50	174	40	234	137
300	70	241	52	244	149
300	100	203	49	211	145

corresponding to 13% (7–21%) of the maximum site yields. In contrast, DayCent simulated higher yields of  $3.2 \text{ t DM ha}^{-1} \text{ yr}^{-1}$  ( $1.6\text{--}4.2 \text{ t DM ha}^{-1} \text{ yr}^{-1}$ ) or 36% (20–46%) of maximum site yields without any fertilization (reflected in the offset in Figure 4b).

The effect of fertilizer on yields depended on the clover proportion in the sward. In APSIM, a fertilizer response was found for grass-clover mixtures of 0–80% clover; for example, grass-clover mixtures of 50% in APSIM largely benefited from fertilization and increased yields from  $3.9 \text{ t DM ha}^{-1} \text{ yr}^{-1}$  without fertilization to  $12.2 \text{ t DM ha}^{-1} \text{ yr}^{-1}$  at  $300 \text{ kg N}$  fertilization, while high-clover mixtures  $\geq 75\%$  and clover monocultures did not benefit from fertilization above  $300 \text{ kg N}$  (Figure 4b, Table 2). DayCent showed a fertilizer response for mixtures of 25% clover, but little fertilization effect on mixtures of  $\geq 50\%$  clover (Figure 4b, Table 2), for example, for 50% clover, an increase in yield from  $7.0 \text{ t DM ha}^{-1} \text{ yr}^{-1}$  without fertilization to  $7.4 \text{ t DM ha}^{-1} \text{ yr}^{-1}$  at  $300 \text{ kg N}$  fertilization (Figure 4b, Table 2). Legume monocultures produced similar yields regardless of fertilizer level, consistently for all models (n.s. in Table 1), while the absolute amount of legume monoculture yields varied largely across models and sites. For example, APSIM predicted an average yield of  $10.7 \text{ t DM ha}^{-1} \text{ yr}^{-1}$  ranging from  $7.0\text{--}12.8 \text{ t DM ha}^{-1} \text{ yr}^{-1}$  across sites and DayCent  $6.5 \text{ t DM ha}^{-1} \text{ yr}^{-1}$  with sites ranging from  $4.9\text{--}8.5 \text{ t DM ha}^{-1} \text{ yr}^{-1}$  (Table 2).

### 3.2.2. Effect of Clover Proportion on Yields

Species mixtures were significantly more productive than grass monocultures in unfertilized swards (Table 2, Figure 4c). Transgressive over-yielding, that is, higher yields for mixtures than for the highest-yielding monoculture, were observed for DC1 and for APSIM (both variants) above  $50 \text{ kg N}$  fertilization. Without fertilization, maximum yields were found at 100% clover. When increasing fertilizer levels, the clover proportion for maximum production (“optimal clover proportion”— $\text{Clo}_{\text{opt}}$ ) shifted to lower clover proportions. For example,  $\text{Clo}_{\text{opt}}$  in APSIM was 82% (site-median) at  $150 \text{ kg N}$ , decreasing to 47% at  $450 \text{ kg N}$ , and in DayCent  $\text{Clo}_{\text{opt}}$  decreased from 58% at  $150 \text{ kg N}$  to 43% at  $450 \text{ kg N}$ .

### 3.3. Symbiotically Fixed Nitrogen

Symbiotically fixed nitrogen reached up to  $465 (\pm 65) \text{ kg N ha}^{-1} \text{ yr}^{-1}$  and  $422 (\pm 153) \text{ kg N ha}^{-1} \text{ yr}^{-1}$  at the scenarios without fertilizer inputs for APSIM and DayCent simulations, respectively (Table 3). With increasing clover proportion,  $N_{\text{symb}}$  increased up to a maximum and decreased at higher clover proportions. This pattern resulted from high productivity in mixtures due to the complementarity between grass and clover growth, which becomes less beneficial at higher clover proportions, resulting in smaller total growth and in less  $N_{\text{symb}}$ .

Consistent with the experimental findings of Nyfeler et al. (2009), the higher the amount of fertilizer, the lower the clover proportion at maximum  $N_{\text{symb}}$ . For example, in APSIM, the maximum  $N_{\text{symb}}$  at 150 kg  $N_{\text{fert}}$  was found at 86% clover (median across sites), while at fertilizer levels of 300 kg N, highest amounts of  $N_{\text{symb}}$  were found at 79% clover proportion. In DayCent, the maximum  $N_{\text{symb}}$  was found at slightly lower clover proportions around 70% clover for 150 kg  $N_{\text{fert}}$  and at 300 kg  $N_{\text{fert}}$ .

Fertilization increased  $N_{\text{symb}}$  for clover mixtures (below 70–80% clover) at moderate fertilizer levels as N addition stimulated growth. For example, grass-clover mixtures of 50% clover showed maximum fixation at around 150 kg N  $\text{ha}^{-1} \text{yr}^{-1}$  fertilization, while  $N_{\text{symb}}$  decreased with further fertilizer increment (Table 3).  $N_{\text{symb}}$  in clover monocultures decreased with increasing fertilization, as fertilized clover swards acquire their N from fertilization if available (Table 3).

#### 3.4. Nitrous Oxide Emission Intensities

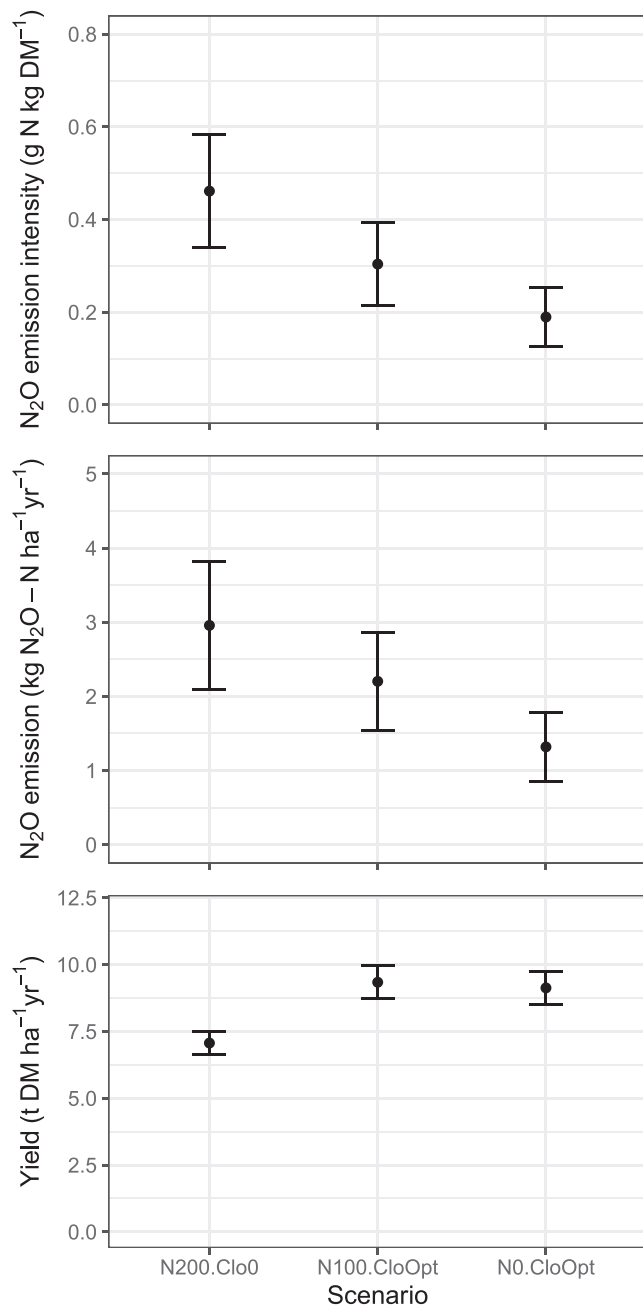
Nitrous oxide emission intensity estimates ( $N_2O\text{-Int}$ ), that is,  $N_2O$  emissions per unit of biomass yield, allow a productivity-based comparison of  $N_2O$  emissions across sites, scenarios, and models.  $N_2O\text{-Int}$  ranged between 0.01 and 0.31 g  $N_2O\text{-N kg}^{-1} \text{DM}$  (APSIM) and 0.63–1.99 g  $N_2O\text{-N kg}^{-1} \text{DM}$  (DayCent) across clover proportions and fertilizer levels on average across sites (Table 2; Figure S3). Maximum  $N_2O\text{-Int}$  were reached at 100% clover and the maximal fertilizer rate (600 kg  $N_{\text{fert}}$ ), except for DayCent at the G3 and G1 sites, where the maximum was already reached at 440 and 260 kg  $N_{\text{fert}}$  due to the leveling-off in  $N_2O$  emissions at higher fertilizer levels. Without fertilizer input,  $N_2O$  emission intensities maximized in grass monocultures which was primarily due to the small denominator (low yields) without fertilizer or  $N_{\text{symb}}$ . Significant variation in  $N_2O\text{-Int}$  among models, fertilizer inputs, and sites were observed. Models agreed that  $N_2O\text{-Int}$  were significantly lower for the sites G1 and G2 compared to G4 and G5 (Table 1). Increasing fertilizer inputs strongly increased  $N_2O$  emission intensity from 0.006 g  $N_2O\text{-N kg}^{-1} \text{DM}$  in unfertilized swards to 0.096 g  $N_2O\text{-N kg}^{-1} \text{DM}$  at 450 kg N fertilization for APSIM and from 0.341 g  $N_2O\text{-N kg}^{-1} \text{DM}$  at unfertilized sward to 1.232 g  $N_2O\text{-N kg}^{-1} \text{DM}$  at 450 kg  $N_{\text{fert}}$  for DayCent (Table 1, Figure S3). Lower  $N_2O\text{-Int}$  was found for mixtures compared to clover and grass monocultures (based on means across fertilizer levels, models, and sites, Figure S3, Table 1). Modeled  $N_2O\text{-Int}$  was highest when N input came from fertilizer and decreased with increasing fraction of the total N input acquired via symbiosis. While the pattern of lower  $N_2O$  intensities at higher fractions of  $N_{\text{symb}}$  was independent of the level of total N input, the magnitude (slope) of the effect on  $N_2O$  intensity was such that a great mitigation effect was achieved in systems with high total N input. Variability in  $N_2O\text{-Int}$  between models exceeded variability across sites and fertilizer levels meaning that there is considerable uncertainty in the estimated magnitude. Still, models agreed on the overall patterns of mixtures resulting in lower  $N_2O\text{-Int}$  compared to monocultures.

#### 3.5. $N_2O$ Mitigation Potential

In order to assess a mitigation potential, a baseline scenario is required. Here, we selected the scenario without clover and with a fertilizer rate of 200 kg N  $\text{ha}^{-1} \text{yr}^{-1}$  (N200.Clo0) as outlined above (see section 2.4). To understand the effect of successive reductions in fertilizer input, we assessed the results by two stepwise reductions of fertilizer to 100 and 0 kg N  $\text{ha}^{-1} \text{yr}^{-1}$  and determined the clover proportion that obtained the maximum yield ( $\text{Clo}_{\text{opt}}$ ) for each site-model combination at the reduced fertilizer input. These were labeled Nx.  $\text{Clo}_{\text{opt}}$  where x is the amount of fertilizer input.

Simulations suggested that N200.Clo0 soils emitted 0.46 ( $\pm 0.27$  SE) g  $N_2O\text{-N}$  per kg of dry matter yield across all models and sites (Figure 5). For the first reduction in fertilizer input, N100.Clo<sub>opt</sub>, models showed consistently lower  $N_2O$  intensities of 0.30 ( $\pm 0.25$ ) g  $N_2O\text{-N}$  per kg of dry matter yield, corresponding to an average reduction of  $N_2O\text{-Int}$  by 34%. The lower  $N_2O\text{-Int}$  was present at all model-site combinations and resulted from a reduction of  $N_2O$  emissions by 25% on average and 32% higher yields compared to the baseline (Figure 5). Even though all model results showed that  $N_2O\text{-Int}$  consistently declined as fertilizer inputs were reduced, the magnitude of these reductions varied strongly across models and considerably across sites (Figure 5). A complete fertilizer omission reduced  $N_2O\text{-Int}$  to 0.19 g  $N_2O\text{-N}$  per kg of dry matter yield, corresponding to a 58% reduction compared to the N200.Clo0 baseline.

In order to define a target for the mitigation management, a key question is how clover proportions need to be adapted in order to produce maximum yields. Clover proportions at which the maximum yields were



**Figure 5.** N<sub>2</sub>O mitigation scenarios for mineral fertilizer replacement: (a) N<sub>2</sub>O emission intensity; (b) N<sub>2</sub>O emission; and (c) yields are displayed for scenario N200.Clo0 (200 kg N, 0% clover), N100.CloOpt (100 kg N, clover % for maximum yield), and N0.CloOpt (no fertilizer, clover % for maximum yield); the points display the mean, and whiskers display the standard error of the mean, calculated from the variability across models and sites.

reached ranged 66–86% (APSIM range across sites) and 0–72% (DayCent) for a N200.Clo<sub>opt</sub> scenario. A decrease in fertilization, N100.Clo<sub>opt</sub>, gave higher optimal clover proportions (72–100% by APSIM and 40–100% by DayCent across sites) as might be expected. Complete fertilizer omission (N0.Clo<sub>opt</sub>) resulted in clover proportions for optimal yield of 88–100% (APSIM) and 46–89% (DayCent).

If a reduction of N<sub>2</sub>O emissions is accompanied by yield losses, its acceptability by farmers is likely to be very low. Thus, our aim was to identify scenarios at which the yield was not compromised by the N<sub>2</sub>O mitigation and distinguish them from scenarios with yield losses. Surprisingly, yield reductions compared with N200.Clo0 were only found for N0.Clo<sub>opt</sub> at five out of 20 model-site combinations. Thus, these results suggest that a reduction of fertilizer from 200 to 0 kg N was possible without yield losses provided it was also possible to obtain an optimal clover proportion in the sward.

The global N<sub>2</sub>O mitigation potential of replacing N<sub>fert</sub> partially with N<sub>symp</sub> was calculated for the optimal clover content and the 30% and 60% clover scenarios. Average N<sub>2</sub>O emissions across our models were 2.96 kg N<sub>2</sub>O-N ha<sup>-1</sup> yr<sup>-1</sup> if business-as-usual fertilizer application was 200 kg N ha<sup>-1</sup> yr<sup>-1</sup>. If the relative mitigation effects assessed our study are assumed to be representative for C3 pasturelands and assuming the baseline fertilizer application of 200 kg N ha<sup>-1</sup> yr<sup>-1</sup>, a reduction to 100 N ha<sup>-1</sup> yr<sup>-1</sup> combined with an increase to 30% or 60% clover would reduce fertilizer N<sub>2</sub>O emissions by 37–40% and would scale to a global mitigation effect of 119–128 Gg N<sub>2</sub>O-N yr<sup>-1</sup>. While N<sub>2</sub>O emissions were similar at 30% and 60% clover, yields were reduced (by 20%) at the 30% clover scenario compared to the baseline, while no yield decrease was found for the 60% clover scenario.

In comparison, a complete omission of N<sub>fert</sub> combined with an increase in N<sub>symp</sub> to its maximum scales to reductions of 258 Gg N ha<sup>-1</sup> yr<sup>-1</sup> for C3 pasturelands globally, when assuming 200 kg N yr<sup>-1</sup> initial fertilization, respectively. Using initial clover percentage of 5%, instead of 0%, did not substantially affect these results.

## 4. Discussion

### 4.1. Yields, N<sub>2</sub>O Emissions, N<sub>2</sub>O Emission Intensities, and Mitigation Potentials

The general pattern of modeled yields increasing with increasing fertilizer input is consistent with experimental data. Yields of grass monocultures and grass-legume mixtures in APSIM increased across the whole range of fertilizer levels from 0 to 600 kg N ha<sup>-1</sup> yr<sup>-1</sup>; however, in DayCent, yields increased only between 0 to 200 kg N ha<sup>-1</sup> yr<sup>-1</sup> of fertilizer. Empirical findings show yields leveling off at high fertilizer amounts (400–500 kg N ha<sup>-1</sup> yr<sup>-1</sup>), depending on the site conditions (Prins, 1983; Reid, 1983; Sparrow, 1979; Whitehead, 2000) or still increasing at the highest N fertilizer levels (Wilkins et al., 2001). We conclude that yield

responses to fertilization were more realistically simulated in APSIM but were unrealistically low in DayCent, which showed only marginally increased yield with increased fertilizer applications above 200 kg N ha<sup>-1</sup> yr<sup>-1</sup> at 50% clover (Table 2 even at the sites with high yield potential (G2–G5) (Prins, 1983; Whitehead, 2000). Underestimating a fertilizer effect on yields has a consequence for calculating mitigation potentials (section 3.5), as this would lead to an overestimation of potential reductions in N<sub>2</sub>O-Int.

Model outputs reflected the influence of clover proportion on yields observed in grass-clover studies (Kirwan et al., 2007; Lüscher et al., 2014; Nyfeler et al., 2009; Suter et al., 2015). Yield maxima for DayCent were observed when the proportion of clover was between 38% and 79% across sites at 150 kg N and at relatively low clover proportions (0–47%) at 300 kg N. APSIM simulated yield maxima at comparably high clover proportions of 74–87% at 150 kg N and 66–76% at 300 kg N. Measured data showed overyielding at 40–60% clover for fertilizer levels of 50–450 kg ha<sup>-1</sup> yr<sup>-1</sup> at a site near G5 (Nyfeler et al., 2009), indicating that models were generally in line with observed data, but APSIM had a tendency to simulate higher optimal clover proportions than observed at this site. APSIM and DayCent differed with respect to plant N uptake at higher fertilizer levels. APSIM simulated highly efficient N uptake at high fertilizer levels, whereas in DayCent grasslands did not profit from increased fertilization above 200 kg N, and DayCent simulated increased leaching. It should be noted that the effect of legume percentage on DayCent's potential N fixation was derived from APSIM simulations (Fitton et al., 2019) while DayCent reduced N fixation from the potential based on its internal simulation of soil nitrogen conditions. Thus, while the two models used here are not completely independent, there are several degrees of separation between them. The development of a more process-based representation of multiple species in DayCent could avoid such a dependency.

Modeled  $N_{\text{symb}}$  was within realistic ranges across all sites and scenarios (Table 1; Ledgard & Steele, 1992; Lüscher et al., 2014; Nyfeler et al., 2011), for example,  $N_{\text{symb}}$  of up to 323 kg N ha<sup>-1</sup> yr<sup>-1</sup> was measured in Nyfeler et al. (2011) and values of up to 545 kg N ha<sup>-1</sup> yr<sup>-1</sup> have been observed by Carlsson and Huss-Danell (2003), both with *T. repens*, in line with our findings of maxima ranging between 424 and 559 kg N ha<sup>-1</sup> yr<sup>-1</sup> across sites. Grass-clover mixtures of <80% clover have the potential to fix more N<sub>2</sub> from the atmosphere compared to clover monocultures due to niche complementarity, while clover proportions >80% are less beneficial due to less growth and therefore less  $N_{\text{symb}}$  (Nyfeler et al., 2011), which was also depicted by the models.

Simulated N<sub>2</sub>O emissions in APSIM were low compared to N<sub>2</sub>O emission estimates using IPCC Emission Factors at all fertilizer levels (i.e., below the IPCC lower confidence interval). These low N<sub>2</sub>O emissions could arise either from APSIM simulating low soil organic matter turnover rates or because APSIM simulates N being taken up by plants too efficiently or a combination of both effects. In contrast, N<sub>2</sub>O emissions in DayCent exceeded IPCC mean estimates at all fertilizer levels but were within the upper confidence interval estimates of the IPCC and were similar to estimates from Shcherbak et al. (2014) for the upper fertilizer application rates (>300 kg N ha<sup>-1</sup> yr<sup>-1</sup>). These high N<sub>2</sub>O emissions in DayCent coincided with increased leaching rates at higher levels of fertilization, reflecting that plants were not able to benefit from additional N fertilization and thus biomass yields from DayCent at these fertilizer rates seemed to be unrealistically low. While the current IPCC Emission Factor (EF1) approach assumes a linear increase in N<sub>2</sub>O emissions with N fertilizer amounts, Shcherbak et al. (2014) found  $N_{\text{fert}}$  to have a nonlinear effect on N<sub>2</sub>O emissions. They analyzed site data across the globe using multiple N application rates, showing that the increases in N<sub>2</sub>O emissions were higher at higher fertilizer amounts (mean +0.0033% change in N<sub>2</sub>O emitted per additional kg N ha<sup>-1</sup> yr<sup>-1</sup> of fertilizer added based on 41 site-years for grassland; Table S3 in Shcherbak et al., 2014). Greater change in N<sub>2</sub>O emissions per  $N_{\text{fert}}$  increment at higher fertilizer input levels implies that N<sub>2</sub>O emission reductions are more effective in high N input systems compared to systems of moderate or low fertilizer N inputs, where a reduction in the same amount of  $N_{\text{fert}}$  is less effective. Kim et al. (2013) conceptualized the relationship between N fertilizer amounts and N<sub>2</sub>O emissions and hypothesized an initial linear increase of N<sub>2</sub>O production at the low fertilizer levels at which microbes compete with plants for available N. Additional N fertilizer increase beyond optimal plant N uptake rates would then lead to an exponential increase in N<sub>2</sub>O production due to an excess of available N, stimulating nitrification and denitrification. Eventually, further fertilizer N increment would lead to N<sub>2</sub>O production leveling off, leading to a steady state where soil C limits N<sub>2</sub>O production (Kim et al., 2013; Rochette et al., 2010). Model outputs showed nonlinear (APSIM) and linear (DayCent) effects of N fertilizer amounts on N<sub>2</sub>O emissions on average, while leveling off was observed for some model-site combinations in DayCent.

In line with the findings reviewed in Lüscher et al. (2014), N<sub>2</sub>O-Int was lower in grass-clover mixtures compared to monocultures, which was expected since N<sub>2</sub>O-Int penalizes yield losses (as observed at low N input), and penalizes high N<sub>2</sub>O emissions (as observed at high N inputs). Thus, grass-clover mixtures optimize the combination of both. As the increase in N<sub>2</sub>O emissions with increases in N fertilizer typically



exceed the yield gain for the same fertilizer increment,  $N_2O$ -Int were minimized at the lowest fertilizer levels (Table 1). The modeled results of yields being maintained under reduced fertilizer might be too optimistic. However, in a recent field experiment at Site G5 (Fuchs et al., 2018),  $N_2O$ -Int was reduced from  $0.42 \text{ g N kg DM}^{-1} \text{ yr}^{-1}$  on a fertilized control obtaining fertilizer N via slurry (240 kg N, 4–15% clover) to  $0.28 \text{ g N kg DM}^{-1} \text{ yr}^{-1}$  at the clover treatment (23–40% clover and no  $N_{\text{fert}}$ ) over a 2-year observation period. Further, lab experiments by Barneze et al. (2019) confirmed the model results that higher clover proportions increase productivity without affecting  $N_2O$  emissions, resulting in reduced  $N_2O$ -Int.

The yields at higher clover percentages (75%) were overestimated by DayCent and APSIM (Table S1b). While DayCent predicted an optimal clover content around 50%, similar to the observations, APSIM predicted higher optimum clover content around 75%. The optimum clover proportion thus could likely be overestimated in this modeling exercise.

Dangal et al. (2019) modeled total emissions from grassland of  $3.11 \text{ Tg } N_2O\text{-N}$ ; other estimates range from  $1.86 \text{ Tg } N_2O\text{-N}$  in the EDGAR database (EDGAR, 2018) to  $2.8 \text{ Tg } N_2O\text{-N}$  in the study of Davidson (2009, referring to 2000). A mitigation potential of  $128 \text{ Gg } N_2O\text{-N yr}^{-1}$  (which resulted from scaling up the results from our study; baseline  $200 \text{ kg N}$ ) resulted in a 4.1% reduction of global grassland  $N_2O$  emissions if the Dangal et al. estimate is used a reference. The tested mitigation option reduced total grassland  $N_2O$  emissions only moderately, since the largest fraction of grassland  $N_2O$  emissions is attributed to the fate of animal excreta and manure (Dangal et al., 2019), while fertilizer application contributes 11%. The estimated average  $N_2O$  emissions for C3 pasture in Dangal et al. (2019) was rather high with an average across C3 pasturelands of  $3.95 \text{ kg } N_2O\text{-N ha}^{-1} \text{ yr}^{-1}$  during 2010–2014 and was comparable to the estimates by our models for grasslands receiving  $300 \text{ kg } N_{\text{fert}} \text{ ha}^{-1} \text{ yr}^{-1}$ .

Replacing not only synthetic fertilizers but also organic N application and applying the organic N on croplands would at least double the estimated  $N_2O$  mitigation effect, since higher emissions come from organic fertilizers compared to mineral fertilizers (Dangal et al., 2019). Modeling the scenarios presented here with organic fertilizer instead of mineral fertilizer resulted in similar reductions of  $N_2O$ . If the surplus of organic fertilizer could replace mineral fertilizer on croplands, the mitigation strategy could have far more impact. Since excreta deposition contributes the large fractions (54%; Dangal et al., 2019) of grassland  $N_2O$  emissions, larger reductions of grassland  $N_2O$  emissions need to target excreta of grazing livestock.

The global estimates of  $N_2O$  emission reductions largely depend on the absolute  $N_2O$  emissions estimated by the applied models. We are more confident in the relative estimates compared to absolute estimates, as the mitigation potentials largely are driven by the  $N_2O$  emission reductions due to omitted N fertilization. Besides the model uncertainties, the global estimates contain uncertainty, first due to the strong dependence on the outcomes of Dangal et al. (2019) and second due to the use of an average mitigation potential for a limited number ( $n = 5$ ) of sites. Thus, these estimates should rather be seen as an indicator for the magnitude, keeping these uncertainties in mind.

#### 4.2. Conclusion and Outlook

In this study, we investigated the effect of reduced fertilizer combined with increased clover proportions as a mitigation strategy and contributed to a mechanistic understanding of the relationships between the proportion of clover and  $N_2O$  emissions and emission intensities. Results from different models gave contrasting perspectives on  $N_2O$  emissions from grassland systems, with quite diverse outcomes in absolute  $N_2O$  emission intensities, but agreement on the overall pattern of  $N_2O$  emission reductions on a per-unit-area, as well as on a per-unit-product basis. Models predicted relative changes in  $N_2O$  emission and  $N_2O$ -Int (relative to the overall site-model reference level) much more consistently than they predicted the absolute changes in site-specific  $N_2O$  emission intensities. Consequently, a more robust statement on the relative reduction in  $N_2O$  emission intensity than on absolute  $N_2O$  reduction potential is possible.

The simulated effects of clover percentage and fertilizer on  $N_2O$  emissions agreed well with the empirical studies, that is, no increase in  $N_2O$  emissions was observed at low to medium fertilizer amounts with increasing clover percentage up to 60% (compare Figure 1b with Figure 3). Assuming 0–5% clover and  $200 \text{ kg N ha}^{-1} \text{ yr}^{-1}$  on average as the baseline scenario, an increase in clover percentage of all fertilized

C3 pastures to 30% combined with a reduction of fertilizer inputs to 100 N ha<sup>-1</sup> yr<sup>-1</sup> would reduce global N<sub>2</sub>O emissions by 128 Gg yr<sup>-1</sup>, around 4% of global grassland N<sub>2</sub>O emissions.

Different biogeochemical models tend to have complementary strengths, leading to their multi-model ensemble median performing better compared to individual models (Asseng et al., 2013; Ehrhardt et al., 2018). Thus, using an ensemble reduces the overall error of the predictions, while also providing an estimate of the uncertainty of those predictions. However, constructing a large ensemble was impossible for this exercise because only few models are capable of simulating species mixtures (Kipling et al., 2016; Snow et al., 2014). Major challenges were the adequate simulation of BNF in grass-clover-mixtures, which finally led to a reduced number of participating models. In order to tackle present shortcomings, models need to be improved and species-interactions need to be considered in order to better and more realistically reflect the ongoing processes. Few models adequately simulate grazing effects (Snow et al., 2014). We simulated cut-and-carry systems in order to tackle the complexity of the modeling and ensure interpretable outputs. In reality, grazing would impact grassland productivity, legume content, and N<sub>2</sub>O emissions. However, the reasoning was to first consider the simpler system in order to enable understanding without confounding effects. An important next step for this work is to test the mitigation under grazing; however, this would entail first overcoming challenges with models and their abilities to reasonably simulate grazed conditions (e.g., see Snow et al., 2014).

**Acknowledgments**

This modeling study was a joint effort of the Models4Pastures project within the framework of FACCE-JPI. Lutz Merbold and Kathrin Fuchs acknowledge funding received for the Swiss contribution to Models4Pastures (FACCE-JPI project, SNSF funded contract: 40FA40\_154245/1) and for the Doc. Mobility fellowship (SNSF funded project: P1EZP2\_172121). Lutz Merbold further acknowledges the support received for CGIAR Fund Council, Australia (ACIAR), Irish Aid, the European Union, the Netherlands, New Zealand, Switzerland, UK, USAID, and Thailand for funding to the CGIAR Research Program on Climate Change, Agriculture and Food Security (CCAFS) as well as for the CGIAR Research Program on Livestock. The NZ contributors acknowledge funding from the New Zealand Government Ministry of Primary Industries to support the aims of the Livestock Research Group of the Global Research Alliance on Agricultural Greenhouse Gases and from AgResearch's Strategic Science Investment Fund (the Forages for Reduced Nitrate Leaching (FRNL) research program). The UK partners acknowledge funding by DEFRA and the RCUK projects: N-Circle (BB/N013484/1), UGRASS (NE/M016900/1), and GREENHOUSE (NE/K002589/1). R.M. Rees and C.F.E. Topp also received funding from the Scottish Government Strategic Research Programme. Lorenzo Brilli, Camilla Dibari, and Marco Bindi received funding from the Italian Ministry of Agricultural Food and Forestry Policies (MiPAAF). The FR partners acknowledge funding from CN-MIP project funded by the French National Research Agency (ANR-13-JFAC-0001) and from ADEME (no. 12-60-C0023). Open access funding enabled and organized by Projekt DEAL

Improved land use mapping of grass-legume mixtures applying machine learning algorithms on remotely sensed high-resolution data may in the future allow the extent of grass-legume mixtures to be identified, and at some point, also the respective percentage of legumes in the mixtures. Such detailed maps will enable a specific baseline on the actual extent and distribution of grass-legume mixtures to be defined and consequently enable a better estimate on the N<sub>2</sub>O mitigation potential of grass-legume mixtures.

Practical limitations potentially restricting the applicability of the investigated N<sub>2</sub>O mitigation strategy were not considered here. These limitations include the challenge of achieving high and persistent legume proportions, particularly under grazing. Further, clover is not well adapted for all locations, and local soil conditions such as low soil pH and waterlogging can severely limit clover production.

Here, we have shown that grass-clover mixtures with the lowest fertilizer input were associated with the lowest N<sub>2</sub>O emission intensities. Experimental studies which systematically assess N<sub>2</sub>O emissions under different clover proportions for different N fertilizer regimes would be highly beneficial. Experimental work should focus on assessing in detail swards with clover proportions of 30–50% and ≤150 kg N input as these seem most beneficial with respect to multiple outputs such as yields, N yields and also feeding values (Lüscher et al., 2014). Additional data on the endpoints (i.e., 0% and 100%) clover would assist in confirming or refuting the modeled outcomes. Legumes have been shown to perform better than non-legumes under drought conditions (Hofer et al., 2017) and show higher water-use efficiencies than non-legumes (Adams et al., 2018). Together with the fact that legumes can replace mineral fertilizer additions make them important species for increasing sustainability and resilience of agricultural production under changing climate, highlighting the role legumes could play not only for climate change mitigation but also for adaptation.

**Data Availability Statement**

The data used in this study are available in the ETH Research collection under <https://www.research-collection.ethz.ch/handle/20.500.11850/395370> as “Dataset: Evaluating the potential of legumes to mitigate N<sub>2</sub>O emissions from permanent grassland using process-based models.”

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