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## A comprehensive assessment of radioactivity monitoring in the Channel Islands

Bruno Fievet, Pascal Bailly Du Bois, Claire Voiseux, Claire Godinot, Olivier Cazimajou, Luc Solier, Anne de Vismes, Catherine Cossonnet, Azza Habibi,  
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## Manuscript Details

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### Abstract

The Channel Islands are located in the Normand-Breton Gulf (NBG), in the mid-part of the English Channel (France, Normandy). In the northern part, off Cap La Hague, controlled amounts of radioactive liquid waste are discharged by the ORANO La Hague nuclear fuel reprocessing plant (RP). Radionuclides were monitored in the NBG to assess the dispersion of radioactive discharges from the RP in the marine environment. The temporal and spatial distribution of the data are consistent with the history of the discharges, with most gamma emitter radionuclide environmental levels being close to or below the current limits of detection. A clear fingerprint of H-3, C-14 and I-129 radionuclides discharged from the RP is measured. The hydrodynamics in the NBG do not yield a simple gradient with linear distance from the outfall of the RP. Modelling tools were used to understand how radioactive discharges spread from the source of input. Dispersion patterns clearly illustrate the different behaviours of soluble and non-soluble radionuclides. The study indicated that the footprint of radioactive liquid discharges by French nuclear facilities was still measurable in species collected from the NBG for the mostly dissolved radionuclides. The less conservative ones, with a high affinity for suspended matter, are potentially influenced by old releases. These pathways could be investigated by dedicated hydrodynamic dispersion models. Overall, in the Channel Islands the levels are low and consistent with the general decrease in liquid radionuclide discharges by the RP since the 1990s.

<b>Keywords</b>	Radioactivity monitoring; Marine environment; English Channel; Channel Islands; Hydrodynamic modelling; Transfer modelling
<b>Taxonomy</b>	Nuclear Waste, Environmental Science, Radioactivity in Marine Environment
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## Reply to Reviewer #1

Reviewer #1 comments:

*After reviewing the article, it would be appreciated to consider by the following minor comments:*

- 1. The title, although precise, may not express the scope of the study, which is broader than what is reflected in it. The comprehensive assessment has been developed not only in the Channel Islands, but also in the whole Normand-Breton Gulf (NBG).*
- 2. In comparing figures 9 and 10 for the year 2015 it is observed that figure 10, which makes reference to mean annual H-3 concentration in seawater (Bq.m<sup>-3</sup>) calculated by the MARS2D model with real discharges, wind and tide conditions, shows values four orders of magnitude higher than figure 9, which makes reference to mean annual seawater concentrations resulting from a theoretical constant discharge of 1 TBq.y<sup>-1</sup> (31 709 Bq.s<sup>-1</sup>) calculated by MARS2D with real wind and tide conditions. Does the latter make reference to the sum of soluble radionuclides? If affirmative, wouldn't it include H-3 and therefore the values be higher than in figure 10? In figure 9 the units are Bq/m<sup>3</sup> but in the text that accompanies figure 9 it doesn't make reference to the units of the figure. Could there be a mistake in the units of this figure?*

Revision:

- To express the broader scope of the study, the title was changed to "A comprehensive assessment of two-decade radioactivity monitoring around the Channel Islands". The reason why "the Channel Islands" was preferred to "the Normand-Breton Gulf" is that the Channel Islands are better known internationally whilst the Normand-Breton Gulf sometimes refers to the Saint-Malo Bay which may be confusing.
- The first dispersion modelling of a constant discharge of 1TBq.y<sup>-1</sup> is purely theoretical and shows the influence of the wind and tide annual changes on the average dispersion of any soluble radionuclide. It is not related to the real content of the discharges from the Reprocessing Plant. The second one however reproduces the dispersion of the real tritium discharges from the RP. Tritium discharges from the RP are in the 1E+16 Bq.y<sup>-1</sup>, so the resulting concentrations are logically four orders of magnitude higher than those resulting from 1E+12Bq.y<sup>-1</sup> discharge.  
This has been clarified accordingly in the text of the revised version.

## Reply to Reviewer #2

### Reviewer #2 comments:

The article entitled as "A comprehensive assessment of radioactivity monitoring in the Channel Islands", presents a large dataset, and use model to understand transfer and dispersion of conservative and non-conservative radionuclides. And I am really glad to have such opportunity to see this work again, since the authors put great efforts to make the big change comparing to the later version submitted to JER two years ago. This work is very important, not only for radioactivity monitoring in vicinity of ORANO RP, but also benefiting for understand the environmental behaviors and using these radionuclides as indicators of water current movement and biological metabolism. For instance, the seasonal cycle of Co-60, Ru-206, Pu and Am isotopes well reflect the resuspension process of sedimentary materials. In addition, the preparation of this manuscript is also very thoughtful and reasonable.

Therefore, I strongly to recommend this work to be published on Journal of Environmental radioactivity after minor revision.

A few comments are listed below.

1. Maybe it is better to change the title as "A comprehensive assessment of two-decade radioactivity monitoring in the Channel Islands".
2. Seaweed grows in the shallow coastal seawater, and the radionuclides concentration in it, of course, reflect the surface level and change of radioactivity. I was wondering how the author consider the vertical distribution and transfer of these radionuclides (both conservative and non-conservative ones) from the outfall
3. Line 59," The limpet parts were finally burned at 450°C to ash." For 129I analysis, the measured sample is the ash of limpet part or the raw sample?
4. Fig. 7 seems no practical meaning, since data and other figure have shown the trend. It is better to delete it for simplification.
5. Line 437-440, please add some references to support this discussion part.

### Revision:

1. The title was changed to "A comprehensive assessment of two-decade radioactivity monitoring around the Channel Islands".
2. In the Normand-Breton Gulf, the vertical distribution in the water column is not a critical issue because it is a very shallow sea. High resolution 3D water sampling around the outlet of the Reprocessing Plant showed that the water column is homogeneous after 1 500 meters distance (Bailly du Bois PB, Dumas F, Voiseux C, Morillon M, Oms PE, Solier L (2020) Dissolved radiotracers and numerical modeling in North European continental shelf dispersion studies (1982-2016): Databases, methods and applications. *Water (Switzerland)* **12**). This was added to the text in the revised version.
3. 129I was quantified on dry seaweed only but it was not measured in limpet.
4. Fig. 7 and related section were deleted for concision purpose.
5. Reference Carvalho FP (2018) Radionuclide concentration processes in marine organisms: A comprehensive review. *Journal of Environmental Radioactivity* **186**: 124-130 was added in the revised version.

## **Radioactivity monitoring around the Channel Islands**

### **Highlights**

Controlled amounts of radioactive liquid wastes are discharged by French nuclear facilities in the Channel Sea. After two decades of seaweed radioactivity monitoring in the Channel Islands, the data are put in perspective with levels in the Normand-Breton Gulf. Hydrodynamic modelling tools were used to yield a comprehensive assessment of radioactive transfer to biota in the area.

## **A comprehensive assessment of two-decade radioactivity monitoring around the Channel Islands**

### **Abstract**

The Channel Islands are located in the Normand-Breton Gulf (NBG), in the mid-part of the English Channel (France, Normandy). In the northern part, off Cap La Hague, controlled amounts of radioactive liquid waste are discharged by the ORANO La Hague nuclear fuel reprocessing plant (RP). Radionuclides were monitored in the NBG to assess the dispersion of radioactive discharges from the RP in the marine environment. The temporal and spatial distribution of the data are consistent with the history of the discharges, with most gamma emitter radionuclide environmental levels being close to or below the current limits of detection. A clear fingerprint of H-3, C-14 and I-129 radionuclides discharged from the RP is measured. The hydrodynamics in the NBG do not yield a simple gradient with linear distance from the outfall of the RP. Modelling tools were used to understand how radioactive discharges spread from the source of input. Dispersion patterns clearly illustrate the different behaviours of soluble and non-soluble radionuclides. The study indicated that the footprint of radioactive liquid discharges by French nuclear facilities was still measurable in species collected from the NBG for the mostly dissolved radionuclides. The less conservative ones, with a high affinity for suspended matter, are potentially influenced by old releases. These pathways could be investigated by dedicated hydrodynamic dispersion models. Overall, in the Channel Islands the levels are low and consistent with the general decrease in liquid radionuclide discharges by the RP since the 1990s.

## **A comprehensive assessment of two-decade radioactivity monitoring around the Channel Islands**

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# 1 A comprehensive assessment of two-decade radioactivity monitoring around 2 the Channel Islands

3 **Keywords** : Radioactivity monitoring / Marine environment / English Channel / Channel Islands /  
4 Hydrodynamic modelling / Transfer modelling

## 6 Introduction

7 The Channel Islands are located in the Normand-Breton Gulf (hereafter NBG), in the mid-part of the  
8 English Channel (France, Normandy). In the northern part of the NBG, controlled amounts of  
9 radioactive liquid are discharged by the ORANO nuclear fuel reprocessing plant off Cap La Hague. The  
10 Flamanville nuclear power plant operated by EDF is also located on the west coast of the Cotentin  
11 peninsula but its liquid radioactive discharges are smaller (2 orders of magnitude lower) than those  
12 of the reprocessing plant. Two decades ago, a complete survey of the consequences of these  
13 radioactive discharges was carried out by the Nord-Cotentin Radioecology Group (NCRG, 1999).  
14 However, the amounts of liquid discharges have substantially declined since the 1990s and an update  
15 of the radionuclide levels in the marine environment is necessary, using modelling tools that have  
16 become available to appraise the dispersion of local radioactive discharges in the marine  
17 environment. This paper presents annual radioactivity measurements in the brown seaweed *Fucus*  
18 *serratus* performed by the Institut de Radioprotection et de Sûreté Nucléaire (IRSN) in the Channel  
19 Islands from 1998 to 2017. The results are compared with radioactivity monthly time-series  
20 measurements in the same seaweed collected at Goury, about 6 km north of the reprocessing plant  
21 outfall for radioactive liquid discharges, between September 2013 and June 2016. They are also  
22 compared with the results from two sampling surveys undertaken by IRSN in 2014 and 2015,  
23 spanning the whole NBG between Paimpol and Goury. The 2014 and 2015 data from the NBG,  
24 including the islands, are compared with radioactivity measurements in reference samples collected  
25 away from the French nuclear facilities in the Cotentin, at Concarneau (North Atlantic) and Roscoff  
26 (the western end of the English Channel) in Brittany. Data for the seaweed *F. serratus* from the  
27 French shoreline were also compared with data collected at the same locations for the soft parts of  
28 limpets *Patella sp.* (a Mollusk Gastropod), as a representative of another biological compartment.  
29 Finally, hydrodynamic modelling tools were used in the NBG to support the data from the biota with  
30 a comprehensive assessment of the general pattern of seawater radionuclide dispersion.

## 31 Material and methods

32 *Annual discharges from the ORANO reprocessing plant at La Hague:* The amounts of controlled  
33 radioactive liquid waste discharges from the ORANO RP at La Hague are published annually  
34 (<https://www.orano.group>). Figure 1 shows the history of liquid discharges from 1982 to 2015 (a  
35 color version is available in the [supplementary material: Fievet et al 2020 Supplementary](#)  
36 [Material.pdf](#)).

37 Figure 1

38 *Bioindicators.* Brown seaweeds and limpets are common bioindicators used to monitor radioactivity  
39 in the marine environment in Western Europe ([OSPAR Commission, 2009](#); [CEFAS, 2016](#)). They are  
40 easily collected manually on the shore all year round. They are known to accumulate radionuclides

41 with respect to seawater ([IAEA, 2004](#)), thereby making them magnifying tools for monitoring levels in  
42 the environment.

43 *Sampling and processing.* Toothed wracks (*Fucus serratus*) and limpets (*Patella sp*) were collected on  
44 shore in different locations of the maritime area between Paimpol and Goury. The sampling locations  
45 are shown in Figure 2 (geographical coordinates are provided in the [supplementary material:](#)  
46 [Fievet\\_etal\\_2020\\_Supplementary\\_Material.xlsx](#)). This included the French shoreline as well as the  
47 Channel Islands and a few other minor islands (Les Écrehou, Chausey, Roches-Douvres). The samples  
48 were collected annually in summer from 1998 to 2017 in the Channel Islands and in spring in 2014  
49 and 2015 at other locations in the NBG. Annual samplings in the Channel Islands were carried out in  
50 Jersey, Guernsey and Alderney. Although Les Écrehou does belong to the Channel Islands, this  
51 sampling location is included in the spring 2014 and spring 2015 NBG group throughout the paper. At  
52 Goury, close to the source of the radioactive liquid discharges, samplings were collected monthly  
53 between September 2013 and June 2016 to address temporal variability with a higher time  
54 resolution. *Fucus serratus* samples were also collected in 2014 and 2015 at reference locations away  
55 from the French nuclear facilities of the Cotentin, at Concarneau and Roscoff in Brittany. About 8 Kg  
56 of *Fucus serratus* and 15 Kg of limpet were collected for each sample. The limpets were frozen at -  
57 20°C for a week and thawed to separate the soft parts from the shell. The whole seaweed thallus and  
58 limpet parts were then dried at 90°C for about a week prior to grinding into a fine powder. The  
59 limpet parts were finally burned at 450°C to ash. Another fresh fraction was freeze-dried to measure  
60 C-14 and H-3 in some of the seaweed and limpet samples.

61 Figure 2

62 *Radionuclide measurements.* Radioactivity measurements were performed at the Laboratoire de  
63 Radioécologie de Cherbourg-Octeville, the Laboratoire de Mesure de la Radioactivité dans  
64 l'Environnement (IRSN, Orsay) and the Laboratoire de Mesure Nucléaire (IRSN, Le Vésinet). Ground  
65 material (dry seaweed or ashes of limpets' soft parts) was transferred to calibrated cylindrical  
66 containers for gamma spectrometry on HPGe germanium detectors. The detectors were equipped  
67 with anti-cosmic devices as described in de [Vismes Ott et al., \(2008; 2012\)](#) and the gamma-gamma  
68 coincidence technique, as described in [Paradis et al. \(2016\)](#), was used. This high-performance gamma  
69 spectrometry guaranteed very low limits of detection (LoD). I-129 was quantified in dried *Fucus*  
70 *serratus* by gamma-X spectrometry according to [Lefèvre et al., \(2003\)](#) and [Bouisset et al., \(1999\)](#).  
71 C-14 was analyzed at LMRE by dry material combustion, CO<sub>2</sub> trapping and conversion into benzene  
72 and C-14 beta scintillation counting on a Tricarb counter (Perkin-Elmer), as described in [Fiévet et al.,](#)  
73 [\(2006\)](#). As of 2010, H-3, in the form of tritiated water molecules (HTO), was measured in the water  
74 extracted from the fresh material by freeze-drying and scintillation counting on a Tricarb counter  
75 (Perkin-Elmer) as described in [Fiévet et al., \(2013\)](#). In the 2014, 2015 and 2017 samples, H-3 as  
76 organically bound tritium (OBT) was measured at LMRE after combustion of dry material and  
77 combustion water scintillation counting on a Tricarb counter (Perkin-Elmer) as described in [Fiévet et](#)  
78 [al., \(2013\)](#). The dry material was finally burned to ash, and tracers (Pu-242 and Am-243) were added  
79 for extraction yield correction. Actinides were then extracted by 3 successive coprecipitations: 1-  
80 calcium oxalate at pH 1.5; 2-iron hydroxide at pH 8.5; 3-calcium oxalate at pH 1.5; and destruction of  
81 the second calcium oxalate precipitate and dissolution of the residue in 8M HNO<sub>3</sub>. Pu(IV) was  
82 separated from Am-Cm using anion exchange chromatography on resin AG1X8 (Biorad), conditioned  
83 in 8M HNO<sub>3</sub>. After loading and washing with 8M HNO<sub>3</sub>, the solution was preserved for the

84 purification of the Am-Cm fraction. Pu was eluted with 12M HCl / 0.1M NH<sub>4</sub>I. Am-Cm fraction  
 85 purification involved 3 successive chromatography columns: 1- separation of Am-Cm/Fe, U, Th on a  
 86 double column with AG1X8/AG50W resin (Biorad); 2- separation of Am-Cm/Cu, Ni metals on a  
 87 column with TRU resin (Biorad); and 3- separation of Am-Cm/rare earth elements on a column with  
 88 AG1X4 resin (Biorad). Pu and Am residues were dissolved in concentrated HNO<sub>3</sub> and diluted in water.  
 89 The 2 sources were prepared by electrodeposition onto stainless steel discs as the cathode (Pt  
 90 anode), current 1 A, time 2 hours, 5 mm between the 2 electrodes. Pu and Am-Cm isotopes were  
 91 determined with a low level background alpha spectrometer (alpha-analyst, MIRION Technologies).  
 92 The detectors were 450 mm<sup>2</sup> PIPS detectors, with an efficiency around 34% for <sup>239</sup>Pu. Counting time  
 93 was 14 days. The data processing software to quantify the radionuclide activities included  
 94 Interwinner™ (Hi-Tech Detection Systems), Genie 2000™ (MIRION Technologies), QuantaSmart™  
 95 (PerkinElmer). C-14 concentrations were expressed in Bq.Kg<sup>-1</sup> Carbon and H-3 concentrations were  
 96 expressed in Bq.L<sup>-1</sup>. Other radionuclide concentrations were reported in Bq.Kg<sup>-1</sup> dry weight. Activity  
 97 results were reported with 2 x Sigma uncertainties. The wet/dry (ww/dw) weight ratios are provided  
 98 for conversion purposes. For clarity, radionuclides are categorized in the paper as gamma, beta and  
 99 alpha emitters according to the spectrometry detection method used for their activity quantification.  
 100 However, for a more comprehensive assessment of the data, radionuclides were categorized in the  
 101 discussion section as soluble and non-soluble.

102 *Seawater circulation and radionuclide dispersion in the NBG:* Previously validated modelling tools are  
 103 available to document the hydrodynamic dispersion of radionuclides in the NBG ([Salomon et al.,](#)  
 104 [1988](#); [Bailly du Bois et al., 2012](#) and references therein). They include a Lagrangian residual model,  
 105 which calculates the tidal residual currents, and an instantaneous current model (MARS2D) which  
 106 calculates currents and dispersion every 60-300s (see details below). These tools were used here to  
 107 provide the following information:

108 - The general water circulation pattern in the NBG was illustrated by the trajectory and  
 109 velocity of Lagrangian residual currents ([Salomon et al., 1988, 1996, Bailly du Bois and Dumas, 2005](#)).  
 110 The residual current is the water's residual movement after one complete back-and-forth tide  
 111 movement has occurred. Residual currents mainly result from tidal forcing, wind surface friction,  
 112 bottom shape and roughness, which are the main driving forces for the long-term dispersion of  
 113 dissolved material in seawater. The dispersion pattern in the NBG is complex due to the presence of  
 114 the Channel Islands and other emerged areas. It is highly dynamic and variable depending on the  
 115 wind and tide conditions and the exact time of the discharge. This pattern of average residual  
 116 currents was calculated by the MARS2D Model using a previously published method ([Salomon et al.,](#)  
 117 [1996; Bailly du Bois et al., 2005](#) and references therein). A theoretical constant wind direction  
 118 (SW=231°) and speed (8 m/s) and a constant medium tide coefficient (70) were applied. This average  
 119 wind condition corresponded to the quadratic mean of the wind direction and speed calculated  
 120 between 1984 and 2017 at a central location in the English Channel (2°W; 50°N).

121 - A further insight was gained with the MARS2D model by calculating the annual mean  
 122 concentrations resulting from a theoretical constant discharge from the RP plant with real wind and  
 123 tide conditions between 1984 and 2015 (radioactive decay was not implemented for this constant  
 124 discharge). The MARS2D model uses two-dimensional horizontal approximation (i.e. shallow-water  
 125 equations) capable of producing a satisfactory representation of dissolved-substance transport in the  
 126 English Channel. These equations were solved using the finite-difference MARS model, with implicit

127 alternate direction time-stepping for gravity-driven inertia waves. Non-linear terms were discretized  
 128 semi-implicitly. Full details concerning the MARS algorithm are given by [Lazure and Dumas \(2008\)](#).  
 129 The model used here involved a nesting strategy, starting from a broad region covering the entire  
 130 North-West European continental shelf (with a 5-km grid resolution) down to a detailed domain  
 131 covering the whole English Channel with a mesh size of 500 m. It accounted for real releases (but  
 132 could be set as theoretical constant), tide and meteorological forcing to simulate instantaneous  
 133 currents with a mean time step of 60 s. Real wind data were provided by Météo France and  
 134 corresponded to the outputs from the European Centre for Medium-Range Weather model ([The ERA  
 135 - Interim reanalysis, 2011](#)). Tide conditions were provided according to [Lyard et al., 2006](#). This  
 136 yielded a normalized dispersion pattern in the NBG corresponding to realistic forcing during the  
 137 period. Annual mean data extend from 1984 to 2015. The aim was to show the influence of the  
 138 wind and tide annual changes on the average dispersion of any soluble radionuclide released by the  
 139 outlet of the Reprocessing Plant (It is not related to the real content of the discharges from the RP).

140 - To link this information on the average dispersion pattern in the NBG with realistic data, the  
 141 H-3 concentrations resulting from real discharges from the RP plant with real wind and tide were  
 142 calculated and annually averaged. The H-3 radionuclide was selected because it is released as HTO  
 143 and thus, as a conservative tracer in seawater, it traces the dispersion of discharges by water  
 144 movements exactly (the half-life of H-3, which is 12.3 years, was implemented in this modelling).

145 *Normalization of biota radionuclide concentrations to the discharges.* For a comprehensive  
 146 assessment of radionuclide transfers into the NBG, the biota data were normalized to discharges. The  
 147 approach was similar to the concept of normalized activity concentrations previously used on data  
 148 from the Irish Sea ([Hunt et al. 1985, 2013; Hunt and Kershaw 1990](#)). In the present study,  
 149 normalization transform was performed according to the following relationship:

$$150 \quad [Rn]_{Norm,s} = \frac{[Rn]_{s,Bq.kg^{-1}dry} * 1000}{wdr_s * CF_s * 12MD_{dat}}, \text{ in Bq} \cdot (1000\text{kg})^{-1} \text{ per TBq} \cdot \text{y}^{-1} \text{ or Bq} \cdot \text{kg}^{-1} \text{ per GBq} \cdot \text{y}^{-1}.$$

151 where

152  $[Rn]_{Norm,s}$  = Concentration of radionuclide Rn in species s normalized as a "dilution&transfer  
 153 factor"

154  $[Rn]_{s,Bq.kg^{-1}dry}$  = Concentration of radionuclide Rn in species s in Bq.kg<sup>-1</sup> dry

155  $wdr_s$  = Wet/dry ratio value for species s

156  $CF_s$  = Concentration Factor for radionuclide Rn in species s

157  $12MD_{dat}$  = 12 months of discharges of radionuclide Rn in TBq preceding the date of sample

158 1000 = converts Bq.Kg<sup>-1</sup> to Bq.1000 kg<sup>-1</sup> in biota because the concentrations in seawater

159 calculated by the hydrodynamic model were expressed in Bq.m<sup>-3</sup>.

160 This normalization transform resulted in the expression of species concentrations as a "dispersion  
 161 factor-like" value resulting from both dilution and transfer to the biota from the source of input,  
 162 assuming a constant discharge of 1 TBq.y<sup>-1</sup>. It can thus be directly compared to the normalized  
 163 average dispersion pattern in seawater calculated by the MARS-2D hydrodynamic model, assuming a  
 164 constant discharge of 1 TBq.y<sup>-1</sup>. It should be pointed out that radioactive decay is implemented by  
 165 MARS-2D in the dilution&transfer process. However, except for Co-60 and Ru-106 (half-lives 5.27 y  
 166 and 373 d, respectively), the influence of radioactive decay is small with regard to the estimated half-

167 time turnover of radionuclides in the NBG (see Results section). The CFs values are given in the  
 168 following table:

169 Table 1

170 For Cs-137 and C-14, the average observation level at Concarneau in 2014 and 2015 was deduced  
 171 from biota concentrations to account for the background level at the entrance to the Channel. This  
 172 background level results from sources other than the ORANO La Hague RP (Cs-137: fallout from past  
 173 nuclear weapons tests, the contribution of discharges from Sellafield, the Chernobyl accident, etc.; C-  
 174 14: natural, fallout, etc.). This data transform made it possible to visualize whether or not the  
 175 radionuclides dispersed in the NBG as soluble radionuclides in seawater before transfer to the biota.  
 176 If not, the discrepancy illustrates the non-conservative behavior of radionuclides in the marine  
 177 environment of the NBG.

## 178 Results

179 *Annual time-series measurements in the Channel Islands.* Radionuclide concentrations measured  
 180 annually in *Fucus serratus* from 1998 to 2017 in the Channel Islands are provided in the  
 181 [supplementary material: Fievet et al. 2020 Supplementary Material.xlsx](#). To illustrate the trends in  
 182 radionuclide concentrations, the data from the location nearest to the source of input in the Channel  
 183 Islands, Alderney, Corblet Bay (NE), and from the westernmost site, Guernsey, Portelet Harbour  
 184 (SW), are displayed graphically for Co-60, Cs-137 and I-129 (Figure 3).

### 185 Figure 3

186 *A wider view over the NBG area.* Data from the Channel Islands were put into perspective through  
 187 comparison with the levels in the wider area of the NBG, as well as reference levels from two  
 188 locations away from the radioactive discharges. For this purpose, radionuclide measurements were  
 189 performed in 12 locations across the NBG (in addition to the locations in Jersey, Guernsey and  
 190 Alderney) on *Fucus serratus* samples collected in the spring of 2014 and 2015. To estimate the  
 191 reference levels, samples from 2 additional remote locations in Brittany (Concarneau) and at the  
 192 western end of the Channel (Roscoff) were collected. The data, including those from the Channel  
 193 Islands for comparison, are illustrated as bar charts sorted by (linear) distance from the outfall of the  
 194 nuclear fuel reprocessing plant (RP) (Figure 4). Spreadsheets with all the data are provided as  
 195 [supplementary material: Fievet et al. 2020 Supplementary Material.xlsx](#).

### 196 Figure 4

197 *Monthly time-series measurements in the near vicinity of the input source.* To provide further insight  
 198 into the temporal variability of radionuclide levels, the same gamma emitter radionuclides were  
 199 analyzed on a monthly basis in the same seaweed and in the soft parts of limpets collected at Goury,  
 200 6 km north of the radioactive liquid discharge outfall at the reprocessing plant. In this location, close  
 201 to the source of input, changes over time are expected to show the widest variations due to the  
 202 proximity of the labelled plume of each release which could influence the coastal measurements. The  
 203 data are presented in Figure 5 and tabulated in the [supplementary material: Fievet et al. 2020](#)  
 204 [Supplementary Material.xlsx](#).

### 205 Figure 5:

206 *Alpha and Beta emitters.* C-14, H-3, Pu-238, Pu-239,240 (the sum of Pu-239 and Pu-240 isotopes) and  
 207 Am-241 were also measured in some of the samples from the Channel Island annual time-series  
 208 (there was no such measurement in 1998). The data are presented in Table 2.

209 **Table 2**

210 The same alpha and beta emitter radionuclides were also measured in seaweed samples collected in  
 211 the NBG area for comparison. The data are reported as bar charts in Figure 6, with locations sorted  
 212 with respect to the distance from the outfall of the RP, using the same presentation as in Figure 4  
 213 and tabulated in the [supplementary material: Fievet\\_etal\\_2020\\_Supplementary\\_Material.xlsx](#). The  
 214 isotopic ratio between Pu-238 and Pu-239,240 was calculated in biota samples collected in the NBG.  
 215 The results are reported in the [supplementary material: Fievet\\_etal\\_2020\\_Supplementary](#)  
 216 [Material.xlsx](#).

217 **Figure 6**

218 *Seawater circulation and radionuclide dispersion in the NBG:* The general water circulation pattern in  
 219 the NBG was illustrated by calculating the trajectory and velocity of average residual currents (Figure  
 220 7).

221 **Figure 7**

222 The map of average residual currents in Figure 7 shows a major northbound pathway in the Alderney  
 223 race, towards the eastward trend of the main water mass in the English Channel (not shown).  
 224 However, persistent counterclockwise recirculation gyres around Guernsey, Jersey and other minor  
 225 islands contribute to another progressive southbound route towards the bottom of the NBG. The  
 226 description of the divergence zone near the RP outfall is accurately documented in [Bailly du Bois et](#)  
 227 [al., \(2012\)](#).

228 The normalized dispersion pattern of radioactive discharges from the RP in the NBG was illustrated  
 229 using the MARS2D model ([Bailly du Bois et al., 2012](#)) to calculate annual mean concentrations  
 230 resulting from a theoretical constant discharge from the RP plant with real wind and tide conditions  
 231 between 1984 and 2015. Only 1985 and 2015 are shown (Figure 8) to demonstrate that the pattern  
 232 was similar throughout the whole period of the study (in 1984, the mixing spins up to reach a stable  
 233 situation). It shows that variations in hydrodynamic forcing have a weak influence at the annual time  
 234 scale. The complete annual color maps time-series is provided in the [supplementary material:](#)  
 235 [Fievet\\_etal\\_2020\\_Supplementary\\_Material.pdf](#).

236 **Figure 8**

237 Since the model's calculation based on a constant discharge yielded a steady state in 1985, the time  
 238 taken to reach this steady state during 1984 provided an estimate of the average turnover of soluble  
 239 radionuclides between the source of input and the different locations in the NBG. Starting from  
 240 concentration "zero", the time taken to reach half of the 1985 mean annual concentration at each  
 241 biota sampling location in the NBG was derived from the model results. The values were added to  
 242 the 1985 map in Figure 8 at the sampling locations, as an estimate of the average half-time turnover  
 243 of soluble radionuclides discharged by the RP (in days).

244 Finally, the H-3 concentrations in seawater resulting from real discharges from the RP plant with real  
 245 wind and tide conditions were calculated, along with the annual mean values. The annual means  
 246 extend from 1984 to 2015 but only the 2014 and 2015 annual mean data are shown, for conciseness  
 247 (Figure 9). 2014 and 2015 maps with a higher color-levels resolution are available in the  
 248 [supplementary material: Fievet\\_etal\\_2020\\_Supplementary\\_Material.pdf](#).

249 Figure 9

250 *Transfer to the biota between the outfall of the RP and the NBG:* For a comprehensive assessment of  
 251 the dispersion pattern of radionuclides in the NBG, the data in the biota were transformed to give  
 252 "dilution&transfer factor-like" values from the source of radioactive input (after background  
 253 subtraction for Cs-137 and C-14, divided by the Concentration Factors, normalized to the discharges,  
 254 see Methods).

255 The first step was to illustrate the annual average dilution factor calculated by the hydrodynamic  
 256 model for the different biota sampling locations. This consisted in extracting the 2014 and 2015 data  
 257 values from the time-series illustrated in Figure 8 (for 2015 and shown in the [supplementary material](#)  
 258 [for 2014: Fievet\\_etal\\_2020\\_Supplementary\\_Material.pdf](#)) for the biota sampling locations. This is  
 259 the reference showing the dispersion of soluble radionuclides constantly discharged by the RP in the  
 260 seawater currents. However, radioactive liquid waste is not discharged constantly by the RP but is  
 261 discharged mainly during a special time window with regard to the tide schedule in order to optimize  
 262 hydrodynamic dispersion to the north and the eastern English Channel (see [Bailly du Bois et al.,](#)  
 263 [2012](#)). This discharge timing reduces dispersion towards the NBG where the water mass residence  
 264 time is longer than in the median area of the Channel. To take this discharge time frame into  
 265 account, the dispersion pattern of real H-3 discharges was used to represent the dispersion of  
 266 soluble radionuclides from the RP. However, to use it as a reference, it had to be normalized to an  
 267 annual amount. We selected 1 TBq.y<sup>-1</sup> because it was the order of magnitude of <sup>137</sup>Cs releases as of  
 268 2000. For this purpose, annual average seawater H-3 concentrations (Bq.m<sup>-3</sup>) resulting from real H-3  
 269 discharges by the RP calculated by the MARS2D model (Figure 9) were divided by the annual H-3  
 270 discharge amount (TBq). This calculation was performed at each biota sampling location in 2014 and  
 271 2015. It yielded dilution factor values that could be used specifically in 2014 and 2015 as a reference  
 272 for the soluble radionuclide dispersion pattern in the NBG for other radionuclides. The dispersion  
 273 patterns of other radionuclides provided by concentrations normalized to discharges were thus  
 274 compared to this reference. Background values estimated on the basis of the average concentrations  
 275 measured at Concarneau were subtracted from NBG data for Cs-137, C-14 and the transuranic  
 276 elements (background was considered negligible for H3). The radionuclides were presented as two  
 277 groups on the basis of their expected behaviour depending on their solubility (Figure 10 and Figure  
 278 11).

279 Figure 10

280 Figure 11

281 *Transfer to the biota near the RP outfall:* A second way to discriminate between soluble and  
 282 particulate radionuclides was to look at the temporal variability of the relationship between the  
 283 observed radionuclide concentrations and the discharges, regardless of the dispersion pattern. At a  
 284 short distance from the source of input, dispersion is just beginning, so the influence of interactions

285 with suspended matter and sedimentation should be less. For this purpose, the monthly time-series  
 286 measurement data for seaweeds from Goury were processed the same way, divided by the CFs  
 287 **(Table 1)** and normalized to the 12-month discharges preceding the sampling dates. The results are  
 288 presented in Figure 12.

289 Figure 12

## 290 Discussion

291 In addition to naturally occurring radionuclides and fallout from past atmospheric nuclear weapons  
 292 tests, the sources of radioactivity in the English Channel mainly include radioactive discharges from  
 293 the nuclear industry. On the French coast, the ORANO reprocessing plant (RP) at La Hague is by far  
 294 the major contributor, even though liquid discharges have fallen considerably since the 1980s. The  
 295 reduction in discharges is due to the constant implementation of the BAT (Best Available Techniques)  
 296 strategy ([OSPAR Commission, 2014](#)). Four EDF nuclear power plants (NPP), at Flamanville, Paluel,  
 297 Penly and Gravelines, are located on the French coast, in addition to the nuclear power plant at  
 298 Nogent-sur-Seine on the Seine River, which flows into the English Channel. The NPPs' contribution to  
 299 radionuclide levels in the NBG is small compared to that of the RP. Since radioactive liquid discharges  
 300 from the French RP into the English Channel occur in the NBG, it was particularly relevant to assess  
 301 the dispersion of those discharges in this area. It was also useful to put the observed data in the NBG  
 302 into perspective by comparing them with reference locations, remote from radioactive discharges.  
 303 This was the purpose of the two reference locations at Concarneau (Brittany, Atlantic) and Roscoff  
 304 (the western part of the English Channel). A comprehensive assessment of the radiological impact of  
 305 radioactive discharges by French facilities in the English Channel was previously carried out by the  
 306 North Cotentin Radioecology Group ([NCRG, 1999](#)). It included a review of environment radioactivity  
 307 concentration measurements up to 1996. Because liquid discharges into the sea have substantially  
 308 decreased since this assessment (Figure 1), it was essential to provide an update of environmental  
 309 radionuclide levels two decades later.

310 The English Channel is a shallow megatidal sea and the outfall of the RP is located in an area well  
 311 known for strong tidal currents. The hydrodynamics of the English Channel are well characterized  
 312 and reliable hydrodynamic models have been developed and extensively validated ([Bailly du Bois and](#)  
 313 [Dumas, 2005](#); [Bailly du Bois et al., 2012](#)). It must be outlined that the vertical distribution of  
 314 radionuclides released by the RP is already homogeneous in the water column at 1 500 m distance  
 315 from the outlet ([Bailly du Bois et al., 2020b](#)). Once the source terms of discharges are known with  
 316 accuracy (time and location), seawater concentrations in the English Channel and up to the North Sea  
 317 can be predicted by models. In addition to residual currents (resulting from back-and-forth tidal  
 318 movements), the general water circulation in the English Channel is also driven by wind friction.  
 319 Because of the tidal residual currents and common prevailing wind directions, the English Channel  
 320 usually flows slowly from west to east, i.e. from the Atlantic to the North Sea. However, when the  
 321 wind blows from the north-east for several weeks, this circulation temporarily reverses. It should be  
 322 added that because of radioactive discharges from the Sellafield facility (United Kingdom),  
 323 radionuclide concentrations in the marine environment in the Irish Sea are substantially higher than  
 324 in the English Channel ([CEFAS, 2016](#)). The potential contribution of dissolved radionuclide inputs at  
 325 the western end of the English Channel coming from the Irish Sea has been estimated to be about 1%  
 326 of discharges from Sellafield ([Bailly du Bois et al., 1995, 2002](#)). This estimate was recently confirmed  
 327 ([Castrillejo et al., 2020](#)). At present, this contribution is considered to be low. The slight difference



328 between the levels of Cs-137 in seaweeds observed in 2014 and 2015 in Roscoff (potentially  
329 influenced by Sellafield discharges) compared to Concarneau (situated on the Atlantic coast of  
330 Brittany and not influenced by Sellafield discharges) was not likely to be significant for two reasons.  
331 First, the uncertainties (2 x sigma) accounted for at least a third of the reported activities and second,  
332 the minimum levels observed in the NBG (Roches Douvres) were lower than at Roscoff and were  
333 similar to those at Concarneau. If we account for error magnitudes, these three locations could not  
334 be distinguished from one another. Annual data were obviously not enough to address this point  
335 accurately.

336 *Hydrodynamics of radionuclide dispersion from the RP in the NBG:* Prior to looking at the data in the  
337 biota, a rapid survey of the general water mass circulation in the NBG was crucial to understand the  
338 dispersion of radionuclides and their transfer to marine species. The hydrodynamics in the NBG are  
339 complex and so an intuitive interpretation of the data was challenging. The only way to achieve a  
340 comprehensive assessment of radionuclide monitoring in the biota of the NBG was to use models to  
341 simulate the local hydrodynamics and water mass pathways (Figure 7). The models used in this study  
342 provided powerful tools for deciphering the major processes governing the fate of radioactive  
343 discharges from the RP in the area and made it possible to attempt to relate the radionuclide  
344 concentrations observed in the biota with their sources of origin. The general pattern of the residual  
345 currents is shown in Figure 7. It was calculated using average conditions, with a constant wind  
346 corresponding to the average for meteorological conditions in the NBG and a medium tide. It should  
347 be emphasized that this average residual current pattern explains the long-term dispersion of  
348 dissolved substances (Salomon *et al.* 1988). But the short-term excursion of radionuclide movements  
349 resulting from real back-and-forth tide currents extends over a much larger spatial magnitude (see  
350 [Bailly du Bois \*et al.\*, 2012](#), for details). After documenting the general pattern of the dispersion, a  
351 good way to assess the exposure levels of the biota was to look at the average concentrations  
352 resulting from the discharges. Many parameters influenced radionuclide concentrations in the  
353 seawater where the biota was monitored. They include the wind, the tide and, last but not least, the  
354 exact time of the discharges. To understand the relative influence of these components, a first step  
355 was to calculate with MARS2D the average concentrations resulting from a theoretical constant  
356 discharge with the real wind and tide conditions. The calculation results in Figure 8 showed that after  
357 one year of discharge starting in 1984, average concentrations in 1985 showed a spatial distribution  
358 which was more complex than just a gradient with respect to linear distance from the outfall of the  
359 discharges. However, this spatial distribution calculated in 1985 turned out to correspond to a  
360 "steady state" at an annual time scale, so the 2015 map in Figure 8 looked very similar. The 1985-  
361 2015 annual color map time-series provided in the [supplementary material](#)  
362 ([Fievet \*et al.\* 2020 Supplementary Material.pdf](#)) support this conclusion. This information  
363 demonstrated that, on a yearly basis, real rapid changes in wind and tide were smoothed and  
364 resulted in similar calculated annual average concentrations. This provided an opportunity to look at  
365 the time-course of changes in concentrations from zero to the steady state observed as of the end of  
366 1984 and to estimate the average half-time turnover of soluble radionuclides discharged in the NBG  
367 by the RP. These estimated average half-time turnover values (in days) completed the average  
368 trajectories information displayed on the 1985 map in Figure 8. Close to the outfall (Goury), the 23-  
369 day average turnover value characterized the time to reach half the average steady state level.  
370 Barneville and Alderney had values of around 2 months whilst central and southern locations of the  
371 NBG had values of between 3 and 4 months. In the innermost area of the NBG, in Granville and

372 Cancale, where a slower average half-time turnover is logically expected, the values were estimated  
373 to be 8 to 9 months, respectively. The spatial modelled average concentration distribution shown in  
374 Figure 8 provided a first insight to understand the distribution of radionuclides in the biota  
375 monitored in the NBG. To make it more realistic, the data should be calculated with real discharge  
376 data. This was the purpose of the third step where all model inputs were "real-world" data for H-3,  
377 i.e. the exact time course of the discharges, and the real wind and tide conditions. It should be  
378 pointed out that the real discharges were not routinely evenly distributed over time but were  
379 released in a particular time frame with regard to the tide schedule. This optimized their dispersion  
380 to the main English Channel water mass flux to the north and east, and prevented them from being  
381 held in the NBG where the water mass renews less rapidly. H-3 dispersion thus provided the best  
382 possible representation of dispersion by water currents in the area. This information was reliable  
383 because this modelling had been extensively validated with many measurements ([Bailly du Bois et](#)  
384 [al., 2005, 2012](#) and references therein). The resulting annual mean concentrations were calculated  
385 from 1984 to 2016 and the maps for the years 2014 and 2015 are given in Figure 9 for illustration  
386 purposes. In Figure 9, the average distribution pattern was similar to that of Figure 8 but the annual  
387 average seawater H-3 concentration values were realistic. Since annual tritium discharges are around  
388  $1\text{E}+04\text{TBq}\cdot\text{y}^{-1}$  (Figure 1), concentrations are logically four orders of magnitude higher than those  
389 resulting from  $1\text{TBq}\cdot\text{y}^{-1}$  theoretical constant discharge. This general pattern of soluble substance  
390 dispersion in the NBG assessed by hydrodynamic modelling tools provided a basis to look at  
391 radionuclide transfers between the outfall of the ORANO La Hague RP and the marine biota.  
392 Nevertheless, it should be emphasized that annual means provided the general scheme but obviously  
393 smoothed out temporal variability and should be used cautiously when comparing with individual  
394 measurement data, particularly in the vicinity of the release outfall.

395 The biota data reported in this study helped to address three main questions about the radiological  
396 consequences of radioactive discharges by French nuclear facilities in the marine environment of the  
397 NBG. Firstly, how did radioactivity increase in the NBG compared to a reference background level?  
398 Secondly, the geographical variability in the whole of the NBG and how did the levels observed in the  
399 Channel Islands compare with this variability? Thirdly, what was the significance of annual  
400 observations with regard to the temporal variability (had we overlooked something?). Analyzing the  
401 same bioindicator, toothed wrack (*Fucus serratus*), all along the shoreline in the marine environment  
402 enabled comparisons to be made to address these three issues. The data for the soft parts of limpets  
403 (*Patella sp*) collected in the vicinity of the radioactive input sources provided an insight into transfer  
404 to mollusks, as a representative of an upper level in the trophic web. Radionuclide concentrations  
405 observed in the soft parts of *Patella* were below those observed in *Fucus*, so emphasis has been  
406 placed on the seaweed in the discussion. The data are presented in the Results section by metrology  
407 categories (gamma spectrometry, liquid scintillation, alpha spectrometry) because this is how  
408 radioactivity monitoring data are usually collected and sorted. However, for a more comprehensive  
409 discussion of radionuclide dispersion in the marine environment, they can be better categorized as  
410 conservative or non-conservative. Conservative radionuclides stay soluble in seawater with no loss  
411 during dispersion by water currents. Some transfer to particulate material and the biota is possible  
412 but the process has little influence compared to the amount of dissolved material (i.e. losses below  
413 5% at the English Channel scale). In the present context, H-3, C-14, I-129 and Cs-137 were included in  
414 this category ([Bailly du Bois and Guéguéniat, 1999](#)). However, due to potential concentration in the  
415 sediment compartment and possible future desorption, as in the Irish Sea ([Jones et al., 2007](#), [Hunt et](#)

416 [al., 2013](#)), caution is required in the case of Cs-137. Loss of radiocaesium was estimated at 17% at the  
 417 English Channel scale ([Bailly du Bois and Guéguéniat, 1999](#)). Conversely, because of their higher  
 418 affinity for suspended matter, non-conservative radionuclides do not behave as soluble substances  
 419 only. They are readily discharged as particulate material or they adsorb to suspended matter and  
 420 thus follow the fate of particulate material. For these radionuclides, the partitioning of the  
 421 radionuclide between the aqueous and the particulate phases must be taken into account.  
 422 Depending on their particulate fraction or the contribution of their adsorption, they may disappear  
 423 from the aqueous phase with time and distance from the source of input. The adsorption capacity of  
 424 particles depends on their specific surface area, so the smaller the particles, the greater their  
 425 capacity to capture radionuclides. This process also potentially influences their capacity to transfer to  
 426 the biota (bioavailability). However, the biota incorporates not only soluble radionuclides but also  
 427 particulate material either by adsorption (sticking to protective mucus) or by ingestion (filter feeders,  
 428 food chain, etc.). When collected in the marine environment, the biota is likely to carry suspended  
 429 matter either adsorbed or inside its digestive tract (if present) ([Carvalho, 2018](#)). Non-conservative  
 430 radionuclides are known to partition mainly in the particulate phase rather than in the aqueous  
 431 phase. Since these radionuclides share the same fate as particulate material, they eventually  
 432 sediment and get temporarily trapped in the sediment. When they do so, they carry the signature of  
 433 the discharges at the time when they adsorbed to the suspended matter. If the sediment resuspends  
 434 for some reason later on, its potential interaction with the biota carries this old discharge signature,  
 435 which may be different from that of the present discharge. It is currently impossible to find out what  
 436 the relative contributions are of desorption following sediment particle resuspension and of the  
 437 "contamination" of biota samples by sediment particles. However, a thorough examination of  
 438 suspended matter is necessary to properly interpret the data for non-conservative radionuclides. In  
 439 the present context, the non-conservative radionuclides included Co-60, Ru-106, Pu-238, Pu-239,240  
 440 and Am-241. At the scale of the English Channel, losses were estimated for Co-60 and Ru-106 to be  
 441 92% and 81%, respectively ([Bailly du Bois and Guéguéniat, 1999](#)).

442 Looking at the biota concentration data sorted by linear distance from the outfall of the RP was a  
 443 convenient way to illustrate graphically the dispersion of the radionuclide discharges (Figure 4 and  
 444 Figure 6). However, because of the very complex water circulation in the NBG, some discrepancies  
 445 between the concentrations observed in the biota sampling locations and their linear distance from  
 446 the source of input were expected. To attempt to link the data for radionuclides in the biota with the  
 447 dispersion pattern in the NBG, the concentrations observed in the biota were normalized to the  
 448 discharges and transformed using Concentration Factor values to yield "Dilution&transfer factor-like"  
 449 data that could be directly compared to the hydrodynamic information (see Methods for details). But  
 450 since hydrodynamic models were used to provide information on the dispersion of soluble material,  
 451 the success of this comparison obviously depended on the conservative vs non-conservative  
 452 behaviour of the different radionuclides. The method shed light on three major components of the  
 453 system: the dispersion throughout the NBG area; the background levels in the seawater entering the  
 454 English Channel; and the putative influence of past discharges trapped in the sediment compartment  
 455 on present biota concentrations revealed by non-conservative radionuclide data in the near vicinity  
 456 of the outfall from the RP.

457 *Conservative radionuclides:* Four mainly soluble radionuclides were routinely detected in seaweed  
 458 collected on the southern shores of the English Channel, namely H-3, C-14, I-129 and Cs-137.  
 459 Regarding H-3, we focused on the HTO form measured in water extracted from seaweed by freeze-

460 drying. The data ranged between our LoD of 2 Bq.L<sup>-1</sup> and 7.4 Bq.L<sup>-1</sup> in the Channel Islands and in other  
 461 locations of the NBG ([supplementary material: Fievet et al. 2020 Supplementary Material.xlsx](#))  
 462 except in the vicinity to the outfall from the RP, where higher levels up to 30 Bq.L<sup>-1</sup> were found  
 463 ([Fiévet et al., 2013](#)). Recent data for the Bay of Biscay indicated that the H-3 background level in  
 464 seawater at the western entrance to the English Channel ranged between 0.07 and 0.33 Bq.L<sup>-1</sup> ([Oms,](#)  
 465 [2018](#)). H-3 is released by the RP as HTO and is a perfect tracer of soluble radioactive discharge  
 466 dispersion. There was a slight difference in the modelled dispersion patterns across the NBG,  
 467 between a theoretical discharge distributed evenly over time, and real discharges, which are  
 468 performed in the right time-frame to optimize dilution to the north towards the general water flow  
 469 of the Channel (Figure 8 and Figure 9, respectively). H-3 dispersion was used as a reference to show  
 470 how soluble radionuclides spread out from the RP into the NBG in 2014 and 2015. Near the outfall of  
 471 the RP, the monthly time-series data normalized to the discharges (Figure 12 A) displayed a noisy  
 472 pattern reflecting the highly dynamic mixing in the Alderney race off Cap La Hague ([Bennis et al.,](#)  
 473 [2020](#); [Bailly du Bois et al., 2012, 2020a](#)). This also gave the reference temporal pattern of soluble  
 474 radionuclide dispersion in that location during that period [Sept. 2013-June 2016], since [Fiévet et al.,](#)  
 475 [\(2013\)](#) have demonstrated that HTO in seaweed very accurately reflects HTO in seawater when  
 476 considering the horizontal resolution of the model. H-3 levels in the NBG clearly resulted from  
 477 discharges by the RP but the levels were low. These observations were consistent with data from our  
 478 previous review of the radiological impact of H-3 discharges from French nuclear facilities in the  
 479 English Channel ([Fiévet et al., 2013](#)). The relationship between H-3 in seawater, H-3 in the biota in  
 480 the form of HTO, and H-3 in the form of OBT (Organically Bound Tritium) was addressed in the same  
 481 publication. OBT was determined in seaweed samples collected in the Channel Islands only in 2017  
 482 but the observed levels were consistent with HTO, in agreement with [Fiévet et al., \(2013\)](#).

483 **C-14** was measured in the Channel Islands in 1999, 2002, 2014, 2015 and 2017 and ranged from 268  
 484 to 360 Bq.Kg<sup>-1</sup> C. This reporting unit was chosen for comparison with the actual background C-14  
 485 level in the marine environment, which is presently around 249 Bq.Kg<sup>-1</sup> C according to [Muir et al.,](#)  
 486 [\(2017\)](#), and ranged from 237 to 242 Bq.Kg<sup>-1</sup> C on the basis of our observations at Concarneau and  
 487 Roscoff ([supplementary material: Fievet et al. 2020 Supplementary Material.xlsx](#)). The dispersion  
 488 pattern of C-14 across the NBG (Figure 10 C) was consistent with that of measured HTO (Figure 10 B)  
 489 and simulated H-3 in seawater (Figure 10 A). It was therefore considered as conservative. However,  
 490 C-14 is mainly released by the RP as dissolved inorganic carbon and this element is incorporated into  
 491 organic matter *via* photosynthesis (and then the food chain). The kinetics of this incorporation into  
 492 organic matter are slow and C-14 changes in the biota were smoothed compared to C-14 in seawater  
 493 ([Fiévet et al., 2006](#)). This was clearly observed in the monthly time-series measurements near the  
 494 outfall of the RP (Figure 12 C). Changes over time looked obviously smoothed with regard to the HTO  
 495 pattern (Figure 12 A). Our results were consistent with measurement results in the whole NBG  
 496 ([supplementary material: Fievet et al. 2020 Supplementary Material.xlsx](#)) as well as previously  
 497 published data ([Fiévet et al., 2006](#)). The fingerprint of C-14 discharges from the RP in the NBG was  
 498 clear. This radionuclide was pointed out as a major relative contributor to the very low dose to  
 499 humans from seafood consumption in the critical group (Goury fishermen) chosen by the Nord-  
 500 Cotentin Radioecology Group ([NCRG, 1999](#)). In short, H-3 and C-14 are definitely good markers of the  
 501 footprint of discharges by the ORANO RP in the NBG. They are both distributed with respect to  
 502 hydrodynamics between Cap La Hague and the NBG (Figure 10 A, B, C) and the present data did not  
 503 challenge the conclusions from previous studies ([Fiévet et al., 2006; 2013](#)).

504 One choice of indicator to assess the fate of discharges from nuclear fuel reprocessing in the marine  
 505 environment was the long-life radionuclide I-129, in particular for seaweed because of their well-  
 506 known capacity to concentrate iodine up to 10,000 times with respect to seawater ([Küpper et al.,](#)  
 507 [1998](#); [IAEA, 2004](#); [Leblanc et al., 2006](#)). Although I-129 was potentially present in background levels  
 508 (naturally occurring and as a remnant from past nuclear weapons tests), it was below our LoD in  
 509 seaweed from the reference locations at Concarneau and Roscoff. However, I-129 was routinely  
 510 detected in seaweed collected on shore in the NBG, as shown by our measurement results in *Fucus*  
 511 *serratus*. Most levels in the NBG were below 25 Bq.Kg<sup>-1</sup> dry while the highest levels, of up to 120  
 512 Bq.Kg<sup>-1</sup> dry, were observed in the vicinity of the RP outfall. The dispersion pattern across the NBG  
 513 (Figure 10 D) corresponded to the family of conservative radionuclides (Figure 10 B). When looking at  
 514 the monthly time-series close to Cape La Hague, the pattern in the seaweed was somewhat  
 515 smoothed out compared to that of HTO. As for C-14, the kinetics of the transfer between seawater  
 516 and seaweed were very likely to be responsible for this smoothing. Iodine transfer between seawater  
 517 and seaweed certainly deserves to be further investigated.

518 In 2014 and 2015 in the NBG, Cs-137 concentrations in seaweed ranged from 0.15 to 0.35 Bq.Kg<sup>-1</sup> dry  
 519 but, during the same period, the levels at Concarneau and Roscoff were in the same range.  
 520 Considering the measurement uncertainties, it can be concluded that the increase in Cs-137 levels  
 521 due to discharges in the NBG was low. In the annual time-series measurements from the Channel  
 522 Islands, the maximum levels of around 0.7 Bq.Kg<sup>-1</sup> dry observed in the late 1990s have dropped to  
 523 below 0.2 Bq.Kg<sup>-1</sup> dry at the present time, following the decrease in discharges (Figure 3 and  
 524 [supplementary material: Fievet\\_etal\\_2020\\_Supplementary\\_Material.pdf](#)). At Goury, about 6 Km  
 525 north of the RP outfall, recent Cs-137 concentrations in seaweed were very close to the background  
 526 level observed at the western end of the English Channel. Nowadays, Cs-137 is obviously not a good  
 527 marker of recent radioactive discharges by French nuclear facilities in the NBG because of the  
 528 existing background level. Figure 11 A confirmed this conclusion because, although Cs-137 can be  
 529 considered soluble, its dispersion pattern in the NBG did not match well that of other conservative  
 530 radionuclides. The subtraction of the background level estimated from average values at  
 531 Concarneau yielded some negative results in the NBG because of uncertainties on measurements.  
 532 Remobilisation from the sediment compartment of past Cs-137 discharges was also a probable  
 533 reason to explain the dispersion pattern illustrated in Figure 11 A. It is only very close to the outfall of  
 534 the RP that the monthly time-series measurements in seaweed displayed a noisy dispersion pattern  
 535 like that of HTO (Figure 12 D and A, respectively) and typical of soluble radionuclides.

536 *Non-conservative radionuclides: Co-60* is a local marker of radioactive liquid discharges from the  
 537 ORANO RP in the NBG (discharges from Flamanville NPP are two orders of magnitude lower) and it is  
 538 an example of a non-conservative radionuclide. Though it has a half-life of 5.27 years, it was  
 539 continuously present in the discharges from the plant. Co-60 was not detected in seaweed from  
 540 Concarneau and Roscoff but was routinely observed at Goury and occasionally in the northern area  
 541 of the NBG. The levels were mostly below 1 Bq.Kg<sup>-1</sup> dry and the highest levels observed were up to  
 542 3.4 Bq.Kg<sup>-1</sup> dry in the late 1990s in Alderney. The annual time-series in the Channel Islands reflected  
 543 the decrease in the amounts of discharge from the RP and two decades later, Co-60 in seaweed had  
 544 dropped to around 0.5 Bq.Kg<sup>-1</sup> dry. A wider view with the 2014-2015 data shows that Co-60 was only  
 545 detected in the northern area of the NBG. The dispersion pattern of Co-60 depicted in Figure 11B  
 546 differed drastically from that of typical soluble radionuclides (Figure 10), with much lower  
 547 "Dilution&transfer factor-like" values. Co-60 is known to be released mainly as particulate material

548 ([Gaudaire, 1996](#)) and a closer look at the fate of suspended matter would be crucial to interpret  
 549 properly that of Co-60 concentrations in the marine environment. Its interaction with particulate  
 550 material results in a potential reduction in its bioavailability to seaweed (the same applies to the soft  
 551 parts of limpets, [supplementary material: Fievet et al. 2020 Supplementary Material.xlsx](#)) and,  
 552 together with its radioactive decay, a likely major contributing process was its disappearance  
 553 because of sedimentation to the seabed. The further from the source of input, the longer it took to  
 554 reach the location, the less it remained in seawater and was available for transfer to the biota. The  
 555 monthly time-series at Goury (Figure 5), close to the source of input, very clearly displayed annual  
 556 cycling, although the monthly discharges did not (ORANO, personal comm.). Once Co-60 has reached  
 557 the sediment compartment, the potential return of sedimentary material to the water column would  
 558 bring up the signature of possibly ancient discharges. Since most radionuclide discharges from the RP  
 559 fell considerably in previous decades, the imprint of previous discharges is likely to influence the  
 560 signal in present suspended matter where resuspension occurs. The resuspension of sedimentary  
 561 material is thus a component to consider because contamination of marine biota samples by  
 562 suspended matter is likely. This seasonal cycle was confirmed when normalizing the concentrations  
 563 to the discharges. The profile displayed for Co-60 (Figure 12 E) clearly differed from that of a typically  
 564 soluble radionuclide (Figure 12 A, for example). Seaweed Co-60 concentrations were obviously partly  
 565 disconnected from actual discharges and were influenced by a seasonal cycle with highs in winter  
 566 and lows in summer. Recent specific studies provided evidence that resuspension of sedimentary  
 567 material displays annual cycling in the Channel ([Gohin, 2010](#); [Rivier, 2013](#)). Interestingly, the cycling  
 568 of Co-60 levels in seaweed and that of the concentration of resuspended sedimentary material in  
 569 seawater coincided with highs in winter and lows in summer. Correlation does not mean  
 570 demonstration but it suggests that further investigating would be worthwhile into the relationship  
 571 between Co-60 in suspended matter and subsequently in marine biota samples, with potential Co-60  
 572 contamination of the latter by the former.

573 **Ru-106** is another radionuclide which may be non-conservative because of its complex chemical  
 574 speciation and affinity to particulate forms ([Gandon et al., 1993](#)). However, even back in the late  
 575 1990s, it was not detected in seaweed from the Channel Islands and in 2014-2015, it was only  
 576 detected in seaweed and soft parts of limpets from Goury near the RP (Figure 5). Ru-106 discharges  
 577 from the RP dropped by two orders of magnitude in the early 1990s and this reflected accordingly on  
 578 the levels in the marine environment. Due to its short half-life (373 d), only recent discharges could  
 579 be detected. Ru-106 concentrations in the biota from Goury were normalized to discharges but the  
 580 data for seaweed included a majority of values below the LoDs and only data for the soft parts of  
 581 limpets were informative. As for Co-60, the Ru-106 data for the soft parts of limpets displayed a  
 582 profile with different "Dilution&transfer factor-like" values than HTO and a seasonal cycle with highs  
 583 in winter and lows in summer, typical of non-conservative radionuclides ([supplementary material:  
 584 Fievet et al. 2020 Supplementary Material.xlsx](#)).

585 **Pu-238, Pu-239,240** and **Am-241** levels were measured in seaweed samples in the NBG. These long-  
 586 lived radionuclides are well known to strongly adsorb to suspended matter and small mineral  
 587 particles, and to persist in the sediment compartment of the marine environment. Sediment acts as  
 588 an historical record of radionuclide levels ([Boust et al., 1996](#)) and the remobilisation of transuranic  
 589 elements from sediment can result in seawater concentrations that are not connected to present  
 590 discharges. This is especially crucial in the current context, given that discharges fell substantially  
 591 during the 1980s and 1990s, because the contribution of this remobilisation process can become the

592 main source of labelling for the seawater column. Since biological samples from the marine  
593 environment may include adsorbed (or ingested) mineral material with high concentrations of these  
594 radionuclides. Their total concentrations may thus be influenced by past discharges. The dispersion  
595 profiles of Pu-239,240 and Am-241 shown in Figure 11 were typical of non-conservative  
596 radionuclides. This was further supported by the normalized data from the monthly time-series for  
597 Goury which showed seasonal cycling profiles (Figure 12 F, G, H), also typical of non-conservative  
598 radionuclides like Co-60 (Figure 12 E) and Ru-106 ([supplementary material:  
599 Fievet et al. 2020 Supplementary Material.xlsx](#)). Nevertheless, the levels observed in seaweed  
600 collected in the NBG were in the  $[10^{-1}-10^{-2}]$  Bq.Kg<sup>-1</sup> dry order of magnitude, in agreement with  
601 previous reports ([Germain et al., 2000](#)). It is noteworthy that the isotopic ratio Pu-238/Pu-239,240  
602 was clearly inversely correlated with the distance from the source of radioactive input in the NBG  
603 ([supplementary material: Fievet et al. 2020 Supplementary Material.xlsx](#)). Values close to 0.7 in the  
604 biota from Goury agreed with the isotopic signature of radioactive discharges from the RP. The  
605 lowest values, close to 0.2, were observed in Roches Douvres and the SW of Guernsey whilst values  
606 around 0.3 were observed along the south coast of the NBG. This gradient with distance to the RP  
607 outfall reflected mixing with other sources. They include the background level resulting from fallout  
608 of past atmospheric nuclear tests with an isotopic signature of around 0.04 ([Boust et al., 1996, 1997;](#)  
609 [Boust 1999;](#) [Germain et al., 2000](#)). They also include the potential influence of resuspended  
610 sedimentary material, which carries the signature of past discharges from the RP. This issue deserves  
611 to be studied more thoroughly by characterizing the potential contribution of biota contamination by  
612 suspended matter.

613 *Conclusion:* Although most radionuclide discharges from the ORANO RP at La Hague decreased in the  
614 1980s, three radionuclides are discharged in proportion to nuclear fuel recycling activity, namely H-3,  
615 C-14 and I-129. The temporal and spatial distribution of the data are consistent with the history of  
616 the discharges, with most gamma emitter radionuclide levels being close to or below the current  
617 LoDs and a clear fingerprint of the ORANO RP at La Hague for the three aforementioned  
618 radionuclides. A comparison of the data for *Fucus serratus* I-129 distributions in the Channel Islands,  
619 in the whole NBG and close to the source of the radioactive discharge input (Goury) showed a sharp  
620 gradient, with the highest levels in close proximity to the RP and dropping rapidly through dilution in  
621 the NBG. The strong hydrodynamics in this area result in rapid mixing and dilution. The lowest levels  
622 of radionuclides were observed in the westernmost part (Guernsey and Roches Douvres) of the NBG.  
623 Nevertheless, the hydrodynamics in the NBG are non-intuitive and do not yield a simple gradient  
624 with linear distance from the outfall of the RP. Modelling tools were essential to understand how  
625 radioactive discharges spread from the source of input. Using comparison with modelling, dispersion  
626 patterns clearly illustrated the different behaviours of soluble and non-soluble radionuclides. Despite  
627 the high magnitude of the back-and-forth water currents related to the tide, soluble radionuclides  
628 spread in the NBG according to mean residual water currents, and the quantitative relationship with  
629 the discharges is consistent. Conversely, this quantitative relationship is not consistent for non-  
630 soluble radionuclides because their concentrations resulted from two sources: the present  
631 discharges and the sediment compartment, which is labelled by past discharges and by fallout from  
632 past nuclear tests. Non-soluble radionuclides associate with particulate material sediment of the  
633 seabed where they accumulate and potentially remobilise. Moreover, biota samples can be  
634 contaminated by small mineral particles (adsorbed or ingested). Since discharges of these non-  
635 soluble radionuclides have substantially declined since the 1980s, the contribution of this labelling of

636 sedimentary origin has potentially overwhelmed that of present discharges. The synchronicity of the  
637 signal in the biota for non-soluble radionuclides and the seasonal cycle of mineral suspended matter  
638 in the Channel supports this hypothesis. In depth investigation of the relative contributions of the  
639 two sources of labelling remains to be specifically carried out.

640 In summary, this comprehensive assessment of biota monitoring data indicated that the imprint of  
641 radioactive liquid discharges by French nuclear facilities was measurable in the Normand-Breton Gulf  
642 and the Channel Islands but the levels are low and generally declining. The temporal and spatial  
643 distribution of the data are consistent with the history of the discharges, with most gamma emitter  
644 radionuclide environmental levels being close to or below the actual limits of detection. This is in line  
645 with the conclusions from the Nord-Cotentin Radioecology Group assessment carried out two  
646 decades ago ([NCRG, 1999](#)) and is consistent with the subsequent decrease of most liquid  
647 radionuclide discharges by the ORANO La Hague reprocessing plant.

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661

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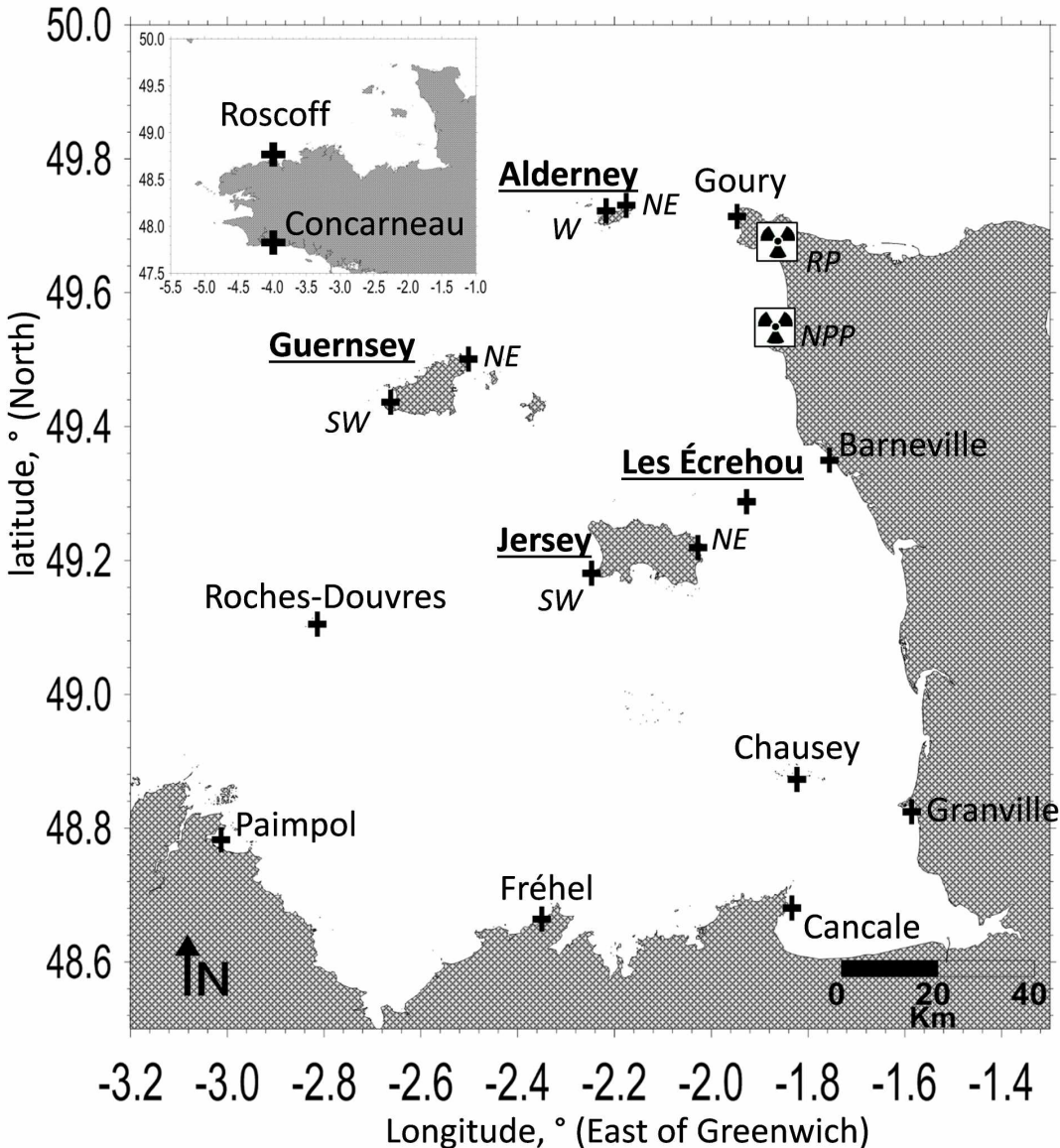
794 **Supplementary material**

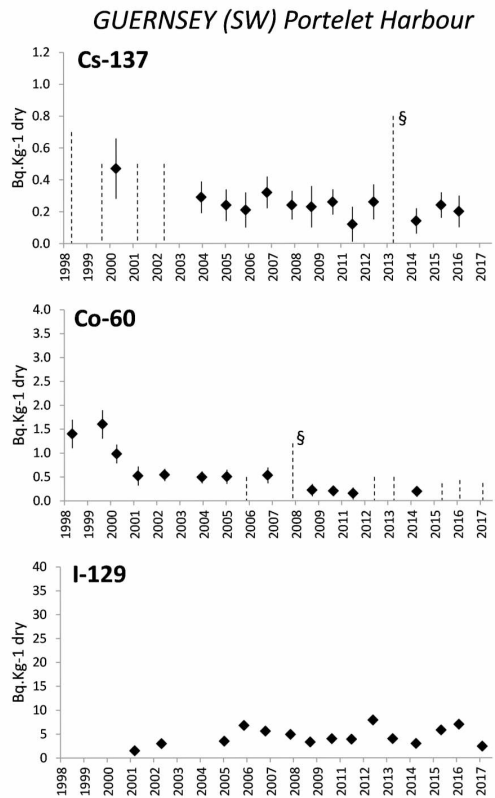
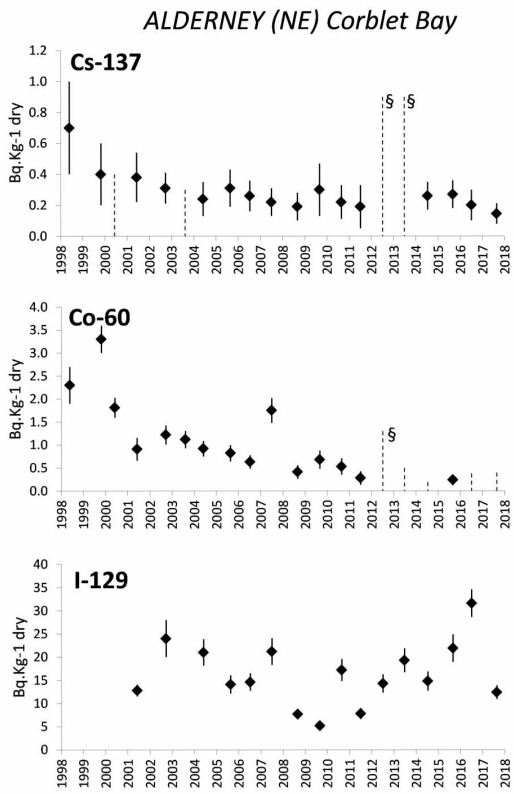
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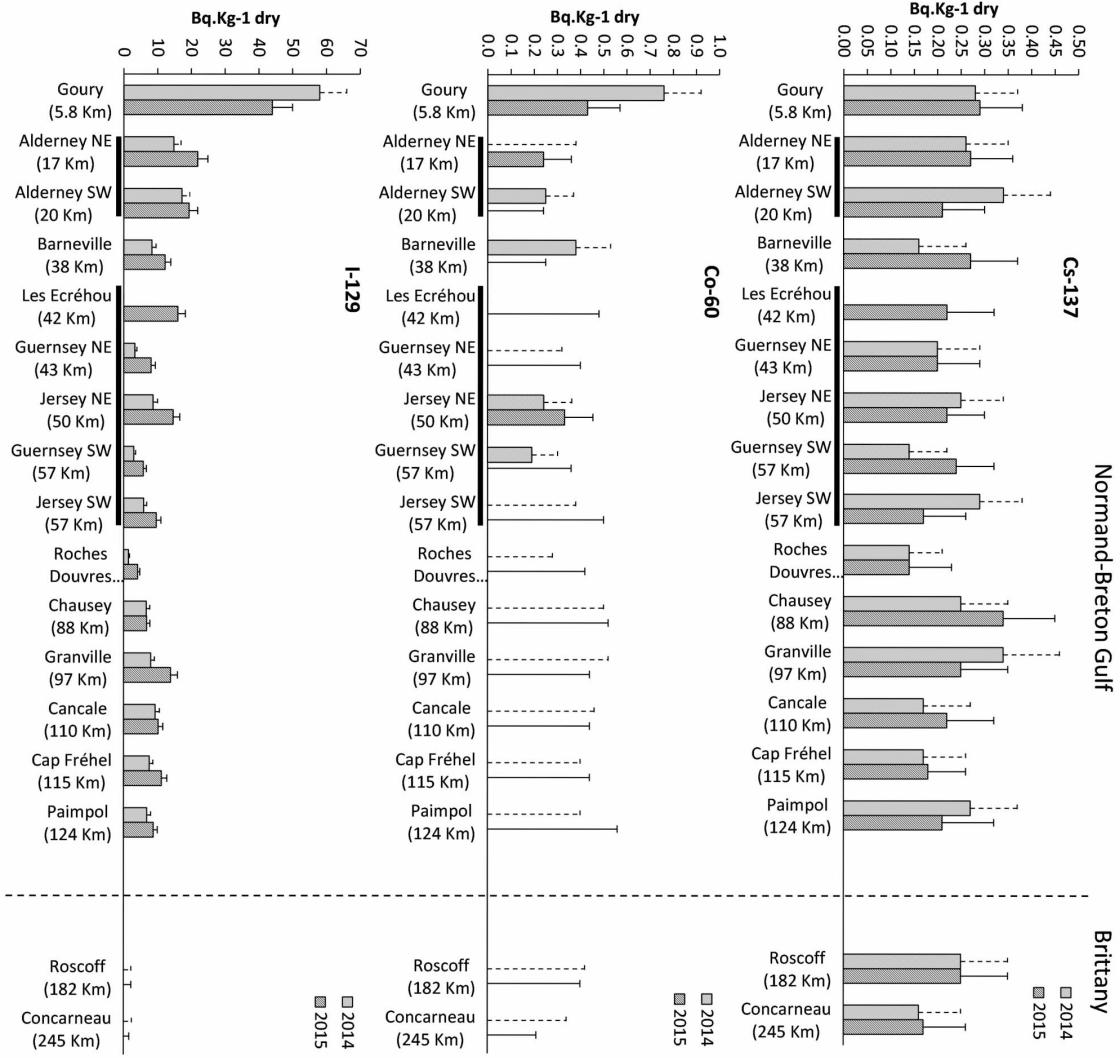
797 Additional figures in pdf format: Fievet\_etal\_2020\_Supplemental\_Material.pdf



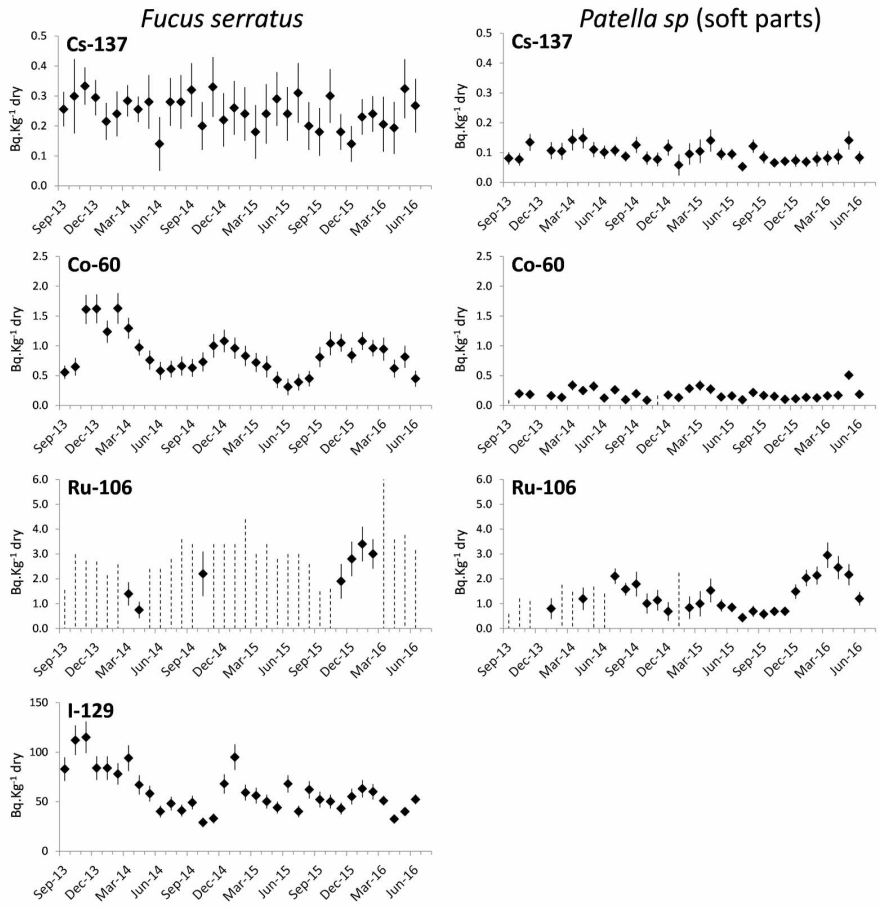


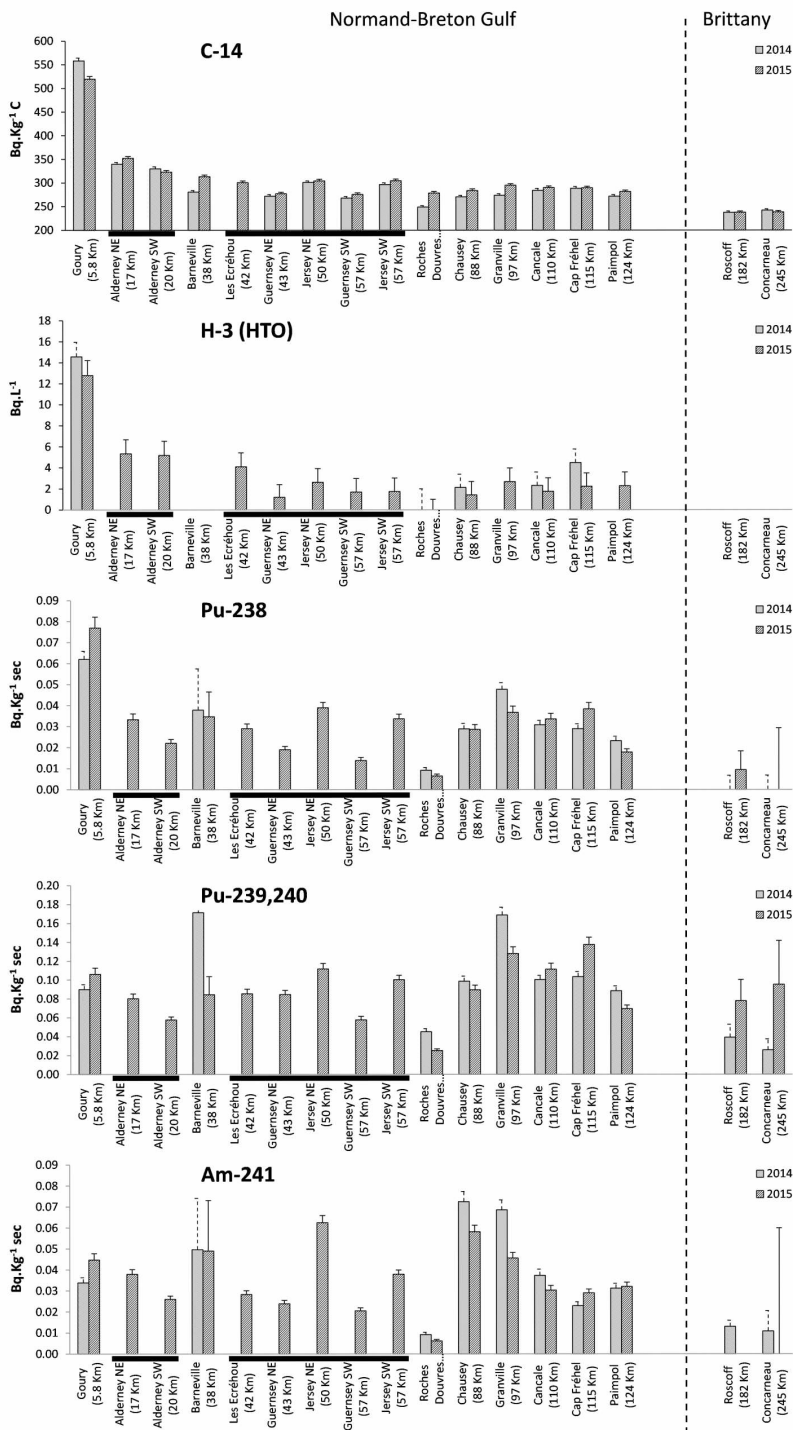




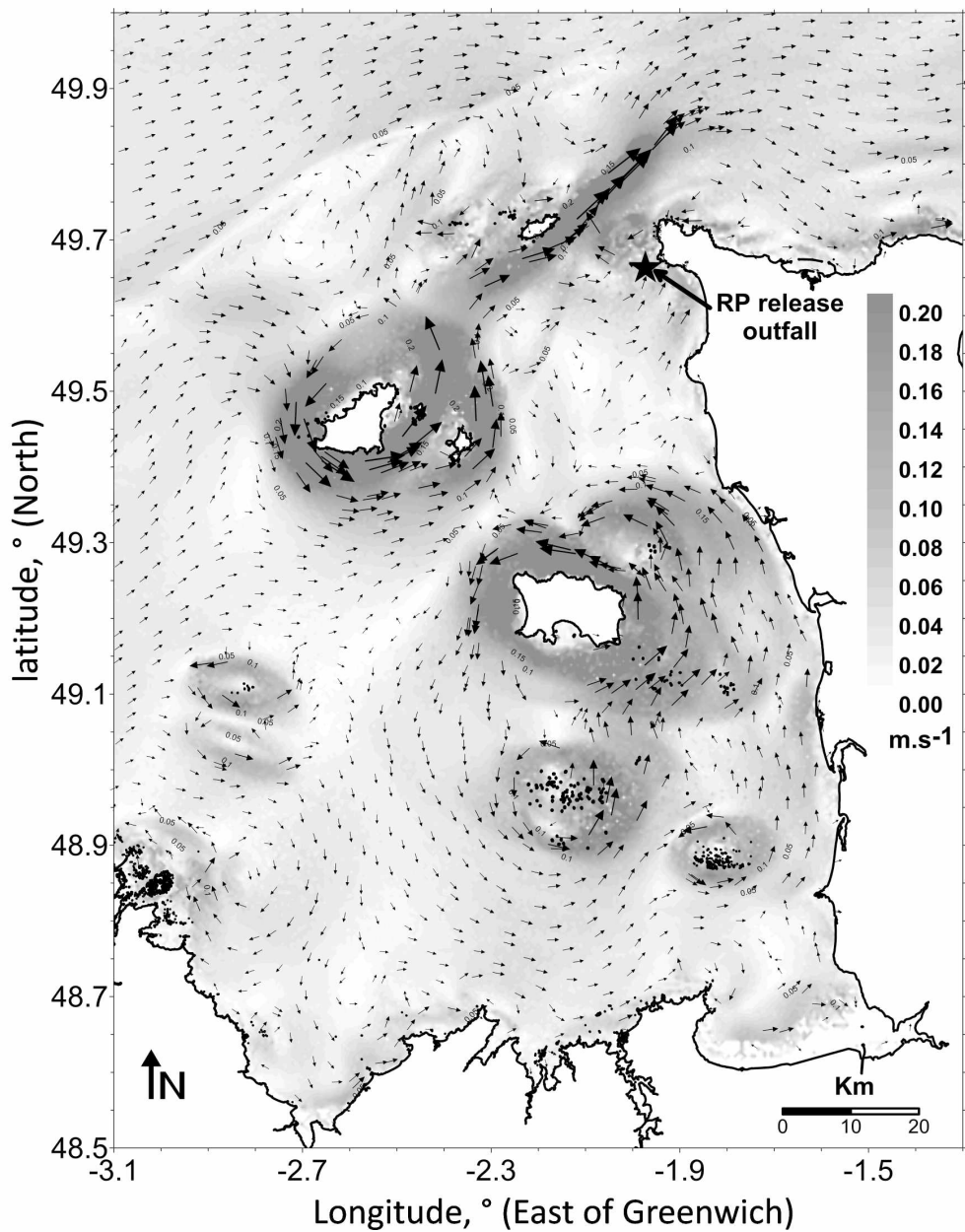


Radioactivity in the Channel Island  
Figure 4

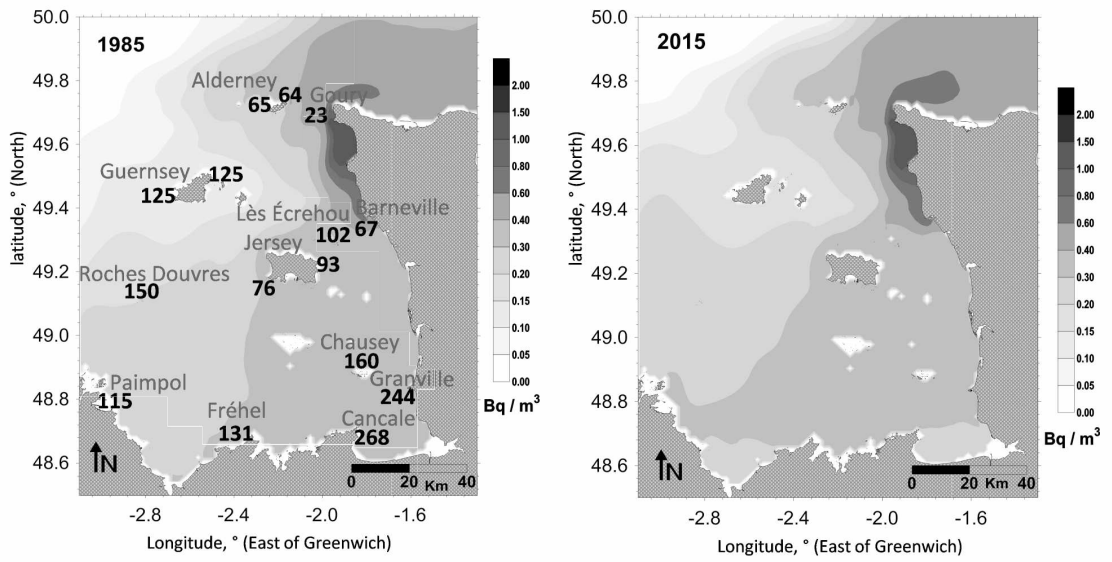




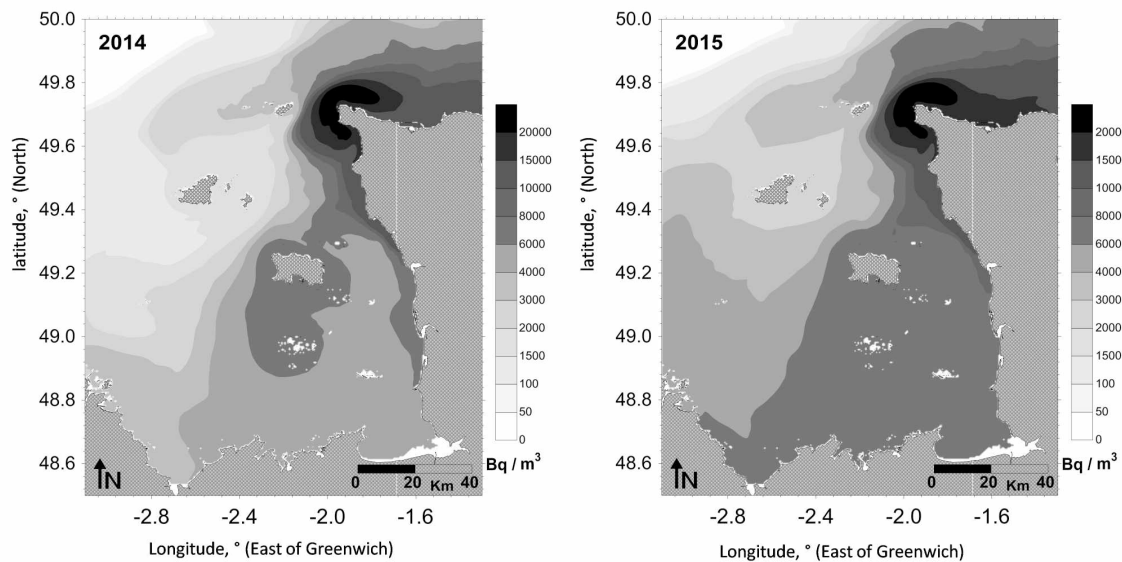
Average residual trajectory and velocity  
constant wind: SW(231°), 8m/s; medium tide: coeff. 70)

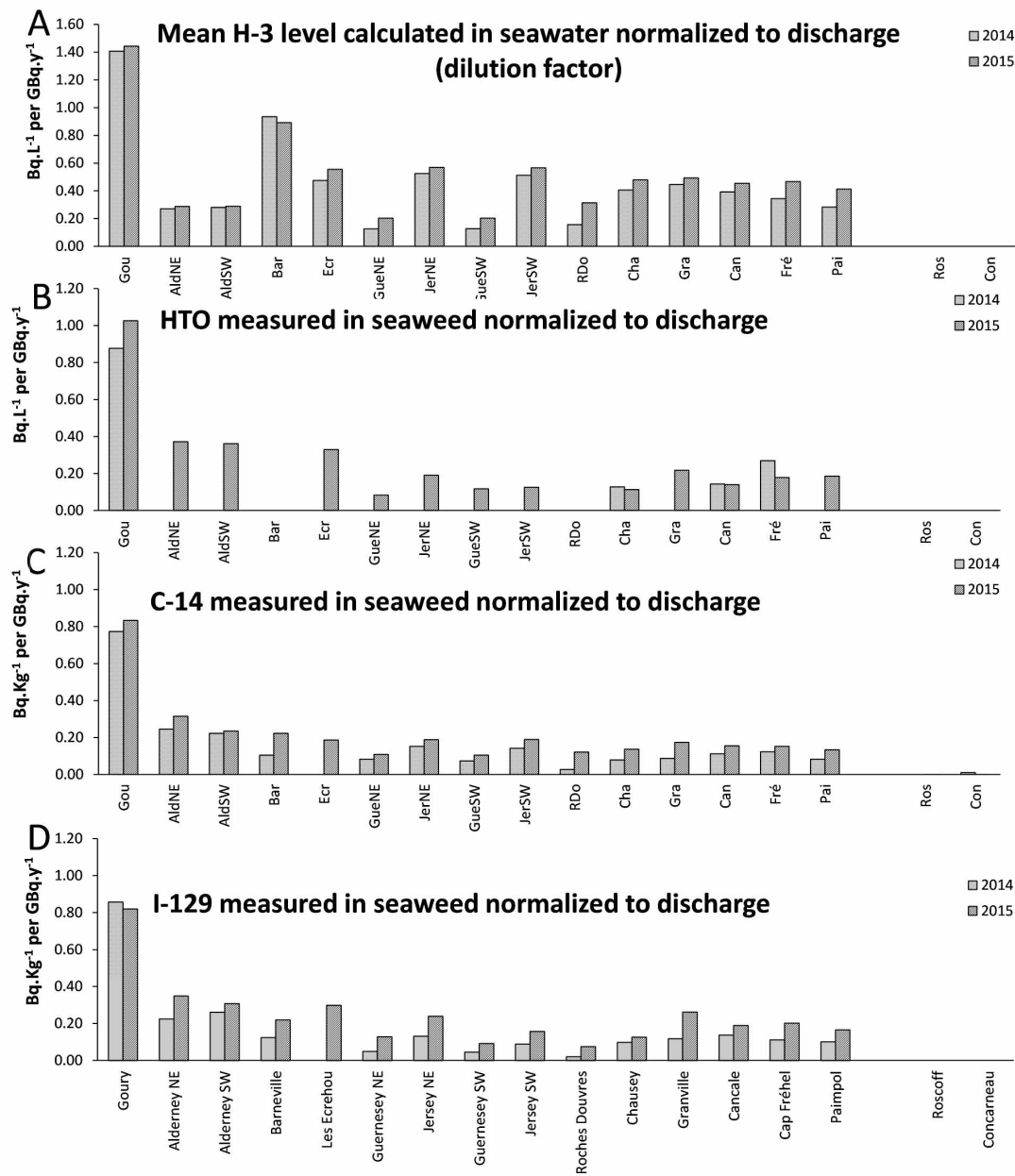


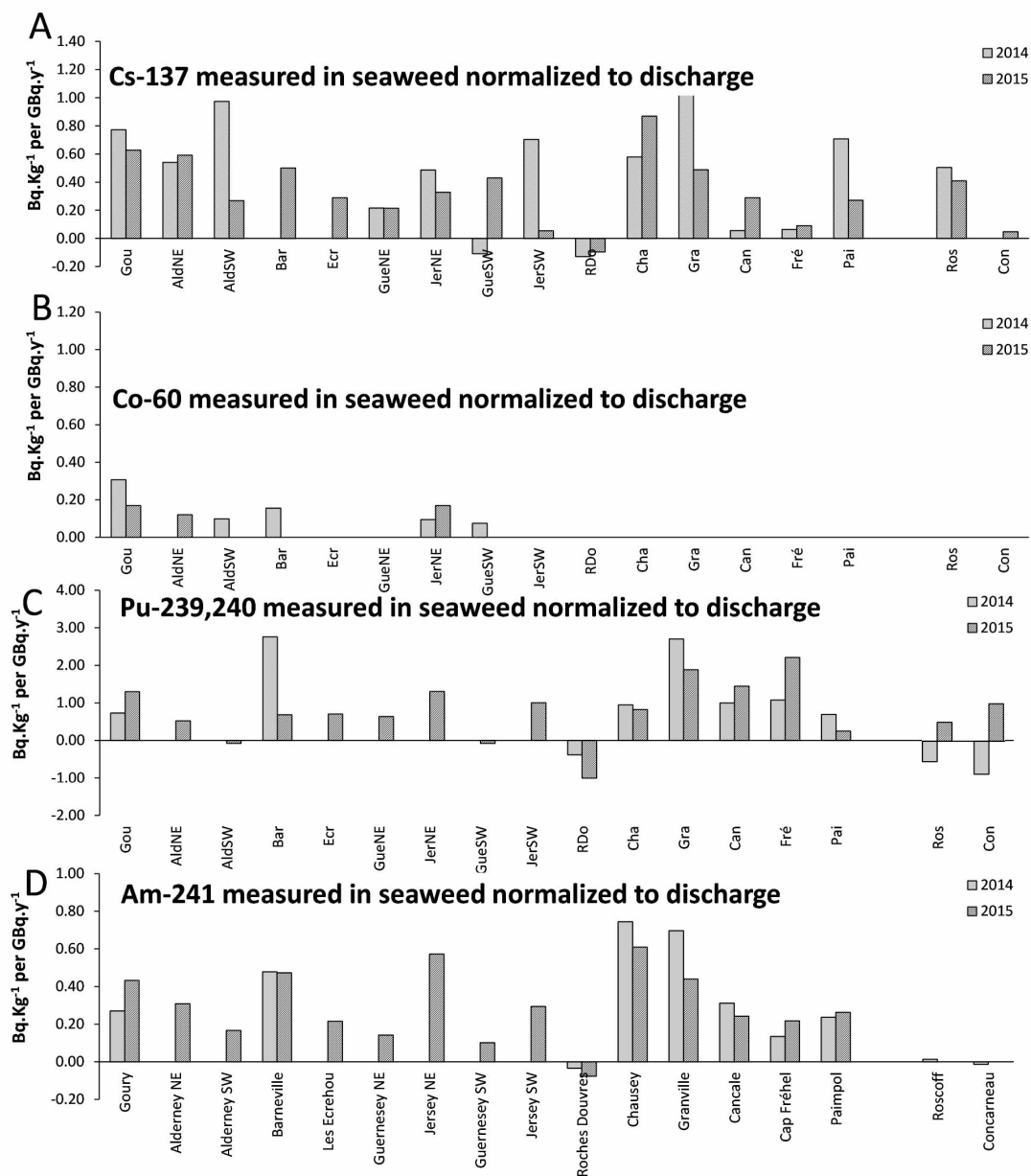
Radioactivity around the Channel Island  
Figure 8



Radioactivity around the Channel Island  
Figure 9

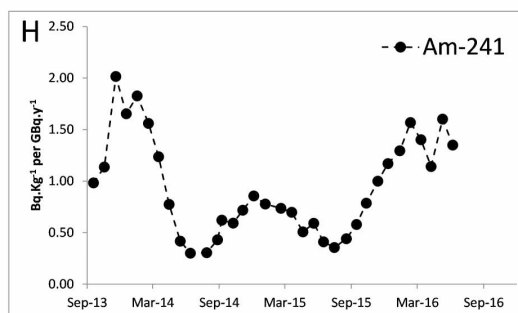
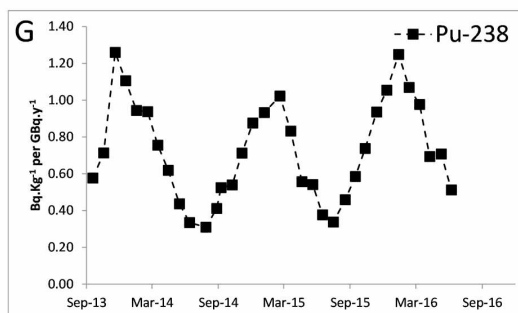
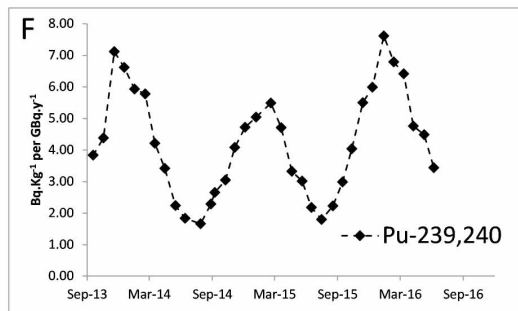
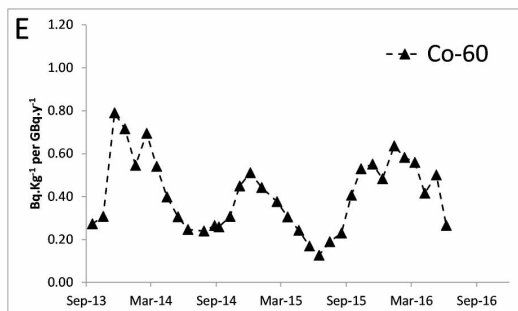
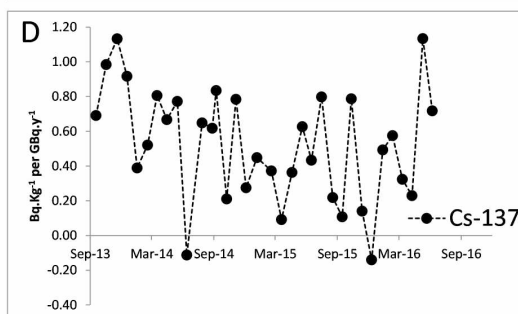
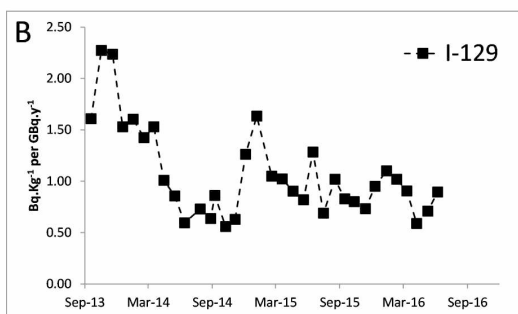
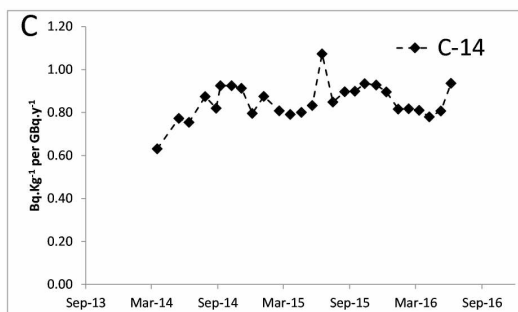
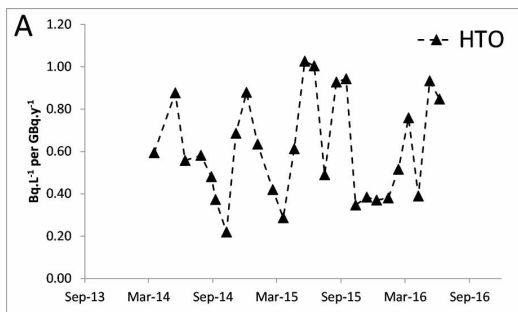








Radioactivity around the Channel Island  
Figure 12



## Radioactivity monitoring in the Channel Islands

### Table 1

Species	wet/dry ratio	Cs-137	Cs-134	Co-60	I-129	H-3	C-14	Pu-238	Pu-239,240	Am-241	Cm-244
Fucus sp.	5.04	50	50	6000	7000	1	2800	4000	4000	8000	5000
Patella/Mussel soft parts	5.00	60	60	1000	10	1	2300	3000	3000	8000	1000

### Table 2

Sampling date	ww/ dw	C-14 Bq.Kg <sup>-1</sup> C	H-3 HTO (OBT) Bq.L <sup>-1</sup>	Pu-238	Pu-239,240 Bq.Kg <sup>-1</sup> dry	Am-241
<i>ALDERNEY (NE) Corblet Bay</i>						
27/10/1999	4.7	360 ± 9				
06/06/2000	6.0			0.0399 ± 0.0038	0.0915 ± 0.0071	0.0415 ± 0.0067
24/09/2002	4.6	330 ± 5				
26/08/2010	5.3		<2.0			
05/07/2011	5.6		4.6 ± 1.2			
05/07/2012	5.7		5.9 ± 1.3			
25/06/2013	5.8			0.0151 ± 0.0016	0.0435 ± 0.0030	0.0174 ± 0.0016
02/09/2015	5.2		5.3 ± 1.3	0.0333 ± 0.0028	0.0802 ± 0.0051	0.0379 ± 0.0023
06/07/2016	5.6		6.7 ± 1.4			
24/08/2017	3.9	356 ± 3	7.4 ± 1.4 (5.8 ± 0.4)			
<i>ALDERNEY (SW) Fort Tourgis</i>						
28/10/1999	4.8	344 ± 9				
06/06/2000	5.8			0.0333 ± 0.0031	0.0839 ± 0.0058	0.0328 ± 0.0027
24/09/2002	4.6	319 ± 4				
26/08/2010	5.2		<2.0			
05/07/2011	5.6		5.4 ± 1.2			
05/07/2012	6.0		5.5 ± 1.3			
25/06/2013	5.2			0.0131 ± 0.0018	0.0527 ± 0.0043	0.0127 ± 0.0014
02/09/2015	5.4		5.2 ± 1.3	0.0221 ± 0.0018	0.0576 ± 0.0032	0.0260 ± 0.0016
06/07/2016	5.1		6.6 ± 1.4			
24/08/2017	3.9	331 ± 3	5.3 ± 1.4 (5.0 ± 0.4)			
<i>GUERNSEY (NE) Mielette</i>						
29/10/1999	4.7	297 ± 8				
23/09/2002	4.6	284 ± 4				
06/06/2000	6.2			0.0299 ± 0.0042	0.0894 ± 0.0081	0.0803 ± 0.0071
25/08/2010	5.2		1.9 ± 1.1			
04/07/2011	4.9		4.2 ± 1.2			
04/07/2012	5.8		1.9 ± 1.3			
24/06/2013	6.0			0.0164 ± 0.0018	0.0660 ± 0.0047	0.0135 ± 0.0013
01/09/2015	5.3			0.0189 ± 0.0016	0.0845 ± 0.0043	0.0239 ± 0.0016
01/09/2015	5.3		1.2 ± 1.2			
05/07/2016	5.0		1.5 ± 1.3			
23/08/2017	4.1	278 ± 3	<2.0 (2.5 ± 0.2)			

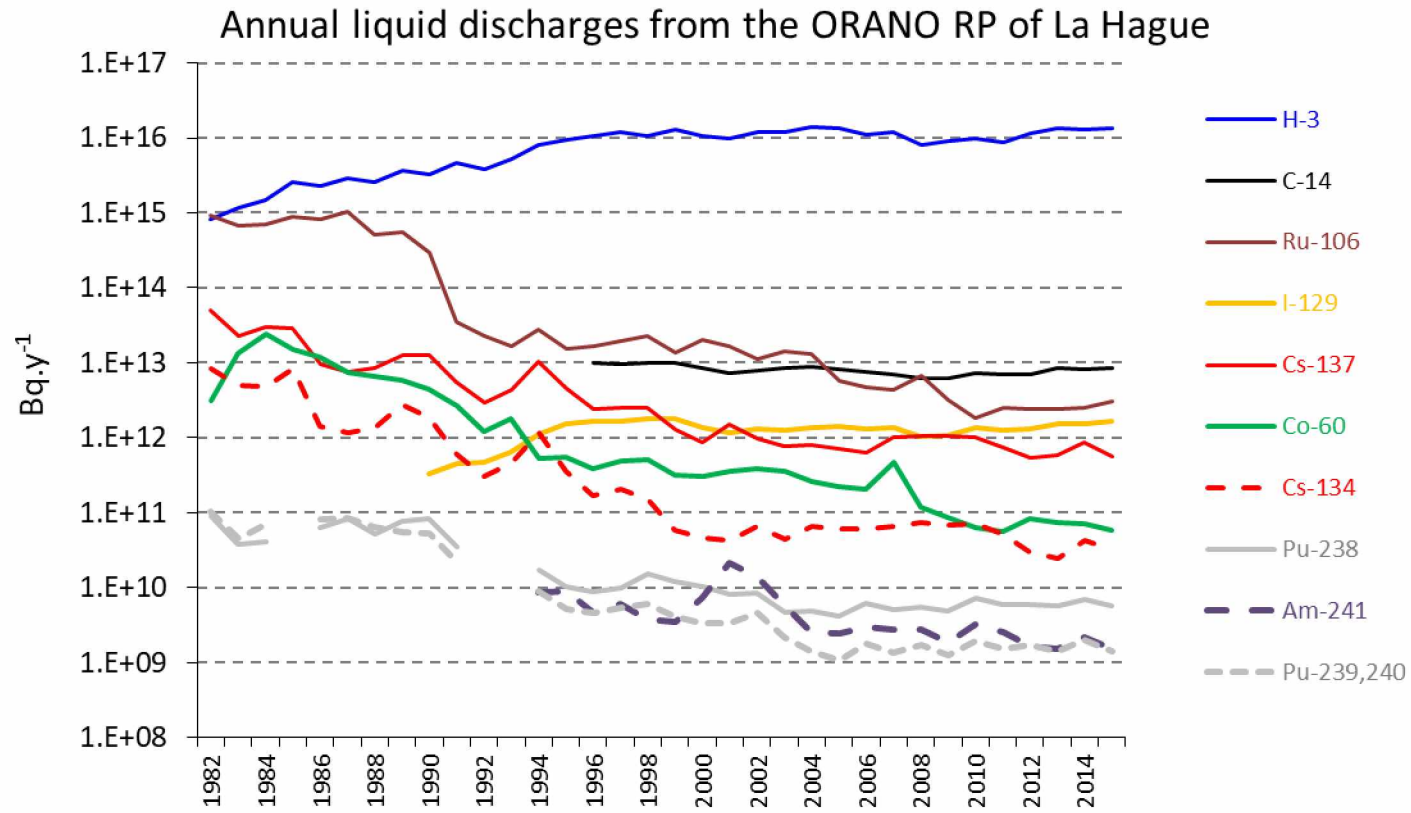
Table 2 (Cont.)

<i>GUERNSEY (SW) Portelet Harbour</i>						
29/10/1999	4.7	298 ± 8				
23/09/2002	4.1	274 ± 3				
06/06/2000	5.2		0.0204 ± 0.0031	0.0627 ± 0.0062	0.0298 ± 0.0023	
25/08/2010	5.1		1.7 ± 1.1			
04/07/2011	4.7		3.6 ± 1.2			
24/06/2013	6.5		0.0111 ± 0.0012	0.0529 ± 0.0029	0.0121 ± 0.0012	
01/09/2015	4.8		0.0138 ± 0.0014	0.0576 ± 0.0037	0.0205 ± 0.0014	
04/07/2012	5.8		1.5 ± 1.2			
01/09/2015	4.8		1.7 ± 1.3			
05/07/2016	4.1		1.3 ± 1.3			
23/08/2017	3.3	268 ± 3	<2.0 (2.4 ± 0.2)			
<i>JERSEY (NE) Anse de Ste Catherine</i>						
06/07/2012	5.5		2.7 ± 1.3			
31/08/2015	5.3		0.0390 ± 0.0025	0.1119 ± 0.0057	0.0624 ± 0.0034	
31/08/2015	5.3		2.6 ± 1.3			
04/07/2016	4.5		2.7 ± 1.3			
22/08/2017	3.6	298 ± 3	2.6 ± 1.3 (3.5 ± 0.3)			
<i>JERSEY (SW) Corbière Lighthouse</i>						
06/07/2012	5.4		2.4 ± 1.3			
26/06/2013	5.7		0.0327 ± 0.0023	0.1016 ± 0.0049	0.0308 ± 0.0023	
31/08/2015	5.8		1.7 ± 1.3	0.0336 ± 0.0022	0.1001 ± 0.0051	0.0379 ± 0.0020
04/07/2016	4.5		4.0 ± 1.3			
22/08/2017	3.8	297 ± 3	2.9 ± 1.3 (3.4 ± 0.2)			

## Declaration of interests

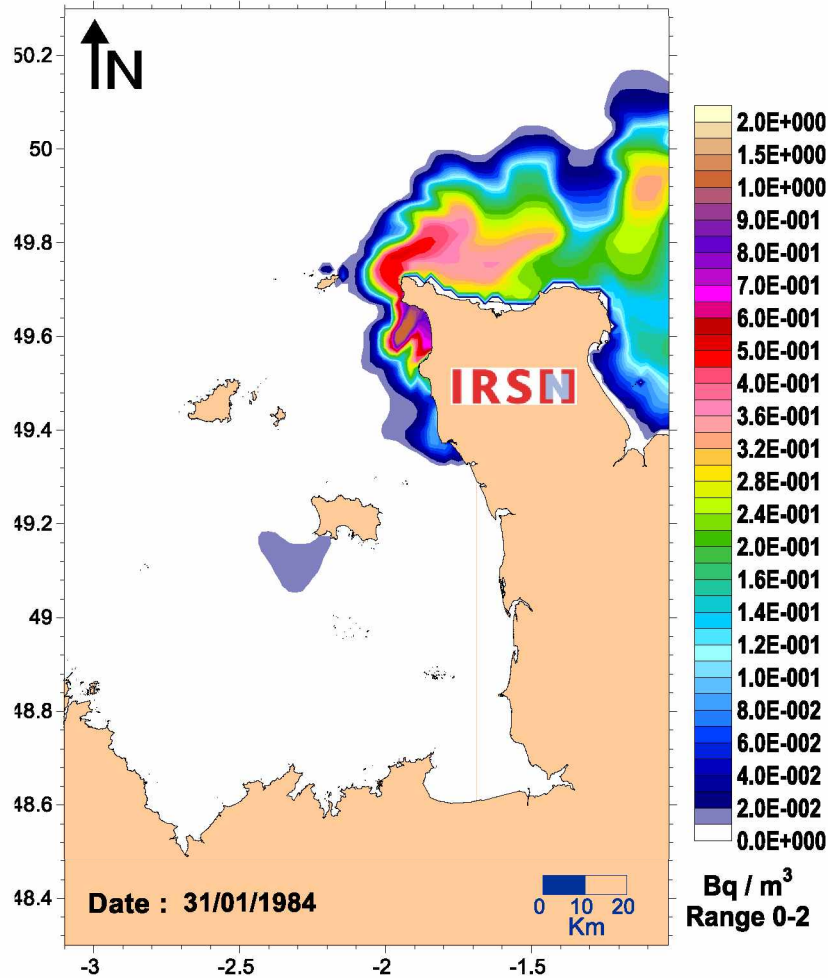
The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:



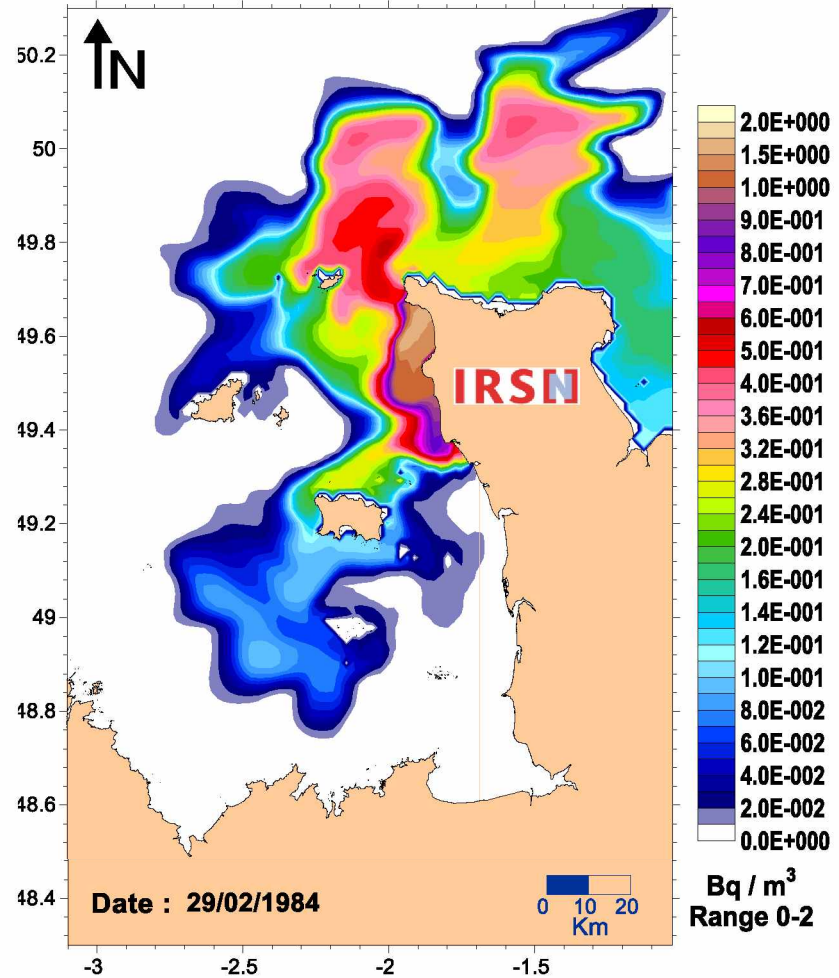
# Reaching a steady state

Concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )



Hydrodynamic modelling (MARS 2D)

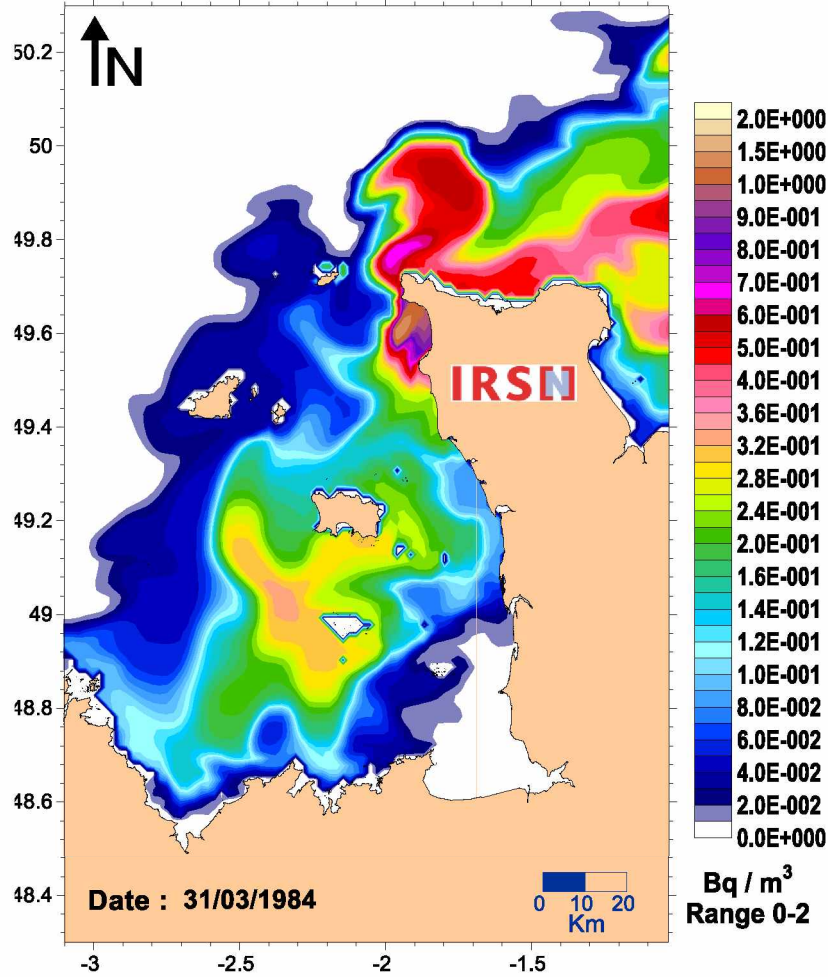
Concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )



Hydrodynamic modelling (MARS 2D)

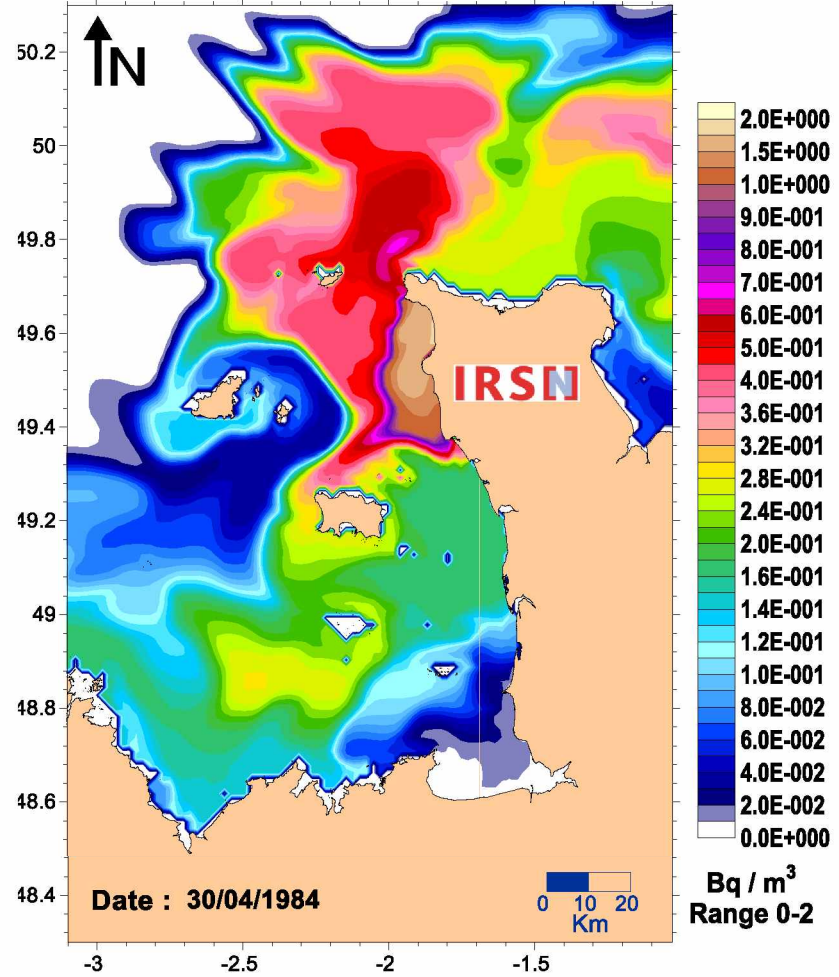
# Reaching a steady state

Concentrations resulting from constant discharge of 1 TBq.y<sup>-1</sup> (31 709 Bq.s<sup>-1</sup>)



Hydrodynamic modelling (MARS 2D)

Concentrations resulting from constant discharge of 1 TBq.y<sup>-1</sup> (31 709 Bq.s<sup>-1</sup>)

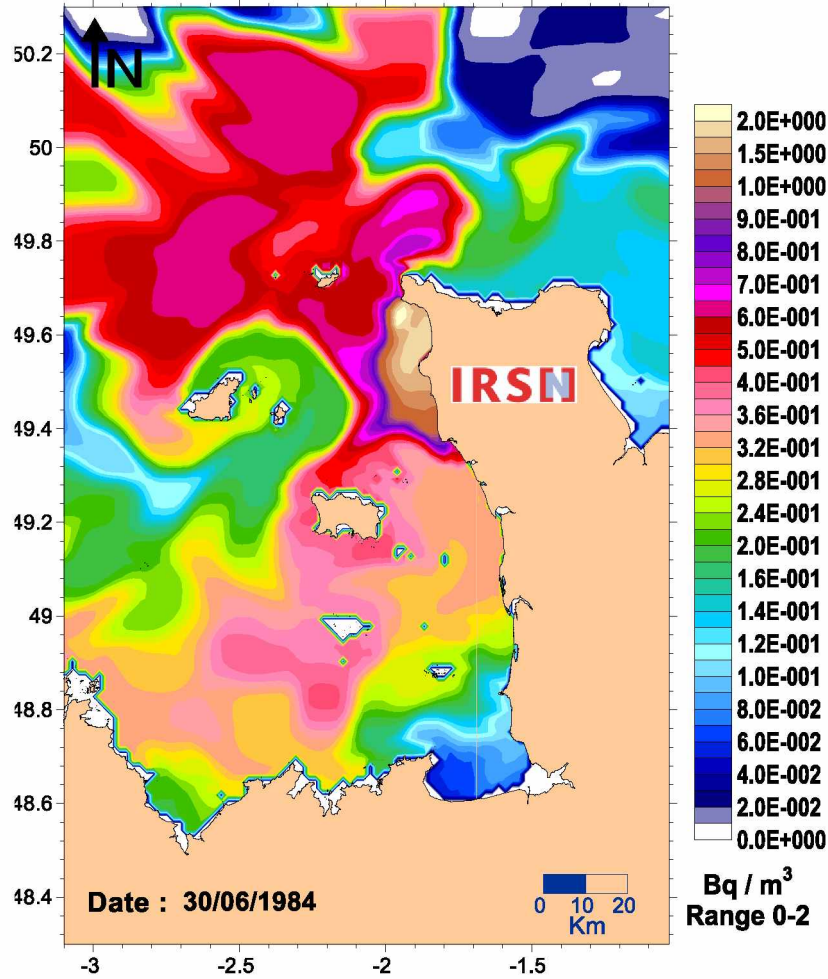


Hydrodynamic modelling (MARS 2D)

# Reaching a steady state

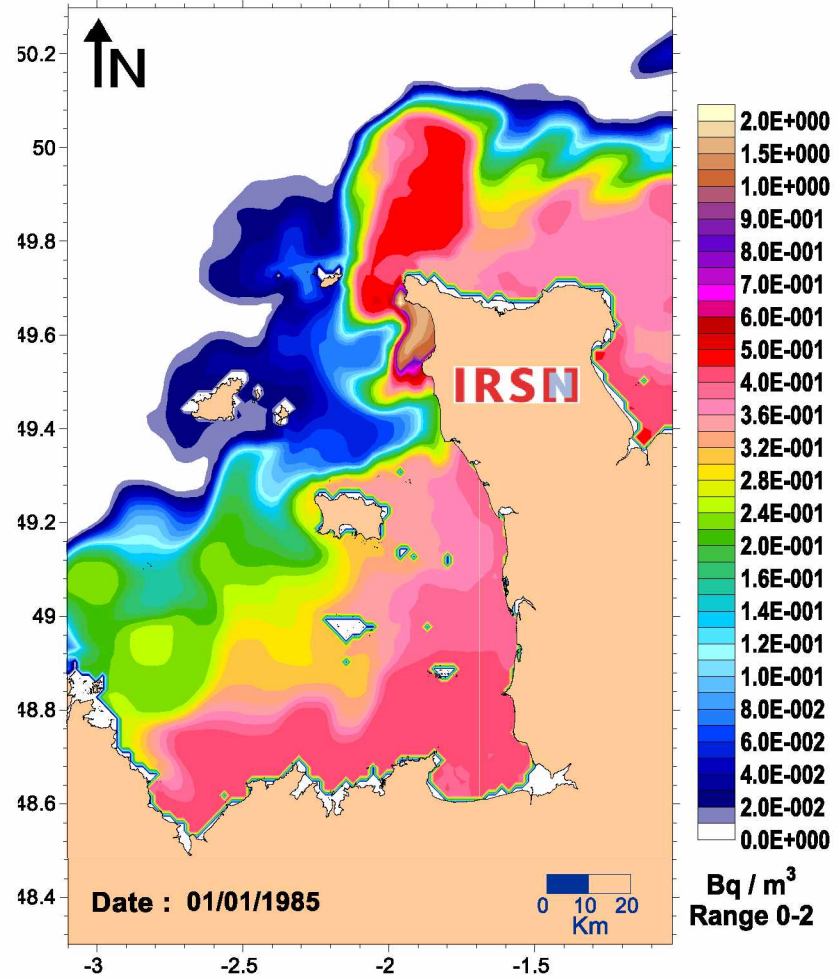
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Hydrodynamic modelling (MARS 2D)

Concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )

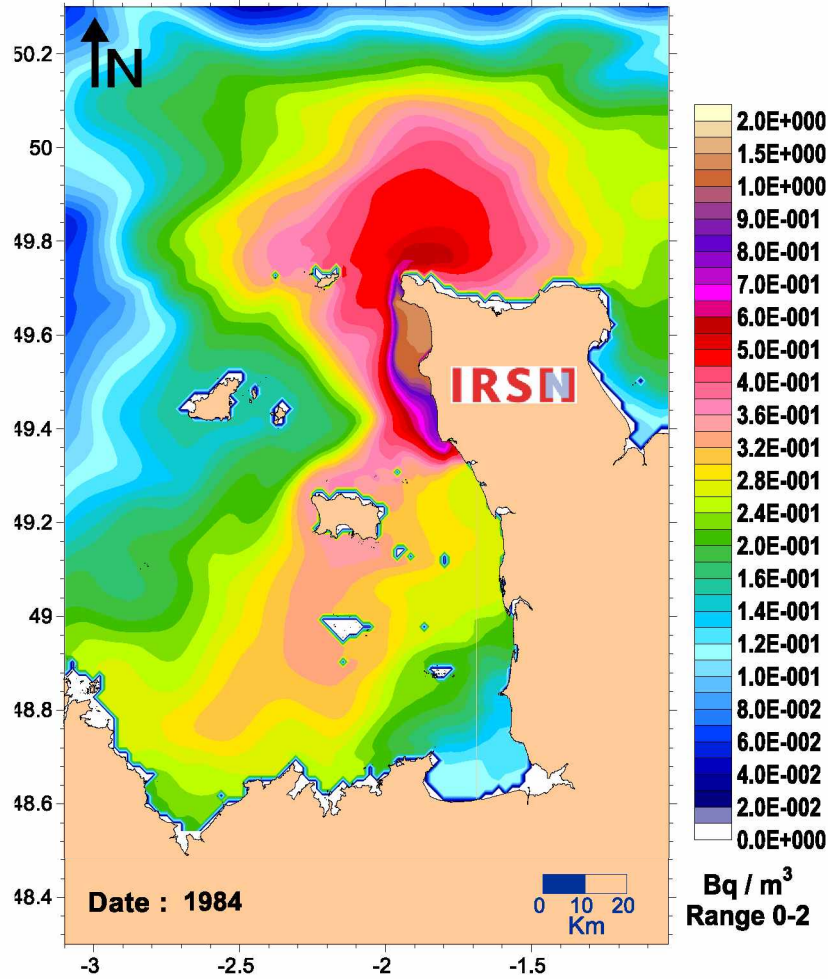


Hydrodynamic modelling (MARS 2D)



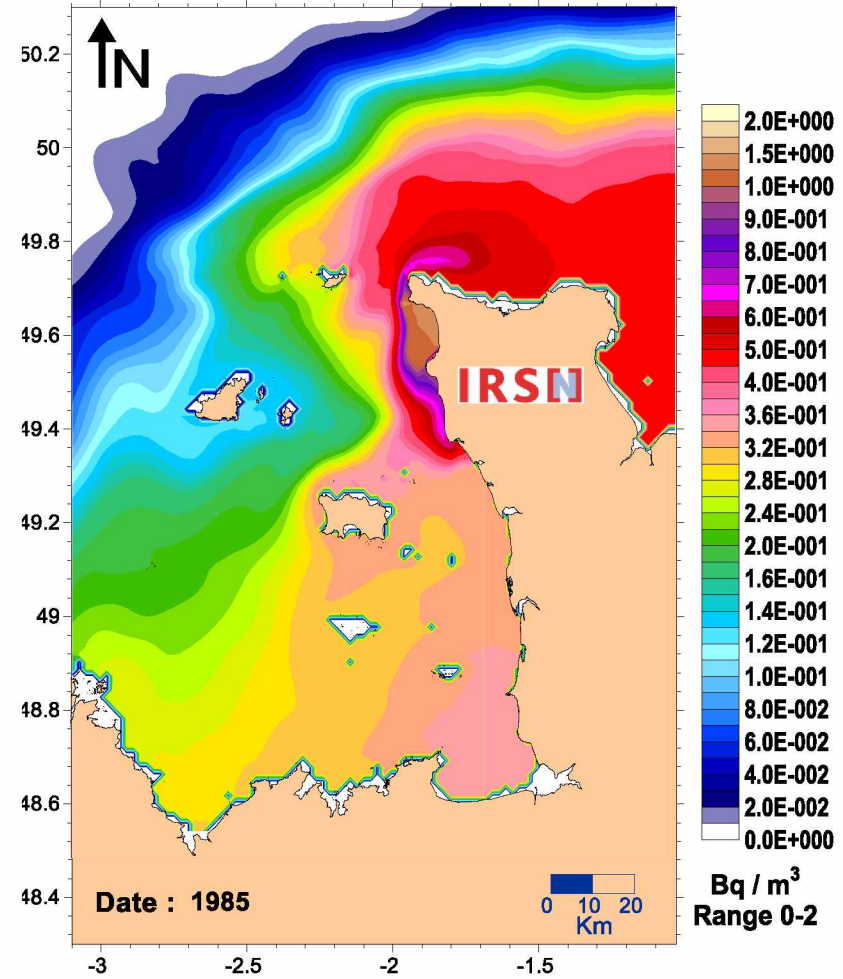
# Annual means

Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )



Hydrodynamic modelling (MARS 2D)

Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )

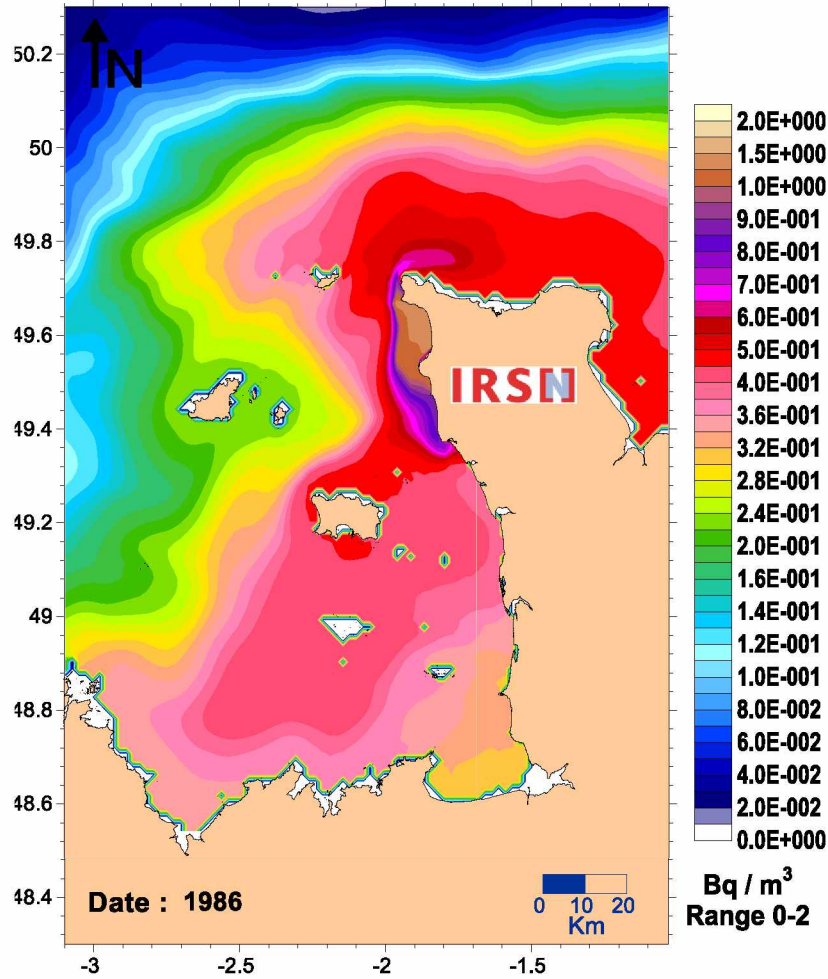


Hydrodynamic modelling (MARS 2D)

# Annual means

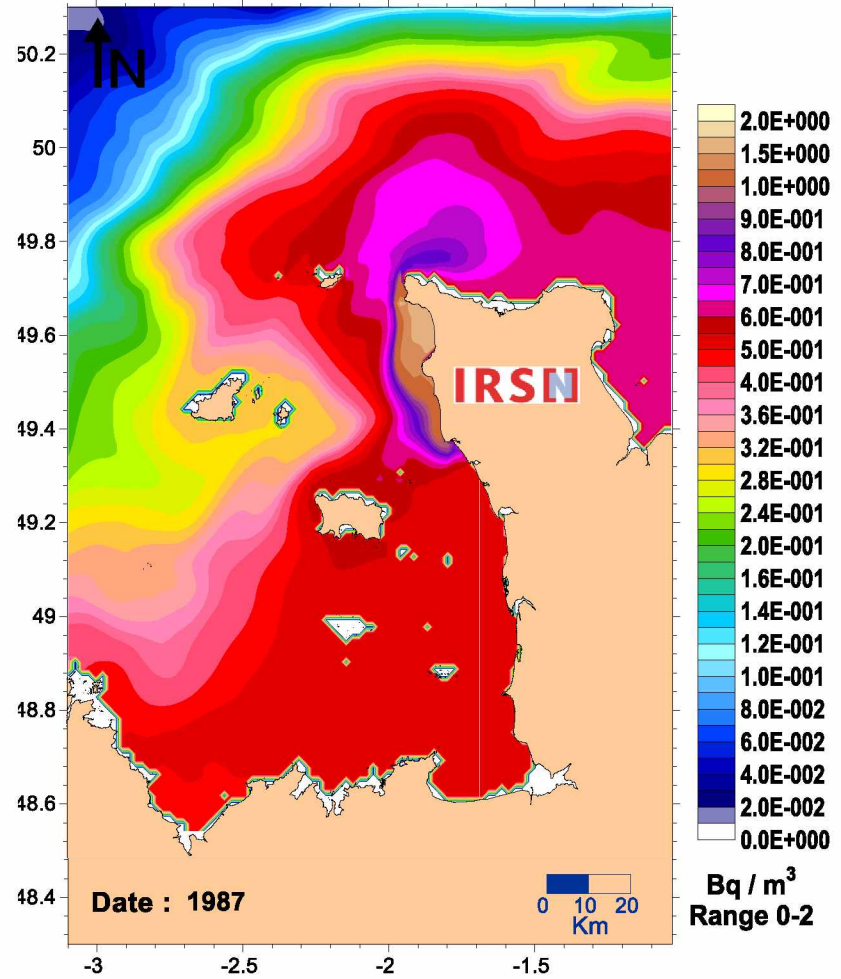
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Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )



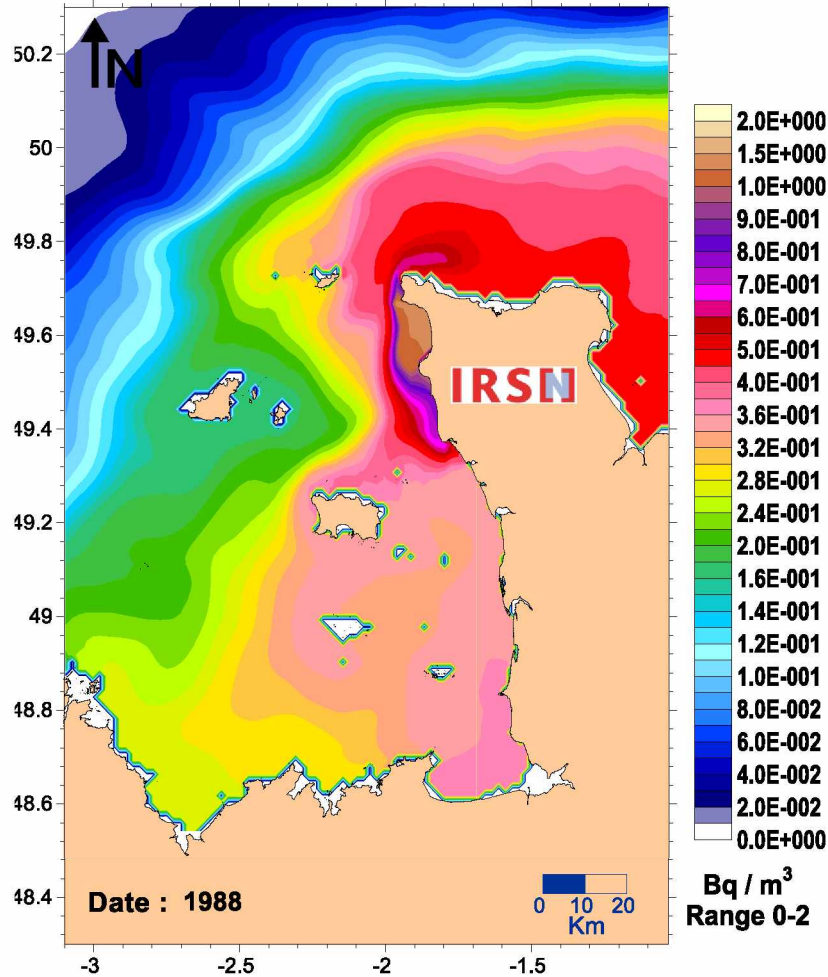
Hydrodynamic modelling (MARS 2D)

Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )

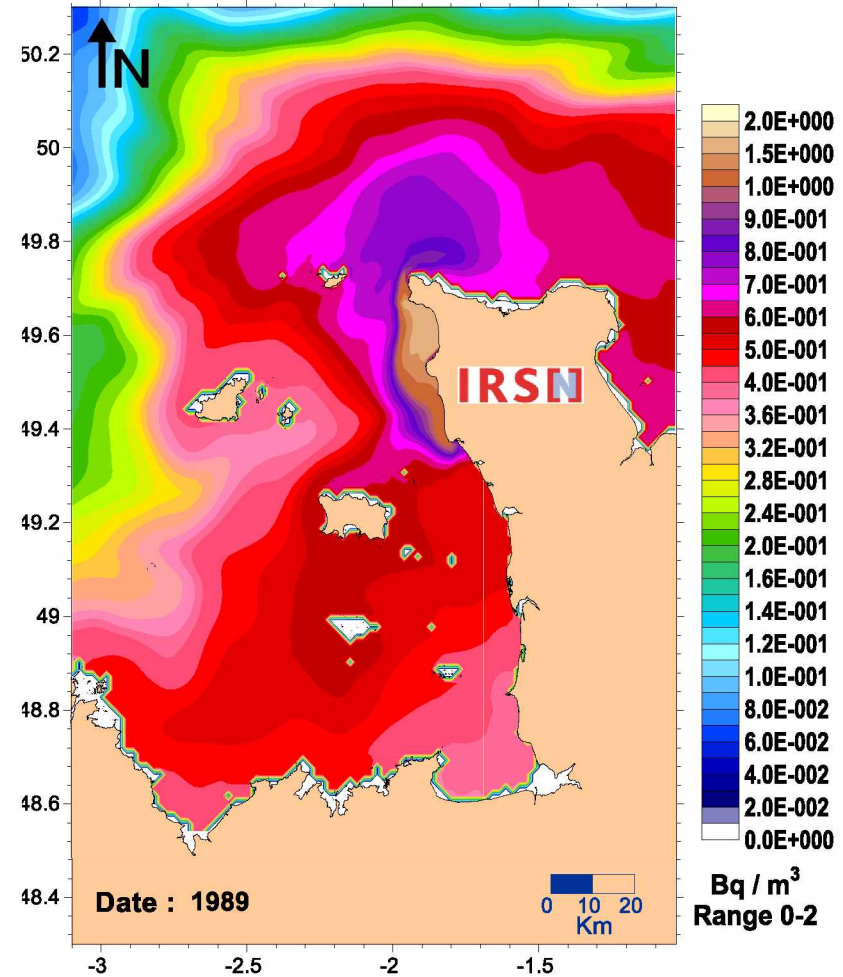


Hydrodynamic modelling (MARS 2D)

## Annual means

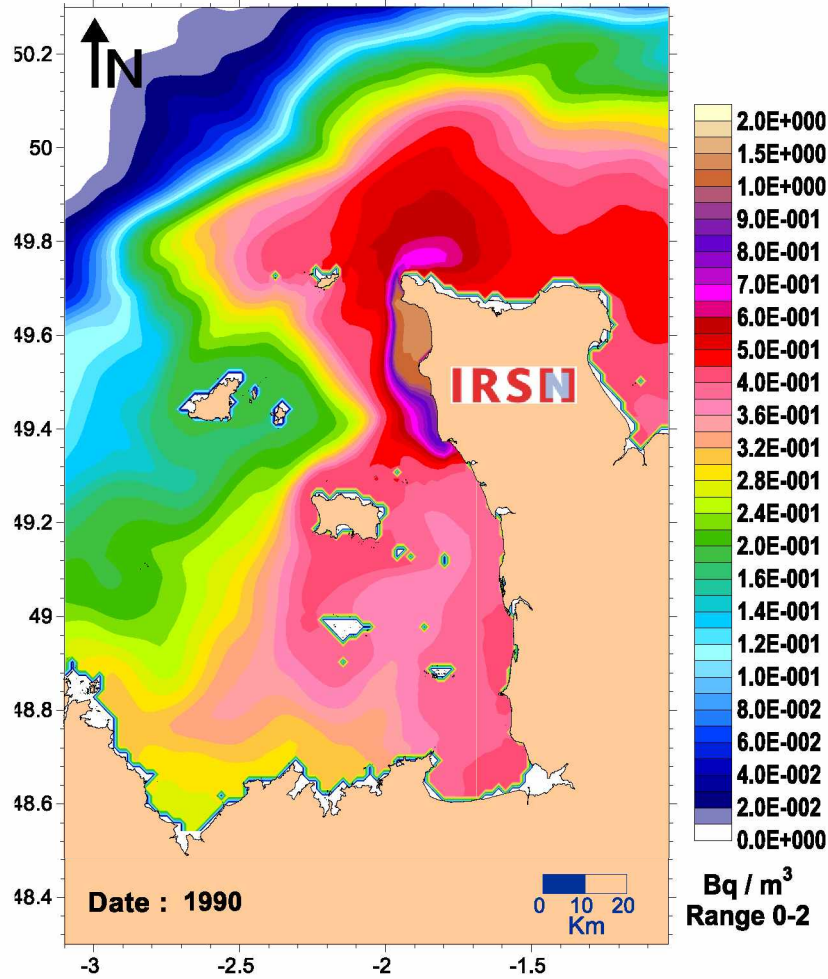
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Hydrodynamic modelling (MARS 2D)

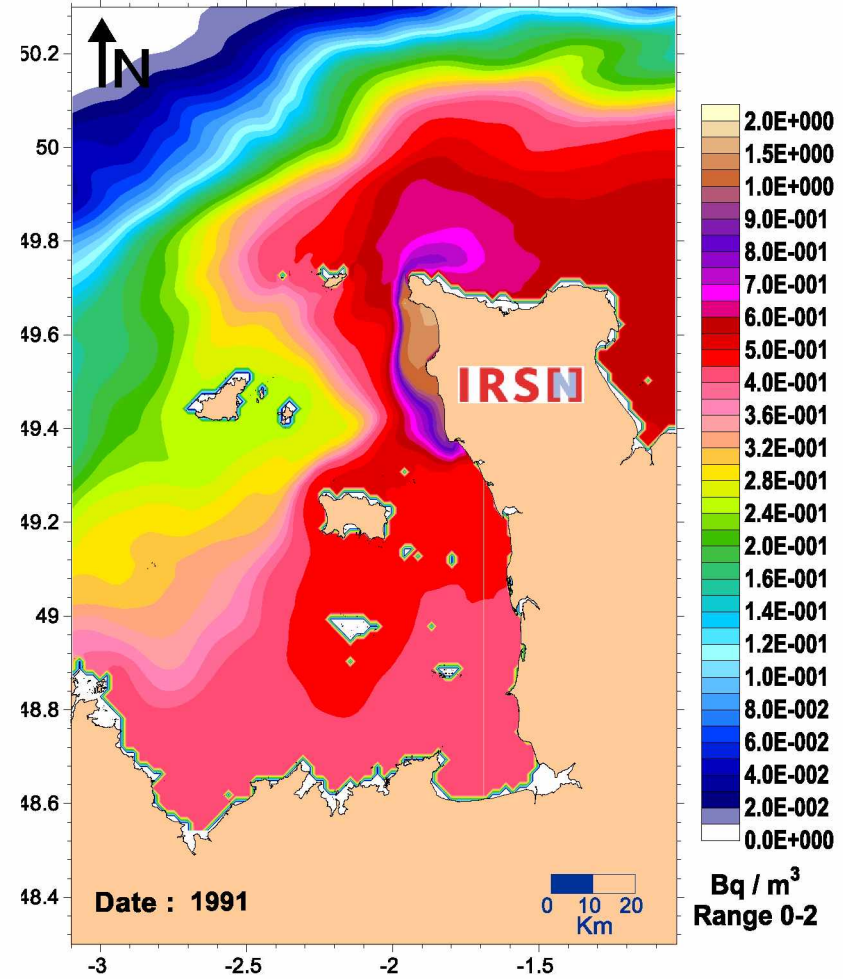
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Hydrodynamic modelling (MARS 2D)

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Hydrodynamic modelling (MARS 2D)

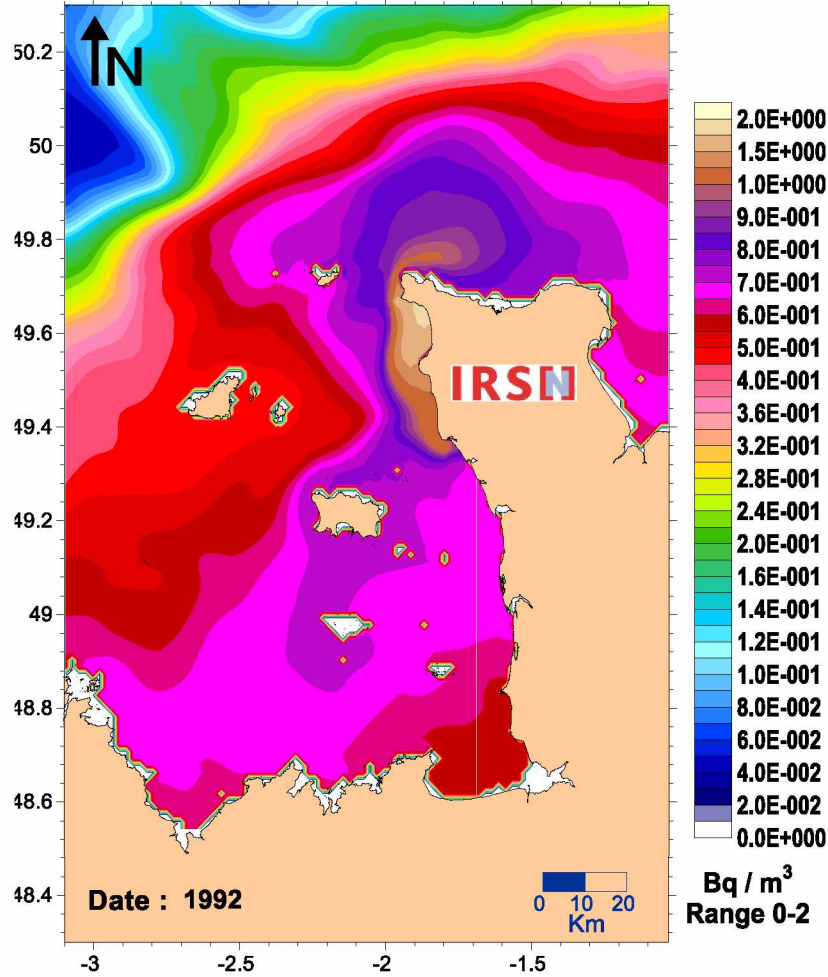
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Hydrodynamic modelling (MARS 2D)

# Annual means

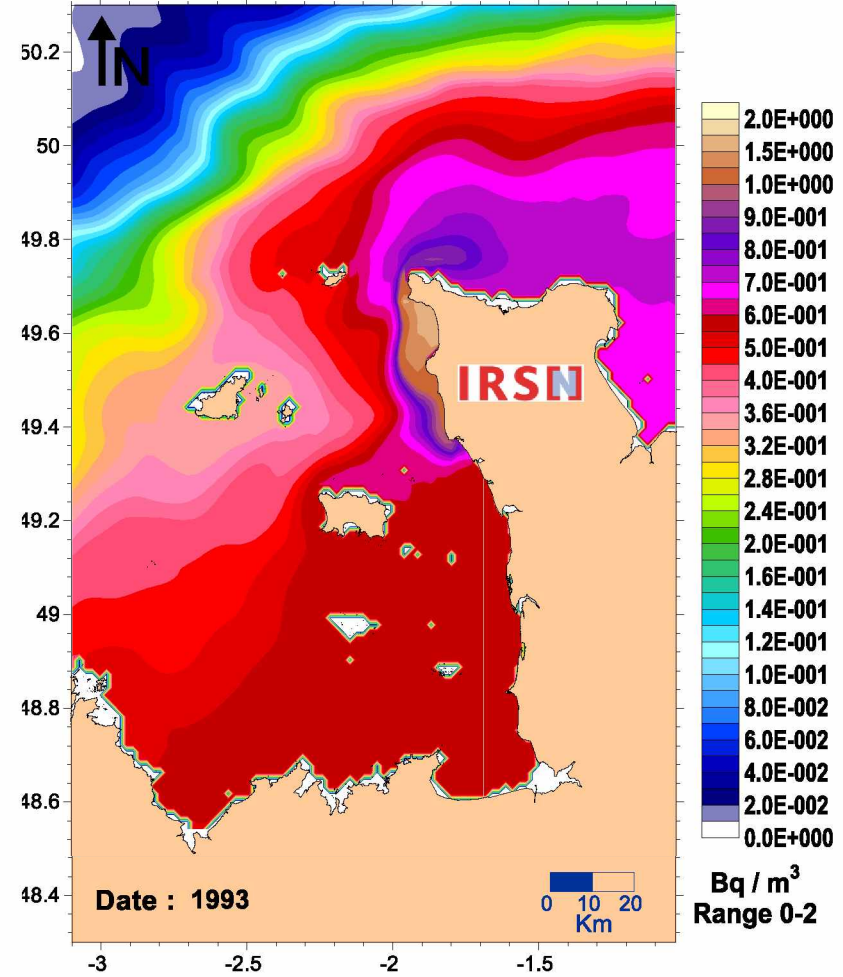
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Hydrodynamic modelling (MARS 2D)

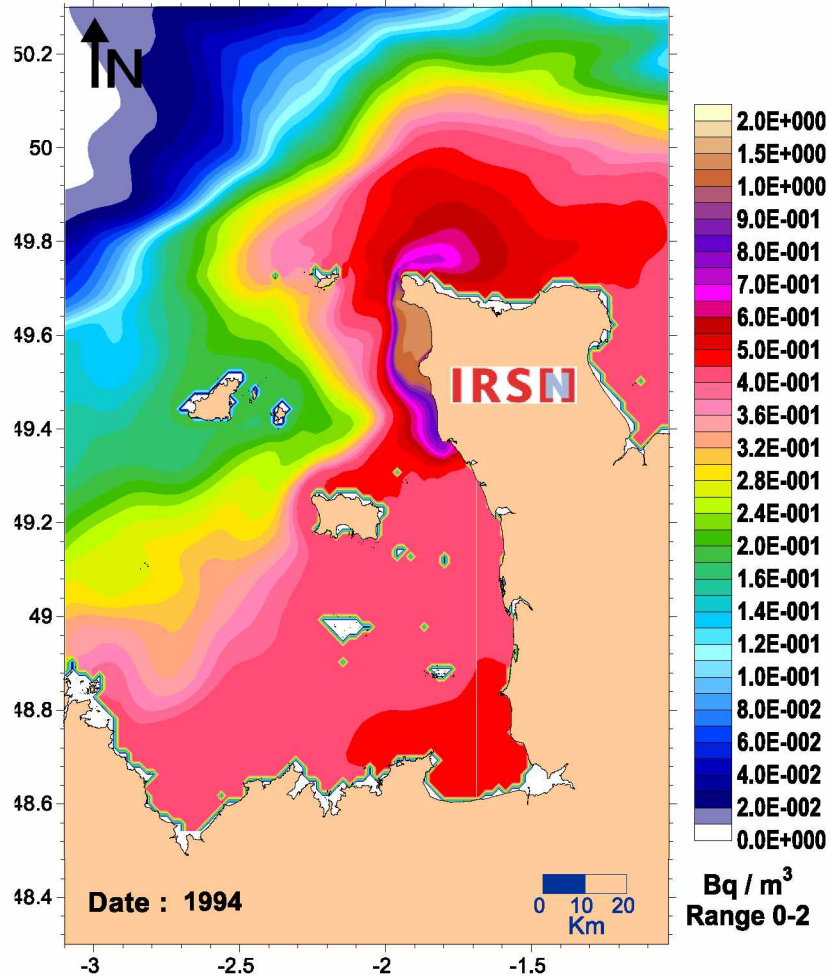
Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )



Hydrodynamic modelling (MARS 2D)

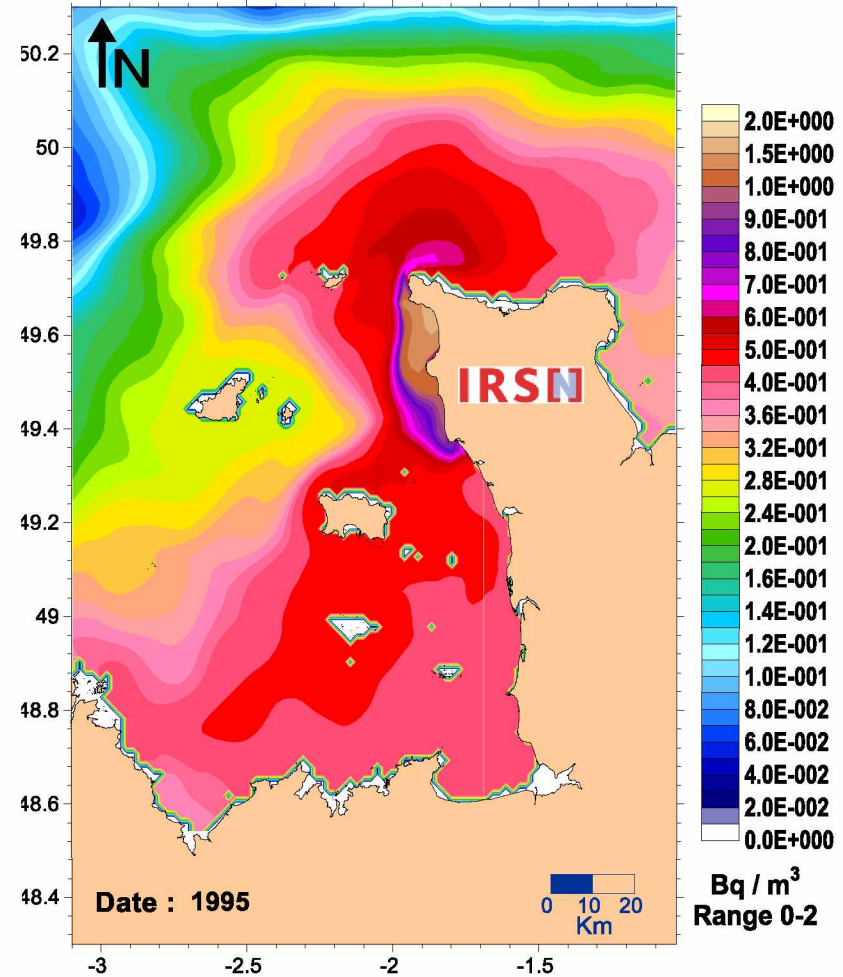
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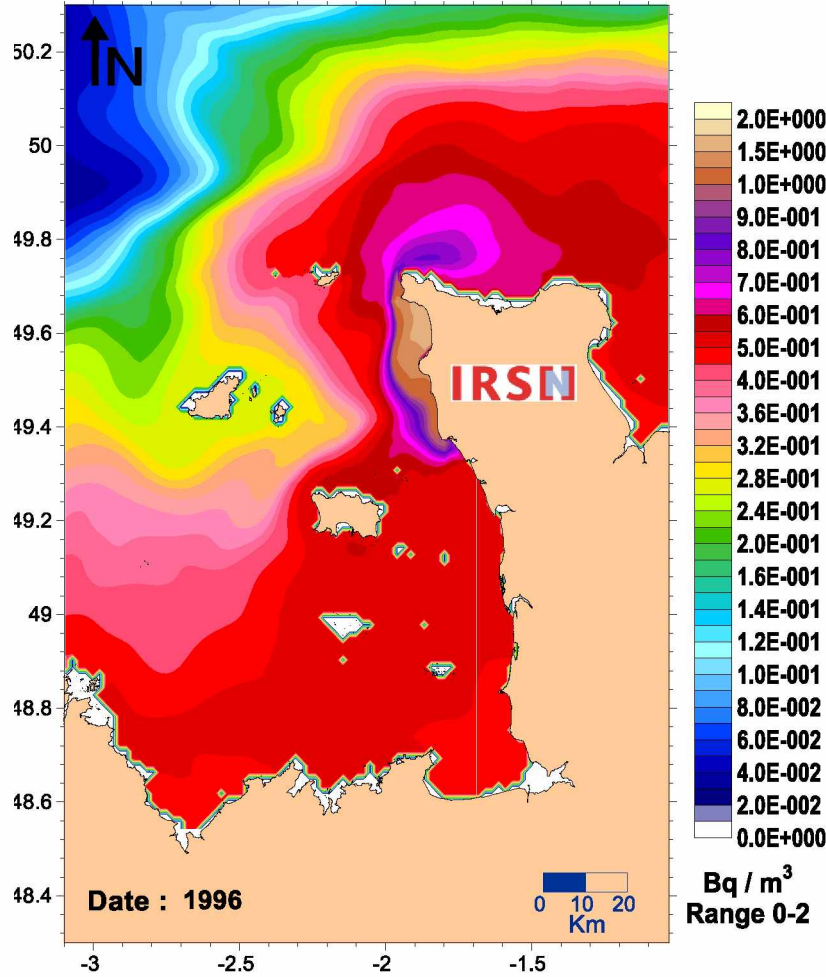


Hydrodynamic modelling (MARS 2D)

# Annual means

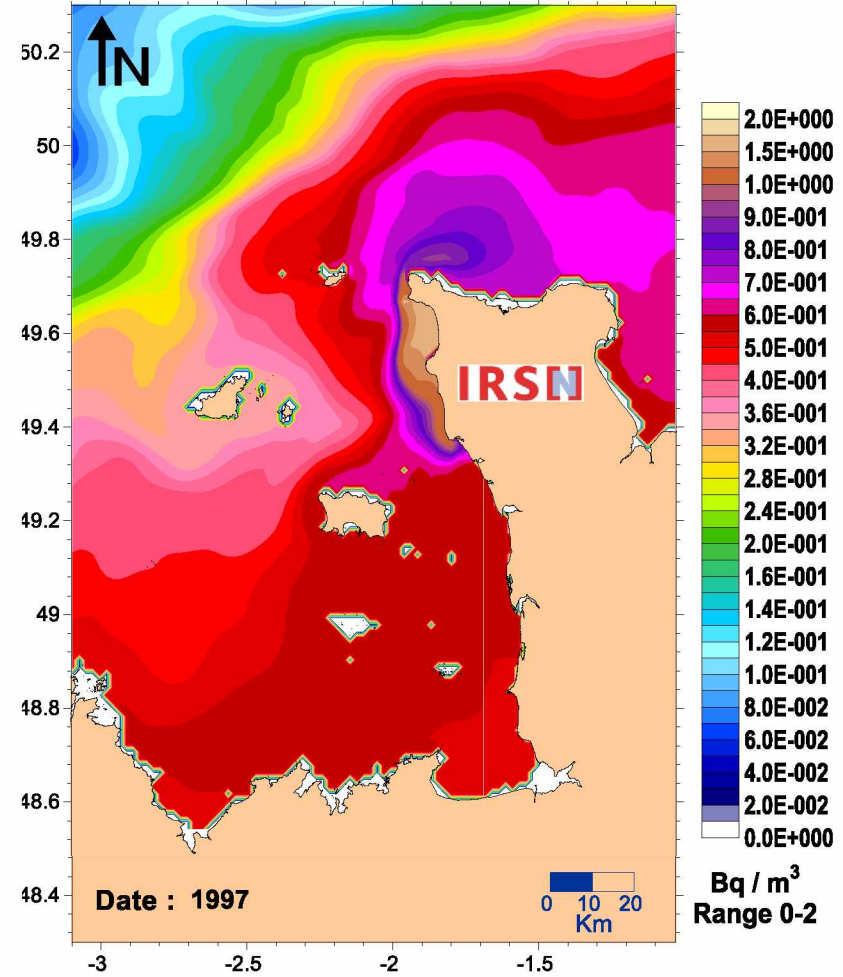
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Hydrodynamic modelling (MARS 2D)

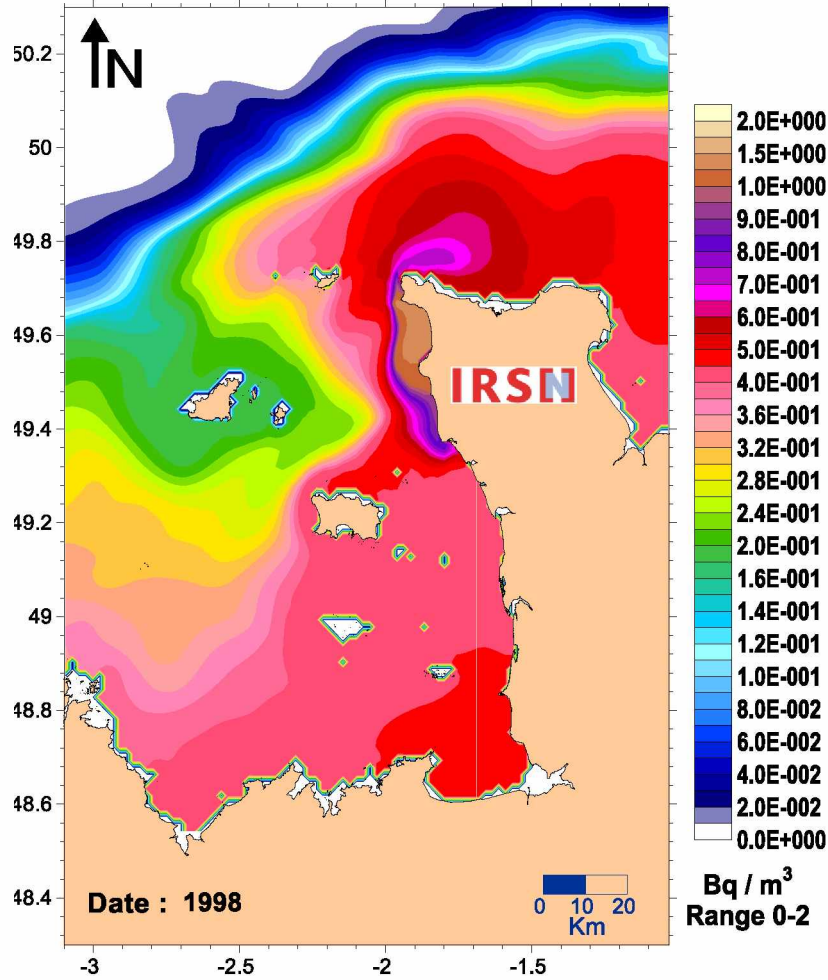
Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )



Hydrodynamic modelling (MARS 2D)

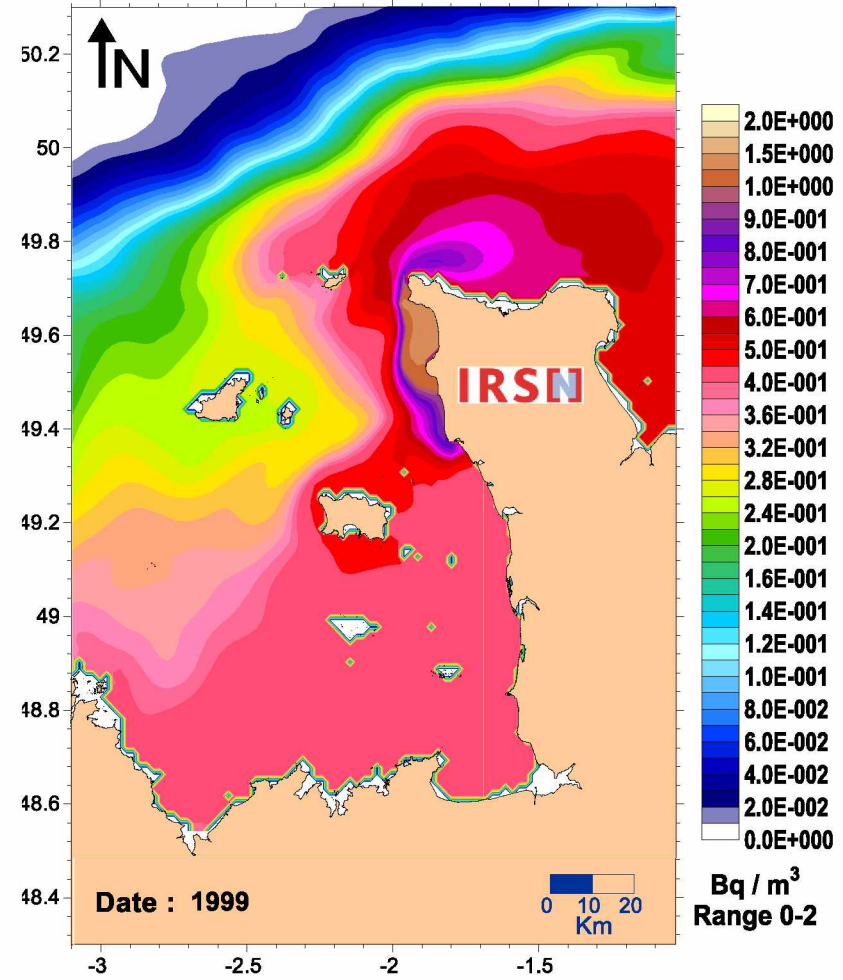
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Hydrodynamic modelling (MARS 2D)

Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )

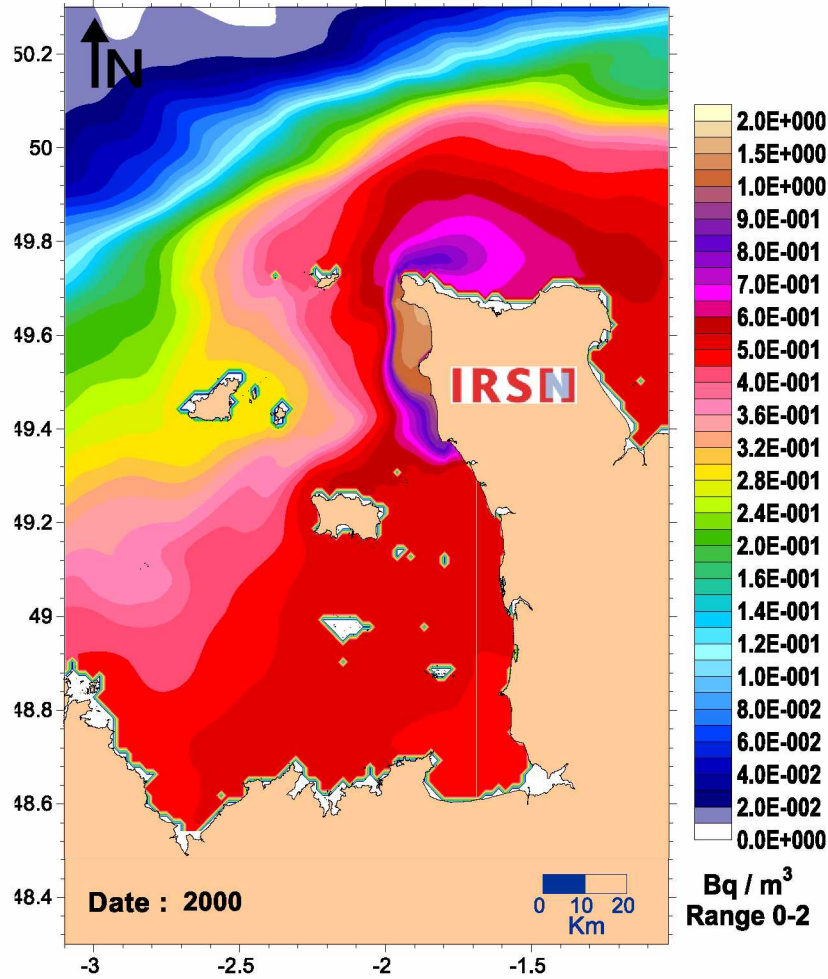


Hydrodynamic modelling (MARS 2D)



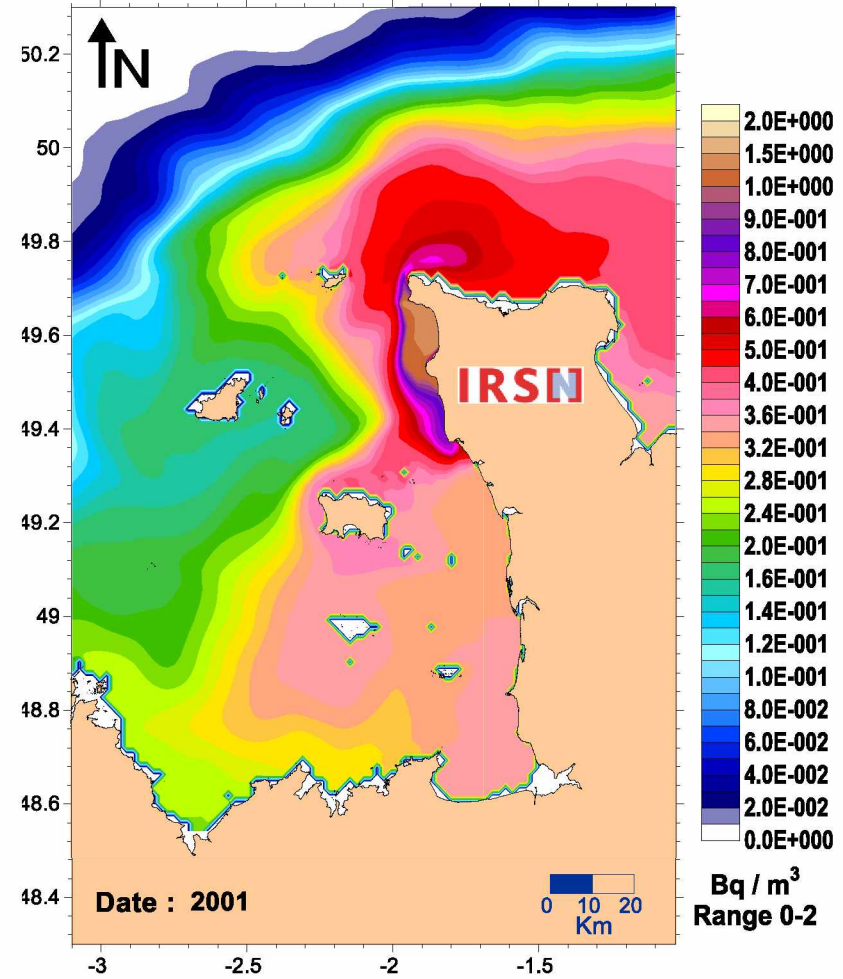
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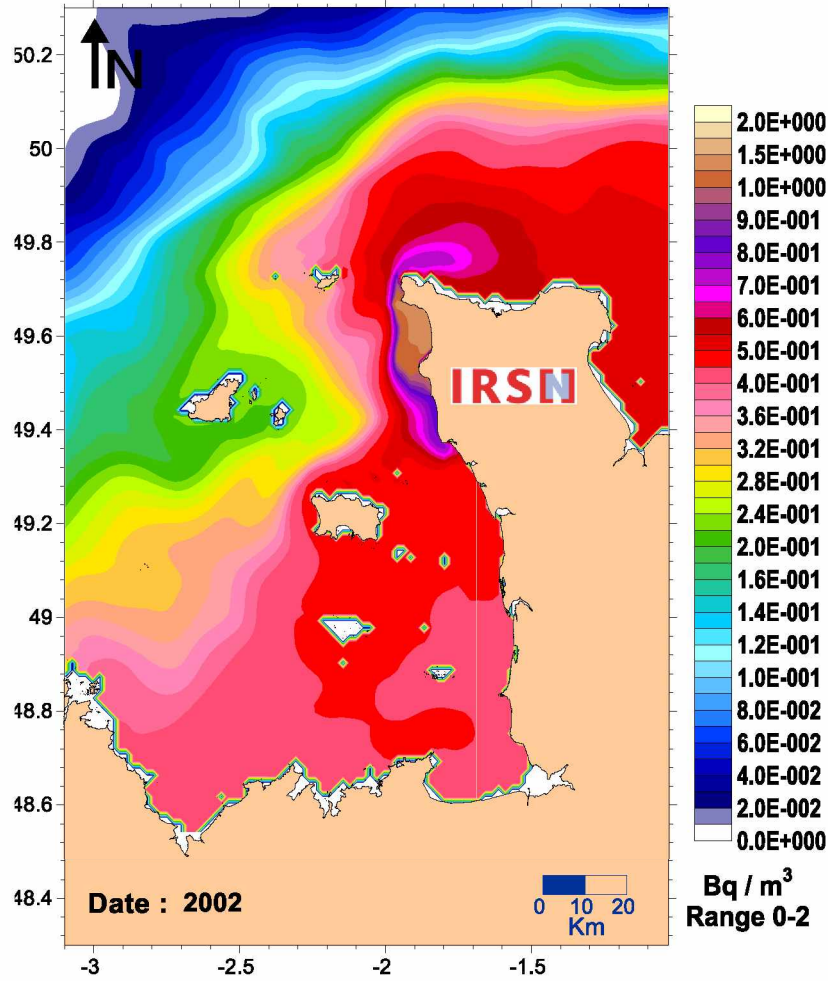


Hydrodynamic modelling (MARS 2D)

# Annual means

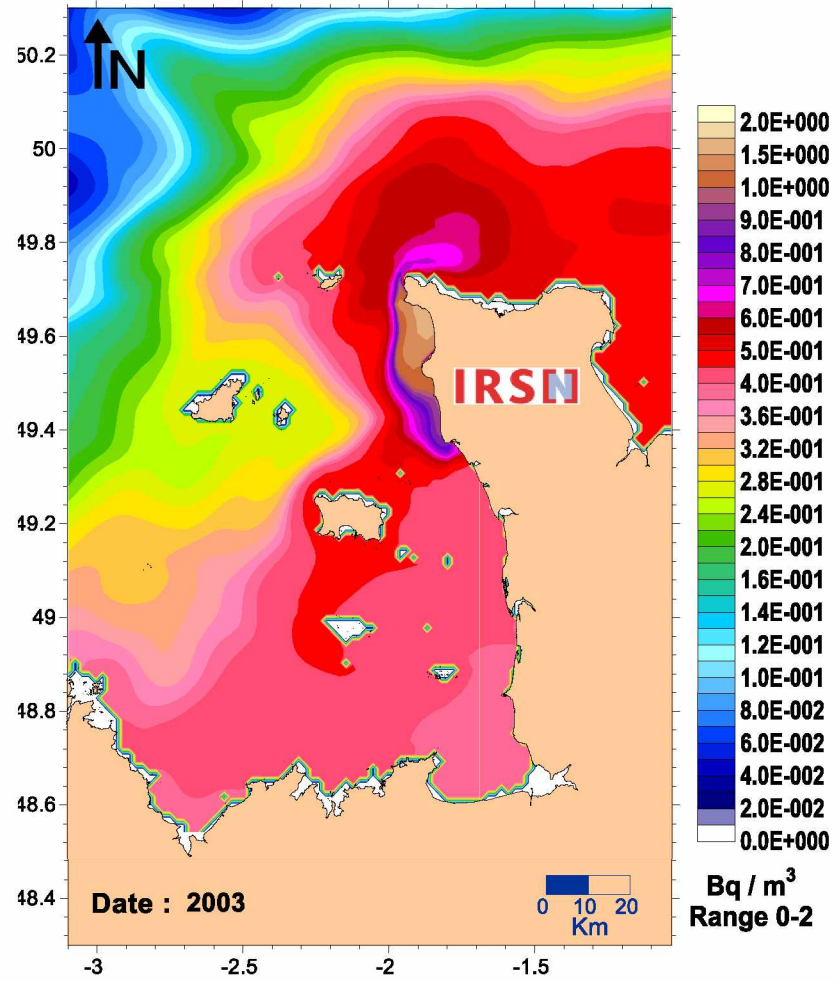
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Hydrodynamic modelling (MARS 2D)

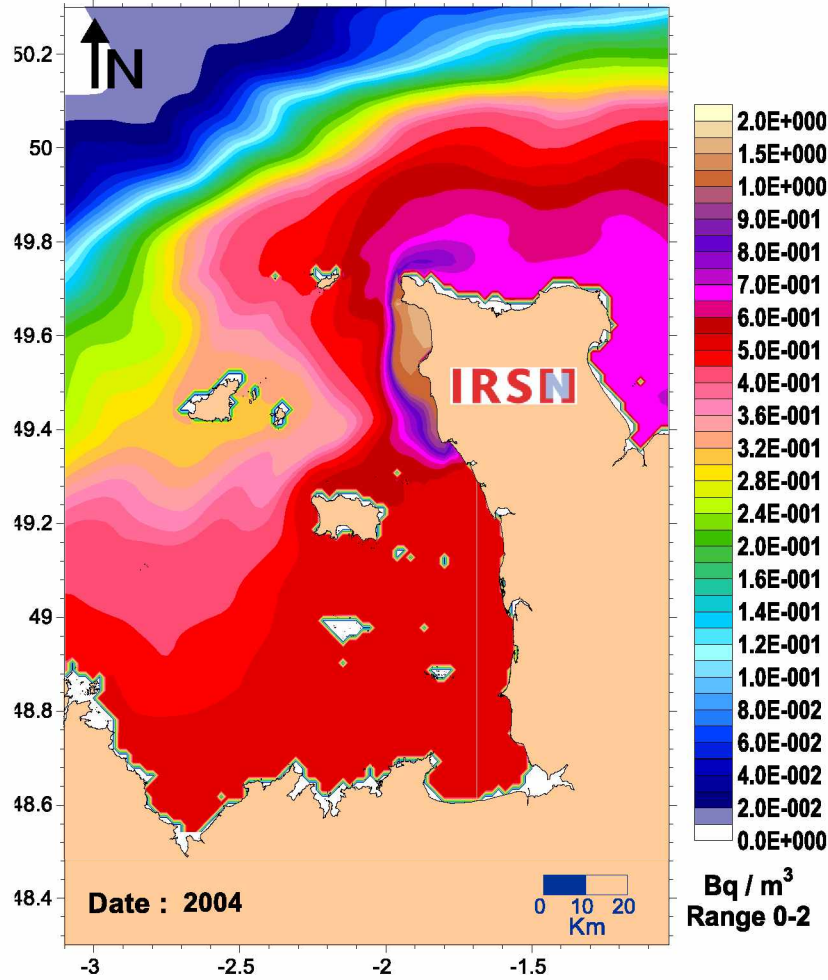
Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )



Hydrodynamic modelling (MARS 2D)

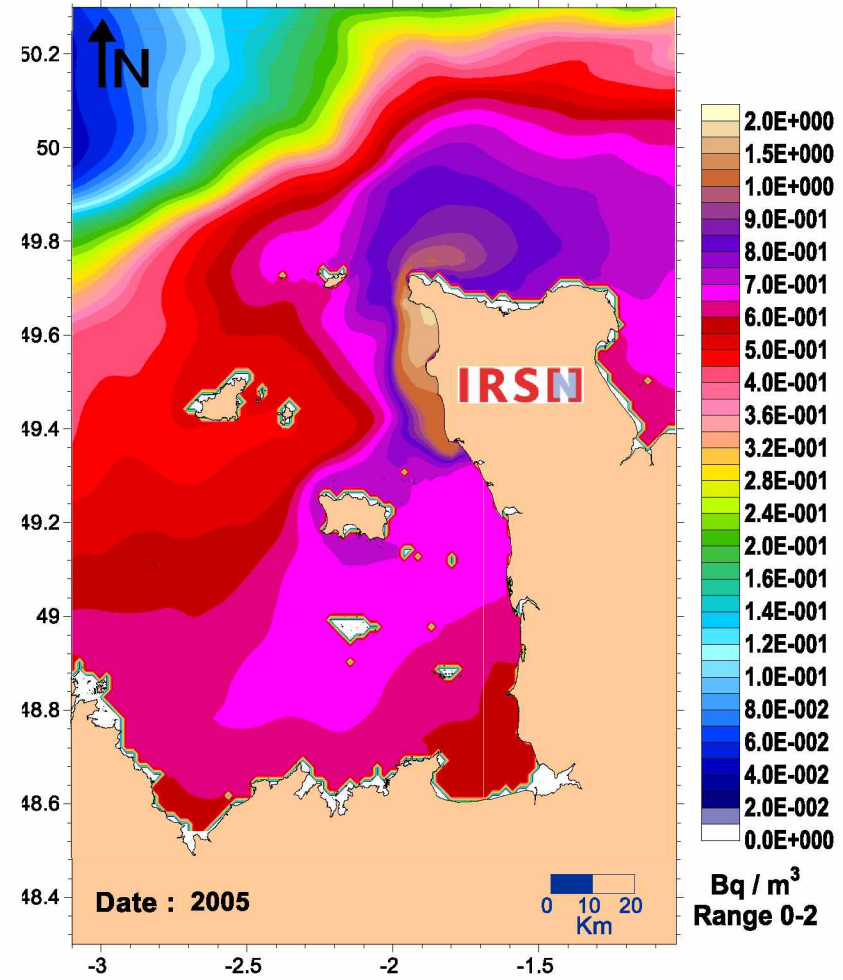
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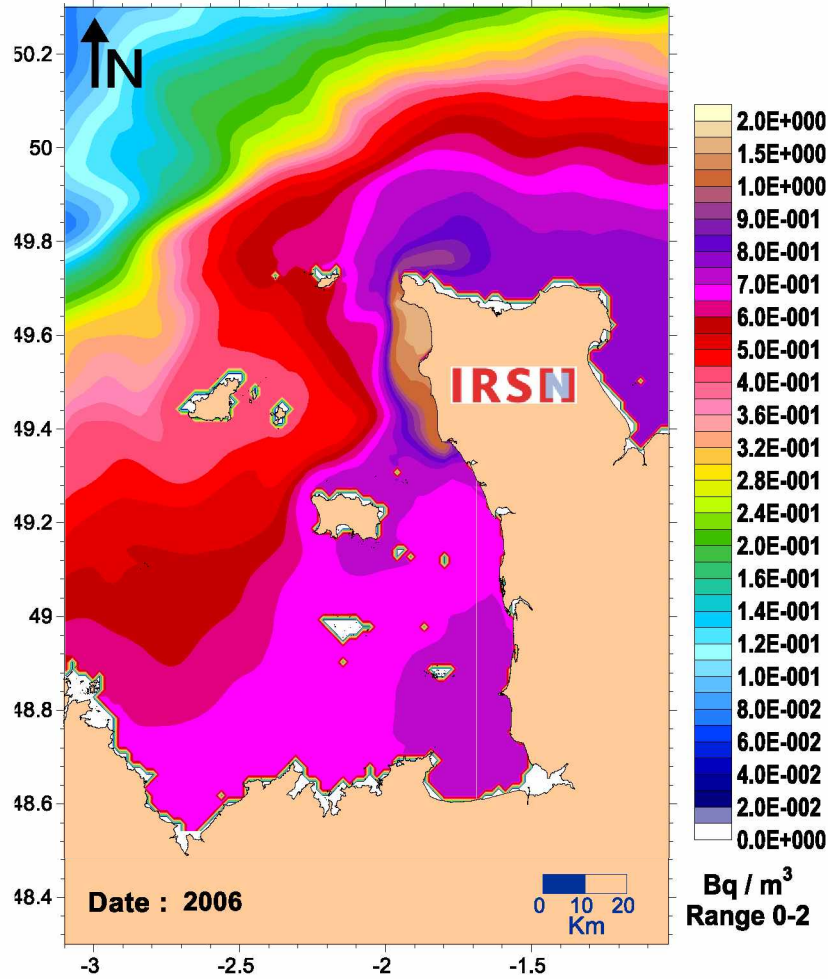


Hydrodynamic modelling (MARS 2D)

# Annual means

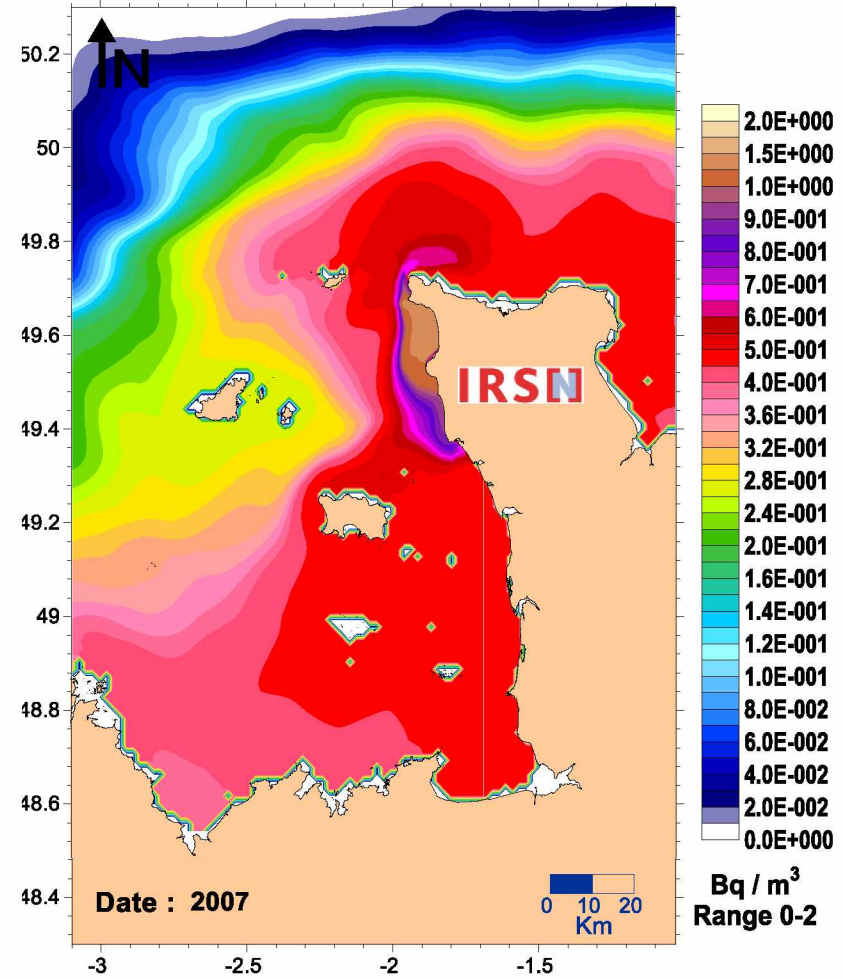
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Hydrodynamic modelling (MARS 2D)

Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )

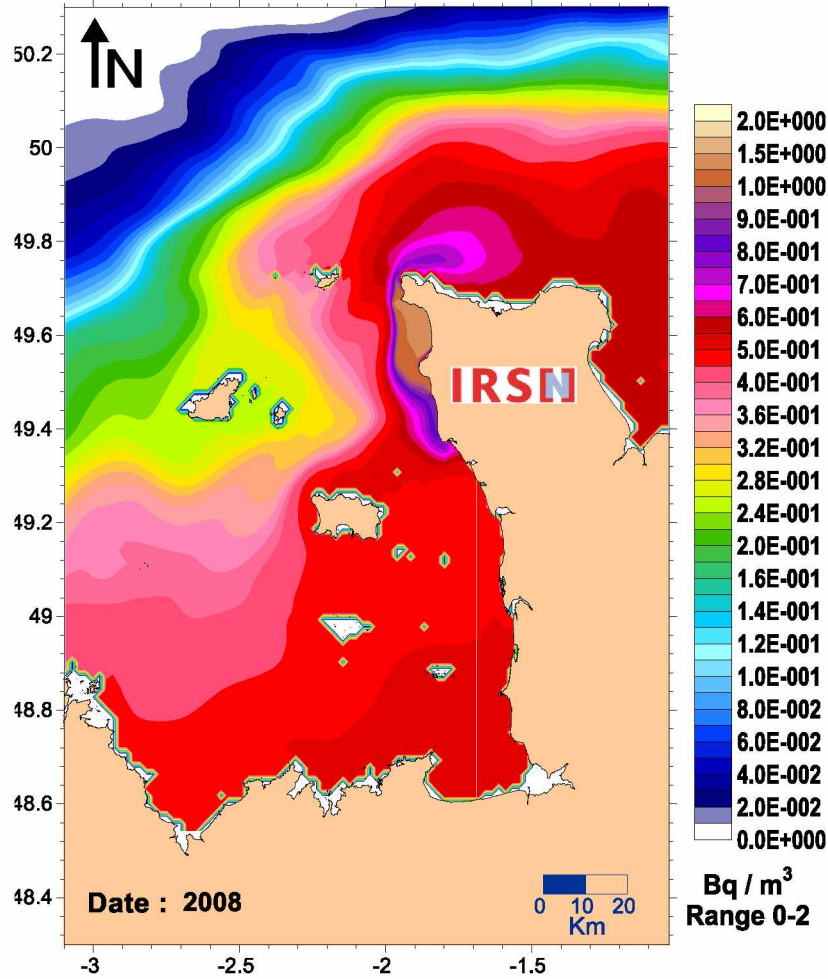


Hydrodynamic modelling (MARS 2D)

# Annual means

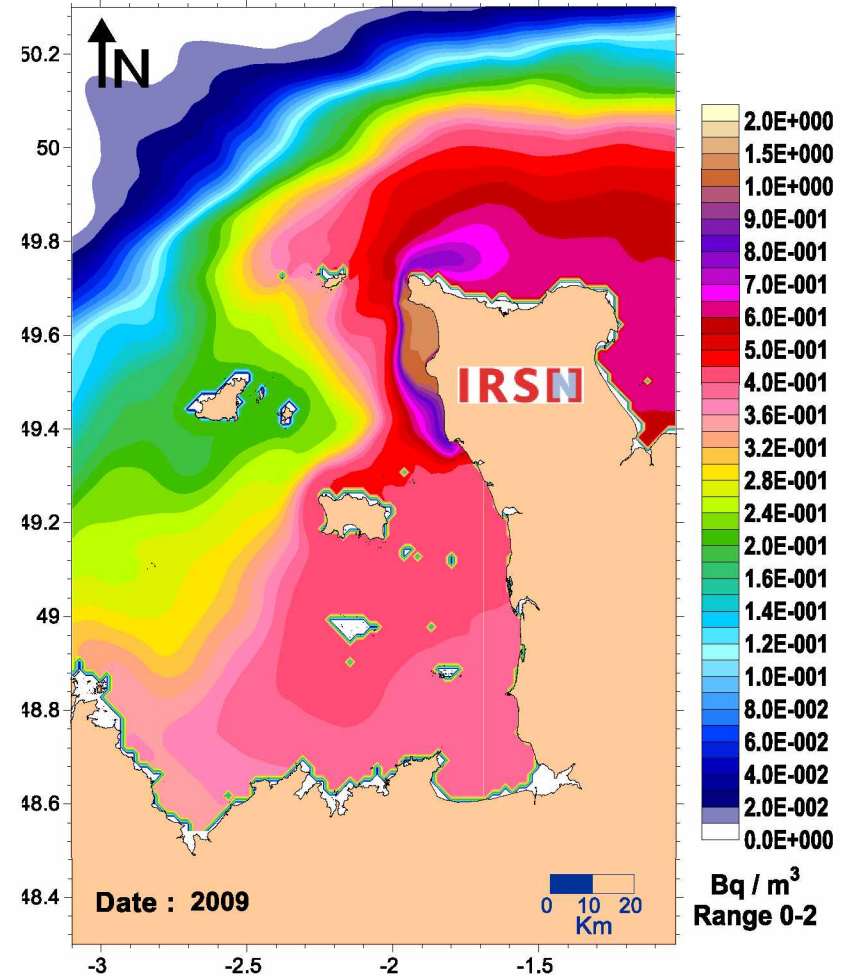
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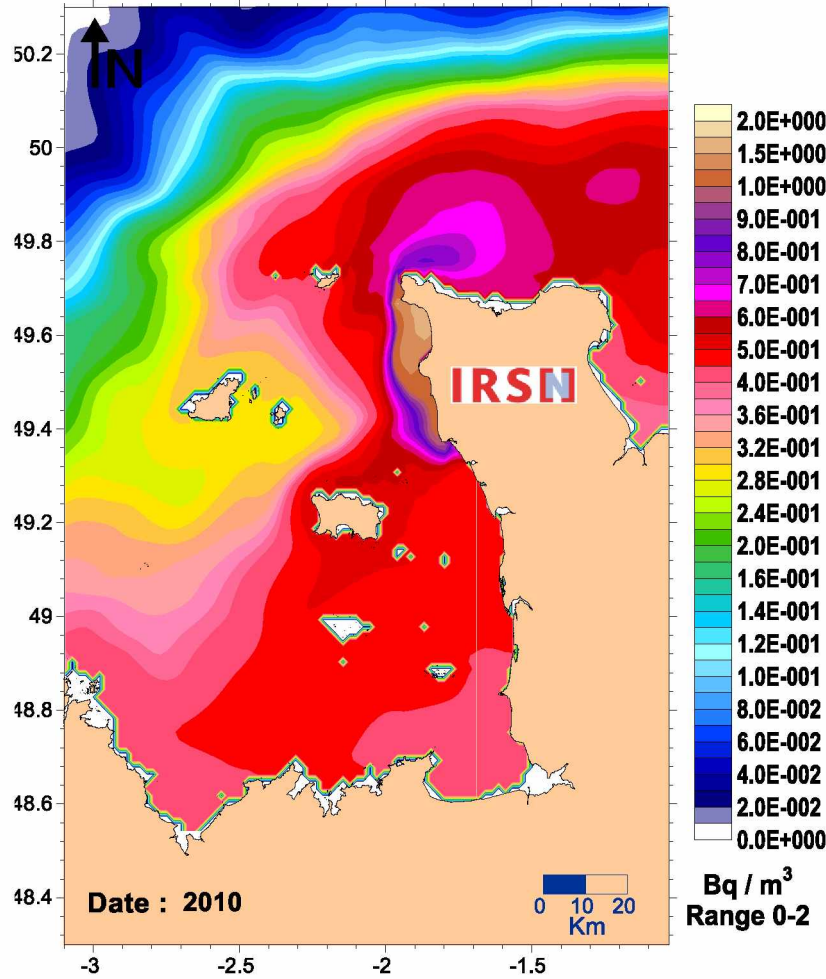
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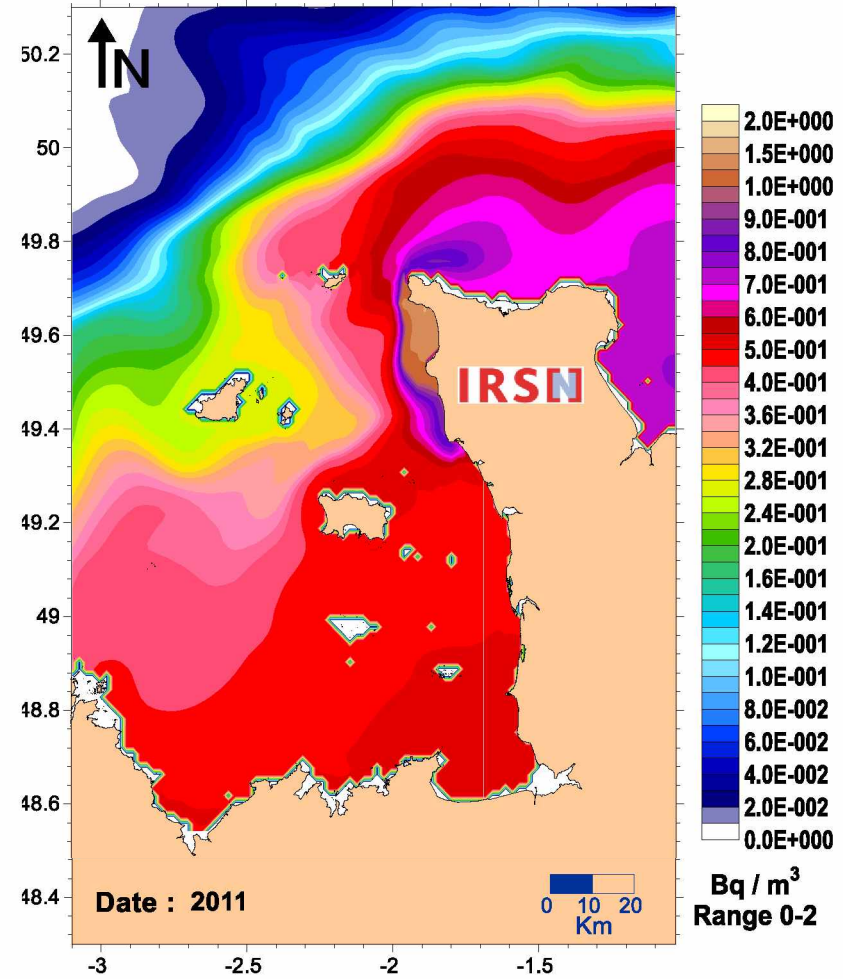
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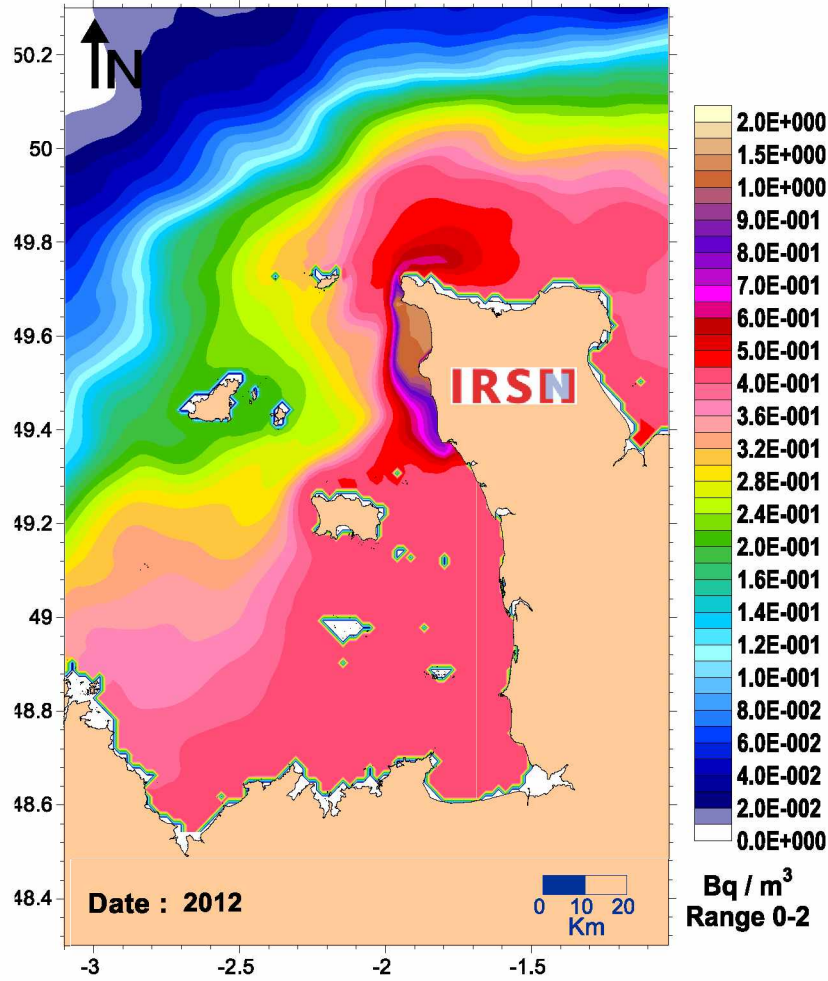


Hydrodynamic modelling (MARS 2D)

# Annual means

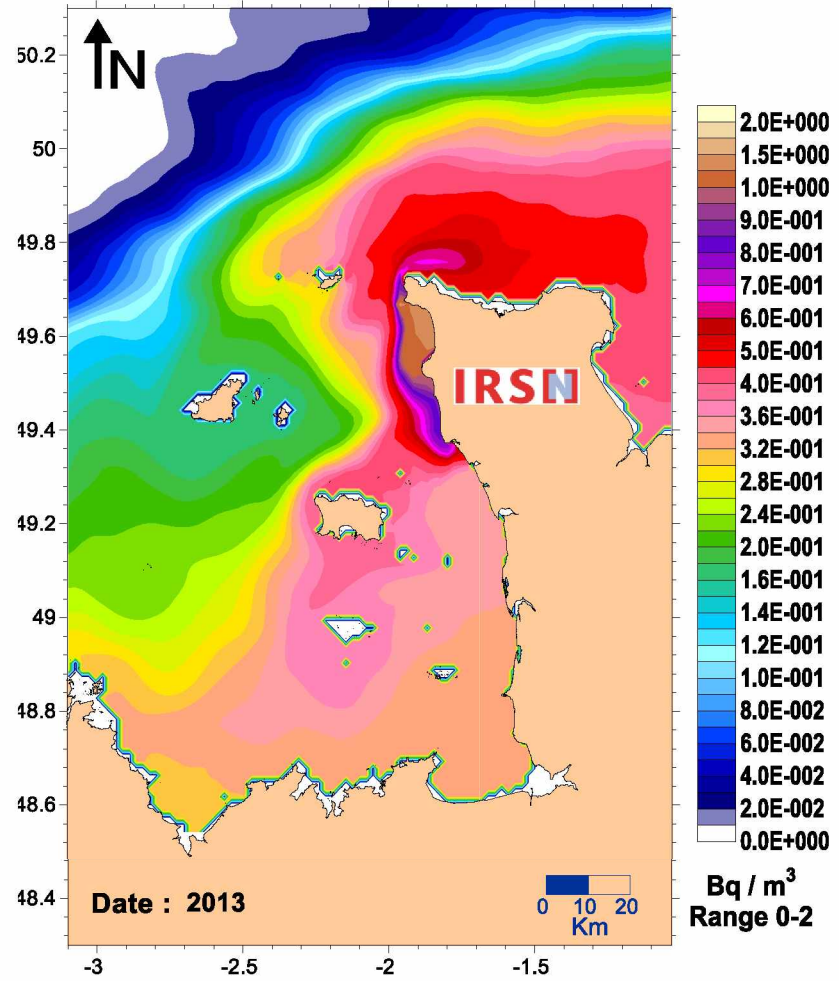
Dispersion modelling of a Constant discharge ( $1 \text{ TBq.y}^{-1}$ ), under average wind ( $\text{dir} = \text{SW}$ ,  $\text{speed} = 8 \text{ m.s}^{-1}$ ) and tide (coeff. 70) conditions.

Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )



Hydrodynamic modelling (MARS 2D)

Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )

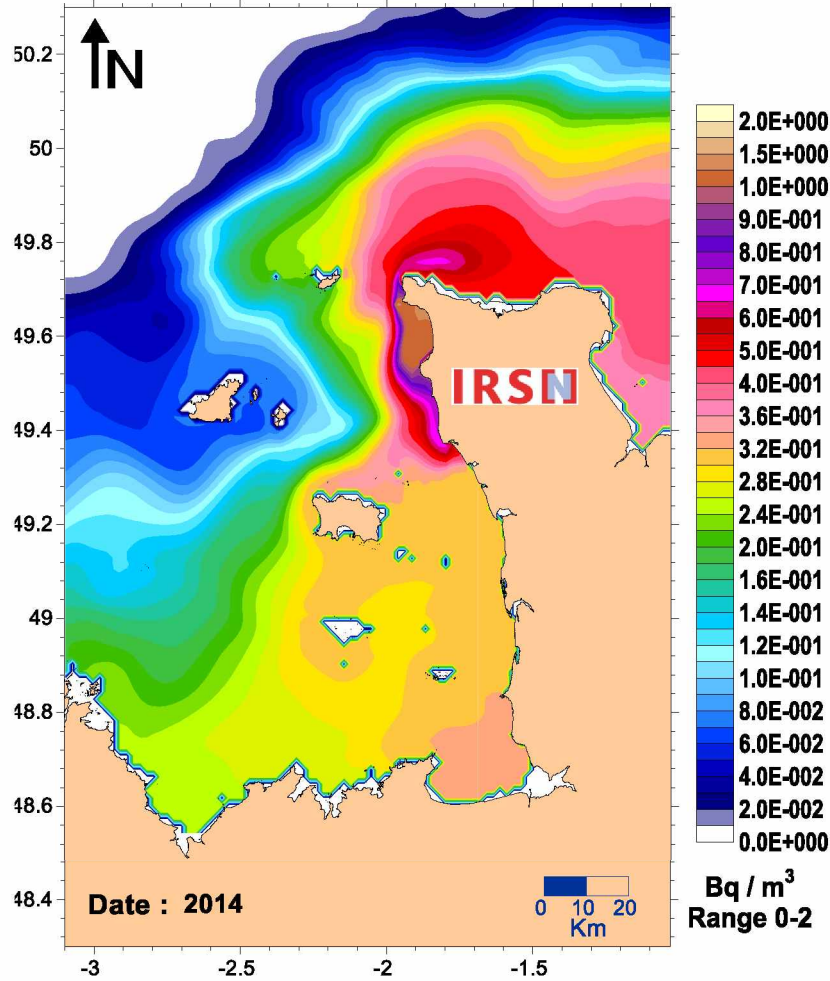


Hydrodynamic modelling (MARS 2D)

# Annual means

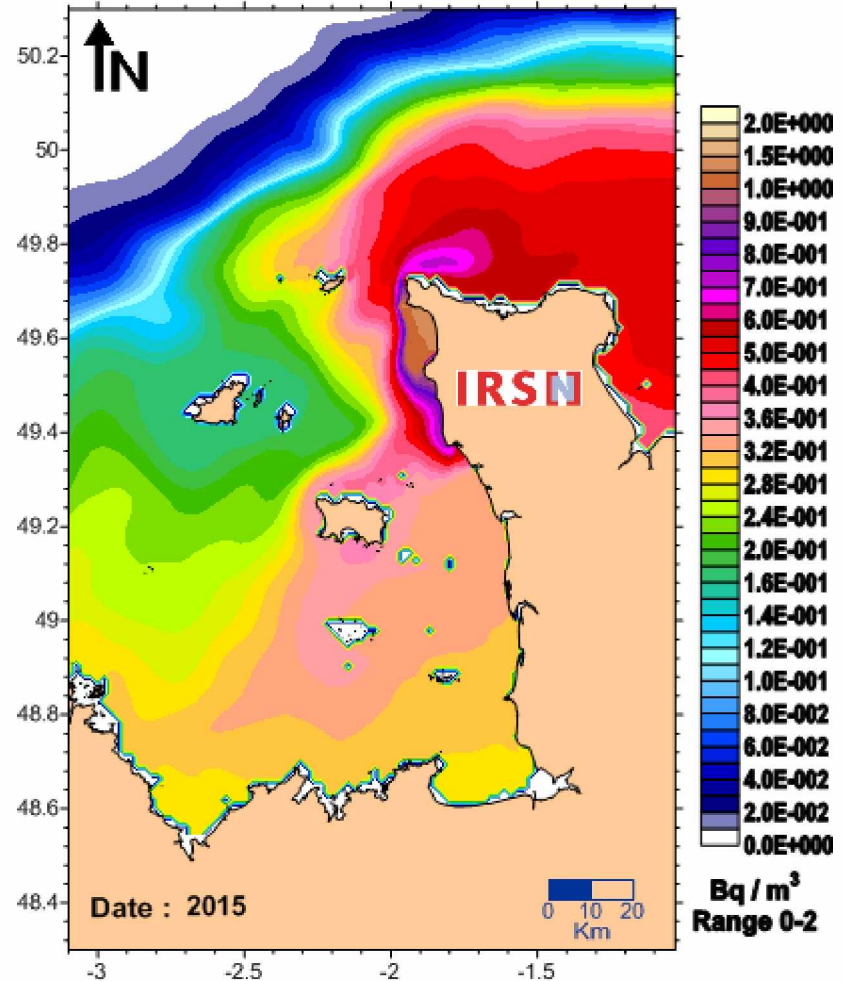
Dispersion modelling of a Constant discharge ( $1 \text{ TBq.y}^{-1}$ ), under average wind ( $\text{dir} = \text{SW}$ ,  $\text{speed} = 8 \text{ m.s}^{-1}$ ) and tide (coeff. 70) conditions.

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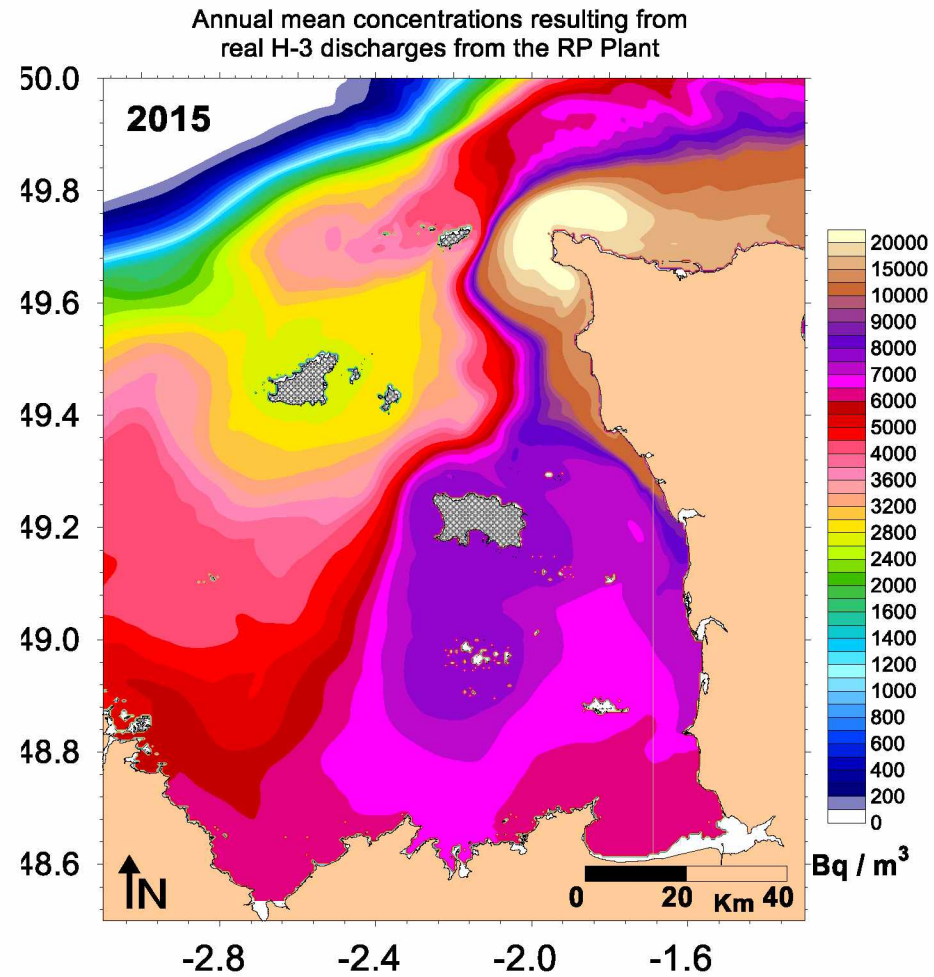
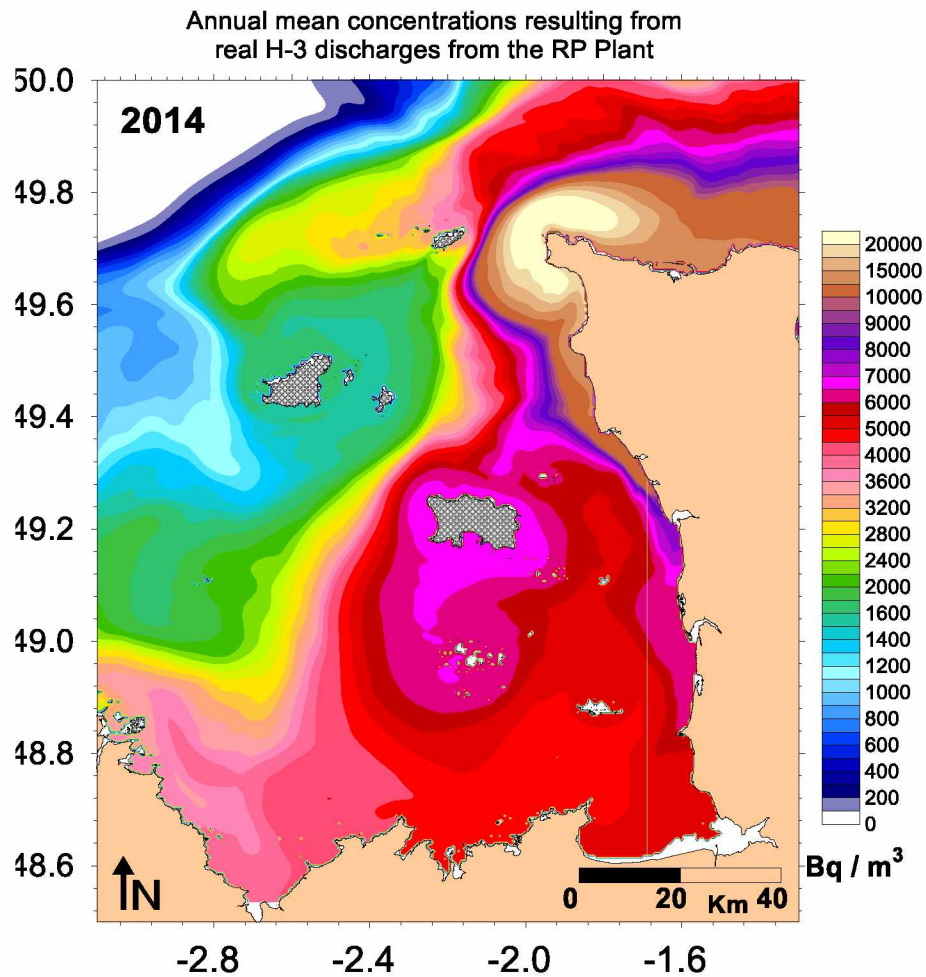
Hydrodynamic modelling (MARS 2D)

Annual mean concentrations resulting from constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ )



Hydrodynamic modelling (MARS 2D)





### Figure Caption

Figure 1: Annual controlled amounts of liquid radioactive discharges from the reprocessing plant of ORANO La Hague (line gaps: missing data. The dashed vertical line splits the graph into pre- and post-GRNC, 1999 publication data).

Figure 2: Normand-Breton Gulf (NBG) map showing the locations of sampling (cross), the ORANO La Hague reprocessing plant (RP) and the Flamanville EDF nuclear power plant (NPP). The names of the Channel Islands are written in bold, underlined. Inset: Western English Channel and Brittany showing the Concarneau and Roscoff reference locations, remote from French nuclear industry radioactive discharges.

Figure 3: Change in radionuclide activity with time in *Fucus serratus* collected in Corblet Bay (Alderney) and Portelet Harbour (Guernsey) over two decades since 1998. The black diamonds correspond to radionuclide activity in Bq.Kg<sup>-1</sup> dry with error bars of 2 x Sigma uncertainty. For activities below the limits of detection (LoD), a dashed vertical line spans from zero to the values of the LoD. §: The LoD was occasionally higher due to a reduction in counting time, depending on measurement laboratory availability constraints.

Figure 4: Change with respect to distance from the RP outfall in gamma emitter radionuclide concentrations measured in *Fucus serratus* collected in the NBG area and in reference locations (see map in Figure 2) in spring 2014 (thin hatching) and spring 2015 (thick hatching) . The vertical bars correspond to radionuclide activity in Bq.Kg<sup>-1</sup> dry and the error bars to 2 x Sigma uncertainty. For activities below the limits of detection (LoD), the error bars span from zero to the values of the LoD. The vertical dashed line splits the graphs between data from the NBG and from the reference locations in Brittany. The horizontal thick bars underline the British Channel Islands.

Figure 5: Changes over time in gamma emitter radionuclide concentrations measured monthly in *Fucus serratus* and *Patella sp* (soft parts) collected at Goury (6 Km from the RP outfall) from September 2013 to June 2016. The same symbols are used as in Figure 3.

Figure 6: Changes with respect to the distance from the RP outfall in alpha and beta emitter radionuclide concentrations measured in *Fucus serratus* collected in the NBG area and in reference locations (see map in Figure 2) in spring 2014 and spring 2015. The same symbols are used as in Figure 4.

Figure 7: General water mass circulation in the NBG resulting from Lagrangian residual currents for a constant wind direction (SW=231°) and speed (8 m/s) and a constant medium tide coefficient (70). The shades of grey show the current speeds (m.s<sup>-1</sup>) and the black arrows show the current directions and speeds. The star symbol shows the outfall of the radioactive releases from the RP.

Figure 8: 1985 and 2015 mean annual seawater concentrations resulting from a theoretical constant discharge of  $1 \text{ TBq.y}^{-1}$  ( $31\,709 \text{ Bq.s}^{-1}$ ) calculated by MARS2D with real wind and tide conditions. The numbers on the 1985 map (left) are the estimated average turnover values (d) of soluble radionuclides released from the RP outfall.

Figure 9: 2014 and 2015 mean annual H-3 concentration in seawater ( $\text{Bq.m}^{-3}$ ) calculated by the MARS2D model with real discharges from the ORANO La Hague RP, and real wind and tide conditions.

Figure 10: Comparison of the dispersion patterns of radionuclides in the NBG. A: Dilution factors based on 2014 and 2015 annual mean H-3 levels in seawater calculated by the MARS2D model and normalized to H-3 discharges from the RP. B, C, D: Concentrations of radionuclides in seaweed measured in the NBG in 2014 and 2015 divided by the Concentration Factors (Table 1), normalized to the 12-month discharges preceding sampling. Average concentrations from Concarneau were subtracted as background values from NBG data for C-14.

Figure 11: Concentrations of radionuclides in seaweed measured in the NBG in 2014 and 2015 divided by the Concentration Factors (Table 1), normalized to the 12-month discharges preceding sampling. Average concentrations from Concarneau were subtracted as background values from all data for Cs-137 (A), Pu-239,240 (C) and Am-241 (D).

Figure 12: Monthly time-series measurement data for *Fucus serratus* from Goury divided by the Concentration Factors (Table 1) and normalized to the 12-month discharges preceding the sampling date.

**Table Caption**

Table 1: parameters used to normalize the biota data.  
The CF values were taken from IAEA TRS 422 (2004), except for C-14 (Fiévet *et al.*, 2006).

Table 2. Alpha and beta emitter radionuclide concentrations in *Fucus serratus* collected on shore in the Channel Islands between 1999 and 2017. Activities are expressed as  $\pm$  uncertainty (2 x sigma), and in  $\text{Bq.Kg}^{-1}$  C for C-14, in  $\text{Bq.L}^{-1}$  for H-3 and in  $\text{Bq.Kg}^{-1}$  dry for alpha emitters. In the 2017 samples, OBT was measured in parallel with HTO and the OBT values in  $\text{Bq.L}^{-1}$  of combustion water are given in brackets next to the HTO values.