

Indoor heterogeneous photochemistry of furfural drives emissions of nitrous acid

A. Depoorter, C. Kalalian, C. Emmelin, C. Lorentz, C. George

▶ To cite this version:

A. Depoorter, C. Kalalian, C. Emmelin, C. Lorentz, C. George. Indoor heterogeneous photochemistry of furfural drives emissions of nitrous acid. Indoor Air, 2020, —, 10.1111/ina.12758 . hal-02999119

HAL Id: hal-02999119 https://hal.science/hal-02999119v1

Submitted on 18 Nov 2020

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 Indoor heterogeneous photochemistry of furfural drives emissions of nitrous acid

2 A. Depoorter[†], C. Kalalian [†], C. Emmelin[†], C. Lorentz[†], C. George[†]

3 † Univ Lyon, Université Claude Bernard Lyon 1, CNRS, IRCELYON, F-69626, Villeurbanne,
4 France.

5 Abstract

6 People spend approximately 80% of their time indoor, making the understanding of the indoor 7 chemistry an important task for safety. The high surface-area-to-volume ratio characteristic of 8 indoor environments leads the semi-volatile organic compounds (sVOCs) to deposit on the 9 surfaces. Using a long path absorption photometer (LOPAP), this work investigates the formation of nitrous acid (HONO) through the photochemistry of adsorbed nitrate anions and 10 11 its enhancement by the presence of furfural. Using a high-resolution proton transfer reactiontime of flight-mass spectrometer (PTR-TOF-MS), this work also investigates the surface 12 emissions of VOCs from irradiated films of furfural and a mix of furfural and nitrate anions. 13 Among the emitted VOCs, 2(5H)-furanone/2-Butenedial was observed at high concentrations, 14 leading to maleic anhydride formation after UV irradiation. Moreover, the addition of 15 potassium nitrate to the film formed NO_x and HONO concentrations up to 10 ppb, which scales 16 to ca. 4 ppb for realistic indoor conditions. This work helps to understand the high levels of 17 HONO and NO_x measured indoors. 18

19

20 Keywords

21 Indoor air quality, Furfural, Nitrous Acid, Surface reaction, Photochemistry

22

24 **Practical implications**

- 25 Furfural is a ubiquitous pollutant indoors, which in the adsorbed state, induces photochemical
- 26 properties to indoor surfaces. This photochemistry is shown here to significantly enhance the
- 27 formation of nitrous acid and therefore impact the oxidation capacity indoors.

28 **1. Introduction**

Current atmospheric chemistry knowledge, built over the last decades, was found to be 29 insufficient when applied to indoor air chemistry¹. The indoor environment is a complex mix 30 of sources and sinks due to biogenic, human and microbial activities under unique conditions 31 (e.g., light, humidity and temperature)^{2,3}. For instance, volatile organic compounds (VOCs) are 32 often present in higher concentrations indoors than outside, affecting its air quality^{4–7}. Furfural 33 is a VOC emitted from multiple materials, generally produced under thermal stress from the 34 degradation of polyoses (hemicelluloses)^{3,8,9} through a mechanism previously described¹⁰. 35 Furfural is expected to be present indoor, emitted from surfaces of wood based products¹¹ which 36 are known to be an important source of indoor pollutants¹². Concentrations measured in Italy 37 and Finland reveal average levels of few $\mu g m^{-3}$ with levels reaching up to 39 $\mu g m^{-3}$ 9,13,14 in 38 homes, offices and public buildings, and was observed in dust in houses at about 20 μ g g^{-1 15} in 39 Sweden. Recently, experiments were conducted within the House Observations of Microbial 40 and Environmental Chemistry, aiming to understand the indoor chemistry in the average US 41 home. Average indoor concentration measured were similar to the concentrations observed in 42 Europe at about 4 μ g m⁻³ and decreased by a factor 10 when the room was ventilated. The 43 44 presence of furfural at comparable levels in Europe and the US highlights its importance independently of the architectural styles i.e., it is not limited to wooden houses. It was 45 determined that the time required for the furfural to reach steady state concentration indoor is 46 shorter than the average air renewable rate, indicating that surfaces act as an important reservoir 47 of this VOC¹¹. As high concentrations measured in some buildings can be linked to the presence 48 of books and wooden furniture^{9,14}, incense burning is also a major source of furfural leading to 49 harmful concentrations of the order of magnitude of 10¹ mg m^{-3 16}. The occupational exposure 50 limit was set to 8 mg m⁻³ over a period of 8 hours¹⁷. Furfural is regarded as a key product in the 51 chemical industries, hence its production was deeply researched in the 20th century¹⁸. It 52

currently has many uses in the industry as a solvent. It could also be used as a marker for indoor
air quality control due to his abundance and volatility ¹⁹.

Indoor surfaces are a medium where a wide range of compounds can react, offering a 55 complex chemistry that can be affected by indoor light. While volatile compounds are rapidly 56 ventilated out in most indoor environments, semi-volatile compounds (sVOCs) tend to adsorb 57 on surfaces due to the high surface-area-to-volume ratios (ranging from 1 to $2 \text{ m}^2 \text{ m}^{-3}$)^{20,21}, and 58 are therefore sustained for longer times. It is acknowledged that ozone indoor is lost primarily 59 through reactions on the surface leading to the formation of VOCs²². Other gas phase 60 compounds can react on the surface such as NO₂ through hydrolysis²³ as well as HONO that 61 can interact with indoor surfaces to produce hazardous compounds like nitrosamines²⁴. In 62 additions, nitric acid can be photolyzed when organic coating are present on the surface²⁵... 63 Surfaces with specific coatings such as paintings could also produce organic indoor air 64 pollutants like VOCs, NO_x and HONO when exposed to UV irradiation²⁶⁻²⁹. While indoor 65 irradiation is less intense than outside and high energy wavelengths below 330 nm are scarce, 66 photochemistry is still an important factor indoors^{30–34}. 67

The hydroxyl radical (OH) was found to be a major indoor oxidant, which was thought to 68 be mainly produced from the ozonolysis of unsaturated VOCs³⁵. However, a recent study by 69 Mendez et al.²¹ showed that the photolysis of nitrous acid (HONO) can also be a significant 70 source of OH radicals, contributing to more than 50% of their sources^{31,36}. HONO can be 71 72 produced heterogeneously through the processing of adsorbed nitrate anions or gaseous nitrogen oxides, which are ubiquitous indoors^{32,37}. Arata et. al.³⁸ recently reported direct 73 measurements of indoor N₂O₅ and NO₃, which can participate in reactions producing HNO₃. 74 75 Indoor nitric acid is not readily photolyzed and is likely to deposit onto surfaces, where water layers could lead to hydrolysis and production of surface nitrate. Nitrate was measured on 76 indoor surfaces with concentrations varying from 10^{-3} to 10^{-5} mol m⁻² ^{39,40}. 77

In this paper, the photochemistry of deposited furfural and its implication in converting 78 79 nitrogen oxides or adsorbed nitrate anions into gaseous HONO and NO_x were investigated through laboratory experiments. Furfural seems to be a ubiquitous VOC emitted from 80 surfaces^{9,11,13,14}, and here we suggest that its presence on these surfaces is conferring 81 photosensitizing properties to them. Those properties are tested here by investigating the 82 photoconversion of adsorbed nitrate to HONO, and the formation of surface peroxy radicals. 83 This knowledge will lead to a better description of surface photochemistry indoors, and possibly 84 closing a gap in our understanding of indoor HONO and hydroxyl radical formation. 85

86 2. Materials and Methods

As stated above, our aim is to explore the photosensitizing properties of adsorbed furfural representative of indoor wooden surfaces. We therefore prepared and exposed organic films made of deposited furfural, doped or not with nitrate anions, to light emitted from a LED in a small photo reactor described below. Gaseous products, including VOCs and HONO were monitored, and the production of surface peroxy radicals probed indirectly by their titration with NO.

94

93

2.1.1. Film preparation

2.1. Experimental protocol

Aqueous solutions of furfural and potassium nitrate were prepared in Milli-Q water (18.2 M Ω cm) with concentrations varying from 0.01 M to 0.5 M and 0.01 mM to 1 M, respectively. Two mL of a fresh solution was evenly dispersed on a circular borosilicate glass window (50 mm diameter). The glass window was weighed before the preparation, dried over a period of 12 hours in a desiccator filled with silica gel desiccant and placed in an oven at 323.2 ± 0.1 K. Once dried, the glass windows were weighed again and used in an experiment within one day.

101

2.1.2. Irradiation setup

102 The organic films deposited on the glass substrates were placed at the bottom of a borosilicate 103 glass reactor (height = 4 cm; volume = 80 cm³) and irradiated for 45 minutes (or more) using a 104 LED lamp (M340L4; Thorlabs, $\lambda_{max} = 345$ nm, nominal maximum irradiance 2.22 W/m²) 105 placed 2 cm above the reactor. The humidity within the reactor was adjusted by varying the 106 flows of synthetic air or synthetic N₂ passing through a bubbler or connected directly to the 107 reactor.

108 This reactor was maintained at room temperature and flushed with nitrogen or synthetic air 109 at a flow rate of 90 mL min⁻¹. The outlet of the flow cell was connected to all the analytic 110 devices described above.

For the experiments of NO uptake (section 3.4), a constant gas flow of nitric oxide at [NO] = 68 ± 2 ppb was used. To stabilize the NO concentration, the flow first bypassed the reactor and was then injected into the reactor at the beginning of the film exposure.

114

115 2.2. Analytical devices and protocols

116

2.2.1. UV analysis

The UV-vis absorption of the organic films in the range 200–600 nm by means of Cary 60 UV-117 118 Vis spectrophotometer (Agilent Technologies) in the liquid (after extraction) and solid (adsorbed) phases. The solid-liquid extraction was performed by rinsing the films with 2 mL of 119 Milli-Q water. This solution which was then diluted between 2 and 20 times prior to be optically 120 probed in 3500 µL Macro Fluorescence Cuvette (Thorlabs). The absorption of the solid films 121 was simply probed by placing the circular borosilicate glass window on which the organic film 122 123 was adsorbed perpendicularly to the optical beam within the spectrometer. Each analysis was performed on three different locations of the film and then averaged, and the absorbance from 124 125 the glass window subtracted.

126 2.2.2. PTR-TOF-MS analysis

127 In order to monitor the emissions of volatile organic compounds (VOC) from the furfural films, 128 a proton transfer reaction-time of flight mass spectrometer (PTR-TOF-MS, 8000 - Ionicon 129 Analytik GmbH, Innsbruck, Austria) was deployed, as described in detail by Ponczek and 130 George²⁴. This soft ionization procedure limits, while not preventing as discussed by Ŝpaněl et 131 $al.^{25}$, the fragmentation of the VOC to be analyzed and enables to the detection of a large number 132 of compounds.

All PTR-TOF-MS measurements were performed using the H_3O^+ ionization mode with the following parameters: source current of 4.0 mA, drift tube voltage of 600 V, drift tube temperature of 373 K, inlet temperature at 353 K and a drift pressure of 2.20 mbar. The E/N was about 123 Td. Spectra were collected at a time resolution of 10 s. Mass calibrations were performed using internal calibration based on two ions with known mass: hydronium ion isotope ($H_3^{18}O^+$ m/z 21.022) and protonated acetone (C₃H₇O+ m/z 59.049).

139

2.2.3. High-resolution NMR analysis

140 Characterization of the furfural films by ¹H NMR were carried out using a Bruker AVANCE 141 HD (400 MHz) spectrometer equipped with a 5mm BBFO Z-gradient probe. The films were 142 dissolved in 1 mL DMSO-d6 and 600 μ L of this solution was used for the NMR tube 143 preparation. Quantitative ¹H NMR was carried out at a frequency of 400.13 MHz using a 30° 144 flip angle (3.1 μ s) with a recycle delay of 1 s and an acquisition time of 5 s. Sixteen scans of 145 single pulse ¹H excitation were collected over a spectral width of 20 ppm. Data were acquired 146 and processed with Bruker topspin 3.2 software.

147

2.2.4. HONO and NO_x gaseous analytics

A Long Path Absorption Photometer (LOPAP, QUMA) was used to measure the concentrations of gaseous HONO²³. In brief, HONO was sampled in a stripping coil by a fast-chemical reaction and converted into an azodye, which was photometrically detected in a long path absorption cell. The inlet gas flow was set to 500 mL min⁻¹ and the integrated time resolution was set at 7 min with a detection limit less than 10 ppt. The liquid flow used for the reactants was set to 20 ml min⁻¹ and the sampling unit was maintained at 293 ± 0.02 K using a (Thermo Scientific ARCTIC Refrigerated/Heated Bath SC100 A10 - 230V), which was calibrated each week.

A standard chemiluminescence NO_X analyzer (CLD 88 CY analyzer, Eco Physics) was used to monitor NO and NO₂ concentrations, with a detection limit of 50 ppt. In the first channel, NO reacts with O₃ to produce NO₂*, which emits a radiation ($\lambda_{max} = 1200$ nm) detected by a photomultiplier. A second canal allows NO₂ to be converted to NO using a reductor agent with a yield of 70%. The NO₂ concentration was obtained by calculating the difference measured between the two channels.

Furfural (99%), potassium nitrate (\geq 99.0%) and dimethyl sulfoxide-d₆ (99.9%) were provided by Sigma Aldrich; NO (1 ppm in N₂) was from Linde, synthetic air (>99.9%) was from Air Products and N₂ was synthetic (>99.9992%). The silica gel desiccant used was provided by Alfa Aesar (indicating, ACS, -4 + 10 Meh Granules) and Drying Pearls Orange was from Sigma Aldrich. The glass windows were made of borosilicate except one made of quartz and were used for the UV analyses of the films.

168

169 **3. Results and discussion**

170 *3.1. Film characterization*

During the film preparation, more than 99% of the mass of furfural was lost by evaporation of the product. For instance, after depositing 96.1 mg of furfural from 2 mL of a solution at 0.5 M, only 0.7 ± 0.4 mg remained after the drying procedure. In other words, all the experiments were performed on very thin films made only from furfural residues. To characterize these films, UV and ¹H NMR analyses were carried out.

To understand the influence of the preparation process on the structure of furfural, UV 176 analyses were performed on deposed and extracted films in 2 mL of Milli-O water and 177 compared to two control solutions of pure furfural solutions at different concentrations (0.01 178 179 and 0.02 mM) (Figure 1). The control solutions of furfural presents UV absorption bands at 277 nm and 228 nm, characteristic of the aldehydic and furan functional groups, respectively. We 180 observed that the UV spectra of films in liquid and solid phases indicated a transformation of 181 the furfural on the surface with light absorption in the visible range (<600 nm) and presented a 182 new band at 340 nm. The intense absorption between 200 and 250 nm indicates the conservation 183 of the furan function while the aldehydic band is shifted, indicating a transformation, which 184 was later confirmed by ¹H NMR and PTR-TOF-MS analyses. We hypothesized that the 185 structure of the film could correspond to polymeric structures, as seen in the literature^{44,45}. 186 Furfural can indeed undergo self-condensation under anhydrous conditions or thermal 187 treatment at 373–523 K^{18,44} and could also form humins^{46,47}; however, the mechanism is still 188 unclear^{46–48}. Such chemical processes result in the formation of polyfurylic structures (Figure 189 2). The band at 340 nm could correspond to maleic anhydride and/or 2-acetylfuran; both of 190 these are products emitted in the gas-phase from furfural films, as discussed in section 3.2. 191

The UV analysis of a furfural film that was irradiated by the LED lamp at 340 nm over time (see Figure S1) showed a change in the shape of the band centered at 340 nm, along the irradiation time, implying a possible degradation of the sample and a chemical transformation of the surface. This change in composition upon irradiation will be important for describing the temporal behavior of HONO emissions discussed in section 3.3.



Figure 1: The UV spectra of a furfural-extracted film, a solid film and aqueous solutions ofpure furfural (0.01 mM, 0.02 mM). The films were not irradiated.

200



201

Figure 2: The chemical structure of the suggested polyfurylic oligomer (left)⁴⁴ and humins
 (right)⁴⁵

204

Moreover, the ¹H NMR analyses showed broad peaks, confirming the complex nature of furfural films⁴⁹. The proton peak from the aldehyde (9.4 ppm) group in furfural was not found, while the other proton signals could not be identified due to the complexity of the spectra

between 6 to 8 ppm (see Figure S2a). An important peak at 6.1 ppm was associated with the 208 209 presence of an enone group in the structure of the oligomer/humin. The low intensity of the signal at 9.5 ppm is also in agreement with the hypothesized structure of the oligomer without 210 211 aldehydic or carboxylic groups. When potassium nitrate is added to the film, the spectra changed and new peaks appeared. A peak at 4.9 ppm, absent from samples without potassium 212 nitrate, indicated the presence of nitro groups (see Figure S2b). The presence of these groups 213 214 involved a reaction between nitrate anions and furfural during the film preparation process. Tests performed on the influence of humidity and irradiation of the film (45 min, with and 215 without nitrate) did not result in a significant modification of the ¹H NMR spectra (see SI Figure 216 217 2a).

218

3.2. VOC production

The PTR-TOF-MS analyses were performed on furfural films at two different relative 219 humidities (<1 and 90%) under dark and irradiation conditions without nitrate. When the films 220 221 were exposed in the dark to dry synthetic air (RH < 1%), no significant VOC formation was observed, however when the humidity was high (RH = $90 \pm 1\%$), numerous VOCs were 222 observed compared to a blank experiment without furfural (see Figure 3 and Table S1). The 223 224 presence of the 2,5-Furandicarbaldehyde (also known as 2,5-Diformylfuran; mass-to-charge 225 ratio of m/z = 125.03), 2-Acetylfuran or 5-Methylfurfural (m/z = 111.04) and more complex structures involving derivatives of furan (m/z = 123.05 and 139.04) confirms the transformation 226 227 of furfural during the film preparation, as suggested from the hypothesized oligomeric structures. Some compounds were unsaturated and possibly conjugated including 2,5-228 Diformylfuran, Furylacrolein, 1-(2-Furyl)-1,2-propanedione, explaining the increase of the 229 maximum of UV absorbance toward visible light, as seen in section 3.1. The presence of 230 enones, as suggested by the ¹H NMR analyses, was confirmed by the presence of 231 232 malondialdehyde, furylacrolein and arolein. The compound emitted from the surface with the

highest relative abundance had a mass-to-charge ratio of m/z = 85.02. This ion could be 233 attributed to 2(5H)-furanone, which forms from the furfural oxidation under harsher 234 conditions⁵⁰. It could also correspond to the isomer 2-butenedial; however, its formation during 235 the film preparation is unclear. Under UV irradiation, maleic anhydride (m/z = 99.01) was 236 formed at high relative abundance (see Figure S3). Photochemistry of both 2(5H)-furanone and 237 2-butenedial could explain the formation of maleic anhydride⁵⁰⁻⁵². However, only the 238 239 photochemistry of 2-butenedial is activated at the wavelengths used in this study. Experiments carried with varying concentrations of oxygen in the gas phase have shown that maleic 240 anhydride formation requires the presence of O₂ as described in the mechanism by Newland et 241 al.⁵¹ (see Figure S4). Nineteen compounds were found to be released at higher concentrations 242 when the sample was irradiated (see Table S2). Direct emissions through the scission of small 243 molecules from the suggested structures described in the literature could be the origin of these 244 VOCs⁵³. 245

The presence of an organic compound (furfural in this study) modifies the molecular structure of adsorbed nitrate anions, leading to a red-shift in the actinic region and increasing the light absorbed by nitrate anions⁵⁴. Due to the presence of high levels of acids on the surface, nitrate is transformed into nitric acid. Adsorbed nitrate/nitric acid ($HNO_{3(s)}$) photolysis can directly produce HONO and NO₂ following reactions R1 and R2. It can produce HONO indirectly through reactions R3 and R4^{54,55}, with Org_(s) corresponding to furfural in this study, as follows:

$$HNO_{3(s)} \xrightarrow{hv} NO_{2(s)} + OH_{(s)}$$
(R1)

$$HNO_{3(s)} \xrightarrow{h\nu} HONO_{(s)} + O({}^{3}P)_{(s)}$$
(R2)

$$NO_{2(s)} + H_2O_{(s)} \xrightarrow{h\nu} OH_{(s)} + HONO_{(s)}$$
(R3)

$$NO_{2(s)} + Org_{(s)} \xrightarrow{hv} Org - oxidized_{(s)} + HONO_{(s)}$$
(R4)



Figure 3: Relative abundance of the most emitted VOCs from furfural films under dark condition and at high relative humidity $(90 \pm 1\%)$

258 The addition of nitrate to the film changed the emission profile of the surface. Concentrations 259 of the products emitted from the film with and without irradiation were enhanced up to a factor of 10 for maleic anhydride and furfural. Concentrations of compounds of higher mass than the 260 furfural were also greatly enhanced; these compounds included succinic/fumaric acid, 2,5-261 Diformylfuran and 1-(2-Furyl)-1,2-propanedione. Unidentified compounds (m/z = 151.05 and 262 169.05) C₈H₆O₃ and C₈H₈O₄, respectively, were also found at higher concentrations. The 263 formation of a new compound was also observed at m/z = 124.01 and was identified as C₆H₃O₃. 264 Experiments with different concentrations of nitrate showed an evolution of VOC 265 concentrations in relation to the quantity of nitrate on the surface. This increase showed that 266

nitrate interacts with the film and decreases the loss of furfural by evaporation during the samplepreparation.

269

270 3.3. Addition of nitrate anions – HONO formation

Nitrate photochemistry is unlikely to occur under indoor conditions due to the lack of
wavelengths below 330nm. However, a photosensitization pathway could certainly promote its
photo conversion to HONO. In this study we explored this possibility on films made of
deposited furfural

When potassium nitrate was added to the furfural films, an enhanced production of HONO was observed. It is important to underline that under our conditions, in the absence of furfural, irradiated potassium nitrate on the surface was unable to produce HONO. In fact, the range of wavelengths (335-380 nm) emitted from the LED lamp used in this experiment does not allow the activation of the photochemistry of potassium nitrate. KNO₃ does not absorb significantly after 335nm^{56,57}.

The amount of HONO produced was dependent on the quantity of furfural deposited, as 281 measured by its initial concentration in the parent solution. An increase in the concentration of 282 furfural from 0.05 to 0.5 M with a concentration of 10⁻¹ M of KNO₃ increased the formation of 283 HONO by a factor of 10 (Figure 4a). A similar trend can be measured for the absorbance at 284 345nm of extracted films in water. However, the deposited nitrate mass had only a slight effect 285 on the formation of HONO from 10⁻⁴ to 10⁻¹ M, which could be expected from a stoichiometric 286 reaction between the furfural oligomer, the nitrate and a stabilization of the furfural present on 287 the surface by the nitrate, as the PTR analyses have shown (Figure 4b). Also, when nitrate is 288 highly concentrated (1 M in the parent solution), the formation of HONO decreases; this is 289 possibly due to the fact that gaseous products formed in the photolysis process may hardly 290

escape from the nitrate crystal layer formed on the top of the film. This phenomenon was
previously observed by Yang et al ⁵⁸. The UV analyses of this film showed a broad absorption
band from 300 to 600 nm, which can prevent the UV irradiation from reaching the furfural film
below the nitrate layer.

295



Figure 4: (a) Formation of nitrous acid from an irradiated film of furfural + KNO_3 (0.1 M) and the absorbance of the film at 345 nm, as a function of the used furfural concentration. (b) Formation of nitrous acid from an irradiated film of furfural (0.1 M) + KNO_3 as a function of

300

301

303

To explain the photochemical formation of NO, NO₂ and HONO, we propose a mechanism 304 based on our observations and the literature⁵⁹⁻⁶² (Figure 5). 2-nitrofuran is used as a proxy of 305 the nitro compounds formed in the film and observed during the NMR analyses. The 306 degradation of 2-nitrofuran leads to two possible products: $[M+H]^+ = 84.02$ and $[M+H]^+ =$ 307 69.02 corresponds to $C_4H_3O_2$ and $C_4H_4O_2$, respectively. Furan is already a product observed 308 309 when irradiating a film made from furfural only; however, when nitrate is added, the photochemistry is enhanced according to the signals detected during the PTR-TOF-MS 310 analyses. Those products could result from the loss of NO and NO₂ as shown in Figure 5. In 311 312 addition, NO₂ could be further reduced to HONO on the surface containing reducing functionalities such as soot⁶³ or humic acids^{64,65}. 313

potassium nitrate concentration. The average HONO concentration was calculated by

integrating the signal during the first 45 minutes of irradiation. Errors are $\pm 2\sigma$.



316

Figure 5: Suggested mechanism of formation of NO, NO₂ and HONO from 2-Nitrofuran under
UV irradiation.

319

This photochemistry was observed under pure nitrogen. The addition of O_2 while maintaining a constant flow in the system resulted in decreases of HONO, NO and NO₂ (60%) formation rates (Figure 6). The O_2^- formation at the surface could explain the formation of nitrate, leading to losses of nitrogen-containing compounds from the gas phase. O_2^- production can result from the formation of peroxy radicals from the surface which were probed in this study (see next section)^{66–68}. As O_2^- is formed, the surface becomes more oxidant, potentially leading to an evolution in the mixing ratios of nitrogen containing compounds at the gas-solid interface.



Figure 6: The formation of HONO, NO and NO₂ from a film of furfural + KNO₃ under UV irradiation (yellow section) at RH = $81 \pm 2\%$. The vertical dashed line separates the two conditions: N₂ and N₂ +O₂.

332

333 3.4. NO uptake experiments

A typical result for a NO photo-enhanced uptake experiments on irradiated furfural films is presented in Figure 7. The NO was initially chosen as a scavenger of the surface-produced peroxy radicals, similar to the study of Hayeck et al⁴⁶. These radicals can react with NO to produce NO₂ following reaction (R7):

338

$$Sample + hv \to R^{\cdot} \tag{R5}$$

$$R^{\cdot} + O_2 \to RO - O^{\cdot} \tag{R6}$$

$$RO - O' + NO \to RO' + NO_2 \tag{R7}$$

Tests performed in the absence of O₂ and UV light did not result in any NO uptake due 340 to the absence of peroxy radicals. Under high relative humidity ($RH = 81 \pm 2\%$), furfural films 341 were unable to uptake NO. This observation could be linked to the presence of a thin layer of 342 water on the surface of the film, which formed under high relative humidity^{39,70}. NO is insoluble 343 in water⁷¹ so no interaction between the surface and NO was observed. Peroxy radicals could 344 also be trapped in the liquid phase, preventing their release in the gas phase to react with NO. 345 However, under indoor-relevant relative humidity (RH = $45 \pm 2\%$), NO is mostly converted 346 into HONO with a yield lower than 30%. It should be noted that the NO_y-mass balance is not 347 closed. That can be explained by the surface formation of nitrogen-containing products. 348

The radical formation from the surfaces could occur due to the presence of furanone / butenedial under irradiation at $\lambda \ge 235$ nm, 2(5H)-furanone forms an aldehyde-ketene through a ringopening mechanism⁵⁰. Butenedial is able to form radicals under irradiation⁵¹. Radicals could also come from the formation of H-abstracted furan and furfural, which were detected in the PTR-TOF-MS analyses.



Figure 7: NO uptake and HONO formation from a furfural film (prepared from 2 mL of 0.5 M of furfural solution) at RH = $45 \pm 2\%$. Yellow section corresponds to UV irradiation

358

359 **3.5.** Influence of relative humidity

360 The relative humidity effect on the gas-phase formation of HONO from mixed furfural and 361 nitrate anions films is shown in Figure 8. The maximum relative humidity was below the 362 deliquescence of potassium nitrate $(>92\%)^{73}$.

363 During the transition from low to high relative humidity, the HONO formation increased due

to nitrate photolysis ⁵⁵. However, pure potassium nitrate films were unable to produce nitrate.

- 365 This result could imply that there was a photosensitizer effect from the organic film, which
- 366 enabled nitrate photochemistry at the surface.



Figure 8: The concentration of HONO formed as a function of relative humidity (%) from a
film prepared with 0.1 M KNO₃ and 0.1 M Furfural. Irradiation time = 45 min.

370

371 **3.6.** HONO source indoor

To estimate the concentration of HONO emitted from a surface in a simulated room, HONO emission fluxes (E_{HONO}) in μ g m⁻² h⁻¹ were calculated according to Potard⁵²:

$$E_{\text{HONO}} = \frac{F_{air} \times ([HONO]_{\text{films}} - [HONO]_{\text{blank}}) \times M_{\text{HONO}}}{V_{\text{mol}}^{\text{air}} \times S_{\text{exposed}}}$$
(Eq. 1)

374

where F_{air} is the airflow ($F_{air} = 0.0054 \text{ m}^3 \text{ h}^{-1}$), [HONO]_{films} and [HONO]_{blank} are the concentrations of HONO emitted from coated samples and clean glass support (ppb), respectively, M_{HONO} is the molecular mass of HONO (0.047 kg mol⁻¹), V_{mol}^{air} is the air molar volume at standard temperature and pressure (25 L mol⁻¹ at 25°C and 1 atm) and S_{exposed} is the exposed surface of nitrate and furfural to light (0.002 m²).

HONO indoor level depends on the production rate, the surface-area-to-volume ratio ofthe room, the photolysis rate and the ventilation rate. Thus, the indoor concentration of HONO

382 can be determined according to Willem et al. 75 :

$$\frac{d[HONO]}{dt} = \frac{E_{HONO} \times S}{V} - N_h[HONO] - J_{HONO}[HONO]$$
(Eq. 2)

383

where E_{HONO} is the emission rate of HONO in μ g m⁻² h⁻¹, S is the surface area of the room (m²), V is the room volume (m³), N_h is the air exchange rate (h⁻¹) and J_{HONO} is the photolysis rate of indoor HONO. Loss on surfaces was not taken into account here as the reactor used in these experiments present a similar surface-to-volume ratio as the one used in this simulation. Hence, HONO measured is already impacted by the loss on the surfaces.

Assuming there is no influence of outdoor concentration of HONO on the indoor concentration due to relatively low outdoor levels⁷⁶, the steady-state HONO concentration can be simplified as:

$$[HONO] = \frac{E_{HONO} \times S}{V \times (J_{HONO} + N_h)}$$
(Eq. 3)

392

For indoor-relevant relative humidity (RH = 45%), with an air exchange rate of 0.5 h⁻¹ and a photolysis rate of HONO equal to 0.25 h^{-1 31,32} and considering the surface of a room of 50 m³ and 94 m² is covered from a film obtained from a solution of 0.1 M KNO₃ and 0.1 M Furfural, HONO concentrations are 3.9 ppb.

Previous studies were performed on the emission of HONO from the uptake of NO₂ on photocatalytic painted surfaces 28,77,78 and with various household surfaces 23 . Conversion yields up to 90% were observed leading to steady state concentrations of HONO ranging from 1.6 to 6.3 ppb. This indicates that furfural film could lead to an increase in HONO production in presence of NO₂.

402 Typical indoor levels of HONO range from 1 to 10 ppb^{33,79,80}; as such, indoor HONO

levels could be significantly affected by surface chemistry as it was previously suggested^{11,80}.
HONO can react with indoor pollutants to form carcinogens like nitrosamines⁸¹ and produce
hydroxyl radicals by photolysis. While the direct effect of hydroxyl radicals on health is still
uncertain^{82,83}, its role in the formation of toxic compounds like ozone or aerosols indicates a
potential negative effect on indoor air quality.

408 **4.** Conclusions

Furfural has been identified as an indoor pollutant^{9,15}. Here, we show that this molecule can 409 react on surfaces, producing complex adsorbed products that initiate photochemistry indoors. 410 It especially leads to the formation of VOCs and the humidity-dependent formation of nitrous 411 acid in the presence of co-adsorbed nitrate anions. O₂ can also form peroxy radicals leading to 412 organic nitrate formation on the surface. As wood-based furniture at home is ubiquitous, 413 surfaces coated with furfural oligomer are to be expected. Extrapolation of our results to an 414 indoor environment lead to an estimation of 3.9 ppb at indoor-relevant relative humidity of 415 416 45%. HONO reacts with indoor pollutants to form carcinogens compounds as well as producing OH radicals through photolysis. The presence of indoor OH radicals leads to the formation of 417 418 irritating and toxic compounds, altering the indoor air quality.

419

420 Acknowledgments

- 421 This work was supported by the Alfred P. Sloan Foundation under its Chemistry of Indoor422 Environments program.
- 423 Data Availability Statement
- 424 The data that support the findings of this study are available from the corresponding author425 upon reasonable request.

426

427 **References**

- Morrison G. Recent Advances in Indoor Chemistry. *Curr Sustainable Renewable Energy Rep.* 2015;2:33–40.
- Gao P, Korley F, Martin J, Chen BT. Determination of Unique Microbial Volatile
 Organic Compounds Produced by Five Aspergillus Species Commonly Found in
 Problem Buildings. *AIHA Journal*. 2002;63:135–140.
- 433 3. Uhde E, Salthammer T. Impact of reaction products from building materials and
 434 furnishings on indoor air quality A review of recent advances in indoor chemistry. In:
 435 ; 2007.
- 436 4. Brown SK, Sim MR, Abramson MJ, Gray CN. Concentrations of Volatile Organic
 437 Compounds in Indoor Air A Review. *Indoor Air*. 1994;4:123–134.
- 438 5. Edwards RD. VOC source identification from personal and residential indoor, outdoor
 439 and workplace microenvironment samples in EXPOLIS-Helsinki, Finland. *Atmospheric*440 *Environment*. 2001;35:4829–4841.
- 441 6. Carslaw N. A new detailed chemical model for indoor air pollution. *Atmospheric*442 *Environment*. 2007;41:1164–1179.
- Waring MS, Wells JR. Volatile organic compound conversion by ozone, hydroxyl
 radicals, and nitrate radicals in residential indoor air: Magnitudes and impacts of oxidant
 sources. *Atmospheric Environment*. 2015;106:382–391.
- 8. Yrieix C, Maupetit F, Ramalho O. Determination of VOC emissions from French wood
 products. In: *4th European Wood-Based Panel Symposium*. Hanover, Germany; 2004:10
 p. September 2004. https://hal.archives-ouvertes.fr/hal-00688527. Accessed July 15,
 2019.
- 450 9. Cincinelli A, Martellini T, Amore A, et al. Measurement of volatile organic compounds
 451 (VOCs) in libraries and archives in Florence (Italy). *Science of The Total Environment*.
 452 2016;572:333–339.
- Mathew AK, Abraham A, Mallapureddy KK, Sukumaran RK. Lignocellulosic
 Biorefinery Wastes, or Resources? In: *Waste Biorefinery*. Elsevier; 2018:267–297. 2018.
 https://linkinghub.elsevier.com/retrieve/pii/B9780444639929000094. Accessed January
 8, 2020.
- 457 11. Wang C, Collins DB, Arata C, et al. Surface reservoirs dominate dynamic gas-surface
 458 partitioning of many indoor air constituents. *Science Advances*. 2020;6:eaay8973.
- 459 12. Shin S-H, Jo W-K. Longitudinal variations in indoor VOC concentrations after moving
 460 into new apartments and indoor source characterization. *Environ Sci Pollut Res.*461 2013;20:3696–3707.
- 462 13. Salonen HJ, Pasanen A-L, Lappalainen SK, et al. Airborne Concentrations of Volatile
 463 Organic Compounds, Formaldehyde and Ammonia in Finnish Office Buildings with
 464 Suspected Indoor Air Problems. *Journal of Occupational and Environmental Hygiene*.
 465 2009;6:200–209.

- 466 14. Martellini T, Berlangieri C, Dei L, et al. Indoor levels of volatile organic compounds at
 467 Florentine museum environments in Italy. *Indoor Air*. 2020;n/a. February 23, 2020.
 468 https://onlinelibrary.wiley.com/doi/abs/10.1111/ina.12659. Accessed July 3, 2020.
- 15. Nilsson A, Lagesson V, Bornehag C-G, Sundell J, Tagesson C. Quantitative
 determination of volatile organic compounds in indoor dust using gas chromatographyUV spectrometry. *Environment International*. 2005;31:1141–1148.
- 472 16. Zhang J, Chen W, Li J, Yu S, Zhao W. VOCs and Particulate Pollution due to Incense
 473 Burning in Temples, China. *Procedia Engineering*. 2015;121:992–1000.
- Hoydonckx HE, Rhijn WMV, Rhijn WV, Vos DED, Jacobs PA. Furfural and
 Derivatives. In: *Ullmann's Encyclopedia of Industrial Chemistry*. American Cancer
 Society; 2007. 2007.
 https://onlinelibrary.wiley.com/doi/abs/10.1002/14356007.a12_119.pub2. Accessed
- 477 https://oninenorary.wney.com/doi/abs/10.1002/14350007.a12_115.pdb2. Accessed 478 January 6, 2020.
- 479 18. Mariscal R, Maireles-Torres P, Ojeda M, Sádaba I, Granados ML. Furfural: a renewable
 480 and versatile platform molecule for the synthesis of chemicals and fuels. *Energy Environ*481 Sci. 2016;9:1144–1189.
- 19. Nagda NL. *Design and Protocol for Monitoring Indoor Air Quality*. ASTM
 International; 1989.
- Liu Q-T, Chen R, McCarry BE, Diamond ML, Bahavar B. Characterization of Polar
 Organic Compounds in the Organic Film on Indoor and Outdoor Glass Windows. *Environ Sci Technol.* 2003;37:2340–2349.
- 487 21. Weschler CJ, Nazaroff WW. Growth of organic films on indoor surfaces. *Indoor Air*.
 488 2017;27:1101–1112.
- 489 22. Kruza M, Lewis AC, Morrison GC, Carslaw N. Impact of surface ozone interactions on indoor air chemistry: A modeling study. *Indoor Air*. 2017;27:1001–1011.
- 491 23. Gómez Alvarez E, Sörgel M, Gligorovski S, et al. Light-induced nitrous acid (HONO)
 492 production from NO2 heterogeneous reactions on household chemicals. *Atmospheric* 493 *Environment*. 2014;95:391–399.
- 494 24. Sleiman M, Gundel LA, Pankow JF, Jacob P, Singer BC, Destaillats H. Formation of
 495 carcinogens indoors by surface-mediated reactions of nicotine with nitrous acid, leading
 496 to potential thirdhand smoke hazards. *Proceedings of the National Academy of Sciences*.
 497 2010;107:6576–6581.
- 498 25. Ye C, Zhang N, Gao H, Zhou X. Matrix effect on surface-catalyzed photolysis of nitric
 499 acid. *Scientific Reports*. 2019;9. December 2019.
 500 http://www.nature.com/articles/s41598-018-37973-x. Accessed July 20, 2020.
- Salthammer T, Bednarek M, Fuhrmann F, Funaki R, Tanabe S-I. Formation of organic
 indoor air pollutants by UV-curing chemistry. *Journal of Photochemistry and Photobiology A: Chemistry*. 2002;152:1–9.

- Schwartz-Narbonne H, Jones SH, Donaldson DJ. Indoor Lighting Releases Gas Phase
 Nitrogen Oxides from Indoor Painted Surfaces. *Environ Sci Technol Lett.* January 2019.
 January 15, 2019. https://doi.org/10.1021/acs.estlett.8b00685. Accessed January 22,
 2019.
- Bartolomei V, Sörgel M, Gligorovski S, et al. Formation of indoor nitrous acid (HONO)
 by light-induced NO2 heterogeneous reactions with white wall paint. *Environ Sci Pollut Res.* 2014;21:9259–9269.
- 511 29. Gandolfo A, Marque S, Temime-Roussel B, et al. Unexpectedly High Levels of Organic
 512 Compounds Released by Indoor Photocatalytic Paints. *Environ Sci Technol*.
 513 2018;52:11328–11337.
- So. Gandolfo A, Gligorovski V, Bartolomei V, et al. Spectrally resolved actinic flux and photolysis frequencies of key species within an indoor environment. *Building and Environment*. 2016;109:50–57.
- 517 31. Alvarez EG, Amedro D, Afif C, et al. Unexpectedly high indoor hydroxyl radical
 518 concentrations associated with nitrous acid. *PNAS*. 2013;110:13294–13299.
- 519 32. Zhou S, Young CJ, VandenBoer TC, Kowal SF, Kahan TF. Time-Resolved
 520 Measurements of Nitric Oxide, Nitrogen Dioxide, and Nitrous Acid in an Occupied New
 521 York Home. *Environ Sci Technol.* 2018;52:8355–8364.
- 522 33. Gligorovski S. Nitrous acid (HONO): An emerging indoor pollutant. *Journal of*523 *Photochemistry and Photobiology A: Chemistry*. 2016;314:1–5.
- 524 34. Gómez Alvarez E, Wortham H, Strekowski R, Zetzsch C, Gligorovski S. Atmospheric
 525 Photosensitized Heterogeneous and Multiphase Reactions: From Outdoors to Indoors.
 526 *Environ Sci Technol.* 2012;46:1955–1963.
- 527 35. Gligorovski S, Weschler CJ. The Oxidative Capacity of Indoor Atmospheres. *Environ* 528 *Sci Technol.* 2013;47:13905–13906.
- 36. Mendez M, Amedro D, Blond N, et al. Identification of the major HOx radical pathways
 in an indoor air environment. *Indoor Air*. 2017;27:434–442.
- 37. Weschler CJ, Shields HC, Naik DV. Indoor Chemistry Involving O3, NO, and NO2 as
 Evidenced by 14 Months of Measurements at a Site in Southern California. *Environmental Science & Technology*. 1994;28:2120–2132.
- 38. Arata C, Zarzana KJ, Misztal PK, et al. Measurement of NO3 and N2O5 in a Residential
 Kitchen. *Environ Sci Technol Lett.* 2018;5:595–599.
- Schwartz-Narbonne H, Donaldson DJ. Water uptake by indoor surface films. *Scientific Reports*. 2019;9:1–10.
- 40. Baergen AM, Donaldson DJ. Seasonality of the Water-Soluble Inorganic Ion
 Composition and Water Uptake Behavior of Urban Grime. *Environ Sci Technol*.
 2019;53:5671–5677.

- 41. Ponczek M, George C. Kinetics and Product Formation during the Photooxidation of
 Butanol on Atmospheric Mineral Dust. *Environ Sci Technol.* 2018;52:5191–5198.
- 42. Ŝpaněl P, Ji Y, Smith D. SIFT studies of the reactions of H3O+, NO+ and O2+ with a series of aldehydes and ketones ScienceDirect. 1997. https://www-sciencedirectcom.docelec.univ-lyon1.fr/science/article/pii/S0168117697001663?via%3Dihub.
 Accessed November 27, 2019.
- 43. Heland J, Kleffmann J, Kurtenbach R, Wiesen P. A New Instrument To Measure
 Gaseous Nitrous Acid (HONO) in the Atmosphere. *Environ Sci Technol*. 2001;35:3207–
 3212.
- 44. Gandini A, Belgacem MN. Furans in polymer chemistry. *Progress in Polymer Science*.
 1997;22:1203–1379.
- 45. van Zandvoort I, Wang Y, Rasrendra CB, et al. Formation, Molecular Structure, and
 Morphology of Humins in Biomass Conversion: Influence of Feedstock and Processing
 Conditions. *ChemSusChem.* 2013;6:1745–1758.
- Shi N, Liu Q, Ju R, et al. Condensation of α-Carbonyl Aldehydes Leads to the Formation of Solid Humins during the Hydrothermal Degradation of Carbohydrates. *ACS Omega*.
 2019;4:7330–7343.
- 47. Sumerskii IV, Krutov SM, Zarubin MY. Humin-like substances formed under the
 conditions of industrial hydrolysis of wood. *Russ J Appl Chem.* 2010;83:320–327.
- Filiciotto L, Miguel G de, Balu AM, Romero AA, Waal JC van der, Luque R. Towards
 the photophysical studies of humin by-products. *Chem Commun.* 2017;53:7015–7017.
- 562 49. Bovey FA, Mirau PA. NMR of Polymers. Academic Press; 1996.
- 563 50. Li X, Lan X, Wang T. Highly selective catalytic conversion of furfural to γ butyrolactone. *Green Chem.* 2016;18:638–642.
- 565 51. Newland MJ, Rea GJ, Thüner LP, et al. Photochemistry of 2-butenedial and 4-oxo-2566 pentenal under atmospheric boundary layer conditions. *Phys Chem Chem Phys*.
 567 2019;21:1160–1171.
- 568 52. Breda S, Reva I, Fausto R. UV-induced unimolecular photochemistry of 2(5H)-furanone
 and 2(5H)-thiophenone isolated in low temperature inert matrices. *Vibrational* 570 *Spectroscopy*. 2009;50:57–67.
- 571 53. Smets G. Organic Polymer Photochemistry. *Polymer Journal*. 1985;17:153–165.
- 572 54. Ye C, Zhang N, Gao H, Zhou X. Matrix effect on surface-catalyzed photolysis of nitric
 573 acid. *Scientific Reports*. 2019;9:1–10.
- 574 55. Zhou X, Gao H, He Y, et al. Nitric acid photolysis on surfaces in low-NOx
 575 environments: Significant atmospheric implications. *Geophysical Research Letters*.
 576 2003;30. December 1, 2003. https://agupubs-onlinelibrary-wiley-com.docelec.univ577 lyon1.fr/doi/abs/10.1029/2003GL018620. Accessed December 5, 2018.

- 56. Bravo M, Olwiert AC, Oelckers B. Nitrate determination in Chilean Caliche samples by
 UV-Visble absorbance measurements and multivariate calibraron. *Journal of the Chilean Chemical Society*. 2009;54:93–98.
- 581 57. Sager EE, Byers FC. Spectral absorbance of some aqueous solutions in the range 10582 degrees to 40-degrees-C. *Journal of Research of the National Bureau of Standards*.
 583 1957;58:33.
- 584 58. Yang W, Han C, Yang H, Xue X. Significant HONO formation by the photolysis of
 585 nitrates in the presence of humic acids. *Environmental Pollution*. 2018;243:679–686.
- 586 59. Chapman OL, Heckert DC, Reasoner JW, Thackaberry SP. Photochemical Studies on 9 Nitroanthracene1. *J Am Chem Soc*. 1966;88:5550–5554.
- Feilberg A. Atmospheric Chemistry of Polycyclic Aromatic Compounds with Special
 Emphasis on Nitro Derivatives. January 2020.
- 61. Ramazan KA, Syomin D, Finlayson-Pitts BJ. The photochemical production of HONO during the heterogeneous hydrolysis of NO2. *Phys Chem Chem Phys.* 2004;6:3836–3843.
- 593 62. Saliba NA, Mochida M, Finlayson-Pitts BJ. Laboratory studies of sources of HONO in
 594 polluted urban atmospheres. *Geophysical Research Letters*. 2000;27:3229–3232.
- 63. Gutzwiller L, Arens F, Baltensperger U, Gäggeler HW, Ammann M. Significance of
 Semivolatile Diesel Exhaust Organics for Secondary HONO Formation. *Environ Sci Technol.* 2002;36:677–682.
- 598 64. Stemmler K, Ndour M, Elshorbany Y, et al. Light induced conversion of nitrogen
 599 dioxide into nitrous acid on submicron humic acid aerosol. *Atmos Chem Phys.* 2007:12.
- 600 65. Wall KJ, Harris GW. Uptake of nitrogen dioxide (NO2) on acidic aqueous humic acid
 601 (HA) solutions as a missing daytime nitrous acid (HONO) surface source. *J Atmos*602 *Chem.* 2017;74:283–321.
- 603 66. Sonntag C von, Schuchmann H-P. The Elucidation of Peroxyl Radical Reactions in
 604 Aqueous Solution with the Help of Radiation-Chemical Methods. *Angewandte Chemie*605 *International Edition in English.* 1991;30:1229–1253.
- 606 67. Tan Y, Lim YB, Altieri KE, Seitzinger SP, Turpin BJ. Mechanisms leading to oligomers
 607 and SOA through aqueous photooxidation: insights from OH radical oxidation of acetic
 608 acid and methylglyoxal. *Atmospheric Chemistry and Physics*. 2012;12:801–813.
- 609 68. Schuchmann MN, Zegota H, Sonntag C von. Acetate Peroxyl Radicals, ·O2CH2 CO2-:
 610 A Study on the γ-Radiolysis and Pulse Radiolysis of Acetate in Oxygenated Aqueous
 611 Solutions. *Zeitschrift für Naturforschung B*. 1985;40:215–221.
- 612 69. Hayeck N, Rossignol S, Alpert PA, et al. Photochemical reaction at the air-water
 613 interface and effect on atmospheric nitric oxide. In: *European Aerosol Conference, EAC*614 2016. Tours , France; 2016. September 2016. https://hal.archives-ouvertes.fr/hal615 01364534. Accessed March 10, 2020.

- 70. Dubowski Y, Sumner AL, Menke EJ, et al. Interactions of gaseous nitric acid with
 surfaces of environmental interest. *Phys Chem Chem Phys.* 2004;6:3879–3888.
- 71. Zacharia IG, Deen WM. Diffusivity and solubility of nitric oxide in water and saline.
 Ann Biomed Eng. 2005;33:214–222.
- Freda S, Reva I, Fausto R. UV-induced unimolecular photochemistry of 2(5H)-furanone
 and 2(5H)-thiophenone isolated in low temperature inert matrices. *Vibrational Spectroscopy*. 2009;50:57–67.
- Freney EJ, Martin ST, Buseck PR. Deliquescence and Efflorescence of Potassium Salts
 Relevant to Biomass-Burning Aerosol Particles. *Aerosol Science and Technology*.
 2009;43:799–807.
- Potard K. Les émissions de composés organiques volatiles (COVs) des sols dans les
 paysages agricoles: identification des sources et incidences sur la qualité de l'air. :182.

628 75. Willem H, Hult E, Hotchi T, Russell M, Maddalena R, Singer B. Ventilation Control of
629 Volatile Organic Compounds in New U.S. Homes: Results of a Controlled Field Study
630 in Nine Residential Units.

- 631 76. Lee Kiyoung, Xue Jianping, Geyh Alison S, et al. Nitrous acid, nitrogen dioxide, and
 632 ozone concentrations in residential environments. *Environmental Health Perspectives*.
 633 2002;110:145–150.
- 634 77. Gandolfo A, Bartolomei V, Gomez Alvarez E, et al. The effectiveness of indoor
 635 photocatalytic paints on NOx and HONO levels. *Applied Catalysis B: Environmental*.
 636 2015;166–167:84–90.
- 637 78. Gandolfo A, Bartolomei V, Truffier-Boutry D, et al. The impact of photocatalytic paint
 638 porosity on indoor NOx and HONO levels. *Phys Chem Chem Phys.* 2020;22:589–598.
- 639 79. Leaderer BP, Naeher L, Jankun T, et al. Indoor, outdoor, and regional summer and
 640 winter concentrations of PM10, PM2.5, SO4(2)-, H+, NH4+, NO3-, NH3, and nitrous
 641 acid in homes with and without kerosene space heaters. *Environ Health Perspect*.
 642 1999;107:223–231.
- 80. Collins DB, Hems RF, Zhou S, et al. Evidence for Gas–Surface Equilibrium Control of
 Indoor Nitrous Acid. *Environ Sci Technol*. 2018;52:12419–12427.
- 81. Sleiman M, Destaillats H, Smith JD, et al. Secondary organic aerosol formation from
 ozone-initiated reactions with nicotine and secondhand tobacco smoke. *Atmospheric Environment*. 2010;44:4191–4198.
- 648 82. Gligorovski S, Wortham H, Kleffmann J. The hydroxyl radical (OH) in indoor air:
 649 Sources and implications. *Atmospheric Environment*. 2014;99:568–570.
- 83. Weschler CJ, Wells JR, Poppendieck D, Hubbard H, Pearce TA. Workgroup Report:
 Indoor Chemistry and Health. *Environ Health Perspect*. 2006;114:442–446.
- 652