

Human-driven atoll island expansion in the Maldives Virginie Duvat

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Human-driven atoll island expansion in the Maldives

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This study assesses the evolution of 104 inhabited and 82 resort islands across the Maldives between 2004-2006 and 2014-2016 and investigates the contribution of human drivers. Using multi-date satellite imagery, we show that 59.1% of these islands expanded as a result of human intervention, including mainly land reclamation (often associated with harbor development) and the construction of engineering structures. Land reclamation drove island expansion on 93.5% of inhabited islands, whereas land reclamation and/or the effects of engineered structures (especially groins and breakwaters) were involved in 79.2% of the resort islands. Almost 52% of the islands that increased in size experienced growth rates \geq 10%, with 13 inhabited and 6 resort islands exhibiting growth rates \geq 50%. As a result of widespread human intervention, these islands behave very differently from both the documented Pacific islands and the Maldivian islands of Gaafu-Alifu Dhaalu Atoll (most of which are 'natural'), 15.5% and 19.5% of which underwent expansion over the past decades, respectively. Changes on the islands were driven by political-institutional (one-size-fits-all risk reduction policy based on hard engineering in inhabited islands), demographic (population growth) and socioeconomic (developing and securing outer inhabited islands and post-tsunami tourism growth) factors. The Maldivian island model of 'reclamationfortification' has detrimental impacts on the environment while also questioning the future capacity of the Maldives to face a hard path under increasing risks. Results from the Maldives more broadly offer lessons for other island countries and territories by highlighting the importance of distinguishing between inhabited and exploited and natural islands, and accounting for path dependencies, when designing adaptation solutions.

Keywords

- 27 Land area change, shoreline change, human disturbances, socioeconomic development,
- sustainability, Indian Ocean.

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1. Introduction

31 Atolls are ring-shaped reefs that enclose a central lagoon and support islands composed of 32 unconsolidated or poorly lithified carbonate sand and gravel (Woodroffe & Biribo, 2011). Because they are small (generally <1 km²), low-lying (generally <4 m in elevation), 33 34 comprising unconsolidated material (Woodroffe, 2008), atoll islands are increasingly at risk 35 of marine flooding (Storlazzi et el., 2018; Giardino et al., 2018) and coastal erosion (Shope et 36 al., 2017; Shope and Storlazzi, 2019) under climate change. The risk arises because a range of 37 environmental changes will likely alter the attenuation of wave energy and sediment supply to 38 islands by the reef ecosystem (Quataert et al., 2015; Perry et al., 2018). These changes include 39 sea-level rise (Oppenheimer et al., 2019), increased storm wave heights (Mentaschi et al., 2017; Vitousek et al., 2017), coral reef degradation caused by ocean warming (illustrated by 40 41 mass coral bleaching caused by the 2015-2016 ENSO event; MacClanahan et al., 2019) and 42 acidification (Gattuso et al., 2015), and local human disturbances (Duvat et al., 2019; Duvat 43 and Magnan, 2019). Since the pioneer study by Webb and Kench (2010), however, increasing 44 assessments of atoll island change over the recent decades proved the prevailing persistence 45 of the land area of individual islands and atolls. This occurred despite higher (i.e. up to 46 5.1 ±0.7mm/y in Tuvalu over the 1950-2009 period; Becker et al., 2012) than average 47 (~1.8 mm/y for the twentieth century; Church et al., 2013) rates of sea-level rise in some atoll 48 regions. Collectively, these assessments showed that, over the past three to five decades, only 49 11.4% of the 709 documented Indo-Pacific islands decreased in size, whereas 15.5% 50 increased and 73.1% remained stable in area (Duvat, 2019).

51 Multidecadal assessments of atoll island change, however, may not reflect very recent 52 changes in island behavior caused by changes in the magnitude or nature of controlling 53 factors. These factors include, in particular, accelerating sea-level rise (IPCC, 2019) and 54 increasing local human disturbances observed on some atolls (e.g. Duvat et al., 2017; Duvat 55 and Magnan, 2019). To fill this gap, some studies investigated recent (since the mid-1990s or 56 early 2000s) short-term (over 4 to 12 years) changes. These studies, which mainly covered 57 uninhabited (that is, either natural or remote and cultivated) islands, found: (1) higher rates of 58 change for short periods compared to multidecadal time periods (Rankey, 2011; Mann et al., 59 2016; Duvat and Pillet, 2017); (2) differing trends between atoll regions over recent years, with erosion prevailing on some atolls – e.g. Aranuka in Kiribati (2005-2009; Rankey, 2011) 60 61 and Wotje in the Marshall Islands (2004-2012; Ford, 2013) – while stability (e.g. Tuvalu, 62 2005-2014; Kench et al., 2018) and accretion (Maiana, Kiribati, 2005-2009; Rankey, 2011; 63 Takapoto, French Polynesia, 1995-2013; Duvat and Pillet, 2017) predominated on others. 64 Changes were mainly attributed, first, to climate variability (e.g., absence of major storm 65 explaining island expansion on Takapoto; Duvat and Pillet, 2017); second, to autogenic 66 processes driving island adjustment to variations in external energy and/or sediment supply 67 (e.g., Maiana; Rankey, 2011; Taku, Papua New Guinea; Mann et al., 2016); third, to human 68 factors, generally activities near villages (e.g., Maiana and Aranuka; Rankey, 2011; Takapoto; 69 Duvat and Pillet, 2017). These studies investigating short-term island changes thus 70 emphasized two key points: first, the high variability of island behavior (erosion, stability, 71 accretion) depending on the local context; and second, the prominent contribution of natural 72 drivers to observed changes. 73 As a complement to these studies, the present study investigates recent short-term (from 74 2004-2006 to 2014-2016) changes in island land area and the drivers of these changes on 104 75 inhabited and 82 resort islands of the Maldives, using satellite imagery. Building upon a

previous paper (Duvat and Magnan, 2019) that highlighted the rapidity of human-driven island change resulting from land reclamation, harbor dredging and the construction of coastal protection structures, this study provides new data for the changes in 82 resort islands. It also compares the directions and modes of change of inhabited and resort islands and investigates the main political, demographic and socioeconomic drivers of change on these two types of islands. The present article thus addresses the following questions: (1) To what extent did the Maldivian inhabited and resort islands behave differently from 'natural' (that is, unsettled and unexploited, or agricultural) Indo-Pacific and inhabited Pacific atoll islands? (2) What were the main modes and drivers of island areal change on these two types of islands? (3) Which underlying political-institutional, demographic and socioeconomic factors drove island change? (4) What are the future implications of recent changes?

2. The Maldivian context

The Maldives Archipelago extends 890 km between 0°34'S and 6°57'N in the central Indian Ocean. It comprises 21 atolls and 4 oceanic reef platforms supporting 1,149 islands that cover a land area of 227.4 km² (Figure 1; GoM, 2015; Naseer and Hatcher, 2004). Maldivian islands formed on a foundation of lagoonal sediments between 5500 and 2500 yr B.P., depending on atolls (Kench et al., 2005; Perry et al., 2013). Because they are very low-lying (80% of land area <1m above mean sea level) and very small (0.20 km² on average), these islands are especially and increasingly (Wadey et al., 2017) vulnerable to flooding (e.g. flood events of 1987, 2007, 2015) under contemporary sea-level rise (Harangozo, 1992; Woodroffe, 2005; Woodworth, 2005; Khan et al., 2010). Between 1950 and 2009, relative sea-level trend values were of 1.3 ± 0.4 and 1.4 ± 0.4 mm/yr at Male' (North Kaafu Atoll) and Gan (Seenu Atoll), respectively (Palanisamy et al., 2014). Between 1989-1991 and 2015, the rates of sea-level rise accelerated, reaching values of 4.38 ± 0.4 mm/yr and 3.46 ± 0.2 mm/yr at Male' (1991-2015) and Gan (1989-2015), respectively (Wadey et al., 2017). Because subsidence is low

(i.e. of between 0.09 m and 0.16 mm/kyr), it is a minor contributor to relative sea-level rise in the Maldives (Gischler et al., 2018). The Maldivian climate is characterized by a reversal in winds and waves between the northeast monsoon (December to February) and the southwest monsoon (May to September), which causes a reversal in longshore sediment drift and marked changes in shoreline position between seasons (Kench and Brander, 2006; Kench, 2012). In addition, the Maldives Islands are affected by tropical depressions forming in the eastern part of the Indian Ocean and by distant-source swells (e.g. May 2007 event) originating from depressions forming in the temperate latitudes of the southern hemisphere (Aslam and Kench, 2017; Wadey et al., 2017). In 2016, the Maldives had 188 inhabited islands hosting 402,071 inhabitants and 126 resort islands that received the majority of the 1,286,135 international tourists visiting the country (GoM, 2017). Both population growth (doubling of the population every ~25 years since the 1960s; GoM, 2018) and tourism development (increase in total bed capacity by 7% from 2012 (25,691 beds) to 2017 (38,737 beds); GoM, 2017) put high pressure on the limited land area of the country and of individual islands. The small size of islands resulted in the scattering of the population into small communities, with 61.7% of the inhabited islands hosting less than 1,000 inhabitants in 2014 (GoM, 2015). It also led to a dramatic increase in the number of resort islands (from 126 in 2016 to 142 in 2018) to host continuously growing international tourists (1,484,272 in 2018) (ADB, 2019). Since the 1990s, the small size and the physical instability of the shoreline of these islands have increasingly appeared as a constraint and been addressed through engineering solutions. The latter include raising islands (that is, the creation of artificial elevated islands; Brown et al., 2020), artificially expanding existing islands (Naylor, 2015; Aslam and Kench, 2017; Fallati et al., 2017; Duvat and Magnan, 2019) and armoring and stabilizing shorelines (Shaig, 2011; Kench, 2012; Hinkel et al., 2018). Following the 2004 Indian Ocean Tsunami, which

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killed 82 people and displaced 30,000 people (half of whom had returned to their islands after two weeks) due to the heavy damage to 8,500 houses (Kan et al., 2007; GoM, 2009), island expansion and fortification have been enhanced. The National Population Consolidation Program (NPCP) has led the fortification of the islands. This program was already in place before the Indian Ocean Tsunami but accelerated in response to this event. Following the tsunami, the Seventh National Development Plan was adopted. This plan aimed at concentrating the country's population on a limited number of 'safer' (i.e. raised and protected by engineered structures) islands to better ensure the reduction of risks, while also boosting socioeconomic development through the provision of infrastructure and public facilities (Shaig, 2008; Suppasri et al., 2015). The long-term (1969-2006/2014) study of shoreline changes conducted on Gaafu Alifu-Dhaalu Atoll (see location in Figure 1), which considered 184 islands, most of which are uninhabited, found that 45% of the islands decreased in size while 55% remained stable or increased in area (Aslam and Kench, 2017). Erosional patterns were mainly observed on small islands and interpreted as resulting from changes in the nearshore process regime (i.e. increased sediment mobilization and transport) caused by sea-level rise and/or from a deficit in the sediment budget of islands. Island expansion was attributed principally to land reclamation, which occurred on twelve islands and caused a 93.8 ha increase in land area on the atoll scale. The constructive role of ocean swell on the southern rim of the atoll was a secondary cause of the expansion of islands.

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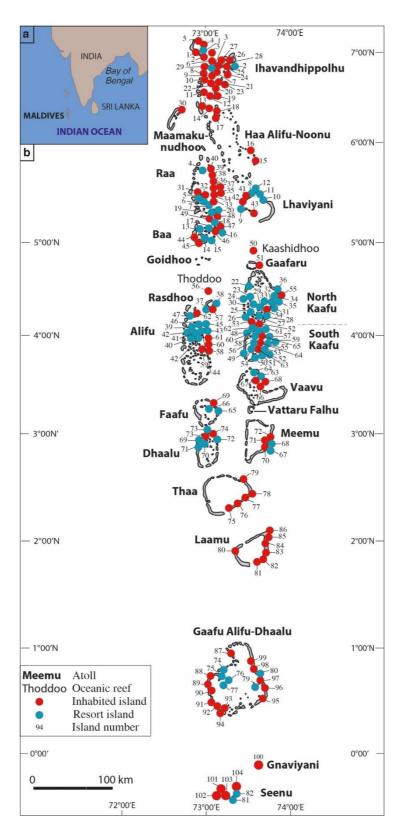


Figure 1. Location of the study islands.

Panels a and b respectively show the location of the Maldives in the Central Indian Ocean and the spatial distribution of the 104 inhabited and 82 resort study islands. The latter are numbered from north to south. The Supplementary Material uses the same numbering.

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3. Materials and Methods

3.1. Short-term change in the land area of islands

In line with a study analyzing recent land use change in the Maldives (Fallati et al., 2017), we used Google Earth historical satellite images (resolution of 0.5 m) from 2004-06 and 2014-16 to assess recent change in the land area of islands. Due to the limited availability of images, which is common on atolls (e.g., Webb and Kench, 2010; Ford and Kench, 2015; Ford, 2013; Aslam and Kench, 2017), we used two different sets (i.e., 2004-2006 and 2014-16) to document the each island at the two periods considered. The limitations encountered were variations in the dates at which the images were taken, even across a given atoll, and cloud cover masking some shoreline sections and islands. To optimize image coverage for a country-wide assessment, we therefore had to use a large number of images. As a result, the initial and final image used differ between islands (see Supplementary Tables 2 and 3). The images allowed assessment of areal change on the islands over the last decade ± 1 year for a total of 186 islands. Of these, 104 are inhabited while 82 are resort islands. The 186 sample islands are distributed across the entire archipelago (Figure 1), and respectively represented 55.3% and 65.1% of the Maldivian inhabited (188) and resort (126) islands in 2016. Together, they covered a land area of ~9,24 ha, which represented ~31.1% of the land area of the country, in 2014-16. At this date, the study islands had an average size of 49.7 ha, and minimum and maximum land areas of respectively 1.4 ha and 693 ha (Supplementary Tables 2 and 3). In line with previous studies including inhabited islands (Duvat et al., 2017), we used the stability line as a shoreline proxy. This indicator is appropriate to capture medium to longterm changes on islands change, as it is not affected by seasonal beach change. It corresponds to the vegetation line along natural shoreline and to the outer limit of human-built structures (e.g., seawalls, infrastructures, reclaimed areas) along artificial shoreline. Because many of the Maldivian inhabited and resort islands have artificial shorelines, using this proxy was particularly relevant. Because of the widespread removal of the coastal vegetation, digitizing artificial shorelines was much easier than those on natural islands with overhanging tree branches (Purkis et al., 2016). The only difficulty faced when digitizing the stability line was, on some islands, uncertainty on the status of bare ground—that is, whether it corresponded to reclaimed (i.e., stabilized) area. On the few islands where uncertainty was high, we excluded bare ground from the stabilized part of the island. For each island, we controlled the accuracy of shoreline digitization by checking inter-date consistency, i.e. always providing the same solution to the same problem. We used the stability line to calculate island land area at each date. The two sources of uncertainty included image resolution (0.5 m) and shoreline digitization, estimated to be <1 m based on repeated (n: 10 times) shoreline digitization by the same operator on a representative sample of 20 islands. These uncertainty values are consistent with those obtained in previous studies (Biribo and Woodroffe, 2013; Ford, 2013; Mann and Westphal, 2014). In line with previous assessments (Webb and Kench, 2010; McLean and Kench, 2015; Aslam and Kench, 2017; Duvat, 2019; Kench et al., 2018), we used the ±3% threshold to determine significant change. We therefore considered that a change falling within -3% and +3% was not significant, i.e., indicated areal stability, whereas changes <-3% and >+3% respectively corresponded to a reduction and to a gain in island land area.

3.2. Attribution of areal change on islands

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3.2.1. Potential drivers of island areal change

Over the timeframe of analysis, three main drivers may have caused marked island change.

199 These factors include (1) the Indian Ocean Tsunami, for the eight inhabited islands and nine

resort islands for which we used pre-tsunami images for the initial date (see Supplementary

Tables 2 and 3); (2) the May 2007 and May 2015 swell events (Wadey et al., 2017); (3) human intervention, especially harbor dredging and extension, island artificial expansion and the effects of engineered structures (Aslam and Kench, 2017; Fallati et al., 2017; Duvat and Magnan, 2019). A study investigating the morphological impacts of the Indian Ocean Tsunami on Baa (South Maalhosmadulu) Atoll (see Figure 1) concluded that it had limited impact on the land area of islands on this atoll (Kench et al., 2006). Pertaining to the May 2007 and 2015 swell events, available studies reported extensive flooding in highly-exposed western and southern islands, but no erosional impact. We therefore assumed that these events likely caused limited change to island land area over the period of analysis. Because human interventions have influenced changes in the Maldives islands over short periods of time (Kench, 2012; Aslam and Kench, 2017; Fallati et al., 2017; Duvat and Magnan, 2019), we paid special attention to this driver of change.

3.2.2. Contribution of anthropogenic drivers to areal change of islands

Because the Maldives exhibited rapid population growth and tourism development over the period of analysis, we made a special effort to detect the contribution of local human disturbances to island areal change. The latter include land reclamation, harbor development, engineered protection structures, quays and jetties, beach nourishment, vegetation destruction or planting, and dredging in reef flat (Webb and Kench, 2010; Ford, 2012; Hamylton and East, 2012; Biribo and Woodroffe, 2013; Yates et al., 2013; Purkis et al., 2016; Aslam and Kench, 2017; Duvat et al., 2017; Duvat et al., 2019). As in most previous studies, this contribution focused on interpreting images. We first classified islands into two categories, based on the occurrence (coded A) or non-occurrence (coded n.a.) of direct (i.e., visually detectable on images) human disturbances in 2014-2016. Some islands which exhibited human disturbances in 2004-2006 showed no additional human disturbances in 2014-2016 (coded A/). We then focused on the islands that exhibited additional human disturbances in

226 2014-2016 compared to 2004-2006, and classified them based on the human disturbance(s) 227 observed (see categories in Tables 1 and 2). 228 We assessed land reclamation on a binary basis: 1. Yes; 2. No. We did not estimate the 229 surface of reclaimed areas because it was difficulty to (i) discern pre-2004-06 reclaimed areas, 230 and (ii) establish if land gains were entirely due to reclamation or also to other factors (e.g., 231 current-driven sediment deposition, beach nourishment). The absence of comprehensive and 232 updated official data on land reclamation did not allow us to overcome these uncertainties. As 233 in previous studies (Aslam and Kench, 2017; Duvat et al., 2017), however, we were able to 234 assess with a high confidence level the occurrence of reclamation and its overall contribution 235 (i.e. high vs. low) to island change. 236 Concerning harbor development, we inventoried "proper" harbors having facilities (i.e., quay, 237 jetty or breakwaters), which are the major infrastructure found on inhabited and resort islands. 238 A proper harbor disrupts longshore and cross-shore sediment transport and deposition 239 (Kench, 2012), which can cause a change in island area over time through a marked increase or decrease in sediment accumulation along the shoreline. In addition, the extension of 240 241 harbors or dredging in reef flat provides important volumes of materials used for land 242 reclamation (Aslam and Kench, 2017; Duvat and Magnan, 2019). The latter is generally 243 carried out either at harbor ends or in harbor area (qualified as "land reclamation at harbor end 244 or in harbor area"), or along the entire shoreline (qualified as "extensive reclamation"). The 245 large-scale reclamation projects implemented by the Government to support the development 246 of regional centers (meant to reduce inequalities between the capital region and outer islands) 247 and implement the National Population Consolidation Program (Shaig, 2008) were also 248 classified as "extensive reclamation". 249 The third human disturbance examined was the establishment of coastal (longitudinal or 250 transversal) and marine (longitudinal) engineered protection structures along island shoreline

and reef crest. While longitudinal coastal (mostly seawalls and riprap) and marine (breakwaters) structures are common in inhabited islands, groins and foreshore breakwaters are widely used on resort islands (Shaig, 2011). Whereas groins are employed to intercept sediments and to stabilize artificially-nourished beaches, breakwaters are used to reduce wave energy and thereby promote sediment deposition along the shoreline. All of these structures have a high influence on island areal change, by inhibiting sediment deposition, which may lead to land loss (e.g. wave reflection by seawalls), or promoting sediment deposition and thereby causing land gain (e.g. foreshore breakwaters). For each island, we qualitatively assessed changes in the presence of such structures based on the comparison of the two sets of images.

Other detectable human disturbances (vegetation removal, jetty construction, beach nourishment, etc.), which played a secondary role in island areal change, were also inventoried based on visual detection.

4. Results

265 4.1. Change in island area

The sample islands comprise 104 inhabited islands and 82 resort islands, with the former generally much larger than the latter in 2014-16. The proportion of islands exceeding 50 ha in size was respectively nil and of 46.1% for the resort and inhabited sample islands (Supplementary Table 1). In addition, 65.8% of the resort islands had a land area <10 ha in 2014-16 against 1.9% of the inhabited islands. Over the past decade, 59.1% of islands (110 islands, including 62 inhabited and 48 resort islands) increased in size ($x \ge 3\%$), whereas 38.2% (71 islands, including 42 inhabited and 29 resort islands) were stable and only 2.7% (5 resort islands) decreased in size (Figure 2; Supplementary Tables 2 and 3). These data mean that 97.3% of islands either expanded or were stable in area between 2004-2006 and 2014-2016. The proportions of expanding and

stable islands were almost similar for inhabited and resort islands—that is, respectively of 59.6% and 58.5% for expansion, and of 40.4% and 35.3% for stability. The five resort islands that decreased in size exhibited rates of change comprised between −3.6% (Kanifushi, Baa Atoll) and −17.7% (Kandoomafushi, South Kaafu Atoll). These islands had areas between 2.4 (islands Nos. 11 and 33 in Supplementary Table 3) and 15.8 ha (island No. 53) in 2004-06. The results highlight the physical instability of the smallest islands, especially those of less than 10 ha (Duvat, 2019), which predominantly underwent rapid expansion (Figure 2). A correlation between island size and the amplitude of change also showed the smallest islands undergoing the largest changes. Importantly, 51.8% of the islands that increased in size (57 islands out of 110, including 33 inhabited and 24 resort islands) experienced growth rates ≥10%, with 13 inhabited and 6 resort islands exhibiting rates ≥50%.

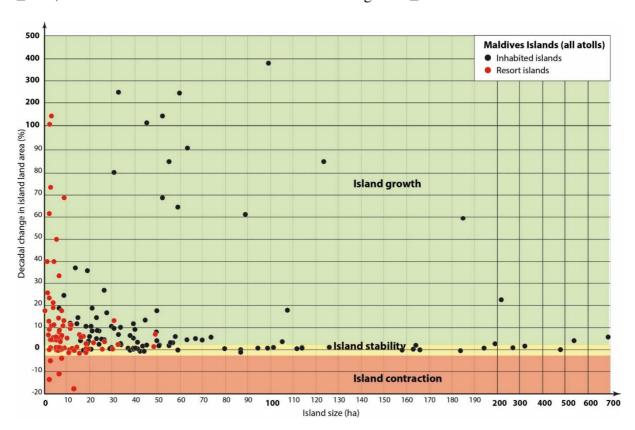


Figure 2. Decadal change in island land area on 186 inhabited and resort islands in the Maldives.

The yellow area corresponds to areal stability (3%<x<-3%), while the green and red areas respectively indicate island expansion (x \ge 3%) and contraction (x \le -3%). The sample islands, including islands >100 ha, predominantly exhibited growth (110 islands; 59.1%) and secondarily experienced stability (71 islands; 38.2%). Only 5 small resort islands showed contraction.

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4.2. Contribution of human activities to island areal change

More than 90% of the inhabited islands (94 out of 104) and 85% of the resort islands (70 out of 82) exhibited additional human disturbances in 2014-16 compared to 2004-2006. Direct human disturbances contributed to island areal change on 98.1% of inhabited islands (i.e. 102 islands out of 104; Table 1) and 93.9% of resort islands (i.e. 77 islands out of 82; Table 2). Only 8.1% of islands (7 islands) experimented areal change without any detectable human influence. The main types of human disturbances involved in the areal change of islands were different between inhabited and resort islands (Tables 1 and 2). Land reclamation occurred on 96.7% (89 out of the 92) of the inhabited islands that experienced additional human disturbances in 2014-16 compared to 2004-06. On 91.0% (81) of these islands (A1 to A5 in Table 1), land reclamation was associated with the dredging or extension of an existing harbor, or with the dredging of one or more new harbor(s) in the island reef flat. Land reclamation occurred either in the harbor area only (49 islands; A1, A2, A4), or on a larger scale, i.e. along most of the harbor coast or even along other shoreline sections (32 islands; A3, A5). Except for the two islands of Seenu Atoll that exhibited areal gains of 8.0 ha and 24.0 ha (islands Nos 101 and 104, respectively; Supplementary Table 2), the islands that experienced land reclamation in harbor area only exhibited areal gains of less than 4 ha. Areal gains due to land reclamation in harbor area caused a maximum island expansion of 11.6% (island No 76). In contrast, the islands that experienced extensive land reclamation (i.e. including outside the harbor area) showed much greater areal gains (i.e. >20 ha for 12 islands, with a maximum gain of 76.8 ha

for island No75; Supplementary Table 3). Gains in land area were ≥10% of initial island area for 33 of these islands (31.7%). Thirteen of these islands (12.5%) experienced gains $\geq 50\%$, with 5 islands (4.8%) even experiencing gains $\geq 100\%$ (Nos 41, 44, 74, 75 and 79, with a maximal value of 384% for island No 75). Most atolls had islands exhibiting such high growth rates. All in all, land reclamation (either associated with harbor dredging or not) was involved in 58 out of the 62 islands that expanded over the period of analysis. These results emphasize the major role of harbor dredging in providing the materials needed for artificially expanding inhabited islands. This means that the considerable and unprecedented effort made by the Government of the Maldives to equip these islands with harbors largely explains the widespread (i.e. in most atolls) and dramatic island expansion observed. Among the 82 resort islands sampled in 2014-16, 29.3% (24 islands) were still unexploited in 2004-06. These data reflect the rapid growth of tourism between 2004-06 and 2014-16 (101 resort islands in 2011; 115 in 2015; 126 in 2016; GoM, 2017). On resort islands, human intervention consisted of the establishment of hard protection structures along island shoreline (noted on 41 islands; Table 2 and Supplementary Table 3) and/or on island reef flat (foreshore breakwaters, noted on 17 islands). Among the former, 19 showed no other detectable human intervention, while 22 also experienced other disturbances such as land reclamation, dredging or the construction of a quay or jetty. Human intervention secondly consisted of land reclamation, which affected 39.0% of islands (32 islands). Land reclamation was associated with harbor extension or dredging on 8 islands, but also noted on 23 islands where no harbor works were present. On these latter islands, we often observed sediment extraction from the island reef flat or lagoon, and sometimes associated with beach nourishment (see A11 in Table 2). Although no quantitative data are available on sediment extraction, we assume that the latter has significantly contributed to island growth. Focusing on the 48 resort islands (58.5% of the total) that exhibited expansion over the period of analysis, we note that

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respectively 14 (29.2%) and 13 (27.1%) experienced land reclamation and the construction of new protection structures, whereas 11 additional islands (22.9%) exhibited both types of interventions. Such works were also carried out on islands that remained stable or that decreased in size. Among the 29 resort islands that exhibited areal stability, for example, three experienced land reclamation whereas 13 had new protection structures. Likewise, two out of the five islands that decreased in size were equipped with protection structures. The maximum areal gain, noted on two islands (Nos 36 and 50 in Supplementary Table 3), reached up to 3.7 ha. Gains in land area were \geq 10% of initial island area for 24 islands (50.0% of expanding resort islands). Six of these islands experienced gains \geq 50%, with two islands even exhibiting gains \geq 100% (maximum value: 140%, No 52). These findings emphasize the major contribution of rapid tourism growth to island areal gain, even though resort islands experienced smaller areal gains compared to inhabited islands. The proportion of islands exhibiting high growth rates was higher in the central tourist atolls of North Kaafu, Alifu and South Kaafu, compared to other atolls. On South Kaafu Atoll, for example, 10 out of the 14 resort islands experienced gains \geq 10%, 4 of which even exhibiting rates \geq 50%.

Table 1. Human disturbances involved in inhabited island areal change between 2004-2006 and 2014-2016 (see Supplementary Table 2 for details).

Driver	Number of islands affected
A1. Land reclamation at harbor end or in harbor area	5
A2. New harbor(s) dredged in reef flat associated with land reclamation at harbor end	32
or in harbor area	
A3. New harbor(s) dredged in reef flat associated with extensive land reclamation	16
A4. Harbor extension with limited or moderate land reclamation at harbor end or in	12
harbor area	
A5. Harbor extension with extensive land reclamation	16
A6. Extensive land reclamation	8
A7. Other changes (e.g. breakwater construction or extension)	3
Sub-total Sub-total	92
n.a. Inhabited island exhibiting no significant human disturbance in 2014-2016	2
A/ Inhabited island exhibiting influential human disturbances in 2004-2006 showing	10
no additional human disturbances in 2014-2016	
Total	104

Ninety-two inhabited islands (88.46%) exhibited areal change due to human disturbances between

2004-2006 and 2014-2016. This table emphasizes the major contribution of harbor dredging

associated with land reclamation in inhabited island areal change.

Table 2. Human disturbances involved in resort island areal change between 2004-2006 and 2014-2016 (see Supplementary Table 3 for details).

Driver	Number of islands
	affected
A1. Land reclamation at harbor end or in harbor area	4
A2. New harbor dredged in reef flat associated with land reclamation at harbor end or	3
in harbor area	
A3. Extensive land reclamation	6
A4. Engineered protection structures (breakwaters, groins, seawalls)	19
A5. Land reclamation + engineered protection structures	14
A6. New harbor dredged in reef flat associated with land reclamation at harbor end or	1
in harbor area + engineered structures	
A7. New harbor dredged in reef flat	1
A8. Dredging in reef flat	1
A9. Dredging in reef flat + engineered structures	6
A10. Dredging in reef flat + land reclamation	3
A11. Beach nourishment*	2
A12. Vegetation destruction	2
A13. Quay or jetty	2
A14. Quay or jetty + engineered protection structures	1
Sub-total Sub-total	65
n.a. Resort island exhibiting no significant human disturbance in 2014-2016	5
A/ Resort island exhibiting influential human disturbances in 2004-2006 showing no	12
additional human disturbances in 2014-2016	
Total	82

This table highlights the major contributions of land reclamation (A1, A2, A5, A6, A10, A15) and

engineered protection structures (A4, A5, A6, A9, A14, A15) in resort island areal change. * Underestimated due to detection difficulties.

4.3. Direction and modes of change in inhabited and resort islands

Beyond the general trend described above, inhabited islands exhibited contrasting degrees of human modification over the decade of analysis. Whilst land reclamation was the main driver of change on these islands, its influence varied substantially between islands. Four types of situations are possible. First, islands experience only limited and localized (i.e. in harbor area) land reclamation that did not generate any gain (x<3%) in land area (e.g. +0.9% in the case of

Kodey; Figure 3a). These islands represent 36.7% of the inhabited islands affected by land reclamation. Second, islands exhibit a moderate to significant (3%≤x≤10%) gain in land area, which represent 26.7% of reclaimed islands. Third, islands exhibit a marked areal gain (e.g. +17.9% for Ukulhas; Figure 3a) due to extensive land reclamation associated with harbor dredging. These islands represent 22.2% of reclaimed islands. Fourth, islands exhibit total transformation in shape, size and sometimes position on the reef flat, following extensive land reclamation, like Thulhaddhoo and Vilufushi (Figures 3c and 3d). These islands are present in nine atolls and represent 14.4% of inhabited islands affected by land reclamation. Thimarafushi (Thaa Atoll, No 75) was considerably extended for airport construction between 2010 and 2012 (+76.8 ha, from 20 ha in 2005 to 96.8 ha in 2016), for example, which led to the incorporation of the nearby islet of Hiriyanfushi.



Figure 3. Patterns of change on inhabited islands.

a shows an island that underwent limited shoreline and areal change (+1 ha corresponding to a gain of 0.9%). The main change observed was localized lagoon shoreline advance caused by harbor dredging-induced land reclamation. b shows an island that experienced marked changes (+3.2 ha corresponding to a gain of 17.9%), as a result of harbor extension and associated extensive land reclamation. c and d show islands that underwent dramatic shoreline and areal change (+22.5 ha corresponding to a gain of 105.1% for Thulhaadhoo; +42 ha corresponding to a gain of 241.4% for Vilufushi) due to extensive land reclamation. On such islands, the entire shoreline is now artificial, i.e. fixed by engineered structures.

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Over the period of analysis, resort islands also experienced various degrees of human-induced areal change. Some islands underwent limited change in area and configuration. This occurred both on islands that were already exploited in 2004-06 and affected by additional human disturbances in 2014-16 (e.g. Alidhoo; Figure 4b), and on islands that were natural in 2004-06 and then became a resort island (e.g. Havoddaa, Figure 4c). At the other end, some resort islands have undergone marked changes in area and configuration, as a result of the cumulative effects of land reclamation, additional coastal protection structures and beach nourishment. Halaveli (No44) expanded by 40.0%, for example, due to land reclamation and increased sediment deposition. Extension of foreshore breakwaters along the reef crest caused the sediment deposition, with the breakwaters now forming a nearly-continuous protection around the island (Figure 4a). In some cases, extensive land reclamation caused a marked increase in land area and a complete change in the shape of the islands, as illustrated by Maadhoo, South Kaafu Atoll (+140%; Figure 4d). The islands that expanded through land reclamation were often equipped with groins, meant to stabilize beaches (Figures 4a and 4d). The impacts of tourism development on island areal change were highly variable, with some of the most recently-developed resort islands (e.g. Maadhoo, opened in 2016; Figure 4d) undergoing total transformation.



Figure 4. Patterns of change on resort islands.

a shows an island that experienced major shoreline and areal change (+1.2 ha corresponding to a land area gain of 40%), as a result of land reclamation. The establishment of foreshore breakwaters has increased sediment deposition on the reef flat and at the coast, which has allowed the development of the coastal vegetation. b illustrates limited shoreline and areal change (+0.3 ha corresponding to an areal gain of 2.2%). It also shows the use of foreshore breakwaters to protect water bungalows from waves. On this island, breakwaters have so far not caused major changes to island configuration. c shows that despite its change from natural to a resort island in 2015, Havoddaa (8.4 ha) experienced limited shoreline change and no areal change to date. d shows the landreclamation-induced expansion of Maadhoo (3.6 ha in 2016; + 140%). c and d respectively illustrate environment-friendly and environmentally-destructive tourism development.

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5. Discussion

5.1. Originality of the trajectory of change of inhabited and resort Maldivian islands

Study islands behaved very differently from both inhabited Pacific atoll islands and Maldivian rural and unsettled islands, which mainly exhibited stability over the past decades. Fifty-nine percent of study islands underwent expansion, against 15.5% of the documented Indo-Pacific 709-island sample mainly comprising rural and natural islands (Duvat, 2019) and 19.5% of the 184 predominantly unsettled Gaafu Alifu-Dhaalu islands (Aslam and Kench, 2017). In one decade, 31.7% and 50.0%, respectively, of the inhabited and resort islands that increased in size experienced growth rates ≥10%. Respectively, 21% and 12.5% of these islands exhibited growth rates ≥50%. These rates contrast with those of naturally accreting islands, which rarely exceed 10%. Land reclamation contributed to island expansion on 93.5% of inhabited islands and 52% of resort islands.

5.2. Political-institutional, demographic and socioeconomic drivers of inhabited island

441 expansion

Over the period of analysis, the M.A. Gayoom (1978-2008, Progressive Party of Maldives) and A. Yameen (2013-2018, Progressive Party of Maldives) administrations – shortly interrupted by the 2008-2012 M. Nasheed Presidency (Maldivian Democratic Party) – gave priority to socioeconomic development to reduce tensions between the capital region (including Male', Hulhumale' and Villingili islands) and peripheral atolls. These tensions result from high disparities with regards to access to staple goods, jobs, education and health facilities (Arnall and Kothari, 2015; Hirsch, 2015; Speelmann et al., 2016). To reduce these tensions and especially put an end to chronic shortages of products affecting human wellbeing and economic activity in outer atolls (Hirsch, 2015), the Government made unprecedented nation-wide investments in harbor development between 2004-2006 and 2014-2016. These investments also aimed at repairing the tsunami-induced damage caused to harbors and jetties on 104 inhabited islands (GoM, 2009). Harbor development provided large

amounts of construction materials used to expand densely populated islands through land reclamation (Naseer, date unknown). This investment, in turn, led to increased shoreline armoring due to (i) the stabilization of newly-reclaimed areas by protection structures, and (ii) the erection of defenses in areas adjacent to harbors and affected by shoreline destabilization (Shaig, 2011; Kench, 2012). Political-institutional factors likely explain why one single island development model (the 'reclamation-fortification model') was implemented across the whole archipelago. The centralization of power at the highest levels of government leads to the standardization of risk reduction action on the ground. The non-consultation and non-involvement of local level governments, NGOs and concerned communities in decision-making and implementation reduce opportunities for the diversification of experiments and explain the spread of the 'reclamation-fortification model' across the whole country. This model, which originates in the initial choice of hard protection by Maldivian politicians, has since then persisted due to institutional inertia (Ratter et al., 2019, p. 175). Shoreline and island fortification generally meet the preference of inhabitants, who view hard engineering as a "true" and permanent solution providing value-for-money (Shaig, 2011). This preference comes from: (1) the inhabitants' view that the Government is responsible for and should therefore fund coastal protection; (2) the fact that the Government funds hard protection, but not soft measures, which reinforces the view that hard engineering is the only appropriate solution; (3) the disproportionate damage caused by the 2004 Indian Ocean Tsunami to unprotected outer islands compared to armored Male', which reinforced this view (Shaig, 2008; Sovacool, 2012; Ratter et al., 2019). Since the 1980s, the socioeconomic tensions highlighted above have caused massive population movements from outer atolls to the capital area, including Male' (128,768 inhabitants in 2016) and, since the mid-2000s, the artificial and raised island of Hulhumale'

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(from 2,866 in 2006 to 15,769 inhabitants in 2014), created to meet housing, commercial and development needs of the capital area (Mohit and Azim, 2012). These population movements have led to the continued population and economic growth of the Male' region, which has further exacerbated the preexisting socioeconomic gap and thus reinforced the attractiveness of the capital region (Speelmann, et al., 2016). To reduce these population movements, the Government has, since the 1960s, designed successive population consolidation policies aimed at regrouping the population in a limited number of large islands that would provide public facilities and economic opportunities (Shaig, 2008). These policies have all failed until the 2004 Indian Ocean Tsunami, which opened a 'policy window' for population consolidation through the Safe Island Development Program (SIDP) aimed at (i) improving public service delivery (including transportation, through harbor and airport development) and social development, (ii) increasing economic growth, and (iii) reducing environmental vulnerability, on a limited number of outer islands. Even though financial and technical constraints have limited the implementation of the SIDP, this program is responsible for the transformation and high expansion rates of those islands designated regional or growth centers. This process is illustrated by Vilufushi, Thaa Atoll (Figure 3d), which is one of the five new regional centers of the country (Plewa, 2018). During the tsunami, Vilufushi suffered widespread destruction and its population was evacuated to nearby Buruni. Its reconstruction included its expansion through reclamation (+42 ha; +241.4%), the armoring of the entire shoreline and the construction of 309 new houses, which allowed the population to be back on the island in May 2009 (GoM, 2009). This new enlarged and safer island, aimed at acting as an attractive regional center, illustrates the emergence of a new inhabited island type, represented by thirteen islands in this study.

5.3. Post-tsunami tourist recovery drives resort island change

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By causing an unprecedented economic crisis – with damages equaling to 62% of the country's GDP, making the Maldives the hardest hit country in Asia (Luetz, 2017) – the Indian Ocean Tsunami pushed for the prioritization of economic recovery through tourism growth. Twenty-one resort islands totaling 4,000 beds (23% of the total) were devastated and closed after the Tsunami, which had all reopened by April 2009. To speed up the recovery process and generate additional economic opportunities, many new resorts have been authorized after the tsunami (GoM, 2009), which explains the marked increase in the number of resort islands and engineering works contributing to island expansion between 2004-06 and 2014-16. Over this period, tourism recovery and growth led to the construction of airports in peripheral atolls, which required the extensive reclamation of some inhabited islands. Thimarafushi Island, Thaa Atoll (No 75; Supplementary Table 2), illustrates the process of extensive reclamation (+76.8 ha) for constructing an airport.

5.4. Implications of the Maldivian 'reclamation-fortification island model'

The 'reclamation-fortification island model' raises environmental and financial issues, with major implications on the security of future populations. First, it compromises (through reef flat destruction, obstruction of sediment transport and annihilation of coastal accommodation space) the capacity of the 'self-maintained' (Temmerman et al., 2013) reef-island system to adjust to sea-level rise through vertical growth, causing a path-dependency to engineering (Duvat and Magnan, 2019). In turn, this path-dependency to engineering questions the future capacity of the Maldives to ensure (technically and financially) the maintenance and upgrading of defense structures to contain increasing sea-related risks. Internal funding will be increasingly challenged by the impacts of increasing disaster frequency on tourism revenues (flood-related and potentially Covid-like), while international funding is uncertain. Lastly, the accelerated human-induced degradation of the reef ecosystem will amplify the negative impacts of climate change on food supply. The impacts are all the more urgent as

fish consumption is key to population diet (per capita consumption of 181 kg/yr; Hohne-Sparboth, 2013), with consumption of reef fish increasing at household level (Sattar et al., 2014).

Conclusions

In summary, analysis of 104 inhabited and 82 resort islands across the Maldives archipelago produced the following answers to the research questions posed. First, this study shows that the Maldivian inhabited and resorts islands behave very differently from both 'natural' Indo-Pacific and Pacific inhabited atoll islands between 2004-2006 and 2014-2016. Fifty-nine percent of the former underwent expansion, compared to 15.5% of the documented Indo-Pacific 709-island sample (Duvat, 2019) and 19.5% of the predominantly natural islands of Gaafu-Alifu Dhaalu Atoll, Maldives (Aslam and Kench, 2017). Respectively, 31.7 and 50.0% of expanding inhabited and resort islands experienced growth rates \geq 10%, with respectively 21 and 12.5% of these islands even exhibiting growth rates \geq 50%, while naturally accreting islands show growth rates \leq 10%.

Second, results show that the main modes of areal changes on the islands were different between inhabited and resort islands. Land reclamation, often associated with harbor development, drove island expansion on 93.5% of inhabited islands, whereas land reclamation and/or the effects of engineered structures (especially groins and breakwaters) were involved in 79.2% of resort islands.

Third, the factors underlying the expansion of islands were different between these two islands types. The implementation of one single island model, the 'reclamation-fortification island model', in inhabited islands was likely due to a combination of political-institutional

(effort made by the Government to develop and secure outer islands), demographic (population growth) and socioeconomic (reduce disparities between the capital region and outer islands) factors. Simultaneously, the unprecedented economic crisis caused by the Indian Ocean Tsunami pushed for the prioritization of economic recovery through tourism growth over the 2004-2006/2014-2016 period. This process was achieved through the expansion of already operating and the development of new resort islands, which encouraged airport development, all of which pushed for land reclamation and increased use of engineered structures.

Fourth, this study highlights the long lasting implications of the 'reclamation-fortification island model', which compromises the capacity of the reef-island system to naturally adjust to sea-level rise. This limitation creates a path dependency to engineering (which has heavy technical and financial implications under increasing risks implying structure maintenance and upgrading over time). It is also detrimental to local, including reef-based, fish supply, which plays a major role in outer island population diet.

Finally, results from the Maldives Islands provide lessons for other island countries and territories worldwide. They show that in these settings, densely urbanized and developed coastal areas may have undergone rapid and marked physical transformation over the past decade, as a result of the widespread use of engineering solutions to face both climate (i.e., to reduce increasing coastal risks) and socioeconomic (i.e., to face population growth and boost economic development) challenges. The underlying endogenous (i.e., political-institutional, socioeconomic, cultural, etc.) and exogenous (e.g., impact of the Indian Ocean Tsunami on international tourism in the Maldives) drivers involved in the widespread use of engineering solutions have, at least in some islands, already generated a hard path dependency that

reduces the range of options available to face future climate challenges. Such results advocate first, for distinguishing between developed (i.e., inhabited and exploited) and natural islands, and second, for accounting for path dependencies, when designing adaptation strategies. This means that adaptation solutions must absolutely be island-specific in these settings.

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Data availability

- All data generated during this study are included in this published article and its
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