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1	Can a herbivore affect where a top predator kills its prey by modifying woody
2	vegetation structure?
3	
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The first and second author contributed equally to this paper. NF carried out the statistical analyses and drafted the manuscript. MM collected field data and drafted the manuscript. MV conceived and designed the study, coordinated the study, and critically revised the manuscript. AJL designed the study, coordinated field data collection and critically revised the manuscript. SD helped with the statistical analyses, the interpretation of the results and revised the manuscript. HF and DWM helped coordinate the study, interpret the results and revised the manuscript.

16 Abstract

17 In large mammal communities, little is known about modification of interspecific 18 interactions through habitat structure changes. We assessed the effects of African 19 elephants (Loxodonta africana) on features of woody habitat structure that can affect 20 predator-prey interactions. We then explored how this can influence where African lions 21 (Panthera leo) kill their prey. Indeed, lions are stalk-and-ambush predators and habitat 22 structure and concealment opportunities are assumed to influence their hunting success. 23 During two years, in Hwange National Park, Zimbabwe, kill sites (n=167) of GPS-24 collared lions were characterized (visibility distance for large mammals, distance to a 25 potential ambush site and presence of elephant impacts). We compared characteristics of 26 lion kill sites with characteristics of random sites (i) at a large scale (i.e, in areas 27 intensively used by lions, n=418) and (ii) at the microhabitat scale (i.e., in the direct 28 surrounding available habitat, < 150m, n = 167). Elephant-impacted sites had a slightly 29 higher visibility and a longer distance to a potential ambush site than non-impacted sites, 30 but these relationships were characterized by a high variability. At large scale, kill sites 31 were characterized by higher levels of elephant impacts compared to random sites. At 32 microhabitat scale, compared to the direct nearby available habitat kill sites were 33 characterized by a reduced distance to a potential ambush site. We suggest a conceptual 34 framework whereby the relative importance of habitat features and prey abundance could 35 change upon the scale considered.

36 Key words: ecosystem engineer, indirect effects, apex predator, megaherbivores,

37 predator-prey relationships.

38 Introduction

39 Species within an ecosystem are linked by a network of interspecific interactions (e.g. 40 predation, competition, facilitation), which ultimately drives ecosystem functioning 41 (Wardle et al. 2004). There is an increasing awareness that these interactions are dynamic 42 and can be mediated by abiotic (e.g. climate change - Tylianakis et al. 2008, van der 43 Putten et al. 2010) and biotic factors (e.g. parasitism - Hatcher et al. 2006, non-lethal 44 effects of predators that mediate interspecific prey competition - Preisser & Bolnick 45 2008). In these cases, the interaction between two species can be modified by a third 46 species (hereafter "interaction modification", Wootton 1993). This process can arise from 47 a change of a plastic trait of one of the two main species interacting (i.e. trait-mediated interaction modification) or through alteration of the environment in which the 48 49 interaction takes place (i.e. environment-mediated interaction modification, Wootton 50 1993, 2002, Dambacher & Ramos-Jiliberto 2007). 51 Questions have arisen about how habitat changes (diversity and/or physical 52 structure) may affect interspecific interactions (Petren & Case 1998). In the current 53 context of biodiversity loss, many studies have focused on anthropogenic alterations of 54 the habitat (e.g. Tylianakis et al. 2007), but other ecosystem engineer species (Jones et al. 55 1994) can affect habitats (Crooks 2002), and ultimately interspecific interactions 56 (Marquis & Lill 2007). Arditi et al. (2005) even designated ecosystem engineer species as "interaction modifiers" due to their capacity to modulate their environment. Interaction 57 58 modifications were shown to drive community dynamics in systems with few species 59 (Werner & Peacor 2003, Preisser et al. 2007, Abrams 2010). There is now growing 60 impetus from other recent works to address the challenges of detecting, measuring and

testing the potential role of interaction modifications in complex systems such as natural
communities (e.g. Wootton 1994, Peacor & Werner 2001, Okuyama & Bolker 2007).
Such an understanding is key to improve our ability to forecast how ecosystems will
respond to global changes (Kéfi et al. 2012) as interaction modifications are often
identified as the cause of unexpected responses to perturbation (Terry et al. 2017 and
references therein).

67 The African elephant (Loxodonta africana) is an ecosystem engineer (Bond 1994). 68 While the effects of elephants on vegetation structure through their foraging activity start 69 to be well understood (review in Guldemond & van Aarde 2008), the consequences of 70 elephant-induced vegetation changes on the whole ecosystem remain unknown as a 71 diversity of indirect effects is documented (Pringle 2008, Valeix et al. 2011, Coverdale et 72 al. 2016). In particular, little is known about the environment-mediated modifications of 73 predator-prey relationships by elephants. Yet, elephants affect the vegetation structure, 74 especially in the understory (Coverdale et al. 2016, Ferry 2018). Further, predation is 75 mediated by physical features of habitats (Bell 1991, Kauffman et al. 2007) and has 76 cascading effects down the food chain (Estes et al. 2011). To our knowledge, two studies 77 highlighted elephant-induced modification of predator-prey relationships. Tambling et al. 78 (2013) showed that elephants, by fragmenting very dense vegetation, improve access for 79 lions, which may ultimately lead to an increased predation by lions on the small prey 80 hiding in this very dense vegetation. Fležar et al. (2019) simulated elephant-induced 81 habitat change at two spatial scales: (i) at the "patch" scale, by comparing high-quality 82 grassland sites with high visibility against ones with low visibility (due to dense woody 83 vegetation) and (ii) at the "within-patch" scale by adding coarse woody debris, potential

84 escape impediment for prey, in open areas. They then assessed the perceived predation 85 risk by different herbivores. They revealed different responses of prey at the two scales 86 and argue that depending on the scale, elephants' impact on the risk landscape could be 87 both to hamper kill success (by opening up vegetation, improving visibility and lowering 88 ambush opportunity) as well as facilitate kill success (by dropping woody debris that may 89 lower visibility and create escape impediments). Elephants are thus able to modify 90 predator-prey interactions by altering habitats and different manifestations of elephantinduced changes on the vegetation (e.g., visibility and coarse woody debris) could act at 91 92 different spatial scales.

93 Here, we investigated whether elephants, through their impacts on vegetation 94 structure (that lead to changes in visibility distance for large mammals and changes in the 95 distance to a potential ambush site), can influence predator-prey interactions between 96 African lions (Panthera leo) and their prey in a woody savanna ecosystem. Lions are 97 stalk-and-ambush predators that rely on features of the habitat providing concealment 98 (typically dense vegetation) to approach and attack their prey (Hopcraft et al. 2005, 99 Loarie et al. 2013, Davies et al. 2016). Therefore, habitat characteristics are expected to 100 play an important role in selecting areas that may increase hunting success (the *ambush*-101 *habitat* hypothesis - Hopcraft et al. 2005). This has been illustrated in Kruger National 102 Park, South Africa, where lions kill their prey within nine meters of a potential ambush site (Loarie et al. 2013). Elephants are thus likely to affect where lions hunt and/or 103 104 successfully hunt (i.e. kill) their prey. 105 The aim of this study is two-fold: (1) to assess whether elephant impacts on

106 woody vegetation are associated with an increased visibility and a change in the distance

107 to a potential ambush site, and (2) to test the hypothesis that lions kill less in areas 108 impacted by elephants (as we expect them to be more successful hunters in areas with 109 denser vegetation thus greater opportunities for concealment). This second aim was 110 investigated at two different scales: (i) we first compared lion kill sites with random sites 111 in areas intensively used by lions to assess if among all habitats used by lions, kill sites 112 were characterized by denser vegetation and less elephant impacts (the "large" scale 113 hereafter), and (ii) we then compared the characteristics of lion kill sites with 114 characteristics of the direct surrounding available habitat (< 150 m) to assess if lions 115 killed more in closed microhabitats that were less impacted by elephants (the 116 "microhabitat" scale hereafter). Together, the results will allow an assessment of the 117 extent to which elephants can induce environment-mediated trophic interaction 118 modification between lions and their prey in woodland savannas and if this modification 119 is scale-dependent.

120

121 Materials and methods

122 Study site - Hwange National Park covers ~15 000 km² of semi-arid dystrophic (low nutrient soil) savanna in western Zimbabwe (19°00' S, 26°30' E). The vegetation is 123 124 primarily woodland and bushland savanna. The east and southern parts of the park are 125 dominated by open wooded savannas on Kalahari sands, primarily teak woodland (Baikiaea plurijuga) and Combretum/Terminalia woodlands. Batoka basalt and Karoo 126 127 sediments in the north and north-west of the park are dominated by *Colophospermum* 128 mopane woodlands interspersed with grassland vleis. The long-term mean annual rainfall 129 is ~ 600 mm, which falls primarily between October and April. The surface water

130 available to animals is found in natural as well as artificial waterholes. The study area is 131 located in the northern region of Hwange National Park (~7 000km²) where lion density 132 is estimated around 4.3 individuals/100 km² (Loveridge et al. 2016), and elephant density 133 is estimated above 2 individuals/km² (Chamaillé-Jammes et al. 2008). 134 Data – We collected data between 2014 and 2015 from 12 female and 15 male lions from 135 different coalitions and prides equipped with 2D size AWT GPS radio-collars. The lions' 136 locations were available hourly and for some lions every two hours, day and night. 137 Potential lion kills were attained by identifying clusters of coordinates that had more than 138 4 hours of sequential locations within a defined proximity (150 m, see also Tambling et 139 al. 2010). In the field, these clusters were searched for a carcass or the remains of a 140 carcass and classified as kill sites based on the evidence of a kill. We confirmed lion kills 141 when the presence of a carcass was associated with indications of a hunt / struggle from 142 animal tracks (observed by skilled field trackers) and / or broken and tramped vegetation 143 and / or from the condition of any remaining hide bearing claw and bite marks typical of 144 lion predation. Carcasses found were classified to species based on the body size of the 145 animal killed and the presence of identifiable material, such as horns, jaws, bones, and 146 hair. We made the assumption that the kill site is a good proxy of the environment within 147 which the lion decided to start the hunt, as lion is a stalk-and-ambush predator attacking 148 and killing prey at short distances (van Orsdol 1984, Haas et al. 2005). This assumption 149 has been made in several previous works (Davidson et al. 2012, 2013, Loarie et al. 2013, 150 Davies et al. 2016). In total, 705 clusters were monitored among which 167 were identified as kill sites and 538 were not (called "non-kill sites" hereafter). For the 167 kill 151 152 sites and for 251 non-kill sites, we identified a paired random site (with a random

153 direction, a random distance between 50 and 150m from the kill for kill sites and from 154 the GPS point identified as the start of the cluster for non-kill sites). In total, 418 random 155 sites were characterized and represented habitats intensively used by lions. Among these 156 random sites, 167 were associated to a kill site and represented the direct surrounding 157 available habitat. For each kill site, non-kill site and random site, we measured the 158 distance to a potential ambush site (DPAS hereafter, a potential ambush site was any 159 habitat feature able to conceal a lying lion, i.e. most of the time a dense bush in the study 160 ecosystem) and the visibility. Visibility at each site was assessed by using two 50 cm x 161 50 cm white boards. The two boards were set so that one board was at 10-60 cm 162 (representative of the height of a crawling lion) and the other was at 100–150cm 163 (representative of a standing lion). One person stood at the location of the kill or at the 164 centre of the random site, while another person held the boards, walked away from the 165 centre in the four cardinal directions and recorded the distance at which the person at the 166 centre of the site could not see each board anymore. The four distances obtained from the 167 four cardinal directions were then averaged ("visibility" hereafter). As lions are more 168 successful at capturing prey when attacks are launched at short distance (<7.6m for 169 Thomson's gazelle, 15m for wildebeest and zebras, Haas et al. 2005), elephant impacts 170 were assessed within a 25m radius of the kill for the kill sites, of the random point for the 171 random sites and of the GPS point identified as the start of the non-kill sites. The extent 172 of elephant impact was determined by the definition of five classes of percentage of trees 173 impacted by elephants (broken, coppiced and/or uprooted): class 0: no impact; class 1: [1-174 25%]; class 2: [26%-50%]; class 3: [51%-75%]; and class 4 : [76%-100%].

175 Analyses – Proximity to water is commonly thought to influence the level of herbivore 176 impacts on the vegetation (i.e. the "piosphere effect", Lange 1969), but this has recently 177 been debated in wild protected areas (Chamaillé-Jammes et al. 2009). We therefore 178 preliminarily checked the existence of a link between distance to water and the existence 179 of elephant impacts on the vegetation and found that sites (random sites and kill sites) 180 impacted by elephants were not located closer to waterholes than sites not impacted by 181 elephants (Kruskal-Wallis test, $\chi^2 = 5.51$, df = 3, p-value = 0.14). 182 *Effect of elephants on woody habitat structure* - Visibility at 100-150 cm was highly 183 correlated to visibility at 10-60cm (r = 0.91, t = 75, df = 1121, p < 0.001), so only results 184 on the visibility at 10-60 cm (visibility hereafter) were considered in the subsequent 185 analyses. We assessed the effect of the level of elephant impacts on (1) the visibility with 186 a simple linear model performed on log-transformed visibility data and on (2) the DPAS 187 with a truncated linear regression as data distribution was left truncated at 0 m on log-188 transformed DPAS data ('truncreg' package from open source Software R 3.3.1 R. 189 Development Core Team, 2014). All kill sites, all non-kill sites and all random sites were 190 included in this analysis to best describe the link between the level of elephant impact 191 and the vegetation characteristics. 192 *Lion kill site characteristics* - For the subsequent analyses, non-kill sites were excluded 193 as they could have represented any lion's activity (e.g., resting site). These sites could

194 have been under selection by lions (e.g., habitat with higher woody cover for shadow

195 preferred) and thus led to a bias in our results/interpretation. At the large scale, we

196 compared the characteristics of lion kill sites with characteristics of the habitats of all the

197 random sites (associated to kill sites and to non-kill sites), representing areas intensively

198 used by lions. We used logistic regressions to develop resource selection functions (RSF), 199 with the dependent variable being 1 for kill sites and 0 for random sites. We performed a 200 first logistic regression to assess if lions kill more in low visibility environments where 201 prey can be closely approached thanks to low DPAS. For this first logistic regression, the 202 explanatory variables are visibility and DPAS. No strong correlation was observed 203 between these two variables, which were therefore kept for the analyses (Pearson's 204 correlation coefficient visibility-DPAS = 0.38). We performed a second logistic 205 regression to assess if the level of elephant impacts on vegetation structure influences 206 lion kill site location. In this second logistic regression, the explanatory variable was the 207 level of elephant impacts. A model selection was performed using the function "dredge" 208 ('MuMin' package) using the Bayesian Information Criterion (BIC) for a compromise 209 between the explanatory power and the parsimony of the models and model averaging 210 was performed on all the models (Burnham & Anderson 2004). Variables considered as 211 important were those for which $\beta \pm 1.96$ *SE did not include zero. At the microhabitat 212 scale, we compared the characteristics of lion kill sites with the characteristics of the 213 direct surrounding available habitat (represented by the random site associated to each 214 kill site). A paired Generalized Estimating Equations (GEE) model was performed using 215 the package "gee" to remove all the variability between the different pairs and focus only 216 of the variability within each pair (Liang & Zeger 1986). We conducted the same two 217 regression analyses as above. For this analysis, the quasi-likelihood criterion (QIC) was 218 used (Liang & Zeger 1986) and a model averaging was performed on all the models. As 219 no difference between lion sexes was observed (Online Resource 1), all kill sites 220 identified were used and pooled together independently of whether the kill site was found

221	using GPS-collar data from a female or a male lion. Further, our data did not allow
222	assessing if the collared individual was the one that made the kill, and male and female
223	lions were regularly observed together (70.1% of all lions' observations) in Hwange
224	National Park at the time of the study.
225	
226	
227	Results
228	Kills were not evenly distributed over the different classes of shrub layer cover and over
229	the different prey species (Online Resource 2). The main prey of lions were greater kudu
230	Tragelaphus strepsiceros (27%), followed by African buffalo Syncerus caffer (20%) and
231	plains zebra Equus quagga (12 %, Online Resource 2). DPAS and visibility at kills for
232	each prey species are presented in Online Resource 3.
233	
234	Effect of elephants on woody vegetation structure – For each class of level of elephant
235	impacts (0: no impact; 1: [1-25%]; 2: [26%-50%]; 3: [51%-75%]; and 4: [76%-100%]),
236	the number of study sites (including all kill sites, non-kill sites and random sites) was
237	respectively: 453, 275, 205, 132, and 55. The log-visibility increased as the level of
238	elephant impacts increased (estimate \pm SE = 0.14 \pm 0.015, t = 9.04, p < 0.001, Table 1a,
239	Fig. 1a), and the log-transformed DPAS also increased as the level of elephant impacts
240	increased (estimate \pm SE = 0.17 \pm 0.02, t = 7.5, p < 0.001), Table 1b, Fig. 1b). On
241	average, there was a difference of 14m for the visibility (mean $_{\text{Level 0}} = 16.7$ m, mean $_{\text{Level 4}}$
242	= 30.7m) and 3m for the DPAS (mean $_{\text{Level 0}}$ = 2.4, mean $_{\text{Level 4}}$ = 5.4m) between habitats

not impacted by elephants and those with the highest level of elephant impacts. It is
noteworthy that there exists a high variability in the visibilities and the DPAS (Fig. 1).

246 *Lion kill site characteristics* - In the first analyses at large scale, comparing kill sites to 247 the all the random sites, representing available habitat in areas intensively used by lions, 248 we revealed that the level of elephant impacts was the only variable to explain lion kill 249 site characteristics (Table 2a). Lion kills were located in habitats with higher levels of 250 elephant impacts (estimate \pm SE = 0.27 \pm 0.09, Fig. 2a, see Online Resource 4 for raw 251 data). At the microhabitat scale, when we compared the characteristics of lion kill sites to 252 the characteristics of the direct surrounding habitat (within-pair comparison approach), 253 we revealed that the DPAS was the only variable to explain lion kill site characteristics 254 (Table 2b). Lion kill sites were preferentially located in habitats characterized by a 255 reduced DPAS compared to the direct nearby available habitat (estimate \pm SE = -0.44 \pm 256 0.19, Fig. 2b). In the kill sites, the mean DPAS value was 5.86 m, whereas it was 7.56 m 257 in the random sites representing a decrease of 1.7 m (22% of the mean DPAS value of 258 random sites).

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260

261 Discussion

In this study, we first assessed the effects of elephants on features of woody habitat structure that can be key for the ecology of predator-prey interactions, i.e. visibility and distance to a potential ambush site. Elephant-induced vegetation changes tended to be associated with an increase in visibility (as observed by Valeix et al. 2011). Regarding 266 distance to a potential ambush site, elephants could either increase it (e.g., by removing 267 large bushes or by reducing the crown diameter of bushes – see Ferry 2018) or reduce it 268 (e.g., by uprooting or breaking trees, which can create ambush sites behind the trunk, 269 branches and foliage on the ground). Overall, in Hwange National Park, elephant-induced 270 vegetation changes tended to be associated with an increase in distance to a potential 271 ambush site. Even though these average differences were not very large, they can make a 272 difference in dense habitats considering the hunting behaviour of lions, which kill their 273 prey close to dense vegetation (e.g. within 9 meters of a potential ambush site - Loarie et 274 al. 2013). Hence, elephants, by altering visibility and distance to potential ambush site, 275 are likely to affect where lions choose to hunt and/or where they hunt successfully in 276 woodland. Following the *ambush-habitat* hypothesis (Hopcraft et al. 2005), we initially 277 expected lions to kill more in habitats with lower level of elephant impacts and 278 characterized by lower visibility and a shorter distance to potential ambush site, thus 279 more favourable to lion hunting success (Fig. 3A– expected pattern). This assumption can 280 appear to be in opposition with the results from Tambling et al. (2013) and Davies et al. 281 (2016). This can be explained by the fact that, in these studies, habitats not impacted by 282 elephants were actually so dense (average distance to cover < 1 m) that lions were not 283 able to move and hunt inside this dense vegetation, which could be therefore used as a 284 refuge by small prey species (e.g., the duiker *Sylvicapra grimmia*). In this study, we were limited on the inferences we could make because of two main 285 286 limitations in our data. The first one is that we were not able to identify hunts in which 287 lions failed, which prevented us from assessing whether there were more kills in a habitat

288 because lions hunted more in this habitat or had a higher hunting success there. The

289 second limitation is the lack of information about the contextual abundance and 290 distribution of herbivores during the hunt, which could influence the kill site location as 291 expected under the *prey-abundance* hypothesis. To partly fill these gaps, we suggest a 292 conceptual framework with different scenarios that could explain the patterns observed 293 based on three different parameters: the probability of prey presence, the probability to 294 hunt (depending either on prey presence or on habitat openness), and the probability to 295 kill a prey (i.e. to hunt successfully) (Fig. 3B). Patterns 3,9 and 11 represent our initial 296 hypothesis, without assumption on prey distribution and with the probability to hunt 297 and/or kill being linked to habitat features only (following the *ambush-habitat* 298 hypothesis, with more hunt/kills in habitats less impacted by elephants, less open). 299 Contrary to our expectations, at the large scale, when we compared the 300 characteristics of lion kill sites to the characteristics of random sites in areas intensively 301 used by lions, kills were more located in woody habitats characterized by higher levels of 302 elephant impacts, but we did not detect a selection for a lower visibility and a shorter 303 distance to a potential ambush site. This result suggests that other factors than habitat 304 structural features drive lion hunting behaviour at this scale, such as the presence and 305 abundance of prey (i.e., the *prey-abundance* hypothesis, Hopcraft et al. 2005). If this is 306 the case, it assumes that herbivores select habitats impacted by elephants (representing all 307 the even numbered patterns in Fig. 3). This selection pattern may arise from (i) a 308 coincidence with elephants and other herbivores using the same habitats, (ii) a reduced 309 perceived risk of predation in elephant-impacted habitats due to the higher visibility 310 caused by elephants in these habitats for all herbivore species (Underwood 1982, Valeix 311 et al. 2011), and/or (iii) a facilitative effect of elephants that may increase browse

312 availability at lower heights within reach of smaller browsers by stimulating tree 313 coppicing, a mechanisms known as "browsing lawns" (Rutina et al. 2005, Fornara & du 314 Toit 2007). Hence, the fact that lion kills were preferentially found in elephant-impacted 315 habitats at the large scale could be explained by a selection for areas where prey are 316 abundant (*patterns* 6, 8, 14 and 16, Fig. 3) and elephants could be considered as 317 interaction modifiers if they influence prey habitat selection. Evidences about the role of 318 elephants in other herbivore woody habitat selection at this scale still need to accumulate 319 (e.g., herbivore distribution data thanks to camera traps placed on contrasted elephant-320 impacted habitats).

321 At the microhabitat scale (the within-pair comparison between a kill site and its 322 paired random site), results revealed that lion kills were not preferentially located in 323 habitats impacted by elephants anymore. At this scale, lion kill sites were preferentially 324 located in habitats characterized by a shorter distance to a potential ambush site (patterns 325 3, 4, 7-16, Fig. 3), supporting here our hypothesis of the role of prey catchability 326 (ambush-habitat hypothesis). Interestingly, the visibility did not seem to be a factor as 327 important as the distance to a potential ambush site. An explanation could be that, 328 whatever the visibility, the presence of a few large bushes / broken trees as potential 329 ambush sites is sufficient to lead to a higher probability of kill even in woody habitats 330 with a high visibility. Finally, when combining the two different scales, the only patterns 331 to explain the observed pattern (Fig 3A – observed pattern) with both more kills in 332 impacted habitat at the large scale and more kills in closed habitat at the microhabitat 333 scale are *Patterns* 8, 14 and 16. These patterns share the same processes: prey select 334 elephant-impacted habitats and a higher probability to hunt in habitat with more prey

(*prey-abundance* hypothesis). However, they differ in terms of probability to hunt or to
kill in closed habitats. *Pattern 8* needs a higher probability to kill in closed habitats, *Pattern 14* needs a higher probability to hunt in closed habitats and *Pattern 16* needs both
of them, suggesting therefore that lions are influenced by habitat structure during the
hunting process at the microhabitat scale (*ambush-habitat* hypothesis).

340 Therefore, our results suggest that the main driver of kill site location for lions is 341 likely to be prey abundance at a first scale of selection, and prey catchability at the scale 342 of the direct nearby available habitat (<150 m). As suggested in previous studies, the 343 *prey-abundance* and the *ambush-habitat* hypotheses are not exclusive and could interact 344 with one another to explain lion hunting behaviour (Davidson et al. 2012). Therefore, by 345 affecting the woody vegetation structure, elephants could play an important role in the 346 intensity of predator-prey relationships although in complex ways, as they could act on 347 both predators and prey's behaviour, with different mechanisms involved depending on 348 the scale considered (as suggested by Fležar et al. 2019). We encourage future research to 349 confirm that herbivores select woody habitats impacted by elephants because of the 350 elephant's engineering process and not because of simple coincidence or shared 351 resources. Further, a focus on identifying unsuccessful hunts will be needed to 352 disentangle the roles of the probability to hunt and the probability to kill in closed 353 habitats. This would ultimately help to know which process is influenced by the 354 vegetation structure during the lion hunting behaviour in woodland areas. This task is 355 both conceptually and practically a challenging one, although perhaps it can be 356 accomplished through the deployment of GPS-collars with integrated tri-axial 357 accelerometer-magnetometer (see for example Fröhlich et al. 2012, Wilmers et al. 2017).

358 Despite remaining questions regarding the underlying mechanisms, our study 359 suggests that elephants have the potential to influence predator-prey interactions in their 360 ecosystem. In a context of rapidly changing elephant populations worldwide (Chase et 361 al., 2016), it is of importance to understand their indirect role on interspecific 362 interactions. Our results reinforce the idea that elephants, through ecosystem engineering, 363 could act on a multitude of broad-scale ecological processes in wooded savannas (Kerley 364 & Landman 2006). Further, whereas previous studies of ecosystem engineers have 365 highlighted their effects on other species abundance and richness (Jones et al. 1997), our 366 findings demonstrate the importance of their indirect effect on interspecific interactions 367 (see also Arditi et al. 2005, Marquis & Lill 2007 and references therein). Finally, we 368 highlighted the importance of multi-scale consideration in interspecific interactions and 369 their modification (see also Fležar et al. 2019). We therefore hope these findings will 370 promote studies on interaction modification, with a multi-scale component (Tylianakis & 371 Morris 2017) in large mammal communities.

372

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Tables

a	% of impacted trees	2.5%	Mean	97.5%
0	0	15.7	16.7	17.8
1	[1,25]	17.5	20.5	24
2	[26,50]	20	23.7	28
3	[51,75]	20.6	24.8	29.9
4	[76,100]	24	30.7	39.1
b	% impacted trees	2.5%	Mean	97.5%
0	0	2.1	2.4	2.7
1	[1,25]	2.5	3.5	4.7
2	[26,50]	3.5	4.8	6.5
3	[51,75]	3.2	4.5	6.3
4	[76,100]	3.5	5.4	8.1

<u>Table 1</u>: Estimated mean and confidence interval at 95% for each level of elephant impacts of (a) the visibility (m) and (b) the distance to a potential ambush site (DPAS) (m).

Table 2: Logistic models examining (1) the effect of visibility (Vis) and distance to a potential ambush site (DPAS) on lion kill site location and (2) the effect of the level of elephant impacts (Ele) on lion kill site location. a) Approach at the large scale, comparing the characteristics of kill sites to characteristics of all random sites in areas intensively used by lions. b) Approach at the microhabitat scale, comparing the characteristics of kill sites to characteristics of paired random site representing the direct surrounding available habitat (< 150 m). Models are ranked according to their BIC or QIC. Model-averaged estimates for the variables \pm standard error are shown at the bottom of each table. Variables considered as important were those for which $\beta \pm 1.96$ *SE did not include zero.

$(1) - Kill \ sites / Random \ sites \sim DPAS + Vis$							
	Candidate models		df	BIC	Δi	Wi	bcc w _i
1	Null		1	702.2	0.00	0.453	0.453
2	DPAS		2	702.7	0.45	0.361	0.814
3	Vis + DPAS		3	704.3	2.06	0.162	0.976
4	Vis		2	708.1	5.89	0.024	1
Variable	Average β	SE					
Vis	0.22	0.25					
DPAS	-0.06	0.13					
(2) - Kill sites / Random sites ~ Ele							

a) Large scale - Kill sites VS All random sites

Candidate models df BIC Δi Wi bcc *w*_i 1 Ele 2 696.6 0 0.943 0.943 2 Null 1 702.2 5.62 0.057 1

Variable	Average β	SE
Ele	0.25	0.09

b) Microhabitat scale - Kill site VS Paired random site

(1) – Kill site / Paired random site ~ DPAS + Vis

	Candidate models		QIC	Δi	Wi	bcc w_i
1	Vis + DPAS		310.5	0.00	0.436	0.436
2	DPAS		310.6	0.13	0.408	0.844
3	Vis		312.7	2.22	0.144	0.988
4	Null		317.6	7.1	0.013	1
Variable	Average β	SE				
Vis	-0.33	0.27				
DPAS	-0.44	0.19				

(2) – Kill site / Paired random site ~ Ele

	Candidate models		QIC	Δi	Wi	bcc w _i
1	Null		317.6	0	0.596	0.596
2	Ele		318.4	0.78	0.404	1
Variable	Average β	SE				
Ele	0.004	0.07				

1 Figure legends

2	Figure 1: Boxplot distribution of a) the visibility and b) the distance to a potential ambush
3	site (DPAS) according to the five classes of level of elephant impacts, i.e. of percentage
4	of trees impacted by elephants (broken, coppiced and/or uprooted): class 0: no impact;
5	class 1: [1-25%]; class 2: [26%-50%]; class 3: [51%-75%]; and class 4 [76%-100%]. The
6	notch represents the 95% confidence interval of the median. Points represent raw data
7	using geom_jitter function from ggplot2 package (Wikcham 2016).
8	
9	Figure 2: (a) Relationship between the level of elephant impacts and the strength of this
10	factor on lions' kill site location at the large scale. (b) Relationship between the log-
11	transformed DPAS (for DPAS ranging from 0 to 50m) and the strength of this factor on
12	lions' kill site location at the microhabitat scale. The selection strength is
13	exp ($\beta 0+\beta 1*$ level of elephant impacts) at the large scale and exp ($\beta 0+\beta 1*$ log(DPAS+1))
14	at the microhabitat scale, where $\beta 0$ is the intercept estimate and $\beta 1$ is the estimated
15	coefficient for the level of elephant impacts at the large scale and for log(DPAS+1) at the
16	microhabitat scale. Dotted lines represent the standard errors.
17	
18	Figure 3: A) Representation of the expected pattern under our initial hypotheses and the
19	observed pattern. 1) Expected pattern - Under our initial hypotheses, we expected higher
20	visibilities and DPAS in habitats with higher levels of elephant impacts, as well as more
21	lion kill sites in habitats characterized by a lower visibility and a shorter DPAS, and thus
22	more kills in non-impacted habitats. 2) Observed pattern - An increased visibility and
23	DPAS were effectively observed with the increase of the level of elephant impacts but

24 not as strongly as expected (see the shape of the green area). At the large scale, lion kills 25 were, unexpectedly, more in highly elephant-impacted habitats. At the microhabitat scale, 26 lion kill sites were more in habitat characterized by a shorter DPAS. B) Representation of 27 the different scenarios envisaged to explain the observed pattern. We played on the 28 combination of three variables: the probability of prey presence, the probability that a 29 hunt will occur (with lions hunting more in high prey abundance habitat and/or with lions 30 hunting more in closed habitats), and 3) the probability of a kill, i.e. of a successful hunt 31 (with lions having a higher success rate in closed habitat). Patterns 8, 14 and 16 appear to 32 be the most likely to explain the observed pattern.

33

Figures

<u>Figure 1</u>



<u>Figure 2</u>



<u>Figure 3</u>

