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1      Combination of two methodologies, artificial neural network and linear  
2                   interpolation, to gap-fill daily nitrous oxide flux measurements

3

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8

9    **Keywords.**

10    Greenhouse Gases, Soil, Chambers, Crops, Agriculture

11

12    **Abstract.**

13    Continuous N<sub>2</sub>O flux acquisition is crucial to enrich our knowledge of the complex  
14    mechanisms underlying the annual greenhouse gas budget and to refine their estimation. N<sub>2</sub>O  
15    flux measurement methodologies at high temporal resolution, i.e. micro-meteorology  
16    methodologies, are still scarce and may exacerbate the lack of important data, especially during  
17    the night if the required turbulent conditions are not met. The static and automated chamber  
18    methodologies also lead to numerous gaps in a time series due to low sampling frequency,  
19    hardware malfunctions, chambers removal during field operations or filtering of low-quality  
20    measurements. There is a strong need to define a generic and realistic N<sub>2</sub>O flux gap-filling  
21    methodology, especially since there is no consensus on the methodology to be used.

22    In this study, we investigated the effect of using either the traditional linear interpolation  
23    methodology alone, either an Artificial Neural Networks (ANN) methodology alone or the  
24    combination of both on gap-filled daily N<sub>2</sub>O flux dynamics and annual budget. All three  
25    methodologies were tested on daily N<sub>2</sub>O flux time series measured with automated chambers

26 over 5 years from 2012 to 2016 on a southwestern France crop site following a wheat – maize  
27 rotation.

28 On average over the studied period, the results showed better statistical scores using the ANN  
29 methodology alone than using the linear interpolation methodology alone, with  $R^2$  and RMSE of  
30 0.84 and  $12.4 \text{ gN ha}^{-1} \text{ d}^{-1}$  and of 0.68 and  $17.4 \text{ gN ha}^{-1} \text{ d}^{-1}$ , respectively. However, whereas the use  
31 of ANN methodology reproduced well high measured  $\text{N}_2\text{O}$  fluxes, it induced overestimation on  
32 low measured  $\text{N}_2\text{O}$  fluxes where the use of the linear interpolation methodology was relevant. To  
33 overcome that issue and to take advantages of both methodologies we propose a new one which  
34 mixes both. On average, using the mixed methodology did not increase statistical scores  
35 compared to the ANN one, with a  $R^2$  and a RMSE of 0.84 and  $12.4 \text{ gN ha}^{-1} \text{ d}^{-1}$  respectively for  
36 both, but for periods with low measured  $\text{N}_2\text{O}$  fluxes using the mixed methodology improved the  
37 statistical scores and the observed daily flux dynamic.

38

39

## 40 **1. Introduction**

41

42 Nitrous oxide (N<sub>2</sub>O) is an important greenhouse gas as it presents a high global warming  
43 potential, around 300-fold higher than carbon dioxide (CO<sub>2</sub>), for an approximate residence time  
44 of 120 years in the atmosphere (Intergovernmental Panel on Climate Change (IPCC), 2013).  
45 Moreover, N<sub>2</sub>O participates in stratospheric ozone depletion (Ravishankara et al., 2009). At the  
46 global scale, N<sub>2</sub>O is responsible for 6% of the current global warming, its atmospheric  
47 concentration has increased at a rate of 0.93 ppb per year from 2007 to 2017 (WMO Greenhouse  
48 gas bulletin, 2018) and represents 46% of the total greenhouse gas (GHG) emissions from  
49 agriculture (United Nations Environment Program (UNEP), 2012). Agricultural soils constitute  
50 the main anthropogenic source of N<sub>2</sub>O, mostly because of an increase in land conversion for  
51 agriculture and the intensified use of nitrogen fertilizers (Ussiri and Lal, 2013; Lognoul et al.,  
52 2017; Davidson, 2009; Snyder et al., 2009). The formation of N<sub>2</sub>O in agricultural soils is mainly  
53 due to two microbial processes: nitrification and denitrification. Ammonium (NH<sub>4</sub><sup>+</sup>) and nitrate  
54 (NO<sub>3</sub><sup>-</sup>) concentrations together with soil water content, temperature, organic matter and pH, are  
55 the principal physical-chemical factors modulating the production of N<sub>2</sub>O from agricultural soils  
56 (Hénault et al., 2005; Li et al, 2000; Parsons et al, 1993; Wiljer et Delwiche, 1954). N<sub>2</sub>O  
57 emissions are thus determined by a combination of factors. The magnitude of each of these direct  
58 controls is subject to its own set of biological and abiotic controls (agricultural practices),  
59 showing the non-linearity of N<sub>2</sub>O emissions.

60 In the aim of mitigating climate change by reducing N<sub>2</sub>O emissions from soils, it is important  
61 to determine and understand N<sub>2</sub>O flux dynamics in response to variability in the climate and  
62 agricultural practices. To this end, different technologies have been developed and it is common

63 to find studies with low sampling frequency for measuring N<sub>2</sub>O fluxes. The static chamber  
64 method has been widely used (Reeves and Wang, 2015) as it presents the advantages of: (1)  
65 having low initial set-up costs, (2) being easily deployable from site to site and (3) being usable  
66 in small experimental plots, but it is also very demanding in manpower and offers a lower, more  
67 discontinuous sampling frequency than automated chambers (Zhang et al., 2014; Tellez-Rio et  
68 al., 2015; Reeves et al., 2016; Delon et al., 2017; Vinzent et al., 2017; Tallec et al., 2019). The  
69 use of a static chamber requires the operator to be well synchronized with N<sub>2</sub>O flux events, which  
70 are not known a priori. Automated chambers have been developed to increase sampling  
71 frequency (e.g. four times a day) and to reduce the uncertainty related to temporal N<sub>2</sub>O flux  
72 variability (Peyrard et al., 2016; Tallec et al., 2019). Nevertheless, due to field operations,  
73 instrument failure and/or quality control filtering, long periods of data loss may occur as for the  
74 static chamber method. Although the recent development of eddy-covariance methodology to  
75 measure N<sub>2</sub>O fluxes (Nemitz et al., 2018) presents the advantage of sampling at very high  
76 frequency (with a half-hourly time step) and at larger scale (from several m<sup>2</sup> to several hectares)  
77 than with chambers, it also presents the disadvantage of losing a lot of data because of low  
78 turbulence, especially during the night (Tallec et al., 2019). So, whatever the methodology used,  
79 data loss is inevitable, which could be critical for studies of how the ecosystems function and for  
80 the calculation of annual N<sub>2</sub>O budget. While standardized methodologies for CO<sub>2</sub> flux gap-filling  
81 (mean diurnal variation, look up table, regression application) are well recognized and validated  
82 thanks to clear, robust relationships between the fluxes and drivers, those for N<sub>2</sub>O flux remain  
83 challenging and non-consensual (Nemitz et al., 2018). The methodologies applied are often site-  
84 specific and depend on the availability of ancillary data. A widely used gap-filling method in  
85 N<sub>2</sub>O emission studies is the linear interpolation technique (Tellez-Rio et al., 2015; Vinzent et al.,  
86 2017). However, N<sub>2</sub>O fluxes have high temporal variability (Tallec et al., 2019) and sporadic

87 events can lead to dynamics totally different from the smoother diurnal variation observed for  
88 CO<sub>2</sub>. N<sub>2</sub>O flux dynamics are too complex to use the same gap-filling methodology as that applied  
89 for CO<sub>2</sub> fluxes. Low sampling frequency of N<sub>2</sub>O fluxes combined with a linear interpolation gap-  
90 filling method may thus result in the real N<sub>2</sub>O flux dynamics being missed, under- or  
91 overestimation of the annual N<sub>2</sub>O budget and misinterpretation of the effects of climate and  
92 management.

93 To overcome this issue, and in the context of the ICOS (Integrated Carbon Observation  
94 System; Franz et al., 2018) network, Nemitz et al. (2018) suggested using and testing the  
95 Artificial Neural Networks (ANN) methodology that had proved to perform well for CH<sub>4</sub> and  
96 CO<sub>2</sub> flux gap-filling (Papale and Valentini, 2003; Melesse and Hanley, 2005; Moffat et al., 2007,  
97 Dengel et al., 2013). Moffat et al. (2007), who reviewed 15 different techniques tested on a  
98 dataset compiled from six different European forest sites to fill gaps in values of the half-hourly  
99 Net Ecosystem Exchange (NEE), showed that ANN had a consistently good gap-filling  
100 performance and low annual sum bias, with R<sup>2</sup> between 0.41 to 0.90, and recommended its use  
101 even in different ecosystems. Melesse and Hanley (2005) tested ANN to gap-fill hourly CO<sub>2</sub>  
102 fluxes from three different ecosystems (forest, wheat, grassland) in the United States. They found  
103 that this technique could successfully predict observed values with R<sup>2</sup> values between 0.75 to  
104 0.94 and therefore considered it reliable, efficient and highly significant to estimate CO<sub>2</sub> fluxes.  
105 However, the very fleeting N<sub>2</sub>O fluxes are more difficult to simulate and gap-fill as, unlike CO<sub>2</sub>  
106 fluxes, they do not present any diurnal dynamics (Tallec et al., 2019). ANN is a statistical method  
107 that provides a relatively easy way to approach non-linear functions. It has the advantage of being  
108 usable without any knowledge of the processes that have an impact on the targeted variable(s).  
109 Some knowledge on the key drivers for thematic applications, as for N<sub>2</sub>O fluxes gap-filling, is  
110 still needed. Today, few studies have used and evaluated the performance of ANN methodology

111 with other methodologies simulating N<sub>2</sub>O fluxes (Ryan et al., 2004; Villa-Vialaneix et al., 2012;  
112 Taki et al., 2018). Ryan et al. (2004) undertook to simulate daily N<sub>2</sub>O emissions from an intensive  
113 grassland ecosystem in New Zealand using ANN methodology and obtained good statistical  
114 results, with R<sup>2</sup> ranging from 0.67 to 0.85. Taki et al. (2018) compared the use of ANN and linear  
115 interpolation methodologies to gap-fill a 6 years time series of N<sub>2</sub>O flux measured on a Canadian  
116 crop site following a corn/soybean/winter wheat rotation. On average, whatever the cropping  
117 year, they found a lower R<sup>2</sup> with the use of ANN than Ryan et al. (2004). Depending on the  
118 cropping year and the crop species, R<sup>2</sup> ranged from 0.19 (for a soybean crop) to 0.65 (for a maize  
119 crop). This strong variation in the ANN performance may question the relevance of using that  
120 methodology and underlines the need to more investigate it in different context and  
121 configuration. However, Taki et al. (2018) recommended the use of ANN instead of the classical  
122 interpolation method, as on average the former scored a higher R<sup>2</sup> (0.41) than the latter (0.34)  
123 when considering the entire studied period.

124 The study proposed here is part of this dynamics, as comparisons between gap-filling  
125 procedures on N<sub>2</sub>O fluxes are still lacking. Its objectives are twofold: (1) to evaluate and compare  
126 linear interpolation and ANN as gap-filling methodologies to reproduce daily low and high N<sub>2</sub>O  
127 fluxes and (2) to propose a third gap-filling approach combining both methodologies using a  
128 criterion on the method's choice, either linear interpolation or ANN, based on the duration of the  
129 missing data period and the N<sub>2</sub>O flux magnitude. All three methodologies were tested on a long  
130 time series of measurements taken on an agricultural plot (maize/wheat rotation) in the south-  
131 west of France from 2012 to 2016 and compared together according to each crop plot functioning  
132 period (bare soil, growing season, low or high N<sub>2</sub>O fluxes intensity, etc.). The effect of the  
133 methodology's choice on the estimated annual N<sub>2</sub>O budget is discussed.

134

## 135 2. Materials

### 136 2.1 Experimental site

137 The study is based on a dataset collected at the Lamasquère site (FR-Lam) in the South West  
138 of France (43°29'47''N, 1°14'16''E, 180 m elevation). This crop site is part of an experimental  
139 dairy farm (Domaine de Lamothe, Instituts Nationaux Polytechniques). The site is located in a  
140 plain having an oceanic climate with Mediterranean and continental influences, a mean annual  
141 rainfall of 680 mm and a mean annual temperature of 12.6 °C. The soil is mainly clayey (50.3%  
142 clay, 35.8% silt, 11.2% sand, 2.8% organic matter). The plot is located near the river Touch, a  
143 tributary of the river Garonne, in the Adour-Garonne water catchment area.

144 This experimental plot is part of the Regional Spatial Observatory, the regional ZA PYGAR  
145 (Zone Atelier Pyrénées-Garonne), the national research infrastructure OZCAR (Critical Zone  
146 Observatories: Research and Applications; Gaillardet et al., 2018) and the European network  
147 ICOS (Integrated Carbon Observation System). The site is therefore fully equipped with  
148 instruments to monitor greenhouse gas fluxes, meteorological, radiation and soil variables (Béziat  
149 et al, 2009; Tallec et al, 2013). Vegetation dynamics are also monitored in detail with green area,  
150 biomass and height measurements.

151 Management of the crop plot is intensive, with exportation of all aboveground biomass for  
152 mulching the cowsheds and/or feeding the herd. An annual irrigation of approximately 150 mm is  
153 applied when maize is cultivated. The total amount of nitrogen (N) applied varies from 100 to  
154 200 kgN ha<sup>-1</sup> and 110 to 145 kgN ha<sup>-1</sup> for mineral and organic fertilization respectively. Input  
155 modalities vary according to the crop species. Mineral fertilization is split into 3 applications for  
156 wheat and applied once for maize. The FR-Lam site occupation followed the rotation maize  
157 (2012), wheat (2013), cover crop (CC) (2013), maize (2014), maize (2015), wheat (2016), CC  
158 (2016).



159

## 160 2.2 $N_2O$ flux measurement

161 To measure  $N_2O$  emissions, 6 stainless steel automated chambers (covering an area of 0.161  
162  $m^2$ ) were installed on the plot according to a closed dynamic set-up. The chambers were  
163 elongated (70cm×23cm×10cm) so as to be easily inserted in the crop inter-rows, where they were  
164 placed at a 10 cm soil depth. A previous comparison study between automated chambers and  
165 eddy-covariance methodologies (Tallec et al. 2019) on the same site showed that a set of 6  
166 automated chambers was sufficient to integrate spatial heterogeneity of the site and to well  
167 capture the mean daily  $N_2O$  flux dynamics. The set-up measured  $N_2O$  accumulation in each  
168 chamber alternately during 17.5 minutes every 6 hours, i.e. four cycles a day (00:00, 06:00,  
169 12:00, 18:00). A pump maintained continuous air circulation in and out of the chamber at a  
170 constant flow rate (1  $L \cdot min^{-1}$ ). Sampled air was supplied from the chamber to a gas analyser that  
171 measure  $N_2O$  molar fraction every 10 seconds (Thermofisher 46i, Megatec, France). The  
172 calculated  $N_2O$  fluxes (exponential fitting) were filtered by means of a mixture of goodness-of-fit  
173 statistics and visual inspection (Tallec et al., 2019). A fan also enabled air homogeneity in the  
174 chamber.

175 Daily  $N_2O$  fluxes were calculated as the mean value of all available fluxes during the day.  
176 The choice to work at a daily timescale was made as no significant diurnal  $N_2O$  flux intensity  
177 variation was found on FR-Lam site in a previous study (Tallec et al., 2019). Because of  
178 hardware dysfunction and/or technical operations, such as tillage and harvest, chambers were  
179 removed many times, leading to missing data which means that the number of available  $N_2O$  flux  
180 measurement per day to calculate a daily  $N_2O$  flux varied from 0 to 24. In this study, in order to  
181 maximize the size of the dataset for the need of that study we kept all available daily fluxes, even  
182 if calculated with only one measurement a day. Finally, from 2012 to 2016, out of 1827 daily

183 N<sub>2</sub>O flux values, 1442 were acquired. The whole dataset included 63% of “highly” representative  
184 daily values (calculated with 12 to 24 measurements a day), 22% of “moderately” representative  
185 daily values (calculated with 6 to 12 measurements a day) and 13% of potentially “poorly”  
186 representative daily values (calculated with 1 to 6 measurements a day).

187

### 188 *2.3 Meteorology, soil variables and ancillary data measurements*

189 The rich dataset acquired on FR-Lam offers the possibility to test different combinations of  
190 possible explanatory variables for the input layer in the ANN and to choose the most explanatory  
191 (see section 3.1.2.3). Matter and energy fluxes, meteorological data, radiation and soil variables  
192 were available at a half-hourly time step and are described in Table 1. The measurement  
193 methodology for each variable is described in Beziat et al. (2009) and Tallec et al. (2013). Each  
194 variable was aggregated at a daily time scale. Vegetation data were collected from 5 to 6 times a  
195 year during the growing season and a simple linear interpolation was applied to obtain the daily  
196 dynamics.

197

<b>Variable names</b>	<b>Variable description (unit)</b>
ABG	Above Ground Biomass (kg m <sup>-2</sup> )
E	Measured Evapotranspiration (kg m <sup>-2</sup> d <sup>-1</sup> )
ETP	Potential Evapotranspiration (mm)
GPP	Green Primary Production (gC m <sup>-2</sup> d <sup>-1</sup> )
H	Sensible heat flux (watt m <sup>-2</sup> d <sup>-1</sup> )
h_veg	Vegetation height (m)
LAI	Leaf Area Index (m <sup>2</sup> m <sup>-2</sup> )
LE	Latent heat flux (W m <sup>-2</sup> d <sup>-1</sup> )
LW	Incoming long wave (W m <sup>-2</sup> d <sup>-1</sup> )
NEE	Net Ecosystem Exchange (gC m <sup>-2</sup> d <sup>-1</sup> )
P	Rain (mm)
Pa	Pressure (kPa)
PPFD	Photosynthetic Photon Flux Density (μmol m <sup>-2</sup> d <sup>-1</sup> )
PWC	Plant Water Content (%)
Reco	Ecosystem respiration from soil (gC m <sup>-2</sup> d <sup>-1</sup> )
RH	Relative humidity (%)
Rg	Global radiation (MJ m <sup>-2</sup> d <sup>-1</sup> )
Rn	Net radiation (W m <sup>-2</sup> d <sup>-1</sup> )
SWC	Soil Water Content at 0, 5, 10, 30, 50 and 100 cm (%)
Ta_max	Maximum air temperature of the day (°C)
Ta_mean	Mean air temperature of the day (°C)
Ta_min	Minimum air temperature of the day (°C)
Ts	Soil temperature at 0, 5, 10, 30, 50 and 100 cm (°C)
Waterfall	Daily rain and irrigation event (mm)
WS	Wind speed (m s <sup>-1</sup> )

199

**Table 1.** Description of variables potentially available for gap-filling

200

201

202 Unfortunately, soil mineral nitrogen content was not available for all years and with low  
203 temporal resolution. Thus, to include the fertilization effect on N<sub>2</sub>O emissions in the ANN  
204 construction, a decreasing exponential function  $N_t$  (Eq. 1) was developed to account for the effect  
205 of fertilizer on N<sub>2</sub>O fluxes. The choice of coefficient 0.10 was determined based on the  
206 visualisation of the function's temporal dynamic. The aim of this function is to memorize a  
207 fertilisation event over multiple days after a fertilisation as we noticed, based on observations,  
208 that N<sub>2</sub>O fluxes appears several days after a fertilisation event if the environmental conditions are  
209 met. No optimisation process was undertaken to find the most accurate coefficient value.

$$210 \quad N_t = N_0 \times e^{-0.10 \cdot t} \quad (1)$$

211  
212 Here  $t$  represents the number of days after fertilization,  $N_t$  is the nitrogen function value at  
213 day  $t$  after fertilization (kgN ha<sup>-1</sup>) and  $N_0$  is the quantity of nitrogen fertilizer applied by the  
214 farmer on day 0 (kgN ha<sup>-1</sup>).

215

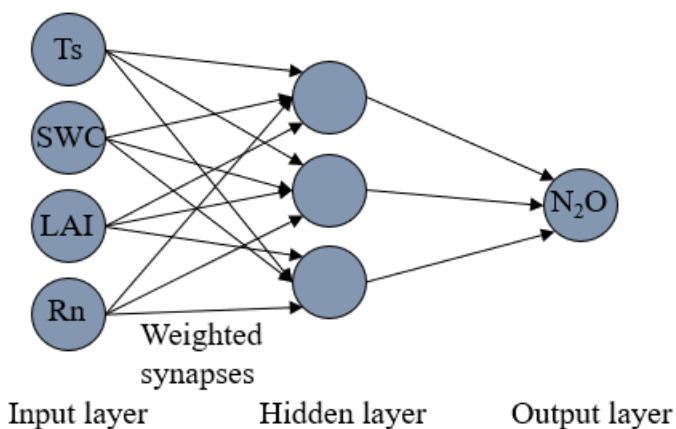
### 216 **3. Methods**

#### 217 *3.1 ANN gap-filling methodology*

##### 218 *3.1.1 General principle*

219 In this study, the ANNs created are based on the algorithm proposed by Bishop (1995) (i.e.,  
220 multi-layer perceptrons implemented in the nnet R package) and are composed of three layers of  
221 neurons (input, hidden and output) connected by synapses (Fig. 1). The input layer represents the  
222 predictive variables selected by the user to explain the targeted variables of the output layer. The  
223 number of input parameters and hidden neurons is determinant in the creation of a neural network  
224 and to avoid over fitting. Not enough hidden neurons will make it difficult to reach an accurate

225 estimation of the targeted variable, while too many will lead to noisy estimates (Delon et al.,  
226 2007). The relationship between input and output variables is modelled by sigmoid functions  
227 connecting neurons, the weights of interactions being determined through an iterative procedure.  
228



**Fig. 1.** Schematic MLP with 4 input variables, 3 hidden neurons and 1 output variable

229  
230 During the model training, nnet uses supervised learning algorithms. The synapse weights are  
231 modified to fit the network in order to reduce an error function (i.e., Sum of Squared Errors  
232 (SSE)) corresponding to the difference between predicted and observed output (Guenther and  
233 Fritsch, 2010; Ryan et al, 2004) following the Broyden-Fletcher-Goldfarb-Shanno algorithm. The  
234 values of the weights are initialized with random values and progressively adjusted.

235  
236 *3.1.2 Creation of the neural network step by step*

237 *3.1.2.1 Neural Network performances*

238 The data preparation procedure is established in accordance with the following studies Lek et  
239 al. 1966 and Olden et al. 2002, which recommend: (1) to standardize the measurement scales of  
240 the network inputs (by subtracting the mean value and dividing by the standard deviation)  
241 especially when the measurements are of different orders of magnitude, and (2) to convert the

242 range of the output variable to the interval [0-1] to conform to the interval of variation of the  
243 sigmoid transfer function (by subtracting the minimal value and dividing by the range of  
244 variation).

245 After normalization, the data set was randomly partitioned into two independent sets of equal  
246 data amount, 50% of the data to train and 50% of the data to test the created ANN, that enables to  
247 test the ANNs generalisation efficiency on an important amount of data. This 50/50 partitioning  
248 operation was repeated 40 times with the objective of obtaining 40 different random draws and  
249 selecting the draw with the highest performance.

250 The performance levels of the ANNs were evaluated using the determination coefficient ( $R^2$ ,  
251 Eq. 2.), the Root Mean Squared Error (RMSE, Eq. 3.) and the Relative Root Mean Squared Error  
252 (RRMSE, Eq. 4.).

$$253 \quad R^2 = \frac{(\sum_{i=1}^n (p_i - \bar{p}_i)(o_i - \bar{o}_i))^2}{\sum_{i=1}^n (p_i - \bar{p}_i)^2 \sum_{i=1}^n (o_i - \bar{o}_i)^2} \quad (2)$$

$$254 \quad RMSE = \sqrt{\frac{\sum_{i=1}^n (o_i - p_i)^2}{n}} \quad (3)$$

$$255 \quad RRMSE = \frac{RMSE}{\bar{y}} \times 100 \quad (4)$$

256 With  $n$  the number of observed values,  $o_i$  the observed value,  $p_i$  the predicted value and  $\bar{o}$  and  $\bar{p}$   
257 the mean of observed and predicted data, respectively.

258 Different ANN configurations were tested. An overall neural network was implemented first  
259 on the whole dataset and then by discriminating each crop growing season and bare soil periods.

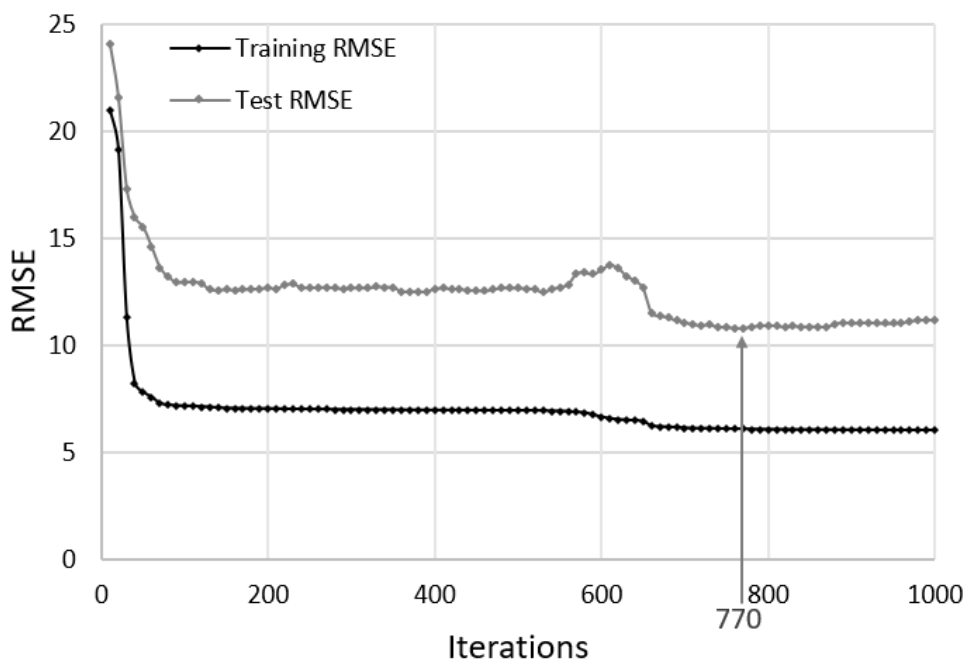
260

### 261 3.1.2.2. *Early stopping procedure*

262 In order to avoid over or under fitting during ANN training, an early stopping procedure was  
263 setup to determine the best number of iterations required to build the ANN model. For every 40

264 draws, ANNs were created and trained throughout 1 000 iterations and saved every 10 iterations  
265 (in total 4000 created ANNs for one functioning period). Then the 4000 trained ANNs were  
266 tested on the testing datasets. The  $R^2$  and RMSE between observations and simulations on the  
267 training and testing dataset were also saved every 10 iterations for the 40 draws. The number of  
268 required iterations to build the best ANN model and select the best draw was determined in  
269 function of the minimum value of RMSE obtained on the testing datasets. Fig. 2 illustrates the  
270 case of draw number 5 for maize 2012 period.

271



**Fig. 2.** RMSE ( $\text{gN ha}^{-1} \text{d}^{-1}$ ) between ANN training (black) and testing (grey) datasets in function of iterations for draw number 5 during maize 2012 period. Here the minimum RMSE for the test simulation is reached after 770 iterations.

272

273 For that draw, the ANN created after 770 iterations was selected as it corresponds to the  
274 minimum RMSE of the testing dataset. Below or above that value the ANNs may be under or  
275 overfitted.

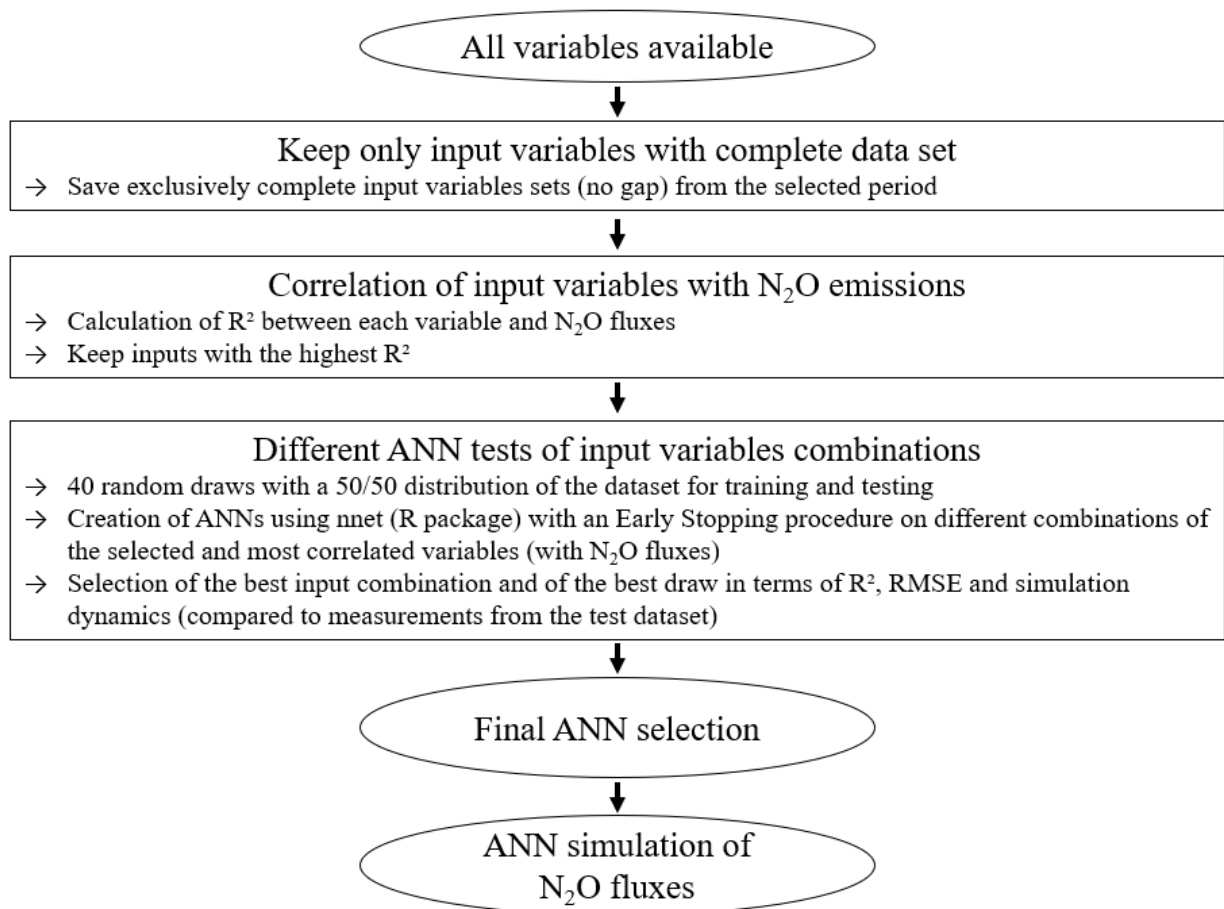
276 The best ANN determined from the iteration procedure was then kept and saved for every 40  
277 draws of each functioning period. Then a classification of the 40 random draws was done in  
278 function of  $R^2$  in order to select the draw showing the best  $R^2$ . A final test to select the best ANN  
279 for a functioning period was then made by comparing the ANN simulation on the test dataset  
280 with the corresponding observations dynamic. If this visual dynamic comparison was not  
281 satisfying, the second ANN and draw set with the best  $R^2$  result was chosen, and so on.

282

### 283 *3.1.2.3 Selection of input variables*

284 The methodology used to select the most appropriate combination of variables for each period  
285 is presented in Fig. 3. Only input variables (Table 1) without any missing values were considered.  
286 A linear regression calculation between each variable and  $N_2O$  fluxes was then performed to  
287 select a combination of variables showing the highest correlation with  $N_2O$  fluxes (highest  $R^2$   
288 values). Afterwards, numerous ANNs were created by testing different combinations of the  
289 selected variables for each functioning period. The final input combination was selected based on  
290 the previously described process using the criterion  $R^2$ , RMSE and visual checking. Finally, a  
291 specific ANN was built for each functioning period using a combination of 4 or 5 input variables,  
292 and 3 hidden neurons to fit the output variable, i.e. the daily soil  $N_2O$  fluxes. Increasing the  
293 number of input parameters and/or hidden neurons did not improve the results significantly (data  
294 not shown).





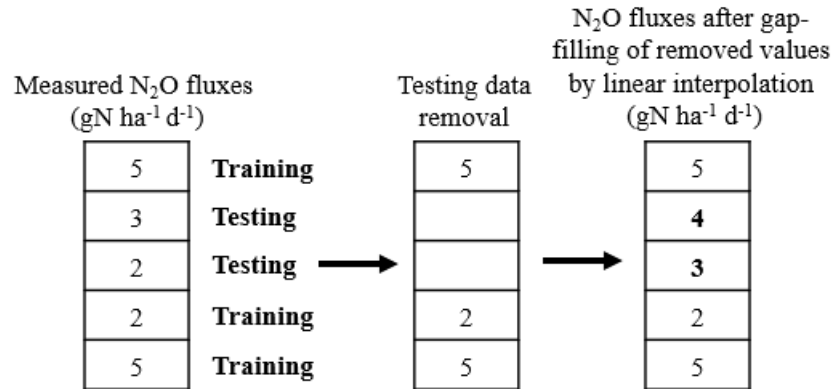
**Fig. 3.** Diagram outlining the successive steps to implement the ANN methodology.

296

297

### 298 3.2 Comparison between ANN and linear interpolation gap-filling methodologies

299 In order to evaluate the linear interpolation methodology and compare it with the ANN  
 300 methodology on exactly the same days of observations, artificial gaps were created into the  
 301 dataset used for the testing procedure (Fig. 4) for each functioning period. Each artificial gap was  
 302 then filled using the linear interpolation method between the surrounding training values. The  
 303 linear interpolation methodology was evaluated statistically using the same indices as described  
 304 in section 3.1.2.1 ( $R^2$ , RMSE, RRMSE).



**Fig. 4.** Method used to create and linearly interpolate artificial gaps from the N<sub>2</sub>O observations.

306

### 307 3.3 Determination of the GMD (Gap Magnitude and Duration) coefficient

308 The comparison between ANN and linear interpolation methodologies showed that each  
 309 technique may have benefits and limitations depending on the gap condition (see section 4.5).

310 Therefore, the best gap-filling approach would be to combine the two techniques with the help of  
 311 a new criterion, the coefficient GMD (Eq. 5) which would allow the most suitable methodology  
 312 to be chosen. The value of GMD depends on the length of the gap and on the difference between  
 313 the N<sub>2</sub>O fluxes intensity bracketing the gap. GMD thus allows the best gap-filling methodology  
 314 to be chosen between linear interpolation and ANN.

315

$$316 \quad GMD = D * abs(FN_2O_{before\ gap} - FN_2O_{after\ gap}) \quad (5)$$

317 Where  $D$  is the number of consecutive missing days in the gap and  $FN_2O_{before\ gap}$  and  
 318  $FN_2O_{after\ gap}$  are the N<sub>2</sub>O flux values observed at the beginning and the end of the gap,  
 319 respectively.

320 A decision tree to guide the choice to use either the linear interpolation or the ANN gap-  
321 filling methodology to fill a specific gap, depending on the GMD value, is developed in Fig. 5.  
322

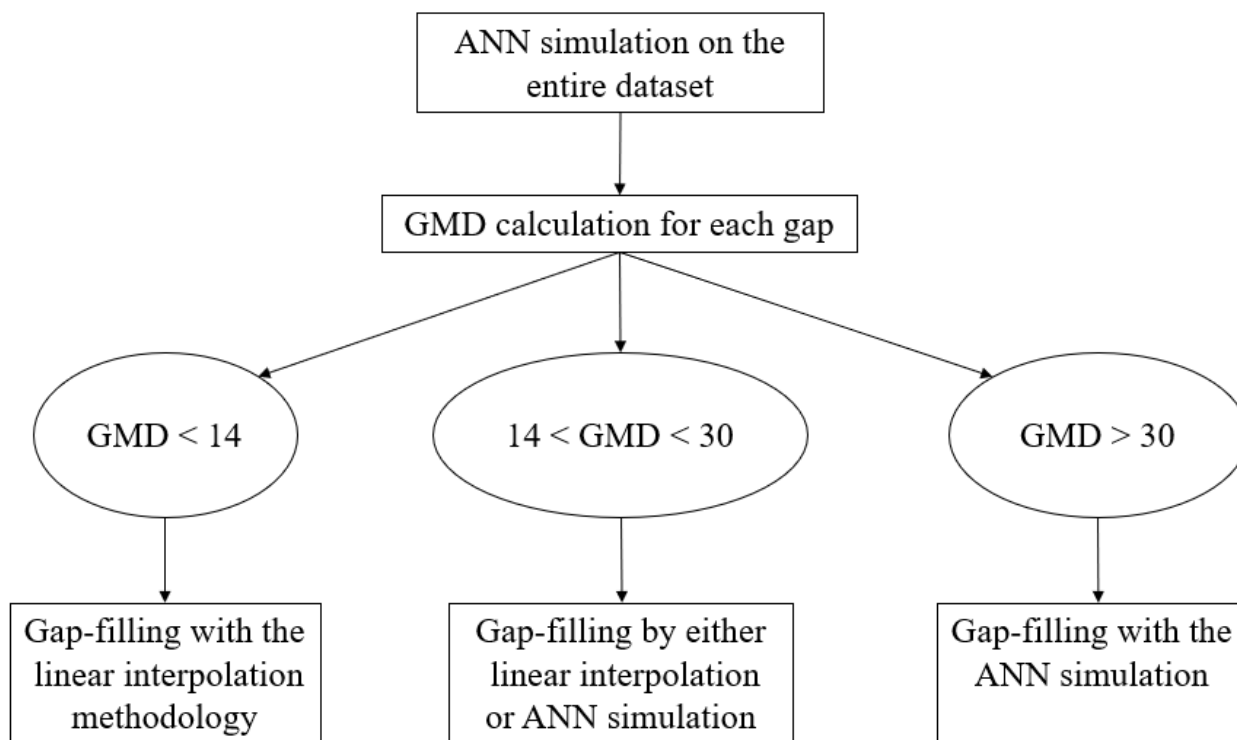
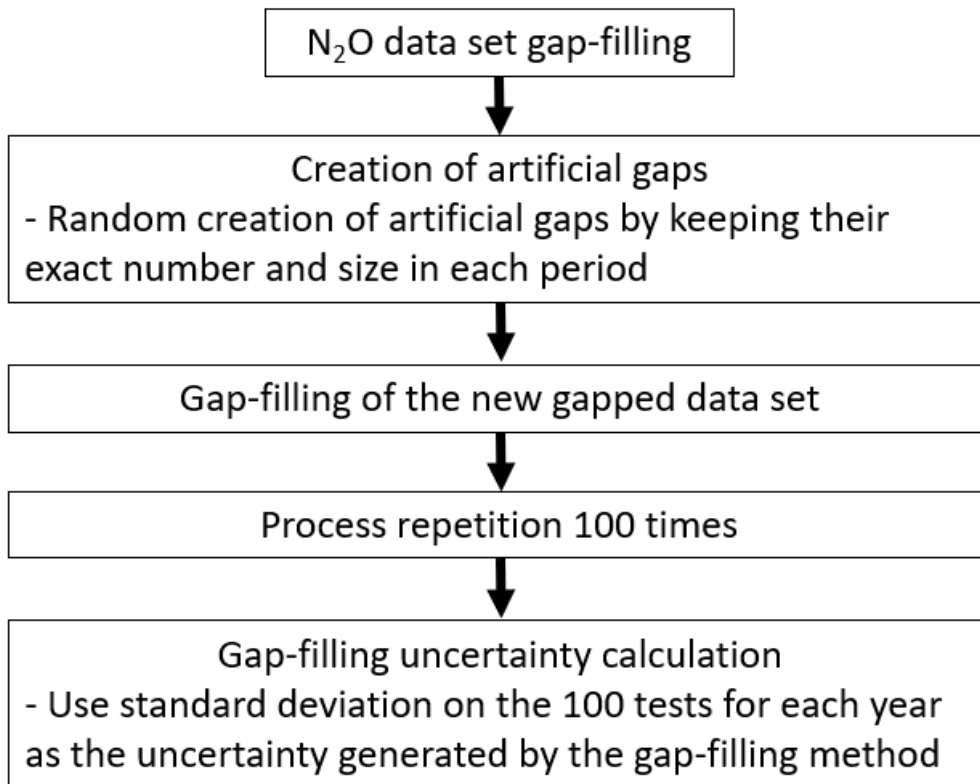


Fig. 5. Decision tree helping to select the more suitable gap-filling method according to GMD value.

323  
324 *3.4 Estimating annual N<sub>2</sub>O budget uncertainty due to gap-filling*  
325 The uncertainty of the annual N<sub>2</sub>O budget related to the gap-filling was investigated using a  
326 methodology adapted from Richardson and Hollinger (2007). This methodology is usually used  
327 to estimate annual NEE gap-filling uncertainty (Richardson and Hollinger, 2007; Schmidt et al.,  
328 2012). It is calculated by adding random gaps as actually observed in each real site-year of data  
329 gaps and then by adding small and long gaps (the maximum gap length is 8 days). However, in  
330 this methodology, the data set specificity is not taken into account, i.e. the number, length and  
331 locations of gaps are not respected. To include this information, a new methodology was

332 proposed as follows (Fig. 6): (1) the N<sub>2</sub>O datasets were first gap-filled according to each  
333 functioning period by using the combined methodology exposed in section 3.3; (2) new artificial  
334 random gaps were then created on the gap-filled data sets to test the effect of different locations  
335 of missing data while ensuring the reproduction of the exact number and size of each gap for each  
336 annual period (for example, the year 2013 had exactly 3 gaps of 1, 6 and 1 consecutive days of  
337 missing N<sub>2</sub>O fluxes). No artificial gaps were created on previous real gap locations to avoid  
338 replacement of a former gap-filled location by the same value; (3) afterwards, gap-filling was  
339 again applied to the N<sub>2</sub>O flux dataset containing the new gaps; (4) this operation was repeated  
340 100 times on the whole period and allowed 100 annual N<sub>2</sub>O budgets to be calculated for each  
341 year. The standard deviations calculated over the 100 repetitions indicated the uncertainty related  
342 to the gap-filling on each annual N<sub>2</sub>O budget.

343



**Fig. 6.** Diagram outlining the successive steps to calculate the uncertainty due to gap-filing methodology.

344

## 345 **4. Results and Discussion**

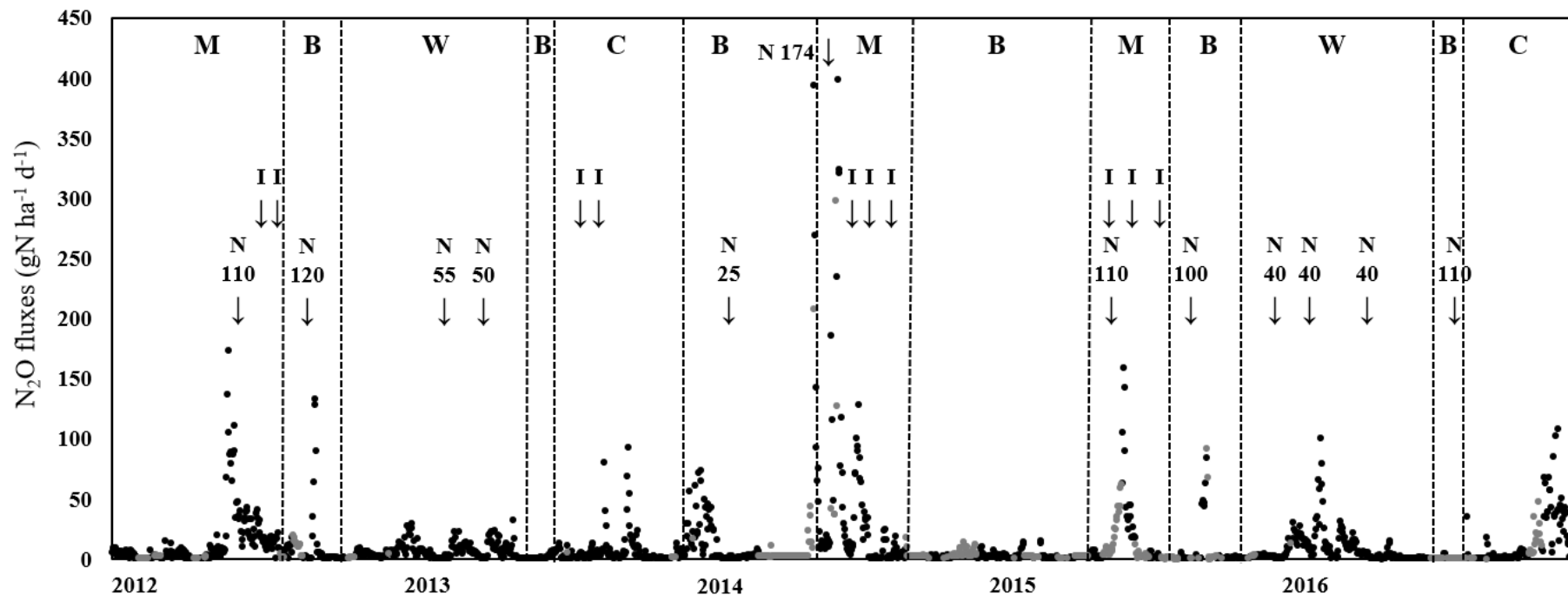
### 346 *4.1 N<sub>2</sub>O flux range and dynamic at FR-Lam crop site*

347 Over the five years of monitoring, the magnitude and timing of measured N<sub>2</sub>O fluxes varied  
 348 according to the soil occupation, the agricultural practices and the climatic year (Fig. 8). Typical  
 349 variability of N<sub>2</sub>O fluxes occurred according to rain, irrigation and fertilisation events. Wheat,  
 350 cover crops and bare soil periods emitted less N<sub>2</sub>O, on average (from 6.6 to 13.0 gN ha<sup>-1</sup> d<sup>-1</sup>),  
 351 than the maize periods (16.3 to 42.9 gN ha<sup>-1</sup> d<sup>-1</sup>) (Table 2). Intensity of N<sub>2</sub>O fluxes was also  
 352 higher in the maize periods than in the others, with maximum values ranging between 160 and  
 353 399, compared to 33 and 133 gN ha<sup>-1</sup> d<sup>-1</sup> (Table 2, Fig. 8). Maximum fluxes measured on maize  
 354 2012 (174 gN ha<sup>-1</sup> d<sup>-1</sup>), maize 2015 (160 gN ha<sup>-1</sup> d<sup>-1</sup>) and wheat 2013 (33 gN ha<sup>-1</sup> d<sup>-1</sup>) were in the

355 range of those given for wheat – irrigated maize crop rotations in Dhadli et al. (2016) for northern  
356 India on a loamy sand soil site (143 and 50 gN ha<sup>-1</sup> d<sup>-1</sup> for maize and wheat respectively) and in  
357 Han et al. (2016) for northern China (96 and 50 gN ha<sup>-1</sup> d<sup>-1</sup> for maize and wheat respectively).  
358 Maximum fluxes measured on wheat 2016 (100 gN ha<sup>-1</sup> d<sup>-1</sup>) and maize 2014 (399 gN ha<sup>-1</sup> d<sup>-1</sup>)  
359 were significantly above those generally reported in the literature. Taki et al. (2018) nevertheless  
360 measured a maximum N<sub>2</sub>O flux of 225 gN ha<sup>-1</sup> d<sup>-1</sup> on a maize crop, which is still half that of  
361 maize 2014 in this study. The strong N<sub>2</sub>O flux peak observed at the end of June 2014 could be  
362 explained by optimal conditions for N<sub>2</sub>O production: the field operations during that period  
363 resulted in easily available N substrate, with two considerable mineral fertilization of 102 and 72  
364 kgN ha<sup>-1</sup> applied at an interval of only one week, and high soil water content, with two irrigation  
365 events one week after N application. We can assume that N<sub>2</sub>O fluxes on our site were within the  
366 average range of other wheat-maize crop rotation sites but atypical management practices or  
367 climate may have led to temporally unusual and important N<sub>2</sub>O fluxes.

368 The bare soil and wheat 2013 periods were the ones with the most and the least missing data,  
369 21% and 1% respectively. The number of consecutive missing daily N<sub>2</sub>O fluxes varied from 1 to  
370 62, with a minimum of 1 to 2 on wheat 2013 and a maximum of 1 to 62 on bare soil periods  
371 (Table 2). These gaps were due to the removal of chambers for field operations or disturbance of  
372 the aero-dynamic conditions, leading to a filtering of the data.

373



**Fig. 7.** Daily N<sub>2</sub>O fluxes dynamic observed (cross) and gap-filled (dot) at FR-Lam from 2012 to 2016. M: maize; B: bare soil; W: wheat; C: cover crop; N: nitrogen application (kgN ha<sup>-1</sup>); I: irrigation.

374

375

376

377

<b>Vegetation periods</b>	<b>Mean daily N<sub>2</sub>O flux (gN ha<sup>-1</sup> d<sup>-1</sup>)</b>	<b>Maximum daily N<sub>2</sub>O flux (gN ha<sup>-1</sup> d<sup>-1</sup>)</b>	<b>Period length (days)</b>	<b>Number of daily N<sub>2</sub>O fluxes missing</b>	<b>Length of consecutive missing daily N<sub>2</sub>O fluxes</b>
<b>Maize 2012</b>	25.6 (± 30.0)	174	119	15 (13%)	1 - 10
<b>Maize 2014</b>	42.9 (± 67.0)	399	124	14 (11%)	1 - 7
<b>Maize 2015</b>	16.3 (± 36.2)	160	127	63 (50%)	17 - 25
<b>Wheat 2013</b>	6.6 (± 6.4)	33	257	3 (1%)	1 - 2
<b>Wheat 2016</b>	7.9 (± 16.7)	100	254	7 (3%)	1 - 5
<b>CC 2013</b>	9.3 (± 18.1)	93	90	6 (7%)	6
<b>CC 2016</b>	13.0 (± 22.8)	109	118	25 (21%)	1 - 17
<b>Bare soil periods</b>	7.5 (± 20.7)	133	738	252 (34%)	1 - 62
<b>Lamasquère 2012-2016</b>	12.3 (± 28.8)	399	1 827	385 (21%)	1 - 62

378

**Table 2.** Description of the daily N<sub>2</sub>O flux dataset at FR-Lam per functioning period.

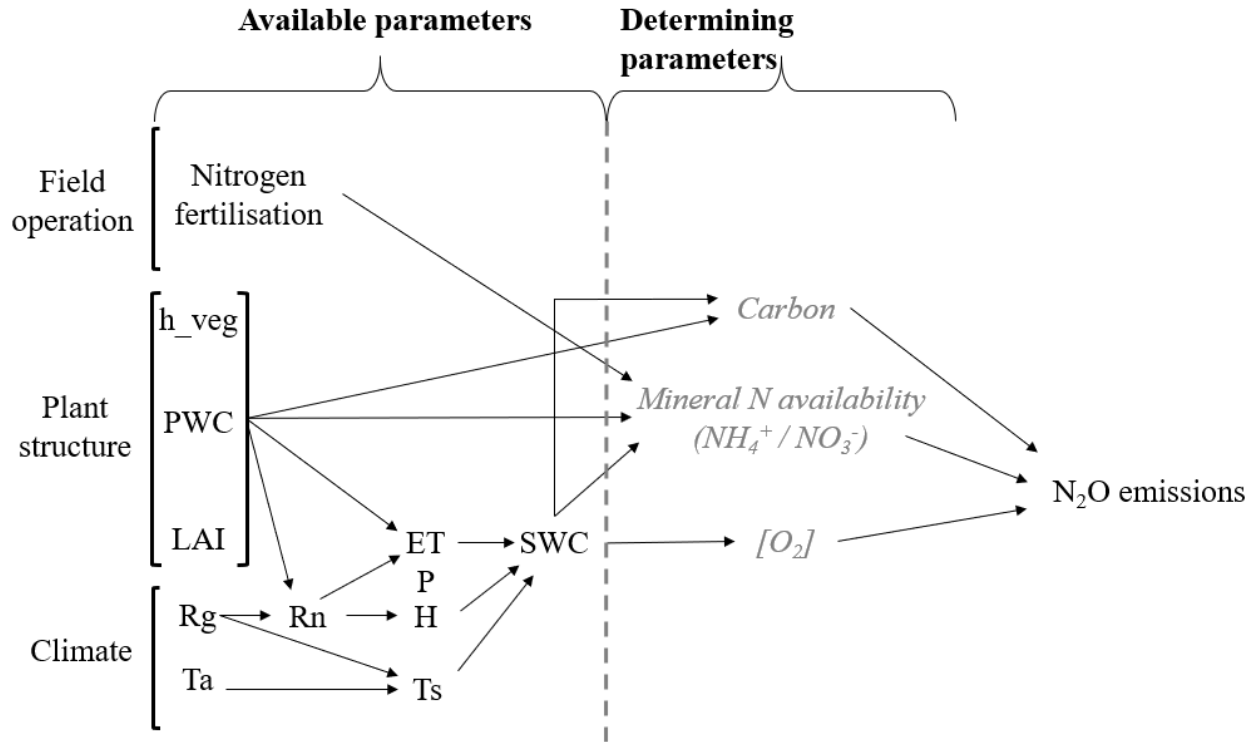
379

#### 380 4.2 ANN simulation and selected variables

381 The selected input variables varied among the study periods (Table 3). Whatever the  
382 combination, all input variables selected for gap-filling (Table 3) had a direct or indirect  
383 influence on N<sub>2</sub>O emissions (Robertson et al., 1989). Fig. 8 illustrates connections between input  
384 variables and the direct proximal factors of denitrification regulation (inspired from Robertson et  
385 al., 1989).

386





**Fig. 8.** Illustration of the possible interactions between selected variables and proximal factors of N<sub>2</sub>O emissions. Variables in grey italics were not measured or not available.

388

389 Variables related to vegetation structure make the greatest contribution to modulating  
 390 proximal factors and thus to N<sub>2</sub>O emission modulation. LAI and h\_veg are naturally related to the  
 391 soil mineral nitrogen availability. These also modulate and reflect the soil dioxygen content via  
 392 root respiration, the soil water content via potential evapotranspiration (ETP) and the soil carbon  
 393 content via rhizodeposition and physical disruption by the roots, etc. Net radiation (Rn)  
 394 corresponds to the available radiative energy in the top of soil – top of canopy continuum, which  
 395 naturally depends on Rg but also on the crop coverage and structure, and it potentially modulates  
 396 the heat fluxes that are implicated in soil water loss and soil heating. Sensible heat flux (H) may  
 397 be linked to soil temperature as an increasing H is associated with surface heating coming from  
 398 water stress. Finally, the fertilization parameter uses the function established in Eq. 1. to

399 represent nitrogen application in the field. Taki et al. (2018) partly used the same parameters in  
400 all their single-year ANNs, together with air temperature, soil temperature, soil water content,  
401  $\text{NO}_3^-$  concentration,  $\text{NH}_4^+$  concentration with snow depth, rainfall, the season and the time of the  
402 year. In contrast with Ryan et al. (2004) and Taki et al. (2018) who both used rainfall as  
403 controlling factor, waterfall (rain and irrigation), even tested in applying a lagtime from 0 to 5  
404 days, did not appear to be a significant input parameter in the ANN constructions at FR-Lam.  
405 SWC at 50 cm depth appeared more explanatory than rainfall. Moreover, since the water table  
406 rose frequently near the soil surface at FR-Lam, SWC at 50 cm parameter reflects environmental  
407 conditions' variation related to sub-ground phenomenon on that site in addition to being  
408 integrative of the rainfall/irrigation amount. The season and the time of the year were not  
409 necessary parameters, as we separated our dataset according to the functioning period, and snow  
410 depth was not useful as no significant snowfall occurred during the monitoring.  $\text{NO}_3^-$  and  $\text{NH}_4^+$   
411 concentrations would have been valuable as determining parameters in the formation of  $\text{N}_2\text{O}$   
412 (Fig. 8) but they were not continuously monitored.

413  
414 An ANN using LAI,  $T_a$ \_mean, Rn, ETP and SWC\_50 as input variables was first tested for  
415 the whole period from 2012 to 2016 and led to poor statistical scores with an  $R^2$  of 0.28, an  
416 RMSE of  $23 \text{ gN ha}^{-1} \text{ d}^{-1}$  and an RRMSE of 189%. We thus separated the data set according to the  
417 functioning periods (Table 3) to create a neural network specific to each crop growing season  
418 period and a single one combining all bare soil periods with the most explanatory variables. The  
419 combination of variables giving the best estimation of  $\text{N}_2\text{O}$  emissions varied from one period to  
420 another (Table 3).

421

422

<b>Growing season</b>	<b>Variables used to develop ANNs</b>
<b>Maize 2012</b>	LAI, Ta_mean, Rn, ETP, SWC_50
<b>Maize 2014</b>	LAI, Ta_mean, Rn, ETP, SWC_50
<b>Maize 2015</b>	LAI, Rn, ETP, SWC_50, H
<b>Wheat 2013</b>	Fertilization, Ts_50, PWC, h_veg, H
<b>Wheat 2016</b>	Fertilization, Ta_mean, Rn, H, h_veg
<b>CC 2013</b>	Ta_mean, Rn, PWC, h_veg
<b>CC 2016</b>	Ts_50, Rn, SWC_50, h_veg, ETP
<b>Bare soil periods</b>	Ta_mean, Rg, ETP, Fertilization

**Table 3** Summary of the selected variables used to develop the different neuronal network at FR-Lam.

424

425

#### 426 *4.3. ANN performances*

427 Creating ANNs relating to each functioning period proved to be relevant, with  $R^2$ , RMSE and  
428 RRMSE ranging from 0.54 to 0.94, from 4.0 to 33.9 gN ha<sup>-1</sup> d<sup>-1</sup> and from 62 to 115, respectively  
429 (Table 4), far better than the single ANN over the whole period. The  $R^2$  scores are highly  
430 improved compared to those of Taki et al. (2018) who found on average  $R^2$  values ranging from  
431 0.19 to 0.65, depending on the cropping year. Two main reasons can explain that important  
432 improvement. First, in our study, an ANN model was built specifically for each functioning  
433 period and by discriminating bare soil from growing season periods, whereas Taki et al. (2018)  
434 built ANN model without discriminating functioning periods. Even if discriminating functioning  
435 period makes more difficult the genericity of the ANN algorithm, it is in favour of improving the  
436 fineness of the gap-filling. Then, Taki et al. (2018) used scores and results averaged over 44

437 scenarios for each cropping year whereas in our study, only the best ANN model was kept from  
438 the 40 random draws for each studied period and shown here.

439 Considerable disparity occurred on statistical results according to the functioning period  
440 (Table 4). The RRMSE appeared to be high when looking at all functioning periods together and  
441 highlighted the fact that the RMSE was of the same magnitude as the average N<sub>2</sub>O fluxes in  
442 average. With R<sup>2</sup> ranging from 0.85 to 0.94, periods with maize or cover crop showed better  
443 performance than periods with wheat or bare soil, showing an R<sup>2</sup> ranging from 0.54 to 0.66. Even  
444 though wheat 2016 and bare soils had an RMSE of the same order of magnitude as the other  
445 periods, the difference between the observations and the test simulations turned out to be  
446 significant during these two periods when related to the average N<sub>2</sub>O fluxes with the highest  
447 RRMSE values, 115 and 114 % respectively. Wheat and bare soil periods also corresponded to  
448 the lowest mean daily N<sub>2</sub>O fluxes: 6.6 – 7.5 gN ha<sup>-1</sup> d<sup>-1</sup> against 9.3 – 42.0 gN ha<sup>-1</sup> d<sup>-1</sup> during  
449 maize/cover crop periods (Table 2). These results show that the ANN method would simulate  
450 fluxes with high intensity variation, as observed in maize crops, better than fluxes with low  
451 intensity variation as observed in winter wheat. However, these scores are close to those reported  
452 in the literature for CO<sub>2</sub> flux gap-filling on a forest site (Melesse and Hanley, 2005; Moffat et al.,  
453 2007). Accordingly, ANN performances for N<sub>2</sub>O flux estimation proved to be relevant even  
454 without any diurnal cycle when compared to other CO<sub>2</sub> flux modelling studies.

Growing season	Methodology	R <sup>2</sup>	RMSE (gN ha <sup>-1</sup> d <sup>-1</sup> )	RRMSE (%)
Maize 2012	ANN	0.86	10.8	43
	Linear interpolation	0.85	12.2	49
	Linear interpolation and ANN combine	0.87	10.6	43
Maize 2014	ANN	0.85	33.9	73
	Linear interpolation	0.58	55.6	119
	Linear interpolation and ANN combine	0.85	34.6	74
Maize 2015	ANN	0.94	13.8	82
	Linear interpolation	0.93	14.3	79
	Linear interpolation and ANN combine	0.97	13.4	77
Wheat 2013	ANN	0.54	4.0	65
	Linear interpolation	0.66	3.6	59
	Linear interpolation and ANN combine	0.67	3.5	57
Wheat 2016	ANN	0.60	8.8	115
	Linear interpolation	0.79	6.6	86
	Linear interpolation and ANN combine	0.69	7.8	102
CC 2013	ANN	0.93	4.0	62
	Linear interpolation	0.53	9.2	145
	Linear interpolation and ANN combine	0.93	4.0	62
CC 2016	ANN	0.86	10.0	76
	Linear interpolation	0.64	14.7	111
	Linear interpolation and ANN combine	0.85	10.1	77
Bare soil periods	ANN	0.66	8.6	114
	Linear interpolation	0.76	7.3	96
	Linear interpolation and ANN combine	0.68	8.6	113
All functioning periods together	ANN	0.84	12.4	101
	Linear interpolation	0.68	17.4	141
	Linear interpolation and ANN combine	0.84	12.4	100

455

456

457 *4.4 Evaluation of ANN and linear interpolation methodologies gap-filling performances*

458 Considering all functioning periods together, ANN methodology gave better statistical scores

459 (see Table 4) than the linear interpolation with higher R<sup>2</sup> (0.84 vs 0.68) and lower RMSE (12.4 vs

460 17.4 gN ha<sup>-1</sup> d<sup>-1</sup>) and RRMSE (101 vs 141 %). In the same trend, Taki et al. (2018) obtained

461 better statistical scores for the ANN methodology than with the linear interpolation methodology,

462 finding an average R<sup>2</sup> of 0.41 and 0.34 respectively.

**Table 4.** R<sup>2</sup> and RMSE calculated on testing dataset for ANN, linear interpolation and both methodologies combined for each functioning period.

463 Even if the linear interpolation might give comparable scores, i.e. for maize in 2012 and  
464 2015, it proved to perform better when the averaged N<sub>2</sub>O fluxes were the lowest, i. e. wheat  
465 2013, wheat 2016 and bare soil periods.

466

#### 467 *4.5 Effect of gap-filling methodologies on mean N<sub>2</sub>O emissions according to functioning period*

468 To analyse the effects of the two different gap-filling methodologies on the mean N<sub>2</sub>O fluxes  
469 per period, we applied them separately to gap-fill the real gaps from the studied data set from  
470 2012 to 2016. In the case of the ANN methodology, there are no significant differences on the  
471 means and standard deviations of the N<sub>2</sub>O fluxes before and after gap-filling whatever the period  
472 considered, which shows no large N<sub>2</sub>O flux variability was created after gap-filling (Table 5). In  
473 contrast, large variability appears after gap-filling when the linear interpolation method is used  
474 during maize 2014 and bare soil periods : the mean N<sub>2</sub>O flux for the maize 2014 period before  
475 gap-filling is  $42.9 \pm 77.7 \text{ gN ha}^{-1} \text{ d}^{-1}$  and becomes  $64.5 \pm 107.2 \text{ gN ha}^{-1} \text{ d}^{-1}$  after gap-filling; for  
476 the bare soil periods, the mean and standard deviations are  $7.5 \pm 15.0 \text{ gN ha}^{-1} \text{ d}^{-1}$  before gap-  
477 filling and  $21.3 \pm 58.0 \text{ gN ha}^{-1} \text{ d}^{-1}$  after gap-filling. In both cases, the linear interpolation  
478 methodology tends to increase mean N<sub>2</sub>O fluxes. Based on these results, it appears that either the  
479 ANN underestimates, or the linear interpolation overestimates N<sub>2</sub>O fluxes, or both are wrong.  
480 However, it questions the use of one methodology or the other. To explain these strong  
481 differences between the two methodologies, we analysed these specific periods in greater detail.

482

Vegetation periods	Before gap-filling	Gap-filling with ANN	Gap-filling with linear interpolation	Gap-filling with combine methodologies
<b>Maize 2012</b>	25.6 (± 29.9)	24.0 (± 28.2)	23.4 (± 28.5)	23.7 (± 28.4)
<b>Maize 2014</b>	42.9 (± 77.7)	45.5 (± 78.7)	64.5 (± 107.2)	45.5 (± 78.7)
<b>Maize 2015</b>	16.3 (± 32.2)	12.8 (± 25.1)	15.8 (± 25.9)	12.8 (± 25.1)
<b>Wheat 2013</b>	6.6 (± 6.3)	6.6 (± 6.3)	6.5 (± 6.3)	6.5 (± 6.3)
<b>Wheat 2016</b>	7.9 (± 12.5)	7.8 (± 12.4)	7.8 (± 12.4)	7.8 (± 12.4)
<b>CC 2013</b>	9.3 (± 17.2)	8.8 (± 16.7)	9.0 (± 16.6)	8.8 (± 16.7)
<b>CC 2016</b>	13.0 (± 23.4)	13.1 (± 21.5)	10.9 (± 20.8)	12.6 (± 21.5)
<b>Bare soil periods</b>	7.5 (± 15.0)	6.3 (± 13.0)	21.3 (± 58.0)	6.1 (± 13.0)

**Table 5.** Effect of ANN and linear interpolation gap-filling methodologies on mean N<sub>2</sub>O emissions (gN ha<sup>-1</sup> d<sup>-1</sup>).

483

484

485 In some specific case, applying a linear interpolation gap-filling on the bare soil periods

486 would introduce an unrealistic continuity such like the period from 06. March 2014 to 19. May

487 2014 as shown in Fig. 9 (since maize 2014 begins on the 20. May and would have been gap-filled

488 with ANN methodology). It could be explained by the presence of a very long gap of 69 days (

489 from 19. March to 26. May 2014) surrounded by two periods with high N<sub>2</sub>O flux amplitudes: 6.7

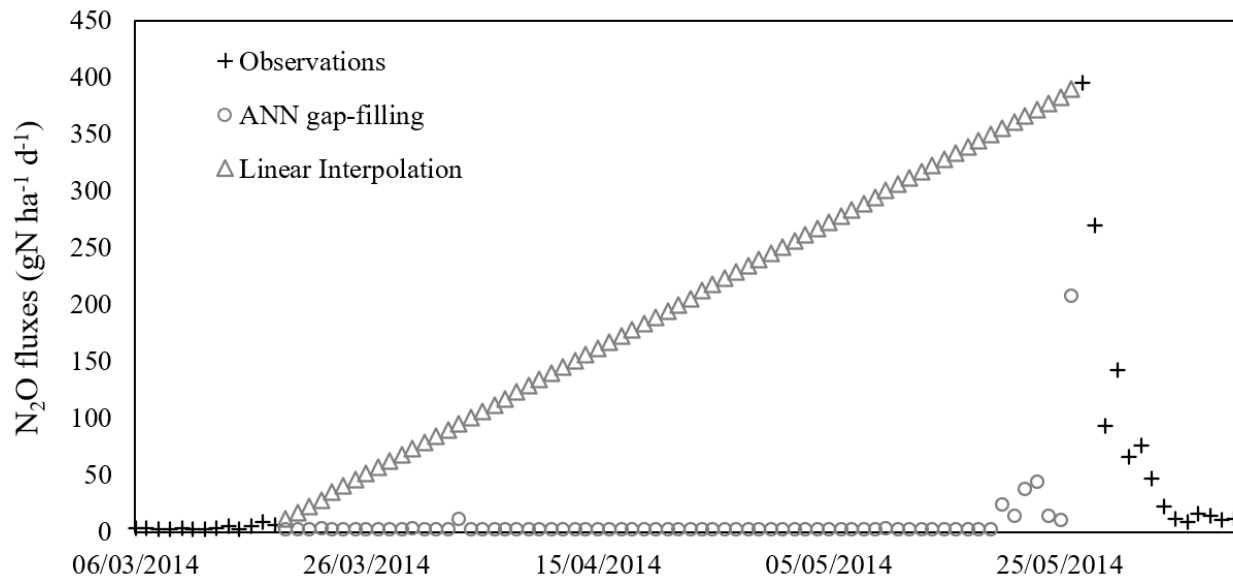
490 gN ha<sup>-1</sup> d<sup>-1</sup> on 18. March 2014 and 394.6 gN ha<sup>-1</sup> d<sup>-1</sup> on 27. May 2014. The linear interpolation

491 dynamics is clearly not suitable during this period. Even if the ANN gap-filling can be wrong, it

492 is more consistent with the dynamics of the observations, in particular since ANN takes the

493 environmental conditions into account, especially since environmental conditions (data not

494 shown) were not in favour of N<sub>2</sub>O production: only 75 mm of rain in 67 days, where the ANN  
495 simulated very low N<sub>2</sub>O fluxes.



**Fig. 9.** N<sub>2</sub>O fluxes (+) gap-filled by ANN (o) and linear interpolation (Δ) on bare soil period from 05/03/2014 to 19/05/2014 and on maize 2014 period from 20/05/2014 to 09/06/2014.

496  
497  
498 *4.6 Combination of ANN and linear interpolation methodologies*  
499 As the linear interpolation method also gives satisfactory statistical results (Table 4), it would  
500 have been a loss to set it aside since we have seen that, in some circumstances, this method is  
501 better than ANN, especially on periods with low averaged N<sub>2</sub>O fluxes. Moreover, in some  
502 specific case, like long period of missing data or period with environmental conditions in favour  
503 of N<sub>2</sub>O emissions, applying a linear interpolation gap-filling could introduce unrealistic N<sub>2</sub>O  
504 fluxes. In order to fix this, we propose to combine the two gap-filling methodologies depending  
505 on the value of GMD ((Eq. 5.), see section 3.3). The RMSE are then compared with GMD values,  
506 and used to highlight the potential threshold GMD values on which the selection of the best  
507 methodology to gap-fill the data could be based (Table 6).



	<b>Methodology</b>	<b>GMD &lt; 14</b>	<b>14 &lt; GMD &lt; 30</b>	<b>GMD &gt; 30</b>
<b>RMSE (gN ha d<sup>-1</sup>)</b>	<b>Linear interpolation</b>	3.9	11.6	43.7
	<b>ANN</b>	8.8	10.9	27.2

**Table 6.** RMSE according to gap-filling methodology and GMD coefficient.

509

510 Two GMD threshold values were highlighted (Table 6): when  $GMD < 14$ , the RMSE of the  
 511 ANN method was higher than the one from linear interpolation; when  $14 < GMD < 30$ , RMSE  
 512 were quite similar for both; when  $GMD > 30$ , the RMSE of the ANN method was lower than  
 513 linear interpolation one.

514 These results suggest that the suitability of a methodology for gap-filling depends on the  
 515 GMD value. The linear interpolation is preferable when  $GMD < 14$  and the ANN method is  
 516 preferable when  $GMD > 30$ . When  $14 < GMD < 30$ , the use of either interpolation or ANN is  
 517 possible as both methodologies give approximately the same score. We decided to use the ANN  
 518 methodology in this case as it allowed capturing the effect of environmental variables on  $N_2O$   
 519 fluxes. However, in the case of long gaps (when a gap contains more than 15 missing data)  
 520 bracketed by two low  $N_2O$  fluxes with  $GMD < 14$ , the decision would be to use the linear  
 521 interpolation methodology. It would be wise to look at possible important environmental  
 522 modifications during the gaps, such as heavy rain, irrigation or fertilization events, where it could  
 523 be more accurate to use the ANN methodology instead.

524 The combined methodology gave practically the same statistical results over the whole period  
 525 as those from the ANN methodology with a  $R^2$  and a RMSE of 0.84 and  $12.4 \text{ gN ha}^{-1} \text{ d}^{-1}$  for both  
 526 (Table 4). However, an improvement is seen on the three periods where the linear interpolation

527 method has better statistical results with the combined methodology compared to the ANN one  
528 with a  $R^2$ , a RMSE and a RRMSE of 0.68, 8.6 and 113 and of 0.66, 8.6 and 114 respectively for  
529 bare soil periods, of 0.69, 7.8 and 102 and of 0.60, 8.8 and 115 respectively for wheat 2016 and  
530 of 0.67, 3.5 and 57 and of 0.54, 4.0 and 65 respectively for wheat 2013. These results support the  
531 use of a combined methodology instead of applying either the linear interpolation alone or the  
532 ANN methodology alone for gap-filling.

533  
534 *4.7 Comparison of the three gap-filling methodologies: linear interpolation, ANN and the*  
535 *combination of both*

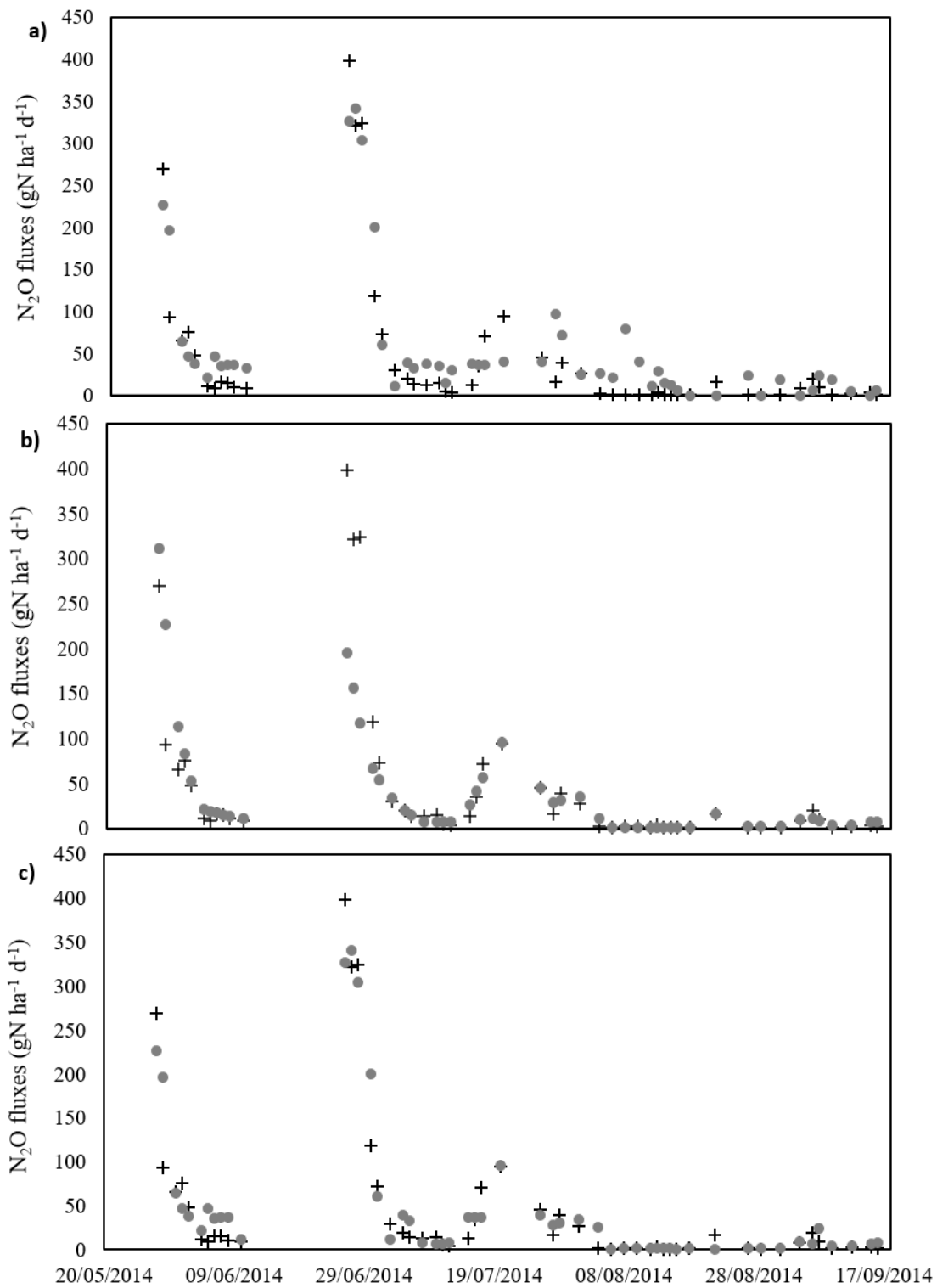
536 The simulated  $N_2O$  flux dynamic were compared between the three methodologies applied  
537 for maize 2014 period (Fig. 10). It showed that the ANN methodology gave on average better  
538 statistical scores than the linear interpolation (Table 4) and better reproduced the important peak  
539 than did the linear interpolation around the 29 June 2014. However, ANN over estimated low  
540  $N_2O$  fluxes measured in early July (Fig. 10 a. and b) where the linear interpolation reproduced  
541 much better their low intensity (Fig. 10 b). Although the combined methodology gave  
542 approximately the same statistical scores than those of ANN alone (Table 4), it allowed a better  
543 catching of  $N_2O$  flux dynamic (Fig. 10 c).

544 In the same way, during periods where the linear interpolation methodology gave better  
545 results than the ANN one (Table 4), combining both methodologies helped to better reproduce  
546 low fluxes (Fig. 11) and to improve statistical results compared to the ANN with a  $R^2$ , a RMSE  
547 and a RRMSE of 0.69, 7.8  $gN\ ha^{-1}\ d^{-1}$  and 102 for wheat 2016, respectively.

548 The ANN and the combined methodologies gave on average comparable statistical scores  
549 whatever the functioning period probably due to a control by high  $N_2O$  flux values. However,

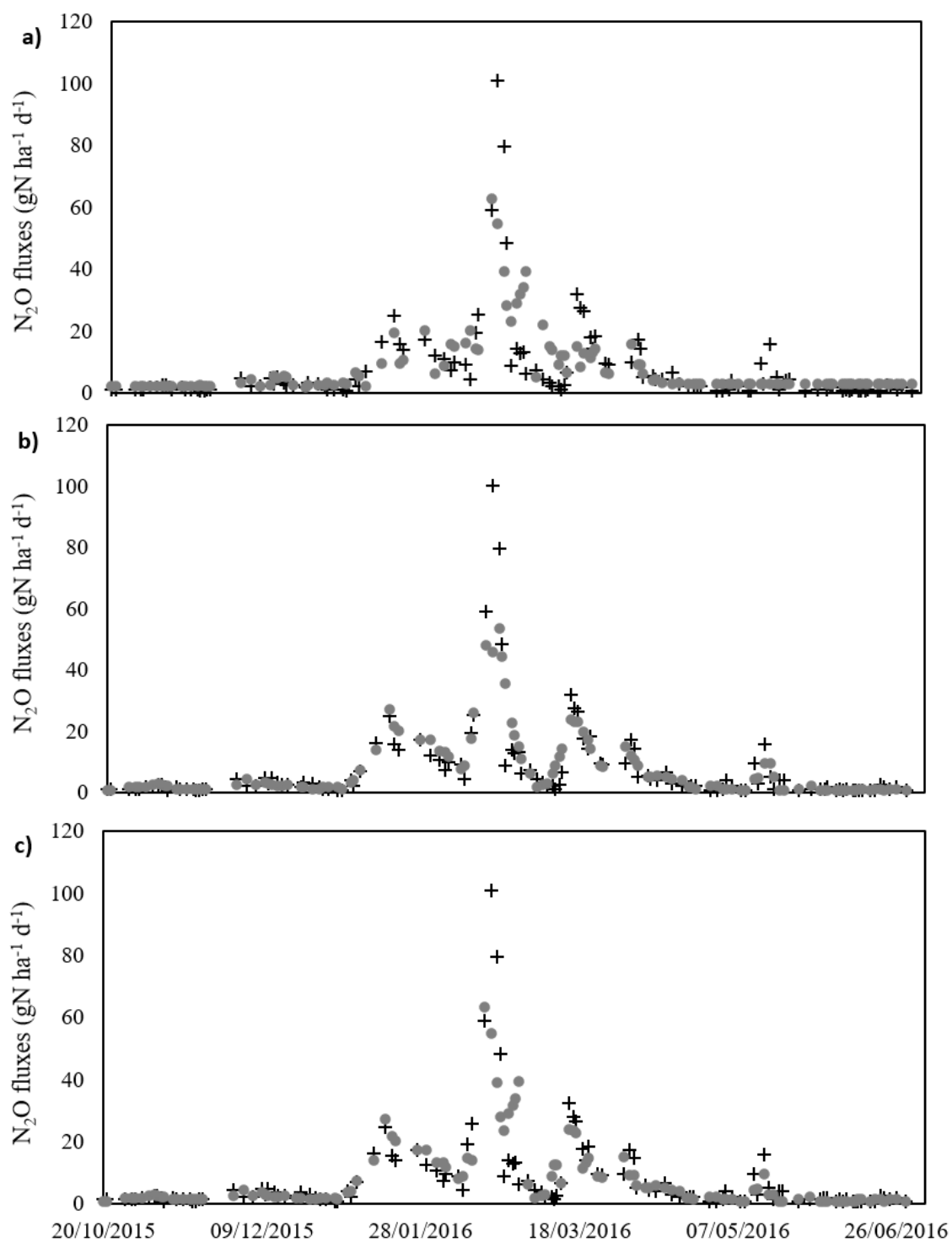
550 because the combined methodology proved to be more efficient to capture N<sub>2</sub>O intensity dynamic  
551 it was used to gap-fill missing data from 2012 to 2016 (Fig. 7).

552



**Fig. 10.** Comparison between observed (cross) and simulated (dot)  $N_2O$  fluxes dynamic according to gap-filling methodology ANN (a), linear interpolation (b) and mixed methodology (c) on Maize 2014. Only the days used for the test dataset are shown here.





**Fig. 11.** Comparison between observed (cross) and simulated (dot)  $N_2O$  fluxes dynamic according to gap-filling methodology ANN (a), linear interpolation (b) and mixed methodology (c) on Wheat 2016. Only the days used for the test dataset are shown here.



557 *5. Evaluation of gap-filling uncertainty on N<sub>2</sub>O budget*

558 The uncertainties on the N<sub>2</sub>O annual budget related to the gap-filling varied between ± 25 and  
559 ± 198 gN ha<sup>-1</sup> among years, and between 0.9 and 4.1 % when looking at the relative uncertainty  
560 (Table 7) calculated via the methodology described in section 3.4. Gap-filling had a particularly  
561 low impact on year 2013 (Table 7) as there were only few gaps during this period: 3 gaps with  
562 respectively 1, 6 and 1 missing value(s). These results are relatively low compared to the annual  
563 uncertainties calculated by Taki et al. (2018), who obtained an uncertainty variation among years  
564 of ± 70 to ± 810 gN ha<sup>-1</sup> and 7 to 24 % when looking at the relative uncertainty using ANN as  
565 gap-filling method, and an uncertainty variation among years of ± 40 to ± 470 gN ha<sup>-1</sup> and 5 to 15  
566 % when looking at the relative uncertainty using linear interpolation as the gap-filling method.  
567 These differences with our results could be explained by the different methods adopted. Their  
568 results also prove that a combination of both methodologies may improve the gap-filling as their  
569 N<sub>2</sub>O annual budget uncertainties were lower with the linear interpolation than with the ANN  
570 methodology, despite the fact that they found a better R<sup>2</sup> for the ANN. These uncertainty  
571 differences among years can be explained by the disparities in the number and size of gaps and in  
572 the intensity of N<sub>2</sub>O emissions among years: years with a high N<sub>2</sub>O budget (related to high N<sub>2</sub>O  
573 fluxes), especially observed during maize periods, and with a small number of gaps lead to a low  
574 gap-filling uncertainty whereas the uncertainty increases when the length of gaps increases and  
575 N<sub>2</sub>O flux intensity is low, especially for wheat and bare soil periods. Nevertheless, the results  
576 show that the proposed mixed gap-filling methodology performs well when used on annual N<sub>2</sub>O  
577 flux datasets with a relative uncertainty below 5 % for each year.

578

579

580



	Annual N <sub>2</sub> O budget after gap-filling (kgN ha <sup>-1</sup> )	Sd (kgN ha <sup>-1</sup> )	Relative uncertainty (%)	Number of missing values (days)
2012	4.21	0.128	3.0	55
2013	2.75	0.025	0.9	8
2014	7.46	0.198	2.8	137
2015	3.01	0.124	4.1	127
2016	3.59	0.127	3.4	58

**Table 7.** N<sub>2</sub>O budget gap-filling uncertainty per civil year. Mean and standard deviation were calculated using the 100 repetitions.

## 585 **Conclusion**

586 Three different gap-filling methodologies were examined with respect to their performance  
587 for filling gaps in long-term N<sub>2</sub>O flux data series. This study has demonstrated that the use of an  
588 artificial neural network could be a helpful tool to gap-fill daily N<sub>2</sub>O fluxes as already  
589 recommended for other greenhouse gases such as CO<sub>2</sub> and CH<sub>4</sub>. On average, ANN proved to give  
590 better gap-filling performance than linear interpolation when looking at the R<sup>2</sup> and RMSE scores,  
591 even though linear interpolation performed better when gap-filling N<sub>2</sub>O fluxes of low intensity.  
592 Thus, the study proposes a new approach to gap-fill N<sub>2</sub>O emissions by combining the advantages  
593 of both methodologies. Such a procedure, which is not very costly in computing time and in  
594 requested variables (4 to 5 in our case), may be very helpful for future works on N<sub>2</sub>O emission  
595 analysis, especially on sites where numerous continuous variables are monitored (fluxnet and  
596 ICOS sites for example). In particular, it may improve the calculated N<sub>2</sub>O budget with limited  
597 gap-filling uncertainty, below 5% in this study. Moreover, given the strong non-linearity of N<sub>2</sub>O  
598 emissions due to multiple interactions between controlling environmental variables, ANN has the  
599 advantage of reproducing a complex process without any a priori knowledge of the system, and is

600 flexible and easy to use. The GMD thresholds developed in this study figures to be site specific  
601 as they highly depend on the ANNs performances and thus on quality and number of explanatory  
602 variables that may strongly vary among study sites. That saying, as the eddy covariance  
603 methodology is becoming more and more used across agricultural sites and for processes studies,  
604 besides the fact that higher available dataset could improve the methodology performance, further  
605 evaluation of the mixed gap-filling methodology should be carried out for N<sub>2</sub>O fluxes measured  
606 and compiled at higher resolution than the day (e.g. 6-hourly data).

607

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723  
724

<b>Variable names</b>	<b>Variable description (unit)</b>
ABG	Above Ground Biomass ( $\text{kg m}^{-2}$ )
E	Measured Evapotranspiration ( $\text{kg m}^{-2} \text{d}^{-1}$ )
ETP	Potential Evapotranspiration (mm)
GPP	Green Primary Production ( $\text{gC m}^{-2} \text{d}^{-1}$ )
H	Sensible heat flux ( $\text{watt m}^{-2} \text{d}^{-1}$ )
h veg	Vegetation height (m)
LAI	Leaf Area Index ( $\text{m}^2 \text{m}^{-2}$ )
LE	Latent heat flux ( $\text{W m}^{-2} \text{d}^{-1}$ )
LW	Incoming long wave ( $\text{W m}^{-2} \text{d}^{-1}$ )
NEE	Net Ecosystem Exchange ( $\text{gC m}^{-2} \text{d}^{-1}$ )
P	Rain (mm)
Pa	Pressure (kPa)
PPFD	Photosynthetic Photon Flux Density ( $\mu\text{mol m}^{-2} \text{d}^{-1}$ )
PWC	Plant Water Content (%)
Reco	Ecosystem respiration from soil ( $\text{gC m}^{-2} \text{d}^{-1}$ )
RH	Relative humidity (%)
Rg	Global radiation ( $\text{MJ m}^{-2} \text{d}^{-1}$ )
Rn	Net radiation ( $\text{W m}^{-2} \text{d}^{-1}$ )
SWC	Soil Water Content at 0, 5, 10, 30, 50 and 100 cm (%)
Ta_max	Maximum air temperature of the day ( $^{\circ}\text{C}$ )
Ta_mean	Mean air temperature of the day ( $^{\circ}\text{C}$ )
Ta_min	Minimum air temperature of the day ( $^{\circ}\text{C}$ )
Ts	Soil temperature at 0, 5, 10, 30, 50 and 100 cm ( $^{\circ}\text{C}$ )
Waterfall	Daily rain and irrigation event (mm)
WS	Wind speed ( $\text{m s}^{-1}$ )

<b>Growing season</b>	<b>Variables used to develop ANNs</b>
<b>Maize 2012</b>	LAI, Ta_mean, Rn, ETP, SWC_50
<b>Maize 2014</b>	LAI, Ta_mean, Rn, ETP, SWC_50
<b>Maize 2015</b>	LAI, Rn, ETP, SWC_50, H
<b>Wheat 2013</b>	Fertilization, Ts_50, PWC, h_veg, H
<b>Wheat 2016</b>	Fertilization, Ta_mean, Rn, H, h_veg
<b>CC 2013</b>	Ta_mean, Rn, PWC, h_veg
<b>CC 2016</b>	Ts_50, Rn, SWC_50, h_veg, ETP
<b>Bare soil periods</b>	Ta_mean, Rg, ETP, Fertilization

table 4.png

<b>Growing season</b>	<b>Methodology</b>	<b>R<sup>2</sup></b>	<b>RMSE (gN ha<sup>-1</sup> d<sup>-1</sup>)</b>	<b>RRMSE (%)</b>
<b>Maize 2012</b>	ANN	0.86	10.8	43
	Linear interpolation	0.85	12.2	49
	Linear interpolation and ANN combine	0.87	10.6	43
<b>Maize 2014</b>	ANN	0.85	33.9	73
	Linear interpolation	0.58	55.6	119
	Linear interpolation and ANN combine	0.85	34.6	74
<b>Maize 2015</b>	ANN	0.94	13.8	82
	Linear interpolation	0.93	14.3	79
	Linear interpolation and ANN combine	0.97	13.4	77
<b>Wheat 2013</b>	ANN	0.54	4.0	65
	Linear interpolation	0.66	3.6	59
	Linear interpolation and ANN combine	0.67	3.5	57
<b>Wheat 2016</b>	ANN	0.60	8.8	115
	Linear interpolation	0.79	6.6	86
	Linear interpolation and ANN combine	0.69	7.8	102
<b>CC 2013</b>	ANN	0.93	4.0	62
	Linear interpolation	0.53	9.2	145
	Linear interpolation and ANN combine	0.93	4.0	62
<b>CC 2016</b>	ANN	0.86	10.0	76
	Linear interpolation	0.64	14.7	111
	Linear interpolation and ANN combine	0.85	10.1	77
<b>Bare soil periods</b>	ANN	0.66	8.6	114
	Linear interpolation	0.76	7.3	96
	Linear interpolation and ANN combine	0.68	8.6	113
<b>All functioning periods together</b>	ANN	0.84	12.4	101
	Linear interpolation	0.68	17.4	141
	Linear interpolation and ANN combine	0.84	12.4	100

table 5.png

<b>Vegetation periods</b>	<b>Before gap-filling</b>	<b>Gap-filling with ANN</b>	<b>Gap-filling with linear interpolation</b>	<b>Gap-filling with combine methodologies</b>
<b>Maize 2012</b>	25.6 (± 29.9)	24.0 (± 28.2)	23.4 (± 28.5)	23.7 (± 28.4)
<b>Maize 2014</b>	42.9 (± 77.7)	45.5 (± 78.7)	64.5 (± 107.2)	45.5 (± 78.7)
<b>Maize 2015</b>	16.3 (± 32.2)	12.8 (± 25.1)	15.8 (± 25.9)	12.8 (± 25.1)
<b>Wheat 2013</b>	6.6 (± 6.3)	6.6 (± 6.3)	6.5 (± 6.3)	6.5 (± 6.3)
<b>Wheat 2016</b>	7.9 (± 12.5)	7.8 (± 12.4)	7.8 (± 12.4)	7.8 (± 12.4)
<b>CC 2013</b>	9.3 (± 17.2)	8.8 (± 16.7)	9.0 (± 16.6)	8.8 (± 16.7)
<b>CC 2016</b>	13.0 (± 23.4)	13.1 (± 21.5)	10.9 (± 20.8)	12.6 (± 21.5)
<b>Bare soil periods</b>	7.5 (± 15.0)	6.3 (± 13.0)	21.3 (± 58.0)	6.1 (± 13.0)



	<b>Methodology</b>	<b>GMD &lt; 14</b>	<b>14 &lt; GMD &lt; 30</b>	<b>GMD &gt; 30</b>
<b>RMSE (gN ha d<sup>-1</sup>)</b>	<b>Linear interpolation</b>	3.9	11.6	43.7
	<b>ANN</b>	8.8	10.9	27.2

	Annual N <sub>2</sub> O budget after gap-filling (kgN ha <sup>-1</sup> )	Sd (kgN ha <sup>-1</sup> )	Relative uncertainty (%)	Number of missing values (days)
2012	4.21	0.128	3.0	55
2013	2.75	0.025	0.9	8
2014	7.46	0.198	2.8	137
2015	3.01	0.124	4.1	127
2016	3.59	0.127	3.4	58

fig. 1.png

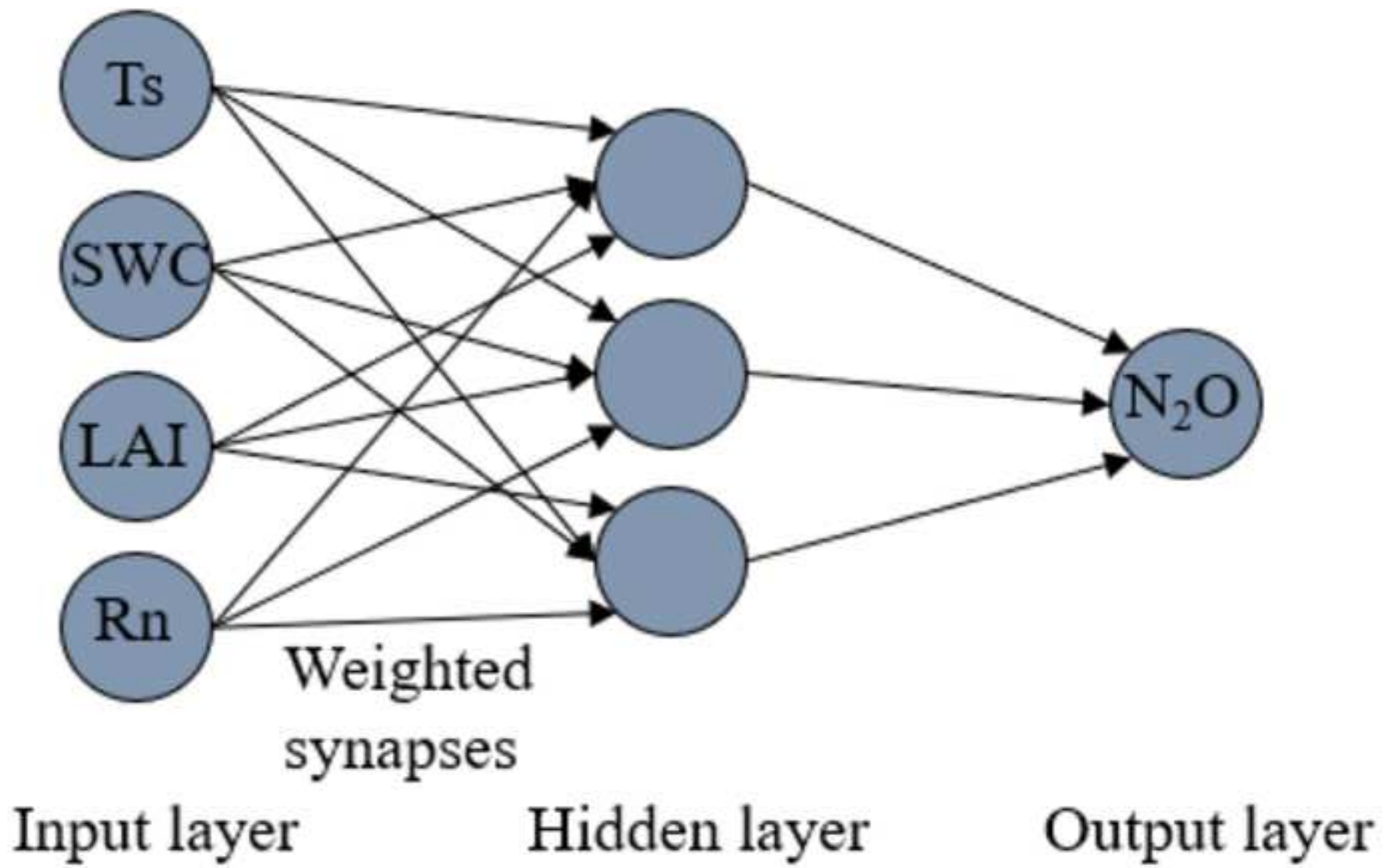


fig. 10. a.png

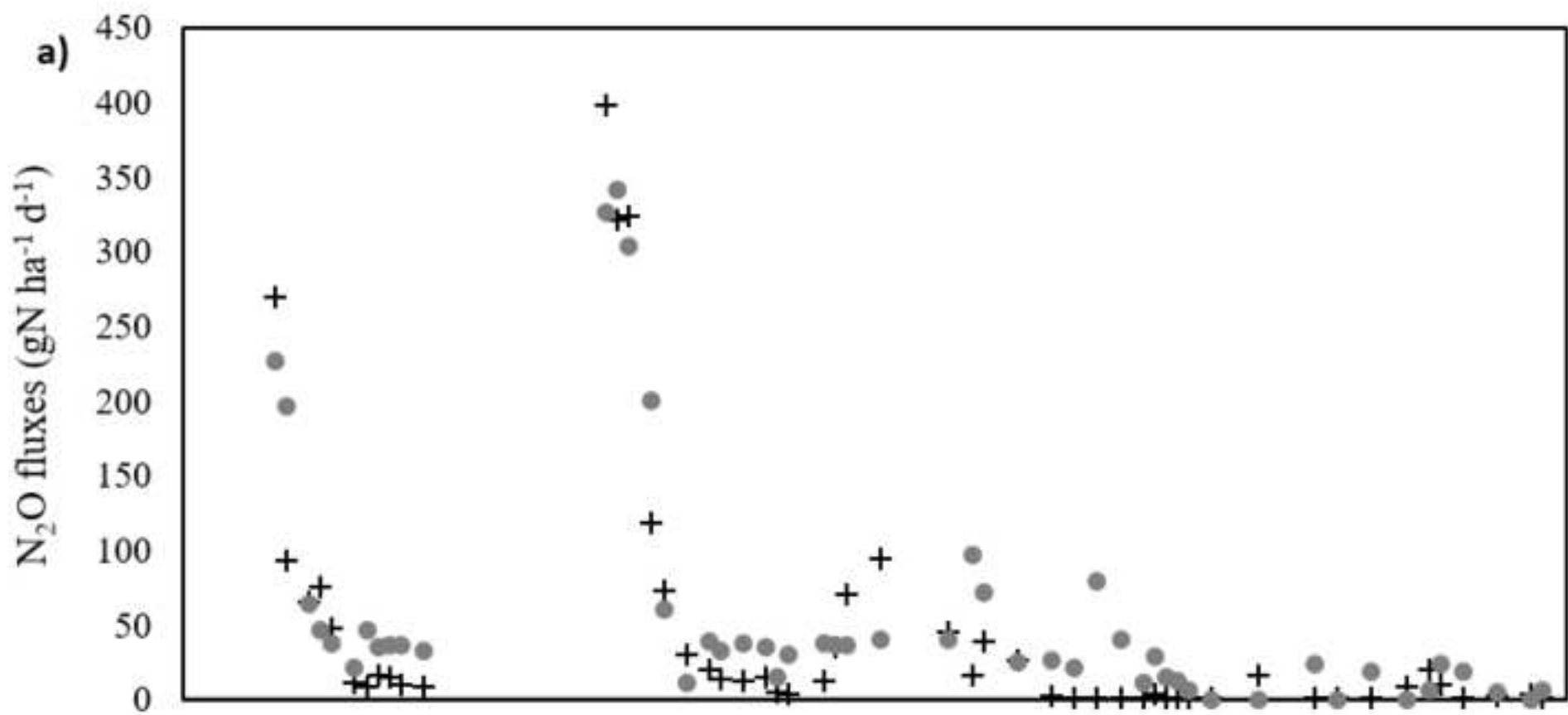






fig. 11. a.png

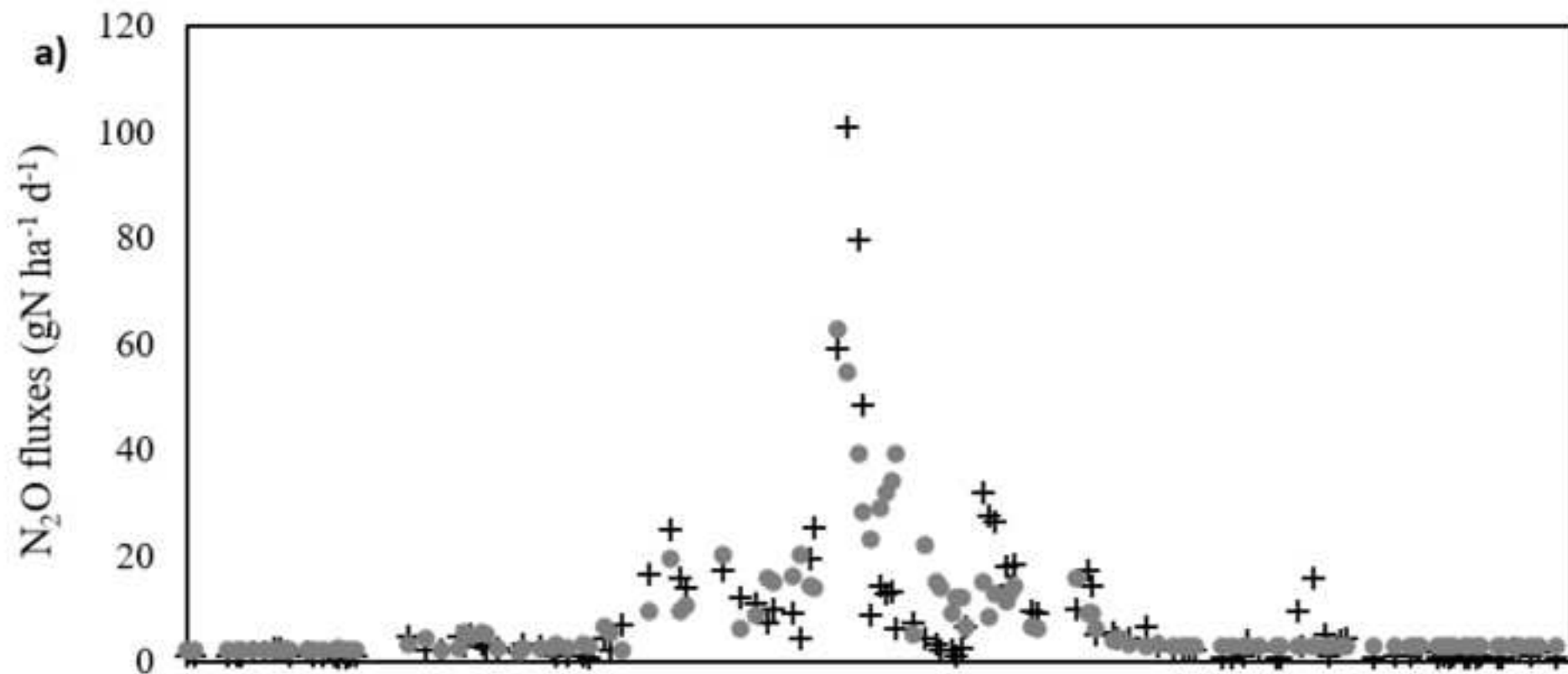


fig. 11. b.png

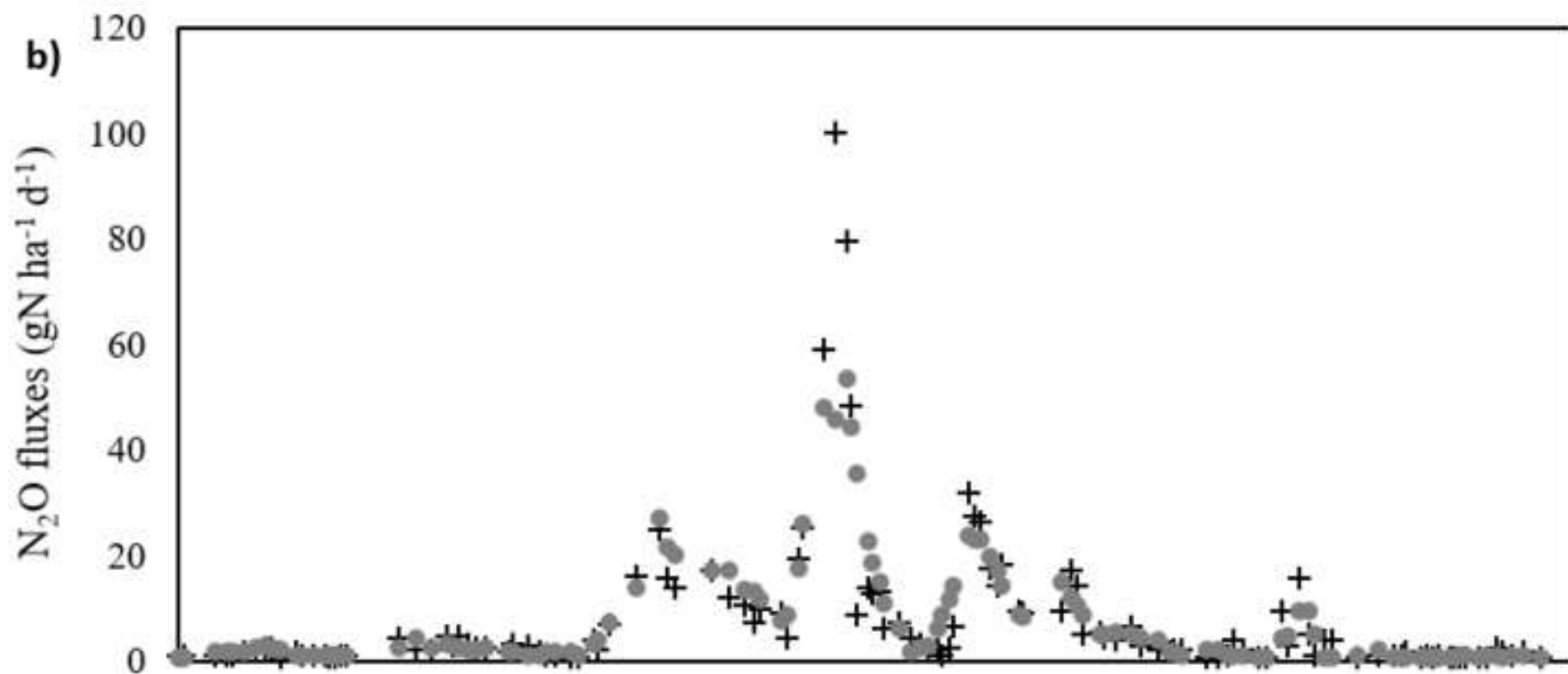




fig. 11. c.png

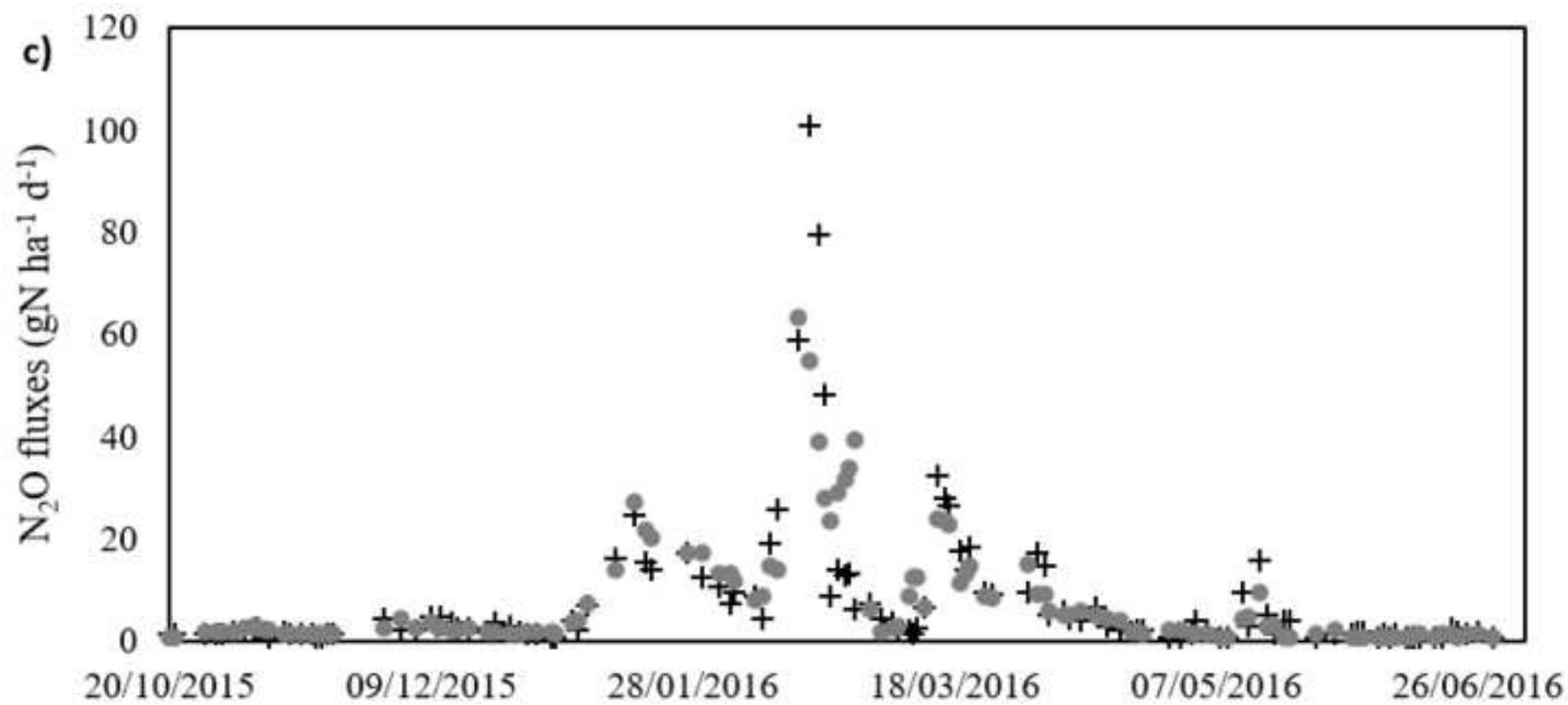
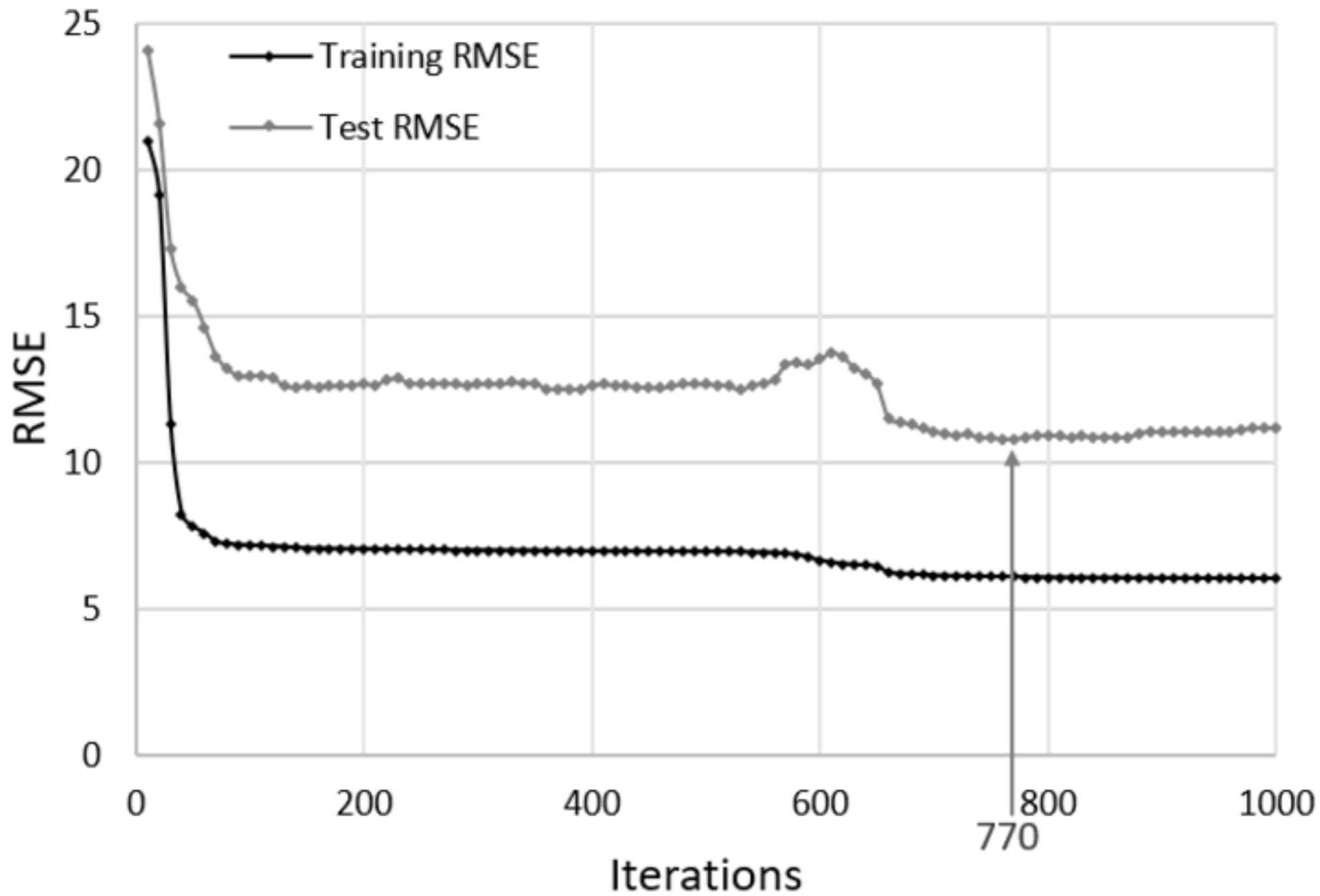
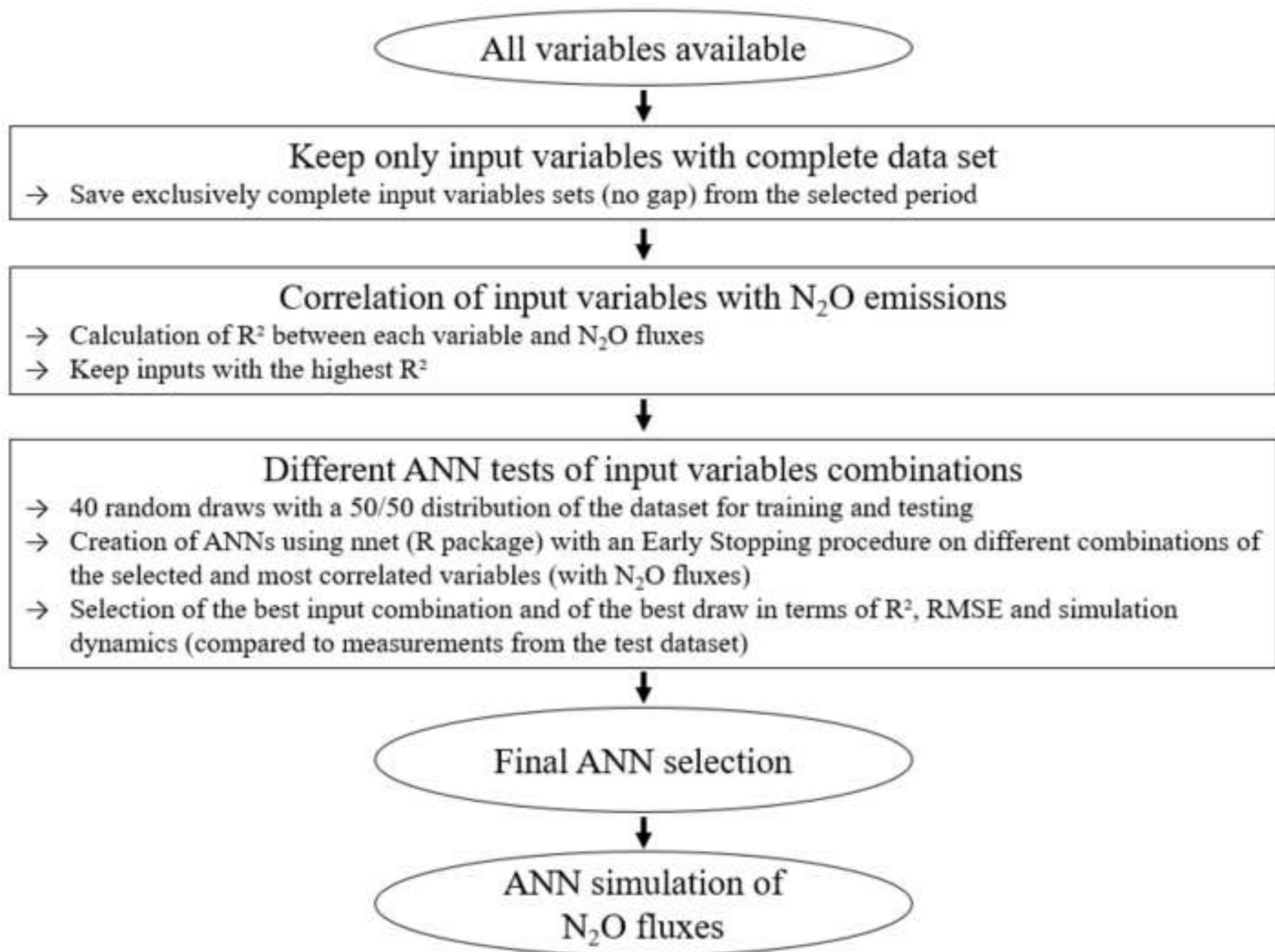


fig. 2.png





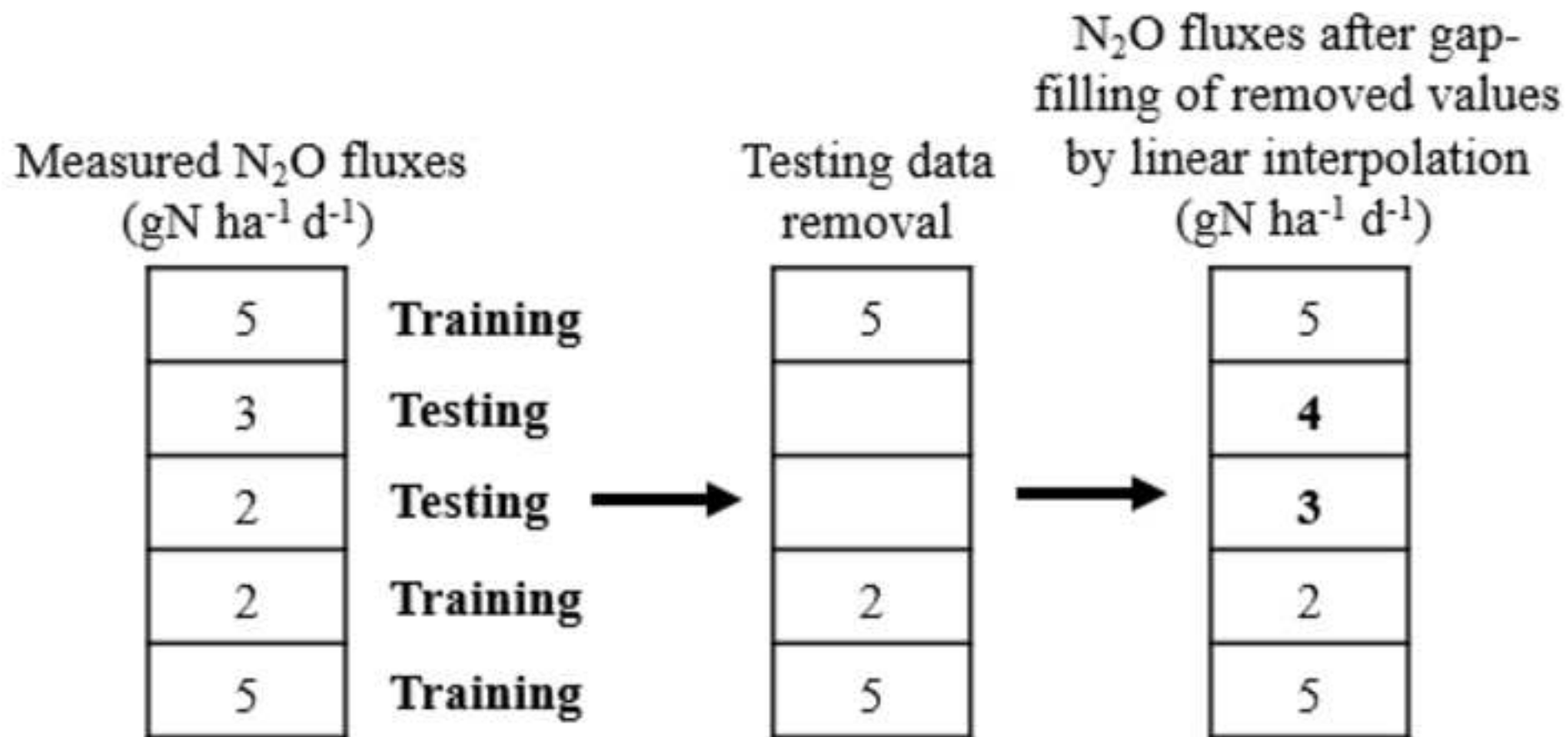
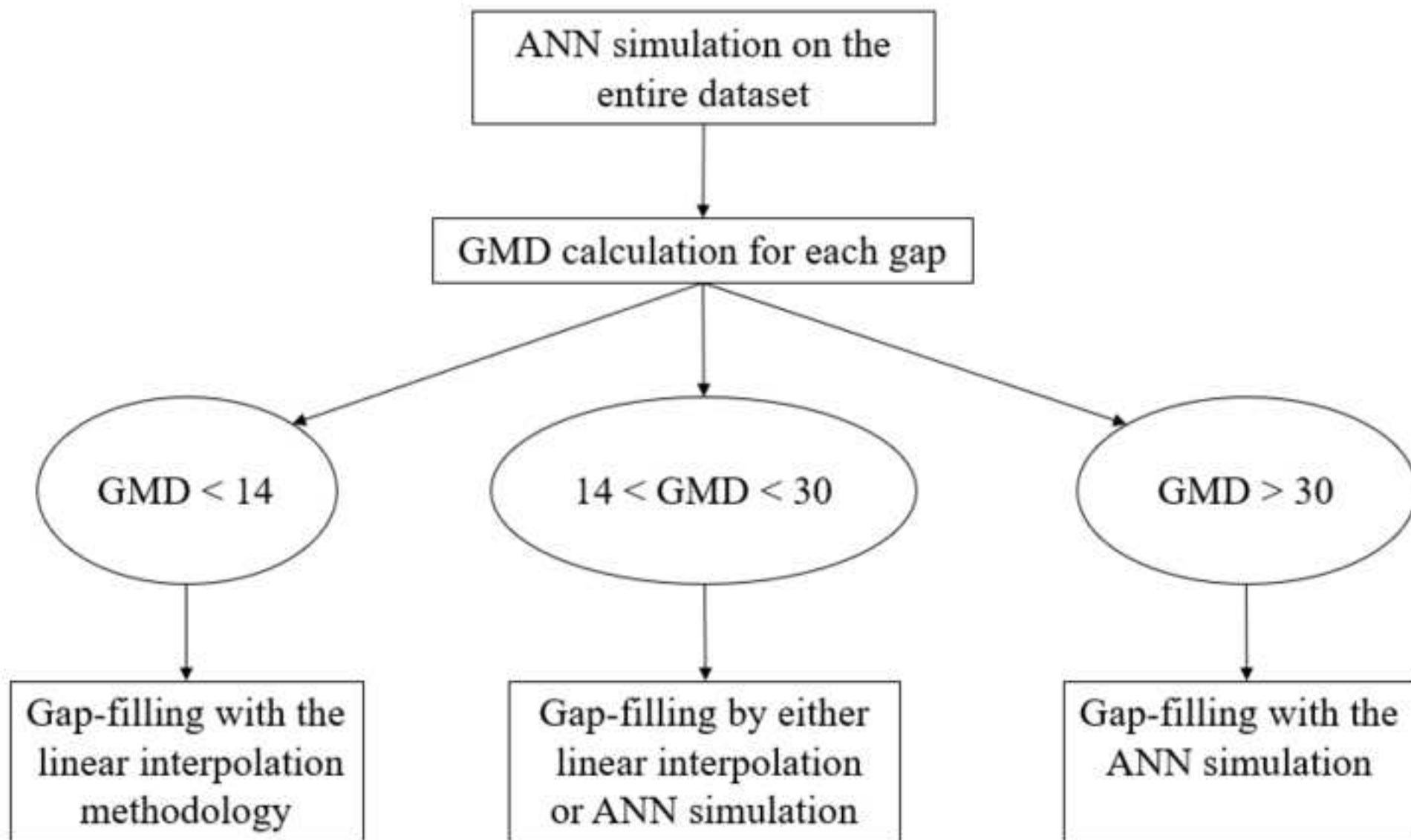


fig. 5.png



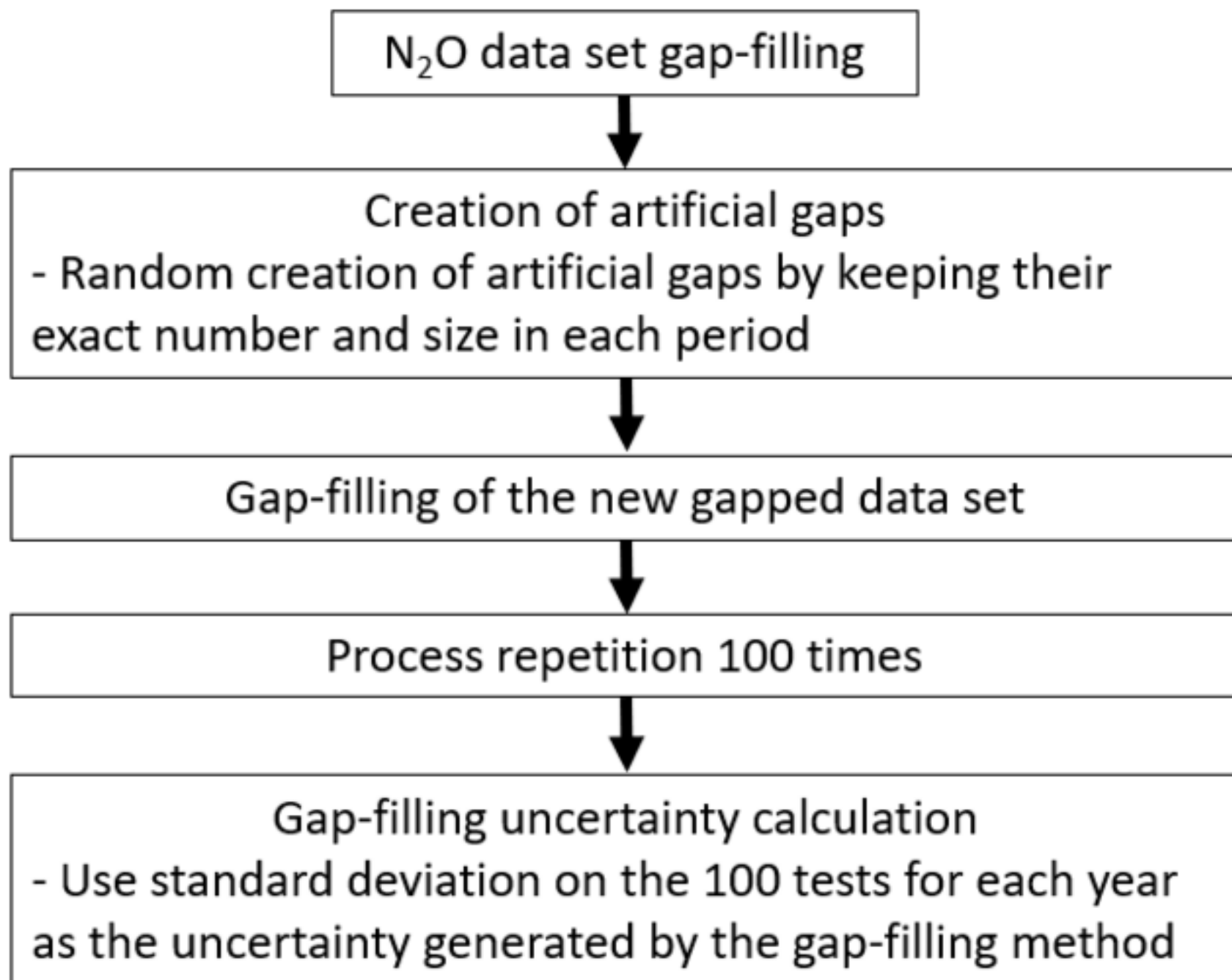


fig. 7.png

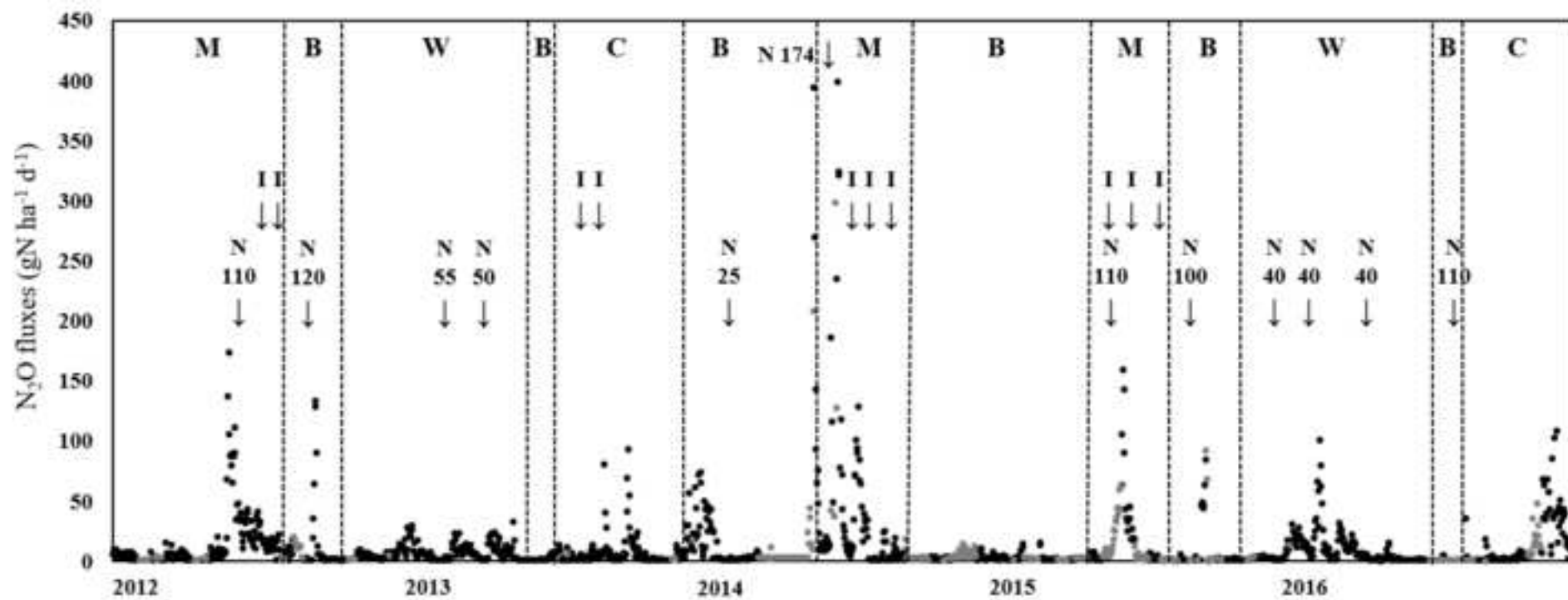


fig. 8.png

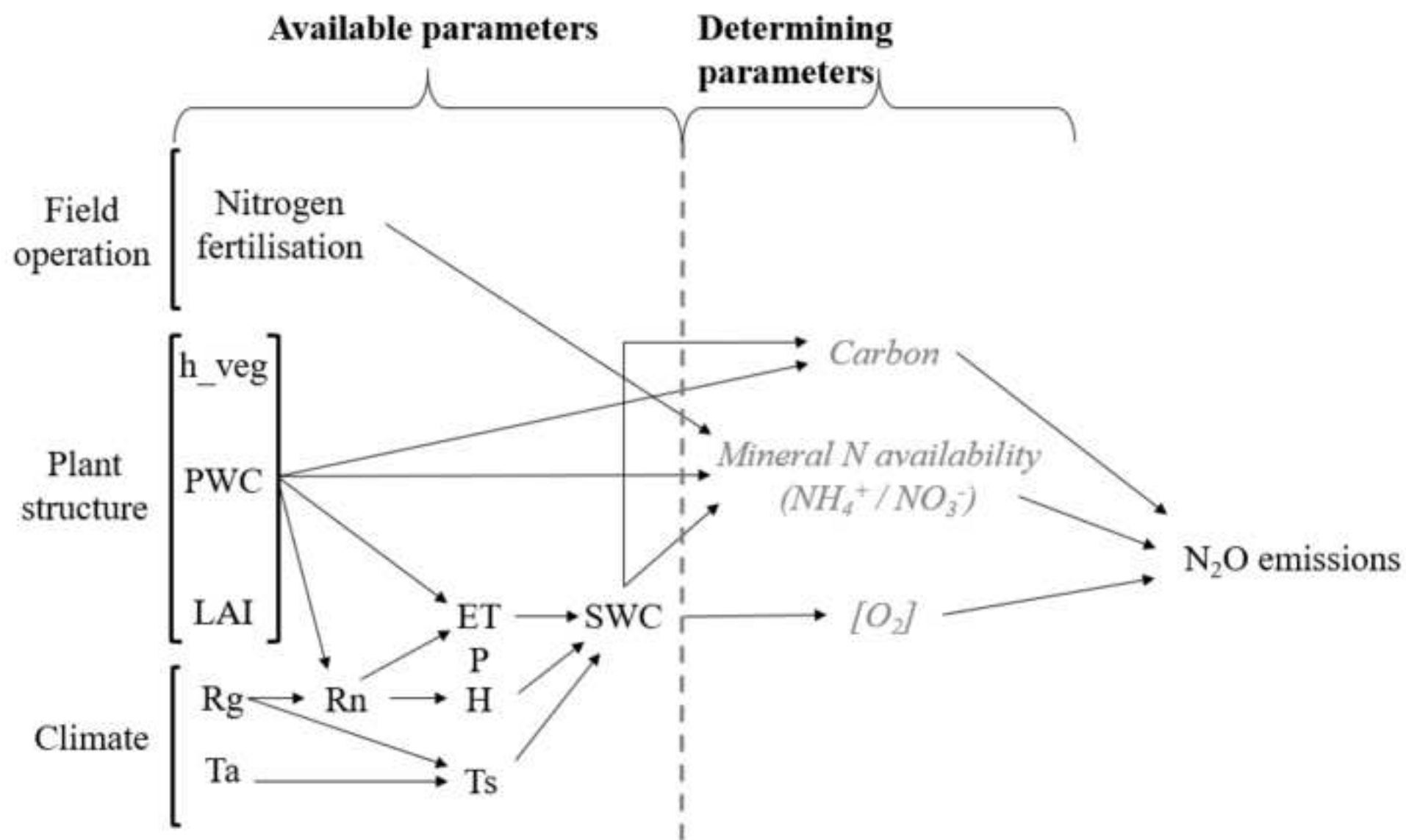




fig. 9.png

