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Particulate pollution in Korhogo and Abidjan (Cote d'Ivoire) during the dry season.

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10 11

12 Abstract

Particulate pollution in West African cities seriously impacts public human health. Assessing 13 14 the levels of pollution to which populations are exposed is problematic, as to date very few 15 countries in Africa have an air quality monitoring network in place. However, given the specific anthropogenic sources present in West African countries and the increase in their projected 16 emissions in the coming years if no regulations are put in place, solutions must be found. This study 17 18 evaluates with a simple existing methodology particulate air pollution in two West African cities 19 (Korhogo and Abidjan in Cote d'Ivoire) which have different population practices and local 20 meteorology.

From two measurement campaigns at about ten sites during the dry season and using the inverse 21 distance-weighted interpolation method, maps of spatial variation in PM10 and PM2.5 22 concentrations are obtained at the scale of each of these cities, which provide a snapshot of the 23 24 concentration levels to which populations are actually exposed. The results of this study showed 25 that populations in low- and middle-income areas are exposed to higher concentrations than those 26 in high-income areas. The analysis of the relative contribution of fine particles to coarse particles shows that PM_{2.5} contributes to PM₁₀ at less than 40% in Korhogo but more than 50% in Abidjan. 27 Daily average concentrations of PM_{2.5} are all above WHO standards, and are 2 to 8 times higher in 28 29 Korhogo, while in Abidjan they are up to 4 times higher. The daily profiles of the mean hourly 30 concentrations of the sites are similar across all sites in Korhogo, while in Abidjan, that of some 31 sites show some differences. 32 This study, by the methodology used, offers an opportunity for different research teams to assess 33 at lower cost the population's exposure levels to urban particulate pollution.

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Keywords: Urban air quality; Aerosols; Africa cities; PASMU project.

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36 INTRODUCTION

37 Air pollution is a global public health problem and its extent in the cities of many developing 38 countries is still poorly known (Kirenga et al., 2015). Among all pollutants, particulate pollution is 39 highly harmful to health, especially in low- and middle-income countries (Kalisa et al., 2019). 40 Recent studies have shown alarming levels of particulate pollutants in West African cities (Dieme 41 et al., 2012 and Djossou et al., 2018), due to various sources of pollution that are generally 42 uncontrolled and relatively close to living areas Val et al., (2013) and Adon, (2019) highlighted the 43 inflammatory nature of particles rich in organic carbon (OC) and small in size emitted by urban 44 west African sources. Also, a few studies (Cohen et al., 2005 and Liousse et al., 2019) have shown the importance of premature death estimations for cardiovascular, lung cancer and respiratory 45 46 diseases causes, due to anthropogenic fine particles in West Africa.

47 Anthropogenic sources are harmful in developing countries because of the absence or 48 ineffectiveness of regulations on pollutant emissions. The typical anthropogenic sources in sub-49 Saharan Africa are domestic fires (biomass and gas for cooking), road traffic (combustion emissions and dust resuspension), waste combustion, industrial activities (petrochemicals, 50 51 agriculture, construction. More precisely, air quality in Cote d'Ivoire and in West Africa, is 52 impacted by emissions from a very old vehicle fleet, on average 20 years according to the Ministry of Transports, (2017), using poor quality fuel and from the use of traditional stoves with wood 53 54 and/or charcoal incomplete consumption. Furthermore, open burning of waste in urban areas of 55 Africa is a very common phenomenon due to the ineffectiveness of municipal waste collection and 56 management policies (Keita et al., 2020). In addition to these local sources, there are important 57 imported sources of aerosols: (1) desert dust and natural savannah fire sources from prevailing 58 Harmattan wind (north wind) as well as (2) marine aerosols brought by monsoon flow (southwest 59 wind).

60 Unlike developed countries, most cities in the developing countries do not have the capacity to 61 monitor urban air concentrations in order to effectively solve the problem of air pollution. In West 62 Africa, Dakar (in Senegal), is the only city with an air quality monitoring network. Considering the 63 anthropogenic source diversity and their strength which is expected to increase if no regulation is 64 taken (Liousse *et al.*, 2014), it is necessary to expand pollution studies in African cities.

65 Studies carried out in West Africa have shown high concentrations of PM_{2.5} and PM₁₀ in several cities: Accra in Ghana (Dionisio et al., 2010 and Arku et al., 2008); Dakar in Senegal (Dieme et 66 67 al., 2012 and Val et al., 2013) and Bamako in Mali (Val et al., 2013 and Antonel and Chowdhury, 2014). In the European program DACCIWA (Dynamic-Aerosol-Chemistry-Cloud Interactions in 68 69 West Africa), Djossou et al., (2018) observed average concentrations (February 2015 to March 70 2016) in Abidjan (Cote d'Ivoire) and Cotonou (Benin), well above WHO standards. Similarly, in 71 the POLCA program (POLlution of African Capitals), Doumbia et al., (2012) measured PM_{2.5} 72 concentrations at an urban site in Dakar from June 2008 to May 2009, which were above the WHO 73 standard. These various studies highlight overexposure of populations to atmospheric particles 74 whose economic impacts were assessed by Stern and van Dijk, (2017), and mortality by Cohen et 75 al., (2005) and in Global Burden of Disease Study 2010 by Murray et al., (2012). Also, within the 76 framework of DACCIWA, Keita et al., (2020) have calculated emission factors (EF) for particulate 77 pollutants based on sources and considering the specificities of African regions in order to improve 78 emission inventories for these pollutants. These results showed very high EF values, highlighting 79 the importance of particulate pollutant emissions from sources such as domestic fires and traffic, 80 which are very important sources of pollution in West Africa. These different results are a challenge 81 for us and we must act.

82 The lack of air pollution monitoring networks, even in the most urbanized cities, due to the high83 cost of acquiring and maintaining air quality measurement equipment can be offset by the

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84 installation of low- and medium-cost sensors offering simple and inexpensive alternatives. 85 Although accessible, these sensors cannot be deployed always and everywhere. To overcome this 86 problem, an experimental existing methodology based on the use of low-cost sensors during 87 intensive measurement campaigns and a geostatistical treatment tool is used to assess particulate air pollution in two cities of Cote d'Ivoire. Specifically, this study quantifies the level of PM₁₀ and 88 89 PM_{2.5} concentrations and their spatial and temporal distribution in these cities of different 90 geography and morphology (Korhogo and Abidjan). This work is carried out within the framework of the project Air Pollution and Health in Urban Areas, Cote d'Ivoire (PASMU). 91

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93 METHODOLOGY

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95 **Description of the experimental sites**

In this study, two measurement campaigns were conducted successively in Korhogo (January 96 3rd to January 16th) and in Abidjan (January 20th to February 10th, 2019), whose economic activities, 97 98 transport specificities and types of roads, cooking fuels, population practices and local meteorology 99 differ. As shown in Fig. 1, the population densities of these two cities are also very different, with 100 a population of Abidjan 10 times that of Korhogo. It can also be noticed that population densities 101 from south to center of Cote d'Ivoire are generally higher compared to areas further north. The 102 population density of Korhogo remains high for a northern region due to agricultural industries. In 103 each city, the measurement sites are chosen to allow representative mapping of the variability of 104 pollution including different districts. Generally, measurement sites can be of 3 types: (1) Local 105 stations, generally used to control classified establishments, are located in proximity of the polluted 106 source to be monitored and therefore under the direct influence of this pollution source; (2) Urban 107 background stations, used to monitor urban pollution, are located in an urban environment but 108 under the indirect influence of pollution sources, and (3) Background stations, are located in a peri-109 urban environment. It is important to note that local stations are therefore not included in this study 110 as they only provide very local information (Jeannée et al., 2006). In our study, urban background 111 stations are mainly influenced by major west-African urban sources, i.e. the domestic fires, the 112 waste burning and the road traffic (from combustion or non-combustion). In addition, dust 113 resuspension of particulate matter is also important because of road conditions and urban 114 developments. In this study, 14 non-permanent sites in Korhogo and 20 in Abidjan were sampled 115 for 3 days (2 working days and 1 weekend day). The 2 cities and their sampling sites are described 15 116 below.

117 Korhogo sampling sites

Korhogo is the third largest city in Cote d'Ivoire in terms of population. A large part of the city's 118 119 area is not built. Korhogo is a city still marked by agricultural activities with a few rare agricultural 120 product factories for processing and packaging. The principal sources of anthropogenic emissions 121 in Korhogo are traffic sources with intense activities of heavy vehicles, national bus line but also 122 motorcycle traffic, which is a particularity of Korhogo. Furthermore, in this city, only main roads 123 are paved and represent, 10% of the urban roads (AGEROUTE, 2020). This implies important road resuspension of aerosols which is also a major source of PM emissions. Another important 124 125 anthropogenic source is the domestic fire with the use of firewood for economic reasons or for 126 convenience as because these resources are still available in the surrounding villages. In addition 127 to the local anthropogenic sources present in Korhogo, the harmattan wind is a major contributor 128 of PM in the dry season with desert dust, biomass burning and aerosol transport. Fig. 2 shows the 129 geographical distribution of the 14 measurement sites as well as the boundaries of the city of 130 Korhogo. Also, the specificities of each sites are summarized in Table S1 (in Supplementary files), 131 where are mentioned site localizations as well as roads infrastructure and possible sources of

pollution around. Additional information like that geographic positions of Korhogo measurement sites is given in Table S2 (in Supplementary files). Sites K₂, K₁₀, K₁₁ and K₁₄ are surrounded by paved roads (i.e. under the influence of the traffic source) and far from other urban sources. The other sites are located in residential areas where wood is used extensively for cooking, with the exception of K₁ site which is in a peri-urban area but close to a major traffic road.

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138 Abidjan sampling sites

Abidjan is the economic capital regrouping the main industrial activities with a population of
4,707,404 inhabitants for an area of 2,119 km² according to the General Census of Population and
Housing (RGPH, 2014).

Abidjan is an Autonomous District with different levels of urbanization and includes 142 143 populations of different social classes. Its principal anthropic sources are also traffic, domestic fires 144 and waste burning sources with different practices depending on the communes. For instance, wood 145 is intensively used in the commune of Yopougon whereas the use of gas dominates in Cocody. 146 Less road dust particles compared to Korhogo are expected since more than 50% of roads are paved 147 according to MCLAU, (2015), (MCLAU for Ministère de la Construction, du Logement, de 148 l'Assainissement et de l'Urbanisme). However, this value also depends on the communes. In 149 Cocody, Le Plateau and Marcory, a higher percentage of paved roads are observed compared to 150 Abobo and Yopougon, where a large part of the city population lives.

Fig. 3 shows the geographical distribution of the measurement sites in Abidjan and boundaries of the sampled communes, which represent the study area of this city. In Supplementary files, Table S5 provides information on the 20 sites sampled in Abidjan. In this study, only the sub-prefecture of Bingerville (Abidjan peripheral municipality) is associated with the 10 municipalities of Abidjan, because Bingerville city is completely integrated into Abidjan. As shown in, in Supplementary files, in Table S5, measurement sites are located in neighborhoods inhabited by high-income populations
(A1, A2, A11, A20, A4, A6, A8, A14, A15), middle (A18) and low (A3, A5, A7, A9, A10, A12, A13, A16,
A17, A19). The latter sites are generally close to unpaved roads. In addition, Table S6 (in
Supplementary files), give latitude and longitude of Abidjan measurement sites.

160

161 Meteorological data and surrounding pollution

To assess the impacts of weather parameters on PM concentration levels measured at each site during the 3-day sampling period, temperature, wind intensity and direction are analyzed at local and regional scales. These meteorological parameters are obtained from the data of the synoptic stations managed by the Agency for the Safety of Air Navigation in Africa and Madagascar (ASECNA) at Korhogo Airport (station No. 65536) and at Felix Houphouet-Boigny Airport in Abidjan (station No. 65578).

Average daily temperatures (available on <u>https://www.accuweather.com/fr/ci/cote-divoire-</u> <u>weather</u>) show a range of 32 to 36 °C in Korhogo and 31 to 35 °C in Abidjan. These daily averages do not vary by more than one degree over several consecutive days. This may suggest stability of the atmosphere.

172 Wind roses presented in Fig. 4 highlight important north (N) and north-east (NE) winds in 173 Korhogo (Fig. 4 (a)) and south-west (SW) winds in Abidjan (Fig. 4 (b)). The regional climate is 174 defined by the position of the Intertropical Convergence Zone (ITCZ), which regulates the succession of weather seasons (Nouaceur and Murarescu, 2020). During measurement campaign, 175 176 the ITCZ is located above the latitude of Abidjan, favoring the arrival of air masses from the south 177 to Abidjan and air masses from the north to Korhogo where high temperature and low relative 178 humidity are typical for dry seasonal conditions. Such features are confirmed by the use of 179 HYSPLIT trajectories (https://www.ready.noaa.gov/HYSPLIT_traj.php) as shown by Fig. S1 (in Supplementary files). HYSPLIT trajectories represent the movement of air masses arriving at the study areas during the measurement campaigns. Fig. S1 shows 14- and 21-days back-trajectories before the end dates of each measurement campaign respectively for Korhogo and Abidjan respectively. It gives an overview of the dominant air mass origins over the measurement period. As expected, Korhogo is under the influence of northern trajectories whereas Abidjan is under the influence of southern air masses, associated to air masses coming from the Nigerian coast.

trajectories, emphasis was put on MODIS active fire maps and Aerosol Optical Thickness data (AOD), obtained from <u>https://neo.sci.gsfc.nasa.gov</u>. Figs. S2 present actives fires (a and b) and Aerosol Optical Depth (c and d) in January (a and c) and February (b and d) 2019, representative of the study period. Associated with Figs. S1, this shows that Korhogo is under the influence of regional winds that carry desert dust and biomass burning aerosols coming from the north and south of the city. While at Abidjan these regional winds show that only marine aerosols are expected.

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194 Material and Methods

195 Measurement campaigns and equipment

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The first step of the methodology deals with intensive data collection through measurement campaigns conducted at sites in the two cities, as presented in previously. The measurement campaigns of PM_{10} and $PM_{2.5}$ concentrations were carried out with 4 portables PM measuring devices (Aeroqual monitor S500 with $PM_{10}/PM_{2.5}$ sensor). Particulate pollution is measured over 3 days at each site, with 2 consecutive working days and 1 weekend day, because these 2 periods (working and weekends days) present different activities levels in urban areas. Indeed, studies have shown different levels of pollution at these 2 periods of the week (Adame *et al.*, 2014; *Doumbia et* *al.*, 2012; Adon *et al.*, 2020). During the campaigns, the measurements are not simultaneous on all
sites of each city because rotations are organized. During the campaigns, the measurements are not
simultaneous on all sites of each city. Rotations are organized with at maximum 4 sites running in
parallel. This is possible due to atmospheric stability detailed in the meteorological paragraph. This
technique of measuring on non-permanent sites during short periods allows to maximize
measurements on a large number of non-permanent sites with a small number of equipment.

210 In this study, optical PM₁₀/PM_{2.5} sensor on S500 Aeroqual monitor is used to measure 211 particulate pollution, because it is easy to use and take up very little space. This sensor covers a 212 concentration range from 0 to 1000 μ g m⁻³, and simultaneously measures PM₁₀ and PM_{2.5}. These types of sensors are based on light scattering and have an accuracy that can be influenced by the 213 214 physical and chemical characteristics of aerosols (Wang et al., 2015). The measuring device has 215 been placed in a box to protect it from natural elements. The device is installed at a height of 216 between 2 and 3 m. The time step chosen is 3 minutes, in order to allow fine sampling and to store 217 the data over the entire measurement period. After data retrieval, the data logger of the equipment 218 is formatted before being installed at another site.

219 Quality Control data: The evaluation of the uncertainties associated with low- and middle-cost 220 sensor measurements is the subject of several studies, as shown by Wang et al., (2015) who tested 221 3 low-cost sensors in combustion chambers with different results. Consequently, it is necessary to 222 evaluate and correct optical sensor concentration measurements against a reference method 223 (Clements et al., 2017). The reference method here proposed, is the gravimetric method. At some 224 of the selected sites, aerosol sampling lines are installed, including a fixed flow pump, a volume 225 meter, a valve, an aerosol size selection system (cane or cut-off head). These lines allow sampling 226 by filtration of aerosols on quartz filters, over the same periods the same sites than the low-cost 227 sensor measurements. The filters are weighed on a submicron balance before and after exposure in order to deduce the collected aerosol masses. This reference method has been used and validated
in several research projects and programs such as the International Network to study Deposition
and Atmospheric chemistry in Africa (INDAAF) program (ref.) and the DACCIWA program
(Djossou, 2018).

232 Geometrical means of the measurements obtained with low-cost sensors are compared with 233 those obtained by the gravimetric method (reference method). Geometrical means concentrations 234 of optical sensors are consistently lower than the gravimetric concentrations of 2.7 ± 1.1 and 3.5 ± 0.9 235 for PM₁₀ and PM_{2.5} respectively. The standard deviation of these coefficients is higher for PM₁₀ 236 than for PM_{2.5}, the latter having more points of comparison. It is worth noting that these values are 237 in agreement with those of Zuidema et al., (2019) who showed the need for a correction factor of 238 2.9 for an optical photometric sensor (GP2Y1010AU0F, Sharp Electronics, Osaka, Japan) 239 compared to the gravimetric method. Similarly, Egondi et al., (2016) also showed overestimated 240 concentration values given by the DustTrak II 8532 portable samplers compared to a reference 241 method. These factors (2.7 and 3.5) are then applied to PM₁₀ and PM_{2.5} concentrations at all 242 measurement sites.

243

244 Spatial mapping

Geographic Information System (GIS) provide a robust method of spatial analysis through interpolation. The most commonly used methods, kriging (geostatistical method) and reverse weighting of IDW (deterministic method), have already been compared in works. For example, Shukla *et al.*, (2020) in the Delhi megacity showed that kriging and IDW methods are associated to errors close to 22% and 24% respectively. In addition, Bahino *et al.*, (2018) carried out a spatial distribution of gaseous pollutants (NO₂, SO₂, NH₃, HNO₃ and O₃) in Abidjan (Cote d'Ivoire) and Cotonou (Benin) during the dry season within DACCIWA Program. The latter study included also a comparison between kriging and IDW, and showed that IDW method is associated to lower uncertainty than kriging method. In addition, our measurement network (number of points and geographic locations) is developed on the basis of the network of Bahino *et al.*, (2018); therefore, in this study, the IDW method is chosen and applied to the collected data.

ArcGIS version 10.3 software (ESRI Inc.) is used because it combines geostatistical data processing and map edition. The shapefiles for map editing are available for free on OpenStreetMap (https://www.openstreetmap.org). Briefly, IDW interpolation allows to predict the concentrations around a sampled point from the following formula presented in Eq. (1).

$$260 \quad Z_{p_j} = \sum_{i=1}^N \lambda_i * Z_{O_i} \tag{1}$$

where,

 Z_{O_i} is the observed value at location *i*; form geometrical mean of measured values at each site during the different measurement campaigns. These geometrical means, averages deviation, minimums and maximum are available, for Korhogo in Table S3 (for PM₁₀) and Table S4 (for PM_{2.5}) and for Abidjan in Table S7 (for PM₁₀) and Table S8 (for PM_{2.5}), in Supplementary files for each site.

267 Z_{P_i} is the predicted value for location *j*;

N is the number of measured sample points surrounding the predicted location that will be used inthe prediction;

270 λ_i are weighting factors applied at each measured value at location *Oi*. They decrease according 271 to the separation distance.

The geostatistical analysis of the data enables the values to be evaluated in order to develop the spatial variation maps. Thus, the IDW method calculates values from the collected data for each point (unsampled and sampled) and gives average predicted values everywhere (Rivera-González *et al.*, 2015; Zamri *et al.*, 2012; Bahino *et al.*, 2018). The Root Mean Square Error (RMSE) (McKeen *et al.*, 2005; Savage *et al.*, 2013; Chai and Draxler, 2014) and the Mean Absolute Error (MAE) (Taylor *et al.*, (2013); Chatterjee *et al.*, (2013) and Jerez *et al.*, (2013) are the two parameters widely used to evaluate the quality of the predictions of interpolation methods. In order to assess the quality of the prediction in the two cities, both RMSE and MAE values are calculated for each measurement campaign following the formulas present in Eqs. (2) and (3).

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$$RMSE = \sqrt{\frac{1}{N} \sum_{i=1}^{N} (Z_{P_i} - Z_{O_i})^2}$$
 (2)
283 $MAE = \frac{1}{N} \sum_{i=1}^{N} |Z_{P_i} - Z_{O_i}|$ (3)

where,

285 Z_{P_i} and Z_{O_i} are respectively the predicted and observed value for a location *i*;

286 *N* is the number of measured sample point.

Table 1 presents such values for the 2 cities as well as the coefficient of variation of error. In Abidjan, the MAE values are lower than the RMSE values due to the high uncertainties associated with some sites whereas no marked differences appear in Korhogo. Also, both coefficients are higher in Korhogo than in Abidjan: that could be due to the high concentration values found in Korhogo. As shown in Table 1, the coefficients of variation of error in Abidjan are higher than in Korhogo. These low values obtained in Korhogo may be due to a smaller study area than in Abidjan. This analysis highlights the impact of the distance between measurement sites on the uncertainties.

295 **RESULTS AND DISCUSSIONS**

296

297 Spatial distribution of PM concentrations

298 Korhogo PM concentrations maps

As previously described, the PM₁₀ and PM_{2.5} concentrations obtained by geographical averages allowed the mapping of spatial variations for the Korhogo study area. These maps are presented in Figs. 5(a) and 5(b), respectively for the PM₁₀ and PM_{2.5}. Concentrations are very high, between 212.2 μ g m⁻³ and 534.7 μ g m⁻³ for PM₁₀ and 54.3 μ g m⁻³ and 218 μ g m⁻³ for PM_{2.5} and well above the World Health Organization (WHO) standards which stipulate that PM_{2.5} should not exceed 25 μ g m⁻³24-hours mean and that PM₁₀ not exceed 50 μ g m⁻³24-hours mean.

As shown in Fig. 5(a), the measurement sites in the city downtown have higher PM₁₀ concentrations, namely K₂ with 534.7 μ g m⁻³, K₆ with 374.4 μ g m⁻³, K₁₁ with 475.1 μ g m⁻³ and K₁₄ with 446.7 μ g m⁻³ compared to the sites in the northern and southern outskirts of the city such as K₁ with 241.4 μ g m⁻³, K₈ with 245.1 μ g m⁻³ and K₁₂ with 261.2 μ g m⁻³. This can be explained by the economic activities mainly occurring in Korhogo downtown, associated with intense road traffic dominated by 2-wheel vehicles and the presence of the urban bus stations. The highest concentrations observed in K₂, K₁₁ and K₁₄ sites are also due to the proximity of busy highways.

 K_{13} and K_5 sites have also high PM_{10} concentrations because they are in neighborhoods with unpaved roads. Moreover, even if downtown roads are mostly paved, a high resuspension of dust by intense traffic is also observed. Indeed, as mentioned earlier, dust deposition on roads are also observed at Korhogo due to northerly air masses transporting desert dust aerosols at the period of this study (Fig. S1). Such aerosols significantly contribute to the pollution level of Korhogo.

Finally, Fig. 5(a) shows that in Korhogo, PM_{2.5} and PM₁₀ spatial variation maps roughly present the same pattern of high and low concentrations despite some differences which are detailed in the following section.

320

321 Abidjan PM concentrations maps

322 The spatial variation maps of PM₁₀ and PM_{2.5} concentrations (obtained from the geographical 323 averages of the measured concentration time series) in Abidjan study area are given respectively in Figs. 6(a) and 6(b). The average concentrations are between 38.1 μ g m⁻³ and 160.4 μ g m⁻³ for 324 PM₁₀ and 23.8 µg m⁻³ and 113.4 µg m⁻³ for PM_{2.5}. For PM₁₀, the highest concentrations are found 325 at sites A7 (160.4 μ g m⁻³), A₁₈ (125.3 μ g m⁻³) and A₂₀ (120.7 μ g m⁻³). The A7 and A₂₀ sites are 326 327 located nearby unpaved roads (Table S2) therefore, dust from road resuspension contribute to the 328 increase of concentrations. In A₁₈ site, the high concentrations may be explained by the proximity 329 of important road traffic and industrial activities. In addition, A_{18 and} A₁₉ sites are located on the sea 330 front, and are therefore subject to a high contribution of marine aerosols.

331 Sites included in areas with better quality infrastructures and with middle- or high-class 332 populations, have generally lower PM_{10} concentrations (e.g. A_2 , A_{11}) compared to sites in 333 neighborhoods with poor quality infrastructures (e.g. A_3 , A_5 , A_7 , A_8 , A_{16} , A_{18} and A_{20}).

334 Compared to PM₁₀, PM_{2.5} spatial variation (Fig. 6(b)) presents high concentrations over most of the study area, despite some very localized low concentrations in A1 (32.9 μ g m⁻³), A2 (24.9 μ g m⁻ 335 ³), A₅ (30.8 µg m⁻³), A₁₄ (23.8 µg m⁻³) and A₁₉ (24.5 µg m⁻³). PM_{2.5} concentrations in northern 336 337 Abidjan are higher than in southern Abidjan. Indeed, the wind rose of Abidjan shows a dominance 338 of south, southwest winds, which could lead to a transport and an accumulation of pollutants in the northern area of the city, as for Bahino et al., (2018). Values for these sites (A7 with 113.4 µg m⁻³, 339 A₉ with 91.7 μ g m⁻³, A₁₂ with 73.2 μ g m⁻³ and A₂₀ with 91 μ g m⁻³) are 2 to 6 times higher than the 340 WHO daily standard ($25 \mu g m^{-3}$). Moreover, these sites mainly appear in low income neighborhood. 341 342 Indeed, in such areas, high concentrations result from the combination of the use of wood for 343 cooking, the proximity of public transport (small bus: gbaka, municipal taxis: wôrô-wôrô) with 344 very old and poorly maintained vehicles and the burning of solid domestic waste due to collection

failure. In contrary, the less densely populated areas (high-class districts) have lower concentrations (sites A₁, A₂, A₁₄) than the medium and low-class districts. In these areas (sites A₁, A₂, A₁₄), people have higher incomes and are more likely to use gas for cooking than wood and charcoal for cooking activities. Lower concentrations are observed in A₅ and A₁₉ sites, although located in popular districts. That may be due to less intensive combustion activities than usually observed in these districts for the A₅ site and to the proximity of the sea front for A₁₉ site.

351

352 **Relative importance of fine particles**

353 In order to highlight the relative contribution of particulate size and aerosol sources, spatial 354 variation maps of the percentage of relative contribution of PM2.5 to PM10 are presented in Figs. 7 355 and 8 for Korhogo and Abidjan respectively. Low values represent a high contribution of PM2.5 to PM₁₀, while high percentages represent a low contribution of PM_{2.5} to PM₁₀. It is important to note 356 that these figures show totally different values for the two cities. In Korhogo, more than 50% of 357 358 PM₁₀ are coarse particles, with the highest values in the north and south of the city and the lowest 359 values in the city center, where most urban activities take place. In Abidjan, the situation is totally different: coarse particles in PM₁₀ represent less than 40%, with the highest values in A₄, A₅, A₁₄, 360 361 A₁₈ and A₁₉ sites and the lowest values in A₆, A₉, A₁₁, A₁₂, A₁₃ and A₁₅.

In both cities, the lowest values associated with high PM_{2.5} are found particularly in densely populated sites or in economically very active neighborhoods with intense combustion sources and fine particle emissions. In addition, the percentages of coarse particles (PM_{2.5-10}) relative to PM₁₀ are high near unpaved roads and sometimes in areas with heavy traffic.

366 In Abidjan, low contributions of $PM_{2.5}$ to PM_{10} are also observed in industrial areas (site A_{18}). 367 In these areas, heavy truck traffic damages the roads and causes dust resuspension. In addition, a 368 contribution from marine aerosols is possible due to the south-westerly (SW) wind direction. These 369 values are in line with our expectations and confirm the good representation of the measurement 370 sites and the good performance of the spatial distribution, because they reflect the main sources of 371 pollution that influence each city.

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373 Daily distribution of PM concentrations

Daily profiles of fifteen minutes average concentrations of PM are shown in Fig. 9. The average daily profiles of PM₁₀ and PM_{2.5} concentrations in Korhogo are presented in Figs. 9(a) and 9(b) respectively. For Abidjan, Figs. 9(c) and 9(d) represent the sites with similar patterns of PM_{2.5} concentrations and the other sites respectively. Since PM₁₀ and PM_{2.5} daily profiles are similar in Abidjan, only PM_{2.5} concentrations are presented here. Note that the daily profiles of PM₁₀ concentrations in Abidjan are shown in Figs. S5(a) and S5(b).

380 In Korhogo, all the sites show the same daily average concentration profiles for PM₁₀ and PM_{2.5} 381 respectively in Figs. 9(a) and 9(b), with two very important daily peaks. One peak in the morning 382 and one in the evening. These peaks are due to the mix of several local sources: mainly, household sweepings (twice a day), use of firewood (morning and evening), and road traffic (dominated by 383 384 2-wheelers). Focusing on these peaks, PM_{2.5} morning and evening peaks are of similar importance 385 whereas PM₁₀ peak is higher in the evening than in the morning. This could be due to favorable 386 conditions for dust resuspension due to the higher dryness of the soil in the evening caused by 387 sunshine during the day. According to Cheng and Li (2010) and Gualtieri et al., (2015), 388 meteorological parameters such as temperature and relative humidity influence particulate matter 389 concentrations. Moreover, the low values observed during the day, twenty times lower than in 390 morning could be due to a decrease of urban activities. However, it is interesting to note that such 391 low average concentrations observed in Korhogo are still higher than the WHO daily standard.

392 In Abidjan, Fig. 9(c) shows morning peaks of concentrations which are more important than 393 those of the evening. Traffic source are mainly responsible of these peaks. The difference between 394 the morning and evening peaks can be explained by the fact that the flow of road traffic in the 395 evening is more spread out in time compared to that of the morning. In Fig. 9(d), some sites show 396 no clear daily variation. For example, A₁₈ and A₂₀ sites show high concentrations throughout the 397 day, characteristic of local economic activities. Indeed, the A₈, A₁₈ and A₂₀ sites are respectively 398 located in a low-income neighborhood (mainly using wood and charcoal for cooking), in a 399 neighborhood under construction and in an industrial area respectively. The daily minimum average concentrations are 19.3 μ g m⁻³ for PM₁₀ and 14 μ g m⁻³ for PM_{2.5}. These low values are 400 compensated by the high values observed during peaks (beyond 350 μ g m⁻³ as well as PM₁₀ and 401 402 PM_{2.5}), thus providing a daily average concentration well above WHO standards. To conclude, this data collection methodology provides daily profiles of PM concentrations, which are important in 403 the analysis of population exposure levels. 404

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406 Comparison with other Sub-Saharan Africa cities

407 Concentrations of PM_{10} and $PM_{2.5}$ in Korhogo and Abidjan are generally above the WHO daily 408 exposure standard. As shown previously the local variability of these concentrations is strongly 409 related to the predominant local sources.

Concentrations vary on average from 212.2 μ g m⁻³ to 534.7 μ g m⁻³ for PM₁₀ and from 57.4 μ g m⁻³ to 230.4 μ g m⁻³ for PM_{2.5} in Korhogo and in Abidjan, from 38.1 μ g m⁻³ to 160.4 μ g m⁻³ for PM₁₀ and from 23.8 μ g m⁻³ to 113.4 μ g m⁻³ for PM_{2.5}. The mean standard deviations of measured concentrations are of 89 μ g m⁻³ and 23.5 μ g m⁻³ for PM₁₀ and of 41.5 μ g m⁻³ and 17.9 μ g m⁻³ for PM_{2.5} respectively in Korhogo and Abidjan. The standard deviation values are lower in Abidjan compared to Korhogo, which shows a greater variability in the latter. The maximum concentrations

in Korhogo are respectively more than 3 to 5 times higher than in Abidjan whereas the minimum
are 2 to 3 times higher. As already mentioned, this may be explained by additional aerosol sources
occurring in Korhogo such as dust aerosol coming from northern African areas and/or from
unpaved roads and the 2-wheel emissions.

Table 2 presents the daily concentrations measured in African urban areas by this study, compared to literature results for Sub-Saharan Africa. $PM_{2.5}$ concentrations measured in this study are lower than those measured with the individual exposures of Xu *et al.*, (2019) in Abidjan. Personal exposure measurements are generally closer to the emission sources than our sites.

424 The PM₁₀ measurements performed by Ana et al., (2014) from January to March 2008 in Ibadan (Nigeria) are on the same order of magnitude as the concentrations measured at Korhogo in January 425 426 2019. These PM₁₀ concentration levels are characteristic of the dry season in West Africa. For the 427 same period of the year, Antonel and Chowdhury (2014) found lower PM₁₀ concentrations (105 µg 428 m⁻³) in Bafossam (Cameroon), which is further south, in a forest area, therefore less exposed to 429 dust. Furthermore, the results obtained by Dionisio et al., (2010) showed that concentrations in 430 low-income areas are higher than in high-income areas in Accra, which is in agreement with the 431 results of this study. Finally, the ranges of concentrations obtained at Korhogo (in Tables S3 and 432 S) and Abidjan (in Tables S7 and S8) measurement sites are on the same order of magnitude as 433 those obtained in other studies carried out in Sub-Saharan Africa. Values from other studies 434 conducted in Africa are presented in Table S9 and agree very well with the results of this study.

435

436 CONCLUSIONS

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This study provides an assessment of air quality in the cities of Korhogo and Abidjan in Cote
d'Ivoire by proposing for the first time to our knowledge, spatial variation maps of PM₁₀ and PM_{2.5}

440 concentrations. It is part of the objectives of the PASMU project, which aims to characterize urban441 air pollution and its impact on health in Cote d'Ivoire.

The use of low-cost sensors at 14 sites in Korhogo and 20 in Abidjan resulted in daily average PM₁₀ concentrations ranging from 212.2 μ g m⁻³ to 534.7 μ g m⁻³ in Korhogo and from 38.1 μ g m⁻³ to 160.4 μ g m⁻³ in Abidjan. PM_{2.5} concentrations range from 57.4 μ g m⁻³ to 230.4 μ g m⁻³ and from 23.8 μ g m⁻³ to 113.4 μ g m⁻³ in Korhogo and Abidjan respectively.

446 The spatial variation maps obtained through the application of GIS in Korhogo show that the 447 high activity areas (market and national bus station) have higher PM₁₀ and PM_{2.5} concentration 448 levels than other areas. Our study also highlights that the relative contribution of PM_{2.5} to PM₁₀ is very different in Korhogo and Abidjan, varying from 50 to 100% in Abidjan, while 20 to 50% in 449 450 Korhogo. In Abidjan, a difference between the concentrations is observed in high- and low-income 451 neighborhoods, with higher concentration levels in low- and middle-income neighborhoods. One 452 likely reason is that, high-income areas have better services (roads, waste collection, ...), and also 453 use less polluting cooking methods (e.g. gas). This result can be also explained by the status of the 454 city: (1) in Korhogo, the higher presence of coarse particles is due to anthropogenic (road dust 455 resuspension) and natural sources (desert dust transport); (2) Abidjan is an economic capital with 456 a higher population density, road traffic and generally a higher activity density than in Korhogo. 457 The concentrations measured in this work are comparable to previous studies conducted in Abidjan and in Africa. 458

In conclusion, this study indicates that this simple and reproductive method provides a quality assessment of particulate pollution at the city level. Note that this method could be also used for other purposes such as performing a quick assessment of the impact of a new industrial unit or infrastructure on the surrounding air quality. Finally, the original results obtained on urban

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particulate pollution in Korhogo and Abidjan, relative contribution of fine particle to coarse and
daily variations of PM concentrations will be very useful for urban modelling in West African cities.

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- 478

479 **DISCLAIMER**

- 480
- 481 The authors decline any commercial link with the companies producing equipment or software482 used in this study.

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List of Table titles

Table 1. Interpolation error statistics, where RMSE is root mean absolute error in μ g m⁻³; MAE is mean average error in μ g m⁻³ and CV is mean of coefficient of variation of error in %.

Table 2. Comparison between PM₁₀ and PM_{2.5} daily concentrations obtained in this study and other studies Sub-Saharan Africa.

Т	ื่ล	bl	e	1.
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Dellutente	Kor	hogo	Abic	ljan
Pollutants	PM ₁₀	PM _{2.5}	PM ₁₀	PM _{2.5}
Root Mean Absolute Error (RMSE)	97.0	43.2	41.5	24.9
Mean Average Error (MAE)	105.6	40.0	31.0	19.98
% Coefficient of Variation of error	33.4 ± 12.6	36.7 ± 31.7	42.5 ± 41.8	39.7 ± 29

City, Country	Site characteristic	Samplings period	PM _{2.5} (μg m ⁻³)	PM ₁₀ (μg m ⁻³)	Reference	
Korhogo, Cote d'Ivoire	14 urban sites	Jan-2019	54.3 to 218	212.2 to 534.7	This study	
Abidjan, Cote d'Ivoire	20 urban sites	Jan. to Feb- 2019	23.8 to 113.4	38.1 to 160.4		
Abidjan, Cote d'Ivoire	Urban (WB ^{<i>a</i>})		494.3	-		
	Urban (DF^b)	Jan. and July-	358.8	-	Xu et al.,	
Cotonou, Benin	Traffic	2016	335.1	-	(2019)	
Accra, Ghana	Traffic-highest SES ^c	2007	21	49	Dionisio et	
	Traffic-lowest SES	2007	39	96	al., (2010)	
Dakar, Senegal	Urban	Jun-2008 to Jun-2009	44.4	3-	Doumbia <i>et</i> <i>al.</i> , (2012)	
	Commercial area			422.3		
Ibadan,	Traffic	Jan. and Mar-	-	328.1	Ana <i>et al.</i> ,	
Nigeria	Industrial	2008	-	319.7	(2014)	
	Residential		-	257.4		
Bamako, Mali	Urban	Sep. to Oct- 2012 and Jul- 2013	43	210	Garrison <i>et al.</i> , (2014)	
Ouagadougou, Burkina Faso	Urban	Nov. to Dec-	86	-	Boman et al., (2009)	
	Suburban	2017	69	-		
Accra, Ghana	Traffic	Jun. to Jul- 2006	27.4	71.8	Arku <i>et al.,</i> (2008)	
Bafoussam, Cameroun			67	105		
Bamenda, Cameroun	Urban	Jan. to Mar- 2012	132	141	Antonel and Chowdhury (2014)	
Yaoundé, Cameroun			49	65	()	
Libreville, Gabon		Jun. to Jul- 2015	35.8	-	Ngo et al	
Port-Gentil, Gabon	Urban		60.9	-	(2019)	

Table 2.

(^{*a*}) **WB:** Waste Burning; (^{*b*}) **DF:** Domestic Fire; (^{*c*}) Social Economic Status; "-": not measured

Figure Captions

Fig. 1. Map of the population density of Cote d'Ivoire, the position of two application cities in Cote d'Ivoire is shown by green diamond. Green diamond represents each measurement cities. Data from https://www.diva-gis.org/gdata.

Fig. 2. Localization of measurement sites in Korhogo (Cote d'Ivoire). Green diamond represents each measurement site with its corresponding ID, described in the Table S1 and S3 and the black lines represent the boundaries of Korhogo city.

Fig. 3. Localization of measurement sites in Abidjan (Cote d'Ivoire). Green diamond represents each measurement site with its corresponding id site, described in the Table S2 and S4 and black lines represent the boundaries of communes included in this study.

Fig. 4. Wind rose from synoptic station measurements during the study periods in Korhogo (a) and Abidjan (b) in Cote d'Ivoire.

Fig. 5. Spatial variation of PM_{10} (**a**) and $PM_{2,5}$ (**b**) concentrations in Abidjan (Cote d'Ivoire). Black diamond represents each measurement site, line represent the main roads of Abidjan and dotted line represent administrative boundaries.

Fig. 6. Spatial variation of PM_{10} (**a**) and $PM_{2.5}$ (**b**) concentrations in Abidjan (Cote d'Ivoire). Black diamond represents each measurement site, line represent the main roads of Abidjan and dotted line represent administrative boundaries.

Fig. 7. Spatial variation of relative difference between PM₁₀ and PM_{2.5} in percentage in Korhogo (Cote d'Ivoire). Black diamond represents each measurement site.

Fig. 8. Spatial variation of relative difference between PM₁₀ and PM_{2.5} in percentage in Abidjan (Cote d'Ivoire). Black diamond represents each measurement site.

Fig. 9. Daily averages of PM_{10} and $PM_{2.5}$ concentrations in Korhogo (**a**, **b**) and $PM_{2.5}$ concentrations in Abidjan (**c**, **d**).



Fig. 1.

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Fig. 4.



Fig.5.



Fig. 6.







Fig.	9.
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