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Silica-clay nanocomposites for the removal of antibiotics in
the water usage cycle
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### 20 Abstract

21 The increasingly frequent detection of resistant organic micropollutants in waters calls for better 22 treatment of these molecules that are recognized to be dangerous for human health and the 23 environment. As an alternative to conventional adsorbent material such as activated carbon, 24 silica-clay nanocomposites were synthesized for the removal of pharmaceuticals in 25 contaminated water. Their efficiency with respect to carbamazepine, ciprofloxacin, danofloxacin, 26 doxycycline and sulfamethoxazole was assessed in model water and real groundwater spiked 27 with the five contaminants. Results showed that the efficacy of contaminant removal depends 28 on the chemical properties of the micropollutants. Among the adsorbents tested, the 29 nanocomposite made of 95% clay and 5% SiO<sub>2</sub> NPs was the most efficient and was easily 30 recovered from solution after treatment compared to pure clay, for example. The composite is

thus a good candidate in terms of operating costs and environmental sustainability for theremoval of organic contaminants.

### 33 Introduction

34 Antibiotics are massively used in human and veterinary therapeutic applications. Their 35 production and use lead to increasing detection of antibiotics in the environment, especially in 36 effluents from wastewater treatment plants, but also in surface water and even groundwater. 37 The presence of these compounds in the water cycle can affect the quality of drinking water. 38 The resistance of some of these compounds to the conventional processes applied in drinking 39 water supply chains has been evidenced in several national surveys. For example, in the United 40 States, a federal survey identified 118 pharmaceutical micropollutants at the outlet of 25 41 drinking water production plants (FURLONG et al., 2017).

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In this context, it is indispensable to strengthen the monitoring and treatment of these pharmaceutical pollutants at different stages of the water cycle. Two main treatment routes for pharmaceutical compounds are during the production of drinking water or during tertiary treatment of wastewater refining: the adsorption route and the chemical or biological oxidation route that were also widely explored for the elimination of other organic contaminants (PAVITHRA *et al.*, 2019).

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50 Activated Carbon (AC) is the most commonly used adsorbent medium for water treatment in 51 industrial scale applications. The performance of AC in retaining a large number of pollutants is 52 undeniable but is nevertheless sometimes insufficient. A wide range of other media are currently 53 available for direct or polishing water treatment, with different characteristics (origin, activity, 54 porosity, size, density, etc.). The best media for water treatment can be selected based on 55 these properties in order to optimize the removal of the targeted pollutant and to be integrated in 56 the existing treatment plant. On one hand, powdered activated carbon (PAC) in suspension can 57 be directly added to the water to be treated, which has the advantage of being flexible as the 58 dose can be adapted to the concentration of the pollutant. However, the main disadvantage of 59 PAC is its high operating cost when used continuously, and the difficulty involved in separating 60 the medium from the sludge or waste by settling or filtration at the disposal step. Granular 61 activated carbon (GAC), on the other hand, uses a fixed-bed filter, usually located in one or two 62 steps of the water treatment plant, for clarification and polishing. Compared with PAC, GAC is

the most economical solution in larger systems or where organic contamination has to be controlled continuously. GAC performance and treatment reliability for the removal of organics are closely linked to the frequency of the carbon regeneration or replacement. In turn, the frequency is highly dependent on the adsorption affinity of the pollutants and on competition with the organic matter present in the water. This is particularly true with GAC used at the clarification step (KNAPPE *et al.*, 2003).

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Considering the economic and technical limitations involved in the use of activated carbon in water treatment, developing an alternative process or material would be of great interest for water treatment suppliers. To be viable and applicable, such new processes should fulfill numerous conditions, including efficient pollutant elimination, easy integration into existing water treatment plants as an additional refining step, easy removal from the treated water after use, environmental sustainability, and low cost (PRASANNAMEDHA AND KUMAR, 2020).

76 Clay-based materials appear to be good alternative candidates in this context as they have a 77 good adsorption capacity and ion exchange properties that can efficiently and selectively retain 78 many micropollutants (BESNAULT et al., 2014; ZHU et al., 2016). Clays can be modified, e.g. by 79 bridging or grafting nanoparticles, to increase the surface exchange by avoiding the stacking of 80 the clay layers, to adjust the interlayer distance to the size of the targeted micropollutants and in 81 turn, improve the performance and adsorption selectivity (LEE AND JANG, 1996; JOLIVET et al., 82 2004; HORNER et al., 2009; BANKOVIĆ et al., 2013; BRINDLEY AND SEMPELS, 2018; GUÉGAN et al., 83 2020).

84

85 In this work, we developed and tested clay-silica nanoparticle hydrid nanostructures that can be 86 an economic alternative to the activated carbon conventionally used in water treatment. The 87 silica to clay ratio was adjusted to optimize the adsorption capacity of the nanocomposites 88 (NCs). The efficiency of the resulting silica-clay NCs for the removal of ciprofloxacine was first 89 tested on a model water. In the second step, real water contaminated by a selection of 90 pharmaceutical micropollutants that are currently among the most recalcitrant to conventional 91 treatments used in industrialised countries was used to assess the performance of the different 92 NCs in realistic conditions.

# 93 Materials and Methods

#### 94 Targeted pharmaceutical molecules

The pharmaceutical substances studied here were chosen based on consumption, occurrence in the water cycle, recalcitrance to conventional treatments, and potential risk for the environment. Based on these criteria, four antibiotics, doxycycline, ciprofloxacin, danofloxacin and sulfamethoxazole, and an anticonvulsant, carbamazepine, were selected (Table 1).

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#### Table 1. Main use and characteristics of the pharmaceuticals studied here

Pharmaceutical class	Antiepileptic	Antibiotics				
Compounds	Carbamazepine	Ciprofloxacine	Danofloxacine	Doxycycline	Sulfamethoxazole	
Chemical formula	$C_{15}H_{12}N_{2}O$	C 17 H 18 FN 3 O3	$C_{19}H_{20}FN_{3}O_{3}$	$C_{22}H_{24}N_{2}O_{8}$	$C_{10}H_{11}N_{3}O_{3}S$	
Chemical structure	O NH2	не стран			H2N C H2N H2N	
Molecular mass (g/mol)	236.27	331.34	357.38	444.43	253.28	
рКа	13.9	5.86 ; 8.24	6.07 ; 8.56	3.09 ; 8 ; 9.2	1.83 ; 5.57	
Solubility (mg/L)	18	30 000	20 000	630	610	
Log Kow	2.45	0.28	-0.3	-0.02	0.89	
Type of use	Human use	Human use	Veterinary use	Human and veterinary use	Human and veterinar use	

101 102

103 The treatability of chemical compounds in water by adsorption on an hydrophobic medium, such 104 as activated carbon, is often classified according to their respective log Kow (measurement of 105 the differential solubility of chemical compounds in two different media), higher log Kow giving 106 higher expected adsorption. Carbamazepine and sulfamethoxazole are thus expected to 107 undergo strong adsorption, while ciprofloxacine is expected to adsorb more weakly, and 108 danofloxacine and doxycycline should be less adsorbed. Performance tests have been carried 109 out at laboratory, pilot and industrial scale for the elimination of carbamazepine (SUDHAKARAN et 110 al., 2013; AKPINAR AND YAZAYDIN, 2017; BIZI, 2019; HE et al., 2020) and sulfamethoxazole 111 (STACKELBERG et al., 2007). Efficient removal of both compounds was obtained by adsorption 112 on PAC (MARGOT et al., 2013; BESNAULT et al., 2014).

AC is also well known as a non-specific but efficient treatment solution for removal of a wide range of hydrophobic compounds in various matrixes (resources, drinking water, wastewater, industrial effluents, etc.). However, the cost of AC is high and finding les costly alternative media is a real challenge for water treatment. That is why several works on different alternative 117 adsorbent materials applied for micro-pollutants treatment have been undertaken. The materials 118 include clay-based media (YU et al., 2008; ZHANG et al., 2010) which have shown some 119 adsorption capacity. The elimination of ciprofloxacine has been studied at laboratory scale on 120 various media including carbon, clay, and zeolithes (WANG et al., 2015; LI et al., 2017; 121 ROSTAMIAN AND BEHNEJAD, 2017; BIZI AND EL BACHRA, 2020). However, these authors did not 122 study the application or the treatment efficiency of clay-based nanocomposites in realistic 123 conditions for the removal of several compounds with different solubility, size, hydrophobicity 124 and charge, such as the ones targeted here. Finally, few laboratory-scale studies on 125 danofloxacine and doxycycline treatments have focused on the difficulty of treating doxycycline 126 by adsorption (ALI AND AHMED, 2017; LIU et al., 2017).

#### 127 Nanocomposite synthesis

128

129 The clay material used in this work was collected from Nador under the trade name Monador 130 (North East Morocco, North Africa). It consists of a relatively pure Na, Ca montmorillonite. The 131 cationic exchange capacity of the raw clay sieved to 20 µm was determined at 141 meg/100g. 132 For this study, the finest fraction of the clay (< 2 µm) was recovered for the synthesis of the 133 silica-clay NCs as follows: 20 g of the sieved clay were dispersed in 500 mL of ultrapure water 134 by magnetic stirring at 500 rpm for 24 h. After 30 minutes, ultrasonic treatment was performed 135 for 10 minutes. The clay suspension was then poured into a 25 cm glass column and left to settle for 20 h. According to Stokes law and considering a clay density of 2.6 g.cm<sup>-3</sup>. 20 h 136 137 correspond to a sedimentation cutoff close to 2 µm at 25 cm in height. The supernatant 138 containing the size fraction below 2 µm was recovered and freeze dried, for further preparation.

The silica nanoparticles used in this work were obtained from "Clariant Corporation" under the trade name Klebosol<sup>®</sup> 30R50. They were characterised by a primary particle size of 80 nm (Bizi, 2012), also confirmed in hydrodynamic size in pure water measured by dynamic light scattering (82.4  $\pm$  0.6 nm).

Silica-clay nanocomposites (NCs) were synthesized as follows. Clay was dispersed in Milli-Q water at a concentration of 20 g/L. The suspension was stirred for 24 h at 500 rpm to favor layer exfoliation. Then, a volume of SiO<sub>2</sub> nanoparticles was added to the suspension to reach 5 or 50 wt% of the clay substrate. At this neutral pH, both clay and silica are negatively charged and are stable in suspension. The mixture was homogenized by magnetic stirring for one minute, and then immediately frozen in liquid nitrogen to maintain the homogenous distribution of silica and clay components. Water was then removed by freeze drying, and the recovered solid material was heated at 500°C for 4 hours to consolidate the architecture of the aggregates. In order to obtain material reference, separate pure silica and clay materials were also heated at 500°C for comparison with NCs referred as SiO<sub>2</sub>-NP-500°C and Clay-500°C.

#### 153 Characterization of the materials

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Scanning electron microscopy (SEM) analysis was carried out using a Zeiss Gemini 500 high resolution field emission scanning electron microscope equipped with an in-lens secondary electron detector. Images of uncoated samples were acquired at low voltages for surface sensitivity and also to avoid beam damage. X-ray diffraction (XRD) analysis was performed using a Panalytical X'Pert Pro diffractometer running at 40kV and 40mA equipped with Co K $\alpha$ radiation ( $\lambda = 1.79$  Å). Samples were deposited on low background silicon plates and scanned in a 2 $\theta$  range of 4-75° with a step size of 0.033° and a time step of 5.1s.

#### 162 Removal efficiency of ciprofloxacin in a model effluent

163 Elimination of ciprofloxacin from a model effluent was studied as a preliminary approach to 164 optimize the synthesis of the sorbent material. Evian water (pH= 7.5-7.8) was used to mimic the 165 ionic composition of a typical surface water. Ciprofloxacin in the aqueous medium was 166 quantified by measuring the specific absorbance peak at 272 nm wavelength. A Jasco V-650 167 UV-visible spectrophotometer (Jasco, Japan) was used for this purpose, with quartz cells of 1 168 cm optical pathway. A calibration curve was first measured showing good linearity from 0.5 to 20 mg/L ( $R^2 = 0.9999$ ). Every absorbance measurement was processed in triplicate. 169 170 Ciprofloxacin was first spiked into water at 5 mg/L. The sorbent material was then added at a 171 concentration of 50 mg/L and the reactor was kept under agitation for three hours by magnetic 172 stirring at 650 rpm. Aliquots of the mixture were sampled at different times and filtered at 0.45 173 um to remove the sorbent material and the retained ciprofloxacin. The supernatant was then 174 analysed by UV absorbance to measure the concentration of ciprofloxacin remaining free in 175 solution. X-ray Photoelectron Spectroscopy (XPS) was performed using a Kratos - Axis NOVA 176 XPS equiped with a monochromatic AI K $\alpha$  source to assess adsorption mechanism (150 Watts).

#### 177 Nanocomposite dispersibility and potential for recovery after treatment

178 After use as a sorbent material during a tertiary or polishing treatment step, the NCs should be 179 easily recovered from the liquid medium in an additional quaternary step before the treated 180 water is released. This can be achieved by the rapid sedimentation of the material in the liquid 181 batch when agitation is stopped, which depends on the size and density of the dispersed units. 182 Dispersion or sedimentation of the synthesized materials in aqueous medium was measured 183 using laser diffraction and turbidity profiles. Firstly, 200 mg/L of NC was introduced in Evian 184 water, that mimicked the ionic composition of a typical surface water, and maintained under 185 constant agitation by magnetic stirring for four hours. A Malvern Mastersizer 3000 (Malvern 186 Instruments, Worcester, UK) connected to the agitated system via a peristaltic pump in 187 recirculation mode was used to measure the size distribution of the different NCs over time. For 188 comparison with reference clay materials, the same procedure was used for pure 189 montmorillonite clay before and after calcination at 500°C. Secondly, at t = 4h, a 20 mL aliquot 190 of the medium was sampled under agitation and poured into the sedimentation tube. A 191 Turbiscan apparatus (Formulaction, Toulouse, Fance) was used to measure the turbidity profile 192 along the tube height according to time. Measurements were recorded every 25 sec during 30 193 minutes. The suspension clarification was analyzed via the intensity of the transmitted light 194 increasing with time, while the formation of sediments was measured via the backscattered light 195 intensity in the bottom of the tube (lower 1 mm). A 2% relative increase was reasonably used to 196 distinguish the top of the solid deposit from the initial background backscattered by the liquid 197 suspension.

#### 198 Removal efficiency in real groundwater

After optimization of the synthesis and tests of NCs efficiency in model water, clay-silica NCs were tested in real groundwater spiked with the five selected micropollutants at realistic concentrations. Activated carbon, currently used in conventional water treatment, was also studied for the purpose of comparison.

Groundwater from the Paris region (France) was used for this experiment (pH = 7.5, dissolved organic carbon: 1.5 mg/L, UV absorbance at 254 nm = 4 m<sup>-1</sup>, expressed as optical density). The main chemical elements constituting the geochemical composition were analyzed. Results are given in mg/L: AI (< 0.05), Ca (72.4), K (3.3), Mg (8.59), Na (15.3), P (< 0.05) and Zn (0.04). In order to control the presence of residual pollutants in this groundwater before exposition, the concentration of the five studied micropollutants was determined. Only carbamazepine and sulfamethoxazole were detected in the groundwater, at respectively 10 and 5 ng/L, giving thechemical background.

211

The five micropollutants ciprofloxacine, danofloxacine, doxycycline, carbamazepine and sulfamethoxazole were spiked into water at individual concentrations of 2 µg/L. The sorbent material was then added to the medium at a concentration of 50 mg/L, as processed in preliminary tests on the model effluent, and the mixture was agitated continuously for 48 h. Filtration at 0.45 µm on a glass fiber filter was then processed to remove the sorbent material and other particulate matter from the liquid medium.

218 An analytical method was developed to quantify the targeted micropollutants in this matrix and 219 also in other complex waters (surface water, wastewater, etc.). Solid phase extraction on HLB 220 cartridge (Waters) at pH 3 was performed. Then, after washing and drying, an elution was 221 realised with 5 mL methanol. Analyses were liquid chromatography coupled with mass 222 spectrometry. A triple quadrupole mass spectrometer, TSQ Quantum (Thermo Fisher Scientific) equipped with a HSS T3 column (Waters®) was used. The mobile phase followed a 223 224 water/methanol gradient + 0.1% formic acid, circulating at a solvent flow rate of 200 µL/min, with 225 an injected volume of 10 µL and a temperature of 30°C. Detection was achieved by two MS/MS 226 transitions in positive ionization mode with electrospray source. Quantification was performed by internal calibration using compounds labelled with D or <sup>13</sup>C. The limits of detection obtained 227 228 ranged between 1 and 5 ng/L.

## 229 **Results**

#### 230 1. Characterization of materials

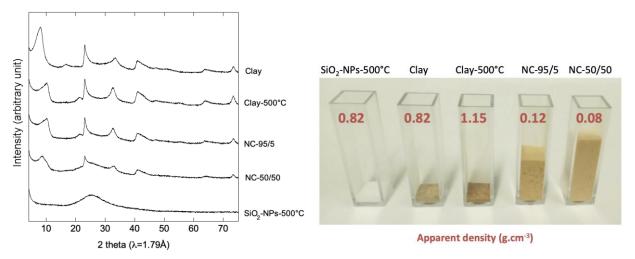
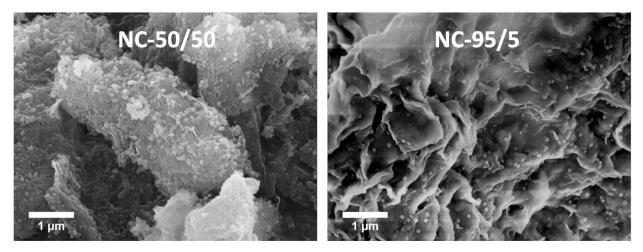




Figure 1. X-ray diffraction patterns (left) and apparent density measured in a volume of 230 mg of material (right) of the initial clay, nanocomposites and references heated at 500°C



235 Since silica is amorphous, the X-ray diffraction pattern of the clay-based materials is consistent 236 with a montmorillonite (Figure 1, left). The first peak corresponds to the (001) reflection (basal 237 spacing) and is characteristic of layer stacking. For raw clay, the interlayer distance is 12.5Å, 238 typical of hydrated Na-montmorillonite. When heated to 500°C, the lamellar structure is 239 preserved and the (001) reflection is shifted to 9.6Å, indicating interlayer dehydration. Regarding 240 NCs, although NC-95/5 was very similar to Clay-500°C, the first NC-50/50 peak became less 241 intense and broader with a slight shift at lower angles, suggesting increased disorder in layer 242 stacking and increased interlayer distance. However, no change in the order inside the layer 243 was observed. The increase in structural desorder is in good agreement with the measured 244 apparent density (Figure 2, right) that is more than one-order lower for NC-50/50 than for Clay-245 500°C (decrease in the apparent density by a fator of 14). In addition, -although not observed by 246 X-ray diffraction, 5% of SiO<sub>2</sub> Nps in NC-95/5 is enough to strongly affect apparent density by 247 one order of magnitude compared to Clay-500°C.



251

Figure 2. SEM images of NC-50/50 (left) and NC-95/5 (right)

252 SEM analysis was performed on NC-95/5 and NC-50/50 (Figure 2). Although SEM cannot 253 provide evidence for the intercalation of SiO<sub>2</sub>-NPs in the interlayers, it can give an idea of the 254 surface coverage and dispersion of SiO<sub>2</sub>-NPs at the surface of the clay sheets. For the lowest 255 loading of SiO<sub>2</sub>-NPs (NC-95/5)(Figure 2, right), NPs are well-dispersed in the clayey matrix and 256 are isolated from each other while NC-50/50 (Figure 2, left) exhibits a relatively high surface 257 coverage of the SiO<sub>2</sub>-NPs at the clay surface. The randomly dispersed distribution of the SiO<sub>2</sub>-258 NPs observed on the clay substrate for NC-95/5 confirms that the method of preparation used 259 here satisfactorily delivered an apparently homogenous hybrid material. Moreover, irregularities, 260 microcavities and the non-compact and very porous nature of the observed structures in 261 addition to the low apparent density (probable high macroporosity) will favor the rapid diffusion 262 of water and micropollutants. Beyond these structural considerations, the reactivity of the NCs 263 was assessed in both model effluents and in realistic conditions.

264

### 265 2. Removal efficiency of ciprofloxacin in model effluents

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267 Ciprofloxacin was used for this first proof of concept thanks to the ease of UV-visible 268 spectroscopy.

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- 270

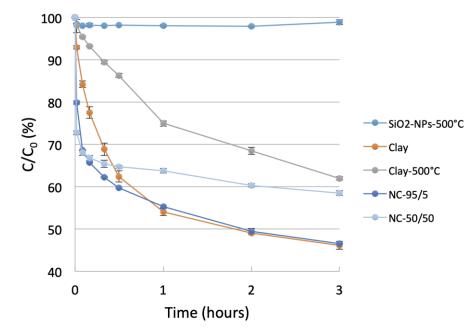




Figure 3. Removal kinetics of ciprofloxacin from a model of surface water by adsorption to the sorbent materials: clay, SiO<sub>2</sub>-NPs-500°C and NCs

275 The time-resolved removal of ciprofloxacin from the model water (Figure 3) provided insights 276 into the adsorption kinetics on clay (raw or heated at 500°C), SiO<sub>2</sub>-NPs-500°C and NCs. While 277 no removal of ciprofloxacin was obtained with SiO<sub>2</sub>NP-500°C, significant but contrasted 278 adsorption behaviors were observed in clay and NC sorbents. The behaviors were 279 characterized by a first removal step lasting around 10 minutes, followed by a slower reaction 280 step that may correspond to saturation of the material. The slope of the curves measured over 281 the 10 first minutes (first 4 points) was used to obtain insight into the initial kinetic of the reaction 282 (Figure 4, blue bars). Three behaviors can be distinguished as (i) a rapid initial adsorption rate 283 with NC-95/5 and NC-50/50, (ii) a moderate initial adsorption rate obtained with clay and (iii) a 284 relatively slow adsorption rate with clay-500°C. Thus, the presence of SiO<sub>2</sub>-NPs in NC-95/5 and 285 NC-50/50 significantly increased the adsorption kinetics of ciprofloxacin.

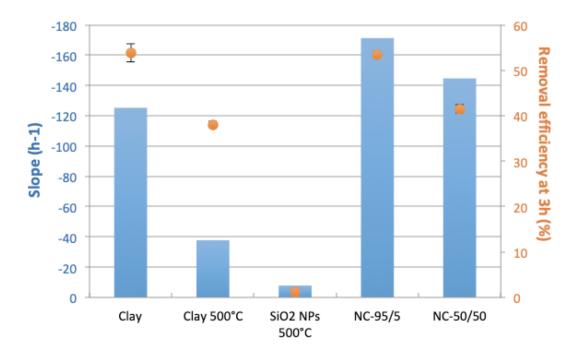


Figure 4. Removal kinetics (Figure 3) calculated on the first 4 experimental points (after 10 minutes of reaction) are represented on the primary y-axis (blue bars). Removal efficiencies after 3 hours are represented on the secondary y-axis (orange dots).

290

291 In addition to the initial kinetics, the removal efficiency measured after 3 hours provides insight 292 into sorbent capacity (Figure 4, orange dots), which ranged from 1% for SiO<sub>2</sub>-NPs-500°C to 293 53.5% for clay and NC-95/5. The best results were obtained for raw clay and NC-95/5 with a 294 removal efficiency > 50%. Both products were well-hydrated after 3 hours and probably had 295 similar surface areas and therefore similar adsorption capacity. This was not the case of NC-296 50/50, which showed reasonable but lower removal efficiency than for NC-95/5 and clay that 297 can be attributed to the high fraction of SiO<sub>2</sub>-NPs in the nanocomposite that has no affinity for 298 ciprofloxacin, as shown in Figure 3.

The reason why NCs are more efficient sorbent material than clay-500°C is certainly linked to the disordered and porous structure obtained in the NCs compared to clay-500°C, which is more favorable for fast diffusion of water and pollutants inside the material. In absence of NP inside the clay, the heating step at 500°C leads to sheet stacking that results in a dense structure that does not have the desired properties.

Although not responsive to ciprofloxacin,  $SiO_2$ -NPs play a significant positive role in the kinetics of ciprofloxacin adsorption. Based on kinetic and removal efficiencies, NC-95/5 appears to be the most appropriate candidate for efficient removal of ciprofloxacin. Further analysis of the NC- 307 95/5 sample was performed to check its crystallinity after adsorption of ciprofloxacin using XRD308 and to assess adsorption mechanism using XPS.

309

The mineralogy of NC-95/5 is preserved after adsorption as shown in Figure S1. The (001) reflection distance is shifted from 9.6Å (before adsorption) to 14.4Å (after adsorption). This basal reflection observed after adsorption of ciprofloxacin is typical of hydrated Camontmorillonite which is consistent with the fact that Evian water contains 80 mg/L of Ca<sup>2+</sup>. In addition to NC-95/5, calcium carbonate phases were detected which is, once again, consistent with the ionic composiiton of Evian water that contains bicarbonates (360 mg/L) which has favored the precipitation of calcite and aragonite (Figure S1).

317

318 XPS analysis was performed on NC-95/5 before and after sorption of ciprofloxacine to 319 investigate sorption mechanism. However, although XPS confirms the presence of ciprofloxacin 320 on the NC-95/5 substrate (specific signature of Fluor), it does not allow to identify the sorption 321 mechanisms mainly because of the presence of a significant fraction of adventious carbon 322 classicaly observed in XPS measurements (see Figures S2, S3, S4 and S5 in SI).

323

In addition to surface reactivity, the recoverability of the particles after treatment is an important
aspect to be taken into consideration and was tested on the different sorbents NC-95/5,
NC50/50, clay and clay-500.

327

328 3. Nanocomposite and clay recovery from liquid media

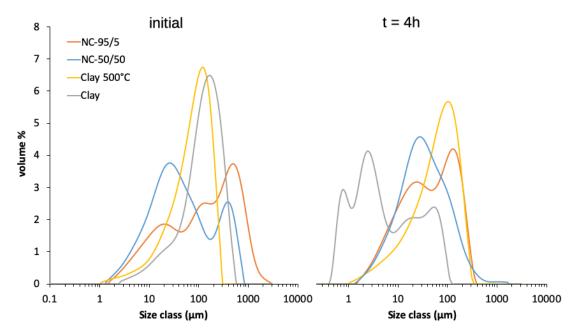


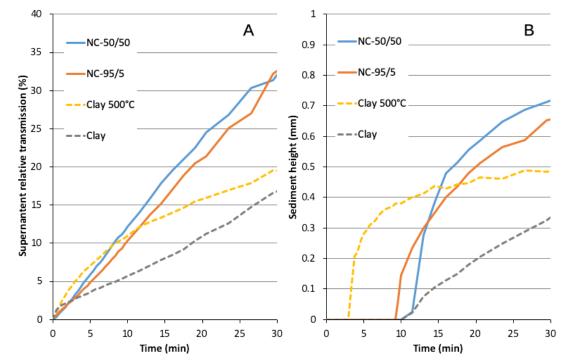
Figure 5: Size distribution of the clay and NCs in synthetic surface water, at  $t_0$  (left, just after dispersion) and after 4h of agitation (right).

330

334 Laser diffraction was used to measure the dispersion of the hybrid material in water compared 335 to raw clay and to clay-500°C. It proved that heating at 500°C causes the formation of 336 irreversible aggregates that can withstand the dispersion step in aqueous medium. Indeed, after 337 4 h of agitation in water, neither clay-500°C nor the NCs displayed a particle size below 1 µm, 338 which is roughly the upper limit of the colloidal size range. These three materials remained in an 339 aggregated form with sizes ranging from 5 to 200 µm, while after 4 h of agitation, the clay was 340 rapidly dispersed in sub-micrometer units. Colloidal dispersion of the montmorillonite clay 341 usually implies continuous swelling, due to hydration of the inter-lamellar cations toward multiple 342 layers of water molecules. Here the calcination step at 500°C led to dehydration of those cations 343 and shrinkage of the interlamellar space, which likely reduced access to the water molecules 344 from the outside when the material was redispersed in water. This resulted in the irreversible 345 aggregation of the clay particles with no noticeable difference between the NCs and the clay-346 500°C.

The larger aggregate size obtained here certainly favors more rapid sedimentation of the NCs, which should facilitate recovery from the water treatment plant after use. This was confirmed by the sedimentation trends shown in figure 6. After 4 h of agitation of the materials in synthetic surface water, the sedimentation rate was measured in terms of supernatant clarification and of formation of a sediment layer. The NCs clearly showed the highest sedimentation rates, 352 reaching 30-35% of average supernatant clarification in 30 minutes (Figure 6A). In comparison, 353 the pure clay, which still contained sub-micrometer dispersed units, sedimented at the lowest 354 rate, reaching only 15-20% in the same time period. Clay-500°C displayed different 355 sedimentation behavior characterized by a higher rate in the first 5 minutes, followed by a 356 slower rate. This suggests coexistence of high-density aggregates that settled out rapidly, and 357 of lower density aggregates that remained in suspension longer. These respective behaviors 358 are corroborated by the formation of the sediment layer shown in Figure 6B. The change in 359 sediment height over time points to higher kinetics of the nanocomposites than of pure clay, 360 reaching approximately 0.7 mm for NCs compared with 0.35 mm in 30 min for pure clay. Again, 361 clay-500°C underwent more rapid sediment formation in the first 5 minutes, after which the 362 process slowed down. It is important to note that the change in sediment height over time 363 results not only from the sedimentation rate but also from apparent sediment density. As the 364 NCs are characterized by lower apparent density than pure clay, the sediments formed are 365 expected to result in a thicker layer at a given mass of material. Such a contribution of the 366 material density cannot be distinguished from that of the sedimentation rate in Figure 6B.

Of note, in the water treatment, the sedimentation efficiency of particulate matter is usually adjusted by the adding chemicals such as coagulants and/or flocculants. So any increase in sedimentation caused by the sorbent material, as obtained in our case, could also lead to lower chemical dosage and new sludge characteristics and behavior (more compact, easier to treat and to dispose of).



374 Figure 6: Kinetics of supernatant clarification (A) and sediment formation (B) measured

375 respectively by light transmission and backscattering in the upper 39 mm and bottom 1 mm of

376 the sedimentation tube for clay-based materials

373

#### 377 4. Removal efficiency in groundwater

Table 2. Adsorption performances of the 5 micropollutants on activated carbon, NCs and thereference materials

Sorbent	Cipro.	Dano	Sulfa	Doxy	Carba
(50 mg/L)					
Activated	100%	100%	100%	100%	100%
carbon					
MMT/SiO <sub>2</sub>	99%	99%	0%	34%	3%
(50/50)					
MMT/SiO <sub>2</sub>	100%	100%	0%	60%	0%
(95/5)					
clay-500	94%	95%	0%	41%	0%

SiO <sub>2</sub> NP-	0%	0%	0%	11%	0%
500					

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382 SiO<sub>2</sub>-NPs-500°C showed no reaction to the five micropollutants tested in this study. NC-95/5 383 and NC-50/50 removal efficiencies were ranked in the following decreasing order Ciprofloxacin 384 = Danofloxacin > Doxycycline > Sulfamethoxazole = Carbamazepine. Further mechanistic 385 studies are now needed to explain this trend. However, based on the chemical composition of 386 the different compounds, one can hypothesize that some properties may control the affinity 387 between NCs and organics. The two antibiotics that are efficiently removed by NCs are the only 388 ones that have a carboxylic functional group (Table 1), which is known to be a good complexing 389 group, with pKa values of 5.86/8.24 and 6.07/8.56 for Ciprofloxacin and Danofloxacin, 390 respectively. At neutral pH, the carboxylic groups are deprotonated and most probably react 391 with the positive AI groups of the clay (LABILLE et al., 2005). Na-Montmorillonite has also been 392 shown to have good affinity for phenolic compounds (DJEBBAR et al., 2012), which may explain 393 the relatively good removal efficiency of doxycycline. Finally, the last two contaminants that 394 were not removed at all have neither carboxylic nor phenolic groups. They both have amino 395 groups that were neutrally charged at the pH of our experiment. Adsorption of aniline (the 396 phenyl group attached to an amino group) on Montmorillonite has been studied in the past and 397 showed maximum adsorption when the pH of the solution was approximately equal to the pKa 398 of the anilinium ion deprotonation reaction (pH 4.5-5) (ESSINGTON, 1994). Minimum adsorption 399 was observed at pH above 7, which could partially explain the low affinity of clay for 400 sulfamethoxazole and carbamazepine.

## 401 **Conclusion**

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Silica-clay nanocomposites were designed specifically for the removal of antibiotics in potable water treatment plants as an alternative to powder activated carbon. Among the five pharmaceuticals tested, the silica-clay nanocomposites enabled efficient removal of three molecules, ciprofloxacin, danofloxacin and doxycyclin, making them promising sorbent materials. In the hybrid structure of the nanocomposites, although non-reactive, the SiO<sub>2</sub>-NPs are homogenously distributed and strongly affect clay stacking by creating disorder which appears to favor the diffusion of water and micropollutants in the material. NC-95/5 showed the 410 highest reactivity among the synthesized products. In addition, the design of the 411 nanocomposites includes a calcination step, which resulted in the formation of non-reversible 412 aggregates > 1  $\mu$ m. After being used use as a sorbent in suspension, the material can thus 413 easily be recovered from the water using normal filtration or sedimentation techniques. Finally, 414 the use of natural clay is beneficial from an environmental point of view and the low loading of 415 SiO<sub>2</sub>-NPs is advantageous in terms of the cost of synthesis.

416 As a possible next step in this work, NCs could be tested with other specific pollutants that are 417 not easily treated by activated carbon due to their unique properties (hydrophilic, polar, mobile, 418 small molecules), such as persistent and mobile organic compounds (PMOCs) (ARP et al., 419 2017). PMOCs, such as pesticide metabolites, per- and poly fluoro alkyl substances etc., are 420 sufficiently mobile in an aquatic environment to enter drinking water resources, and sufficiently 421 persistent to survive current water treatment processes. Treatment of such compounds is a real 422 challenge for environmental and sanitary risk management. Another possible perspective is to 423 increase the efficiency of NCs by replacing the poorly reactive SiO<sub>2</sub>-NPs by Al- or Fe-based NPs 424 that have a stronger affinity for these micropollutants. Finally, as an alternative to micropollutant 425 removal through adsorption, recent publications pointed to the interest of oxydoreductases 426 (such as peroxidases, polyphenol oxidases), hydrolases (such as proteases, esterases, lipases 427 and cellulases) and lyases for wastewater treatment and especially for the degradation of 428 phenol, aromatic compounds and hormones (DE CAZES et al., 2014a). The interest of using 429 laccases and esterases for the degradation of antibiotics has recently been demonstrated and 430 more generally, the biodegradation treatment by enzymes fixed on an adsorbent support like 431 clay or membrane has been proposed as promising treatment solution (MAGNAN et al., 2004; DE 432 CAZES et al., 2014a; DE CAZES et al., 2014b; DE CAZES et al., 2015). In this perspective, NCs 433 could be used as a support for enzyme grafting.

# 434 Ethics approval and consent to participate

435 Not applicable

# 436 **Consent for pubication**

437 Not applicable

# 438 Availability of data and materials

- 439 The datasets used and/or analysed during the current study are available from the
- 440 corresponding author on reasonable request.

# 441 Competing interests

442 The authors declare that they have no competing interests

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# 446 Authors' contributions

- 447 CL, IB, AG and JL have designed the experiment, CL, KH-A, DB, AC, IB, AG and JL have
- 448 performed the experiments, All authors have actively participated to the interpretation of the
- 449 data, writting of the manuscript and approved the final manuscript.

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