

Artificial neural network modeling of cefixime photodegradation by synthesized CoBi(2)O(4) nanoparticles

Oussama Baaloudj, Noureddine Nasrallah, Mohamed Kebir, Bouzid Guedioura, Abdeltif Amrane, Phuong Nguyen-Tri, Sonil Nanda, Aymen Amin Assadi

▶ To cite this version:

Oussama Baaloudj, Noureddine Nasrallah, Mohamed Kebir, Bouzid Guedioura, Abdeltif Amrane, et al.. Artificial neural network modeling of cefixime photodegradation by synthesized CoBi(2)O(4) nanoparticles. Environmental Science and Pollution Research, 2021, 28 (12), pp.15436-15452. 10.1007/s11356-020-11716-w. hal-03037964

HAL Id: hal-03037964 https://hal.science/hal-03037964v1

Submitted on 19 May 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1	Artificial neural network modeling of Cefixime photodegradation by
2	synthesized CoBi₂O₄ nanoparticles
3	Oussama Baaloudj ^a , Noureddine Nasrallah ^a , Mohamed Kebir ^{a,b} , Bouzid
4	Guedioura ^c , Abdeltif Amrane ^d , Phuong Nguyen-Tri ^{e*} , Sonil Nanda ^f , Aymen Amin
5	Assadi ^{d⁺}
6 7	^a Laboratory of Reaction Engineering, Faculty of Mechanical Engineering and Process Engineering USTHB, BP 32, Algiers, Algeria
8 9	^b Research Unit on Analysis and Technological Development in Environment (URADTE-CRAPC), BP 384, Bou-Ismail Tipaza, Algeria
10	^c Reactor division, Nuclear Research Center, Draria, Algeria
11 12	^d Univ Rennes - ENSCR / UMR CNRS 6226 "Chemical Sciences of Rennes" ENSCR, Campus de Beaulieu, av. du Général Leclerc, 35700 Rennes, France
13	^e Université du Québec à Trois-Rivières (UQTR), Trois-Rivières, Québec, G9A 5H7, Canada
14 15	^f Department of Chemical and Biological Engineering, University of Saskatchewan, Saskatoon, Saskatchewan, S7N 5A9, Canada
16 17 18 19	* Corresponding authors: E-mail: aymen.assadi@ensc-rennes.fr (A. A. Assadi); Tel.: +33 223238152 Phuong.Nguyen-Tri@uqtr.ca (P. Nguyen-Tri); Tel: +819 376-5011
20	Abstract
21	CoBi ₂ O ₄ (CBO) nanoparticles were synthesized by sol-gel method using
22	Polyvinylpyrrolidone (PVP) as a complexing reagent. For a single-phase with the spinel
23	structure, the formed gel was dried and calcined at four temperatures stages. Various
24	methods were used to identified and characterized the obtained spinel, such as X-ray
25	diffraction (XRD), Scanning Electron Micrograph (SEM-EDX), Transmission electron
26	microscope (TEM), Fourier transform infrared (FT-IR), X-ray fluorescence (XRF),
27	Raman and UV-Vis spectroscopies. The photocatalytic activity of CBO was examined
28	for the degradation of a pharmaceutical product Cefixime (CFX). Furthermore, for the
29	prediction of the CFX degradation rate an artificial neural network model was used.
30	The network was trained using the experimental data obtained at different pH with

different CBO doses and initial CFX concentrations. To optimize the network, various

algorithms and transfer functions for the hidden layer were tested. By calculating the

31

32

mean square error (MSE), 13 neurons were found to be the optimal number of neurons and produced the highest coefficient of correlation R² of 99.6 %. The relative significance of the input variables was calculated and the most impacting input was proved to be the initial CFX concentration. The effects of some scavenging agents were also studied. The results confirmed the dominant role of hydroxyl radical OH• in the degradation process. With the novel CoBi₂O₄/ZnO hetero-system, the photocatalytic performance has been enhanced, giving an 80% degradation yield of CFX (10 mg/L) at neutral pH in only 3 hours.

Keywords: CoBi₂O₄ Spinel, Characterization, Cefixime, Artificial neural network,
 Optimization.

1. Introduction

33

34

35

36

37

38

39

40

43

44

45

46 47

48

49

50

51

52

53

54

55

56

57

58

59

60

61

62

63

In recent years, pharmaceutical compounds have sullied the aqueous environment, among those compounds, antibiotics are the major concern which are extremely dangerous because they are widely used in both human and veterinary medicines (Kaczala and E. Blum 2015). Antibiotics are also used to improve feed quality and thus increase rates in the poultry industries (World Health Organisation 2011). Analytical study of antibiotic use in 76 countries from 2010-2015 found that global antibiotic consumption has increased by 65% for the fifteen years studied (Sample 2018). The majority of antibiotic products are not biodegradable and their environmental contamination can increase aquatic toxicity (Mostafaloo et al. 2019). Generally, antibiotics are present at amounts ranging from ng/L to µg/L in treated wastewater (Ibáñez et al. 2017; Shokoohi et al. 2017; Mirzaei et al. 2018). But even in small amounts, the presence of these compounds in water supplies will increase bacterial resistance, which produces a kind of micro-organisms called superbugs or antibiotic resistance (ABR) that are considered to cause very dangerous health effects (Elder et al. 2020). ABR is the ability to survive and multiply in the presence of antibiotic doses that are eventually deemed effective against a bacterial population or other microorganisms (Elder et al. 2020). A statistics study indicates that antibiotic resistance causes 700,000 deaths globally each year (Liu et al. 2019). ABR is now, in the 21st century, widely recognized as a grave threat to global health in the 21st century (Littmann et al. 2015).

As a typical example of antibiotics, Cefixime (CFX) was chosen which belonging to broad-spectrum, third-generation cephalosporin antibiotic derived semi-synthetically from the marine fungus Cephalosporium acremonium (Shooshtari and Ghazi 2017). CFX with molecular formula C₁₆H₁₅N₅O₇S₂ (Fig. 1) is (6R,7R)-7-[[(2Z)-2-(2-amino -1,3-thiazol -4- yl) -2- (carboxymethoxyimino) acetyl] amino] -3- ethenyl -8- oxo -5- thia -1-azabicyclo [4.2.0] oct-2-ene-2-carboxylic acid (Sheydaei et al. 2018). A study detected CFX as one of the most prevalent antibiotics observed in different water environments with concentrations of 278.65 to 422.1 ng/L (Mirzaei et al. 2018).

64

65

66

67

68

69

70

71

72

73

74

75

76

77

78

79

80

81

82

83

84

85

86

87

88

89

90

91

92

93

94

95

96

Fig. 1 Molecular structure of the antibiotic Cefixime.

Most of the conventional approaches that are commonly used for the removal of antibiotics have no positive findings for the degradation of antibiotic residues from contaminated waters or have substantial inconveniences such as by-products of the degradation (Xiong et al. 2018; Zhu et al. 2018; Zhang et al. 2020). On the other hand, Advanced Oxidation Processes (AOPs) have proven to be an effective solution for the rapid degradation of non-biodegradable substances in polluted water (Lou et al. 2017; Aboudalle et al. 2018; Kamagate et al. 2018). Particularly, the Heterogeneous photocatalysis process which is one of the best approaches for aqueous environment treatment as it is safer and a natural technology (used green energy), photocatalysis has proven on a previous investigations a successful degradation for various types of pharmaceutical drugs and organic compounds in the wastewater (Kamagate et al. 2018; Nguyen et al. 2018; Li et al. 2019b; Sayadi et al. 2019). Photocatalysis is based generally on the principle of the activation of a catalyst using an energy provided by a source of light (Homem and Santos 2011). As a catalyst, semiconductors are commonly used as they are very effective and easy to synthesize. Recently, many bismuth-based semiconductors have become very significant because of their high chemical stability, their excellent ability to absorb light and its narrow optical band gap (Tho et al. 2020). Cobalt bismuthate with the chemicals formula CoBi₂O₄ (CBO) was chosen as a typical semiconductor. To our knowledge, only one study has dealt with the synthesis and use of CBO photocatalyst (Jagadeesh and Sailaja 2019). Moreover, no study has been made yet concerning CoBi₂O₄/ZnO hetero-system.

Artificial neural network (ANN) is one of the most powerful tools for modeling linear and non-linear systems (Mohammadzadeh Kakhki et al. 2020). It an effective way of predicting experimental patterns in various systems, especially for photocatalysis

(Zhou et al. 2020) Moreover, ANN has the self-learning ability and to work with incomplete knowledge, storing information on the entire network, having fault tolerance and a distributed memory (Kaur and Kaur 2014). Thus, modeling and optimization can be accomplished without the rigor of the experimental information via ANN (Ayodele et al. 2020). Owing to that, ANN has been employed in the prediction of photocatalysis CFX efficiency. The challenge in the ANN modeling is to build a network with the smallest possible number of random experimental points to modulate and achieve the expected results approach to real cases without designing an experimental model. Only a few studies have attempted this challenge (Hassani et al. 2015; Tabatabai-Yazdi et al. 2019; Zulfiqar et al. 2019). To find the most efficient and effective network, various algorithms and transfer functions for the hidden layer has been attempted.

In this study, we report the synthesis of CoBi₂O₄ nanoparticles which prepared by sol-gel auto combustion method using polyvinylpyrrolidone (PVP) as a complexing reagent. Various methods were used to identified and characterized the obtained spinel, such as XRD, SEM-EDX, TEM, FT-IR, XRF, Raman and UV-Vis spectroscopies. After that, Cefixime was considered as a model of antibiotic compound to tested the photocatalytic performance of this catalyst. The effect of pH, irradiation time, the concentration of CFX and concentration of catalyst on the degradation of CFX was examined and modelized using ANN. Also, a Garson formula was introduced to measure the relative significance of different input variables. After that, the effects of some scavenging agents have been studied. Finally, the new hetero-system CoBi₂O₄/ZnO has been tested in order to enhance the photocatalytic performance.

2. Material and methods

2.1 Chemicals

Chemicals used in the present study were bismuth nitrate pentahydrate [Bi(NO₃)₃·5H₂O] (98.5% Chem-Lab), cobalt nitrate hexahydrate [Co(NO₃)₂·6H₂O] (98% Biochem), Zinc nitrate hexahydrate [N₂O₆Zn ·6H₂O] (98% Biochem), Cefixime trihydrate and Polyvinylpyrrolidone were obtained from a pharmaceutical company Pharmalliance Algeria, ethanol from Biochem, Sodium hydroxide, Hydrochloric acid, isopropanol and nitric acid were supplied by Sigma Aldrich. All chemicals were used as received without further purification. Distilled water has been used as a solvent.

2.2 Synthesis of CoBi₂O₄ nanoparticles

CoBi₂O₄ was synthesized by the aqueous PVP sol-gel method. The gel was prepared using stoichiometric amounts (1:2 ratio) of cobalt nitrate hexahydrate [Co(NO₃)₂·6H₂O] and bismuth nitrate pentahydrate [Bi(NO₃)₃·5H₂O] respectively. They were then dissolved in ethanol by using a magnetic stirrer; an excess of 5% citric acid was added in the solutions. After total solubilization, both solutions had been mixed and heated at 50 °C for 1h.

In the following step, Polyvinylpyrrolidone (PVP K30), used as a complexing agent was dissolved in 50 mL water and added to the reaction solution, applied about 15% w/w to obtain the complexing role (Giannopoulou et al. 2015). After concentrating the solutions by evaporation at 80 °C under stirring, the mixed solution turned into a reddish transparent gel. After drying and combustion at 200 °C for 6 h a precursor gel powder was formed (Xerogel).

The amorphous powder was subjected to calcination at 300°C, 400°C, 600°C and 800°C in an air oven each one for 3 h. The synthesized sample was annealed to improve the degree of crystallinity and to eliminate any leftover carbonated waste after the combustion reaction. Then, it was subjected to phase identification, microstructural analysis and photocatalytic studies.

2.3 Characterization

The XRD analysis of the products was recorded by a Phillips PW 1730, X-ray diffractometer using monochromatized Cu K radiation (λ = 0.15417 nm the morphology of the sample SEM was studied using the electron scanning microscope FEI Quanta 650 equipped with EDX and Mapping analysis with 20 kV accelerated voltage. Transmission electron microscope (TEM) images of the nanostructures were taken on a JEOL-24011BU. ATR-FTIR spectra were obtained in the range of 400–4000 cm⁻¹ with an Alpha Bruker spectrometer (One FTIR model). The chemical analysis was carried out by using Xray fluorescence (PANalytical epsilon3-XL spectrometer). UV–visible diffuse reflectance spectrum (DRS) of the sample was produced by dry pressed disk sample using Cary 5000 UV–vis. The Raman analysis was carried out using LabRAM HR Evolution system Horiba Jobin-Yvon.

2.4 Photocatalytic degradation study

158

159

160

161

162

163

164

165

166

167

168

169

170

171

172

173

174

175

176

177

178

179

180

181

182

183

184

185

186

We selected (CFX) as a pharmaceutical model contaminant to investigate CoBi₂O₄'s photocatalytic performance. The photocatalytic degradation of CFX was conducted by mixing an appropriate quantity of CoBi₂O₄ in 100 mL of CFX solution with known concentration. The pH was set to the desired value with NaOH and HCl. The solution was continuously stirred for 2h in dark conditions in a closed reactor to reach the balance of adsorption and desorption before switching on the light. As a visible light source, a tungsten bulb 200 W manufactured by Osram was applied and positioned at 10 cm from the solution surface. The intensity (8.25 mW/cm²) and the maximum emission (630 nm) of the lamp were measured by a digital flux meter (Testo 545). A cooling system consisting of a double-walled reactor attached to a thermostatic bath (Daihan Labtech, LCB-11D) has been used to maintain the temperature of the solution nearly constant (25 ± 1 °C) during the photocatalytic experiments. At a regular time, 2 mL of the solution was centrifuged at the speed of 5,000 rpm for 10 min to filter the suspended CoBi₂O₄ particles from the sample and directly analyzed by UV-Visible spectrometry (OPTIZEN, UV-3220UV) at the maximum absorption wavelength of CFX (288 nm). It should be mentioned that the small positive error of UV-Visible spectrometry in comparison with HPLC was ignored. The following equation was used to calculate the degradation rate of CFX.

Photodegradation% =
$$\frac{(C_O - C)}{C_O} \times 100$$
 (1)

Where Co is the initial concentration of CFX after adsorption and C is the concentration of CFX at time t.

As already mentioned, the CFX concentration was monitored by measuring the absorbance of the samples using a UV–Visible spectrophotometry. The degradation of the pollutant was followed by the disappearance of the band corresponding to CFX at 288 nm as shown in figure 2. The experiment was conducted in the presence of 0.1 g CBO at pH 6, room temperature and an initial CFX concentration of 10 mg/L.

To evaluate the CFX mineralization by the catalyst CBO, the amount of total organic carbon (TOC) in the solution at a regular time with the same conditions was measured by TOC analyzer (Shimadzu TOC-L, Japan) using the catalytic combustion

method at 680°C. The results of TOC analysis are presented in Fig. 2 (b). The following equation was used to calculate the TOC removal (%):

$$TOC_{removal}\% = \frac{(TOC_O - TOC)}{TOC_O} \times 100$$
 (2)

where TOC_0 (mg) is the initial TOC values after adsorption, TOC (mg) is the remaining TOC in solution.

the TOC removal reached 67.23% in the presence of CBO catalyst after 6 h, which confirmed that cefixime could be effectively mineralized in the presence of CBO photocatalyst under visible light irradiation.

Fig. 2 (a) UV-visible spectra (B)TOC removal of CFX (10 mg/L).

2.5 Experimental design and modeling

The input layer consisted of four numbers of neurons, which are pH, irradiation time (min), the dose of CBO (g/L) and concentration of CFX (mg/L), while photodegradation of CFX (%) was selected as the output layer. The application was performed in the laboratory to obtain the actual responses. As it is a challenge to build a network with the smallest possible number of random experimental points to approach to real cases without designing an experimental model. For that, a total of 40 random experimental points has been randomly divided into 70% training, 15% testing and 15% validation data. Training data were presented to the network during training. The use of validation data was to test network generalization and to stop training if generalization stops. The test data provided a fully independent network performance measurement during and following training (Ayoubi-Feiz et al. 2019).

The modeling of the photodegradation of CFX was carried out by MATLAB R2014a which has been widely used (Medarević et al. 2016; Ayoubi-Feiz et al. 2019; Tabatabai-Yazdi et al. 2019; Zulfiqar et al. 2019; Garg et al. 2020). The model parameters are described in Table 1. For the model development, the feed-forward neural network with backpropagation method was used. Furthermore, a supervised learning technique called Levenberg-Marquardt with a training function 'trainlim' has been used for training the network.

Table 1 ANN model parameters.

The number of hidden neurons is a critical element of modeling experimental processes with ANN (Hazrati et al. 2017). For its optimization, a set of topologies have been used in which the number of hidden neurons changed from 2 to 20. The mean square error MSE was performed to investigate the relation between the hidden neurons and the network error (Ayoubi-Feiz et al. 2019). The MSE function was used as Eq.3. The coefficient of correlation R (Eq.4) used to predict model conformity and applicability, the more the value is close to one, the better it is applicated.

$$MSE = \frac{1}{n} \sum_{i=1}^{n} (y_{i,NN} - y_{i,exp})^{2}$$
 (3)

$$R^{2} = 1 - \frac{\sum_{i=1}^{n} (y_{i,NN} - y_{i,exp})^{2}}{\sum_{i=1}^{n} (y_{m} - y_{i,exp})^{2}}$$
(4)

where n is the total number of data points, $y_{i,NN}$ and $y_{i,exp}$ are the network prediction and the experimental response and i is an index of the data, y_m is the average of the actual values.

3. Results and discussion

3.1 Characterization of the catalyst

To confirm the single phase of $CoBi_2O_4$, an X-ray diffraction (XRD) pattern after each calcination was taken. Figure 3 illustrates the effect of different annealing temperatures on the structure of the synthesized samples. The pure $CoBi_2O_4$ nanopowders were obtained at 800 °C (after annealing at 300 °C, 400 °C, 600 °C) almost a single phase of CBO without impurity was observed. The phase of the $CoBi_2O_4$ could be indexed with the five main diffraction peaks at $2\theta = 27.88^{\circ}$ (111), 33.12° (200), 45.72° (220), 52.76° (311) and 56.04° (222) which were in good agreement with the reference peaks (Jagadeesh and Sailaja 2019) almost the same, and also in line with the standard card (JCPDS 50-0369) from Joint Committee of Powder Diffraction Standards of the face-centered cubic of $CoBi_2O_4$.

The crystallite size was calculated from the main diffraction peaks of the last calcined product (at 800 °C) with the following Debey- Scherer formula Eq. (5). The crystalline sizes have been estimated between 19.45 and 46.39 nm.

$$D = \frac{K\lambda}{\beta\cos(\theta)} \tag{5}$$

where D is the crystallite size of the phase-in nanometers, K is the Scherrer constant. (K = 0.94 for cubic symmetry), λ is the wavelength of the X-ray beam used, β is the full width at half maxima (FWHM) in radians and θ is the Braggs angle.

Fig. 3 XRD diffractogram of CoBi₂O₄ calcined at (a) 300°C, (b) 400°C, (c) 600°C and (d) 800°C.

For the XRD patterns of the ZnO nanoparticles prepared via co-precipitation. As shown in Figure 4, the ZnO nanoparticles displayed a Wurtzite structure (hexagonal phase, space group P63mc). All standard peaks observed for ZnO nanoparticles were in line with those obtained from (JCPDS 36-1451) (Raou 2017).

Fig. 4 XRD diffractogram of ZnO.

The spinel structure may be considered as a continuously bound crystal in which the atoms are bonded by equivalent forces (ionic, covalent or van der Waals) to all nearest neighbors. The spinel shaped metal ions are found in two separate sub-lattices known as A-site (tetrahedral) and B-site (octahedral), as stated in the geometric arrangement of the nearest oxygen neighbors (Sabri et al. 2016). Fourier transform infrared spectra of CoBi₂O₄ recorded within the range of 4000–400 cm⁻¹ are shown in Figure 5. It is evident in the IR spectrum that CBO had five intensive absorption peaks in the region of 400-900 cm⁻¹, which could be related to the typical metal-oxygen (M–O) vibrations (Jasaitis et al. 2011). The presence of peaks between 570 cm⁻¹ and 669.9 cm⁻¹ belonged to the vibration of the Bi–O bond and the vibration peak at 824.28 cm⁻¹ can be assigned to Bi–O–Bi (Yang et al. 2018; Viruthagiri and Kannan 2019). Furthermore, the appearance of Peaks 446.13 cm⁻¹ and 532.31 cm⁻¹ can probably be related to the tetrahedral (CoO₄) in the spinel (Sousa et al. 2019).

Fig. 5 FTIR spectra of CoBi₂O₄.

Figure 6 shows the Raman spectra of CoBi₂O₄. There are five Raman-activated bands; all these modes were observed at ambient conditions. The bands above 600 cm⁻¹ were due to the motion of oxygen in the tetrahedral AO₄ octahedral BO₆ groups (Li et al. 2019a), suggesting a certain amount of Co and Bi disturbance in the tetrahedral and octahedral sites. The low-frequency modes observed at 54.13, 80.5 and 123.73 cm⁻¹ can be related to the Bi³⁺ ion motion in the spinel layer and the mode at 521.03 cm⁻¹ can be attributed to the variation of bond distances between Bi atoms

and the spinel block apical O atoms (Zhang et al. 2018). The strongest band observed at 264 cm⁻¹ corresponded to O–Bi–O vibration (Zhang and Saxena 2013). The Raman modes of 193.29, 476.05 ,600 and 670 cm⁻¹ were related to the vibration of the Co atoms (Diallo et al. 2015). Besides, a small Raman band was observed at 151.35 cm⁻¹; the origin of this band is still unknown.

Fig. 6 Raman spectra of CoBi₂O₄.

The morphology and structures of the prepared sample were visualized by Scanning Electron Micrograph (SEM) and transmission electron microscope (TEM) techniques were used. Figure 7 (a) shows a typical SEM image of the sample. It can be observed that the sample consisted of particles exhibiting cubic morphology. Due to the ultrafine nature of the sample, these particles displayed a slightly agglomerated morphology. The TEM images of CoBi₂O₄ shown in figure 7 (b) demonstrate clear information regarding the size, shape and size distribution of nanoparticles. Based on the size distribution histogram in the inset figure, the catalyst was composed of particle size in the range of 22 nm. It was found that the size of the crystallite obtained from XRD analysis and the particle size obtained from TEM analysis were in the same size order. This picture also shows the existence of several aggregations within the range of nanometres.

The electron diffraction pattern of the nanoparticles are shown in figure 7 c, This pattern has a cubic shape and right angles (K/H = 1, (\vec{H}, \vec{K}) = 90°) which is defined as a cubic face-centered CFC type of crystalline spinel (Andrews KW , Dyson DJ 1967) and that confirmed the XRD results.

Fig. 7 SEM (a) TEM (b) Diffraction pattern (c) of nanoparticles CoBi₂O₄.

The elementary study of the CoBi₂O₄ nanoparticles was carried out by Energy Dispersive X-ray spectrometer (EDX). The peaks for Co, Bi and O elements were clearly observed in the EDX image, as shown in figure 8 (a). The atomic percentage of the elements in the sample are given in the inset Table of the EDX data. The atomic percentages of Co, Bi and O were estimated to be 14.02%, 26.25% and 59.73%, respectively. The quantification of Co, Bi and O confirmed that the atomic ratio of Co:Bi:O was 1.00:1.87:4.26, which agrees with the theoretical stoichiometric composition of CoBi₂O₄.

X-ray fluorescence measurements (XRF) spectra figure 8 (b) showed that Co and Bi were the most abundant elements. Regarding silver (Ag), it was used in the XRF analysis method. The inset Table of the XRF results shows that the weight percentage of CoO and Bi₂O₃ was estimated to be (14.109%) for CoO and (85.89%) Bi₂O₃. The molar mass was then calculated for CoO and Bi₂O₃ and found to be 74.93 and 467.96 g/mol, respectively. This confirmed that the atomic ratio of Co:Bi:O was 1.00:1.95:3.93, which agrees with the theoretical stoichiometric composition of CoBi₂O₄ and also in line with the EDX analysis.

Fig. 8 Energy dispersive X-ray (a) X-ray fluorescence (b) diffractogram of nanoparticles CoBi₂O₄.

The elemental mappings in figure 9 were located in the same EDX region. The colored map of the SEM image (where Co, Bi and O are given in colors like Green, Red and Blue, respectively) revealed that Co, Bi, and O elements coexisted and were uniformly distributed.

Fig. 9 EDX elemental mapping images of CoBi₂O₄ a) All elements b) Cobalt c) Bismuth d) Oxygen.

Optical properties of CoBi₂O₄ nanostructure were measured using the UV–Vis diffuse reflectance spectrum. As shown in figure 10, the bandgap was estimated as 2.5 eV which is lower than that of TiO₂ and ZnO (3.2 eV) (Serpone 2006; Miki-yoshida 2016). This indicates that CBO has an excellent absorption level in the wavelength range from 300 to 800 nm that contains both UV and visible light, which appears more efficient than ZnO and TiO₂, leading to an increase in the formation of electron-hole pairs on the photocatalyst's surface in visible light irradiation.

Fig. 10 UV-Vis optical absorption edges.

3.2 Artificial neural network modeling of the elimination of Cefixime

For the modeling, irradiation time, pH, initial CFX concentration and CBO catalyst dose were selected as the input variables and the related degradation rate of CFX was the output variable. To feed the structure of the ANN, a total of 40 experimental sets were used. The data sets have been divided into 70% training, 15% validation and 15% test, as shown in Table 2.

Table 2 Data sets of experimental results and ANN's predicted results.

The essential part of the neural network architecture is the choice of the number of hidden neurons. The model's best performance comes when the mean square error (MSE) determined using equation (2), is the smallest.

Fig. 11 Impact of the number of hidden neurons on the MSE of the neural network.

Figure 11 shows that the Levenberg-Marquardt with a training function 'trainlim' and 13 hidden-layer neurons produced the highest correlation coefficient $R^2=0.996$ and the smallest MSE value 1.81×10^{-4} . The relation of experimental data to ANN model predictions can be derived from the contrast between the experimental data and the predictions of the ANN model (Fig.12). The training data were given as the model's input values, and the simulation output was compared with the relevant experimental data. The adaptation of the model results to experimental data for the training data set was very successful. The main part of model validation is the testing of network performance for the testing data set access, the correlation coefficient was likewise very high in this case $R_{\rm test}^2=0.977$.

Fig. 12 The relation of experimental data to the ANN model's predictions for each type of data.

3.2.1 Effect of pH

The pharmaceuticals polluted water is discharged in the environment at variable pH values; consequently, it is necessary to investigate the role of pH on the photocatalytic degradation. For this purpose, photodegradation experiments have been carried out in 100 ml of 10 mg/L CFX solution at a fixed CBO amount of 0.1 g and 360 min irradiation. Three different initial pH was selected and kept constant by adding HCl or NaOH. The degradation rate of CFX increased with the increase of the pH value of the solution from 4 to 6 and then decreased from 6 to 8, as observed in figure 13 (a). Based on these results, it can be deduced that the photodegradation of CFX was optimal at pH = 6. It is difficult to grasp pH impact on the performance of the photocatalytic degradation process because of its multiple roles, such as the neutral electrical charge of the CFX molecules, as well as the protonation and the electropositive surface of CBO. This may also be related to the reaction between organic compounds and dissolved oxygen (Abbasi and Hasanpour 2017). The improvement of the photocatalytic activity of the catalyst at pH = 6 can probably be related to the charge accumulation on the external surface (Bai et al. 2017). So, the

pH change may therefore affect the dispersion of the catalyst in the suspension. Moreover, figure 13 (b) shows that the values predicted by the ANN model were well in line with the experimental data.

3.2.2 Effect of the catalyst dose

The effect of CBO catalyst dose on the photodegradation activity of CFX was examined. A set experiment was conducted by using a catalyst dose ranging from 0.05 to 0.15 g in 100 mL of solution for 15 mg/L of CFX at pH = 6 and 360 min irradiation. Figure 13 (c) illustrates the effect of the catalyst dose on the CFX photodegradation removal, showing that the degradation of CFX increased from 66% to 72% for an increase of the catalyst dose from 0.05 to 0.08 g because of the limited active sites available on the CBO catalyst surface. However, the degradation efficiency of CFX decreased beyond this amount of catalyst, namely from 0.08 to 0.15 g. This may be attributed to the fact that the accumulation of free catalysts leads to an increase in the opacity of the solution and a decrease in light penetration. This is due to a scattering effect for an expansion of the dose of catalyst (Zhou et al. 2018). Therefore, the optimal dosage of the catalyst for CFX photodegradation was selected to be 0.08 g. Based on the obtained results shown in figure 13 (d), the results predicted by the ANN and the experimental values were well agreed.

3.2.3 Effect of the initial CFX concentration

For the purpose, the photodegradation studies were performed by varying initial concentrations of CFX from 5 to 15 mg/L in 100 ml of the solution with 0.1 g of the CBO catalyst at pH 6 after 360 min of irradiation time under visible light. Figure 13 (e) shows that by raising the CFX concentration, the degradation yield decreased from 82 to 55%. This behavior can be related to the diminishing number of available sites on the catalyst CBO surface. Therefore, a high number of CFX molecules would have an inhibitory effect on the photodegradation efficiency due to the lack of any direct contact between the CBO catalyst surface and CFX. Besides, many photons could be absorbed by the CFX molecules and hence did not reach the catalyst surface. The comparison between ANN predicted values and the experimental data shown in figure 13 (f) revealed a good agreement between them.

Fig. 13 (a) effect pH (c) effect of catalyst dose (e) effect of initial CFX concentration (b,d,f) Comparison of the experimental results with the predicted ANN values for various operating parameters.

3.2.4 The relative significance

An evaluation process based on the network weight matrix and the Garson's equation (Garson 1991) was applied to calculate the relative significance of the input variables on the network's output. The relative significance of input variables influencing the CFX photodegradation values has been calculated by equation 6 (Ibrahim 2013).

$$RI_X = \sum_{x=1}^{n} \frac{|w_{xy}w_{yz}|}{\sum_{y=1}^{m} |w_{xy}w_{yz}|}$$
 (6)

where RI_X is the relative significance, n and m are respectively the input numbers and hidden neurons. w_{xy} are the connection weights, $\sum_{y=1}^{m} |w_{xy}w_{yz}|$ is the sum of the product of the final weights of the connections from input neurons to hidden neurons along with the links from hidden neurons to output neurons (Ibrahim 2013).

Table 3 ANN model's weight and bias matrix.

The ANN model's weight matrix is given in Table 3, while the relative significance of the input variables derived from equation (4) is displayed in figure 14. As can be seen, the initial CFX concentration showed the most significant impact on the performance of CFX photodegradation (28%) compared with the other operating factors, namely CBO dosage (27%), irradiation time (24%) and pH (21%).

Fig. 14 The relative significance of the inputs on CFX degradation.

3.3 Effect of scavenging agents and the photocatalytic mechanism

Generally, the photodegradation of organic compounds by a catalyst semiconductor can be described by different reaction pathways regulated by different active species such as OH* hydroxyl radical, superoxide O₂*, and photogenerated electrons and holes e⁻/ h⁺ pairs which called reactive oxidant species (ROS) (Zuo et al. 2017). To investigate the dominant ROS on the photodegradation of Cefixime over the CBO catalyst and for an understanding of the photocatalytic mechanism, several active species trapping tests were run in the presence of different concentrations of some scavengers. Each of the scavenging species was chosen for a specific type of photocatalytic radicals. p-Benzoquinone BQ was used as superoxide O₂* radical quencher (a scavenger for superoxide anion radical) (Nithya and Ayyappan 2020), Potassium dichromate K₂Cr₂O₇ was used as a scavenger for electron (e⁻) (Gupta et al. 2020), Disodium ethylenediaminetetraacetic acid (Na₂-EDTA) was used as a scavenger for the hole (h⁺) and as photoinduced h⁺ acceptor (Tho et al. 2020) and isopropanol (IPA) alcohol was employed as an OH* scavenger (hydroxyl OH* radical quencher) (Ghattavi and Nezamzadeh-Ejhieh 2020). The tests were performed with an initial Cefixime concentration of 5 mg/L, at pH 6, and using CBO with a solution ratio of 0.8 g/L. The impact of these scavenging agents (quenchers) on the photodegradation efficiency of CBO nanocomposite was shown in Fig 15.

Fig. 15 Effect of different scavengers on photocatalytic degradation of CFX over CBO photocatalyst.

The obtained results are summarized in the next order of the photodegradation efficiency in the presence of different scavengers: Sample without the scavengers (Free) $(79 \text{ %}) > \text{K}_2\text{Cr}_2\text{O}_7$ (64%) > p-Benzoquinone $(59\%) > \text{Na}_2\text{-EDTA}$ (40%) > Isopropanol (25%). As shown, the most decrease in the degradation efficiency was observed by Isopropanol which confirmed that the generated OH* has a dominant role in the degradation of CFX. OH* radicals are strong, active and unselected oxidizing species for organic substrates in terms of oxidative degradation (Kumar et al. 2020). Then it comes Na $_2$ -EDTA, this indicates that both OH* and h* are the major active species. However, in the presence of p-Benzoquinone, only a small effect has been observed on the CFX photodegradation. This indicates that O $_2$ * was a minor active species. The effect of K $_2$ Cr $_2$ O $_7$ was so small that it neglected the electron e $_1$ role in the CFX degradation. According to these results, the following pathways (Fig. 16 and equations 7,8,9 and 10) were suggested for the photodegradation of CFX onto CBO.

$$CoBi_2O_4 + hv \rightarrow CoBi_2O_4^* + e^- + h^+$$
 (7)

$$H_2O + h^+ \rightarrow OH^* + H^+$$
 (8)

$$O_2 + e^- \rightarrow O_2^{-*}$$
 (9)

$$CFX + OH^* + h^+ + O_2^{-*} \rightarrow CO_2 + H_2O + other small degradation products$$
 (10)

Fig. 16 Proposed mechanism for CFX degradation in the presence of CoBi₂O₄ nanoparticles.

3.4 Comparison of Cefixime photodegradation by different processes

In order to enhance the photocatalytic performance of CFX by synthesized CBO nanoparticles and also to eliminate the photolysis effect, a set of experiments have been tested in the same condition as illustrated in fig.17. The experiment was conducted in the presence of 0.1 g catalyst at pH 6, room temperature and an initial CFX concentration of 10 mg/L. In the absence of the catalyst, a very small degradation rate of CFX by photolysis can be observed, lower than 5%. On the other hand, in the presence of the CBO photocatalyst, the degradation rate of CFX reached 60 % within 3 h. The hetero-systems can increase the rate of photo-activity by extending the spectral photo response of wide band semiconductors toward the visible light region, as already demonstrated in recent works (Belaissa et al. 2016; Boumaza et al. 2016; Rekhila et al. 2017; Boutra et al. 2019). For that, the new hetero-system CoBi₂O₄/ZnO has been tested. Different weight ratios of the two catalysts CBO and ZnO were tested to examine the impact on the CFX photodegradation. As can be seen in fig.17, the high photodegradation efficiency was noticed at ratio of 50%/50% weight percentage of the hetero-system CoBi₂O₄/ZnO, leading to an 80% degradation efficiency within only 3 h.

Fig. 17 Time-course of the CFX degradation by different processes.

4. Conclusion

In summary, in the present work, $CoBi_2O_4$ nanoparticles were successfully synthesized by the sol-gel method and then characterized by both spectroscopic and morphological tools. The catalysts prepared were tested for CFX antibiotic degradation. The photocatalytic degradation of CFX was modeled successfully using an artificial neural network (ANN). A feed-forward neural network with backpropagation method and a Levenberg-Marquardt method was used. The degradation of CFX in different operating conditions was predicted with 13 hidden neurons. The results showed that the $CoBi_2O_4$ nanoparticles led to an 86 % CFX photodegradation yield in the following optimal condition, pH = 6, 5 mg/L CFX, 0.8 g/L catalyst dose and 360 min of irradiation. It was shown that all the parameters of the study (irradiation time, pH, initial concentration of CFX and catalyst dose) have a significant effect on the CFX photodegradation.

However, the main effect was the CFX's initial concentration. Moreover, the effects of some scavenging agents were studied and the results confirmed the dominant role

of hydroxyl radical OH• in the CFX degradation process. Finally, the new CoBi ₂ O ₄ /ZnO
hetero-system was discussed and demonstrated a rapid degradation efficiency of 80
% in just 3 hours. Owing to these promising preliminary results, it will be further
considered in future works.

Acknowledgment

The authors gratefully acknowledge the financial support from the Thematic Research Agency for Science and Technology (ATRST) through the national research program (PM, PRFU Project N°A16N01UN160420190002) and the Directorate-General for Scientific Research and Technological Development (DGRSDT) of Algeria.

491 492

483

484

485

486

487

488

489

490

Declaration

494

495

493

Ethics approval and consent to participate

496 Not applicable

497

498

Consent for publication

499 Not applicable

500

501

502

Competing interest

The authors declare that they have no conflict of interest.

503

504

Authors contributions

- 505 Oussama Baaloudj and Mohamed Kebir. investigation, formal analysis, visualization,
- 506 writing original draft.
- 507 Noureddine Nasrallah and Aymen Amin Assadi: conceptualization, funding acquisition,
- 508 methodology, resources, project administration, supervision, writing-review and
- 509 editing.
- 510 Phuong Nguyen-Tri: writing-review and editing, methodology, conceptualization.
- 511 Bouzid Guedioura: formal analysis
- 512 Abdeltif Amrane and Sonil Nanda: investigation, visualization.

513

514

Availability of data and materials:

515	The datasets used and/or analysed during the current study are available from the
516	corresponding author on reasonable request.
517	
518	References
519	
520	Abbasi S, Hasanpour M (2017) The effect of pH on the photocatalytic degradation of
521	methyl orange using decorated ZnO nanoparticles with SnO2 nanoparticles. J
522	Mater Sci Mater Electron 28:1307–1314. https://doi.org/10.1007/s10854-016-
523	5660-5
524	Aboudalle A, Fourcade F, Assadi AA, et al (2018) Reactive oxygen and iron species
525	monitoring to investigate the electro-Fenton performances. Impact of the
526	electrochemical process on the biodegradability of metronidazole and its by-
527	products. Chemosphere 199:486–494.
528	https://doi.org/10.1016/j.chemosphere.2018.02.075
529	Andrews KW , Dyson DJ KS (1967) Interpretation of electron diffraction patterns
30	Ayodele BV, Alsaffar MA, Mustapa SI, Vo DVN (2020) Backpropagation neural
31	networks modelling of photocatalytic degradation of organic pollutants using
532	TiO2-based photocatalysts. J Chem Technol Biotechnol.
333	https://doi.org/10.1002/jctb.6407
34	Ayoubi-Feiz B, Sheydaei M, Karimi M (2019) Visible light photoelectrocatalysis for
35	wastewater treatment using bifacial N-TiO2/Graphene/Ho2O3/Titanium
36	nanocomposite: Artificial neural network modeling and evaluation of ozone
37	addition. Process Saf Environ Prot 127:56–65.
38	https://doi.org/10.1016/j.psep.2019.04.026
39	Bai Y, Huang H, Wang C, et al (2017) Engineering the surface charge states of
540	nanostructures for enhanced catalytic performance. Mater Chem Front 1:1951-
541	1964. https://doi.org/10.1039/c7qm00020k
542	Diallo A, Beye AC, Doyle TB, et al (2015) Green synthesis of Co3O4 nanoparticles
543	via Aspalathus linearis: Physical properties. Green Chem Lett Rev 8:30-36.
544	https://doi.org/10.1080/17518253.2015.1082646

Elder FCT, Feil EJ, Snape JN, et al (2020) The role of stereochemistry of antibiotic

545

546	agents in the development of antibiotic resistance in the environment. Environ Int 139:105681. https://doi.org/10.1016/j.envint.2020.105681
547	139.103001. https://doi.org/10.1010/j.envint.2020.103001
548	Garg A, Kaur G, Sangal VK, et al (2020) Optimization methodology based on neural
549	networks and box - behnken design applied to photocatalysis of acid red. 0-2.
550	https://doi.org/10.4491/eer.2019.246
551	Garson GD (1991) Interpreting neural-network connection weights. AI Expert 6:46-51
552	Ghattavi S, Nezamzadeh-Ejhieh A (2020) A visible light driven AgBr/g-C3N4
553	photocatalyst composite in methyl orange photodegradation: Focus on
554	photoluminescence, mole ratio, synthesis method of g-C3N4 and scavengers.
555	Compos Part B Eng 183:107712.
556	https://doi.org/10.1016/j.compositesb.2019.107712
557	Giannopoulou I, Saïs F, Thomopoulos R (2015) Handbook-of-pharmaceutical-
558	excipients-6th-edition. Rev des Nouv Technol l'Information E.28:257–262
559	Gupta NK, Ghaffari Y, Kim S, et al (2020) Photocatalytic Degradation of Organic
560	Pollutants over MFe2O4 (M = Co, Ni, Cu, Zn) Nanoparticles at Neutral pH. Sci
561	Rep 10:1-11. https://doi.org/10.1038/s41598-020-61930-2
562	Hassani A, Khataee A, Karaca S (2015) Photocatalytic degradation of ciprofloxacin
563	by synthesized TiO2 nanoparticles on montmorillonite: Effect of operation
564	parameters and artificial neural network modeling. J Mol Catal A Chem 409:149-
565	161. https://doi.org/10.1016/j.molcata.2015.08.020
566	Hazrati H, Moghaddam AH, Rostamizadeh M (2017) The influence of hydraulic
567	retention time on cake layer specifications in the membrane bioreactor:
568	Experimental and artificial neural network modeling. J Environ Chem Eng
569	5:3005-3013. https://doi.org/10.1016/j.jece.2017.05.050
570	Homem V, Santos L (2011) Degradation and removal methods of antibiotics from
571	aqueous matrices - A review. J Environ Manage 92:2304–2347.
572	https://doi.org/10.1016/j.jenvman.2011.05.023
573	Ibáñez M, Borova V, Boix C, et al (2017) UHPLC-QTOF MS screening of
574	pharmaceuticals and their metabolites in treated wastewater samples from
575	Athens, J Hazard Mater 323:26–35.

576	https://doi.org/10.1016/j.jhazmat.2016.03.078
577	Ibrahim OM (2013) A comparison of methods for assessing the relative importance of
578	input variables in artificial neural networks. J Appl Sci Res 9:5692-5700.
579	https://doi.org/10.1067/mmt.2002.123333
580	Jagadeesh C, Sailaja BB V (2019) Photocatalytic Degradation of Rohdamine B Dye
581	using CoBi2O4 Nanocatalyst and Effect of Various Operational Parameters.
582	CIKITUSI J 6:97–107
583	Jasaitis D, Beganskiene A, Senvaitiene J, et al (2011) Sol-gel synthesis and
584	characterization of cobalt chromium spinel CoCr 2O 4. Chemija 22:125-130.
585	https://doi.org/10.1039/c3ce41663a
586	Kaczala F, E. Blum S (2015) The Occurrence of Veterinary Pharmaceuticals in the
587	Environment: A Review. Curr Anal Chem 12:169–182.
588	https://doi.org/10.2174/1573411012666151009193108
589	Kamagate M, Amin Assadi A, Kone T, et al (2018) Activation of persulfate by
590	irradiated laterite for removal of fluoroquinolones in multi-component systems. J
591	Hazard Mater 346:159–166. https://doi.org/10.1016/j.jhazmat.2017.12.011
592	Kaur R, Kaur B (2014) Artificial Neural Network Learning Enhancement using
593	Bacterial Foraging Optimization Algorithm. Int J Comput Appl 102:27-33.
594	https://doi.org/10.5120/17852-8812
595	Kumar R, Barakat MA, Al-Mur BA, et al (2020) Photocatalytic degradation of cefoxitin
596	sodium antibiotic using novel BN/CdAl2O4 composite. J Clean Prod 246:119076.
597	https://doi.org/10.1016/j.jclepro.2019.119076
598	Li K, Luo Y, Liu B, et al (2019a) High-performance NO2-gas sensing of ultrasmall
599	ZnFe2O4 nanoparticles based on surface charge transfer. J Mater Chem A
600	7:5539-5551. https://doi.org/10.1039/c8ta12168k
601	Li Q, Jia R, Shao J, He Y (2019b) Photocatalytic degradation of amoxicillin via TiO2
602	nanoparticle coupling with a novel submerged porous ceramic membrane
603	reactor. J Clean Prod 209:755–761. https://doi.org/10.1016/j.jclepro.2018.10.183
604	Littmann J, Buyx A, Cars O (2015) Antibiotic resistance: An ethical challenge. Int J
605	Antimicrob Agents 46:359–361. https://doi.org/10.1016/j.ijantimicag.2015.06.010

606	Liu M, Ni H, Yang L, et al (2019) Pretreatment of swine manure containing β-lactam
607	antibiotics with whole-cell biocatalyst to improve biogas production. J Clean Prod
608	240:118070. https://doi.org/10.1016/j.jclepro.2019.118070
609	Lou W, Kane A, Wolbert D, et al (2017) Study of a photocatalytic process for removal
610	of antibiotics from wastewater in a falling film photoreactor: Scavenger study and
611	process intensification feasibility. Chem Eng Process Process Intensif 122:213-
612	221. https://doi.org/10.1016/j.cep.2017.10.010
613	Medarević DP, Kleinebudde P, Djuriš J, et al (2016) Combined application of mixture
614	experimental design and artificial neural networks in the solid dispersion
615	development. Drug Dev Ind Pharm 42:389-402.
616	https://doi.org/10.3109/03639045.2015.1054831
617	Miki-yoshida M (2016) Optical Band Gap Estimation of ZnO Nanorods a E = B ^ E -
618	Eg h. 19:33-38. https://doi.org/10.1590/1980-5373-mr-2015-0612
619	Mirzaei R, Yunesian M, Nasseri S, et al (2018) Occurrence and fate of most
620	prescribed antibiotics in different water environments of Tehran, Iran. Sci Total
621	Environ 619–620:446–459. https://doi.org/10.1016/j.scitotenv.2017.07.272
622	Mohammadzadeh Kakhki R, Mohammadpoor M, Faridi R, Bahadori M (2020) The
623	development of an artificial neural network-genetic algorithm model (ANN-GA)
624	for the adsorption and photocatalysis of methylene blue on a novel sulfur-
625	nitrogen co-doped Fe2O3 nanostructure surface. RSC Adv 10:5951-5960.
626	https://doi.org/10.1039/c9ra10349j
627	Mostafaloo R, Mahmoudian MH, Asadi-ghalhari M (2019) BiFeO 3 / Magnetic
628	nanocomposites for the photocatalytic degradation of ce fi xime from aqueous
629	solutions under visible light. J Photochem Photobiol A Chem 382:111926.
630	https://doi.org/10.1016/j.jphotochem.2019.111926
631	Nguyen CH, Fu CC, Juang RS (2018) Degradation of methylene blue and methyl
632	orange by palladium-doped TiO2 photocatalysis for water reuse: Efficiency and
633	degradation pathways. J Clean Prod 202:413-427.
634	https://doi.org/10.1016/j.jclepro.2018.08.110
635	Nithya R, Ayyappan S (2020) Novel exfoliated graphitic-C3N4 hybridised ZnBi2O4
636	(g-C3N4/ZnBi2O4) nanorods for catalytic reduction of 4-Nitrophenol and its

637	antibacterial activity. J Photochem Photobiol A Chem 398:.
638	https://doi.org/10.1016/j.jphotochem.2020.112591
639	Raou D (2017) Synthesis and microstructural properties of ZnO nanoparticles
640	prepared by precipitation method Synthesis and microstructural properties of
641	ZnO nanoparticles prepared by precipitation method. Renew Energy 50:932–
642	937. https://doi.org/10.1016/j.renene.2012.08.076
643	Sabri K, Rais A, Taibi K, et al (2016) Structural Rietveld refinement and vibrational
644	study of MgCrxFe2-xO4 spinel ferrites. Phys B Condens Matter 501:38–44.
645	https://doi.org/10.1016/j.physb.2016.08.011
646	Sample I (2018) Calls to rein in antibiotic use after study shows 65% increase
647	worldwide. In: Guard. https://www.theguardian.com/science/2018/mar/26/calls-
648	to-rein-in-antibiotic-use-after-study-shows-65-increase-worldwide. Accessed 26
649	Mar 2018
650	Sayadi MH, Sobhani S, Shekari H (2019) Photocatalytic degradation of azithromycin
651	using GO@Fe3O4/ ZnO/ SnO2 nanocomposites. J Clean Prod 232:127–136.
652	https://doi.org/10.1016/j.jclepro.2019.05.338
653	Serpone N (2006) Is the band gap of pristine TiO2 narrowed by anion- and cation-
654	doping of titanium dioxide in second-generation photocatalysts? J Phys Chem B
655	110:24287-24293. https://doi.org/10.1021/jp065659r
656	Sheydaei M, Shiadeh HRK, Ayoubi-Feiz B, Ezzati R (2018) Preparation of nano N-
657	TiO2/graphene oxide/titan grid sheets for visible light assisted photocatalytic
658	ozonation of cefixime. Chem Eng J 353:138–146.
659	https://doi.org/10.1016/j.cej.2018.07.089
660	Shokoohi R, Leili M, Dargahi A, et al (2017) Common Antibiotics in Wastewater of
661	Sina and Besat Hospitals, Hamadan, Iran. Arch Hyg Sci 6:152-159.
662	https://doi.org/10.29252/archhygsci.6.2.152
663	Shooshtari NM, Ghazi MM (2017) An investigation of the photocatalytic activity of
664	nano A-Fe2O3/ZnO on the photodegradation of cefixime trihydrate. Chem Eng
665	315:527–536. https://doi.org/10.1016/j.cej.2017.01.058
666	Sousa OM, Lima JS, Lima AF, Lalic M V. (2019) Theoretical study of structural,

667	electronic and magnetic properties of the spinel Co 3 O 4 under the pressure
668	from 0 to 30 GPa. J Magn Magn Mater 484:21–30.
669	https://doi.org/10.1016/j.jmmm.2019.03.122
670	Tabatabai-Yazdi FS, Ebrahimian Pirbazari A, Esmaeili Khalilsaraei F, et al (2019)
671	Photocatalytic treatment of tetracycline antibiotic wastewater by silver/TiO2
672	nanosheets/reduced graphene oxide and artificial neural network modeling.
673	Water Environ Res. https://doi.org/10.1002/wer.1258
674	Tho NTM, Khanh DNN, Thang NQ, et al (2020) Novel reduced graphene
675	oxide/ZnBi2O4 hybrid photocatalyst for visible light degradation of 2,4-
676	dichlorophenoxyacetic acid. Environ Sci Pollut Res 27:11127–11137.
677	https://doi.org/10.1007/s11356-020-07752-1
678	Viruthagiri G, Kannan P (2019) Visible light mediated photocatalytic activity of cobalt
679	doped Bi 2 O 3 nanoparticles. J Mater Res Technol 8:127–133.
680	https://doi.org/10.1016/j.jmrt.2017.06.011
681	World Health Organisation (2011) Tackling antibiotic resistance from a food safety
682	perspective in Europe. World Heal Organ 1-88
683	Xiong W, Zeng G, Yang Z, et al (2018) Adsorption of tetracycline antibiotics from
684	aqueous solutions on nanocomposite multi-walled carbon nanotube
685	functionalized MIL-53(Fe) as new adsorbent. Sci Total Environ 627:235-244.
686	https://doi.org/10.1016/j.scitotenv.2018.01.249
687	Yang J, Xie T, Liu C, Xu L (2018) Facile fabrication of dumbbell-like β-Bi2O3/
688	graphene nanocomposites and their highly efficient photocatalytic activity.
689	Materials (Basel) 11:. https://doi.org/10.3390/ma11081359
690	Zhang FX, Saxena SK (2013) Raman studies of Bi 2 Cu O 4 at high pressures.
691	141926:2-5. https://doi.org/10.1063/1.2189450
692	Zhang H, Ke H, Ying P, et al (2018) Crystallisation process of Bi5Ti3FeO15
693	multiferroic nanoparticles synthesised by a sol-gel method. J Sol-Gel Sci
694	Technol 85:132–139. https://doi.org/10.1007/s10971-017-4530-9
695	Zhang X, Zhang Y, Ngo HH, et al (2020) Characterization and sulfonamide antibiotic
696	adsorption capacity of spent coffee grounds based biochar and hydrochar. Sci

697	Total Environ 716:. https://doi.org/10.1016/j.scitotenv.2020.137015
698	Zhou C, Wang Q, Zhou C (2020) Photocatalytic degradation of antibiotics by
699	molecular assembly porous carbon nitride: Activity studies and artificial neural
700	networks modeling. Chem Phys Lett 750:137479.
701	https://doi.org/10.1016/j.cplett.2020.137479
702	Zhou F, Yan C, Liang T, et al (2018) Photocatalytic degradation of Orange G using
703	sepiolite-TiO2 nanocomposites: Optimization of physicochemical parameters
704	and kinetics studies. Chem Eng Sci 183:231–239.
705	https://doi.org/10.1016/j.ces.2018.03.016
706	Zhu H, Chen T, Liu J, Li D (2018) Adsorption of tetracycline antibiotics from an
707	aqueous solution onto graphene oxide/calcium alginate composite fibers. RSC
708	Adv 8:2616–2621. https://doi.org/10.1039/c7ra11964j
709	Zulfiqar M, Samsudin MFR, Sufian S (2019) Modelling and optimization of
710	photocatalytic degradation of phenol via TiO2 nanoparticles: An insight into
711	response surface methodology and artificial neural network. J Photochem
712	Photobiol A Chem 384:112039.
713	https://doi.org/10.1016/j.jphotochem.2019.112039
714	Zuo S, Chen Y, Liu W, et al (2017) A facile and novel construction of
715	attapulgite/Cu2O/Cu/g-C3N4 with enhanced photocatalytic activity for antibiotic
716	degradation. Ceram Int 43:3324–3329.
717	https://doi.org/10.1016/j.ceramint.2016.11.173

718

Table 1 ANN model parameters.

Parameters	Value
Input Neurons	pH, time, dose CBO,
	concentration of CFX
Output Neurons	Degradation of CFX
Number of Hidden Layer	(2-20)
	selection based on
	minimum MSE
Training function	Levenberg-Marquardt
Number experiment points	40 points
Performance function	MSE
Data division	70%-15%-15%

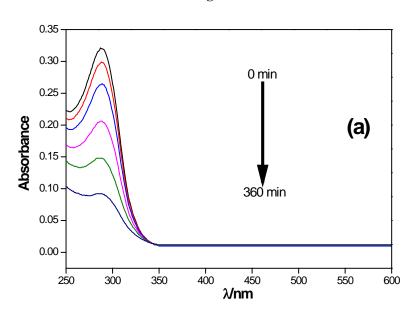
Table 2 Data sets of experimental results and ANN's predicted results.

Run	Time (min)	pН	CFC	photocatalyst (g/l)	photodegradation (%)	
	(11111)		(ppm)	(g/1)	Actual	Predicted
1	30	6	10	1	0.1281	0.1303
2	360	6	15	1	0.5471	0.5501
3	120	6	10	1	0.4812	0.4831
4	30	6	15	0.75	0.1065	0.1087
5	240	6	10	1	0.6813	0.6816
6	240	6	12.5	1	0.5814	0.5742
7	360	6	10	1	0.7031	0.699
8	60	6	15	1.5	0.1238	0.1229
9	240	4	10	1	0.5378	0.5382
10	300	8	10	1	0.6216	0.6205
11	360	8	10	1	0.6445	0.6484
12	300	6	15	1.5	0.4857	0.4864
13	180	4	10	1	0.4538	0.454
14	300	6	15	0.75	0.5017	0.5027
15	60	4	10	1	0.1387	0.1403
16	120	6	15	1.5	0.2286	0.2286
17	360	4	10	1	0.5588	0.5753
18	120	6	15	0.5	0.3464	0.3529
19	360	6	15	1.5	0.5238	0.5521
20	60	6	12.5	1	0.2093	0.2002
21	180	6	15	0.5	0.3954	0.4005
22	120	6	15	1	0.3134	0.3162
23	120	6	15	0.75	0.3471	0.3442
24	360	6	15	0.5	0.4444	0.4585
25	300	6	15	0.5	0.45	0.4448
26	360	6	15	0.75	0.502	0.504
27	300	6	15	1	0.538	0.5336
28	60	8	10	1	0.1972	0.1965
29	60	6	15	1	0.183	0.1828
30	180	8	10	1	0.4954	0.4832
31	60	6	5	1	0.4468	0.4464
32	300	4	10	1	0.5588	0.5592
33	240	6	15	1	0.4928	0.503
34	30	6	5	1	0.1844	0.282
35	240	6	5	1	0.8227	0.8208
36	300	6	10	1	0.6969	0.7056
37	360	6	5	1	0.8227	0.824
38	180	6	10	1	0.6	0.6174
39	60	6	15	0.5	0.2647	0.2456
40	60	6	10	1	0.2469	0.2497

Table 3 ANN model's weight and bias matrix.

Neurons of hidden	Weights and biases between input and hidden layers Input variables				Weights and biases between hidden and output layers		
layer	Radiation time	рН	Initial concentration of CFX	Catalyst dosage	Bias	Output Degradation Rate %	Bias
1	-0.209	-0.155	-2.371	-1.364	2.586	0.258	
2	-0.524	-1.142	-1.34	-1.544	2.318	-0.514	
3	-1.669	1.438	1.281	-1.376	1.922	0.081	
4	1.204	1.295	-1.615	0.709	-0.558	0.515	
5	-1.58	0.183	1.387	-1.792	1.284	0.217	
6	-1.005	-1.551	-1.256	1.273	0.359	0.117	
7	1.407	-1.721	-0.843	-0.515	-0.531	-0.157	-0.29
8	-0.0028	-1.191	-1.811	-0.459	-0.645	0.3	
9	-1.107	-1.116	-0.201	-1.503	-0.872	-0.277	
10	-0.5608	1.804	-0.763	1.813	-1.227	-0.721	
11	2.053	0.782	-0.303	-0.83	1.619	0.71	
12	1.637	-0.136	-1.36	-1.151	2.809	-0.614	
13	1.435	0.1417	-1.93	-1.541	2.271	0.231	

Fig. 1



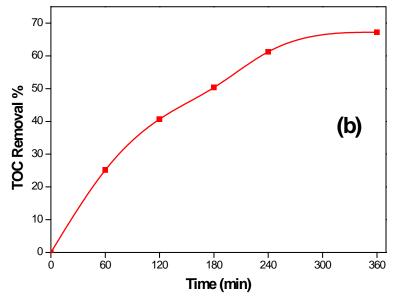


Fig. 2

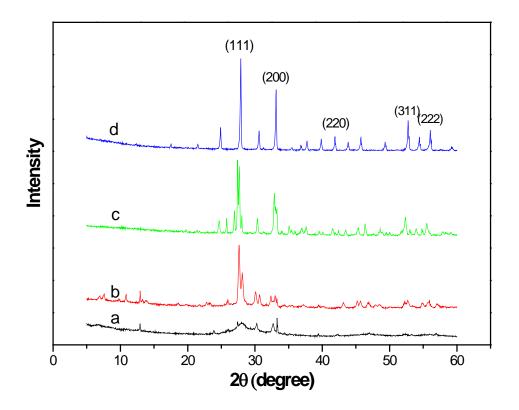


Fig. 3

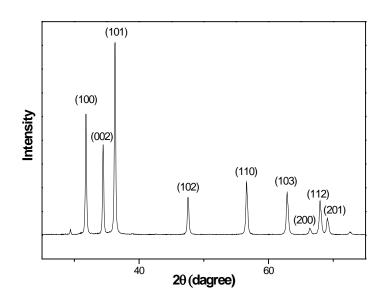


Fig. 4

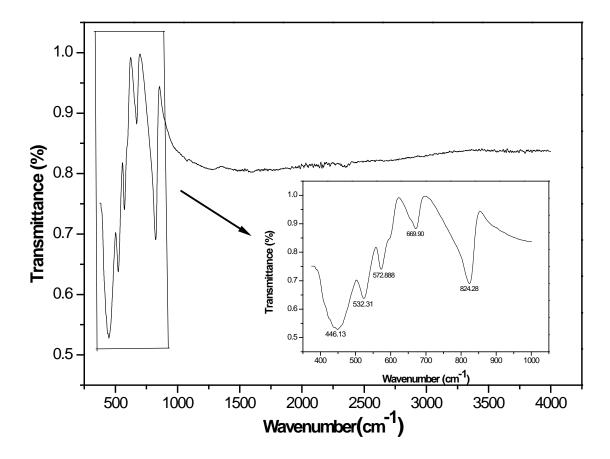


Fig. 5

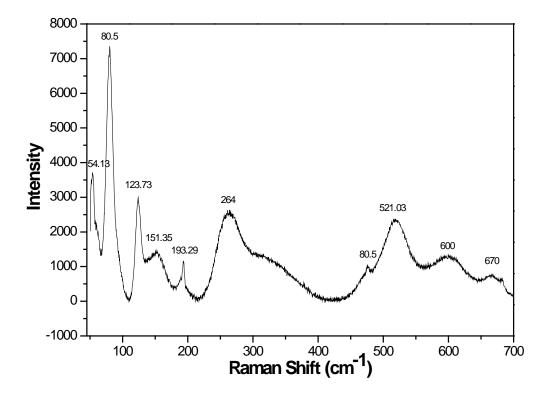


Fig. 6

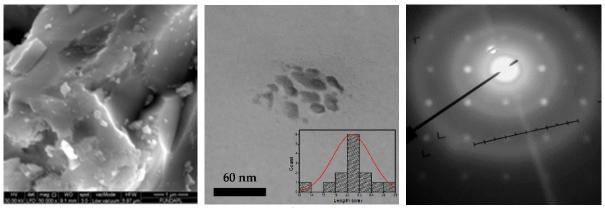
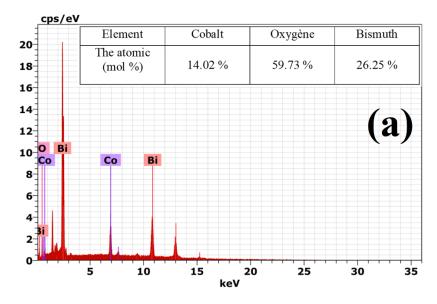


Fig. 7



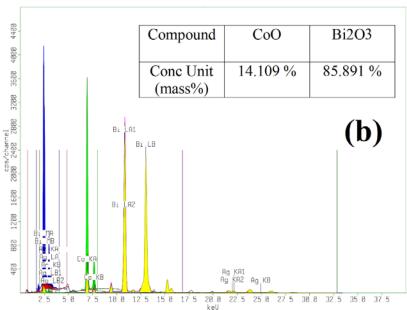
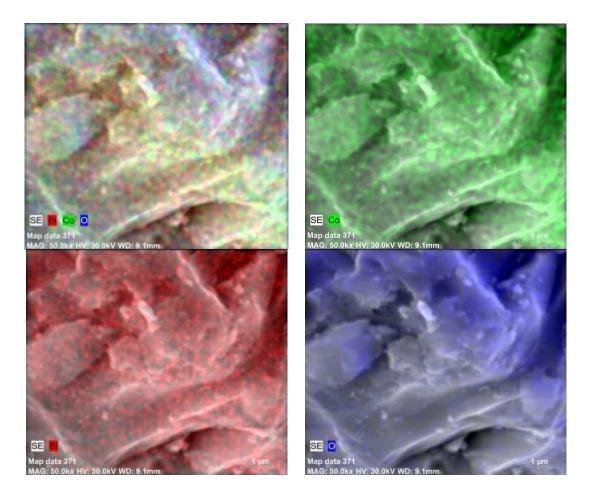
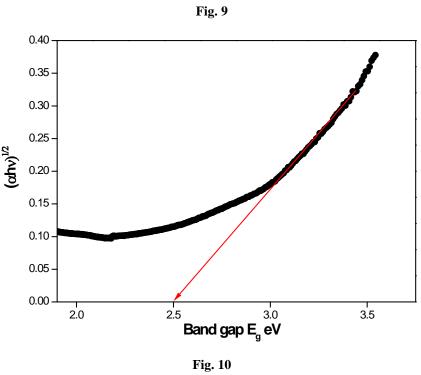


Fig. 8





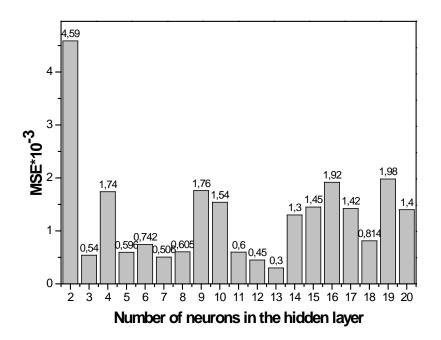


Fig. 11

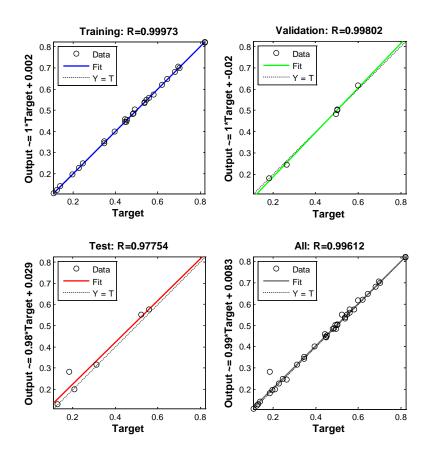


Fig. 12

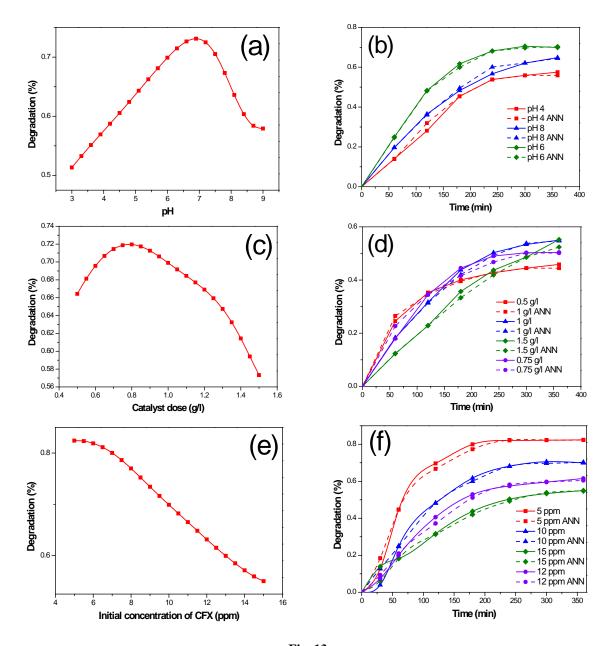


Fig. 13

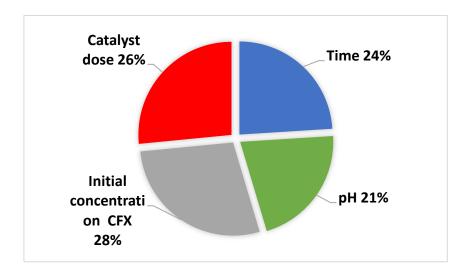


Fig. 14

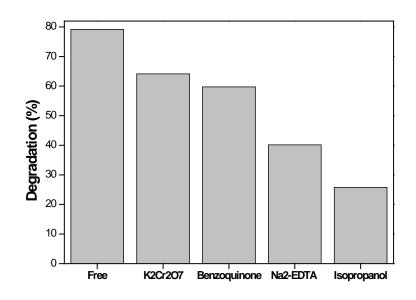


Fig. 15

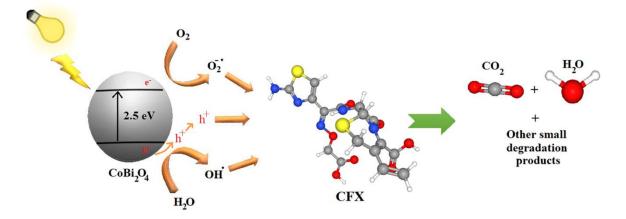


Fig. 16

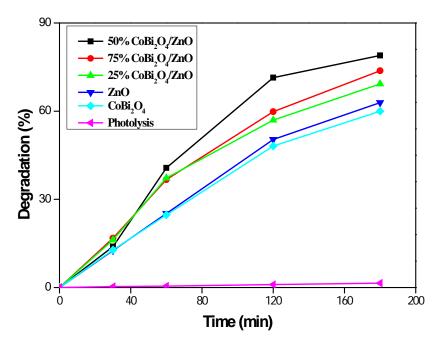


Fig. 17