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1 **Ensemble modelling, uncertainty and robust predictions of organic**
2 **carbon in long-term bare-fallow soils**

3 *Model inter-comparison of soil organic carbon*

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51

52 **Abstract**

53 Simulation models represent soil organic carbon (SOC) dynamics in global carbon (C) cycle
54 scenarios to support climate-change studies. It is imperative to increase confidence in long-term
55 predictions of SOC dynamics by reducing the uncertainty in model estimates. We evaluated SOC
56 simulated from an ensemble of 26 process-based C models by comparing simulations to
57 experimental data from seven long-term bare-fallow (vegetation-free) plots at six sites: Denmark
58 (two sites), France, Russia, Sweden, the United Kingdom. The decay of SOC in these plots has
59 been monitored for decades since the last inputs of plant material, providing the opportunity to test
60 decomposition without the continuous input of new organic material. The models were run
61 independently over multi-year simulation periods (from 28 to 80 years) in a blind test with no
62 calibration (Bln) and with three calibration scenarios, each providing different levels of
63 information and/or allowing different levels of model fitting: a) calibrating decomposition
64 parameters separately at each experimental site (Spe); b) using a generic, knowledge-based,
65 parameterisation applicable in the Central European region (Gen); and c) using a combination of
66 both a) and b) strategies (Mix). We addressed uncertainties from different modelling approaches
67 with or without spin-up initialisation of SOC. Changes in the multi-model median (MMM) of SOC
68 were used as descriptors of the ensemble performance. On average across sites, Gen proved
69 adequate in describing changes in SOC, with MMM equal to average SOC (and standard deviation)
70 of $39.2 (\pm 15.5)$ Mg C ha⁻¹ compared to the observed mean of $36.0 (\pm 19.7)$ Mg C ha⁻¹ (last observed
71 year), indicating sufficiently reliable SOC estimates. Moving to Mix (37.5 ± 16.7 Mg C ha⁻¹) and
72 Spe (36.8 ± 19.8 Mg C ha⁻¹) provided only marginal gains in accuracy, but modellers would need
73 to apply more knowledge and a greater calibration effort than in Gen, thereby limiting the wider
74 applicability of models.

75

LIST OF SYMBOLS AND ABBREVIATIONS

Symbol/abbreviation	Long version	Explanation
<i>System variables</i>		
C	Carbon	Chemical element with atomic number 6
SOC	Soil organic carbon	Carbon stored in soil organic matter
SOM	Soil organic matter	The fraction of the soil that consists of plant, animal or microbial tissue in various stages of decomposition
N	Nitrogen	Chemical element with atomic number 7
<i>Experimentation</i>		
LTE	Long-term field experiment	Research facility providing data for monitoring trends and evaluating different agricultural management strategies over time
LTBF	Long-term bare-fallow experimental site	Research facility providing data for monitoring trends on bare-fallow soils
S1	Site 1	Askov (Denmark) – location 1
S2	Site 2	Askov (Denmark) – location 2
S3	Site 3	Grignon (France)
S4	Site 4	Kursk (Russia)
S5	Site 5	Rothamsted (United Kingdom)
S6	Site 6	Ultuna (Sweden)
S7	Site 7	Versailles (France)
<i>Modelling</i>		
M01, ..., M34	Model 01, ..., model 34	Simulation models (M) anonymously coded from 1 to 34
Bln	Blind	Uncalibrated simulations (blind test)
Gen	Generic	Generic simulation scenario
Mix	Mixed	Mixed simulation scenario
Spe	Specific	Specific simulation scenario
SP	Spin-up	Process of running the model from a set of conditions to initialise the state of C pools
NS	No spin-up	Any function (or analytical procedures) to make an initial partition of C pools (alternative to spin-up runs)
<i>Statistics</i>		
SD	Standard deviation	Variation amount of a set of data
MMM	Multi-model median	Median value of simulated data from different models
Obs	Observations	Observed data
RRMSE	Relative root mean square error	Aggregate magnitude of the errors in predictions relative to the mean of observations

EF	Modelling efficiency	Predictive power of a model with respect to the mean of observations
R^2	Coefficient of determination	Proportion of the variance in the modelled data that is predictable from the observations
r	Pearson's correlation coefficient	Degree to which predictions and observations are linearly related
P(t)	Paired Student t-test probability of I-type error	Probability to reject the true null hypothesis of equal means of two samples of paired data (i.e. predictions and observations)
d	Index of agreement	Ratio of the mean square error and the potential error represented by the largest value that the squared difference of each prediction/observation pair can attain
z	z-score transformation	Number of standard deviations by which the value of a raw score is above or below the mean value of the variable of interest
sd	Standard deviation	Standard deviation units expressing z-scores
sd_{obs}	Standard deviation of observations	Variation amount of a set of observed values
P	Predicted value	Value of a variable that is generated using a model
O	Observed value	Value of a variable that is actually observed
n	Number of predicted or observed values	Number of predicted/observed pairs
i	i^{th} predicted or observed value	Subscript index of each predicted/observed pair
\bar{O}	Mean of observed values	Arithmetic mean of actually observed data
\bar{P}	Mean of predicted values	Arithmetic mean of actually observed data
\bar{D}	Mean difference	Arithmetic mean of the differences between predicted and observed values
S_D	Standard deviation of the differences	Variation amount of a set of differences between predictions and observations
p	Probability of I-type error	Probability to reject the true null hypothesis of null correlation between two variables
<hr/> <i>Agro-climatic metrics</i> <hr/>		
Tamp	Temperature amplitude	Difference between the highest and the lowest temperature in a year

Tmax	Maximum air temperature	Average of the highest daily temperatures in a year
Prec	Precipitation	Annual precipitation total
<i>b</i> ^a	De Martonne-Gottman aridity index	Indicator of aridity including both annual and monthly temperature and precipitation
<i>hw</i> ^a	Heatwave frequency	Number of at least seven consecutive days when the maximum air temperature is higher than the average summer (June, July and August) maximum temperature of a baseline value +3 °C

77 1. INTRODUCTION

78 The ability of soils to sequester and store large amounts of carbon (C) is well known (e.g.
79 Lehmann and Kleber, 2015). Soil organic carbon (SOC) stocks are crucial for maintaining soil
80 fertility and preventing erosion and desertification, and they positively influence the provision of
81 ecosystem services at the local as well as the global scale (e.g. Lal, 2004, 2014). For these reasons,
82 farmers aim to establish and maintain high organic C stocks in agricultural soils, which have often
83 been depleted through historical land use practices (Fuchs et al., 2016; Gardi et al., 2016; Chenu et
84 al., 2018). The continuing studies on SOC sources and biogeochemical processes in the soil
85 environment provide key insights into climate-C feedbacks, and help prioritizing C sequestration
86 initiatives (Gross and Harrison, 2019). In light of the climate change issue, the storage of C and
87 additional sequestration of atmospheric C have received increasing attention recently (Rumpel et
88 al., 2018; Whitehead et al., 2018; Lavallee et al., 2020), promoting land management, and agro-
89 ecosystems in particular, as a key mitigation option (e.g. the ‘4 per mille Soils for Food Security
90 and Climate’ initiative, Minasny et al., 2017; Soussana et al., 2017). However, the slow response
91 of SOC to changes in management and environmental factors hampers our understanding of how
92 SOC can be increased in a sustainable manner, especially under changing climatic conditions.
93 Long-term field experiments (LTEs), in which SOC responses have been observed over several
94 decades, provide this information and deliver reference data on SOC content for knowledge gain
95 and model development (Johnston and Poulton, 2018). However, LTEs are costly to maintain, and
96 it is generally difficult to extrapolate experimental results across space and time (Debreczeni and
97 Körschens, 2003; Mirtl et al., 2018). Simulation models play a prominent role in SOC research
98 because they provide a mathematical framework to integrate, examine and test the understanding
99 of SOC dynamics (Campbell and Paustian, 2015). They can also be used to extrapolate from micro-
100 (e.g. carbohydrate production during photosynthesis) to macro-scale dynamics (e.g. global C
101 cycling) (e.g. Gottschalk et al., 2012; Sitch et al., 2003). In particular, complex agricultural and

102 environmental models incorporate a mechanistic view of processes and system interactions, in
103 which the soil components are often represented by different, operationally defined, pools of
104 different sizes and with different properties (e.g. Parton et al., 2015). The concept of multiple C-N
105 pools represents C-N dynamics with an idealised description (Hill, 2003). The relative proportion
106 of C and N (and sometimes lignin to N ratio) in the plant residue is the primary mode to divide
107 plant inputs (from e.g. leaf litter and root exudates) into fresh litter pools, which then decompose
108 into SOC (or SOM, i.e. soil organic matter) pools, each being modelled with different residence
109 (or turnover) times, varying from months for labile products of microbial decomposition to
110 hundreds to thousands of years for organic substances with firm organic-mineral bonds (e.g. Yadav
111 and Malanson, 2007; Dungait et al., 2012). Plant material and animal manures are often modelled
112 to enter the soil environment as either readily decomposable (carbohydrate-like) or resistant (lignin
113 and cellulose-like) materials. A varying number of pools (often including inert and slow-
114 decomposing organic matter, and microbial biomass) linked by first-order equations is usually
115 simulating both C and N fluxes within and between each pool (Falloon and Smith, 2010). However,
116 different models vary considerably in the underlying assumptions and C processes in current
117 models, e.g. regarding number of pools, type of decomposition kinetics used and processes
118 regulating SOC retention (Manzoni and Porporato, 2009; Cavalli et al., 2019).

119 Each model offers a distinctive synthesis of scientific knowledge (Brilli et al., 2017) and
120 multi-model ensembles developed from several models may reduce uncertainties in biological and
121 physical outputs that occur over large scales, such as regions and continents (e.g. Rötter et al.,
122 2012; Asseng et al., 2013; Ehrhardt et al., 2018). The advantage of using ensemble estimates over
123 individual models is that caused by compensation of errors across models, and a broader
124 integration of model processes (Martre et al., 2015). It has been recommended to use model
125 ensembles for reducing uncertainties in simulations of agricultural production (Asseng et al., 2013;
126 Bassu et al., 2014; Challinor et al., 2014; Li et al., 2015; Ruane et al., 2016; Maiorano et al., 2017)

127 and other biophysical/biogeochemical outputs (Sándor et al., 2017, 2018a; Ehrhardt et al., 2018).
128 However, after the pioneering study of Smith et al. (1997), who evaluated nine SOC models using
129 12 datasets from seven LTEs, other modelling studies targeting SOC dynamics have often been
130 limited in scope. Smith et al. (2012) used four models to assess the effect on SOC of crop residues'
131 removal in 14 experiments in North America. Todd-Brown et al. (2013, 2014) performed global
132 estimates of SOC changes with 11 Earth system models. Kirschbaum et al. (2015) used one
133 simulation model and two years of eddy covariance measurements collected over an intensively
134 grazed dairy pasture in New Zealand to better understand the drivers of changes in SOC stocks.
135 Puche et al. (2019) performed a similar study in France. Using multi-model ensembles in scenario
136 studies at eight sites worldwide, Basso et al. (2018) highlighted the importance of soil feedback
137 effects (C and N) on the prediction of wheat and maize yield. We are not aware of any recent
138 model inter-comparison studies specifically assessing soil C dynamics with several models across
139 a range of experimental sites. This is a field where there is a need for standardised guidance to
140 estimate C stocks at various spatial scales (Bispo et al., 2017). A difficulty in testing and comparing
141 various models (and interpreting model outputs) lies in the interaction between soil and plant
142 processes so that any of the model-data discrepancies could be due to errors in either component
143 (e.g. Ehrmann and Ritz, 2014). A rigorous model testing and comparison would require different
144 model components, e.g. plant and soil modules, to be assessed separately. Bare-fallow plots offer
145 such an opportunity in that they are plots maintained for decades without any plant inputs. The
146 changes in SOC stocks therefore result only from decomposition processes. To assess the function
147 of soil-model components without interaction with plant processes, we conducted a model inter-
148 comparison using a dataset from long-term bare-fallow experiments where plant inputs were zero.
149 In this study, we refer to bare-fallow plots that were kept free of plants by manual and/or chemical
150 means for several decades. We used seven bare-fallow treatments included in six long-term
151 agricultural experiments (>25 years), all located in Europe (Denmark, France, Russia, Sweden and

152 United Kingdom). In these plots, the soils became progressively depleted in the more labile SOM
153 components, as they decomposed, and relatively enriched in more stable SOM (Barré et al., 2010).
154 The soil C concentrations determined at given years in these sites represented a unique opportunity
155 to follow the decay of SOC from a multi-model ensemble perspective, without any interference
156 from new plant C inputs, and conduct a multi-model ensemble comparison. The model inter-
157 comparison included 26 process-based models from an international modelling community. Some
158 models only accounted for soils and used C input from plants as an external input where others
159 were full agro-ecosystem models that explicitly simulate plant growth and resulting C input into
160 soils. These models all simulate interactions between the soil-atmosphere continuums in different
161 ways, but for this comparison all models were run assuming no input of fresh plant-derived C,
162 allowing the comparison of just the soil components of the models.

163 Here, we assess the models, by comparing multi-decadal simulations to experimental data
164 from seven sites in Europe. The primary goal of this study was to assess the multi-model ensemble
165 in simulating SOC dynamics across bare-fallow sites in Europe. To achieve this goal, model
166 evaluation against actual measurements was performed before and after model calibration. In
167 addition, deficient areas in models and their processes were identified, paving the road for future
168 research directions.

169

170 **2. MATERIALS AND METHODS**

171 **2.1. Simulation models**

172 The ensemble of models consisted of 26 process-based models, mainly developed for crop or
173 grassland ecosystems (or focussing just on soils) and covering a broad variety of approaches (Table
174 1). While they are mostly based on first-order decay kinetics of multiple C pools (where C losses
175 are proportional to SOC stocks with additional modifiers to represent the effects of other factors),
176 ESOC1 simulates C fluxes with second-order kinetics equations based on concepts applied in

177 Schimel and Weintraub (2003) and reviewed in Wutzler and Reichstein (2008). In this case,
178 organic matter decomposition includes reactions between SOC and decomposers (i.e. a microbial
179 or enzyme pool). These different approaches depend mainly on alternative ways in which the C
180 pools are linked. For instance, MONICA is one of the most complex models, considering three
181 types of organic matter in six conceptual pools, viz. newly added organic matter, living soil
182 microbial biomass and native non-living soil organic matter, each sub-divided into fast and slowly
183 decomposing sub-pools. It simulates the turnover of C pools by applying first-order degradation
184 to each pool due to microbial growth and maintenance respiration (after Abrahamsen and Hansen,
185 2000). Then, like other models (e.g. CenW), MONICA also includes a coupled N-cycle and
186 sophisticated temperature and water-balance calculations that act as modifiers of degradation and
187 respiration rates. The decomposition rates of individual pools in such multi-pool SOC models are
188 typically controlled by vastly different reaction coefficients that can result in highly nonlinear
189 behaviour of the overall system (e.g. Caruso et al., 2018). The initial list included 34 models, but
190 eight of them were excluded from further analysis because they showed severe limitations to run
191 properly either under bare-fallow soils or under the given climate conditions. For all models,
192 estimates of SOC were compared with measured SOC data.

193 Table 1. The process-based simulation models used. Model names were anonymised in the
 194 reporting of simulation results using model codes from M01 to M34, from the initial list of 34
 195 models, the order of models not being identical to that used in the table.
 196

Model name	Version	C pools ^a	Spin-up	URL or contact for documentation/description	References
AMG	2	2 to 3	None	https://www6.hautsdefrance.inra.fr/agroimpact/Nos-dispositifs-outils/Modeles-et-outils-d-aide-a-la-decision/AMG-et-SIMEOS-AMG/AMG-model-description	Andriulo et al. (1999); Saffih-Hdadi and Mary (2008); Clivot et al. (2019)

APSIM	Apsim 7.9- r4044	3	Simulation from start of climate record (no additional simulation period)	http://www.apsim.info	Keating et al. (2003); Holzworth et al. (2014)

	7.10 r4158		Yes		

CANDY_CIPS	1.0 (but always implemented in newest	4	None	https://www.ufz.de/export/data/2/95948_CANDY_MANUAL.pdf	Kuka, (2005); Kuka et al. (2007)

	version of CANDY 29.06.2018					
CCB	2019.1.16	3	None	https://www.ufz.de/index.php?en=44046		Franko et al. (2011); Franko and Spiegel (2016); Franko and Merbach (2017)
Century	4.0	5 to 7	Yes	https://www2.nrel.colostate.edu/projects/century/MANUAL/html_manual/man96.html		Parton et al. (1987, 1994)
CenW	4.2	5	Uses an automatic spin-up routine to find equilibrium conditions under given environmental variables and specified system properties	http://www.kirschbaum.id.au/Welcome_Page.htm		Kirschbaum (1999); Kirschbaum and Paul (2002)
C-TOOL	2014	3	None (can be run also with spin-up)	http://envs.au.dk/fileadmin/Resources/DMU/Luft/emission/SI_NKS/C-TOOL_Documentation_2015_.pdf		Taghizadeh-Toosi and Olesen (2016); Taghizadeh-Toosi et al. (2014a, b, 2016)
Daily DayCent	4.5 2010 Daily DayCent 4.5 2013	5 to 9	Yes	http://www.nrel.colostate.edu/projects/daycent-home.html		Parton et al. (1994, 1998); Del Grosso et al. (2001, 2002)

 Daily
 DayCent
 August 2014

4.5 2013

DNDC	CAN	6	Yes (10 years recommended)	http://www.dndc.sr.unh.edu	Li et al. (2012); Smith et al. (2020)
DSSAT	...	5	Yes, 20 years prior to beginning of the experiment to estimate the proportions of carbon in each organic matter pool	http://dssat.net	Jones et al. (2003); Porter et al. (2009); Gijssman et al. (2002); White et al. (2011); Thorp et al. (2012)
ECOSSE	5.0.1	5	None	https://www.abdn.ac.uk/staffpages/uploads/soi450/ECOSSE%20User%20manual%20310810.pdf	Smith et al. (2007, 2010a, b); Bell et al. (2010)
ESOC1	1.0	3	Yes	https://doi.org/10.5281/zenodo.3539484 fmoyano@uni-goettingen.de	Moyano et al. (2018)

Exp		1	None	-	Lorenzo Menichetti (lorenzo.menichetti@slu.se)
Exp + inert		2	None	-	
ICBM	...	2	None	martin.bolinder@slu.se https://www.slu.se	Andrén and Kätterer (1997); Andrén et al. (2008)
MONICA	2.0.2	7	None	http://monica.agrosystem-models.com	Nendel et al. (2011); Specka et al. (2016); Stella et al. (2019)
ORCHIDEE	2.0	3	Yes	https://vesg.ipsl.upmc.fr/thredds/fileServer/IPSLFS/orchidee/DOXYGEN/webdoc_2425/annotated.html	Krinner et al. (2005)
RothC	RothC10N ----- 26.3	4 to 5	None	https://www.rothamsted.ac.uk/rothamsted-carbon-model-rothc	Coleman and Jenkinson (1999); Farina et al. (2013)
STICS	9.0	2 to 4	None	http://www6.paca.inra.fr/stics	Brisson et al. (1998, 2003, 2008); Coucheney et al. (2015)

YASSO15

15

5

Yes

<https://en.ilmatiiteenlaitos.fi/yasso>

Tuomi et al. (2009)

197 ^a Some models/model versions include options for varying C pools (this varying number may depend on the fact that the full

198 set of pools including fresh C can be optionally simplified in the case of bare-fallow treatments).

199 2.2. Experimental sites

200 We used data from a network of six long-term bare-fallow experimental sites (LTBF) in Europe
201 (with two fields located in Askov, Denmark; Barré et al., 2010), to test the ability of the models to
202 represent SOC dynamics. The sites were located at a range of latitudes between 48° to 59° North
203 (Table 2; Fig. 1a), with experiments running for at least 28 years, which were used as a test bed
204 for the models to represent SOC dynamics. Table 2 shows the main characteristics of each site and
205 provides a brief description of the historical land use and management of the area (more details
206 are given by Barré et al., 2010 and references therein). The documented history of the experimental
207 sites referred to the presence of agricultural areas (grassland or cropland), without woodlands. Soil
208 texture provides evidence of variability in soil physical properties, with a gradient of intermediate
209 situations between the sandy loam of Askov (Denmark) and the clay loam of Ultuna (Sweden).
210 Water relations (precipitation minus reference evapotranspiration) indicate positive climatic water
211 balance for the two North Atlantic sites only (Askov in Denmark and Rothamsted in the United
212 Kingdom). Mean annual temperatures vary from ~6 °C in the Sweden and Russian sites (Ultuna
213 and Kursk, respectively) to near 11 °C in the two French sites (Grignon and Versailles). Annual
214 air temperature amplitudes - from about 14 °C in Rothamsted to near 30 °C in Kursk - indicate
215 that the study sites span a broad thermal gradient (Fig. 1b), which likely leads to different soil
216 thermodynamics (e.g. Zhu et al., 2019). Two widely used metrics (aridity index and frequency of
217 heatwaves; Sándor et al., 2017, 2018a, b) were also calculated to complete the climatic analysis of
218 study sites (Fig. A, supplementary material).

219

220
 221 Table 2. Long-term bare-fallow experimental sites. Table A in the supplementary material contains
 222 the summary description of the experimental sites.

		Experimental sites (country)					
		S1, S2	S3	S4	S5	S6	S7
General description		Askov (Denmark)	Grignon (France)	Kursk (Russia)	Rothamsted (United Kingdom)	Ultuna (Sweden)	Versailles (France)
Coordinates	Latitude	55.28	48.51	51.73	51.82	59.49	48.48
	Longitude	9.07	1.55	36.19	0.35	17.38	2.08
Soil	Sand/Silt/Clay (%)	78/12/10 (sandy loam)	16/54/30 (silty clay loam)	5/65/30 (silty clay loam)	13/62/25 (silt loam)	23/41/36 (clay loam)	26/57/17 (silt loam)
	Bulk density (Mg m⁻³)	1.50	1.20	1.13	0.94	1.44	1.30
	Experimental period	<i>Bare-fallow years</i> 1956-1985	1959-2007	1965-2001	1959-2008	1956-2007	1929-2008
	<i>N. of data/replicates</i>	30/4, 29/4	11/6	6/0	14/4	18/4	9/6
	Initial/final carbon stocks (Mg C ha⁻¹)	52.1/36.4	41.7/25.4	100.3/79.4	71.7/28.6	42.5/26.9	65.5/22.7
Climate^a	Climate type^b	Dfb (humid continental)	Cfb (oceanic)	Dfb (humid continental)	Cfb (oceanic)	Dfb (humid continental)	Cfb (oceanic)

	Mean annual precipitation total (mm)	890	584	482	723	457	608
	Mean annual cumulative evaporation (mm)^c	578	662	602	630	546	668
	Mean annual air temperature (°C)	7.4	10.7	6.2	9.4	6.0	10.7
	Mean annual air temperature range (°C)^d	17.6	16.8	29.8	14.4	22.8	16.7
Vegetation (historical period)^e	ANPP (g C m⁻² yr⁻¹)	1.7	1.1	0.9	1.3	0.9	1.2
	TNPP (g C m⁻² yr⁻¹)	3.3	2.2	1.7	2.5	1.7	2.2

223 ^a Climatic analysis was performed on longer periods than the experimental periods: 1956-1987/1929-2008/1944-
224 2003/1856-2006/1956-1999/1929-2008.

225 ^b Köppen-Geiger climate classification (Kottek et al., 2006).

226 ^c Mean values over the bare-fallow period. Reference evaporation was estimated based on the Thornthwaite (1948)
227 equation.

228 ^d Mean difference in temperature between the warmest and the coldest month of the year.

229 ^e Estimates of aboveground (ANPP) and total (TNPP) net primary productivity based on the precipitation levels of
230 each site, as provided by Del Grosso et al. (2008) for non-tree dominated systems.

231

232

(Fig. 1 here)

233

235 **2.3. Study design**

236 Model simulations were carried out independently by each modelling team (which included model
237 developers and users, and field experts of soil C dynamics) on commonly formatted data using
238 their own approaches and technical background. Harmonising calibration techniques was out of
239 scope of the inter-comparison exercise. The SOC outputs from each model were compared to data
240 from the study sites before and after calibration. Calibration mostly focussed on parameters related
241 to substrate use, C partitioning among pools and decomposition processes. However, rate
242 equations for C pools often required the calibration of a large number of parameters, which are at
243 the core of key processes responsible for differences among models in the understanding and
244 interpretation of SOC processes (number of pools and type of decomposition kinetics used to
245 represent C turnover). For the uncalibrated (blind test, Bln) simulations, the models were run for
246 each site using the available data of weather, soil texture and bulk density (model inputs), and the
247 initial SOC values, with no parameter adjustment other than initialisation based on historical
248 management and land use. With this information, Bln reflects the ability of the models to simulate
249 SOC decomposition after plant inputs has stopped, using the original parameter settings and
250 calibration, simply by removing their components related to new C inputs. At this stage, default
251 values were mostly used for all decomposition rates. C-pool fraction sizes were adjusted based
252 only on C-input estimates from the information on land use prior to the establishment of the bare-
253 fallow treatments.

254 After the blind simulations were completed, SOC measurements taken during the bare-
255 fallow period were supplied to each modelling group for the calibration work. Details on
256 management (tillage), which may have influenced the SOC dynamics before the bare-fallow
257 treatment, were also provided to improve the initialisation process. It was requested that each
258 modelling group adjust soil parameters to improve the simulations based on the observed data,
259 using whatever techniques they normally use, and to document the changes. At this stage, models

260 were split into two categories: a) with spin-up (SP) and b) without spin-up (NS). Both SP and NS
261 models require an initial estimate for SOC content and/or an adjustment of parameters towards
262 balancing the split between soil C pools. The two classes of models work in the same way using
263 information about plant residues and root growth that provide the C substrate for SOC dynamics
264 simulations. NS-type models (e.g. DNDC and RothC) use the initial measured SOC value, where
265 estimates of C inputs in the background of model runs are obtained with various methods (e.g.
266 Keel et al., 2017) in order to initialise the SOC pools, which can sometimes be calculated
267 analytically. In order to keep the legacy effect of previous land-use and past management practices,
268 in SP models (e.g. DayCent) SOC pools are routinely initialised by running the models to achieve
269 their own states of equilibrium, where change in C stocks is minimised (e.g. Lardy et al., 2011;
270 Huntzinger et al., 2013). However, if soils are not at equilibrium (e.g. after a sudden disturbance),
271 spin-up runs may not always be valid with the risk of starting simulations with biased initial values
272 (e.g. Wutzler and Reichstein, 2007; Nemo et al., 2017) but a fuller discussion on the “spin-up
273 problem” (Reynolds et al., 2007) is not within the scope of this paper. Carbon inputs are usually
274 estimated through sub-models calculating total net primary production (TNPP). As it was not
275 possible to derive TNPP data from local sources at each study-site, TNPP estimates were obtained
276 at each site (Table 2) based on precipitation levels according to the approach of Del Grosso et al.
277 (2008). In this way, the creation of the TNPP database used by modellers was based on an identical
278 methodology, which is widely used worldwide, though the uncertainty in quantifying productivity
279 across ecosystems is highlighted (e.g. Wieder et al., 2014).

280 The distinction between SP and NS models can appear somewhat arbitrary as virtually any
281 model with more than one C pool could be spun-up or, alternatively, a function (or analytical
282 procedures) can be used to make an initial pool partition. We refer here to common modelling
283 practice, as performed by users within the constraints imposed by packaged (operational) solutions
284 of SOC models (for which spin-up procedures may be operationally more difficult) or relying on

285 the procedure suggested by previous experience. For instance, although spin-up equilibrium runs
286 are documented for RothC (e.g. Herbst et al., 2018), it is common practice to initialise three C
287 pools for subsequent simulations through an internal routine over 10,000 years, with limited model
288 inputs including clay fraction and weather, and a pre-defined ratio of decomposable over
289 recalcitrant plant material (e.g. Xu et al., 2011; Weihermüller et al., 2013). Modellers were left to
290 choose one option or the other when both were available for use in their models (e.g. C-TOOL).
291 About 40% of the models (10 models) in the study did not use SP processes and set the initial SOC
292 values manually (using the initial SOC observation).

293 For each model category (SP and NS), two main modelling approaches were identified: site-
294 specific *versus* generic (single set of parameter values for all the sites). For the site-specific
295 approach, at each site users informed models about historical management practices and land uses
296 such as grassland or cropland (with both SP and NS models), SOC decomposition parameters (only
297 for SP models) or the partitioning of C among different soil pools (only for NS models). With the
298 generic (not site-specific) approach, model calibration was not applied separately for each
299 experimental site but simultaneously on all available multi-location datasets to find for each model
300 parameter values that would be applicable at regional scales. In this case, multi-location calibration
301 was used to capture generic model parameter values so that the models could still perform well
302 across a range of climate and management conditions in Europe (Dechow et al., 2019). Site-
303 specific and non-site-specific approaches were variously combined with factors affecting model
304 initialisation/parameterisation (Table 3) to create simulation scenarios Gen (generic), Mix (mixed)
305 and Spe (specific).

306 Scenario Mix uses a site-specific approach for the initialisation of C pools with both SP and
307 NS models and, for each model, a unique calibration of decomposition parameters. Fixed
308 decomposition rate parameters (but not rate modifiers) were maintained at a constant value
309 throughout all sites (e.g. the maximum passive pool decomposition rate in M25 was set to 0.003

310 yr⁻¹ at all sites), while site-specific climate and soil textural conditions provided supplementary
 311 factors driving the actual decomposition curve (likely in the uncalibrated blind simulations as
 312 well). In scenario Spe, decomposition rates could be changed separately at each experimental site,
 313 which constrained the modelling to a fitting exercise, but made it possible to explore the spatial
 314 variability of model parameters. Scenario Gen ignored base histories of each site: arable crops and
 315 grasslands were not distinguished, past climate conditions were disregarded, and this translated
 316 into discounting the variability in the TNPP levels among sites affecting the starting SOC level.
 317

318 Table 3. Modelling approaches and simulation scenarios for spin-up and no spin-up models (Gen:
 319 generic; Mix: mixed; Spe: specific).

Model category	Factors	Approaches	Calibration scenarios ^a		
			Gen	Mix	Spe
Spin-up (SP) based models	Historical management/land use	Site-specific		X	X
		Non-site-specific	X		
	Decomposition processes	Site-specific			X
		Non-site-specific	X	X	
No spin-up (NS) based models	Partitioning of C pools	Site-specific		X	X
		Non-site-specific	X		
	Decomposition processes	Site-specific			X
		Non-site-specific	X	X	

320 ^a The term ‘generic’, which refers to calibration, here means ‘ubiquitous’ or ‘universal’, since the aim of any model
 321 is to work well under all conditions, without the need to adjust decomposition coefficients. In this case, the model
 322 correctly represents the main processes and integrates the main factors to accurately simulate the C cycle. The
 323 ‘specific’ calibration, which aims at improving the model performance, implicitly suggests an incomplete knowledge
 324 of the SOC turnover. The ‘specific’ calibration allow exploring the spatial variability of model parameters, but this

325 amplitude (which is not discussed or reported here) may indicate the extend of degree of the knowledge gap in soil
 326 processes (i.e. model parameters might need a huge adjustment across sites)

327
 328 Twenty-six modelling teams participated in the blind test. At calibration stage, 17 teams
 329 completed scenarios Spe and Mix, and 16 the scenario Gen. Some model packages are set to restrict
 330 access to individual parameter values, which did not allow users to carry out some site-specific
 331 scenarios (Mix and Spe). The same outputs were obtained with some models (e.g. RothC, DNDC),
 332 which run blind and generic simulations with non-specific information like the previous land-use
 333 type (arable crop or grassland) and the historical climate. When results from the blind test were
 334 exactly equal to outputs from Gen scenario, they were not included for further analysis. Estimated
 335 and observed SOC values (Mg C ha^{-1}) were compared at blind test and for each calibration
 336 scenario. The agreement between simulations and observations was evaluated by the inspection of
 337 time series graphs and, numerically, through a set of performance metrics (Table 4) combining
 338 difference- and correlation-based metrics (e.g. De Jager et al., 1994; Moriasi al., 2007;
 339 Confalonieri et al., 2009; Bellocchi et al., 2002, 2010).

340
 341 Table 4. Model performance metrics (P, predicted value; O, observed value; n, number of P/O
 342 pairs; i, each of P/O pairs; \bar{O} , mean of observed values; \bar{D} , average of the differences between
 343 predicted and observed values; S_D , standard deviation of the differences between estimated and
 344 observed values).

Performance metric	Equation	Unit	Value range and purpose
RRMSE, relative root mean square error	$\text{RRMSE} = 100 \cdot \frac{\sqrt{\frac{\sum_{i=1}^n (P_i - O_i)^2}{n}}}{\bar{O}}$	%	0 (optimum) to positive infinity: the closer the values are to 0, the better the model performance

(Willmott and Wicks, 1980)	closer the values are to 1, the better the model
----------------------------	--

345

346 **2.4. Multi-model and ensemble assessment**

347 We first focussed on the quantification of model-data discrepancies and then assessed the

348 uncertainty of the individual models in comparison with the multi-model ensemble. The modelling

349 teams provided deterministic model simulation results according to the protocol established, which

350 meant that: 1) one run was provided for each site; 2) the spread of model results due to parameter

351 uncertainty was not specifically addressed. The latter would have dramatically increased the range

352 of model outputs used within the study and would have confounded the uncertainty in calibrated

353 parameters with the uncertainty in model structure (Wallach and Thorburn, 2017). While the

354 uncertainty in model predictions could be due to parameterisation, model calibration from different

355 users (i.e. ensemble of users within ensemble of models) cannot be regarded as the solution to

356 estimate uncertainty due to parameterization (Confalonieri et al., 2016). As well, different

357 calibration techniques do not seem to be primarily responsible for differences in model

358 performance (Wallach et al., 2020) and the contribution of the initialisation to the uncertainty in

359 SOC changes can be negligible compared to the uncertainty related to the model itself and

360 simulated systems characteristics (Dimassi et al., 2018). As uncertainty could not be associated

361 with any individual simulation, we focussed on the analysis of model residuals. We documented

362 the variability of the multi-model simulation exercise across two stages (blind test and alternative

363 calibration scenarios), while inspecting how the multi-model median (MMM) converged to the

364 observations. We used box-plots to compare the variability of estimates by different models (with

365 focus on multi-year averages) to the observed variability, and we represented model ensembles

366 with MMM, which has the advantage to exclude distinctly biased model members with a

367 disproportionate influence on the mean (Rodríguez et al., 2019). The advantage of using MMM

368 was established in practical studies in crop and grassland modelling but also on a theoretical basis
369 (Wallach et al., 2018).

370 We also quantified the relationship among standardised model residuals of SOC, based on
371 uncalibrated (BIn) and calibrated (Gen, Mix, Spe) simulations. Moreover, we quantified the
372 relationship between residuals of agro-climatic metrics (annual values): temperature amplitude,
373 mean maximum temperature and annual precipitation. Arrays of pairwise scatterplots (scatterplot
374 matrices) were generated with the panel plot option in the R language and environment for
375 statistical computing ('panel.smooth', [https://stat.ethz.ch/R-manual/R-](https://stat.ethz.ch/R-manual/R-devel/library/graphics/html/panel.smooth.html)
376 [devel/library/graphics/html/panel.smooth.html](https://stat.ethz.ch/R-manual/R-devel/library/graphics/html/panel.smooth.html)), which also overlaid a local non-parametric
377 smoother curve (locally estimated scatterplot smoothing) on each plot to give some indication of
378 trends (after Cleveland, 1979).

379 To explore how MMM varied with the number of models in the ensemble, we performed a
380 calculation for each z -score transformed MMM, $z = \frac{MMM - \bar{O}}{sd_{obs}}$, which was obtained by dividing the
381 multi-model data deviation from the mean of observations (\bar{O}) by the standard deviation of the
382 observations (sd_{obs}) (Sándor et al., 2020). A z -score can be placed on the normal distribution curve
383 to indicate how much it deviates from the mean of the distribution. The units of a z -score are sd
384 units: zero equals the mean, positive z -scores exceed the mean, and negative z -scores are less than
385 the mean. A z -score allows comparisons to be made between combinations of models with different
386 distribution characteristics, i.e. different \bar{O} and sd_{obs} (used here as practical descriptors of time-
387 series central tendency and spread). As illustrated in Fig. 2, different sites occupy distinct zones in
388 the sd_{obs} versus \bar{O} space. Low variability and low mean SOC observations were found at Askov
389 (S1, S2), Grignon (S3) and Utuna (S6). The variability was higher at Rothamsted (S5) and
390 Versailles (S7), while the mean was the highest at Kursk (S4). None of the site occupies the upper
391 right quadrant, i.e. high variability and high mean.

392

393 (Fig. 2 here)

394
395 We calculated z -scores for all possible combinations of sets of k out of $n=26$ models ($k=2, \dots n$).
396 The minimum number of models providing plausible estimates at each site was that for which the
397 z -scores lay within the ranges -1 to $+1$ or -2 to $+2$. The arbitrary choice of these thresholds was
398 due to a conventional rule, for which values falling within 1 and 2 times the standard deviation
399 approximate the 68% ($|z|=1$) and 95% ($|z|=2$) confidence limits of a normal distribution,
400 respectively (after Ehrhardt et al., 2018). R software (<https://cran.r-project.org>) was used for
401 statistical analysis and graphical visualization.

402

403 **3. RESULTS**

404 **3.1. Evaluation of SOC dynamics**

405 Fig. 3 show the range of model results (represented by the shaded area) for each scenario and the
406 multi-model median (MMM hereinafter) together with the measured values. In general, the
407 greatest spread of model results was found under the Bln scenario, followed by the Gen scenario.
408 In some cases, the multi-model median of Bln and Gen scenarios overestimate observations (e.g.
409 at S5, S6 and S7 sites). As expected, the tightest range of model results (simulation envelope) was
410 found with site-specific simulations. MMM simulations of Spe came closest to the observations.
411 All the MMM lines were remarkably close to the observations at sites S1, S2 and S3 (Fig. 3),
412 despite the much wider spread of the individual simulations, while the MMM at other sites differed
413 more substantially from the observations (e.g. S5, S6 and S7, Fig. 3). Overall, most of the
414 simulations (Bln, Gen and Mix) tended to overestimate the amount of SOC (e.g. S5, S6 and S7,
415 Fig. 3).

416 SOC stocks decreased under all bare-fallow sites during the investigated period. At S1, S2,
417 S3, S4 and S6 (Fig. 3) sites, the decrease in SOC stock was from minimum to moderate whereas

418 at S5 and S7 (Fig. 3) SOC loss in the top 0.20 m was more rapid, with initial SOC halved during
419 ~30 years. The decay tended to be more rapid in the first years and then the rate of loss decreased
420 (e.g. at S7 site between 1929 and 1962, Fig. 3).

421 (Fig. 3 here)

422

423 **3.2. Ensemble performance by site**

424 Fig. 4 shows a high variability in the multi-model spread of responses at different sites. The results
425 show that Kursk (S4) soil, which stored the highest amount of SOC, 91.8 Mg C ha⁻¹, was
426 approximated well by the models, mainly with calibration scenario Spe, with a MMM value of
427 90.1 Mg C ha⁻¹. For calibration scenario Gen, some underestimation is apparent (84.2 Mg C ha⁻¹).
428 Site S4 had the narrowest variability in the measured values, whilst the Bln simulation and
429 calibration scenario Gen had the highest variability. Measured SOC was well estimated at S1, S2
430 and S3, including with blind simulations, despite several outlying dots, mainly with Bln and Gen
431 scenarios. The MMM tended to overestimate the measured SOC at S5 (42.5 Mg C ha⁻¹) and S7
432 (33.0 Mg C ha⁻¹) with some scenarios: Bln, S5: 56.7 Mg C ha⁻¹, S7: 44.49 Mg C ha⁻¹; Mix scenario,
433 S5: 50.0 Mg C ha⁻¹, S7: 35.5 Mg C ha⁻¹; Gen scenario, S5: 52.1 Mg C ha⁻¹, S7: 40.0 Mg C ha⁻¹.
434 On the other hand, the MMM of Gen scenarios showed the closest values to the observed median
435 at S5 and S7 (Fig. 4.).

436 Overall, with some exceptions, the MMM of calibrated runs were within the range of the
437 25th and 75th percentiles of observations. The Spe scenario provided the best MMM estimation.

438 (Fig. 4 here)

439 **3.3. Individual models versus multi-model ensemble**

440 The scatterplot analysis for both each model and the MMM shows that SOC estimates were
441 improved when moving from the Bln runs (Fig. 5) to the calibration Spe scenario (Fig. 6). Model
442 performances for calibration Mix and Spe scenarios also showed better simulation results than the

443 Bln simulations (see also Appendix A and Appendix B). Considering all the sites and years, the
444 predictions of some of the models (e.g. M02, M13, M22, M24 and MMM) were close to the
445 observations even for the blind level simulations (correlation coefficient >0.9 , Fig. 5). Simulations
446 improved even further (correlation coefficient >0.98 for half of the models, Fig. 6) under scenario
447 Spe.

448 All the correlation coefficients of the simulations by other models also considerably improved with
449 the site-specific data and got closer to the 1:1 line. For instance, for M31, the spread of simulation
450 data in the blind simulations (Fig. 5) was mainly caused by incorrect initial SOC estimates for the
451 different sites. When the model was re-run with correctly set initial SOC amounts (Fig. 6), the
452 subsequent drawdown of SOC over the bare-fallow period was estimated fairly well.

453 Even with blind simulations, MMM gave results in agreement with the observations ($R^2=0.94$).
454 This level of agreement was only exceeded by M22 ($R^2=0.95$) and approached by M02 ($R^2=0.92$)
455 and M13 ($R^2=0.90$). The MMM simulations continued to give the closest agreement with the
456 observations even under the full site-specific calibrations ($R^2=0.99$) with several other models
457 performing equally well (i.e. M02, M05, M09, M13, M23, M26). Overall, with some specific
458 information for model calibration, many models did remarkably well in reproducing the observed
459 patterns of SOC loss over time.

460

461 (Fig. 5 here)

462

463 (Fig. 6 here)

464

465 **3.4. Analysis of model residuals**

466 The plots of the discrepancy between MMM and observations (Fig. 7) as a function of time shows
467 a limited scatter (within ± 1) at each site. While Bln, Gen and Mix scenario overestimated the SOC

468 decomposition rate at Kursk (where the highest SOC content was measured), the standardized
469 residuals were around zero at Grignon and both Askov sites during the whole of experimental
470 period. However, the departure from observations may increase over time especially with Bln and
471 Gen scenarios at some site (e.g. at Rothamsted, Ultuna, Versailles) indicating that models
472 underestimate decomposition rates after a few years/decades.

473

474 (Fig. 7 here)

475

476 Model residuals displayed one versus the other can help establish relationships by exploring the
477 correlation of residuals from different modelling scenarios, both among them and with external
478 drivers. Residuals of blind test and calibration scenarios calculated from MMM (Fig. 8) and
479 individual models (Figs. B1-26 in the supplementary material) were correlated with the mean
480 annual climate indicators such as the precipitations, maximum temperatures and temperature
481 amplitudes. When considering the MMM, residuals of Bln were strongly correlated with Gen
482 ($r=0.90$) and with Mix ($r=0.59$) residuals, but less with Spe ($r=0.25$) residuals, indicating a higher
483 similarity of the first three approaches, while residuals of Spe were more correlated with those of
484 Mix ($r=0.65$) than of Gen ($r=0.39$).

485 The most prominent effect of annual climate indicators was found at the blind test stage, whose
486 residuals were negatively correlated with precipitation ($r=-0.17$) and positively correlated with
487 Tmax ($r=0.41$). Combinations of high maximum air temperature and low precipitation values may
488 thus generate greater errors in blind SOC simulations. Calibration scenario Gen did not show
489 significant correlations to climate indicators. However, calibration scenario Spe and Gen had
490 opposite correlations. The annual precipitation positively correlated with Spe residuals ($r=0.26$)
491 and with scenario Mix ($r=0.15$). Annual maximum temperature and scenario Spe negatively
492 correlated ($r=-0.10$). These correlations with climate indicators hint that the site-specific

493 calibration (scenario Spe) is more sensitive to precipitation than to maximum temperatures. On the
494 contrary, Bln and Gen simulation residuals showed greater sensitivity to maximum temperatures.
495 Residuals of individual models were approximately equally influenced by precipitation and
496 temperature drivers, but with differences among models and scenarios (Figs. B1-26 in the
497 supplementary material). In most of the cases, model residuals were positively correlated with
498 annual maximum temperatures and negatively correlated with annual precipitation totals (e.g.
499 M03, M09, M18, M22 for Bln). In some cases, e.g. M09 (Fig. B8 in the supplement), the
500 correlations among SOC residuals for different scenarios were both positive and negative (r values
501 ranged from -0.043 to 0.36), and even the effect of climate indicators were different (e.g. for T_{max} ,
502 r values ranged from -0.096 to 0.65). In other cases, e.g. M25 (Fig. B18 in the supplement), SOC
503 residuals were more similar to each other (r -values 0.17-0.80) and the effect of precipitation and
504 temperature drivers was often important (with $r > 0.4$). It is interesting in this respect that the Spe
505 residuals had near-zero correlations with climatic drivers, showing a lesser influence of these
506 factors on model results with this scenario, whereas the Bln scenario showed some correlations
507 with T_{amp} ($r=0.13$), T_{max} ($r=-0.44$) and precipitation ($r=0.40$). For M25, Gen scenario residuals
508 (Fig. B18 in the supplement) appeared unrelated with precipitation (r -value near zero), but not with
509 temperature amplitude ($r=0.50$) and maximum air temperature ($r=-0.56$).

510

511 (Fig. 8 here) .

512

513 **3.5. Minimum ensemble size**

514 We attempted to identify the minimum number of models required to obtain reliable results for
515 Bln and calibration scenarios Mix, Spe and Gen (Fig. 9 and Appendix C-E). We observed that
516 there could be large differences in the z -scores obtained across sites with different ensemble sizes
517 and scenarios. Overall, Bln is characterised by greater z -scores than the calibration scenarios. Our

518 analysis suggests that the ensemble size could be reduced to four models (or even fewer) at S3, S6
519 and S7. For the other sites (e.g. S4), only ensemble sizes of at least 9-10 models reduced z -scores
520 to within the range from -2 to +2, but this number should be raised to 20 or higher to comply with
521 the most stringent criterion of $z=|1|$. A minimum ensemble size of 9-10 models was also identified
522 with Gen at S4 (Fig. 9), while with Mix and Spe scenarios the number of models could be reduced
523 down to 7 and 3, respectively (up to about 14 [Gen], 8 [Mix] and 4 [Spe] to comply with $z=|1|$)
524 (Appendix C-E).

525

526 (Fig. 9 here)

527

528 4. DISCUSSION

529 4.1. Scenarios of ensemble SOC estimates

530 For Bln, Mix, Gen and Spe scenarios, the overall differences between the simulated and the
531 observed first-year SOC values were -0.46 , $+3.49$, $+2.40$ and $+1.92$ Mg C ha⁻¹, respectively, for
532 the NS models, and $+0.58$, -0.29 , $+0.95$ and -0.12 Mg C ha⁻¹, respectively, for the SP models.
533 Despite manually setting the initial SOC values (magnitude of first SOC observation for the
534 simulation period), the NS models mostly overestimated SOC content in the initial year of the
535 model run. In first-year estimates of the calibrated (mainly with Spe and Mix scenarios), SP models
536 deviated less from observations than NS models that overestimated SOC stocks for the first year
537 with the exception of M25 ($+8.4$ Mg C ha⁻¹ for Gen), M29 ($+18.6$, $+21.1$ and $+23.7$ Mg C ha⁻¹ for
538 Spe, Gen and Mix, respectively) and M31 ($+25.2$ Mg C ha⁻¹ for Gen). In the case of M25, the
539 model was run with a generic grassland spin-up (i.e. 7,000 years), which was applied to all sites.
540 Thus, a generic history was simulated without considering the cropping history at each site. This
541 spin-up protocol affected the simulated SOC, showing the poor ability of Gen scenario to produce
542 results consistent with observations, which questions the practicality of spin-up processes under

543 generic calibration. With M31, there was a greater difference between simulated and observed
544 SOC values in the initial simulation year and the model gave results that did not correspond to the
545 observations at all sites (Appendix F), especially under the Bln and Gen scenarios. Though M31
546 used the initial SOC observation as default parameter, it failed to reproduce the LTBF dynamics
547 between sites because of large differences in C input to the soil from the former vegetation during
548 the spin-up period. Consequently, the starting points of the LTBF simulations differed greatly from
549 the observations, which were overestimated at S1, S2, S3 and S6, and underestimated at S4.
550 Overall, Mix and Spe calibrations showed better performance indices than the Gen scenario
551 (Appendix F). We note, however, that M13, for which the SOC pool sizes (humads and humus)
552 were generically calibrated across sites, produced low RRMSE for Gen (5.7%).

553 The improved calibration knowledge obtained with the site-specific information also improved
554 model accuracy. Moving from Bln (with knowledge of weather and soil texture, historical land use
555 and management, and initial SOC; section 2.3) to the Gen scenario, we reproduced SOC data in a
556 number of European bare-fallow experimental sites with a single set of calibrated, regional-scale
557 parameter values (regardless of the possible soil, climate and past land-use dissimilarities between
558 different sites). According to performance indicators in Appendix F, in the Bln simulations the NS
559 models performed better than the SP models. For instance, average RRMSE and EF were 19.44%
560 and 0.60, and 26.94% and 0.24, for NS and SP models, respectively. Compared to the Bln scenario,
561 the discrepancy between the measured and estimated SOC values under the Gen scenario was
562 slightly reduced with NS models and increased with SP models. Multi-site calibration can be
563 characterised by lower uncertainty than site-specific calibration, because more data contribute to
564 the calibration process (e.g. Minunno et al., 2014; Ma et al., 2015). The availability of a variety of
565 detailed data from multiple sites thus offers the possibility of a genuine multi-location calibration
566 of the model, assuming that a single calibration across sites is appropriate. The limit of the Gen
567 scenario calibration was that it did not make it possible to explore the spatial variability of model

568 parameters. The latter was done with scenarios Mix and Spe, for which a basic requisite is that
569 model parameters are not hard coded but configuration files are left open to the users. From Gen
570 to Mix, parameters describing initial values of each pool were determined separately for each site.
571 Moving from Mix to Spe, the decomposition parameters became site-specific. Hence, modellers
572 needed to invest increasingly more knowledge (and more time-demanding calibration effort) than
573 in Gen. Under these conditions, the improvement of simulations in SP models was evident (up to
574 70% for some indicators, e.g. RRMSE and EF). On the contrary, NS models only had a slight
575 improvement in accuracy of simulations from Bln (RRMSE=21.5%; EF=0.58) to Mix
576 (RRMSE=18.6%, EF=0.55) or Gen (RRMSE=20.5%; EF=0.45). In our analysis, the two types of
577 models (NS and SP) appear to be suitable for different sets of data. NS-type models, in most cases,
578 can perform well even when data are limited to climate, initial C and historic land use, while SP
579 models generally benefit from the availability of more detailed data. All metrics related to the
580 performance of the SP models were improved with calibration. There were some differences in
581 model performance among the sites, but site-specific soil or climatic conditions cannot easily
582 explain such differences.

583 Overall, across the seven LTEs and using simulated and observed SOC data at the end of the
584 experimental period we observe that the greatest and least differences from observations were
585 approximately +14.3% with Bln and +2.2% with Spe (Fig. 10). The Gen scenario achieved almost
586 half the error (+8.9%) of its closest competitor, i.e. the Bln scenario. More than one-third of the
587 Bln-scenario error is achievable with the Mix scenario (+4.0%).

588

589

(Fig. 10 here)

590

591 This study has shown that it is difficult to define an *a priori* criterion that could be used to select
592 a subset of models that would perform better than others would. In terms of the minimum number

593 of models required to obtain reliable results, our study indicates that the suggested minimum
594 ensemble size (~10 models) proposed by Martre et al. (2015) for crop growth could be a reference
595 also when model ensembles are implemented to blindly simulate SOC in bare-fallow soils, which
596 can be reduced down to 3-4 models with a site-specific calibration. These sizes are lower than that
597 found by Sándor et al. (2020) to provide reliable C-flux estimates in croplands and grasslands (i.e.
598 ~13 models). While the current study applied the same methodology as Sándor et al. (2020), but
599 as the present study focuses on one output variable only, SOC, evaluated in simplified systems
600 (bare-fallow soils), its relative ease of simulation offers great advantages for scenario analyses in
601 the absence of vegetation cover and plant residues, nor farming practices (only occasional tillage
602 operations occurred at some sites and were considered by models which can simulate this option).
603 This is reflected in the several z-scores within the range of -2 and +2, as obtained with a limited
604 number of models, showing that reduced ensemble sizes can satisfactorily estimate the SOC
605 content in bare-fallow systems, mainly when site-specific calibration is possible. However, our
606 analysis of the Russian site (S4), which had low observed variability and high mean ($sd_{obs}=6.9$, \bar{O}
607 $=91.8$ Mg C ha⁻¹), is challenging because it showed that model ensembles that are too small might
608 not always guarantee sufficient accuracy in SOC estimates of C-rich soils. An application to the
609 peatlands located on the Mid-Russian Upland (e.g. Shumilovskikh et al., 2018) should thus be
610 considered with caution.

611

612 **4.2. Possibilities for model inaccuracies**

613 We presented an approach that uses a correlation matrix (with graphical representation) to account
614 for possible correlations between Bln, Mix, Gen and Spe residuals and, additionally, climatic
615 factors (mean air temperature amplitude, maximum air temperature and precipitation total). This
616 residual analysis helps find correlations among alternative scenarios, which might indicate
617 comparable scenarios in which error propagation within models is similar, though the way of error

618 propagation cannot be easily retrieved from the correlation matrix. This is the case of Bln, Gen
619 and Mix, whose residuals are highly correlated, while the weak correlations between Spe and other
620 scenarios highlight the distinct behaviour of the latter. This analysis can also help find correlations
621 between the SOC output and external drivers, and thus suggest additional predictors that may need
622 to be included in the models (e.g. Medlyn et al., 2005). This need emerged especially when specific
623 models were run under Bln, Gen and Mix scenarios, for which some correlations ($r > |0.4|$) were
624 obtained between model residuals and drivers of thermal and moisture conditions. A weaker but
625 significant correlation ($r = 0.26$, $p = 0.02$) was also obtained between Spe residuals and precipitation.
626 These correlations indicate some limitations related to the response functions of SOC
627 decomposition to soil temperature and soil moisture, though the relative uncertainties of our model
628 ensemble are attenuated by the presence in the models of physical and chemical processes that
629 explain the intra- and inter-annual variability of SOC. We add that such biophysical conditions
630 affect the microbial activity (e.g. Blagodatskaya and Kuzyakov, 2008; Guenet et al., 2010; Wutzler
631 and Reichstein, 2013), and care should be taken when extrapolating our results over long time
632 frames (especially without locally calibrated models, Fig. 7) if no corroborating field evidence for
633 long-term decay rates can be obtained (e.g. on how models are dealing such situations in which
634 microbes become increasingly C limited as no new C input by plants occurs; Kuhry and Vitt,
635 1996).

636

637 **5. CONCLUSIONS AND FUTURE DIRECTIONS**

638 This paper on SOC modelling offers a tentative answer to the questions about: (i) whether and to
639 what extent an ensemble of models performs better than single models, (ii) the minimum ensemble
640 size that is required to reduce the error below a given threshold, and (iii) the set of data required
641 to prepare and substantiate ensemble estimates. This study presents a framework for interpretation
642 of model performance and uncertainties obtained with a set of process-based biogeochemical

643 models (individually and in an ensemble) simulating soil C contents in bare-fallow experimental
644 systems at a variety of European sites. One of the features of SOC modelling today is the huge
645 amount and variety of models available. Although our analysis did not take into account all sources
646 of uncertainty (e.g. the influence of the unique choices made by modellers), it enabled the
647 integration of several modelling teams into an ensemble protocol. Classifying and comparing
648 different approaches have revealed great model diversity, and is the basis for the development of
649 dedicated ensemble protocols. In this model inter-comparison, the need to accommodate
650 challenges experienced by modellers (including C pools of different nature, and optional
651 initialisation and calibration procedures) was reflected in the co-creation (with modellers and data
652 providers) of alternative calibration scenarios (Mix, Gen, Spe). As far as we are aware, no previous
653 multi-model inter-comparison studies have examined differences in such calibration scenarios or
654 differences between models with or without spin-up.

655 In our study, we did not aim to identify the best model(s) for simulating SOC dynamics for bare-
656 fallows and no probability of success was assigned to prove the suitability of using one model
657 rather than another. Overall, we showed that a calibration scenario with generic system knowledge
658 was adequate for providing sufficiently reliable output, but additional site-specific knowledge can
659 further improve results under certain circumstances. This is operationally relevant because the
660 effort required to gather calibration data might no longer be feasible for modelling scenarios
661 moving from single sites to increasingly larger spatial scales. Site-specific calibration could help
662 refine model estimates. However, geographical locations have characteristics (e.g. soil and climate
663 conditions, past history) that require specific model structures and local optimisation, and the
664 application of models may be limited by the ability to provide representative parameter values.
665 Soil-C model inter-comparisons including more models and experimental data from other regions
666 should be continued to improve our ability to simulate biogeochemical processes with acceptable
667 accuracy. Additional assessments are also recommended to complete the analysis of model

668 behaviour in the long term (like thousands of years) with constant inputs. While the various models
669 evaluated here did not include all available modelling approaches used to simulate soil C
670 dynamics, the present model inter-comparison was large compared to other studies. As such, it is
671 a distinct improvement over previously published quantitative approaches because it represents a
672 reasonable sub-population of common and current approaches. In this, we offer a method to allow
673 a broad ensemble of models to be implemented using existing datasets and current modelling
674 practices. Overall, this multi-model ensemble sets a precedent for key progress in soil C modelling
675 because it provides essential information about SOC modelling and opens a path to a more in-
676 depth analysis of the response of individual models and their uncertainties against soil and climate
677 drivers. Now that we have examined SOC decomposition in-depth without the difficulties of C
678 input uncertainties, a similar modelling study should be conducted on LTEs that examine both
679 plant derived C inputs as well as C inputs from manures and other organic materials recycled in
680 agroecosystems. In fact, under field conditions, the amount of C input is not only an important
681 factor driving the changes in SOC stocks (including the changes due to tillage), but the amount of
682 C input also drives the mineralization rate of the SOC (Mary et al., 2020). How simulation models
683 compare under such conditions is important for improving our ability to evaluate and achieve
684 climate C goals. With increasing availability of data and computational resources, there are many
685 opportunities for the SOC modelling community to enrich its offering and to keep up with evolving
686 methodologies, which would significantly increase transparency of the underpinning science and
687 modelling practice. A number of recent actions are ongoing under the guidance of international
688 initiatives such as the European Joint Programme (EJP) on Soil (<https://projects.au.dk/ejpsoil>).
689 Started in 2020, the EJP-Soil is undertaking a detailed inventory of models and all available data
690 sources (e.g. world soil maps, satellite images, downscaled weather data), and appears as an ideal
691 arena to facilitate the exchange of information and to further explore SOC model developments
692 and practice.

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713

714 AUTHOR CONTRIBUTIONS

715 R. Farina, R. Sándor and G. Bellocchi coordinated the study, contributed to its design, conducted
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722 A. Taghizadeh-Toosi and E. Tsutskikh performed the model calibrations and runs.
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727 Those interested in the details of the modelling process are encouraged to contact authors.

728

729 **Data Availability Statement**

730 The data that support the findings of this study are available from the corresponding author upon
731 reasonable request and permission of the third parties (i.e. the data holders for the Long Term
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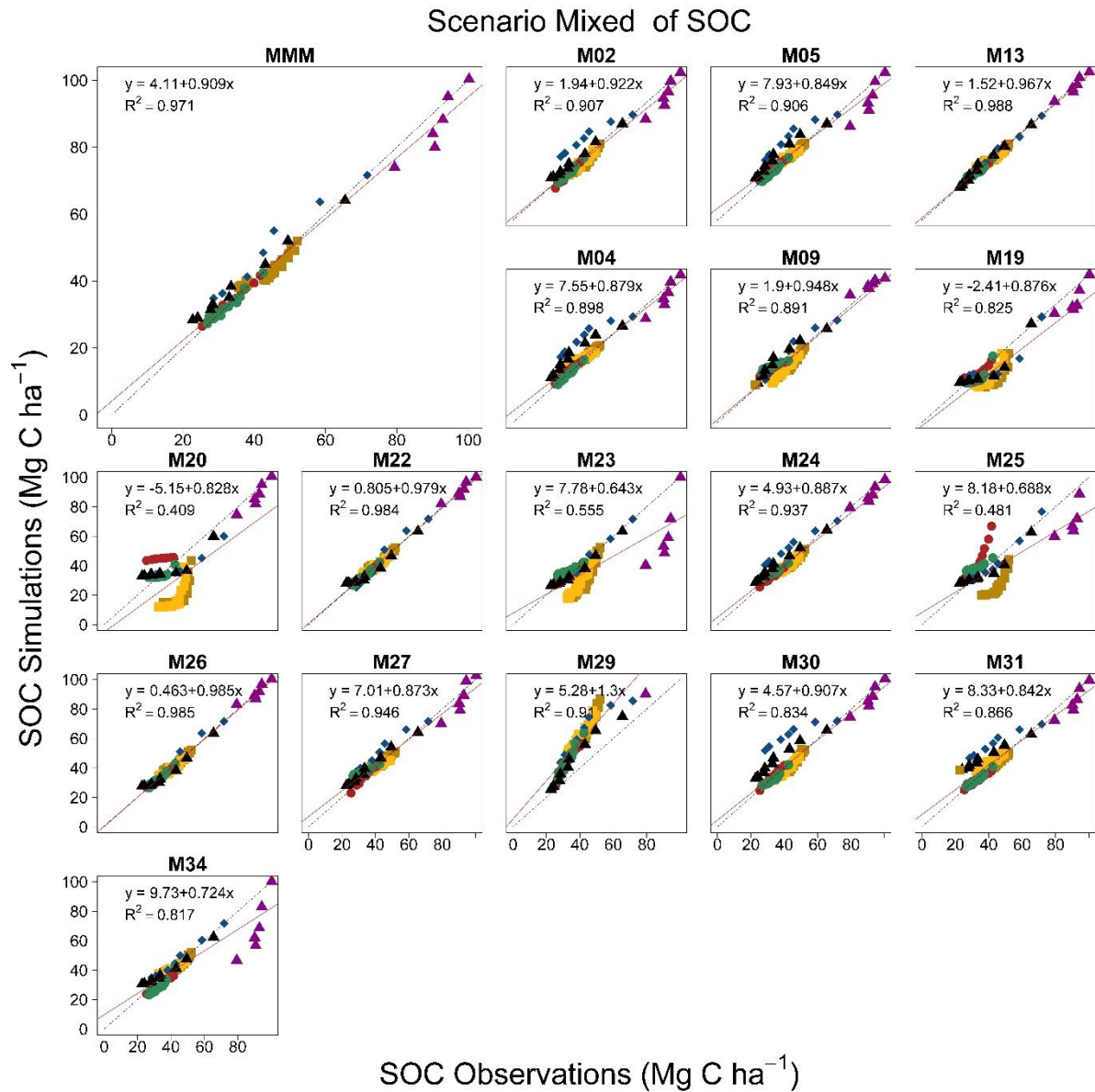
1184 [019-11103-1](https://doi.org/10.1038/s41467-019-11103-1)

1185

1186 **Appendix A**

1187 Multi-year, multi-site comparison of individual model simulation of SOC (Mg C ha^{-1}): multi-
 1188 model medians (MMM) from Mix scenario simulations (17 models) versus observations.
 1189 (coloured symbols represent sites as in Fig. 1).

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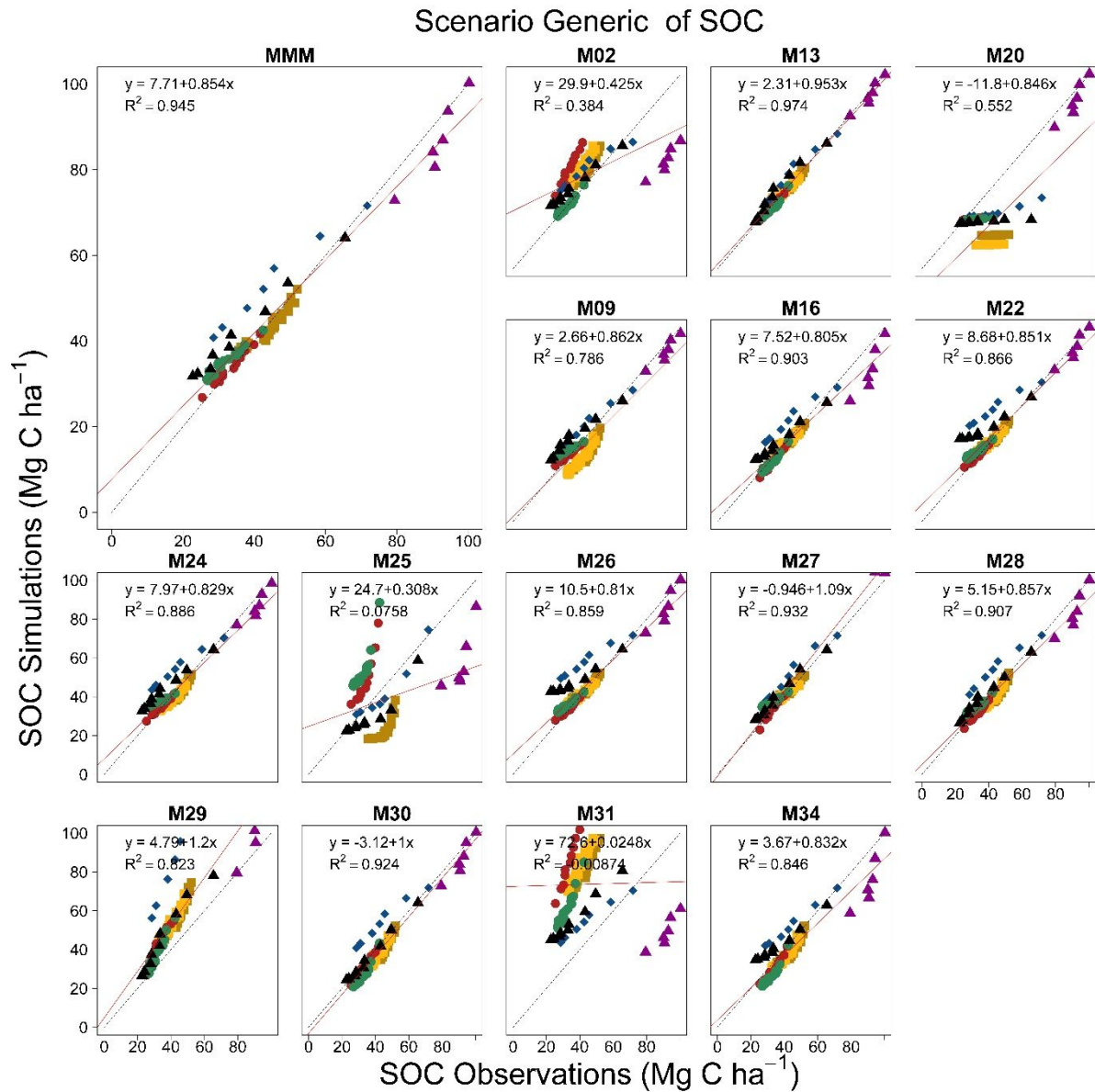


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1192 **Appendix B**

1193 Multi-year, multi-site comparison of individual model simulation of SOC (Mg C ha^{-1}): multi-
 1194 model medians (MMM) from Gen scenario simulations (16 models) versus observations.
 1195 (coloured symbols represent sites as in Fig. 1).

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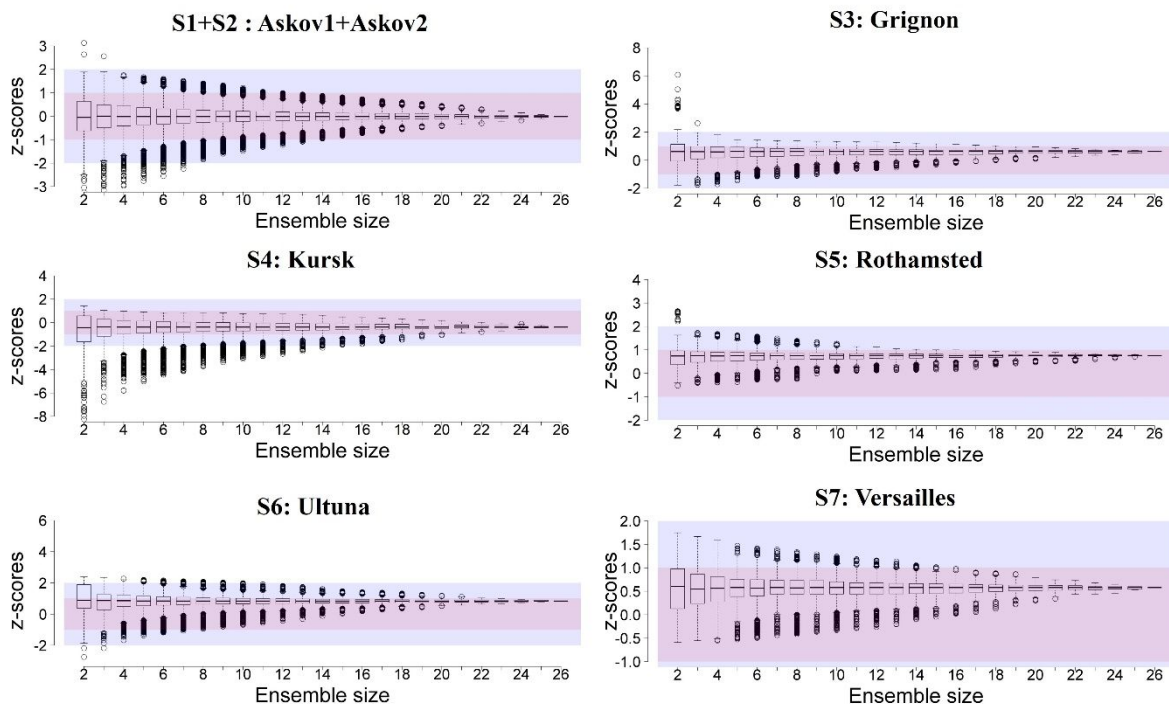


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1199 **Appendix C**

1200 z-scores calculated with different ensemble sizes for SOC estimates obtained with BIn scenario at
 1201 different experimental sites. Black lines show median values. Boxes delimit the 25th and 75th
 1202 percentiles. Whiskers are 10th and 90th percentiles. Circles indicate outliers. Coloured bands mark
 1203 two critical values: $z=|1|$ (light purple) and $z=|2|$ (light blue).

Blind scenarios

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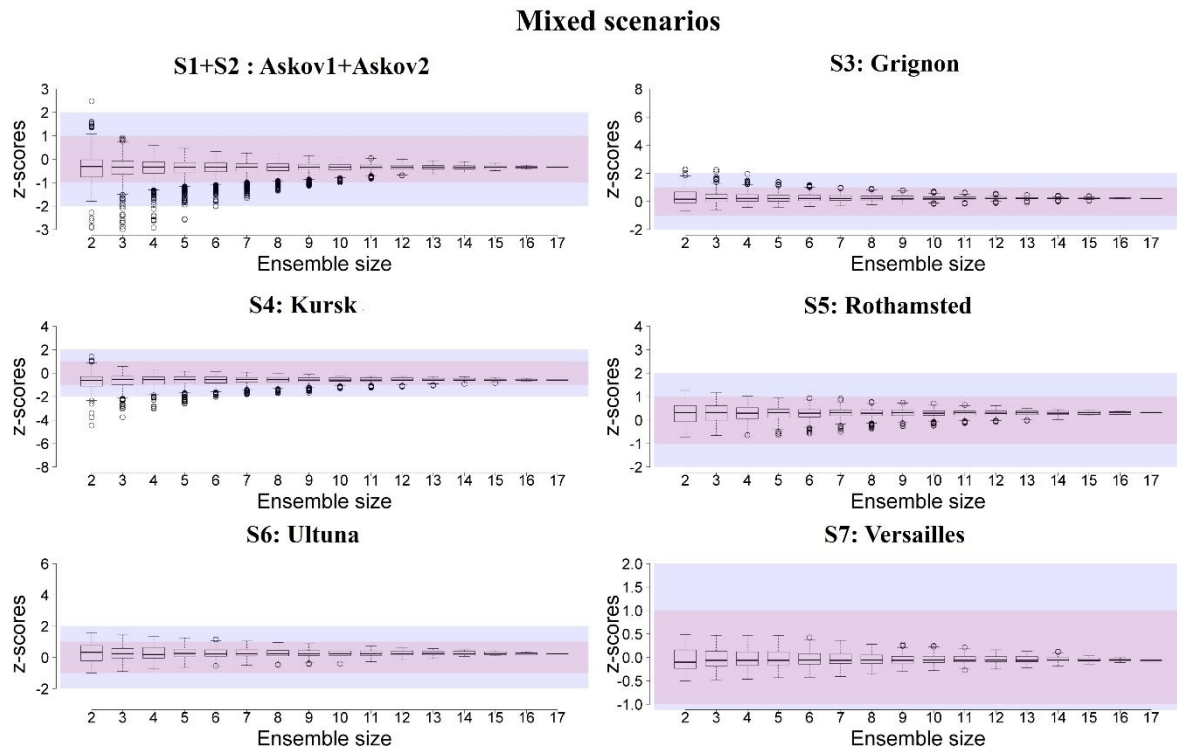
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1208 **Appendix D**

1209 z -scores calculated with different ensemble sizes for SOC estimates obtained with Mix scenario at
 1210 different experimental sites. Black lines show median values. Boxes delimit the 25th and 75th
 1211 percentiles. Whiskers are 10th and 90th percentiles. Circles indicate outliers. Coloured bands mark
 1212 two critical values: $z=|1|$ (light purple) and $z=|2|$ (light blue).

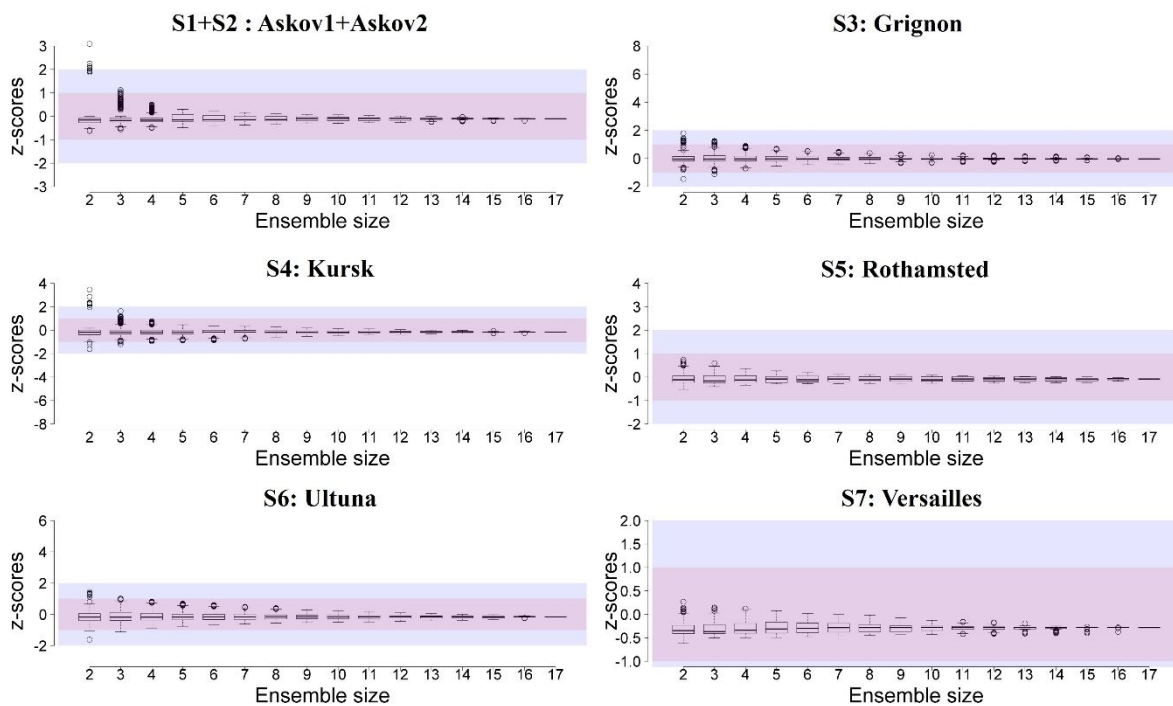


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1215 **Appendix E**

1216 z -scores calculated with different ensemble sizes for SOC estimates obtained with Spe scenario at
 1217 different experimental sites. Black lines show median values. Boxes delimit the 25th and 75th
 1218 percentiles. Whiskers are 10th and 90th percentiles. Circles indicate outliers. Coloured bands mark
 1219 two critical values: $z=|1|$ (light purple) and $z=|2|$ (light blue).

Specific scenarios

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1221

1222

1223 **Appendix F**

1224 Individual and multi-model ensemble (MMM) performance metrics (as in Table 4) for blind (Bln)

1225 and calibration scenarios (Mix, Spe and Gen as in Table 3) across sites. Red (*italic*) and blue (**bold**)

1226 numbers indicate the worst and best performances by metric, respectively.

Performance metric Scenario	Model																									MMM	
	M01	M02	M03	M04	M05	M06	M07	M09	M12	M13	M16	M18	M19	M20	M22	M23	M24	M25	M26	M27	M28	M29	M30	M31	M32		M34
R ²																											
Bln	0.73	0.92	0.67	0.83	0.79	0.86	0.76	0.89	0.83	0.90	0.33	0.81	0.69	0.63	0.95	0.76	0.92	0.41	0.86	0.76	0.92	<i>0.21</i>	0.82	0.35	0.57	0.80	0.94
Gen	NA	0.39	NA	NA	NA	NA	NA	0.79	NA	0.97	0.90	NA	NA	0.56	0.87	NA	0.89	0.09	0.86	0.93	0.91	0.82	0.93	<i>-0.00</i>	NA	0.85	0.95
Mix	NA	0.91	NA	0.90	0.91	NA	NA	0.89	NA	0.99	NA	NA	0.83	<i>0.41</i>	0.98	0.56	0.94	0.49	0.99	0.95	NA	0.91	0.84	0.87	NA	0.82	0.97
Spe	0.97	0.99	NA	0.98	0.99	NA	NA	0.99	NA	0.99	0.96	NA	NA	0.96	0.98	0.99	NA	0.91	0.99	0.97	<i>0.88</i>	0.93	0.98	0.94	NA	NA	0.99

BIn	0.88	0.97	0.84	0.93	0.90	0.89	0.90	0.97	0.93	0.97	0.71	0.94	0.85	0.79	0.99	0.89	0.97	0.73	0.95	0.91	0.95	0.59	0.95	<i>0.52</i>	0.85	0.93	0.98
Gen	NA	0.71	NA	NA	NA	NA	NA	0.93	NA	0.99	0.97	NA	NA	0.66	0.96	NA	0.97	0.53	0.95	0.97	0.97	0.81	0.97	<i>0.23</i>	NA	0.94	0.98
Mix	NA	0.97	NA	0.96	0.97	NA	NA	0.97	NA	-1.00	NA	NA	0.89	<i>0.69</i>	-1.00	0.79	0.98	0.81	-1.00	0.98	NA	0.76	0.96	0.96	NA	0.93	0.99
Spe	0.99	-1.00	NA	-1.00	-1.00	NA	NA	-1.00	NA	-1.00	0.99	NA	NA	0.99	-1.00	0.99	NA	0.97	-1.00	0.99	0.95	<i>0.76</i>	0.99	0.98	NA	NA	-1.00

RRMSE (%)

BIn	24.1	10.9	28.0	18.6	21.9	21.9	23.1	12.5	17.7	11.8	28.6	15.5	27.2	33.1	7.9	25.4	11.0	36.6	14.0	24.0	14.4	48.4	16.3	<i>69.1</i>	27.7	16.3	10.4
Gen	NA	30.8	NA	NA	NA	NA	NA	17.9	NA	5.7	11.5	NA	NA	51.3	14.0	NA	12.1	49.4	14.5	12.7	10.9	37.9	12.4	<i>92.1</i>	NA	15.8	10.6
Mix	NA	11.0	NA	12.6	11.5	NA	NA	11.7	NA	3.8	NA	NA	23.3	45.6	4.4	29.0	8.9	33.0	4.2	9.4	NA	<i>46.5</i>	14.4	13.4	NA	15.9	7.2

Spe	6.5	3.4	NA	5.0	3.2	NA	NA	3.8	NA	3.8	8.2	NA	NA	6.7	4.4	5.0	NA	14.5	4.1	6.2	14.9	46.2	5.5	8.7	NA	NA	3.2

Bln	<i>-0.00</i>	<i>-0.00</i>	<i>-0.00</i>	<i>-0.00</i>	<i>-0.00</i>	<i>-0.00</i>	<i>-0.00</i>	<i>-0.00</i>	<i>-0.00</i>	<i>-0.00</i>	0.64	0.02	<i>-0.00</i>	<i>-0.00</i>	0.31	<i>-0.00</i>	<i>-0.00</i>	<i>-0.00</i>	0.45	0.05	<i>-0.00</i>	<i>-0.00</i>	0.13	<i>-0.00</i>	<i>-0.00</i>	0.01	<i>-0.00</i>
Gen	NA	<i>-0.00</i>	NA	NA	NA	NA	NA	<i>-0.00</i>	NA	0.13	0.17	NA	NA	<i>-0.00</i>	<i>-0.00</i>	NA	0.08	0.04	<i>-0.00</i>	<i>-0.00</i>	0.06	<i>-0.00</i>	<i>-0.00</i>	<i>-0.00</i>	NA	<i>-0.00</i>	<i>-0.00</i>
Mix	NA	<i>-0.00</i>	NA	<i>-0.00</i>	<i>-0.00</i>	NA	NA	0.55	NA	0.31	NA	NA	<i>-0.00</i>	<i>-0.00</i>	0.76	<i>-0.00</i>	0.54	<i>-0.00</i>	0.31	<i>-0.00</i>	NA	<i>-0.00</i>	0.24	<i>-0.00</i>	NA	<i>-0.00</i>	0.49
Spe	0.46	0.99	NA	0.06	0.03	NA	NA	0.85	NA	0.34	<i>-0.00</i>	NA	NA	0.12	0.93	<i>-0.00</i>	NA	<i>-0.00</i>	-1.00	0.29	<i>-0.00</i>	<i>-0.00</i>	<i>-0.00</i>	0.68	NA	NA	0.83

Bln	0.52	0.90	0.49	0.72	0.60	0.60	0.56	0.87	0.74	0.88	0.33	0.80	0.39	0.09	0.95	0.47	0.90	-0.11	0.84	0.53	0.83	-0.93	0.78	-2.95	0.37	0.78	0.91
Gen	NA	0.22	NA	NA	NA	NA	NA	0.73	NA	0.97	0.89	NA	NA	-1.17	0.84	NA	0.88	-0.49	0.83	0.87	0.90	-0.19	0.87	-6.00	NA	0.79	0.93

Mix	NA	0.90	NA	0.87	0.89	NA	NA	0.89	NA	0.99	NA	NA	0.55	-0.72	0.98	0.31	0.93	0.34	0.99	0.93	NA	-0.78	0.83	0.85	NA	0.79	0.97
Spe	0.97	0.99	NA	0.98	0.99	NA	NA	0.99	NA	0.99	0.94	NA	NA	0.96	0.98	0.98	NA	0.87	0.99	0.97	0.82	-0.76	0.97	0.94	NA	NA	0.99

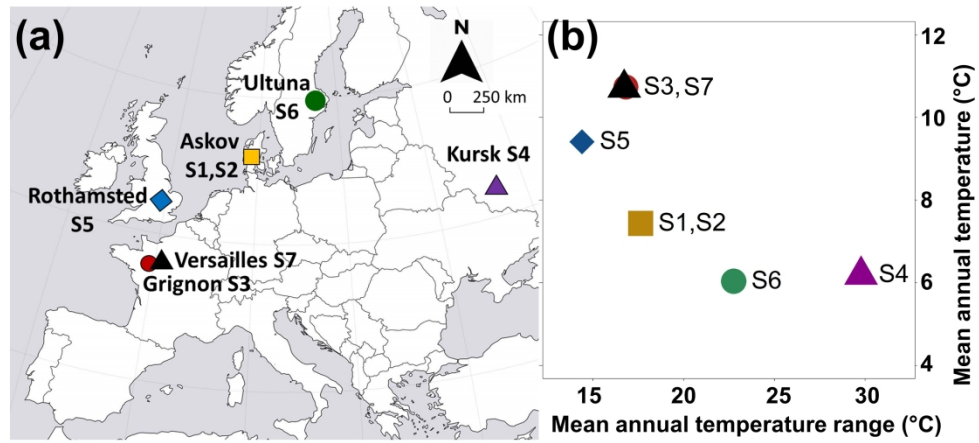


Fig. 1. Location (a) and characterisation of the study sites (b) with respect to mean annual temperature (°C) and mean annual temperature range (°C). Details about study sites are in Table 2.

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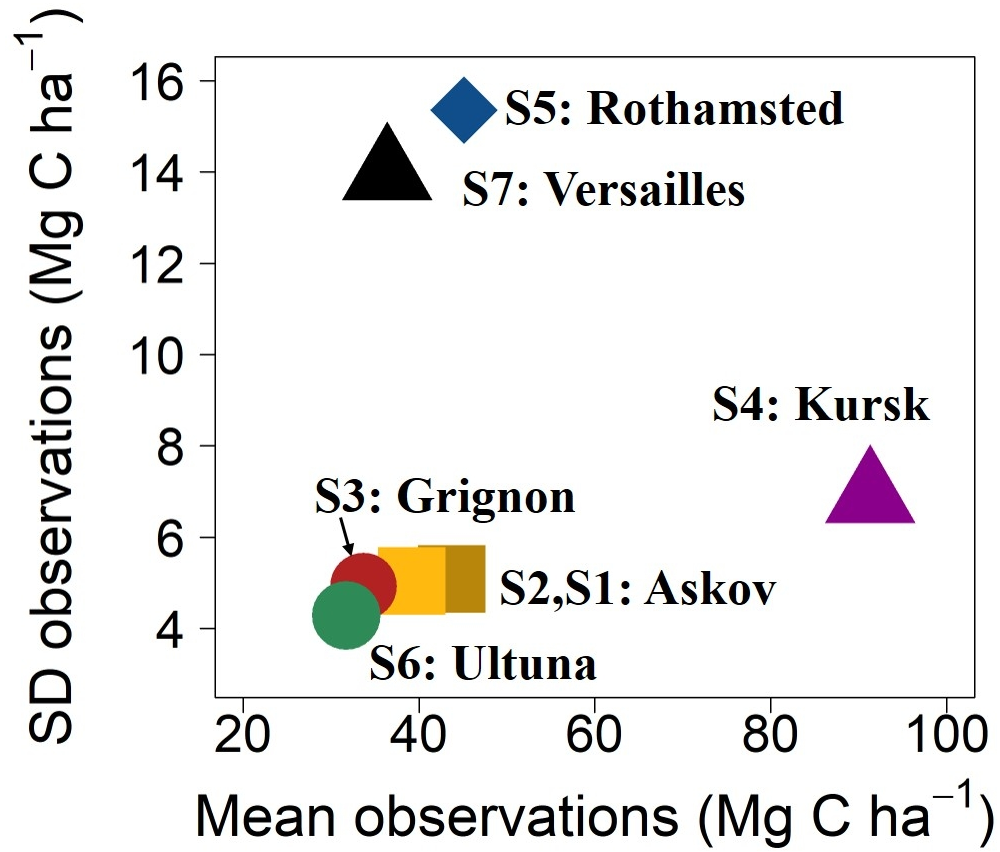


Fig. 2. Standard deviation (SD) and mean of SOC observations at the study sites (details are in Table 2).

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Fig. 3. Temporal changes of soil organic carbon (SOC, Mg C ha⁻¹) observations (Observed, purple square) and simulations: blind (Blind, blue) simulations (26 models); three calibration scenarios, Generic (16 models, pink), Mixed (17 models, green) and Specific (17 models, grey) at all sites (as in Table 2). Lines represent the multi-model median (MMM) of the simulations and shaded area represents the simulation envelope

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Fig. 4. Soil organic carbon (SOC, Mg C ha⁻¹) at each site (as in Table 2), for blind simulations (Blind, (26 models), three calibration scenarios (Mixed, 17 models; Specific and Generic, 16 models) and observations (Observed). For each boxplot, black horizontal lines show the multi-model median. Boxes delimit the 25th and 75th percentiles. Whiskers are 10th and 90th percentiles. Dots indicate outliers.

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Fig. 5. Multi-year, multi-site comparison of individual model simulation of SOC (Mg C ha⁻¹): multi-model medians (MMM) from blind simulations (26 models as in Table 1) versus observations (coloured symbols represent sites as in Fig. 1).

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Fig. 6. Multi-year, multi-site comparison of individual model simulation of SOC (Mg C ha⁻¹): multi-model medians (MMM) from Specific scenario simulations (17 models as in Table 1) versus observations (coloured symbols represent sites as in Fig. 1).

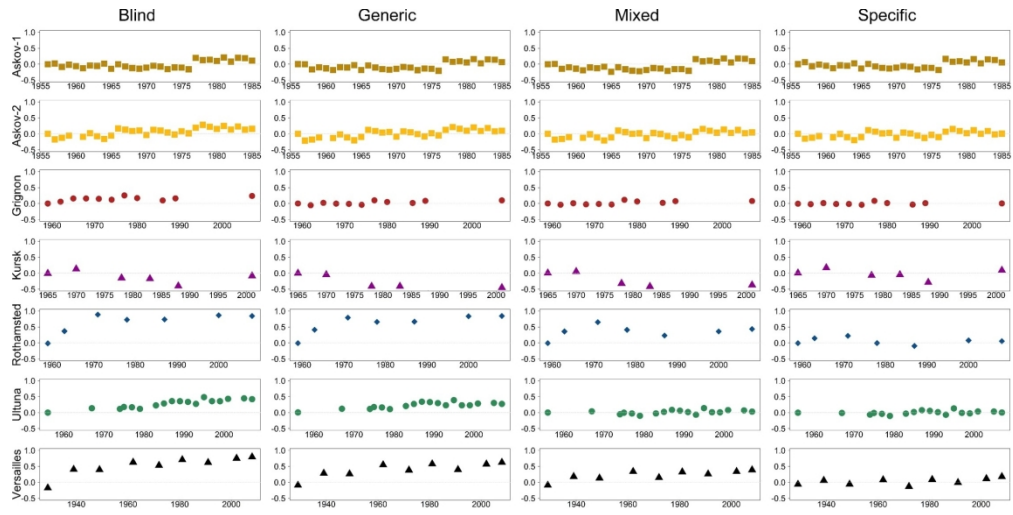


Fig. 7. Standardized model residuals $((MMM-O) / \sqrt{sd})_{obs}$ over time for blind (Blind) simulations and calibration scenarios Mixed, Specific and Generic at each site.

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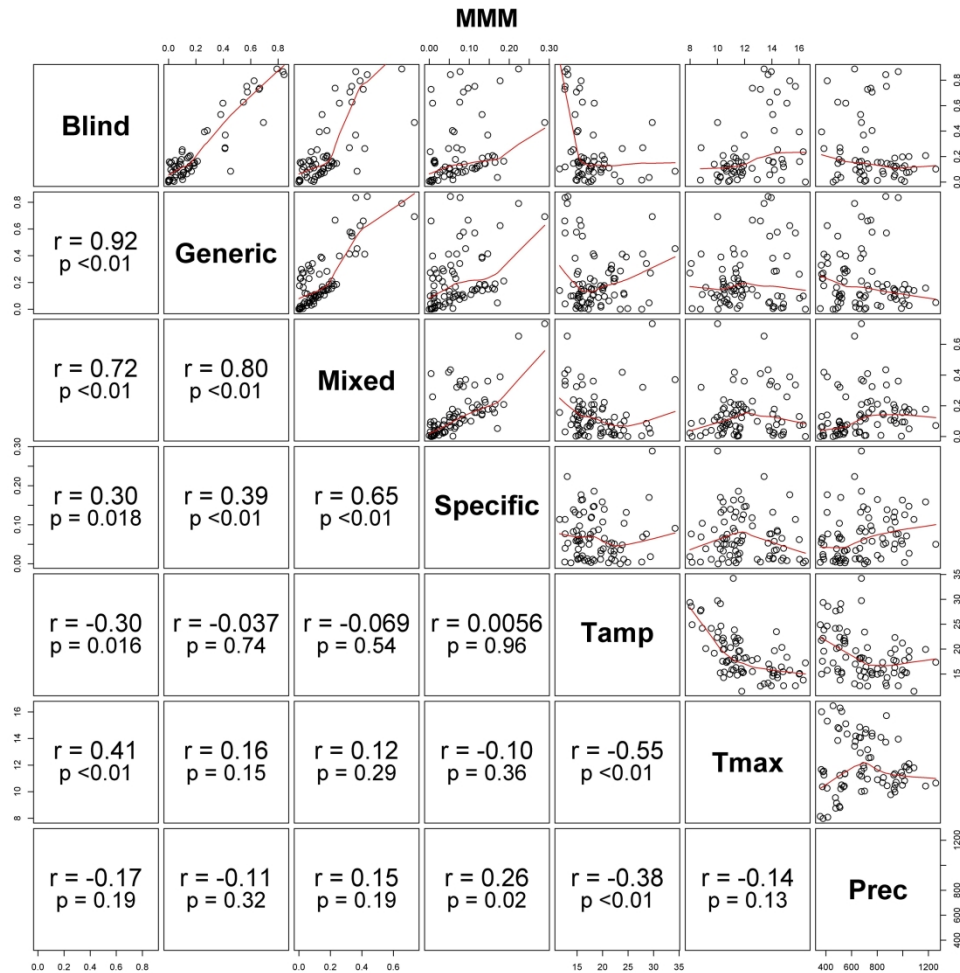


Fig. 8. Scatterplot correlation matrix of SOC (Mg C ha⁻¹) model residuals of multi-model medians (MMM) for blind simulations (Blind) and calibrations scenarios (Generic, Mixed and Specific as in Table 3), and the annual climate metrics maximum temperature (Tmax), mean temperature amplitude (Tamp) and precipitation (Prec). Overlaid (red line) is a local non-parametric smoother curve.

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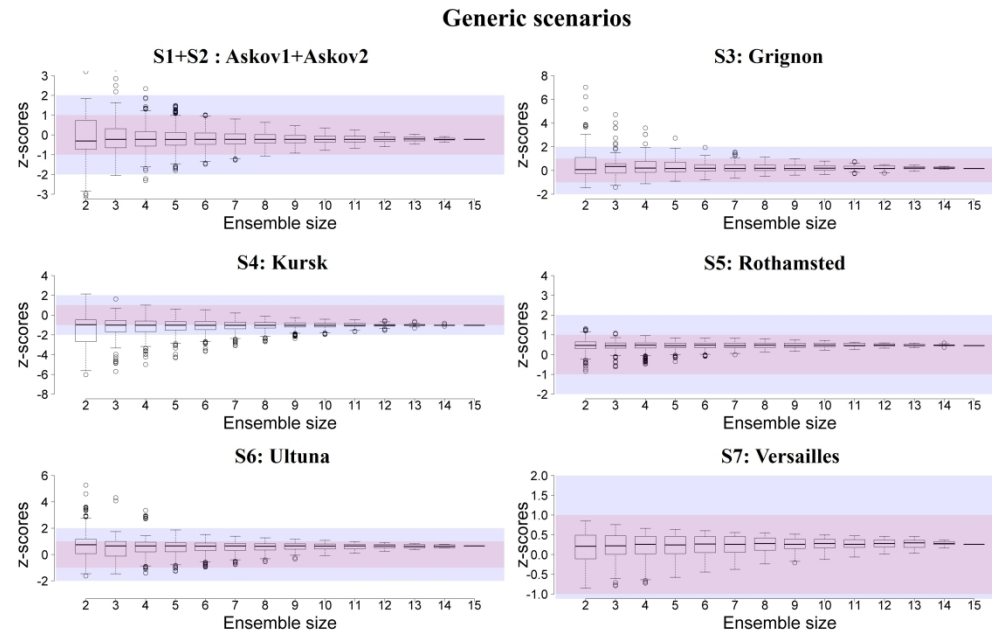


Fig. 9. z-scores calculated with different ensemble sizes for SOC estimates obtained with Generic scenario at different experimental sites. Black lines show median values. Boxes delimit the 25th and 75th percentiles. Whiskers are 10th and 90th percentiles. Circles indicate outliers. Coloured bands mark two critical values: $z=|1|$ (light purple) and $z=|2|$ (light blue).

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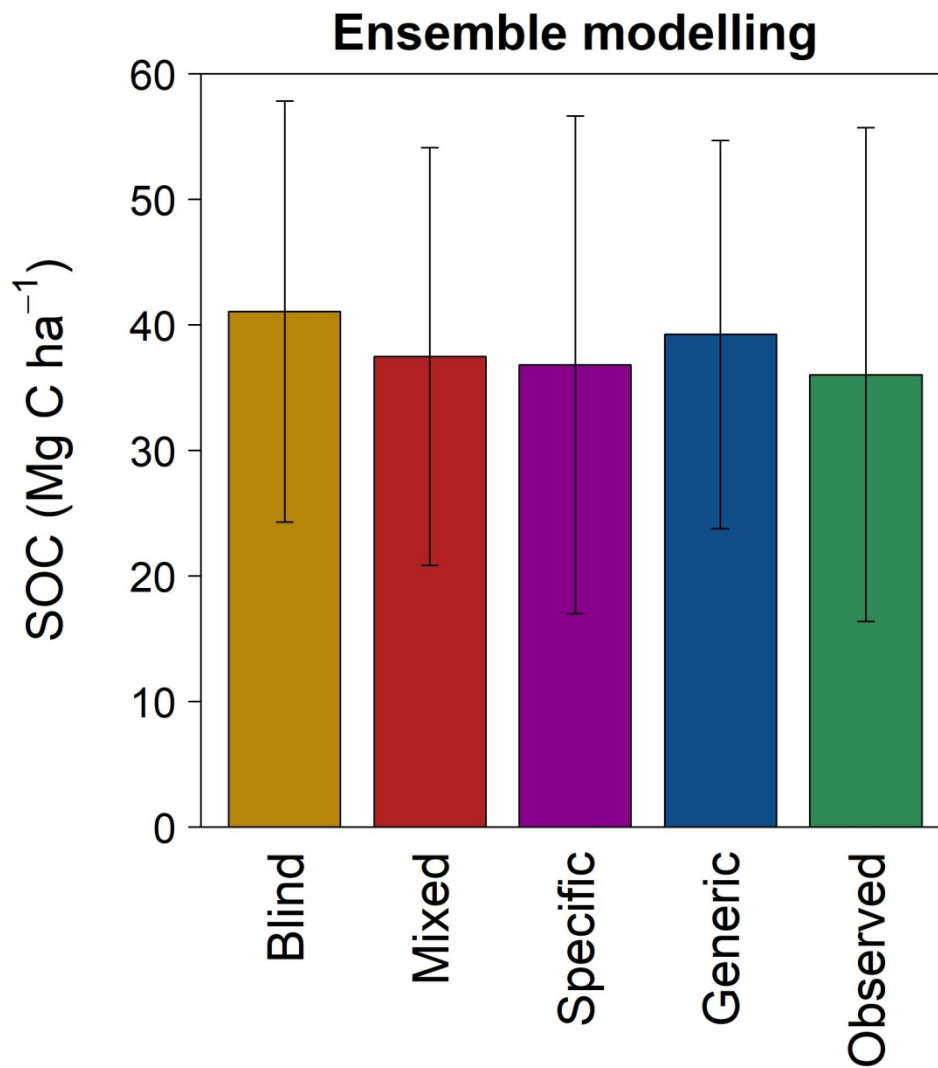


Fig. 10. Multi-site averages (vertical bars) and standard deviations (vertical lines) of observed and estimated (ensemble multi-model median) values of SOC (Mg C ha^{-1}) in the last year of the experimental period. The ensemble modelling was applied with blind simulations (Blind) and calibration scenarios (Mixed, Specific and Generic as in Table 3).

152x169mm (300 x 300 DPI)