

Perceptions and responses of Pacific Island fishers to changing coral reefs

Andrew Rassweiler, Matthew Lauer, Sarah E Lester, Sally J Holbrook, Russell J Schmitt, Rakamaly Madi Moussa, Katrina S Munsterman, Hunter S Lenihan, Andrew J Brooks, Jean Wencélius, et al.

▶ To cite this version:

Andrew Rassweiler, Matthew Lauer, Sarah E Lester, Sally J Holbrook, Russell J Schmitt, et al.. Perceptions and responses of Pacific Island fishers to changing coral reefs. AMBIO: A Journal of Environment and Society, 2020, 49 (1), pp.130-143. 10.1007/s13280-019-01154-5. hal-03034158

HAL Id: hal-03034158 https://hal.science/hal-03034158

Submitted on 9 Dec 2020

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

TITLE

3	AUTHORS
4	Andrew Rassweiler ¹ , Matthew Lauer ² , Sarah E. Lester ³ , Sally J. Holbrook ⁴ , Russell J.
5	Schmitt ⁵ , Rakamaly Madi Moussa ⁶ , Katrina S. Munsterman ⁷ , Hunter S. Lenihan ⁸ , Andrew J.
6	Brooks ⁹ , Jean Wencélius ¹⁰ , Joachim Claudet ^{11,12}
7	¹ Assistant Professor, Department of Biological Science, Florida State University, 319
8	Stadium Drive, Tallahassee, FL, 32306, USA.
9	² Professor, San Diego State University – Anthropology, 5500 Campanile Dr, San Diego, CA,
10	92182, USA.
11	³ Assistant Professor, Department of Geography, Florida State University, 113 Collegiate
12	Loop, Tallahassee, FL, 32306, USA.
13	⁴ Professor, Department of Ecology, Evolution, and Marine Biology, University of California,
14	Santa Barbara, CA, 93106, USA.
15	⁵ Professor, Department of Ecology, Evolution, and Marine Biology, University of California,
16	Santa Barbara, CA, 93106, USA.
17	⁶ Post doctoral Fellow, CRIOBE-USR 3278 CNRS-EPHE-UPVD-PSL, Laboratoire
18	d'Excellence CORAIL, BP 1013, Papetoai, 98729 Moorea, French Polynesia.
19	⁷ Masters Student, Department of Ecology, Evolution and Marine Biology, University of
20	California, Santa Barbara, CA, 93106, USA.

Perceptions and responses of Pacific Island fishers to changing coral reefs

21	⁸ Professor, Bren School of Environmental Science and Management, University of
22	California, Santa Barbara, CA, 93106, USA.
23	⁹ Senior Project Scientist, Marine Science Institute, Building 520, University of California,
24	Santa Barbara, CA 93106-6150.
25	¹⁰ Post-doctoral Research Fellow, San Diego State University – Anthropology, 5500
26	Campanile Dr, San Diego, CA, 92182, USA.
27	¹¹ Senior Research Scientist, National Center for Scientific Research, PSL Université Paris,
28	CRIOBE, USR 3278 CNRS-EPHE-UPVD, Maison des Océans, 195 rue Saint-Jacques, 75005
29	Paris.
30	¹² Laboratoire d'Excellence CORAIL, Moorea, French Polynesia.
31	
31 32	ACKNOWLEGEMENTS
31 32 33	ACKNOWLEGEMENTS We thank T. Atger, M. Strother, A. Bunnell, C. Hunter and O. L. Lenihan for leading
31 32 33 34	ACKNOWLEGEMENTS We thank T. Atger, M. Strother, A. Bunnell, C. Hunter and O. L. Lenihan for leading anthropological field work, K. Seydel, J. Verstaan, A. Dubel, P. Germain, L. Thiault and R. Terai
31 32 33 34 35	ACKNOWLEGEMENTS We thank T. Atger, M. Strother, A. Bunnell, C. Hunter and O. L. Lenihan for leading anthropological field work, K. Seydel, J. Verstaan, A. Dubel, P. Germain, L. Thiault and R. Terai for technical assistance, the staff of University of California Berkeley Gump Research Station
31 32 33 34 35 36	ACKNOWLEGEMENTS We thank T. Atger, M. Strother, A. Bunnell, C. Hunter and O. L. Lenihan for leading anthropological field work, K. Seydel, J. Verstaan, A. Dubel, P. Germain, L. Thiault and R. Terai for technical assistance, the staff of University of California Berkeley Gump Research Station including Ms. Hinano Murphy for logistic support, René Galzin for initiating the roadside
 31 32 33 34 35 36 37 	ACKNOWLEGEMENTS We thank T. Atger, M. Strother, A. Bunnell, C. Hunter and O. L. Lenihan for leading anthropological field work, K. Seydel, J. Verstaan, A. Dubel, P. Germain, L. Thiault and R. Terai for technical assistance, the staff of University of California Berkeley Gump Research Station including Ms. Hinano Murphy for logistic support, René Galzin for initiating the roadside surveys and Dr. Jean-Yves Meyer for assistance with permits. We gratefully acknowledge the
 31 32 33 34 35 36 37 38 	ACKNOWLEGEMENTS We thank T. Atger, M. Strother, A. Bunnell, C. Hunter and O. L. Lenihan for leading anthropological field work, K. Seydel, J. Verstaan, A. Dubel, P. Germain, L. Thiault and R. Terai for technical assistance, the staff of University of California Berkeley Gump Research Station including Ms. Hinano Murphy for logistic support, René Galzin for initiating the roadside surveys and Dr. Jean-Yves Meyer for assistance with permits. We gratefully acknowledge the support of the National Science Foundation (OCE 1637396, OCE 1325652, BCS 1714704), the
 31 32 33 34 35 36 37 38 39 	ACKNOWLEGEMENTS We thank T. Atger, M. Strother, A. Bunnell, C. Hunter and O. L. Lenihan for leading anthropological field work, K. Seydel, J. Verstaan, A. Dubel, P. Germain, L. Thiault and R. Terai for technical assistance, the staff of University of California Berkeley Gump Research Station including Ms. Hinano Murphy for logistic support, René Galzin for initiating the roadside surveys and Dr. Jean-Yves Meyer for assistance with permits. We gratefully acknowledge the support of the National Science Foundation (OCE 1637396, OCE 1325652, BCS 1714704), the Gordon and Betty Moore Foundation and Agence Nationale de la Recherche (ANR-14-CE03-
 31 32 33 34 35 36 37 38 39 40 	ACKNOWLEGEMENTS We thank T. Atger, M. Strother, A. Bunnell, C. Hunter and O. L. Lenihan for leading anthropological field work, K. Seydel, J. Verstaan, A. Dubel, P. Germain, L. Thiault and R. Terai for technical assistance, the staff of University of California Berkeley Gump Research Station including Ms. Hinano Murphy for logistic support, René Galzin for initiating the roadside surveys and Dr. Jean-Yves Meyer for assistance with permits. We gratefully acknowledge the support of the National Science Foundation (OCE 1637396, OCE 1325652, BCS 1714704), the Gordon and Betty Moore Foundation and Agence Nationale de la Recherche (ANR-14-CE03- 0001-01). Permits for field work were issued by the Haut-commissariat de la République en 494

42 2009-10, 2010-11, 2011-12, 2012-13, 2013-14, and 2014-15 to RJS and SJH) for research

associated with the US NSF Moorea Coral Reef Long Term Ecological Research project. Service
d'Observation CORAIL from CRIOBE kindly provided ecological monitoring data.

45

46

47

ABSTRACT (150 words maximum- at 149)

48 The transformation of coral reefs has profound implications for millions of people. However, 49 the interactive effects of changing reefs and fishing remain poorly resolved. We combine 50 underwater surveys (271,000 fishes), catch data (18,000 fishes), and household surveys (351 51 households) to evaluate how reef fishes and fishers in Moorea, French Polynesia responded to a 52 landscape-scale loss of coral caused by sequential disturbances (a crown-of-thorns sea star 53 outbreak followed by a category 4 cyclone). Although local communities were aware of the 54 disturbances, less than 20% of households reported altering what fishes they caught or ate. This 55 contrasts with substantial changes in the taxonomic composition in the catch data that mirrored 56 changes in fish communities observed on the reef. Our findings highlight that resource users and scientists may have very different interpretations of what constitutes 'change' in these highly 57 dynamic social-ecological systems, with broad implications for successful co-management of 58 59 coral reef fisheries.

60

61 **KEYWORDS**

Fisheries, Social-ecological feedbacks, Selectivity, Disturbance, Local knowledge, Coral reef
resilience

65 **INTRODUCTION**

66

pollution, sedimentation, ocean acidification, and rising seawater temperatures (Bellwood et al. 67 2004; Hughes et al. 2018), resulting in unprecedented levels of coral mortality (Hughes et al. 68 2017) and shifts from coral-dominated to macroalgae-dominated community states (Rogers and 69 70 Miller 2006). Beyond biodiversity loss, degraded reefs present challenges for millions of coastal dwellers who rely on healthy reef ecosystems for food, income, and their personal and cultural 71 identities. This has prompted research examining how local communities and resource users 72 73 perceive, adapt to, and manage coral reefs in the Anthropocene (McClanahan and Cinner 2012; McMillen et al. 2014), including a focus on adaptive co-management, whereby management is 74 implemented and adapted based on knowledge about feedbacks between resource users and 75 shifting local ecosystems (Hughes et al. 2005). 76

Coral reef ecosystems are under significant anthropogenic pressures from overfishing,

The Pacific Islands region represents an ideal context to investigate how local communities 77 and changing coral reefs interact. Island peoples have shown the capacity to adapt, cope, and 78 79 innovate in the face of social-ecological change, with positive and negative outcomes for coral reef health (Johannes 2002). In some Pacific Islands, such as Fiji, Vanuatu, and the Solomon 80 81 Islands, marine resources have been effectively managed over long periods through periodic fishing ground closures, gear restrictions, and other socially enforced constraints on harvesting 82 (Cinner et al. 2006). Elsewhere, local responses to changing conditions have had negative 83 84 ecological outcomes, as with poison and dynamite fishing (McManus et al. 1997). The effectiveness of adaptive responses is shaped by local cultural values and power relations that 85 inform decision-making and the range of possibilities available (Cinner et al. 2018). 86

87 Effective adaptive management requires that resource users detect or anticipate shifts in the local environment and alter their activities accordingly. Some empirical studies have 88 demonstrated that Pacific islanders can detect rapid shifts in benthic communities disrupted by 89 90 tsunamis (Lauer and Matera 2016), in addition to more gradual changes such as expanding seagrass beds (Lauer and Aswani 2010). Numerous questions remain, however, about the 91 92 sensitivity of local resource users to change, and in particular whether ecosystem disturbances identified by ecologists are similarly perceived by Pacific islanders. 93 We addressed these issues for a small-scale reef fishery on the island of Moorea, French 94 95 Polynesia. Social and ecological surveys explored how communities perceived and responded to changes in fishery resources associated with a crown-of-thorns sea star (COTS) outbreak 96 followed by a destructive cyclone. In 2004, coral cover around Moorea was near the highest 97 levels observed in the past half century (Trapon et al. 2011), but an outbreak of corallivorous 98 COTS that peaked in 2009, followed by Cyclone Oli in early 2010, reduced live coral cover by > 99 95% (Adam et al. 2011; Trapon et al. 2011; Adam et al. 2014; Lamy et al. 2015). Dead coral 100 101 skeletons and cleared reef substrates provided substantial free space for growth of macroalgae, raising the possibility that a macroalgal phase shift could occur. However, benthic community 102 103 changes were rapidly followed by changes in the fish assemblage, with roving herbivorous fishes such as parrotfishes doubling in density and tripling in total biomass (Han et al. 2016), thus 104 preventing macroalgae from establishing. Moreover, in the years since the disturbances, coral 105 106 cover has increased and even exceeds pre-disturbance levels in some areas (Holbrook et al. 2018). Despite intensive ecological study, it is not known if these changes in the fish 107 assemblages have altered fishable resources, the activities of reef fishers, or how local people 108 109 perceived the changes. Because spearfishing – a highly selective method – is common in

Moorea, a shift in the abundances of fishable resources provides an opportunity to assesswhether fishers alter what they catch as their resource environment changes.

This study addressed four questions: 1) How did residents of Moorea perceive the shifts documented in ecological studies? 2) Do they report changing their fishing behavior or seafood consumption in response to the shift? 3) How did the changes in the fish assemblage affect the availability and taxonomic composition of fishable biomass? 4) Is there evidence for changes in fishing behavior (such as taxonomic selectivity) over time?

To answer these questions, we conducted 351 household surveys documenting fishers' perceptions of the changes and their potential responses via alteration in fishing practices or fish consumption. We analyzed a time series of catch data (~18,000 identified and measured fishes) collected before and after the disturbances, spanning a nine year time period, to determine changes in targeted fish species and sizes, including key groups of herbivores crucial to recovery and resilience of the coral state. Finally, we compared the catch data with extensive surveys that estimated abundance and biomass of fishes on the reef throughout the same time frame.

124 METHODS

125 *Ecological and social context*

Moorea (17° 32' S, 149° 50' W) is a volcanic 'high' island 60 km in perimeter with an offshore barrier reef that encloses a shallow lagoon (Figure 1A). The island has three types of reef habitats: within the lagoon there are fringing reefs and back reefs, while outside the barrier reef crest there is a steeply sloping fore reef. Moorea has over 17,000 inhabitants (Institut de la statistique de la Polynésie française 2012) residing in five *communes associées*: Afareaitu, Ha'apiti, Paopao, Papetoai, and Teavaro. It has undergone substantial economic development over the past half-century, including becoming a major international tourist destination.

Communal land has been supplanted by private land ownership and the state declared that all 133 134 lagoon and marine areas are public property, meaning that customary sea tenure is non-existent. Reefs in Moorea continue to be the focus of widespread fishing activity, although major 135 economic and social changes have shifted household livelihoods away from direct dependence 136 on marine resources for food or income towards wage-earning employment. Over half of 137 households fish, with free-dive spearfishing as the preferred method (Leenhardt et al. 2016). 138 Most people fish so they can eat and share fresh reef fishes, a fundamental marker of Polynesian 139 life. Reef fishes constitute the bulk of the catch and are prized as symbols of Polynesian identity 140 141 and cultural pride. It is notable that Moorea's households are less dependent on marine resources for food security or income than is common in other regions in the Pacific. As citizens of France 142 they have access to state-subsidized healthcare, education, and social services, so poverty levels 143 are lower than in most of Oceania. Although most households contain fishers, only a small 144 number of fishers fish full-time solely for income. 145

146 Household surveys and key informant interviews

In 2014-15, we interviewed 351 (approximately 20%) households in the communes of 147 Afareaitu, Papetoai and Haapiti. In each commune, we identified village boundaries, within 148 149 which we conducted a blend of random and convenience sampling. The 60 to 80 min survey interviews were conducted in French or Tahitian, with local Tahitians assisting in the surveys 150 and translating for household heads more comfortable speaking in Tahitian. Interview topics 151 152 included basic demographic information, fishing effort, livelihoods, catch preferences, consumption patterns, and perceptions of resource conditions. Standardized questions allowed 153 for later comparison, but more open-ended questions were used to discuss important issues and 154 155 perceptions. Sample size for the standardized questions varied, since not every question was

relevant for all respondents. We also conducted 15 semi-structured interviews with fishers from around the island who were considered highly knowledgeable local experts. In 2018, follow-up interviews were carried out with nine key informants to whom results from this paper were presented. Questions explored respondents' perceptions of post-disturbance changes in the fish assemblages.

161 Fish-seller surveys

The sale of most reef fishes takes place from small roadside stands along the perimeter road of the island (which has no fish markets). Fresh reef fishes are strung through the gills and hung from racks (Figure 1B, S1). Each string of fish is sold as a unit, known in Tahitian as a *tui*. A seller, often the fisher, assembles each *tui* and 10 or more may be hung for sale. Any single *tui* may contain a few larger fishes or many small ones of different species. Most fish stands are active early in the morning and by mid-morning most have sold their catch.

168 To sample the fishes being sold at these roadside stands, a researcher drove Moorea's ring

169 road early in the morning on weekends, typically the busiest times for fish sales. At each stand,

the rack of *tui* was photographed with a scale bar of known size (0.5m) and the seller was briefly

171 interviewed. Photographed fishes were later identified to the lowest taxonomic level possible and

the length of each was estimated by comparison with the scale bar (Schneider et al. 2012).

173 Catch surveys were conducted in 5 different years during 2007-15 (2007, 2008, 2012, 2014,

174 2015; Table 1). Three of the five *communes associées* were sampled in all five years (Afareaitu,

175 Paopao, and Teavaro), and so only data from these regions were analyzed to maintain consistent

176 geographical coverage through time, with data pooled across regions in all analyses.

177 *Reef surveys*

We assessed reef fish populations using data from the NSF-funded Moorea Coral Reef Long 178 Term Ecological Research (MCR LTER) project that collects time series data at 18 locations 179 180 around Moorea (Brooks 2017), and the SO CORAIL-PGEM monitoring program that collects data from 13 locations around the island (Lamy et al. 2015). Here we used data collected 181 182 annually from 2007 to 2015 (Table 1), and included data only from transects located on reefs offshore of the three focal *communes* (Afareaitu, Paopao, Teavaro) as most targeted fishes are 183 territorial and most fishers fish near where they live. 184 185 The MCR LTER surveys are conducted by SCUBA divers between 0900 and 1600 hours during late July or early August. Abundances of all mobile taxa of fishes observed are recorded 186 on fixed 5 m x 50 m transects that extend from the surface of the reef through the water column. 187 188 The abundances of all non-mobile or semi-cryptic taxa of fishes are also counted along the same transect lines in a 1 m wide transect. The total length of each fish observed is estimated to the 189 nearest 0.5 cm. The SO CORAIL-PGEM monitoring program has sampled similar habitats in 190 191 each of these years, counting and estimating sizes of all fishes within 5 m x 25 m transects. Fish biomass (kg) is calculated based on species-specific scaling parameters (Brooks 2011). 192

193 Fishing selectivity and fishable biomass

Spearfishing is a highly selective fishing method in which the size and species of targets can be observed before they are harvested. We tested for selectivity in size by comparing the fishes being sold by the roadside to the sizes of fishes observed during reef surveys (pooling data across the five years for which we have catch data). We defined a minimum fishable size (15 cm) across all species based on sizes observed in the catch (<2% of fishes were below this size). 199 We determined which taxa were targeted based on the relative abundances of each genus 200 observed in the catch and on the reef. We defined fished taxa as genera making up more than 0.1% of the total catch, which included 23 genera, constituting 99% of all fishes and 95% of all 201 202 biomass being sold. Parrotfishes from the genera Scarus and Chlorurus were combined in all analyses because species from these genera often could not be reliably distinguished in our 203 photographs of *tuis*. We note that some excluded species may be highly prized but rare in the 204 catch because they are rare on the reef. Subsetting the ecological survey data based on our list of 205 23 targeted genera and the minimum fishable size, we calculated how the total fishable biomass 206 207 and the fishable biomass of different targeted groups changed from 2007 to 2015.

208 *Taxonomic composition of the catch*

We evaluated the degree to which variation in the biomass of each taxon on the reef predicts 209 210 variation in the taxonomic composition of the catch by comparing the relative biomass of the seven most common taxa in the catch with their relative biomass on the reef. We excluded 211 soldierfishes (*Myripristis spp.*) from this analysis because they are nocturnal and were poorly 212 213 sampled in our (diurnal) reef surveys, when they shelter within reef structures. Other species may shift habitats on a daily cycle, but any such movements are well within the spatial scale of our 214 215 sampling. Because sampling effort of the catch (during roadside surveys) was not consistent over time, we cannot determine how total catch changed. 216

217 **RESULTS**

218 Household surveys and interviews

The household surveys revealed that a substantial majority of households reported regular consumption of fish, with 67% reporting that they eat fish at least three times per week, and more than half of those eating fish six to seven times per week. Most households (76%) reported at least one member who actively participated in the local reef fishery. There was great
consistency in the species that households preferred to eat and preferred to catch (Table 2). All of
these species are commonly caught and highly prized for the taste and texture of their meat. An
exception to the focus on reef fishes is tuna (*thon* in French), which has become an increasingly
important component of diets in Moorea but which is caught by a small number of pelagic
fishers operating with specialized boats offshore.

There was considerably more variability in how households reported any changes in their 228 behavior in response to the outbreak of COTS (taramea in Tahitian) and the cyclone (Table 3). 229 230 Although 40% remembered the COTS outbreak and 100% remembered Cyclone Oli, few reported modifying the kinds of fishes they ate or bought (1.5% and 10% respectively). Of those 231 that reported responding to the COTS outbreak, responses included removing COTS from the 232 reefs (30%), avoiding fishing in COTS-dominated areas (18%), or changing their fishing areas 233 (6%). Of those that reported responding to Cyclone Oli, responses included waiting until the 234 lagoon was clean from runoff before resuming fishing (30%), fishing in different locations 235 236 because the fishes moved to different areas of the lagoon (16%), fishing less in the lagoon than prior to the cyclone (13%), or fishing less overall after the cyclone (10%). 237

In-depth interviews with expert fishers revealed that they are aware of COTS outbreaks and they recognize that COTS kill coral. Two expert fishers described how in the past, parts of the seastars' bodies were applied as garden pesticide. Other expert fishers mentioned that the Papetoai school and local fisher organizations (in Haapiti and Afareaitu) organized outings where local people removed COTS from the reefs. One fisher noted that this practice was "a new thing" and that "the oldtimers never mentioned this kind of practice happening in the past." Most fishers acknowledged a relationship between live coral cover and reef fish abundance. However, few indicated that the dramatic loss of live coral cover caused by the COTS outbreak or Cyclone
Oli had an impact on the composition of fish assemblages or the relative abundance of the main
targeted taxonomic groups.

248 *Fishing selectivity*

Roadside fish sellers mostly caught fishes on the reef (77%), largely from the lagoon (69%),
and the most common gear used was the spear gun (83%), followed by fishing with nets (11%)
and hook and line (5%). Fishes sold in the morning were mostly caught at night (90% between
1800 and 0600 hrs), so our surveys of fishes sold by the roadside (hereafter, "the catch") may not
be representative of fishing activities undertaken at other times.

Fishes in the catch represent a nonrandom distribution of sizes relative to those observed on 254 the reef (Figure 2). Harvested fishes were significantly larger on average than fishes on the reef 255 (23 cm vs. 8 cm; P < 0.0001, Wilcoxon rank-sum test). More than 98% of fishes in the catch 256 were at least 15 cm in length, suggesting this is a minimum bound on the size of fishes that are 257 targeted. The relative abundance of taxa observed in the catch also diverged substantially from 258 259 the community found on the reef, even when only individuals of fishable size were considered. More than 99% of the fishes in the catch were from 23 genera (Table 4) with almost 60% of the 260 261 catch made up of unicornfishes (*Naso spp.*), parrotfishes (*Scarus* and *Chlorurus spp.*), soldierfishes (Myripristis spp.), and rabbitfishes (Siganus spp.). The composition of the catch 262 contrasted with the most abundant taxa on the reef (based on fishable sized individuals; Tables 4, 263 S1). In particular, while Scarids and Naso were both abundant on the reef, Myripristis and 264 *Siganus* were rarely observed in the reef surveys (the 38th and 29th most abundant taxa, 265 respectively; Table S1). Furthermore, several of the most abundant taxa on the reef were 266

267 completely absent from the catch, most notably surgeonfishes from the genus *Ctenochaetus*268 (25% of fishes on the reef).

269 Fishable biomass

The amount of fishable biomass (fishes >15 cm in length from 23 targeted genera) on the reef 270 was relatively stable from 2007-2015. Although there was some variation from year to year 271 (Figure 3), including a spike in 2010, there was no sustained shift in fishable biomass coinciding 272 with the disturbances that occurred in 2009-2010. By contrast, there was substantial change in 273 the abundance of some taxonomic groups on the reef over the time period. Most dramatically, 274 *Naso* biomass fell from 21 kg ha⁻¹ to about 4 kg ha⁻¹. This decline was offset to some degree by 275 an increase in the biomass of parrotfishes of the genus Scarus. While the biomass of other taxa 276 varied substantially from year to year, there was no apparent secular trend in their abundances. 277

278 *Taxonomic composition of the catch*

The changes in the taxonomic composition on the reef were roughly mirrored by trends in the 279 catch (Figure 4). For example, Naso comprised more than a third of the catch prior to the 280 281 disturbances, and less than 10% after. By contrast, the proportion of the catch composed of parrotfishes from the genera Chlorurus and Scarus increased over time from 56% to 66%. Naso, 282 283 Chlorurus and Scarus collectively composed the bulk of the fishable biomass on the reef (48-66%) and a roughly similar total proportion of the catch (43-65%). 284 For the taxa that were well-sampled in our reef surveys, there was a surprisingly high 285 286 correlation between the biomass of each taxon on the reef and its annual contribution to the catch, with high correlations observed for the most common taxa (Figure 5). Indeed, the 287

correlation for unicornfishes is above 0.99, which suggests both that our reef surveys captured

variation in their abundance over time and that the variation in abundance within the ecologicalcommunity may explain the observed pattern of variation in catch.

291 **DISCUSSION**

In this study we couple data from intensive sampling of both the ecological community and 292 human resource users to provide new insights on how fishes and Pacific Island fishing 293 communities interact during periods of substantial ecological change, and how the fishing 294 communities perceive the changes. Each method provided a different view on these feedbacks. 295 Household surveys confirmed that residents of Moorea were aware of the major disturbances that 296 297 occurred on the reef, but revealed that little explicit change occurred in fishing behavior or perceptions of resources harvested. This contrasts with the marked shifts in the taxonomic 298 composition of the catch that we observed, particularly the significant decrease of Naso spp., one 299 of the most highly prized fishes due to its palatability. Those taxonomic shifts mirrored changes 300 we observed in fish communities on the reef, implying that the composition of the catch is highly 301 dependent on reef state despite the high selectivity of the fishery and local perceptions that 302 303 fishing and fished resources had not changed.

304 *Fishing selectivity*

Our results revealed high selectivity in the Moorea reef fishery, both in terms of body size and taxonomy, consistent with observations of other spearfishing-focused fisheries in the Pacific (Dalzell et al. 1996). Fishers showed a preference for fishes that are larger on average than those encountered on the reef. Even when size selectivity was accounted for, we found strong taxonomic selectivity for a handful of taxa, with some being disproportionately abundant in the catch relative to their abundance on the reef (e.g., *Naso spp.* and *Myripristis spp.*) while others were greatly under-represented in the catch (e.g., *Ctenochaetus spp.*). This high degree of size and taxonomic selectivity is not surprising given the prevalence of spearfishing on the island.
Spearfishers visually identify and evaluate each fish before it is harvested (Frisch et al. 2008).
The resultant selectivity affords them greater latitude for adapting to ecological shifts than other
capture techniques, such as hook and line or gill netting, in which the fishes are invisible to the
fisher before capture.

The suite of preferred species on Moorea is not limited to larger-bodied species. Soldierfishes 317 (*Myripristis*), for example, are relatively small-bodied but represent the third most fished genus 318 (in terms of numbers and biomass in the catch), as they are prized for the taste and the texture of 319 their meat rather than their large filets. In commercially-oriented fisheries, size selectivity can be 320 linked to higher market demand or value for fishes of particular sizes, e.g., large enough to filet 321 or sized to fit on a dinner plate (Reddy et al. 2013). In Moorea spearfishers commonly describe 322 their fishing decisions through idioms of cooking and eating, and will seek out certain species 323 based on how they want to cook their meal that day, underscoring the non-economic nature of 324 the fishery. 325

326 Elsewhere, Pacific Islanders commonly target piscivores, such as emperors and groupers, but in fisheries where spearfishing is the primary mode of capture, herbivorous fishes such as 327 328 unicornfishes and parrotfishes often dominate the catch (Jennings and Polunin 1995; Gillett and Moy 2006). Contemporary reef fish preferences in Moorea may be the result of the gear type 329 used or an outcome of overfishing and fishing down the food web (Pauly et al. 1998) from 330 331 piscivores to herbivores. More historical work could shed light on this possibility by detailing the trajectory of taxonomic selectivity over the last several centuries. We also note that Moorea 332 333 fishers show a strong selectivity against harvesting *Ctenochaetus* and *Acanthurus (maito* in 334 Tahitian) even though they are some of the most abundant species on the reef. These fish are

known to be ciguatoxic, and the sale of *Ctenochaetus* was banned by the territorial governmentin the 1960s (Walter 1968).

337 *Taxonomic composition on the reef and in the catch over time*

Our roadside surveys indicate that the taxonomic composition of the catch shifted 338 339 substantially after the disturbance (Figure 4). Changes in the catch largely correlated with shifts in the taxonomic composition of the reef community, particularly for species that made up a 340 substantial proportion of the catch (Figure 5). However, there is wide variation in the strength of 341 this relationship. The unexplained variation may stem from analyzing catch at the genus level, 342 343 likely combining species of different desirability within the same category. For example, dynamics on the reef and in the catch were poorly correlated for Acanthurus. There are 5 species 344 commonly observed in the catch within this genus; if some of these are targeted and some are not 345 (possibly based on ciguatera risk), then trends in the biomass of the genus on the reef may not 346 represent trends in the preferred species within that genus, obscuring a tighter relationship at the 347 species level. By contrast, one species (*Naso lituratus*) makes up more than 90% of the fishable-348 349 size individuals of that genus on the reef, so variation in the abundance of that species translates more directly to our genus-level analyses. 350

The composition of the catch is a joint product of the availability of resources and the demand for each from the fishing communities. If the catch primarily reflects demand for different species, we might expect to see little change in the composition of the catch as the ecosystem changes, particularly in such a highly selective fishery. Instead, the high correlations between biomass on the reef and in the catch for unicornfishes (*Naso spp.*) and parrotfishes (*Scarus spp./Chlorurus spp.*) indicate that shifting relative abundances result in different compositions of the catch, and suggest that there is considerable flexibility in harvest and consumption behavior. 358 Perceptions of change

Our household surveys and key informant interviews suggest that Moorea's fishers generally 359 were aware of the COTS outbreak and Cyclone Oli and that they understood the ecological 360 impacts of these disturbances. This in-depth understanding is not surprising given most engage in 361 fishing on a regular basis and thus have frequent experiential contact with the marine 362 environment. It is widely acknowledged that in Pacific Island contexts where communities 363 depend on marine resources, islanders maintain rich, site-specific knowledge of the marine 364 environment as well as sophisticated understanding of ecological processes (Johannes 1981; 365 366 Lauer 2017). Despite their awareness of the disturbances, few households saw these as a change that warranted modification of their fishing strategies, or altering what species of fish they ate. 367 This narrative is in striking contrast to the shifts we documented with our roadside surveys 368 369 conducted before and after the disturbances. Most surprisingly, the significant decrease of *Naso spp.* in the reef counts, while reflected in the catch, was not expressed in informants' responses. 370 There are several possible explanations for this apparent discrepancy. For one, the relative 371 372 abundances of species shifted after the disturbances but the suite of species caught did not, with 373 the same top five species caught before and after the disturbances. It may be that Moorea fishers 374 would only report a more radical shift (e.g., the complete disappearance of a targeted fish) in the taxonomic composition of their diet and catch. Furthermore, fishers speak less of shifts in 375 abundance per se but rather about changes in fishes' behavior and their habitat choices. When 376 377 asked about the decline in abundance of *Naso spp*. in the catch surveys, several fishers stated that unicornfishes have learned, as a result of heavy fishing pressure, to swim into deeper waters. Yet 378 these behavioral changes of *Naso spp.* do not necessarily result in fewer fish caught for the best 379 380 spearfishers. As one fisher stated, "a good spearfisher will find and catch the fish he desires."

381 The discrepancy between what constitutes noteworthy changes for Moorean fishers and western scientists could also be related to the different ways each group conceptualizes marine 382 environments (Johannes 1981; Hviding 1996). Ethnographic material indicates that Pacific 383 Islanders cognize marine and terrestrial environments holistically, with more attention focused 384 on the components and interactions of an integrated whole, than on discrete ecological attributes. 385 386 The most vivid Islander expressions of this ecosystem-like understanding are the wedge-shaped, ridge-to-reef resource management units that have been described across Oceania (Ruddle et al. 387 1992; Lauer 2016). These land-sea concepts emphasize the intrinsic entangling of physical and 388 389 biological components with the social and cultural world.

In addition to a holistic worldview of coral reef social-ecological systems, island societies like 390 Moorea also emphasize the unpredictability and unknowability of these systems. In fact, many 391 non-western societies, including those in Oceania, grasp the nature of ecosystems in ways similar 392 393 to non-equilibrium ecosystem science, a framework that emphasizes surprise and non-linearity, threshold effects, and systems flips instead of predictability, stable states, and homeostasis. The 394 magnitude of ecological and fishing changes we observed likely fall within the bounds of Pacific 395 Islanders' cultural expectations for normal fluctuations in their diets and catch. In other words, 396 397 the disturbances deemed dramatic from a Western scientific perspective, and perceived as significant events to fishers, are also inscribed for Pacific Islanders within a 'normal' cyclical 398 pattern of disturbances and recoveries. Indeed, the ecological observations of the COTS outbreak 399 400 and Cyclone Oli span relatively short timeframes (barely a decade) relative to individuals' own 401 lifespans. In addition, the fore reef of Moorea has proven very resilient to disturbances that 402 reduce coral cover, with several major disturbance events and subsequent recovery of the reef 403 since the 1970s (Adam et al. 2011; Trapon et al. 2011). In the case of the most recent

404 disturbances considered here, many areas of the fore reef regained their pre-disturbance levels of live coral within 5 years (Holbrook et al. 2018). The resilience of the reef ecosystem, when 405 considered at the scale of the individuals' lifespans, may contribute to the perceptions of our 406 informants (whose mean age = 47 y) of the limited impacts the disturbances had on their fishing 407 behavior and dietary choices. Future archeological research, similar to that carried out on Hawaii 408 409 and Rapa Nui (Kirch and Hunt 1997), exploring the long-term socio-ecological dynamics on Moorea, could shed light on the scale and intensity of social-ecological changes on Moorea in 410 the context of disturbance frequency. 411

412 *Conclusions*

Although this study focuses on fisher-fish interactions in Moorea, our results are of general 413 relevance for coral reef ecosystems. Coral reefs globally are experiencing increasing 414 disturbances, in many cases causing major changes in benthic and fish communities (Holbrook et 415 al. 2008). Understanding how fishers conceive and respond to these ecological changes is crucial 416 to predicting how social-ecological feedbacks might enhance or erode ecosystem resilience 417 (Leenhardt et al. 2016; Leenhardt et al. 2017). Such feedbacks are particularly likely in places 418 like Moorea where the most commonly targeted fishes are herbivores, which control macroalgae 419 420 and confer resilience on the coral dominated reef state (Mumby et al. 2007; Holbrook et al. 2016). Fishing on such species has often been linked to switches between coral and algal 421 community states (Hughes et al. 2007; Rasher and Hay 2010), and thus the details of fishing 422 behavior may be critical for understanding the resilience of these alterative states. 423 424 More broadly, our analysis has implications about researching knowledge production and

formulating management initiatives in socio-ecological systems. The disconnect between

426 Moorea's fishers' reporting of change, those apparent in the catch data, and the characterizations

427	of reef change offered by ecologists, highlights a critical issue - western scientists and other
428	stakeholders may produce knowledge grounded in different epistemological and ontological
429	assumptions about the world and what constitutes 'change' (Barnes et al. 2013). In complex
430	social-ecological systems like the one studied here, we should not expect singular,
431	incontrovertible knowledge about the system, and there will be significant differences between
432	and gaps within both local and ecological knowledge that may only widen with the uncertainty
433	of the Anthropocene era. Thus, it is likely to be increasingly useful to understand how all
434	stakeholders (e.g., scientists, conservation practitioners, fishers, tourist operators, etc.) produce in
435	situ site-specific knowledge and form social-ecological relations. Scientist-resource user
436	collaborations for research and resource monitoring can increase trust between stakeholders,
437	improve adaptive management strategies, and help keep pace with unforeseen social-ecological
438	transformations of the Anthropocene.
439	References
440 441 442	Adam, T. C., A. J. Brooks, S. J. Holbrook, R. J. Schmitt, L. Washburn, and G. Bernardi. 2014. How will coral reef fish communities respond to climate-driven disturbances? Insight from landscape-scale perturbations. <i>Oecologia</i> 176: 285-296.
443 444 445	Adam, T. C., R. J. Schmitt, S. J. Holbrook, A. J. Brooks, P. J. Edmunds, R. C. Carpenter, and G. Bernardi. 2011. Herbivory, connectivity, and ecosystem resilience: response of a coral reef to a large-scale perturbation. <i>PloS one</i> 6: e23717.
446 447 448	Barnes, J., M. Dove, M. Lahsen, A. Mathews, P. McElwee, R. McIntosh, F. Moore, J. O'Reilly, et al. 2013. Contribution of anthropology to the study of climate change. <i>Nature Climate Change</i> 3: 541-544.
449 450	Bellwood, D. R., T. P. Hughes, C. Folke, and M. Nyström. 2004. Confronting the coral reef crisis. <i>Nature</i> 429: 827.
451 452	Brooks, A. 2011. Moorea Coral Reef LTER: Reference: Fish Taxonomy, Trophic Groups and Morphometry. knb-lter-mcr.6001.3.
	Dreaks A 2017 Maaroo Carol Deef I TED, Carol Deef, Long term Deputation and Community

Brooks, A. 2017. Moorea Coral Reef LTER: Coral Reef: Long-term Population and Community
Dynamics: Fishes, ongoing since 2005. knb-lter-mcr.6.55
doi:10.6073/pasta/4541694f7928bc7f0d8b604ff9936a81.

- Cinner, J., M. J. Marnane, T. R. McClanahan, and G. R. Almany. 2006. Periodic closures as
 adaptive coral reef management in the Indo-Pacific. *Ecology and Society* 11: 31.
- 458 Cinner, J. E., W. N. Adger, E. H. Allison, M. L. Barnes, K. Brown, P. J. Cohen, S. Gelcich, C. C.
 459 Hicks, et al. 2018. Building adaptive capacity to climate change in tropical coastal
 460 communities. *Nature Climate Change* 8: 117-123.
- 461 Dalzell, P., T. J. H. Adams, and N. V. C. Polunin. 1996. Coastal fisheries in the Pacific Islands.
 462 *Oceanography Marine Biology Annual Revview* 34: 395-531.
- Frisch, A., R. Baker, J. A. Hobbs, and L. Nankervis. 2008. A quantitative comparison of
 recreational spearfishing and linefishing on the Great Barrier Reef: Implications for
 management of multi-sector coral reef fisheries. *Coral Reefs* 27: 85-95.
- Gillett, R., and W. Moy. 2006. Spearfishing in the Pacific Islands: Current status and
 management issues. Rome, Italy: FAO/FishCode Review. No. 19.
- Han, X., T. C. Adam, R. J. Schmitt, A. J. Brooks, and S. J. Holbrook. 2016. Response of
 herbivore functional groups to sequential perturbations in Moorea, French Polynesia.
 Coral Reefs 35: 999-1009.
- Holbrook, S. J., T. C. Adam, P. J. Edmunds, R. J. Schmitt, R. C. Carpenter, A. J. Brooks, H. S.
 Lenihan, and C. J. Briggs. 2018. Recruitment drives spatial variation in recovery rates of resilient coral reefs. *Scientific Reports* 8: 7338.
- Holbrook, S. J., S. R. J., T. C. Adam, and A. J. Brooks. 2016. Coral reef resilience, tipping points
 and the strength of herbivory. *Scientific Reports* 6: 35817.
- Holbrook, S. J., R. J. Schmitt, and A. J. Brooks. 2008. Resistance and resilience of a coral reef
 fish community to changes in coral cover. *Marine Ecology Progress Series* 371: 263-271.
- Hughes, T. P., M. L. Barnes, D. R. Bellwood, J. E. Cinner, G. S. Cumming, J. B. C. Jackson, J.
 Kleypas, I. A. van de Leemput, et al. 2017. Coral reefs in the Anthropocene. *Nature* 546:
 82-90.
- Hughes, T. P., D. R. Bellwood, C. Folke, R. S. Steneck, and J. Wilson. 2005. New paradigms for
 supporting the resilience of marine ecosystems. *Trends in Ecology and Evolution* 20:
 380-386.
- Hughes, T. P., J. T. Kerry, A. H. Baird, S. R. Connolly, A. Dietzel, C. M. Eakin, S. F. Heron, A.
 S. Hoey, et al. 2018. Global warming transforms coral reef assemblages. *Nature* 556:
 486 492-496.
- Hughes, T. P., M. J. Rodrigues, D. R. Bellwood, D. Ceccarelli, O. Hoegh-Guldberg, L. McCook,
 N. Moltschaniwskyj, M. S. Pratchett, et al. 2007. Phase shifts, herbivory, and the
 resilience of coral reefs to climate change. *Current Biology* 17: 360-365.
- Hviding, E. 1996. *Guardians of Marovo Lagoon: Practice, place, and politics in maritime Melanesia*. Honolulu, HI: University of Hawaii Press.
- Institut de la statistique de la Polynésie française. 2012. Premiers résultats du recensement de la population de la Polynésie française 2012. Papeete: ISPF.
- Jennings, S., and N. Polunin. 1995. Comparative size and composition of yield from six Fijian
 reef fisheries. *Journal of Fish Biology* 46: 28-46.

- Johannes, R. E. 1981. Words of the lagoon: Fishing and marine lore in the Palau District of
 Micronesia. Berkeley: University of California Press.
- Johannes, R. E. 2002. The renaissance of community-based marine resource management in
 Oceania. Annual Review of Ecology and Systematics 33: 317-340.
- Kirch, P. V., and T. L. Hunt. 1997. *Historical ecology in the Pacific Islands: Prehistoric environmental and landscape change*. New Haven: Yale University Press.
- Lamy, T., P. Legendre, Y. Chancerelle, G. Siu, and J. Claudet. 2015. Understanding the spatiotemporal response of coral reef fish communities to natural disturbances: insights from beta-diversity decomposition. *PloS one* 10: e0138696.
- Lauer, M. 2016. Governing uncertainty: Resilience, dwelling, and flexible resource management
 in Oceania. *Conservation & Society* 14: 34-47.
- Lauer, M. 2017. Changing understandings of local knowledge in island environments.
 Environmental Conservation: 1-12.
- Lauer, M., and S. Aswani. 2010. Indigenous knowledge and long-term ecological change:
 Detection, interpretation, and responses to changing ecological conditions in Pacific
 Island communities. *Environmental Management* 45: 985-997.
- Lauer, M., and J. Matera. 2016. Who detects ecological change after catastrophic events?
 Indigenous knowledge, social networks, and situated practices. *Human Ecology* 44: 33-46.
- Leenhardt, P., M. Lauer, R. Madi Moussa, S. J. Holbrook, A. Rassweiler, R. J. Schmitt, and J.
 Claudet. 2016. Complexities and uncertainties in transitioning small-scale coral reef
 fisheries. *Frontiers in Marine Science* 3: Article 70.
- Leenhardt, P., V. Stelzenmüller, N. Pascal, W. N. Probst, A. Aubanel, T. Bambridge, M. Charles,
 E. Clua, et al. 2017. Exploring social-ecological dynamics of a coral reef resource system
 using participatory modeling and empirical data. *Marine Policy* 78: 90-97.
- McClanahan, T. R., and J. E. Cinner. 2012. Adapting to a changing environment: Confronting
 the consequences of climate change. New York: Oxford University Press.
- McManus, J. W., J. Reyes, Rodlfo B Reyes, and J. Nanola, Cleto L. 1997. Effects of some
 destructive fishing methods on coral cover and potential rates of recovery. *Environmental Management* 21: 69-78.
- McMillen, H. L., T. Ticktin, A. Friedlander, S. D. Jupiter, R. Thaman, J. Campbell, J. Veitayaki,
 T. Giambelluca, et al. 2014. Small islands, valuable insights: Systems of customary
 resource use and resilience to climate change in the Pacific. *Ecology and Society* 19: 44.
- Mumby, P. J., A. Hastings, and H. J. Edwards. 2007. Thresholds and the resilience of Caribbean
 coral reefs. *Nature* 450: 98.
- Pauly, D., V. Christensen, J. Dalsgaard, R. Froese, and F. Torres. 1998. Fishing down marine
 food webs. *Science* 279: 860-863.
- Rasher, D. B., and M. E. Hay. 2010. Chemically rich seaweeds poison corals when not
 controlled by herbivores. *Proceedings of the National Academy of Sciences* 107: 9683 9688.

- Reddy, S., A. Wentz, O. Aburto-Oropeza, M. Maxey, S. Nagavarapu, and H. M. Leslie. 2013.
 Evidence of market- driven size- selective fishing and the mediating effects of biological and institutional factors. *Ecological Applications* 23: 726-741.
- Rogers, C., S., and J. Miller. 2006. Permanent 'phase shifts' or reversible declines in coral cover?
 Lack of recovery of two coral reefs in St. John, US Virgin Islands. *Marine Ecology Progress Series* 306: 103-114.
- Ruddle, K., E. Hviding, and R. E. Johannes. 1992. Marine resources managment in the context of
 customary tenure. *Marine Resources Economics* 7: 249-273.
- Schneider, C. A., W. S. Rasband, and K. W. Eliceiri. 2012. NIH Image to ImageJ: 25 years of
 image analysis. *Nature Methods* 9: 671.
- Trapon, M. L., M. S. Pratchett, and L. Penin. 2011. Comparative effects of different disturbances
 in coral reef habitats in Moorea, French Polynesia. *Journal of Marine Biology* 2011:
 Article ID 807625.
- Walter, C. 1968. The biology of *Ctenochaetus straitus*. A known ciguateric acanthurid fish of
 Tahiti. *Report* South Pacific Comission.

552 **FIGURES**

553 Figure 1. (A) Map of the island with the focal regions of Afareaitu, PaoPao and Teavaro

marked. (B) Photo of fish being sold by the roadside (note 0.5 m sizing bar).



Figure 2. Size distributions of all fish taxa observed on the reef, the subset of targeted taxa on the reef,
and the taxa found in the catch. Curves are kernel density estimations produced with the geom_density
function of the ggplot package in R (bandwidth smoothing parameter = 0.02).



Figure 3. Fish biomass on the reef through the time period spanning the 2009-2010 disturbances. The
"Small Fish" indicates biomass of fish smaller than 0.15 m, "Fishable Size" represent larger (> 0.15 m)
fish from non-targeted taxa. The remaining areas represents biomass of fish > 0.15 m which are
commonly found in the catch (Table 4), with 8 taxa broken out, and the remainder combined into "Other
Fishable." The timing of the peak disturbance is indicated with a dashed line.



570 Figure 4. The relative biomass of fishable taxa (including only individuals > 15 cm) on the reef (A) and





Figure 5. The relationship between the relative biomass of each taxonomic group on the reef and the
relative biomass of that group in the catch plotted by year. The time-averaged biomass in the catch and on
the reef for each taxon are also plotted (H). In this latter panel, the symbols for each species match those
in A-G and the 1:1 line is plotted.



Biomass on Reef

578 TABLES

Year	Fish counted on reef	Fish sampled in catch
2007	32131	1878
2008	34255	4309
2009	41538	-
2010	30013	-
2011	24231	-
2012	25963	2435
2013	29330	-
2014	30430	4319
2015	23995	4836

Table 1. Number of fish observed in the reef surveys and in the catch, by year.

Tahitian name	Scientific name	Reported commonly eaten	Reported commonly caught
Thon (French)	Thunnus spp.	17%	4%
Pa'ati*	Scarus/Chlorurus spp.	15%	12%
Pahoro	Scarus/Chlorurus spp.	11%	6%
I'ihi	Myripristis spp.	8%	13%
Tarao	Epinephelus spp.	8%	8%
Pa'aihere	Caranx spp.	6%	9%
Ume	Naso spp.	6%	6%
Maito	Acanthurus/ Ctenochaetus spp.	4%	3%
Ature	Selar crumenophtalmus	3%	4%
Roi	Cephalopholis spp.	2%	4%
To'au	Lutianus fulvus	2%	4%

Table 2. Fish most frequently reported eaten or caught in household surveys (N = 326 surveys).

582

* This term denotes terminal phase fish

584	Table 3.	Percentage of households	who responded	affirmatively to	the questions	related to the (COTS

585 outbreak and Cyclone Oli.

	Answered 'Yes'
Do you remember any taramea (COTS) outbreaks? ($N = 348$)	40%
Did taramea outbreaks change how, what, or where you fished? ($N = 339$)	13%
Did it change what fish you ate or bought to eat? ($N = 194$)	1.5%
Do you remember Cyclone Oli? (N = 348)	100%
Did Cyclone Oli change how, what, or where you fished? ($N = 310$)	19%
Did Cyclone Oli change what fish you ate or bought to eat? ($N = 350$)	10%

Table 4. Relative abundance of taxa observed in the catch and their corresponding % contribution to
abundance on the reef (considering only fish of targetable size, > 0.15 m). The top 23 genera observed in
the catch are listed, representing more than 99% of the catch. The genera *Chlorurus* and *Scarus* have been
combined in this table because they can be difficult to distinguish in the photos of the catch. Stars (*)
indicate taxa reported commonly eaten in more than 5% of household surveys.

Genus	% abundance in catch	% biomass in catch	% fishable size abundance on reef	% fishable size biomass on reef
Chlorurus-Scarus*	26.5	35	20.9	22.8
Naso*	18.5	16.1	5.5	4.2
Myripristis*	15.5	10	0.2	0.1
Siganus	8.9	4.5	0.4	0.2
Mulloidichthys	6.4	5.1	3.4	2.1
Parupeneus	5.4	5.4	1.5	0.9
Epinephelus*	3.2	3.1	0.4	0.3
Selar	2.9	1.2	0	0
Cypselurus	2.2	0.9	0	0
Acanthurus	1.8	1.6	10.1	8.5
Cephalopholis	1.4	1.8	3.8	3.3
Cheilopogon	1.4	0.5	0	0
Sargocentron	1.1	0.9	0.6	0.3
Lutjanus	0.8	0.9	0.3	0.2
Monotaxis	0.7	1.4	0.4	0.5
Caranx*	0.5	3.1	0.2	1.1
Lethrinus	0.4	1.3	0.1	1
Calotomus	0.3	0.5	0	0
Heteropriacanthus	0.3	0.2	0	0
Cheilinus	0.2	0.4	0.5	0.8
Gnathodentex	0.2	0.1	2.5	1.5
Kyphosus	0.2	0.4	0	0