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# Non-Fickian dispersivity investigation from numerical simulations of tracer transport in a model double-porosity medium at different saturations

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#### Abstract

It is generally admitted that dispersivity is an indicator of the heterogeneity scale of porous media. This parameter is assumed to be an intrinsic property which characterizes the dispersive behavior during the transport of a tracer in a porous medium. When the medium is saturated by two fluid phases (water and air), dispersivity depends strongly on saturation. "Doubleporosity" medium concept can be attributed to a class of heterogeneous soils and rocks in which a strong contrast in local pore size characteristics is observed. In this work, we characterized non-Fickian dispersivities of a double-porosity medium at different saturations, by performing numerical simulations for a series of one-dimensional experiments of tracer dispersion under different initial and boundary conditions. The physical double-porosity model was composed of solidified clayey spheres, distributed periodically in a more permeable sandy matrix. Using a two-equation macroscopic model, numerical simulations reproduced very well the experimental data, thus allowing to determine the dispersivity for different transport scenarios. For the first time, the existence of a unique dispersivity of a double-porosity medium at a given saturation was demonstrated for different transport scenarios of initial and boundary conditions. The saturation dependence of the dispersivity in the double-porosity medium was established and compared with the trends obtained for the single-porosity soils in previous studies.

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### **Keywords**

1 Non-Fickian dispersivity; experiment; numerical simulation; saturation dependence; double-

2 porosity medium.

#### 1. Introduction

Prediction of solute transport in a geological medium requires a thorough understanding of both physiochemical-biological processes and porous medium properties. One of the key processes is the dispersive transport mechanism that can be characterized by the dispersivity (Bear, 1972). This parameter is generally defined as the ratio between the dispersion coefficient and the flow velocity (Bear, 1972; Fetter, 1999). It is empirically determined from concentration measurements (evolution with time and space) during laboratory and field tracer tests by using conventional/advanced models of convection-dispersion or spatial moments analysis (Aris, 1956; Fried, 1975; Gaudet et al., 1977; Khan and Jury, 1990; Gelhar et al., 1992; Vanderborght and Vereecken, 2007; Tran Ngoc et al., 2011; Robert et al., 2016). The existence of a unique dispersivity within an infinite heterogeneous medium was regarded in theoretical investigations of the scale effect on dispersivity (Gelhar et al., 1979; Matheron and De Marsily, 1980). Although it is largely admitted that dispersivity is an intrinsic soil property, rare studies, especially for heterogeneous media, have shown a unique dispersivity as an intrinsic value which could be simultaneously used to reproduce different breakthrough curves (BTC) under different transport boundary conditions. Solute transport observed in porous heterogeneous media is often non-Fickian, i.e. preferential

transport, earlier arrival and tailing effect in BTCs, due to scale-dependent dispersion/dispersivity (Wheatcraft and Tyler, 1988; Sternberg et al., 1996). A general consensus about the scale-dependent dispersivity has been reported by many laboratory and field researches over the last few decades (Silliman and Simpson, 1987; Gelhar et al., 1992; Irwin et al., 1996; Pachepsky et al., 2003; Huang et al., 2006, among others). It means that during solute transport in heterogeneous media the dispersivity increases with travel distance (thus with time) in a pre-asymptotic stage before reaching a constant value (Gelhar, 1987; Mishra and Parker, 1990; Suresh Kumar et al., 2006; Sharma and Abgaze, 2015). Irwin et al. (1996) showed that dispersion may be dependent upon the scale of a characteristic length (travel distance) from their KCl tracer transport experiment in a medium with periodic heterogeneity. However, the experiments of Sternberg et al. (1996) pointed out that dispersion coefficient could increase or decrease depending on the heterogeneity arrangement of glass particle sets in periodic heterogeneous porous media. Majdalani et al. (2015) conducted several

replicable transport experiments in a periodic heterogeneous medium, composed of a set of PVC columns containing glass spheres surrounding a cylindrical cavity placed in the center of the column, acting as a wide pore that induces preferential water and solute transfer. They found statistically that the dispersion coefficient could be a function of the medium length or not, taking into account sampling effects. Recently, from small- and large-scale experiments with different soil volumes (0.15 – 0.60 m in column length and 0.1 – 0.2 m in diameter) performed in tropical intact soil columns, Godoy et al. (2018) pointed out an increasing trend of the dispersivity with column lengths fitted by exponential functions due to heterogeneous arrangements in soil samples, but no clear dependence of the dispersivity on sample diameters and volumes was observed.

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In the meanwhile, there has been no general conclusion concerning the saturation-dependent dispersivity for unsaturated media. A correlation between the increase of dispersivity with decreasing saturation has been stated according to the investigation of tracer dispersion experiments in homogeneous sand columns (single-porosity media) (Yule and Gardner, 1978; De Smedt et al., 1986; Haga et al., 1999; Padilla et al., 1999; Nutzmann et al., 2002; Latrille, 2013). However, Raoof and Hassanizadeh (2013) proposed a non-monotonic relation of the dispersivity with saturation. This relation is compatible with the results of hydrodynamic dispersion experiments of Toride et al. (2003) in unconsolidated sand columns under variably saturated conditions. They observed for intermediate values of water saturation, an increase of the dispersivity as the saturation decreases, while the contrary is reported for small saturations. The experimental results of other authors (Maraga et al., 1997; Sato et al., 2003; De Witte, 2017) showed also a non-monotonic variation of the dispersivity as a function of the saturation (see in Tran Ngoc et al., 2017). It should be noted that saturation distribution profiles were directly measured in a few studies (Toride et al., 2003; Latrille, 2013). It is also worthwhile to note that non-Fickian transport was observed in homogeneous unsaturated media (singleporosity media) for saturation profiles, stated uniform throughout the columns, established by unit gravitational hydraulic gradient (Yule and Garner, 1978; Maraqua et al., 1997; Toride et al., 2003), but mostly Fickian transport for uniform saturation profiles were established by controlled air suction (Sato et al., 2003). Using a porous micro-model made of a polymer for the dispersion experiments of an ink-water system, Karadimitriou et al. (2016; 2017) confirmed a parabolic behavior of dispersion coefficients with saturations (leading to non-monotonic dispersivities) because of the absence (high saturation) or presence (small saturation) of stagnant zones in the porous medium. Juarez-Morejon (2017) found an increase of the

1 dispersivity with smaller saturations in a series of experiments simulating crude oil recovery 2 by flooding of water and polymer solution in a homogeneous sandstone. Therefore, the practice 3 of performing numerical simulations for real-life transport problems using as the dispersivity 4 value an estimate based on the length of the contaminant plume or the distance between the 5 source and specific measurement points (travel distance) (Sudicky, 1986; Freyberg, 1986), has 6 to be considered with caution. Moreover, this practice would not lead to a good estimate 7 without taking into account the dependency of dispersivity upon saturation, especially for 8 transport problems in a heterogeneous vadose zone.

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Dispersivity studies for unsaturated heterogeneous media are less numerous compared to those for homogeneous media. Indeed, laboratory experiments conducted to investigate saturationdependency of the dispersivity in unsaturated heterogeneous media are scarce in the literature. A class of heterogeneous media can be studied by the concept of "double-porosity" media (Barenblatt et al., 1960) in which a contrast in local macro-micro pore size characteristics is observed. Solute dispersion phenomena in the macro-pores with interactions/exchanges with micro-pores are commonly observed (Gerke and van Genuchten, 1993; Schwartz et al., 2000; Singha, 2017, among others). This leads to non-Fickian behavior characterized by a corresponding dispersivity. The available dispersivity-saturation relationship for singleporosity media may be applicable for double-porosity media if the convective flux and thus dispersion only takes place only in the mobile zone represented by the macro-pore domain. However, a study on the relation between the non-Fickian dispersivity and water saturation for transport in a double-porosity medium is still needed to provide a confirmative proof, because dispersion also depends on heterogeneity of microstructure in such media as well as dispersion in the macro-pore domain and diffusion in the micro-pore domain. To our knowledge, no trend has been established for this relation in a double-porosity medium with a known microstructure. Golfier et al. (2011) used experiments of solute transport in a bimodal medium (fine sand inclusions surrounded by coarse sand) to study the impact of injection conditions on the Fickian or non-Fickian transport, using the same dispersivities for the two material constituents in numerical simulations by macro-scale models. Under saturated conditions, Paterson et al. (1996) found a dispersivity of 0.04 cm at the transition regime of diffusiondispersion (Péclet number ~ 1) for a double-porosity medium made of porous grains with the permeability ratio between intra- and extra-grain domain of 2.5×10<sup>-2</sup>. This dispersivity value is of the same order of magnitude as the heterogeneity size (mean grain diameter of 0.05 cm). Tran Ngoc et al. (2007) carried out tracer dispersion experiments in an unsaturated periodic

double-porosity medium to validate a theoretical macroscopic model of the two-equation type (Tran Ngoc et al., 2011) and to highlight qualitatively double-porosity effects in non-Fickian behaviors (Tran Ngoc et al., 2014). The double-porosity medium is a physical model column at which the REV (Representative Elementary Volume), equivalent to a period, is well controlled by the microstructure of two constituents of strong hydraulic contrast (sintered clay spheres and sand). In a similar periodic double-porosity medium, a KBr transport experiment was done under unsaturated conditions by water infiltration and suction applied at the column inlet and outlet, respectively (Peng et al., 2015). For this heterogeneous medium, a dispersivity value of 0.13 cm was fitted for the obtained breakthrough curve which is smaller than the diameter of the clayey spheres (heterogeneity of medium) of 0.5 cm, although of the same order of magnitude.

Non-Fickian transport appearing in BTC data may possibly be a result of different superimposed physical, chemical and biological effects. In this work, our focus will be on anomalous dispersion due to physical effects and more particularly to the microstructure of the porous medium under study. Thus, it is necessary to carry out tracer transport experiments in a physical double-porosity medium with controlled microstructure under various physical frameworks: different saturation profiles and different exterior loadings. A BTC database constituted from these experiments will allow one to provide a proof of the existence of a dispersivity value that enables characterization of the non-Fickian BTCs for different transport scenarios.

The main objective of this study is to investigate the non-Fickian dispersivity of a double-porosity medium, the dependence of this parameter on the saturation in double-porosity medium and to compare the trends obtained with those observed for the single-porosity media in previous studies. To this end, we analyze the dataset of the experiments presented in Tran Ngoc et al. (2011; 2014) and of a new series of one-dimensional dispersion experiments performed under different initial and boundary conditions. The experimental BTC are fitted with those obtained by numerical simulations using the conventional and upscaled convection-dispersion model in order to identify dispersivity values for single- and double-porosity media with different saturations, respectively. A description of the experiment and new data analysis methods are presented in the two next sections. After that, the results and discussions are presented. The conclusions of the study will be addressed in the final section.

# 2. Materials and experimental methods

#### 2.1. One-dimensional physical model of double-porosity medium

The physical model of the double-porosity medium used in this study was initiated by Lewandowska (2004). It was constructed from fine sand Hostun HN38 and clay spheres which were washed and hand-modeled in a spherical form with an average radius R of 0.32 cm and then fired at 1000°C in a pottery furnace. These two geo-material constituents come from the aquifers of Hostun (Southeastern France) and La Bisbal (Northeastern Spain), respectively. The mineralogic analysis by X-Ray diffraction indicated Si and Si/Al/Ca/Fe for the main chemical composition of the sand and clay (montmorillonite type), respectively (Tran Ngoc, 2008). The physical and hydrodynamic properties of two porous media were investigated by Lewandowska et al. (2005; 2008); Tran Ngoc et al. (2011; 2014) and thus are reported in Table 1.

12 Table 1. Main soil properties of two geo-material constituents used for creating the model double-porosity medium.

Property		Sand	Clay
Grain density, $ ho_{ m g}$	[g/cm <sup>3</sup> ]	2.65	2.91 <sup>a</sup>
Dry bulk density, $ ho_{ m d}$	$[g/cm^3]$	1.59	1.89 <sup>a</sup>
Average grain diameter, $d_{50}$	[µm]	162	_
Pore-size, $\varXi$	[µm]	$10 - 500^{c}$	$0.1 - 1^{a}$
Porosity, n	[-]	0.400	0.348 <sup>a</sup> ; 0.343 <sup>b</sup>
Hydraulic conductivity, $K_{\rm S}$	[cm/s]	$2.87 \times 10^{-3}$	$1.96\times10^{-6}$
Air entry pressure, $h_a$	[cm of H <sub>2</sub> 0]	90	165

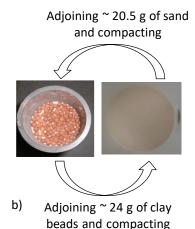
(a Average values obtained by two different mercury injection tests for two different clay spheres; b by gamma ray measurement for a cylindrical sample (25.9 cm high and 5.25 cm diameter) made of the same clay as the solidified clay spheres; Vitorge et al., 2013)

The formation of the double-porosity physical medium follows the same principle as that presented in previous studies (Lewandowska et al., 2005; 2008; Szymkiewicz et al., 2008; Tran Ngoc, 2008; Tran Ngoc et al., 2011; Lewandowska and Pilawski, 2012; Tran Ngoc et al., 2014). The HN38 sand and solidified clay spheres play the role of the macro-pores (between the sand grains) and micro-pores (inside the spheres) in the double-porosity medium, respectively. They were interlayered in a manner keeping a volumetric fraction (ratio between the volume of each

material constituent including solids and pores and the total volume of the sample) of about 50% for each porous material. We filled up about 65 layers of each constituent in an acrylic circular column of 6 cm inner diameter to reach the heights of about 50 cm (Table 2). Note that the first and last layers in the columns were composed of the sandy matrix to envelop clay spheres. In this way, a column of the pseudo-periodic double-porosity medium is constructed (Fig. 1 a and b) with a 3D periodic microstructure of  $0.64 \times 0.64 \times 1.3$  cm<sup>3</sup> modelled by the scheme of Fig. 1 c. This periodic cell is the REV of the medium. The hydraulic conductivity at 100% saturation (by pure water) of the obtained double-porosity medium was expected to be slightly smaller than  $1.50 \times 10^{-3}$  cm/s determined by brine injection (Tran Ngoc, 2008), due to the viscosity difference between water and brine. The hydraulic conductivity of  $0.97 \times 10^{-3}$  cm/s at very close to saturation (about 85%) was determined by using an infiltrometer (Jougnot, 2006; Jougnot et al., 2008). The double-porosity medium here is heterogeneous on the scale of observation but could be considered as a homogeneous medium on the scale of measurement. This may be consistent with natural soils, even if our physical model could not represent natural porous formations (Sternberg et al., 1996).

We also built single-porosity medium columns by filling up only Hostun 38 sand in the same columns used for the double-porosity medium model. The sand was compacted in a controllable manner to obtain the same porosity as the one measured for the sand matrix in the double-porosity medium (Table 2).





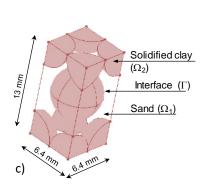


Fig. 1. Model column of the double-porosity medium being filled up periodically up to all the column (a), by interchanging the clay sphere and sand layers b, (the clay sphere layer displayed on the figure would be hidden after pouring in the column a sand layer), and c) its schematized microstructure period.

*Table 2. Characteristics and conditions of the dispersion experiments.* 

Tost Column	Truns	L	$n_1$	n <sub>2</sub>	<b>W</b> 1	W2	$m_1$	<i>m</i> <sub>2</sub>	$S_{\mathrm{w}}$	Tavg	q	$C_{ m ini}$	[g/l]	$C_{ m inj}$	[g/l]	
Test	Test Column Typ	1 ype	[cm]	[-]	[-]	[-]	[-]	[g]	[g]	[%]	[ <sup>0</sup> C]	[cm/s]	C'ini	$C^{"}$ ini	$C'_{ m inj}$	C''inj
1	A	Step	50.4	0.389	0.343	0.484	0.516	1117	1379	100	19.3	5.90E-04	20	0.90	0	0.90
2	В	Step	51	0.378	0.343	0.49	0.51	1168	1381	86	20.7	3.44E-04	0	0.90	5	0.90
3	C	Step	46.9	0.386	0.343	0.483	0.517	1044	1289	*	20.3	3.44E-04	0	0.90	5	0.90
4	В	Step	51	0.378	0.343	0.49	0.51	1168	1381	82	20.1	3.44E-04	0	0.70	5	0.70
5	В	Pulse	51	0.378	0.343	0.49	0.51	1168	1381	*	24.4	3.44E-04	0	1.27	5	1.04
6	C	Pulse	46.9	0.386	0.343	0.483	0.517	1044	1289	*	20.5	3.44E-04	0	0.95	5	0.95
7	C	Step	46.9	0.386	0.343	0.483	0.517	1044	1289	*	19.9	5.90E-04	5	0.90	0	1.10
8	D	Step	50.1	0.399				2413		100	19.1	5.90E-04	20	0.40	0	0.40
9	E	Step	50.3	0.404				2434		86	21.2	3.44E-04	0	0.40	5	0.40
10	F	Step	50.2	0.4				2440		87	22.2	3.44E-04	0	0.40	5	0.40

(Tests 1-7 stand for the double-porosity medium (combination of sand and clay spheres) and Tests 8-10 for the single-porosity medium (pure sand). L is the length of the double-porosity medium column;  $w_1$  and  $w_2$  are the volumetric fractions of the sand (macro- porosity domain) and of the clay (micro-porosity domain), respectively;  $m_1$  and  $m_2$  are the masses of the sand and clay spheres, respectively;  $S_w = \langle \theta \rangle / \langle n \rangle$  is the average saturation degree,  $n_1$  and  $n_2$  are the porosities of the sand and clay spheres;  $T_{avg}$  is the average temperature; q is the imposed water flux (Darcy velocity);  $C_{ini} = C'_{ini} + C''_{ini}$  is the initial concentration where  $C'_{ini}$  is the intentionally imposed salt concentration of minerals in the solution of medium;  $C_{inj} = C'_{inj} + C''_{inj}$  is the applied concentration where  $C'_{inj}$  is the intentionally imposed salt concentration and  $C''_{inj}$  is the estimated concentration of minerals in the influent solution. For the tests of pulsewise type, the pulse duration is of  $t_0 = 4620$  s with a pulse concentration  $C_p = 5$  g/l. \* not available but the  $S_w$  of 86% of Test 2 can be assumed for Tests 3, 5, 6 and 7).

#### 2.2. Dispersion experiments

The experimental setup used for the different stages of the dispersion experiments (Fig. 2) in the double-porosity medium is presented in Tran Ngoc et al. (2011; 2014). The first stage concerns the establishment of saturated/unsaturated steady-state flow using a volumetric pump (Amersham Biosciences Pump P-500) for imposed inflow rate control and a balance linked to a data logger (Campbell Scientific Ltd CR 10X) for outflow rate control. The dispersion tests were only launched when an identical flow rate is obtained for both inflow and outflow. The inlet flux (Darcy velocity) was imposed from the bottom (except for Test 4) by using the volumetric pump to limit possible density-driven effect. For Test 4, the inlet flow was supplied by the pump from the top with an applied suction of h = -85 cm of H<sub>2</sub>O at the column bottom. This suction application induced an average water content of 0.298 for the entire medium (Fig. 3a). The soil columns were placed in the gamma ray mobile rig to measure locally the water contents inside the column (from which the saturation degree is deduced) before performing the dispersion stage. The gamma ray device consists of a radioactive <sup>241</sup>Am source, a scintillator and a photomultiplicator allowing to determine the water content with a precision of 3%. To obtain fully saturated samples for dispersion experiments (Tests 1 and 8), the initially dry 

To obtain fully saturated samples for dispersion experiments (Tests 1 and 8), the initially dry soil columns were saturated by carbon dioxide gas before establishing the flow. For other tests under unsaturated conditions, the  $CO_2$  was not injected into initially dry soil columns. The influx was entered from the column base without maintaining a low pressure at the outlet (Tests 2, 3, 5, 6 and 7 for the double-porosity columns and Tests 9 and 10 for the single-porosity columns). So, the particular unsaturated condition was attained due to air trapping in the medium (water content inside the medium is smaller than total porosity). Dead-end pores may only exist in the micro-porosity or immobile water domain (clay spheres), but this is an acceptable assumption for the dispersion in mobile-immobile systems when we consider no flow in this domain. The measured water content in this unsaturated condition can be considered as the effective porosity for the forthcoming dispersion experiments. Using the gamma attenuation technique, the saturated/unsaturated conditions were confirmed by comparison between water contents  $\theta$  and total porosities (equal to saturated water content  $\theta$ s) (Fig. 3).

The water content profiles measured inside the single- and double-porosity medium columns provided a clear indication on the unsaturated conditions for dispersion experiments, i.e. when  $\theta < \theta_S$ , Fig. 3. Here  $\theta$  can be considered as the effective porosity and  $\theta_S$  as the total porosity.

Note that the measured average  $\theta_S$  (0.3988) is in perfect agreement with the total porosity determined by weighting (0.3989) for Test 8 (single-porosity medium), while a difference of 3% for Test 1 (double-porosity medium) between the measured average  $\theta_S$  (0.353) and the total porosity (0.364) calculated by  $n_1 \times w_1 + n_2 \times w_2$  (see Table 2 for annotations of  $n_1$ ,  $n_2$ ,  $w_1$  and  $w_2$ ). Note also that the measured values of water contents in the vicinity of the top and bottom of the samples were biased due to diffracted gamma ray attenuation in these zones (Fig. 3). The distribution of the water contents is highly influenced by the presence of heterogeneity in the medium (i.e. clay spheres) (Fig. 3b), while it remains fairly uniform for single-porosity medium columns (Fig. 3a). It must be noted that in spite of using gamma ray attenuation technique, it is still very difficult to figure out the water re-distribution induced by the heterogeneities. For the double-porosity medium columns, the water content profiles are available only for Tests 1, 2 and 4.



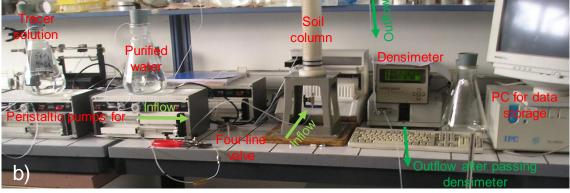
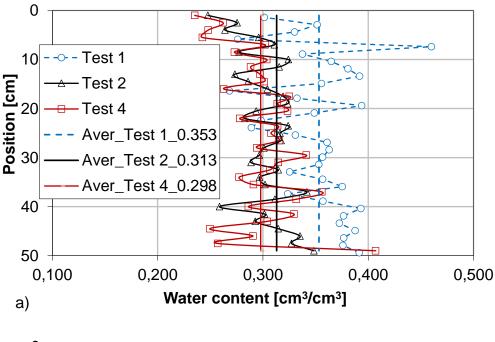


Fig. 2. Apparatus for the experiment: a) Flowing stage with water content measured by gamma ray technique; b) Dispersion stage.



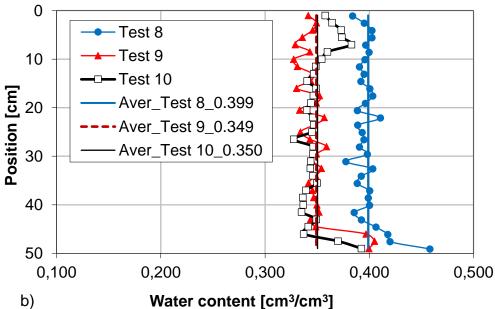


Fig. 3. Distribution of the water content inside the soil columns measured by the non-destructive technique of gamma ray attenuation after end of dispersion test. (The right lines represent for the average value of each test): a) for the double-porosity medium columns in saturated condition (Test 1) and unsaturated condition (Tests 2 and 4) and b) for the single-porosity medium columns (pure sand) in saturated condition (Test 8) and unsaturated condition (Tests 9 and 10) and (Data of these profiles were previously presented in Tran Ngoc et al., 2011; 2014).

The NaCl tracer dispersion was launched in the final stage by switching water to tracer solution or vice-versa, according to the selected boundary and initial conditions (water displaced or displacing tracer solution) (Fig. 2). In almost all dispersion tests (except Tests 1, 7 and 8), the higher concentration solution was used as displacing solution, while the smaller concentration solution was chosen as the displaced solution to avoid viscous fingering since at room temperature viscosity is lower for more concentrated solutions (Table 2). Two boundary conditions were used in this study: i) stepwise-like condition where an imposed salt concentration in the influent solution was continuously injected at the column inlet and ii) pulsewise-like condition where a concentration pulse was injected within a given duration at the column inlet. The later condition was obtained by using a four-line valve in order to switch the water pump into the NaCl solution pump (starting to inject the imposed concentration) and vice-versa (ending concentration pulse). After the tracer dispersion process, the column was moved to the gamma ray device to measure once again the water content. The water used in the experiments was purified by a reverse osmosis apparatus (Milli-ro 3 Plus) but some minerals possibly remained. The evolution of tracer concentration with time (breakthrough curves – BTC) was monitored by measuring the density of the effluents every 150 - 300 s by a densitometer (Anton Paar, mPDS 2000). During the tests, this densitometer recorded also the laboratory room temperature. The precision of the density and temperature is 0.005% and 1% of the measured values, respectively. A temperature difference of 0.1 °C might result in a density error of up to 0.0001 g/cm<sup>3</sup>.

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A series of dispersion tests were carried out in the model double-porosity medium. For the experiments under unsaturated conditions, the tests with stepwise-like and pulsewise-like brine injection were successively performed in the same double-porosity medium: Tests 2, 4 and 5 (same column B) and Tests 3, 6 and 7 (same column C) (Table 2). The NaCl concentration of the tracer solution was higher in the test under the saturated condition than in the tests under the unsaturated condition (Table 2). Considering that the high salt concentration used in the experiments might influence the dispersion behavior, especially for the experiments in the double-porosity medium, the high salt concentration (20 g/l) was used only for the two first tests (Tests 1 and 8 under saturated condition) in the beginning of our dispersion experiments campaign and was changed into low concentration (5 g/l) for other tests.

For the pulse-like tests (Tests 5 and 6), a pulse of the salt tracer solution equal to a volume of about 1/10 of the total pore volume (482 cm<sup>3</sup>) was injected in the medium during 4620 s. The main characteristics of all the dispersion tests are reported in Table 2. Parallel to the dispersion

1 tests in the double-porosity medium, another series of dispersion tests (Tests 8 - 10) were

performed in the single-porosity medium (pure sand) as a constituent of the double-porosity

3 medium, under the same saturation conditions as the double-porosity tests (Tran Ngoc et al.,

4 2011; 2014). Therefore, the required properties of the sand column for later numerical

5 simulations were independently identified.

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The mass balance of all the tests was verified by checking the fact that the sum of the masses of salt in the influent flow,  $M_1$ , and originally in the column,  $M_0$ , is equal to the sum of the masses of salt leaving the column,  $M_2$ , and present in the column at the end of a dispersion test,  $M_3$ . To do so, we first converted the BTCs in density into BTCs in concentration using the linear relation of density – concentration at the average temperature of each test, established from International Equation of State of Seawater 1980 (Unesco-IES80) (Fofonoff and Millard Jr., 1983; Le Calvé, 2002; Millero, 2010). We were conscious that this conversion contains errors but considered it acceptable as a first estimate. The difference between the density measured by the densimeter and calculated by the IES80 is about 0.0003 g/cm<sup>3</sup> when fixing the concentration within the range of 1 - 6 g/l (Fig. 4). The salt mass in the effluent  $M_2$  was then estimated at the final time of each test by using TableCurve® to calculate the area under BTC.  $M_0$ ,  $M_1$  and  $M_3$  were calculated by taking into account mineral mass (different from salt mass imposed intentionally) that was estimated from the beginning and ending density values in the BTCs,  $C_{\rm ini}$  and  $C_{\rm inj}$  (Table 3). The mass error is smaller for the single-porosity columns than for the double-porosity columns. It is very small for almost all tests (< 5%), (except for Test 4 and 7 where they were 9% and 7.5% respectively) and becomes even smaller when calculated for higher of  $S_w$ . Higher saturation values were explained by the fact that the medium was saturated a little more during long lasting dispersion experiments and by the measurement precision (3%). Thus, the estimations of the concentrations  $C_{\rm ini}$  and  $C_{\rm inj}$  by fitting the BTCs in regarding density values at early times are acceptable. This small difference comes from the physical nature of the salt retained in the micropores of the clay spheres. Indeed, retained salt diffuses very slowly and therefore leads to the tailing effect of BTCs. Another reason for the difference comes from the estimations of  $C_{\rm ini}$  and  $C_{\rm inj}$  (see Table 2) based on the mineral mass present in the used water and released by the ion exchange of clay during the experiments and detected by the densitometer. It could also be attributed to the conversion of the measured density into concentration that was impacted by the temperature variations during dispersion tests.

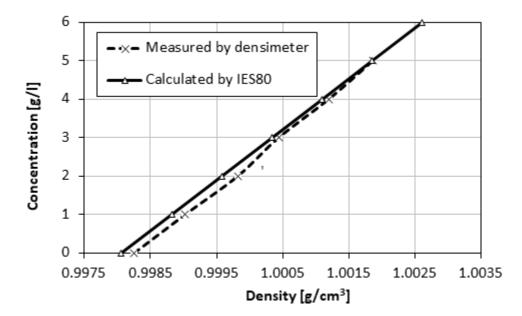


Fig. 4. Comparison of Concentration – Density between the measure and theory at the temperature of  $20.7\,^{\circ}\text{C}$ .

*Table 3. Salt mass balance calculated for the experiments.* 

Test	$M_0[g]$	$M_1$ [g]	$M_2$ [g]	<i>M</i> <sub>3</sub> [g]	Mass error = $   (M_0+M_1) - (M_2+M_3)   $ [g]	$\Delta M = 100\%$ Mass error/ $(M_0 + M_1)$ [%]
1	10.828	1.265	11.414	0.466	0.212	1.76
2	0.404	7.124	4.672	2.649	0.207	2.75
3	0.372	6.109	3.887	2.436	0.158	2.44
4	2.520	0.885	3.402	0.309	0.307	9.01
5	0.729	1.419	1.627	0.597	0.076	3.54
6	0.392	1.561	1.551	0.373	0.030	1.51
7	2.728	1.166	3.769	0.416	0.292	7.49
8	11.524	0.5804	11.953	0.226	0.074	0.61
9	0.202	4.379	1.663	2.732	0.186	4.06
10	0.199	4.536	2.024	2.683	0.028	0.58

(Tests 1-7 stand for the double-porosity medium columns and Tests 8-10 for the single-porosity medium columns.  $M_1$  and  $M_2$  are the salt masses of the influent and effluent solutions;  $M_0$  and  $M_3$  are the salt mass present initially (before dispersion stage) and finally (end of dispersion test) in the medium.  $M_2$  was calculated by the area below the BTC averaged from the experimental data).

# 3. Modelling and dispersivity determination

# 3.1. Governing equations

- 1 The BTCs of the brine dispersion experiments in the single-porosity medium columns (Tests 8
- 2 10) can be modelled by using the conventional Convection-Dispersion transport Equation
- 3 (CDE), written as (Bear, 1972; van Genuchten and Wierenga, 1976; Kutılek and Nielsen,
- 4 1994):

$$5 \quad \frac{\partial c}{\partial t} = D_{SP} \frac{\partial c}{\partial z} - v \frac{\partial c}{\partial z} \tag{1}$$

- 6 where C [ML<sup>-3</sup>] is the salt concentration in the flowing solution;  $\theta$  [-] is volumetric water
- 7 content ( $S_w = 100\% \times \theta/n$ , with *n* porosity);  $D_{SP}$  [L<sup>2</sup>T<sup>-1</sup>] is dispersion coefficient in the liquid
- 8 phase; v [LT<sup>-1</sup>] is pore velocity ( $v = q/\theta$ , q the imposed Darcy velocity); z [L] is the position or
- 9 depth in the soil column and t [T] is the time variable. The subscript SP stands for Simple
- 10 Porosity vs DP used below for Double Porosity porous media. The dispersivity can be written
- 11 as (Bear, 1972):

$$12 \quad \lambda_{SP}(S_w) = \frac{D_{SP}(S_w) - D_e}{v^{\gamma}} \tag{2}$$

- In common practice for the dispersion in soils, the effective diffusion coefficient  $D_{\rm e}$  [L<sup>2</sup>T<sup>-1</sup>], is
- 14 considered negligible and  $\gamma$  is taken equal to 1, i.e. linear increase of  $D_{SP}$  with  $\nu$ . Thus,

$$15 \quad \lambda_{SP}(S_w) = \frac{D_{SP}(S_w)}{v} \tag{3}$$

- 16 For the dispersion experiments in double-porosity medium columns, the BTCs exhibited non-
- 17 Fickian features marked by an early arrival time and a tailing of salt concentration evolution.
- 18 A satisfactory description of this non-standard transport could not be obtained by applying the
- 19 CDE model (van Genuchten and Wierenga, 1976; Berkowitz et al., 2000; Golfier et al., 2011,
- 20 Simhayov et al., 2018, among others). These BTCs can be appropriately described using two-
- 21 equation models (mobile-immobile MIM) obtained from phenomenological (van Genuchten
- and Wierenga, 1976) or upscaling approaches (Ahmadi et al., 1998; Royer and Boutin, 2012).
- 23 In these models, transport is governed by the CDE in the macro-pores and CDE or diffusion
- 24 equation in the micro-pores. Here, we adopt the macroscopic dispersion convection model
- obtained in Tran Ngoc et al. (2011), by applying the asymptotic homogenization technique
- 26 (Auriault, 1991). This choice is motivated by the fact that this model was developed for the
- validation range of the experimental dispersion tests in the double- porosity medium presented
- above with periodic microstructure (Tran Ngoc et al., 2011) and we do not need to fit solute
- 29 transfer coefficients between the macro- and micro-pores that are computed integrating all

- 1 diffusive flux through surfaces of the micro-pore domain. In this model, the flow occurs only
- 2 in the macro-pores (sandy matrix domain 1 denoted  $\Omega_1$ ) and is stagnant in the micro-pores
- 3 (clay spherical inclusions domain 2 denoted  $\Omega_2$ ). This resulted from modelling water flow in
- 4 an unsaturated double-porosity medium (Lewandowska et al., 2004). The transport in the
- 5 domain 1 is described by the dispersion and convection mechanisms, while only diffusion is
- 6 considered in the domain 2. The model is re-written for simulating the dispersion experiments
- 7 mentioned above by a 1D macroscopic model coupled with a 3D problem for diffusion in the
- 8 inclusions that is converted into a 1D problem in a spherical coordinate system. It is given by
- 9 the following system of equations:

10 
$$w_1 \frac{\partial(\theta_1 C_1)}{\partial t} = \frac{\partial}{\partial z} \left( D_{DP} \frac{\partial}{\partial z} C_1 \right) - \frac{\partial}{\partial z} (qC_1) - S$$
 (4)

11 
$$\frac{\partial(\theta_2 C_2)}{\partial t} = D_2 \left( \frac{\partial^2 C_2}{\partial r^2} + \frac{2}{r} \frac{\partial C_2}{\partial r} \right)$$
 in the domain 2 (5)

12 
$$C_1 = C_2$$
 on the interface  $\Gamma$  between the two domains (6)

13 
$$S = \frac{1}{|\Omega|} \int_{\Gamma} \frac{\partial(\theta_2 C_2)}{\partial r} d\Gamma$$
 (7)

- where the subscripts "1" and "2" stand for the domains 1 ( $\Omega_1$ ) and 2 ( $\Omega_2$ ) respectively ( $\Omega$  =
- 15  $\Omega_1 + \Omega_2$ ;  $C_1$  and  $C_2$  [ML<sup>-3</sup>] are the concentration in the domain 1 and 2, respectively;  $w_1$  [-] is
- 16 the volumetric fraction of the domain 1; q [LT<sup>-1</sup>] is the imposed Darcy velocity; z ( $0 \le z \le L$ )
- and r ( $0 \le r \le L$ ) [L] are spatial coordinates;  $D_{DP}$  [L<sup>2</sup>T<sup>-1</sup>] is the dispersion coefficient of the
- double-porosity medium;  $D_2$  [L<sup>2</sup>T<sup>-1</sup>] is the effective diffusion coefficient in the domain 2. Note
- that Eq. (4) could be returned into Eq. (1) if considering  $D_{SP} = D_{DP}/\theta$ , because of the
- 20 calculation definition difference of  $D_{SP}$  and  $D_{DP}$ .
- 21 8In this study the dispersivity of the double-porosity medium is defined by (Bear, 1972):

$$22 \lambda_{DP}(S_W) = \frac{D_{DP}(S_W)}{q} (8)$$

- In Eq. (), we assumed the Darcy velocity q equal to the macroscopically averaged linear
- velocity relevant for the dispersion coefficient  $D_{DP}$  defined over all the porous medium in the
- double-porosity model (see detail in Tran Ngoc et al, 2011). Compared with the single-porosity
- 26 medium (pure sand column), the dispersivity is determined by the ratio of the dispersion
- coefficient in liquid phase and pore velocity (Eq. 3).

- 1 The dispersion coefficients  $D_{SP}$  and  $D_{DP}$  of the dispersion tests at different water saturations
- 2 were determined by fitting and validating the experimental results with numerical solutions of
- 3 the above models that are presented after herein.

#### 4 3.2. Numerical simulations

# 5 **3.2.1. Single-porosity medium**

- 6 The transport parameters of the dispersion experiments in the single-porosity medium columns,
- 7 i.e. Tests 8 10, were found by fitting experimental BTCs with solutions of Eq. (1) using the
- 8 CXTFIT code (Toride et al., 1999). To do so, the experimental BTCs in terms of density were
- 9 first converted into curves in terms of relative concentration by the following relation:

$$10 \quad \frac{C(t)}{C_{ini}} = \frac{\rho(t) - \rho_{min}}{\rho_{max} - \rho_{min}} \tag{9}$$

- where  $\rho(t)$ ,  $\rho_{min}$  and  $\rho_{max}$  [M/L<sup>3</sup>] are the measured densities with time, minimal and maximal
- values of density in the BTCs, respectively.
- Only  $D_{SP}$  was determined to avoid non-uniqueness of solution in the inverse problem, since the
- pore velocity v could be calculated from controlled water contents. However, various water
- 15 contents were tested in the range of measured value by Gamma ray of  $0.399 \pm 3\% \times 0.399$  for
- 16 Test 8,  $0.350 \pm 3\% \times 0.350$  and  $0.350 \pm 6\% \times 0.350$  for Tests 9 and 10, due to measurement
- precision to get the best fit. To reproduce numerically Tests 8 10,  $D_{SP}$  values obtained from
- the above inverse problem were used in direct simulations by HYDRUS (Šimůnek et al., 2013)
- 19 with the following initial and boundary conditions:

20 
$$0 < z < L$$
:  $C(z, 0) = C_{ini}$  (10)

$$21 z = 0, t > 0: vC - D_{SP} \frac{\partial C}{\partial z} = vC_{inj} (11)$$

$$22 z = L, t > 0: D_{SP} \frac{\partial C}{\partial z} = 0 (12)$$

- 23 While the profile discretization was of 100 elements for Test 8, the profile was finely
- 24 discretized into 300 elements for Tests 9 and 10 in order to avoid numerical oscillations of
- simulation BTC. To compare with the experimental results, the evolution of the concentration
- 26 with time at the outlet element of the simulation domain was exported. The simulated BTCs in
- 27 concentration were then converted into the BTCs as a function of density to be confronted
- 28 directly with the measured BTCs. To this end, we used the function giving the density  $\rho$  as a

- 1 function of the concentration C, the pressure P and the temperature T from the International
- 2 Equation of State of Seawater 1980 (Unesco-IES80) (Fofonoff and Millard Jr., 1983; Le Calvé,
- 3 2002; Millero, 2010), taking into account the temperature data recorded simultaneously with
- 4 the density in each test and the pressure value of 1.01325 MPa (equivalent to the case of
- 5 standard seawater at 1 atm). The standard error related to this conversion is about  $3 \times 10^{-6}$  g/cm<sup>3</sup>
- 6 (Millero, 2010). The dispersivity was determined for a saturated case (Test 8) and two
- 7 unsaturated cases (Tests 9 and 10) from Eq. (3). The simulation quality indicated by statistical
- 8 criteria such as the sum of squared residuals  $SSQ = \sum (\rho_{mes} \rho_{sim})^2$ , the root mean square
- 9 error  $RMSE = \sqrt{\frac{\sum (\rho_{mes} \rho_{sim})^2}{N}}$  which minimize the error and the coefficient of determination
- 10  $R^2 = \frac{\left[\sum(\rho_{mes} \overline{\rho_{mes}})(\rho_{sim} \overline{\rho_{sim}})\right]^2}{\sum(\rho_{mes} \overline{\rho_{mes}})^2\sum(\rho_{sim} \overline{\rho_{sim}})^2}$  which tends to 1, where  $\rho_{mes}$  and  $\rho_{sim}$  are the measured and
- simulated values of density respectively;  $\overline{\rho_{mes}}$  and  $\overline{\rho_{sim}}$  are the average values of measured
- and simulated density; and *N* is the number of values.

## 3.2.2. Double-porosity medium

- 14 The macroscopic model given by Eqs (4) (7) was numerically implemented by using the FEM
- 15 code COMSOL Multiphysics®. The implementation, detailed in Tran Ngoc et al. (2011; 2020),
- 16 concerns a two-scale computation for the macroscopically 1D concentration (Eq. 4) coupled
- with the microscopically 1D concentration (Eq. 5, spherical transport) on the interface (Eq. 6).
- 18 The coupling term is the source term in Eq. (4) which is calculated from the diffusive flux on
- 19 the interface of all the microscopic domain (clay spheres) (Eq. 7) by resolving Eq. (5). To do
- so, a 1D geometry (L) representing the length of the double-porosity medium column in which
- 21 the solute transport is described by Eq. (4), was first created (one independent variable z) in
- 22 COMSOL. Then, a 2D geometry  $(L \times R)$  representing a series of the spheres of radius R
- 23 (microscopic domain) side by side along the axis of length L was created with two independent
- variables z and r. Although the created geometry is 2D, the transport is considered by Eq. (5)
- 25 in one-dimensional only in the R direction (radial problem) for all z positions. Through the
- 26 interface (surface of the sphere), each node of the macroscopic 1D geometry is linked with a
- sphere in which the diffusion occurs (in the R direction). The macroscopic domain L was
- meshed into 100 grid blocks of the size  $0.5 \times 10^{-2}$  m, thus 101 nodes. After a mesh sensitivity
- analysis ensuring grid convergence of the results, the microscopic domain  $L \times R$  was discretized
- very finely with 1504 grids (1300 quadrilateral elements and 204 triangular elements). The
- 31 meshing was regular in the z axis while it was highly refined near r = R where the coupling
- 32 with the macroscopic equation is imposed by the maximum element growth rate of 1.3 and

maximum and minimum element size of  $3.38 \times 10^{-2}$  m and  $1.51 \times 10^{-4}$  m, respectively. So that, 1 the mesh distribution consisted of two zones: a zone of 1300 quadrilateral elements (closer to 2 3 the interface r = R smaller elements) and 204 triangular elements (for the farthest elements 4 from the interface). The simulation result remained unchanged for other finer meshes. To solve 5 Eqs (4) - (7), an appropriate set of initial and boundary conditions that correspond to the 6 experimental conditions are required. Table 4 presents the initial and boundary conditions for 7 all the dispersion tests. Note that a third-type boundary condition was used to prescribe the 8 concentration flux at the inlet boundary for better reproduction of the dispersion tests rather 9 than the concentration boundary condition of the first-type (van Genuchten and Parker, 1984; 10 Schoen et al., 1999; Cushman and Tartakovsky, 2016).

The conversion of concentration in the BTCs into density versus time was integrated in the numerical simulations to allow direct comparison with the experimental BTCs which are impacted by the temperature. As for the BTCs of the single-porosity medium, this conversion was made via the Unesco-IES80.

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Table 4. Initial and boundary conditions for the numerical simulation of the dispersion experiments.

		Eq. 4	Eq. 5				
Type	Initial	Boundary	Initial	Boundary			
	condition	condition	condition	condition			
Stepwise tests (1 – 4 and 7)		$z = 0, t > 0:$ $qC_1 - D_{DP} \frac{\partial C_1}{\partial z} = qC_{inj}$ $z = L, t > 0:$ $D_{DP} \frac{\partial C_1}{\partial z} = 0$		r = 0, t > 0:			
Pulsewise tests (5 and 6)	$C_1 = C_{ini}$	$z = 0, t \le t_0:$ $qC_1 - D_{DP} \frac{\partial C_1}{\partial z} = qC_p$ $z = 0, t > t_0:$ $qC_1 - D_{DP} \frac{\partial C_1}{\partial z} = qC_{inj}$ $z = L, t > 0:$ $D_{DP} \frac{\partial C_1}{\partial z} = 0$	$C_1 = C_{ini}$	$D_2 \frac{\partial C_2}{\partial r} = 0$ $r = R, t > 0:$ $C_2 = C_1$			

(For annotation of  $C_{\text{ini}}$ ,  $C_{\text{inj}}$  and  $C_{\text{p}}$  refer to Table 2).

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- 2 The input data required for the numerical simulations were taken from Table 2. The diffusion 3 coefficient in the domain 2 was considered equal to that of a saturated brick,  $D_2 = 7.345 \times 10^{-1}$ <sup>7</sup> cm<sup>2</sup>/s (Ahl, 2003). The water content in the domain 2 (very low hydraulic conductivity) of 4 the double-porosity medium was set at the value  $\theta_2 = n_2 = 0.343$  for all the tests. Because no 5 6 flow occurs in the domain 2 in which the water was only absorbed and retained, thus the 7 saturated value could be reached during the long time of experiments. The "trial and error" approach was used to determine the dispersion coefficient  $D_{\rm DP}$  of the double-porosity medium 8 9 by fitting the experimental BTCs with those obtained by the double-porosity model (Eqs. 4 – 7). The corresponding dispersivity is then determined by Eq. 8. For Test 1 (saturated condition), 10 11 based on the statistical indicators, i.e. SSQ,  $R^2$  and RMSE, the fitted  $D_{DP}$  was chosen from different simulations made with different water contents in the domain  $1, \theta_1$  and different 12 dispersion coefficients  $D_{DP}$ . The initial value of  $D_{DP}$  was the dispersion coefficient found in the 13 saturated single-porosity medium, Test 8. A series of simulations were run by coupling  $\theta_1$  = 14 15 0.364; 0.385 and 0.408 (deduced from the average value measured by the Gamma ray technique  $<\theta>_{\text{mes}} = 0.354; <\theta>_{\text{mes}} + 3\% <\theta>_{\text{mes}} (0.364) \text{ and } <\theta>_{\text{mes}} + 6\% <\theta>_{\text{mes}} (0.374), \text{ respectively)}$ 16 with  $D_{DP} = 2.713 \times 10^{-4}$ ; 1.899 ×10<sup>-4</sup> (decreased by 30%); and 3.527 ×10<sup>-4</sup> (increased by 30%) 17 cm<sup>2</sup>/s. We took the case of  $\theta_1 = 0.385$  (very close to porosity value, Table 2),  $\theta_2 = 0.343$ ,  $D_{DP}$ 18 = 2.713  $\times 10^{-4}$  cm<sup>2</sup>/s and  $D_2 = 7.345 \times 10^{-7}$  as the base case. Knowing that the diffusion 19 coefficient  $D_2$  plays also a key role in the non-Fickian BTC, we made a sensitivity analysis 20 21 from the base case with increasing and decreasing  $D_2$  by the factor of 2 and 5.
- For Test 2, the unsaturated condition was controlled by the macro-porosity domain (Domain 1). Therefore,  $\theta_2$  and  $D_2$  were considered the same as those of the base case in Test 1. The numerical reproduction of Test 2 was done for  $\theta_1$  in the range of water content measured in the single-porosity medium columns (Tests 9 and 10) by gamma ray technique and  $D_{DP}$  in the range of dispersion coefficient fitted for the single-porosity medium columns (Tests 9 and 10) and fitted for the double-porosity medium (Test 1).
  - All the parameters chosen for Test 2, including the obtained dispersivity from the above fitting were used directly in the simulations for Tests 3, 5, 6 and 7. These simulations can be referred to as the validating stage for the obtained dispersivity values. For Test 4,  $\theta_1 = 0.249$  was deduced from the average value of the measured water content profile (Fig. 3a) by  $\langle \theta \rangle = \theta_1 \times \theta_2$

- 1  $w_1 + \theta_2 \times w_2$ . The fitting was done by "trial and error" to determine  $D_{DP}$  for Test 4 by starting
- 2 with the dispersion coefficient of Test 2.

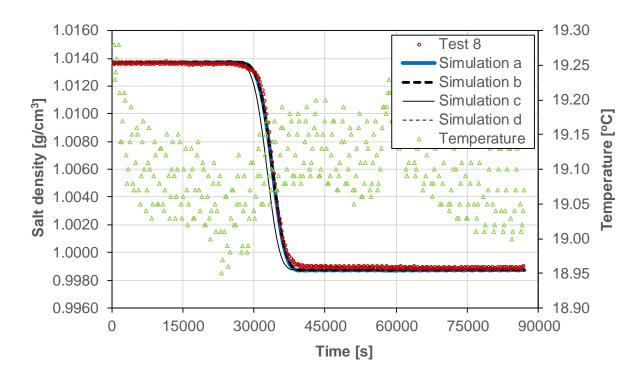
#### **4. Results and discussion**

# 4 4.1. Saturated and unsaturated dispersivity in the single-porosity medium

- 5 The saturated dispersivity for the single-porosity sample in Test 8 is 0.089 cm as the best fit
- from Simulation d (Fig. 5; Table 5) with  $\theta = 0.401$  = porosity determined by gravity, which is
- 7 very close to measured value by Gamma ray (Fig. 3b). The corresponding column Péclet
- 8 number is Pe = 1.5 (Table 5). This dispersivity is in the order of 0.01 0.1 cm found for
- 9 saturated unconsolidated materials in the literature (Padilla et., 1999; Nutzmann et al. 2002;
- 10 Toride et al., 2003; Torkzaban et al., 2006).
- 11 As known the temperature variation influences strongly the density of the outlet solution.
- 12 However, it is difficult to separate its influence due to transport mechanisms during the
- evolution of BTCs, especially at low temperature variation such as in Test 8 (Fig. 5). Normally,
- we can only observe that an increase of the temperature leads to the diminution of the density
- 15 for a stage of transport, earlier times at which the concentration does not break through yet, for
- example (Fig. 6). In this work, the integration of the temperature in conversing of the BTCs in
- 17 concentration into BTCs in density helps to minimize the influence of its variation for fitting
- the dispersion coefficients.
- 19 It should be noted that the BTCs in the single-porosity medium correspond to the Fickian
- 20 transport, i.e. fairly symmetrical BTCs not only in the saturated case (Test 8, Fig. 5), but also
- 21 in the unsaturated case (Tests 9 and 10, Fig. 6), thanks to their uniform saturation profiles (Fig.
- 22 3b). This was observed in Sato et al. (2003) for transport in uniformly unsaturated single-
- 23 porosity medium columns. However, the non-Fickian transport has been often observed for
- 24 unsaturated homogeneous soils (Padilla et., 1999; Nutzmann et al., 2002; Toride et al., 2003;
- 25 Bromly and Hinz, 2004). The best fit of the unsaturated dispersivity is 0.044 cm (Test 9) and
- 26 0.041 cm (Test 10) with corresponding  $\theta = 0.370$  and  $\theta = 0.368$  (Simulation d, Fig. 6 a and b,
- Table 5). The fitted water contents are greater than the measured ones by 6% for Test 9 and by
- 28 5% for Test 10 (Fig. 3b). It can be observed that the saturated dispersivity (Test 8) is greater
- 29 than the unsaturated dispersivity (Test 9 and 10). However, some works in the literature showed
- an increase of dispersivity with a decrease of saturation (Padilla et., 1999; Nutzmann et al.
- 31 2002; Sato et al., 2003). This contradiction cannot be definitively confirmed due to the lack of
- 32 other experiments with low saturation profiles. Moreover, it must be underlined that the

- 1 saturated and unsaturated uniform flow condition in this study occurred for particular values
- of the total porosity ( $\theta = 0.399$ ) and the effective porosity ( $\theta = 0.350$ ) respectively (Fig. 3b).
- 3 From our results, it can be concluded that the dispersivity is greater when transport takes place
- 4 in the total porosity of a single-porosity porous medium than when it takes place in a part of
- 5 the porosity i.e. the effective porosity.

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Fig. 5. Breakthrough curves obtained different simulations (see details for simulation parameters in Table 4) (lines) and the experiment (circles) with the associated temperature data (triangles) for Test 8 (single-porosity medium, saturated case). Converting the density evolution with time into the relative concentration,  $C/C_{ini}$  with pore volume exit time,  $V/V_p$  we have ratio  $C/C_{ini} = 0.5$  at  $V/V_p = 1$ .

1 Table 5. Fitting parameters and errors for different simulations of dispersion tests in the single-porosity medium. (The simulation case leading to the best fit was put in italic)

Experiment	Simulation	$\theta$	q [cm/s]	v [cm/s]	$D_{\rm SP}$ [cm <sup>2</sup> /s]	λ [cm]	$R^2$	SSQ	RMSE
	a	0.399	5.90E-04	1.48E-03	1.360E-04	0.092	0.99922	2.254E-05	2.495E-04
Tost 9	b	0.411	5.90E-04	1.44E-03	1.410E-04	0.098	0.99917	2.317E-05	2.530E-04
Test 8	c	0.387	5.90E-04	1.52E-03	1.500E-04	0.098	0.98650	2.739E-04	8.699E-04
	d	0.401	5.90E-04	1.47E-03	1.310E-04	0.089	0.99925	2.227E-05	2.481E-04
	a	0.349	3.44E-04	9.85E-04	7.710E-05	0.078	0.93334	6.088E-05	4.679E-04
Test 9	b	0.360	3.44E-04	9.56E-04	3.550E-05	0.037	0.96915	2.864E-05	3.210E-04
Test 9	c	0.339	3.44E-04	1.02E-03	1.910E-04	0.188	0.88832	1.063E-04	6.182E-04
. <u> </u>	d	0.370	3.44E-04	9.30E-04	4.044E-05	0.044	0.99049	9.377E-06	1.361E-04
	a	0.350	3.44E-04	9.82E-04	5.890E-05	0.060	0.949225	4.250E-05	1.361E-04
Test 10	b	0.361	3.44E-04	9.53E-04	3.920E-05	0.041	0.987510	1.076E-05	3.953E-04
	c	0.340	3.44E-04	1.01E-03	1.860E-04	0.184	0.909419	7.610E-05	1.989E-04
	d	0.368	3.44E-04	9.35E-04	3.839E-05	0.041	0.995449	5.148E-06	1.055E-04

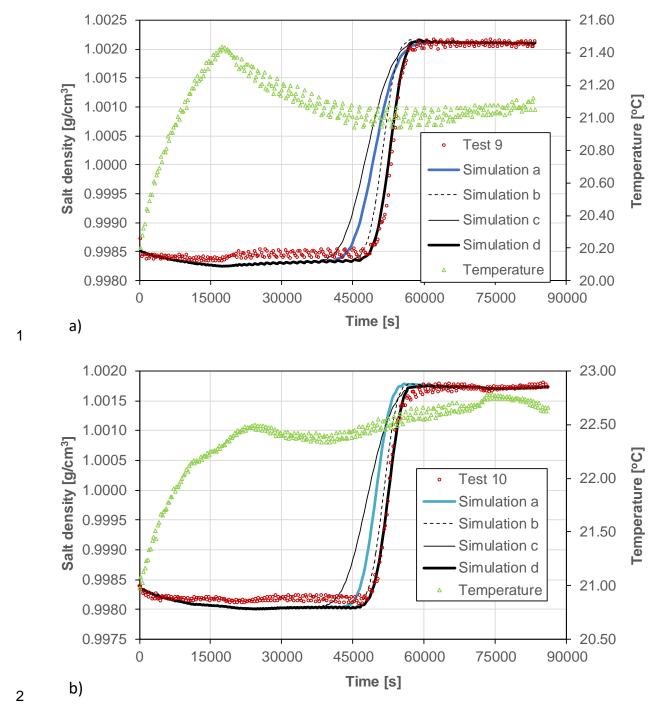


Fig. 6. Breakthrough curves obtained different simulations (see details for simulation parameters in Table5) (lines) and the experiment (circles) with the associated temperature data (triangles) for Tests 9 (a) and 10 (b) (single-porosity medium, unsaturated case). Converting the density evolution with time into the relative concentration,  $C/C_{ini}$  with pore volume exit time,  $V/V_p$  we have ratio  $C/C_{ini} = 0.5$  at  $V/V_p = 0.89$ .

#### 4.2. Saturated dispersivity in the double-porosity medium

Both the experimental BTCs in terms of density and the temperature as a function of time are presented in Fig. 7 for three different simulations of Test 1 which corresponds to the saturated

case. The simulation parameters of the best fits are reported in Table 6. It must be noted that the clay spheres where sintered at a very high temperature ( $1000^{\circ}$ C) to make them resistant to aggregate dispersion and breakdown even for rather high NaCl concentrations (20 g/l). Also, compared to other tests (Tests 2-7) with a lower NaCl concentration (5 g/l), no particular dispersion behavior was detected in the BTC of Test 1. However, there is a slight discrepancy for the period of time t=0-15000 s with 0.5 °C of temperature variation, in spite of taking into account the correction of the density by temperature. In fact, the density for the period t=0-15000 s could have been impacted not only by the room temperature, but also by other factors such as ion exchange of clayey spheres and/or the existence of minerals in the displacing solution (water). The effects of these factors are superimposed during the solute transport, so that it is very difficult to obtain a better fit than the one presented in Fig. 7. It should be noted that the "best" fits chosen in this study ensured to cover the non-Fickian features of experimental BTCs.

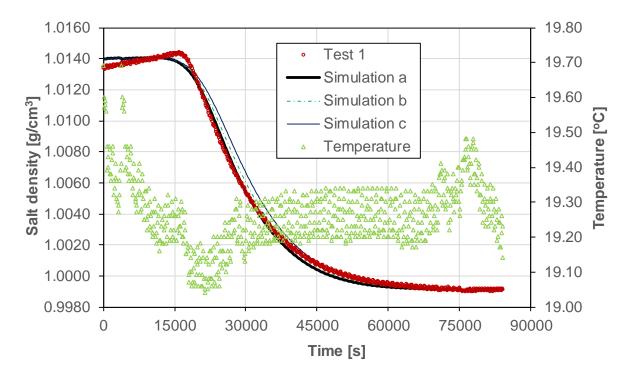


Fig. 7. Breakthrough curves obtained from different simulations (see details for simulation parameters in Table 6) (lines) and the experiment (circles) with the associated temperature data (triangles) for Test 1 (double-porosity medium, saturated case).

Table 6. Fitting parameters and errors for different simulations of dispersion tests in the double-porosity medium. (The simulation case leading to the best fit was put in italic)

Experiment	Simulation	< 0>	$\theta_1$	$D_2$ [cm <sup>2</sup> /s]	$D_{\mathrm{DP}}[\mathrm{cm}^2/\mathrm{s}]$	λ [cm]	$R^2$	SSQ	RMSE
	a	0.353	0.364	7.345E-07	1.899E-04	0.322	0.99821	5.634E-05	2.831E-04
	b	0.364	0.385	7.345E-07	2.713E-04	0.460	0.99729	8.173E-05	3.440E-04
	c	0.374	0.408	7.345E-07	3.527E-04	0.598	0.99417	2.070E-04	5.426E-04
Так 1	d	0.364	0.385	1.469E-07	2.713E-04	0.460	0.95640	1.302E-03	1.361E-03
Test 1	e	0.364	0.385	3.673E-06	2.713E-04	0.460	0.97269	1.016E-03	1.202E-03
	f	0.364	0.385	3.673E-07	2.713E-04	0.460	0.99513	1.615E-04	4.793E-04
	g	0.364	0.385	1.469E-06	2.713E-04	0.460	0.98791	4.285E-04	7.807E-04
	h	0.364	0.385	7.345E-06	2.713E-03	4.598	0.99340	1.907E-04	5.208E-04
Test 2	a	0.346	0.350	7.345E-07	1.662E-04	0.483	0.99677	5.195E-06	1.119E-04
	b	0.349	0.355	7.345E-07	2.374E-04	0.690	0.99614	5.489E-06	1.150E-04
	c	0.352	0.360	7.345E-07	3.086E-04	0.897	0.99517	6.39E-06	1.329E-04

In this case, the dispersion coefficient is equal to  $1.899 \times 10^{-4}$  cm<sup>2</sup>/s chosen by fitting the 1 experimental BTC (Simulation a). The quality of this fitting can be evaluated by the criteria 2 3 reported in Table 4. For the same Darcy velocity imposed, the obtained  $D_{\rm DP}$  of Test 1 is greater 4 than the D<sub>SP</sub> of Test 8. Regarding the Péclet number of 4.03 (of the order of 1) for Test 1 5 calculated using the period size, the Darcy velocity and the dispersion coefficient and Pe = 1.546 for Test 8, the linear relation between the dispersion coefficient and Péclet number is respected 7 (Fetter, 2008). Thus, the saturated dispersivity of 0.322 cm obtained is of the same order of 8 magnitude as the heterogeneity size in the double-porosity medium, i.e. the diameter of the 9 clay spheres of 0.64 cm. It is greater than the dispersivity (0.089 cm) found for the saturated 10 single-porosity medium (Test 8, pure sand, one of two constituents of the double-porosity 11 medium). The heterogeneity (clay spheres) presence in the double-porosity medium is 12 responsible for the increase of the dispersivity (Silliman and Simpson, 1987). So far, it can be 13 clearly seen that the dispersivity of the single-porosity medium cannot be used as the 14 dispersivity of the double-porosity medium in spite of having the same macro-porosity. The fitting of the BTC for the double-porosity medium columns involves two transport 15 16

The fitting of the BTC for the double-porosity medium columns involves two transport properties  $D_{DP}$  and  $D_2$ , which may lead to non-uniqueness of the set of coefficients determined. Therefore, only  $D_{DP}$  was fitted in this study, thus fixing  $D_2$ . Based on the base case, a sensitivity analysis to the variation of  $D_2$  within the possible range of  $1 \times 10^{-7} - 1 \times 10^{-6}$  cm<sup>2</sup>/s for solidified clay (Jougnot et al., 2009) was conducted by performing different simulations (d - g), Fig. 8 and Table 6. Comparing BTCs for different  $D_2$ , the value of  $7.345 \times 10^{-7}$  cm<sup>2</sup>/s corresponds to the best fit between the simulation and the experiment (Fig. 7). Note that the simulation only enables to reproduce the non-Fickian feature of experimental BTC if  $D_{DP}$  and  $D_2$  is enough contrasted (Simulation h).

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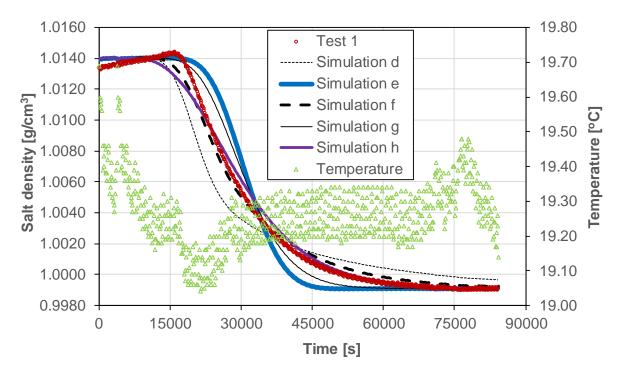


Fig. 8. Breakthrough curves obtained from sensitivity analysis of the salt diffusion coefficient in the clay spheres  $D_2$  (see details for simulation parameters in Table 6) (lines) and the experiment (circles) with the associated temperature data (triangles) for Test 1 (double-porosity medium, saturated case).

# 4.3. Unsaturated dispersivity in the double-porosity medium

For the unsaturated case, the simulations results fitted the experimental ones in a satisfactory manner using  $\theta_I = 0.350$ . This value of the water content leads to  $S_w$  of about 90%, slightly greater than the measured value of  $S_w = 86\%$  and consistent with the mass balance. For Test 2, at the early times of the BTC, the simulation respecting the principle of the density increase with decrease of temperature could not reproduced correctly the experiment at which a high density with a temperature drop was observed (Fig. 9). Together with other reasons aforementioned, this can be explainable by the fact of not complete solute washing before the dispersion test. The initial concentration was not possibly uniform along the soil column. It might be greater at the upper part of medium profile (close to the outlet column). The estimation of the initial concentration for Test 2 (0.9 g/l, Table 1) was not totally satisfactory to simulating this behavior, but globally acceptable to fit the coefficient dispersion.

The dispersivity  $\lambda_{\rm DP} = 0.483$  cm was determined from  $D_{\rm DP} = 1.662 \times 10^{-4}$  cm<sup>2</sup>/s, corresponding to the best fit with the experimental BTC of Test 2 (stepwise test) in the analyzed range of  $\theta_{\rm I}$  and  $D_{\rm DP}$ , (Fig. 9, Table 6). This dispersivity was used in direct numerical simulations for other tests: Test 3 (different macroscopic length), Test 7 (different velocity), Tests 5 and 6 (different

initial and boundary conditions – pulsewise test). The Péclet number and all statistical indicators were listed in Table 7 for all tests. For the same Darcy velocity, the greater value of Pe for the double-porosity medium vs for the single-porosity medium corresponds to a greater  $D_{\rm DP}$  vs  $D_{\rm SP}$ . It is observed that the numerical simulations reproduced very well the experimental BTCs of the tests (Fig. 10 and 11). For the same Darcy velocity and saturation, the breakthrough time,  $t \sim 30000$  s, was longer in Tests 2 (Fig. 9) and 5 (Fig. 11 a) than in Tests 3 (Fig. 10 a) and 6 (Fig. 11 b) where  $t \sim 18000$  s. If we refer the saturated case (Test 1) and unsaturated cases (Tests 2, 3, 5, 6 and 7) to the total and effective porosity, respectively, contrary to the single-porosity medium (Tests 8-10), the dispersivity is smaller when transport takes place in the total porosity of the sample compared to the case where only a part of the porosity (effective porosity) is explored by the salt solution for the double-porosity medium due to the presence of heterogeneity. To our best knowledge, for the first time, there exists a unique dispersivity value at a given saturation which validates different scenarios of transport in a double-porosity medium, regardless of flow and transport conditions.

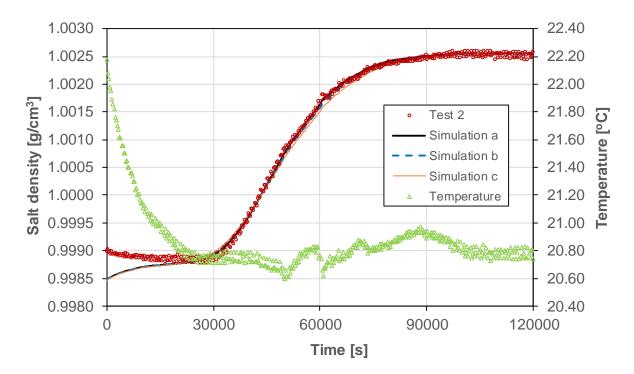
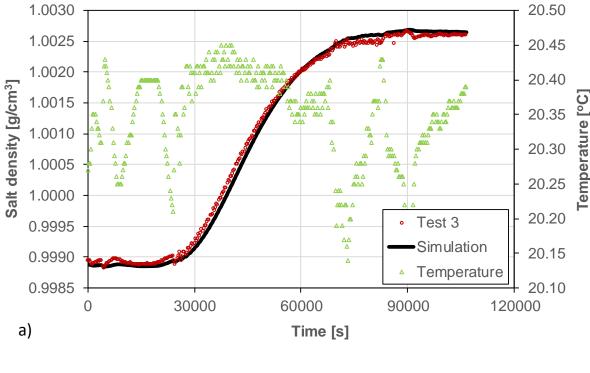


Fig. 9. Breakthrough curves obtained from different simulations (see details for simulation parameters in Table 6) (lines) and the experiment (circles) with the associated temperature data (triangles) for Test 2 (double-porosity medium, case of 86% saturation, fitting stage).



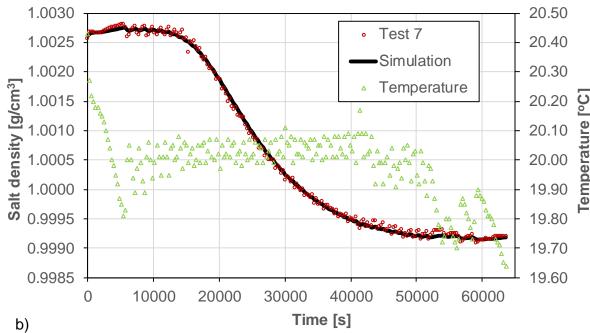
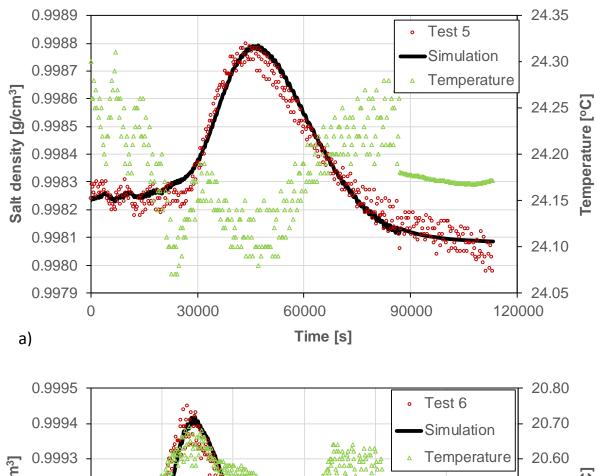


Fig. 10. Breakthrough curves obtained from the simulation (bold line) and the experiment (circles) with the associated temperature data (triangles) for Test 3 (a) and Test 7 (b) (double-porosity medium, case of 86% saturation, validating stage).



Salt density [g/cm<sup>3</sup>] 20.50 0.9992 20.40 0.9991 0.9990 20.30 0.9989 20.20 0.9988 20.10 0.9987 20.00 30000 60000 90000 120000 150000 180000 b) Time [s]

Fig. 11. Breakthrough curves obtained from the simulation (bold line) and the experiment (circles) with the associated temperature data (triangles) for Test 5 (a) and Test 6 (b) (double-porosity medium, case of 86% saturation, validating stage).

Fig. 12 displays the very good agreement obtained between the experimental BTC of Test 4 and those obtained by numerical simulation at  $S_{\rm w} = 82\%$  using the fitted  $D_{\rm DP} = 2.713 \times 10^{-4}$  cm<sup>2</sup>/s. The dispersivity obtained is 0.79 cm for this case. Consequently, the variation of

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dispersivity versus saturation could be established for the double-porosity medium in Fig. 13. This relation can be fitted by the power law:  $\lambda_{DP} = 0.315 S_w^{-4.5}$  (Fig. 13) that was proposed for single-porosity soils (Nutzmann et al., 2002; Toride et al., 2003; Latrille, 2013). Although our experiments concern a limited range of saturation produced by our experimental setup, it can be seen that the dispersivity in the double-porosity medium increases with the diminution of the saturation in the range of this study. This could be explained by the fact that the solute mixing length is much greater not only due to the heterogeneity of the medium (here clay spheres), but also as a result of some disconnections of fluid phases in the medium related to the unsaturated condition of the sample. This trend is in agreement with the findings of previous works in the literature for single-porosity soils (Padilla et al., 1999; Nutzmann et al., 2002; Toride et al., 2003; Sato el al., 2003; Latrille, 2013, among others). For saturation values close to the saturated condition, the dispersivity is of the same order of magnitude as the heterogeneity size in the double-porosity medium (diameter of the clay inclusion), whereas it is of the same order of magnitude as the grain size in the single-porosity medium (Ujfaludi and Maginecz, 1989; Haga et al., 1999; Illangasekare et al., 2010). In the saturation range of 80% - 100%, the dispersivity in the double-porosity medium increases more than the one in the single-porosity soils reported in the literature with an average grain size of 0.20 - 0.28 mm which is of the same order of magnitude as grain size of the sand constituent in the double-porosity medium (0.16 mm) (Fig. 13). The explanation can rely on the increase of tortuosity, in addition to the structural heterogeneity brought by the spherical inclusions.

The BTCs obtained for the dispersion tests in the single-porosity medium feature no delay in the solute's breakthrough (Figs 5 and 6), which can be considered as an indication of the non-existence of dead zones, at least in the sandy matrix, in both types of soil columns. Therefore, the non-Fickian BTCs in the double-porosity medium can be mainly attributed to the presence of heterogeneity, i.e. clay inclusions. The numerical simulations with the obtained dispersivities captured the non-Fickian behaviour of the tracer transport in the double-porosity medium columns. Generally, the agreement between simulated and experimental BTCs is better for the stepwise tests (Tests 3, 4 and 7: Fig. 10 and 12) rather than the pulsewise tests (Tests 5 and 6: Fig. 11). This might be the consequence of the complex non-Fickian mechanism of "tracer flooding water and flooded by water" involved in solute transport during the experiments. In fact, oscillations of density signals in the BTCs of the pulsewise tests were observed. These may be attributed to the switching of the valve twice to change water injection to solute injection and vice-versa, not to mention other accompanying physical mechanisms

(temperature variation, ion exchange in the clay, solute washing for the soil column before the tests). For Test 6, the agreement between simulated and experimental BTCs evaluated by  $R^2 = 0.8453$  is not as good as in the other tests. However, the fitting in terms of SSQ (2.402 × 10<sup>-6</sup>) and RMSE (6.597 × 10<sup>-5</sup>) is of similar quality as previously presented tests. Note that the number of measured values (N) is the greatest in the data of Test 6 with the longest end time of  $18 \times 10^4$  s (Table 7) and this test has the best mass balance (Table 3). The dispersivity obtained in this study may be considered as a reference value for a medium having a similar REV. It may be also used for a field-scale problem if upscaling conditions are respected, i.e. a REV exists. However, if field-scale permeability is so varied that the local solute dispersion is masked, the value obtained may no longer be applicable.

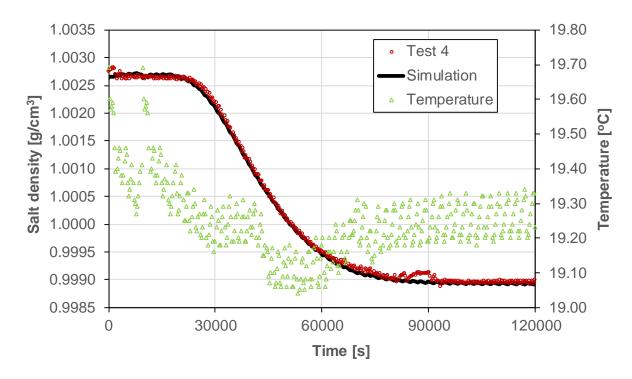


Fig. 12. Breakthrough curves obtained from the simulation (bold line) and the experiment (circles) with the associated temperature data (triangles) for Test 4 (double-porosity medium, case of 82% saturation).

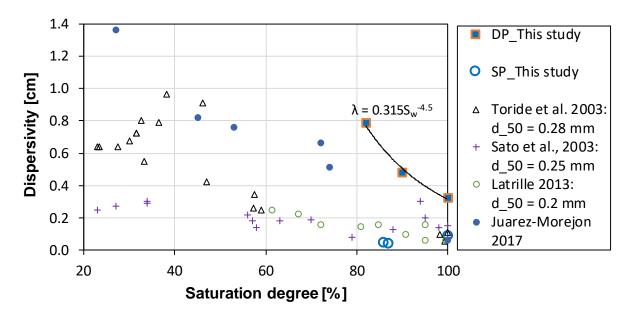


Fig. 13. Dispersivities of the double-porosity medium in function with the saturation obtained in this study, in comparison with the other ones in the single-porosity media available in the literature.

Table 7. Characteristic transport parameters – dispersivity and Péclet number and fitting quality indicators.

Experiment	q [cm/s]	$D [cm^2/s]$	Pe [-]	λ [cm]	$R^2$	SSQ	RMSE
Test 1	5.90E-04	1.899E-04	4.03	0.322	0.99821	5.634E-05	2.831E-04
Test 2	3.44E-04	1.662E-04	2.68	0.483	0.99677	5.195E-06	1.119E-04
Test 3	3.44E-04	1.662E-04	2.68	0.483	0.99816	2.459E-06	7.698E-05
Test 4	3.44E-04	2.713E-04	1.64	0.789	0.99886	1.479E-06	5.956E-05
Test 5	3.44E-04	1.662E-04	2.68	0.483	0.96512	6.677E-07	4.203E-05
Test 6	3.44E-04	1.662E-04	2.68	0.483	0.84531	2.402E-06	6.597E-05
Test 7	5.90E-04	2.850E-04	2.68	0.483	0.99878	5.777E-07	3.223E-05
Test 8	5.90E-04	1.310E-04	1.54	0.089	0.99925	2.227E-05	2.481E-04
Test 9	3.44E-04	4.044E-05	0.97	0.044	0.99049	9.377E-06	1.361E-04
Test 10	3.44E-04	3.839E-05	0.98	0.041	0.99545	5.148E-06	1.055E-04

<sup>(</sup>Péclet number Pe is calculated by  $Pe = \frac{vd_{50}}{D_e}$  for the single-porosity medium and Pe =

It must be noted that the BTC data obtained experimentally in this study may not be convenient for comparison with numerical simulations. Indeed, the successive dispersion tests in the experimental process (although all materials were washed before starting each new test) and

 $<sup>\</sup>frac{(\frac{q}{w_1})\ell}{D_{DP/w_1}}$  for the double-porosity medium, where  $v = q/\theta$ ;  $d_{50} = 0.0162$  cm;  $D_e = 1.545 \times 10^{-5}$ 

<sup>10</sup> cm<sup>2</sup>/s (effective diffusion coefficient of NaCl);  $\ell = 1.3$  cm is the characteristic length of the periodic cell of the double-porosity medium).

the water flow containing other minerals than the used NaCl tracer could have led to modifications of the concentration conditions vs targeted concentrations. The use of a high precision concentration measurement apparatus such as spectrophotometry or electrical conductivity would have allowed a more accurate analysis of the concentration in the porous samples. Indeed, the intermediary step of initial and boundary conditions' concentration estimation and conversion of BTC in concentration into density could also be the sources of further imprecision. Nevertheless, the simulation results using as input the fitted transport parameters reproduce very well all the BTCs and are consistent with the verified mass balance (Table 3). The obtained dispersivities capturing the Fickian and non-Fickian features of the dispersion tests in the single- and double-porosity medium columns, respectively allow us to believe that inherent errors in the experiments, if occurred, like clay particle dispersion, dispersion segregation of high concentration and density flow are negligible.

#### 5. Conclusions

In this study, the non-Fickian transport observed in 1D dispersion experiments in a physical double-porosity medium was quantified. Dispersivity values were obtained by fitting experimental BTCs with ones obtained through numerical simulations with an appropriate twoequation model. Dispersivities of solute transport experiments in single-porosity medium were also determined in order to compare with the non-Fickian dispersivities of the double-porosity medium. These experiments were carried out under different saturation conditions with the control of the water content profiles measured inside the soil samples by a nondestructive gamma-ray technique. In the interpretation of the BTCs, the saturation profiles, therefore, their uniformity or non-uniformity was taken into account. Within the range of the parameters considered in this work, the intrinsic nature of the dispersivity for a given saturation for solute transport in a heterogeneous medium composed of less permeable inclusions in a more permeable matrix has been demonstrated. The intrinsic value found seems independent of exterior loading (different boundary conditions). For high saturation values, the obtained trend between the dispersivity and the saturation in the double-porosity medium is the same as in single-porosity media presented in the literature. In the meanwhile, it could not be concluded definitely whether the trend obtained for the single-porosity medium is in agreement with the findings of previous works.

The present study contributes to the understanding of non-Fickian dispersion as a function of saturation in model heterogeneous double-porosity media. It could be considered as a helpful guide for modelers for estimating the dispersivity in heterogeneous media with variable

- 1 saturations. It has been shown that the dispersivity of the single-porosity medium should not
- 2 be used to replace the dispersivity of the double-porosity medium, even if its mobile zone is
- 3 constituted from the same material as the single porosity medium. Indeed, although the
- 4 transport in clay nodules is only dominated by diffusion, the presence of these zones, therefore
- 5 the heterogeneity of the medium has a great influence on the transport of solutes. Only the
- 6 saturation-dependence of the dispersivity for the single-porosity medium is similar to that of
- 7 the double-porosity medium.
- 8 It is worthwhile to point out that the obtained findings in this study are limited to high saturation
- 9 values. Further experiments of tracer dispersion for smaller values of saturation in single- and
- double-porosity media need to be performed in order to generalize and confirm these findings.
- 11 The results of these experiments will allow to address the validity of the non-monotonic
- behaviour of the dispersivity versus saturation for double-porosity porous media in comparison
- with the trend obtained in single-porosity media.

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