

# Human mercury exposure levels and fish consumption at the French Riviera

Mariia Petrova, Melanie Ourgaud, Joana R.H. Boavida, Aurélie Dufour, Javier Tesán Onrubia, Aurelia Lozingot, Lars-Eric Heimbürger-Boavida

# ▶ To cite this version:

Mariia Petrova, Melanie Ourgaud, Joana R.H. Boavida, Aurélie Dufour, Javier Tesán Onrubia, et al.. Human mercury exposure levels and fish consumption at the French Riviera. Chemosphere, 2020, 258, pp.127232. 10.1016/j.chemosphere.2020.127232. hal-02935253

HAL Id: hal-02935253

https://hal.science/hal-02935253

Submitted on 13 Nov 2020

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1	<b>Human mercury</b>	exposure levels	and fish	consumption a	t the French

- 2 Riviera
- 3 Mariia V. Petrova<sup>a\*</sup>, Mélanie Ourgaud<sup>a</sup>, Joana R. H. Boavida<sup>a,b</sup>, Aurèlie
- 4 Dufour<sup>a</sup>, Javier A. Tesán Onrubia<sup>a</sup>, Aurelia Lozingot<sup>a</sup>, Lars-Eric
- 5 **Heimbürger-Boavida**<sup>a\*</sup>
- a. Aix Marseille Université, CNRS/INSU, Université de Toulon, IRD, Mediterranean
- 7 Institute of Oceanography (MIO) UM 110, 13288 Marseille, France
- 8 b. Center of Marine Sciences (CCMAR), University of Algarve, 8005-139 Faro,
- 9 Portugal

11

10 \*Corresponding author: lars-eric.heimburger@mio.osupytheas.fr

# Abstract

13	Humans are exposed to methylmercury (MeHg), a bioaccumulative neurotoxin,
14	mainly through the consumption of marine fish. Several studies showed that
15	high MeHg exposure can lead to neurological damage. This is particularly
16	relevant for pregnant women, because MeHg exposure negatively impacts foetal
17	development. Populations living near the sea are generally at increased
18	exposure risk due to higher consumption of fish and seafood. Here, we present
19	the first study of MeHg exposure levels of the population living at the French
20	Riviera, using mercury (Hg) concentrations in hair as a proxy for MeHg
21	exposure. We found that older people that consume more fish presented the
22	highest hair Hg concentrations. Compared to other Mediterranean bordering
23	countries and other European countries, the southern France population is
24	among those with high MeHg exposure (median for women of childbearing age
25	is $0.56~\mu g~g^{-1}$ ). A global implementation of the Minamata Convention is
26	necessary to lower MeHg exposure of the population.
27	Key words: hair, methylmercury, Southern France
28	
29	
30	
31	
32	
33	

# 1. Introduction

34

Methylmercury (MeHg) is a bioaccumulative neurotoxin, which is more 35 toxic than inorganic mercury (NRC/NAS, 2000). It has been consistently 36 observed that MeHg is biomagnifying to harmful levels along the marine food 37 webs with increasingly greater MeHg proportion (close to 100%) in top-38 predators (Lavoie et al., 2013, Chen et al., 2018, Ourgaud et al., 2018). 39 Humans are exposed to MeHg mainly through their diet, especially through 40 the consumption of fish and other products from aquatic environment 41 (Schoeman et al., 2009, Castaño et al., 2015, Pérez et al., 2019, Sunderland et 42 al., 2018). There is a global trend of increased fish consumption, which may 43 44 increase the human risk of exposure to trace elements, including Hg (FAO, 2016). Worldwide, annual per capita fish consumption grew from 9.9 to over 45 20 kg between 1960 and 2015, respectively (FAO, 2016). In France, the 46 consumption of fish is high (annual per capita fish consumption ranges from 30 47 to 60 kg, FAO, 2012, Vieira et al., 2015) and seafood has always been an 48 important part of the Mediterranean population diet (Castaño et al., 2015; Faget, 49 50 2009; Metro et al., 2017). The neurotoxicity of MeHg was first described for the Minamata incident, in 51 Japan, in 1956 (Kurland et al., 1960). High-level human MeHg exposure after 52 release of MeHg from a chemical factory led to fatalities and devastating 53 neurological damage in Minamata (Harada, 1995). Particularly, foetuses are 54 considered much more sensitive to MeHg exposure (NRC/NAS, 2000). 55

Consumption of fish contaminated with MeHg by pregnant women has resulted 56 in serious damage of the central nervous system such as mental retardation, 57 cerebral palsy, blindness and deafness (NRC/NAS, 2000). Since 1970, several 58 large-scale epidemiological studies showed that even low-level chronic MeHg 59 exposure by pregnant women leads to poor neurodevelopment of children 60 (decreasing of fine motor-adaptive function, reduction in IQ, attention deficit 61 disorder) later in life (Kjellstrom et al., 1986, Julshamn et al., 1987, Myers et 62 al., 2003, Deroma et al., 2013). Several studies suggested that MeHg exposure 63 by adults may be associated with an increased risk of cardiovascular deseases, 64 such as hypertension and myocardial infarction (Guallar et al., 2002, Kim et al., 65 2014). Despite the risks of MeHg exposure, fish consumption has many benefits 66 and should not be avoided altogether from human diets (Egeland and Middaugh, 67 1997, Salvo et al., 2016, Cammilleri et al., 2017, Metro et al, 2017). Fish is a 68 valuable source of proteins, vitamin D, selenium, and other essential elements 69 (Fox et al., 2004, Holden et al., 2008, Roos et al., 2007). Fish provide 70 71 significant amounts of polyunsaturated fatty acids, which prevent 72 cardiovascular diseases (Simopoulos, 2008, Mozaffarian and Wu, 2011). Thus it is important to balance the risks and benefits of fish consumption (Sonke et al., 73 2013). 74 Several countries and international organisations have established a 75 reference MeHg dose estimated to be safe or without appreciable risk to health 76

(Tab. 1). Generally, all organisations advise pregnant women, women of

- childbearing age and young children to limit their consumption of fish known to have high levels of Hg (high trophic level fish such as tuna, shark, swordfish) to
- 1-2 per week. Since October 2013, the Minamata Convention under the
- auspices of the United Nations Environment Program (UNEP), a global action
- 82 to protect human health and environment from anthropogenic and mercury
- emissions, was signed (#Web1). The Minamata Convention is implemented
- 84 through the control of specific human activities that contribute to widespread
- Hg pollution. The majority of the European countries ratified the Minamata
- 86 Convention, while others countries, especially from Northern Africa, were not
- yet parties to the convention (#Web1).
- Mercury (Hg) concentrations in hair, urine, tissue, toenails and blood have
- been widely used as biomarkers for human MeHg exposure (for example,
- 90 WHO, 2008, Shao et al., 2013, Bonsignore et al., 2015). About 95% of MeHg in
- 91 fish ingested by humans is absorbed in the human gut (Aberg et al., 1969,
- 92 Miettinen et al., 1971). After absorption into the bloodstream, MeHg enters the
- 93 red blood cells and it is distributed throughout the body within hours (Aberg et
- al., 1969, Clarkson, 1997). Hair incorporates MeHg present in circulating blood
- 95 during hair formation in the hair follicule (Clarkson et al., 1983).
- Approximately 95% of measurable Hg in the blood is in methylated form
- 97 (Sherlock et al., 1984), making MeHg also the dominant species in hair (80%,
- 98 Cernichiari et al., 1995, Phelps et al., 1980). There are several advantages of
- 99 using hair: it is non-invasive, samples are easier to obtain without medical

support, to transport and store. Despite all advantages, hair analysis is suspected of having several limitations, such as: inter-individual variability in the pharmacokinetics of Hg uptake from blood to hair shaft (NRC/NAS, 2000); inter- and intra-individual variability in hair growth rate, density, colour, and waving (Sakamoto et al., 2004, Ohba et al., 2008); external contamination in the areas where ambient air Hg concentrations are elevated.

100

101

102

103

104

105

106

107

108

109

110

111

112

113

114

115

116

117

118

119

120

121

A correlation between consumption of fish or marine products and Hg levels in hair is typically observed (e.g., Castaño et al., 2015, Den Hond et al., 2015). With the funding of the Consortium to Perform Human Biomonitoring on a European Scale (COPHES) and its demonstration project DEMOCOPHES (DEMOnstration of a study to COordinate and Perform Human biomonitoring on a European Scale) 17 European countries provided the baseline information on selected contaminants levels including Hg (Den Hond et al., 2015). However, France did not participate in this project. Thus, the aim of the present study was to provide a first baseline information on the Hg concentrations in human hair at the French Riviera and to complete the European assessment. The main investigations addressed are: (i) the influence of age, gender and fish consumption on Hg concentrations; (ii) a comparison of hair Hg concentrations for women of childbearing age (19 – 45 years old) in French Atlantic vs Mediterranean coastal areas; (iii) a comparison of Hg concentrations in hair of women of childbearing age and annual per capita fish consumption along Mediterranean and some European countries (depending if a country is party of

the Minamata Convention or not); and (iv) we used this assessment to illustrate the mismatch between the MeHg exposure and a lack of commitment or political interest to sign and implement the Minamata Convention.

# 2. Materials and methods

# 2.1. Sampling design

We organized a public awareness campaign: "Arctic Mediterranean Mercury - AM2" for societal Hg exposure in September 2017 (Fig. 3b). During the outreach events we informed the participants about Hg in the environment and exposure. The participants were then given the opportunity to join our study, and to be informed about their individual MeHg exposure level. A total 404 volunteers along the French Mediterranean coastline between Marseille and Villefranche-sur-Mer joined the study (58 % female and 42 % male with ages from 1 to 84 years old). No effort was made in selecting the volunteers according to age, gender, diet, etc.

### 2.2. Data collection

Hair samples were collected according to the Standard Operation Procedure (SOP) established within the framework of COPHES/DEMOCOPHES (Esteban et al., 2015). Briefly, hair strands (~10 – 30 mm) from each volunteers were collected from the nape of the neck, close to the occipital region of the scalp using titanium scissors, wearing powder free single use nitrile gloves. The grey hairs were not removed for the adult group (Pozebon et al., 2017). Given that

hair is a stable material, samples were placed into polyethylene bags until chemical analysis (Pozebon et al., 2017).

A standard questionnaire was used to collect data on the age ("Adults": >18 years old and "Child":  $\leq$  18 years old), gender ("Female", "Male") and frequency of fish consumption ('Almost never':  $\leq$  0.5 fish meals per week, "Once": 1 meal per week, "Twice": 2 meals per week, and 'Often'  $\geq$  3 meals per week). We also subdivided all participants in smaller age groups (1 – 10; 11 – 18; 19 – 24; 25 – 29; 30 – 39; 40 – 49,  $\geq$  50 years old) to compare with Shao et al. (2013). Volunteers gave their consent to participate to the study.

# 2.3. Chemical analysis

Mercury measurements were performed on a Hg analyser (atomic adsorption spectrometer, a Leco-AMA 254). A subsample of approximately 0.03-0.05 g of hair was loaded in nickel boats and placed inside a combustion tube, where the subsample was thermally decomposed (~750°C) into a gaseous form. The evolved gases were then cleaned up from all interfering gases passing through the catalytic compounds. Mercury vapour was then transported to the amalgamator, where all Hg was trapped by gold-plated ceramics. The amalgamator was heated to ~900°C essentially releasing all Hg vapour to the detection system of the atomic absorption spectrometer.

The method's detection limit, estimated as three times the standard deviation of the blank samples, was 0.005 µg g<sup>-1</sup>. The certified reference material (IAEA-086, human hair) was run several times per analytical batch and

constantly before starting the measurements, to check the accuracy of the measurements. The measured values were on average, within  $\pm$  5% of the recommended values. Hair Hg concentrations are given on a fresh weight basis in  $\mu g \ g^{-1}$ .

# 2.4. Statistical analysis

Differences in hair Hg concentrations were tested with non-parametric tests (Shapiro-Wilk and Figner-Killeen tests indicated departures from normality and homogeneity of variances, respectively, except for gender). Independent 2-group Mann-Whitney U tests were used for age groups and gender. A Kruskal Wallis test One Way Anova by Ranks, followed by a Dunn's Multiple Comparisons test, was used to test for differences in hair Hg concentrations among fish consumption groups. P-values were adjusted with the Benjamini & Hochberg method for Multiple Comparisons (alpha = 0.05). Statistical analyses were performed in R version 3.5.0 (The R Foundation for Statistical Computing, 2018).

### 3. Results and discussion

A total of 404 hair samples from participants including children, women of childbearing and non-childbearing age and men have been analysed (Tab. 2). Hair Hg concentrations of participants ranged from 0.01 to 3.98  $\mu$ g g<sup>-1</sup> with a median of 0.51  $\mu$ g g<sup>-1</sup> (interquartile range 0.86 – 0.28  $\mu$ g g<sup>-1</sup>). Only one man and one woman of childbearing age exceeded the French ANSES level of 3.2  $\mu$ g g<sup>-1</sup>,

while about 5% of women and 4% of men exceeded the US EPA level of 1.4  $\mu g$  g<sup>-1</sup>. None of the data exceeded the WHO level of 7  $\mu g$  g<sup>-1</sup>. There was no specific Hg contamination source such as gold mining, coal combustion and fluorescent light factories located around the study area, which should exclude elevation of hair Hg concentration through direct MeHg exposure. The measured hair Hg concentrations may be linked to age, gender or diet of the sampled population (Liu et al., 2008, Schoeman et al., 2009, Castaño et al., 2015, Sunderland et al., 2018).

# 3.1. Influence of age and gender on Hg concentration

Mean Hg concentrations increased significantly with age (Fig. 1). The highest mean concentrations of hair Hg (0.91  $\pm$  0.62 and 0.88  $\pm$  0.66  $\mu g$  g  $^{-1}$ ) were observed in the 40 – 49 and  $\geq$  50 age groups, respectively; the lowest hair Hg concentrations were found on the 11 – 18 age group (0.36  $\pm$  0.34  $\mu g$  g  $^{-1}$ ). Adult group samples (>18 years old) had significantly higher hair Hg concentration than child group samples ( $\leq$  18 years old) (W = 28691, P-value < 2.2e-16). It seems apparent that the body burden of Hg increased with age due to regular accumulation. A similar increase with age was observed by Liu et al. (2008), who determined that the hair Hg of a local population in Southern China increased between their twenties and forties. Shao et al. (2013) also showed increasing of mean Hg concentration for Chinese people living at the coast of up to 49 years of age. Moreover, Liu et al. (2008) and Shao et al. (2013) also noted a gradual decreasing of Hg concentrations in people > 50 years old,

suggesting zero Hg content in the grey hair, which resulted from the role of sulphur-containing chemicals in the formation of hair pigment (Bou-Olayan and Al-Yakoob, 1994; Al-Majed et al., 2000). Unlike Wakisaka et al. (1990), no significant difference in Hg concentrations in hair samples between men and women was observed (W = 21429, P-value = 0.01937).

#### Fish consumption rate and Hg concentrations 3.2.

209

210

211

212

213

214

215

216

217

218

219

220

221

222

223

224

225

226

227

228

229

230

The lowest mean hair Hg concentration  $(0.42 \pm 0.45 \,\mu g \,g^{-1})$ , range:  $0.02 - 10.04 \,\mu g \,g^{-1}$ 1.41 µg g<sup>-1</sup>) was observed on participants which 'Almost never' consume fish, corresponding to 23% of the study participants, intermediate (0.59  $\pm$  0.42 and  $0.87 \pm 0.66 \,\mu g \,g^{-1}$ ) concentrations were measured on participants eating fish meals 'Once' (46% of participants) and 'Twice' (22%) per week, respectively. The highest mean hair Hg concentration  $(1.32 \pm 0.89 \ \mu g^{-1})$  was measured for participants that 'Often' eat fish (9%; Fig. 2a). Hair Hg concentrations were significantly different among all fish consumption groups (Kruskal-Wallis chisquared = 73.169, df = 3, P-value = 8.941e-16; multiple comparisons P-values between 1.49e-3 to 1.34e-13). Previous studies reported that fish consumption rates affect Hg

concentrations in human hair. For example, significantly higher Hg concentrations were found for people living in coastal areas from Sicily and China, who often eat fish (4-7 times per week), compared to those who eat fish less frequently (1 – 2 times per week; Giangrosso et al., 2015, Shao et al., 2013). Miklavčič et al., 2014, also noted that the highest exposure levels to Hg were

found in coastal populations of Europe, which consume more fish than inland populations.

231

232

233

234

235

236

237

238

239

240

241

242

243

244

245

246

247

248

249

250

251

252

Despite a statistically significant difference between hair Hg concentration for children and adults overall (W = 28691, P-value < 2.2e-16), such a difference was not evident between children and adults who consume fish with the same frequency (Fig. 2b). Mean Hg concentrations for children and adults who consume fish 'Almost never' and 'Once' were equivalent, while for a weekly consumption of 'Twice' and 'Often', mean Hg concentrations were lower for children. Such a variation of individual mean Hg concentration has been observed even within families, where children and their mothers with the same fish consumption rate showed different Hg concentrations in hair (Den Hond, 2015). Children are generally characterized by lower Hg concentrations than their mothers, however, Castaño et al. (2015) showed that there is strong correlation between the mother and child in the same family. We are aware that questionnaire answers about individual fish consumption rates do not reflect a precise estimation but rather a "best guess", especially for children. Our broader fish consumption categories were chosen to incorporate some of those uncertainties.

# 4. Regional (and coastal) variations in hair Hg concentration and fish consumption in Europe

Mean Hg concentration in hair for women of childbearing age in the French Mediterranean coast was slightly lower (geometric mean 0.52 µg g<sup>-1</sup>) than in the

French Atlantic coast (Brittany and Nantes, geometric mean 0.62 and 0.67 µg g 253 <sup>1</sup>, respectively, Pichery et al., 2012). Previous studies have shown that fish from 254 Mediterranean waters show higher Hg levels than Atlantic fish. For example, 255 256 bluefin tuna (*Thunnus Thynnus*), which is commonly found in the Mediterranean Sea, was reported to be nearly five times as much concentrated 257 in Hg compared with yellowfin tuna (Thunnus albacares), which is absent in 258 Mediterranean Sea (Cammilleri et al., 2018). Significant higher Hg 259 concentrations in Mediterranean fish might lead to higher MeHg exposure by 260 population. Comparison of hair Hg concentrations of the Mediterranean and 261 Atlantic coasts is not obvious in this study due to the absence of quantitative 262 data about fish consumption by populations in the French Atlantic coast. 263 Geometric mean Hg levels in hair for women of childbearing age in the 264 French Mediterranean coast ('Once' per week - 0.62 µg g<sup>-1</sup>, and 'Often' - 1.91 265 μg g<sup>-1</sup>) were higher than for women of childbearing age with similar fish 266 consumption rates sampled inland (Toulouse, South-West of France; 0.44 µg g<sup>-1</sup> 267 and 1.32 µg g<sup>-1</sup>, respectively; Sonke et al., 2013). The French Mediterranean 268 269 coast compared to inland Toulouse may be related to higher fish consumption from the Mediterranean Sea (Vieira et al., 2015; Faget, 2009), but we cannot 270 draw any conclusion based on our study. 271 There is a widespread difference in MeHg exposure in the European 272 population and the difference is very likely related to consumption of fish and 273 other products from marine environment (Fig. 3a, right panel). Comparing with 274

other contries along the Mediterranean Sea, we found that the Southern France hair Hg concentrations were similar to Hg concentrations found in Italy, Croatia and Albania, but lower than in Spain, Morocco and Greece (Fig. 3c; Miklavčič et al., 2013, Babi et al., 1999, Den Hond et al., 2015, Elhamri et al., 2007, Pérez et al., 2019). According to the European Market Observatory for fisheries and aquaculture products (2017), the annual fish consumption in France is about 34 kg per capita, while it is about 45 kg in Spain. Annual fish consumption in Greece and Morocco is much lower than in France, 17.3 kg and 12.5 kg per capita, respectively. Higher MeHg exposure by Greek and Moroccan populations could be explained by Hg contamination of local fishes. For example, Elhamri et al. (2007) found up to 1.2 µg g<sup>-1</sup> of Hg in hair for women of childbearing age in the Tetouan province, Morocco. This was related to Hg release into the environment from several chloralkali plants which used Hg cell technology. The released Hg contributed to contamination of local fishes and lead to higher MeHg exposure for local populations, including in women of childbearing age. To our knowledge, there is no published data of MeHg exposure levels for African Mediterranean countries, except Morocco (Elhamri et al., 2007) and Egypt (El-baz et al., 2009). These countries do not present high fish consumption rates but we found that there are potential health risks connected with the release of industrial Hg to the environment. For example, the Mercurial

Complex of Azzaba, in Algeria, was an area of very active Hg mining until

275

276

277

278

279

280

281

282

283

284

285

286

287

288

289

290

291

292

293

294

295

2005 (200-500 t y<sup>-1</sup>; Hylander and Meili, 2003). This site alone accounts for 1 million tons of Hg waste (Ministry of land planning, environment and the city, Algeria, 2012). Moreover, the cement industry potentially remains an important source for Hg emissions in Northern African countries (#Web2).

Our data showed that France exhibits greater Hg concentrations in hair for women of childbearing age than other European countries (Fig. 3c) with one exception for Portugal (where Hg concentrations are up to 1.2 µg g<sup>-1</sup>, Den Hond et al.,2015). This trend is in accordance with the higher annual fish consumption per capita in Portugal (56 kg; European Market Observatory for fisheries and aquaculture products, 2017).

We noted that the highest hair Hg concentrations (related to high MeHg exposure) were found along the Mediterranean Sea on countries that are not parties to the Minamata Convention (Fig. 3c). For instance, Spain, which has among the highest annual fish consumption per capita of 45.2 kg, and the highest Hg hair concentration in this study of 1.49 μg g<sup>-1</sup> (Den Hond et al., 2015), and Greece, with and annual fish consumption of 17.3 kg per capita and Hg level of 1.20 μg g<sup>-1</sup> (Miklavčič et al., 2013), signed the convention on 10 October 2013 but did not ratify it. We also noted the lack of data on MeHg exposure levels for Northern African and Middle East countries surrounding the Mediterranean Sea. Taking in account that most of these countries did not ratify the Minamata Convention (Fig. 3c), we propose further research of MeHg exposure for Northern African and Middle East countries.

We hypothesize that populations of non-party countries are not informed about the potential health risks of MeHg exposure. Moreover, published data (Elhamri et al., 2007, Miklavčič et al., 2013) showed that, relatively to the Minamata Convention, non-party countries have elevated MeHg exposure even if they have a low annual fish consumption (i.e., Greece and Morocco). In this case, high MeHg exposure may be explained by Hg industrial contamination of local or imported fishes. Non-party countries are characterized by relatively significant Hg emissions to the atmosphere, according to the Data Visualisation on Global Mercury Emissions by Country and Sector, (2018). For example, Greece accounts for more than 6.5 Ton y<sup>-1</sup>, which is 1.5 Ton y<sup>-1</sup> higher than Hg emissions in France (Fig. 3a, left panel). The implementation of the Minamata Convention can benefit party nations in terms of human health and the environment, by banning Hg from industrial processes, using the best available Hg emission-control technologies and wastewater treatment in new plants. The implementation of the Minamata Convention may lead to some economic losses if we consider the transfer from Hg-based to less efficient or costlier industrial applications. For some of the more industrialized European countries, the impact may be more important. For instance, most of Germany's energy demand is supplied by coal-fired power plants, which have now to be modernized to meet the standards according to the Minamata Convention (UNEP GMA, 2018). However, Germany did implement the Minamata Convention early on, since 15 September 2017. Surprisingly, some of the

319

320

321

322

323

324

325

326

327

328

329

330

331

332

333

334

335

336

337

338

339

countries with low anthropogenic Hg emissions (lower than 1 Ton y<sup>-1</sup>, *e.g.*, Albania and Cyprus, Fig. 3a) are not yet a party the Minimata Convention. The adoption of the Minamata Convention for these countries would have little impact on industry and largely be outweighed by the health benefits and associated economics gains. Pichery et al. (2012) highlighted that prenatal MeHg exposure has serious impacts on the life-time productivity and on society due to adverse cognitive and associated economic consequences.

# 5. Conclusion

Humans are mainly exposed to MeHg when consuming marine fish, in simple words: the more and the bigger fish we eat the more MeHg we take up. In France, in Europe and in most Western countries, protection guidelines have been developed following the Minamata Convention. However, protection guidelines are often too technical, and people tend to either avoid fish completely, at the risk of a polyunsaturated fatty acid-deficiency, or simply not respect them, at the risk of an elevated MeHg exposure.

We examined MeHg exposure of people from the French Mediterranean coast, based on Hg concentrations in their hair. We showed that older people that consumed more fish presented higher hair Hg concentrations. We highlighted that very few study participants presented values above the French reference dose and none above the WHO reference dose. Comparison with Mediterranean bordering and European countries showed that the southern France population is among those with higher MeHg exposure.

# Acknowledgments

363

364

365

366

367

368

369

370

371

372

373

374

375

376

377

378

379

380

381

382

This study was carried out as a part of The AXA Research Fund funded a postdoc fellowship project "Levering Knowledge Gaps To Understand And Anticipate Risk From Methylmercury Exposure". A large part of this project was to investigate Hg exposure in the Arctic region. As part of the project we participated in several oceanographic cruises on the research icebreaker Polarstern to the central Arctic Ocean and Greenland. The AXA Research Fund funded then a follow up outreach, the "AM2" project. With the support of the communication teams of Mediterranean Institute of Oceanography (MIO), Aix-Marseille-University (AMU), Sciences of the Universe Observatory Pytheas (OSU Pytheas) and the French National Center for Scientific Research (CNRS), we shared impressions from those arctic expeditions, in concert with an awareness rising campaign for societal Hg exposure with the broad public during the event of Septembre en Mer 2017. This study forms part of the CONTAM project of MIO. The authors are grateful to the expedition sailing boat François and Sophie Cleuxis; the team of Expé 2M and the ports and the cities of Marseille, Les Embiez Islands, Port-Cros and Porquerolles national Park, Cannes, Saint Tropez, Villefranche-sur-Mer. The project leading to this publication has received funding from European Regional Development Fund under project 1166-39417.

# 383 **6. References**

- 1. Aberg B., Ekman L., Falk R., Greitz U., Persson G., Snihs J.O. (1969)
- Metabolism of methyl mercury (<sup>203</sup>Hg) compounds in man. Archives of
- *Environmental Health: an International Journal.* 19(4), 478-484.
- 2. Al-Majed N.B., Preston M.R. (2000) Factors influencing the total
- mercury and methyl mercury in the hair of the fishermen of Kuwait.
- *Environmental Pollution.* 109(2), 239-250.
- 390 3. Babi D., Vasjari M., Celo V., Koroveshi M. (1999) Some results on Hg
- content in hair in different populations in Albania. Science of the Total
- 392 *Environment*. 259(1-3), 55-60.
- 393 4. Bonsignore M., Andolfi N., Barra M., Madeddu A., Tisano F.,
- Ingallinella V., Castorina M., Sprovieri M. (2015) Assessment of mercury
- exposure in human populations: A status report from Augusta Bay (southern
- 396 Italy). Environmental Research. 150, 592-599.
- 5. Bou- Olayan A.H., Al- Yakoob S.N. (1994) Mercury in human hair: A
- 398 study of residents in Kuwait. Journal of Environmental Science & Health
- 399 *Part A.* 29(8), 1541-1551.
- 6. Cammilleri, G., Vazzana, M., Arizza, V., Giunta, F., Vella, A., Lo Dico,
- 401 G., Giaccone V., Giofrè S. V., Giangrosso G., Cicero N., Ferrantelli, V.
- 402 (2017) Mercury in fish products: what's the best for consumers between
- bluefin tuna and yellowfin tuna? *Natural Product Research*, 32(4), 457-462.
- 7. Castaño A., Cutanda F., Esteban M., Pärt P., Navarro C., Gómez S.,
- Rosado M., López A., López E., Exley K., Schindler B.K., Govarts E.,
- Casteleyn L., Kolossa-Gehring M., Fiddicke U., Koch H.M., Angerer J.,
- Hond E.M., Schoeters G.E., Sepai O., Horvat M., Knudsen L.E., Aerts D.,
- Joas A., Biot P., Joas R., Jiménez-Guerrero J.A., Díaz G.C., Pirard C.,
- Katsonouri A., Černá M., Gutleb A.C., Ligocka D., Reis F.M., Berglund M.,
- Lupsa I., Halzlová K., Charlier C.J., Cullen E.A., Hadjipanayis A., Krsková

- 411 A., Jensen J.F., Nielsen J.K., Schwedler G., Wilhelm M., Rudnai P.,
- Középesy S., Davidson F., Fischer M.E., Janasik B., Namorado S., Gurzău
- 413 A.E., Jajcaj M., Mazej D., Tratnik J.S., Larsson K., Lehmann A.D., Crettaz
- 414 P., Lavranos G., Posada M. (2015). Fish consumption patterns and hair
- mercury levels in children and their mothers in 17 EU countries.
- 416 Environmental Research. 141, 58-68.
- 8. Cernichiari E., Brewer R., Myers G.J., Marsh D.O., Lapham L.W., Cox
- 418 C., Shamlaye C.F., Berlin M., Davidson P.W., Clarkson T.W. (1995)
- 419 Monitoring methylmercury during pregnancy: maternal hair predicts fetal
- brain exposure. *Neurotoxicology*. 16, 705-710.
- 9. Chen M.M., Lopez L., Bhavsar S. P., Sharma S. (2018) What's hot about
- mercury? Examining the influence of climate on mercury levels in Ontario
- top predator fishes. *Environmental Research*. 162, 63-73.
- 10. Clarkson T.W. (1983) Mercury. Annual Review of Public Health. 4, 375-
- 425 380.
- 426 11. Clarkson T.W. (1997) The toxicology of mercury. *Clinical Laboratory*
- 427 *Science Journal*. 34(4), 369-403.
- 12. Cossa D., Harmelin-Vivien M., Mellon-Duval C., Loizeau V., Averty B.,
- 429 Crochet S., Chou L., Cadiou J.F. (2012) Influences of Bioavailability,
- 430 Trophic Position, and Growth on Methylmercury in Hakes (*Merluccius*
- 431 *merluccius*) from Northwestern Mediterranean and Northeastern Atlantic.
- 432 Environmental Science & Technology. 46(9), 4885-4893.
- 13. Den Hond E., Govarts E., Willems H., Smolders R., Casteleyn L.,
- Kolossa-Gehring M., Schwedler G., Seiwert M., Fiddicke U., Castaño A.,
- Esteban M., Angerer J., Koch H.M., Schindler B.K., Sepai O., Exley K.,
- Bloemen L., Horvat M., Knudsen L.E., Joas A., Joas R., Biot P., Aerts D.,
- Koppen G., Katsonouri A., Hadjipanayis A., Krskova A., Maly M., Mørck
- T.A., Rudnai P., Kozepesy S., Mulcahy M., Mannion R., Gutleb A.C.,
- Fischer M.E., Ligocka D., Jakubowski M., Reis M.F., Namorado S.,

- Gurzau A.E., Lupsa I., Halzlova K., Jajcaj M., Mazej D., López A., Lopez
- E., Berglund M., Larsson K., Lehmann A., Crettaz P., Schoeters G. (2015)
- First Steps toward Harmonized Human Biomonitoring in Europe:
- Demonstration Project to Perform Human Biomonitoring on a European
- 444 Scale. *Environ Health Perspect.* 123(3), 255-263.
- 14. Deroma L., Parpinel M., Tognin V., Channoufi L., Tratnik J., Horvat M.,
- Valent F., Barbone F. (2013) Neuropsychological assessment at school-age
- and prenatal low-level exposure to mercury through fish consumption in an
- Italian birth cohort living near a contaminated site. *International Journal of*
- 449 Hygiene and Environmental Health. 216(4), 486-493.
- 450 15. Egeland, G.M., Middaugh, J.P. (1997) Balancing Fish Consumption
- 451 Benefits with Mercury Exposure. *Science*. 278 (5345), 1904-1905.
- 16. El-baz F., Elhossiny R.M., Elsayed A.B., Gaber G.M. (2009) Hair
- 453 mercury measurement in Egyptian autistic children. *The Egyptian Journal of*
- 454 *Medical Human Genetics*. 11(2), 135-141.
- 455 17. Elhamri H., Idrissi L., Coquery M., Azemard S., Abidi A.E., Benlemlih
- 456 M., Saghi M., Cubadda F. (2007) Hair mercury levels in relation to fish
- consumption in a community of the Moroccan Mediterranean coast. *Food*
- 458 *Additives & Contaminants*. 24(11), 1236-1246.
- 18. Esteban M., Schindler B.K., Jiménez J.A., Koch H.M., Angerer J.,
- Rosado M., Gómez S., Casteleyn L., Kolossa-Gehring M., Becker K.,
- Bloemen L., Schoeters G, Den Hond E., Sepai O., Exley K., Horvat M.,
- Knudsen L.E., Joas A., Joas R., Aerts D., Biot P., Borošová D., Davidson F.,
- Dumitrascu I., Fischer M.E., Grander M., Janasik B., Jones K., Kašparová L.,
- Larssen T., Naray M., Nielsen F., Hohenblum P., Pinto R., Pirard C., Plateel
- G., Snoj Tratnik J., Wittsiepe J., Castaño A., EQUAS Reference
- Laboratories. (2015) Mercury analysis in hair: Comparability and quality
- assessment within the transnational COPHES/DEMOCOPHES project.
- 468 Environmental Research. 141, 24-30.

- 19. Faget D. (2009) Le milieu marin méditerranéen: conflits, usages et
- représentations: le cas du golfe de Marseille (début XVIIIe-début XXe
- siècles). Thèse de doctorat, Université Aix Marseille 1, 560p.
- 20. FOOD AND AGRICULTURE ORGANIZATION OF THE UNITED
- NATIONS (FAO). (2012) The state of world fisheries and aquaculture.
- 474 Rome. 209p.
- 21. FOOD AND AGRICULTURE ORGANIZATION OF THE UNITED
- NATIONS (FAO). (2016) The state of world fisheries and aquaculture.
- 477 Contribution to food security and nutrition for all. Rome. 200p.
- 478 22. Fox T.E., Van den Heuvel E.G.H.M., Atherton C.A., Dainty J.R., Lewis
- D.J., Langford N.J., Crews H.M., Luten J.B., Lorentzen M., Sieling F.W., van
- 480 Aken-Schneyder P., Hoek M., Kotterman M.J.J., van Dael P., Fairweather-
- Tait S.J. (2004) Bioavailability of selenium from fish, yeast and selenate: a
- comparative study in humans using stable isotopes. *European Journal of*
- 483 *Clinical Nutrition*. 58(2), 343.
- 484 23. Giangrosso G., Cammilleri G., Macaluso A., Vella A., D'Orazio N.,
- Graci S., Lo Dico G.M., Galvano F., Giangrosso M, and Ferrantelli V.
- 486 (2015) Hair mercury levels detection in fisherman from Sicily (Italy) by ICP-
- 487 MS method after Microwave\_Assisted digestion. Hindawi Publishing
- 488 Corporation. Bioinorganic Chemistry and Applications, Volume 2016,
- 489 Article ID 5408014, 5.
- 490 24. Guallar, E., Sanz-Gallardo, M., Veer, P., Bode, P., Aro, A., Gomez-
- 491 Aracena, J., Kark, J., Riemersma, R., Martin-Moreno, J. M., Kok, F.J. (2002)
- Mercury, Fish Oils, and the Risk of Myocardial Infarction. The New England
- 493 journal of medicine. 347. 1747-54.
- 494 25. Harada M. (1995) Minamata disease: methylmercury poisoning in Japan
- caused by environmental pollution. Critical Reviews in Toxicology. 25(1), 1–
- 496 24.

- 497 26. Holden, J. M., Lemar, L. E., & Exler, J. (2008) Vitamin D in foods:
- development of the US Department of Agriculture database. *The American*
- 499 *Journal of Clinical Nutrition*. 87(4), 1092S-1096S.
- 500 27. Hylander L.D., Meili M. (2003) 500 years of mercury production: global
- annual inventory by region until 2000 and associated emissions. Science of
- *the Total Environment*. 304(13), 13-27.
- 503 28. Julshamn K., Andersen A., Ringdal O., Mørkøre J. (1987) Trace elements
- intake in the Faroe Islands. I. Element levels in edible parts of pilot whales
- 505 (Globicephalus meleanus) Science of the Total Environment. 65, 53-62.
- 506 29. Kjellstrom T., Kennedy P., Wallis S. and Mantell, C. (1986) Physical and
- mental development of children with prenatal exposure to mercury from fish.
- Stage 1 preliminary tests at age 4. National Swedish Environmental Board,
- 509 Solna. Report number 3080.
- 30. Kurland L.T., Faro S.N., Siedler H. (1960) Minamata disease: The
- outbreak of a neurologic disorder in Minamata, Japan, and its relationship to
- the ingestion of seafood contaminated by mercuric com- pounds. *World*
- 513 *Neurology*, 1(5), 370-395.
- 31. Lavoie R.A., Jardine T.D., Chumchal M.M., Kidd K.A., Campbell L.M.,
- 515 (2013) Biomagnification of Mercury in Aquatic Food Webs: A Worldwide
- Meta-Analysis. Environmental Science & Technology. 47(23), 13385-13394.
- 32. Liu X., Cheng J., Yuling S., Honda S., Wang L., Liu Z., Sakamoto M.,
- Liu Y. (2008) Mercury concentration in hair samples from Chinese people in
- coastal cities. *Journal of Environmental Sciences*. 20(10), 1258-1262.
- 33. Metro, D., Tardugno, R., Papa, M., Bisignano, C., Manasseri, L.,
- 521 Calabrese, G., Gervasi T., Dugo G., Cicero, N. (2017) Adherence to the
- Mediterranean diet in a Sicilian student population. *Natural Product*
- 523 Research. 32(15), 1775-1781.
- 34. Miklavčič A., Casetta A., Snoj Tratnik J., Mazej D., Krsnik M., Mariuz
- 525 M., Sofianou K., Spirić Z., Barbone F., Horvat M. (2013) Mercury, arsenic

- and selenium exposure levels in relation to fish consumption in the
- Mediterranean area. *Environmental Research*. 120, 7-17.
- 35. Miklavčič A., Kocman, D., Horvat, M. (2014) Human mercury exposure
- and effects in Europe. *Environmental Toxicology and Chemistry*. 33(6),
- 530 1259-1270.
- 36. Miettinen J.K., Rahola T., Hattula T., Rissanen K., Tillander M. (1971)
- Elimination of <sup>203</sup>Hg-methylmercury in man. *Annals of Clinical Research*, 3,
- 533 116-122.
- 37. Myers G. J., Davidson P. W., Cox C., Shamlaye C., Palumbo D.,
- 535 Cernichiari J., Sloane-Reeves E., Wilding G.E., Kost J., Haung Li-S.,
- Clarkson T.W. (2003) Prenatal methylmercury exposure from ocean fish
- consumption in the Seychelles child development study. *The Lancet*.
- 538 361(9370), 1686-1692.
- 38. Mozaffarian D., Wu J. H. (2011) Omega-3 fatty acids and cardiovascular
- disease: effects on risk factors, molecular pathways, and clinical
- events. *Journal of the American College of Cardiology*. 58(20), 2047-2067.
- 39. National Research Council (NRC). (2000) Toxicological effects of
- methylmercury. Committee on the Toxicological Effects of Methylmercury,
- Board on Environmental Studies and Toxicology, Commission on Life
- Sciences, National Research Council. The National Academies Press.
- Washington, DC. 344p.
- 40. Ohba T., Kurokawa N., Nakai K., Shimada M., Suzuki K., Sugawara N.,
- Kameo S., Satoh C., Satoh H. (2008) Permanent waving does not change
- mercury concentration in the proximal segment of hair close to scalp. *The*
- Tohoku Journal of Experimental Medicine, 214(1), 69-78.
- 41. Ourgaud M., Ruitton S., Bourgogne H., Bustamante P., Churlaud C.,
- Guillou G., Lebreton B., Harmelin-Vivien M. L. (2018) Trace elements in a
- Mediterranean scorpaenid fish: Bioaccumulation processes and spatial
- variations. *Progress in Oceanography*. 163, 184-195.

- 42. Pérez R., Suelves T., Molina Y., Corpas-Burgos F., Yusà V. (2019)
- Biomonitoring of mercury in hair of children living in the Valencian Region
- (Spain). Exposure and risk assessment. *Chemosphere*. 217, 558-566.
- 43. Phelps R.W., Clarkson T.W., Kershaw T.G., Wheatley B. (1980)
- Interrelationships of blood and hair mercury concentrations in a North
- American population exposed to methylmercury. *Archives of Environmental*
- *Health: An International Journal.* 35(3), 161-168.
- 44. Pichery C., Bellanger M., Zmirou-Navier D., Fréry N., Cordier S., Roue-
- LeGall A., Hartemann P., Grandjean P. (2012) Economic evaluation of health
- consequences of prenatal methylmercury exposure in France. *Environmental*
- 565 *Health*. 11(1), 53.
- 45. Pozebon D., Scheffler G., Dressler V. (2017) Elemental Hair analysis: A
- review of procedures and application. *Analytica Chimica Acta*. 992, 1-23.
- 46. Roos N., Wahab M.A., Chamnan C., Thilsted S.H. (2007) The role of fish
- in food-based strategies to combat vitamin A and mineral deficiencies in
- developing countries. *The journal of Nutrition*. 137(4), 1106-1109.
- 47. Sakamoto M., Kubota M., Liu X.J., Murata K., Nakai K., Satoh, H.
- 572 (2004) Maternal and fetal mercury and n-3 polyunsaturated fatty acids as a
- 573 risk and benefit of fish consumption to fetus. *Environmental Science and*
- 574 *Technology*. 38 (14): 3860-3863.
- 48. Salvo A, Cicero N, Vadalà R, Mottese AF, Bua D, Mallamace D,
- Giannetto C, Dugo G. (2016) Toxic and essential metals determination in
- commercial seafood: Paracentrotus lividus by ICP-MS. *Natural Product*
- 578 Research, 30(6):657–664.
- 49. Schoeman K., Bend J. R., Hill J., Nash K., Koren G. (2009) Defining a
- lowest observable adverse effect hair concentrations of mercury for
- neurodevelopmental effects of prenatal methylmercury exposure through
- maternal fish consumption: a systematic review. *Therapeutic Drug*
- 583 *Monitoring.* 31(6), 670-682.

- 50. Shao D., Cheng Z., Kang Y., Wong M. H. (2013) Hair mercury levels
- and food consumption in residents from the Pearl River Delta: South China.
- 586 Food Chemistry. 136(2), 682-688.
- 51. Sherlock J, Hislop J, Newton D, Topping G, Whittle K. (1984) Elevation
- of mercury in human blood from controlled chronic ingestion of
- methylmercury in fish. *Human Toxicology*. 3(2), 117-131.
- 590 52. Simopoulos A.P. (2008) The importance of the omega-6/omega-3 fatty
- acid ratio in cardiovascular disease and other chronic diseases. *Experimental*
- *Biology and Medicine*. 233(6), 674-688.
- 593 53. Sonke J.E., Heimbürger L.E., Dommergue A. (2013) Mercury
- biogeochemistry: Paradigm shifts, outstanding issues and research needs.
- 595 *Comptes Rendus Geoscience*. 345(5-6), 213-224.
- 596 54. Sunderland E.M., Li M., Bullard K. (2018) Decadal Changes in the
- Edible Supply of Seafood and Methylmercury Exposure in the United States.
- *Environmental Health Perspectives*. 126(1), 017006.
- 599 55. UNEP (United Nations Environment Programme) WHO (World Health
- Organization). (2008) Guidance for identifying populations at risk from
- mercury exposure. UNEP DTIE Chemicals Branch and WHO Department of
- Food Safety, Zoonoses and Foodborne Diseases, Inter-Organization
- Programme for the Sound Management of Chemicals (IOMC). Geneva,
- 604 Switzerland, August 2008, DTI/1131/GE. 170p.
- 56. UNEP (United Nations Environment Programme). (2019) Global
- Mercury Assessment -2018. Geneva, Switzerland. 400p.
- 57. Vieira, H.C., Morgado, F., Soares, A. M.V.M., Abreu, S.N. (2015) Fish
- consumption recommendations to conform to current advice in regard to
- mercury intake. Environmental Science and Pollution Research. 22(13),
- 610 9595-9602.

- 58. Wakisaka I., Yanagihashi T., Sato M., Nakano A. (1990) Factors
- contributing to the difference of hair mercury concentrations between the
- sexes. Nihon Eiseigaku Zasshi. Japanese Journal of Hygiene. 45(2), 654-664.
- 59. #Web1: Minamata Convention on Mercury, 2020.
- 615 www.mercuryconvention.org
- 616 60. #Web2: Pro Global Media Ltd, 2020. www.globalcement.com