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1 Mechanisms of cadmium accumulation in plants

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18 Mechanisms of cadmium accumulation in plants

19 Abstract

Cadmium is a non-essential trace metal, which is highly toxic to nearly all living organisms. 20 21 Soil pollution causes Cd contamination of crops, thereby rendering plant products responsible 22 for the chronic low level Cd over-exposure of numerous populations in the world. For this reason, Cd accumulation in plants has been studied for about five decades now. The research 23 first focused on the relationships between plant and soil Cd levels, on the factors of the metal 24 25 availability in soil, as well as the root uptake processes. Cd distribution in plant organs was also investigated, first using a macroscopic and eco-physiological approach, then with the 26 27 help of molecular biology tools, at both tissue and cell scale. Cadmium has no biological function and hijacks the transport pathways of micronutrients such as Fe, Mn or Zn, in order 28 to enter the plant through the roots and be distributed to all its organs. The study of the genes 29 that control the influx and efflux of the Cd²⁺ ion in the cytosol, vacuoles and vascular tissues 30 has significantly contributed to the understanding of the metal root uptake and of its transfer 31 32 to the aerial parts. However, the mechanisms responsible for its distribution to the different 33 aboveground tissues and specially to fruits and seeds have yet to be clarified. This review summarizes current knowledge in order to present a detailed overview of Cd transport and 34 storage, from the rhizosphere to the different organs and tissues of the plant. 35

36

37 Keywords

Ion transporter; cell wall; chelation; phytochelatin; vacuolar sequestration; xylem sap;
phloem; inter-element competition; speciation; organic acid; zinc; iron; silicon; manganese;
nitrogen

41 Cadmium is a transition metal considered as a trace element by geochemists, as its concentration in the upper continental crust - around 0.1 mg kg⁻¹ - is well below the limit 42 between trace and major elements (1000 mg kg⁻¹). Cadmium has no known biological 43 44 function in any organism, except for the marine diatom Thalassiosira weissflogii, in which it can replace zinc as cofactor of carbonic anhydrase (Lane et al., 2005). On the contrary, Cd is 45 highly toxic to most living organisms. The metal exerts its toxic effects by competing with 46 essential metals and calcium, as well as inducing oxidative stress through perturbation of the 47 glutathione redox balance. When internalized, it is retained with a biological half-life >30 48 49 years in the human body. It accumulates in kidneys and bones and can cause osteoporosis, kidney failure, heart disease and cancer (EFSA, 2012). This makes low-level chronic Cd 50 exposure a serious threat to human health. This is the reason why Cd use has been severely 51 52 limited by the European REACH regulation

53 (http://ec.europa.eu/growth/sectors/chemicals/reach_en).

Apart from tobacco smoking, food is the main source of human exposure to Cd, through 54 cereals, vegetables, roots and tubers (EFSA, 2012). In these plant products, Cd comes from 55 the soil, through root absorption and translocation to the harvested organs. Most of the soil Cd 56 57 originates from contamination resulting from human activities. The sources of contamination are atmospheric deposition of industrial and urban emissions, phosphate fertilizers, organic 58 and mineral amendments. In Europe, as atmospheric deposition has decreased since the mid-59 60 1960s, Cd from P fertilizer is now the major input into cultivated soils (Six and Smolders, 2014; Sterckeman et al., 2018). The dominant form of the metal in soil is Cd²⁺. The physico-61 chemical properties of this ion lead to its accumulation in the soil's ploughed layer by binding 62 63 to soil solid components. However, this binding is largely reversible (Smolders et al., 1999), making Cd one of most available trace metals for root absorption after desorption into the soil 64 solution (Lin et al., 2016b). 65

66 Based on the weekly intake limit recommended by the US Agency for Toxic Substances and Disease Registry (ATSDR, 2012) of 0.7 µg Cd per kg of body weight per week, most 67 populations of the world would be largely overexposed to this metal through low-level 68 69 chronic exposure due to food consumption (Clemens et al., 2013). Considering the threshold set by the European Food Safety Authority (EFSA, 2012), which is over 3 fold higher than the 70 one set by ATSDR (2.5 µg Cd kgbw⁻¹ w⁻¹), European populations (2.04 µg Cd kgbw⁻¹ w⁻¹) 71 are on average just below the limit, with some of them being overexposed. 72 73 Cadmium chemistry in soil and its accumulation in plants have been studied since the 74 seventies (e.g. (Lagerwerff, 1971; John et al., 1972; Haghiri, 1973). The aim of the investigations carried out until now has been to understand the soil supply of Cd to roots as 75 76 well as the distribution, the toxicity and the tolerance of the metal in the plant organs. Several 77 works have shown that Cd distribution in the plant varies greatly both from one species to another and within a given species, depending on the population or the cultivar (Florijn and 78 Van Beusichem, 1993b; Greger and Lofstedt, 2004; Gonneau et al., 2014; Laporte et al., 79 2015). This highlights the importance of genetic factors in this distribution. Over the two last 80 decades, numerous publications have addressed the biomolecular processes of Cd 81 82 accumulation in plants, with a particular effort being made to identify the genes and related proteins responsible for Cd uptake, transport, storage and detoxification in plant tissues. 83 An underlying application of these works is the creation of cultivars that accumulate less Cd 84 85 in their consumed parts (Clemens et al., 2013). Another one is the decontamination of soils by phytoextraction, i.e. the cultivation of plants accumulating high levels of Cd in their aerial 86 parts, to the point that they would significantly decrease the soil's Cd content. Phytoextraction 87 88 has been considered with low Cd accumulating and high biomass producing species, such as sunflower (Helianthus annuus), maize (Zea mays), poplar or willow. An alternative is the use 89 of species which hyperaccumulate and are hypertolerant to metals such as Cd, Ni or Zn 90

91 (Chaney *et al.*, 2014; Sterckeman and Puschenreiter, 2018). These wild plants are rare and
92 peculiar, as they are able to grow on metalliferous soils that are toxic to most other species.
93 For this reason, they have been studied extensively to the point of becoming experimental
94 models (Cobbett, 2003; Verbruggen *et al.*, 2009b; Honjo and Kudoh, 2019). These include
95 *Arabidopsis halleri*, *Noccaea caerulescens* (Brassicaceae) or *Sedum alfredii* (Crassulaceae).

We have written this review bearing in mind these underlying applications, which both 96 require an in-depth understanding of the mechanisms controlling Cd flow from soil to plant, 97 in order to be able to either reduce the Cd content in the consumed parts (grain, tubers, leaves) 98 or increase it in the harvested shoots of the plant used to decontaminate the soils. This review 99 100 thus focuses both on the processes responsible for the transfer of Cd from soil to plant and on those responsible for the distribution of the metal in the different plant parts under non-toxic 101 conditions. The mechanisms of Cd toxicity and tolerance are not addressed as such in this 102 103 review, but only with regards to their possible contribution to the flow and the distribution of the metal in the plant. For instance, Cd sequestration in cell walls and in vacuoles is a way to 104 reduce Cd toxicity in plants, but it will be regarded here more as a factor controlling Cd 105 distribution and storage among tissues and organs. The reader may refer to other literature 106 reviews regarding the management of Cd toxicity by plants (Moulis, 2010; Lin and Aarts, 107 108 2012; Verbruggen et al., 2013; Choppala et al., 2014; Kupper and Andresen, 2016).

109

I. Overview of the plant Cd accumulation processes

110 Cadmium exists mainly in the soil as the divalent cation Cd^{2+} . Cadmium concentration as free 111 hydrated ions in soil solutions is in equilibrium with dissolved Cd complexes and Cd 112 adsorbed onto the soil solid phase. By the means of proton excretion, root cells maintain a 113 negative gradient across their plasma membrane, which drives cation absorption, including 114 that of Cd^{2+} . The absorption of Cd^{2+} by root cells reduces the ion concentration at the root

surface, generating a concentration gradient, which in turn drives diffusion of Cd²⁺ towards 115 the root. Cadmium is also transported by advection, where the ion moves towards the root at 116 the same speed as the water taken up by the plant for transpiration. However, in the case of 117 Cd, advection is most frequently negligible when compared to diffusion. In addition, Cd²⁺ 118 desorbed from the solid phase replenishes the soil solution as the root depletes it (Figure 1). 119 Some of the Cd ions in the solution adsorb onto the pecto-cellulosic walls of the root cells 120 (root apoplast). Others diffuse through the apoplast until they meet the first apoplastic barrier, 121 i.e. a domain of the extracellular matrix that prevents the diffusion of ions between cells due 122 123 to deposits of impermeable polymers such as lignin (Casparian strip) or suberin. Depending on the plant species or on the location along the root, the first apoplastic barrier is the 124 exodermis, located just below the root absorptive epidermis in rice for example, or 125 126 endodermis, delimiting the central cylinder where the root vasculature is located (Figures 2-4). From the apoplast, the ions cross the cell membrane (plasma membrane) using the 127 transport proteins. Under a membrane potential of -120 mV, as usually measured in root cells, 128 Cd²⁺ may accumulate passively by 4 orders of magnitude. Therefore, active transport is not 129 required to drive Cd²⁺ uptake into root cells. Once in the cytosol, Cd ions are complexed by 130 organic ligands. This reduces the concentration of free Cd²⁺ in the cytosol and therefore the 131 metal toxicity. Some of the Cd in the cytosol diffuses towards the xylem, following the 132 133 plasmodesmata to pass from one cell to another to reach the xylem parenchyma (Figures 2-4). 134 The metal can then be loaded into the xylem vessels by the way of an active transport system, against its electrochemical gradient. Other Cd compounds in the cytosol are distributed to 135 intracellular organelles such as the mitochondria, the Golgi apparatus and mainly the vacuole, 136 137 through transport proteins. In the vacuole lumen, Cd is mostly chelated by organic ligands. It can therefore be considered that once Cd^{2+} is absorbed into the root symplast, its 138 translocation to the aerial parts is governed by three main processes: sequestration in the root 139

140 cells, transport in the symplast to the stele and xylem loading. The competition between these three processes leads most higher plants to accumulate much more Cd in their roots than in 141 their shoots. Nonetheless, some species, such as hyperaccumulating plants, accumulate a 142 higher concentration of the metal in their above ground parts than in their roots. 143

Once in the xylem sap, Cd^{2+} eventually forms complexes with ligands. These complexes, 144 together with free ions, are transported to the plant aerial parts with the xylem sap flow being 145 driven by transpiration (advection). During this transport, the ions interact with the cell walls 146 of xylem vessels and partly adsorb on them. They also migrate through xylem cell walls to 147 those of the stem and leave cells, under the influence of the transpiration flow. From the 148 apoplast, Cd²⁺ crosses cell membranes *via* transport proteins (Figure 5). Once in the leaf cell 149 cytosol, the metal is generally chelated by organic ligands and migrates from cell to cell. 150 151 Some of the free ions and of the complexes enter the organelles and particularly the vacuole, where they are stored as chelates. In the shoot, Cd^{2+} is finally distributed in variable amounts 152 according to the cell compartment (wall, cytosol, vacuoles, etc.) and to the tissue (vascular 153 tissues, parenchyma, epidermis...), with a speciation differing with the location and with the 154 plant species. 155

Although less documented, it is likely that part of the Cd accumulated in the leaves is 156 redistributed to other plant organs *via* the phloem, or even to the roots where Cd could be 157 excreted. The phloem flow also allows Cd transport to the reproductive organs and its 158 accumulation in the seeds. Some works suggest that phloem could be involved in Cd transport 159 before it reaches the leaves, thanks to exchanges from xylem to phloem in the stem. 160

II. 161

Chemistry of Cd in the rhizosphere

Plant roots take up Cd from the surrounding soil solution. The chemistry of Cd in the 162 rhizosphere can usefully be considered as the reactions that control the species and the 163

164 concentrations of the metal in solution. These reactions take place in the soil solution and at the surface of the solid phase (Figure 1). With an ionic potential (IP = Z/r, the valence divided 165 by the ionic radius in nm) of 21 nm⁻¹, Cd²⁺ is a solvating cation, whose behaviour can be 166 differentiated from that of hydrolysing cations, such as Al^{3+} ($IP = 56 \text{ nm}^{-1}$), and oxyanion-167 forming cations, such as Mo^{6+} ($IP = 143 \text{ nm}^{-1}$) (Sposito, 1989). Based on this parameter, Cd^{2+} 168 behaves similarly to Ca^{2+} ($IP = 20 \text{ nm}^{-1}$). However, it is differentiated from this cation by its 169 Misono softness parameter (Y, nm), whose value is 0.31 nm (Misono et al., 1967; Sposito, 170 1989). Values of Y above 0.32 nm correspond to metal ions with high polarizability and low 171 172 electronegativity, which tend to form covalent chemical bonds, while values below 0.25 nm correspond to ions with low polarizability and high electronegativity tending to form ionic 173 bonds. Cadmium, together with Cu²⁺, Mn²⁺, Ni²⁺ and Fe²⁺ fall into the "borderline" zone 174 (0.25 < Y < 0.32), while Ca²⁺ has a Y value of 0.16 nm and Hg and Pb have Y values of 0.40 175 nm. This makes Cd²⁺ chemical behaviour intermediate between that of Ca²⁺, forming ionic 176 bonds, and those of Pb^{2+} and Hg^{2+} , often involved in covalent bonds. 177 In the solution of non- or low contaminated soils, Cd concentration is generally below 10 nM, 178 while in Cd polluted soils, it ranges between 10 nM to 200 nM, and can even reach 179 180 micromolar concentrations under extreme contamination and solubility conditions (Helmke, 1999; Sauvé et al., 2000b; Weng et al., 2002; Nolan et al., 2003; Maxted et al., 2007). The 181 free hydrated ion (Cd^{2+}) is in equilibrium with organic and inorganic complexes. Cd^{2+} reacts 182 with water to form Cd(OH)₂, Cd(OH)⁺ and Cd(OH)₃⁻ (Gu et al., 2010). In the presence of 183 chloride, particularly in salt affected soils, Cd²⁺ can form soluble chloride complexes, i.e. 184 CdCl⁺, CdCl₂, CdCl₃⁻ and CdCl₄²⁻. Cd²⁺ reacts with sulphate, to form CdSO₄ complexes and, 185 in alkaline soil, with carbon dioxide to form CdCO₃ and CdHCO₃⁺ in solution (Alloway, 186 1995; Helmke, 1999). Free Cd²⁺ is considered to dominate in soil solutions (Degryse et al., 187 2009; Ren et al., 2015a), although some measurements and modelling have shown that this is 188

189 not always the case (Schneider, 2008; Lin et al., 2016b). Indeed, the free hydrated ion can associate to a large extent with soluble humic compounds (Schneider, 2008; Ren et al., 2015a; 190 Ren et al., 2015b), as well as with well-defined molecules such as organic acids (Schneider 191 and Nguyen, 2011). For instance, Lackovic et al. (2004) proposed several reactions to model 192 Cd complexation by citrate (noted L): 193 $Cd^{2+} + I^{3-} + U^{+} \leftarrow CdI U^{0} (loc K - 7.85)$ 10/ (1)

$$194 \quad Cd^{-2} + L^{-2} + H^{-2} \rightarrow CdLH^{-2} (logK = 7.85) \tag{1}$$

- $Cd^{2+} + L^{3-} CdL^{-} (\log K = 3.71)$ 195 (2)
- $Cd^{2+} + 2L^{3-} CdL_2^{4-} (\log K = 5.3)$ 196 (3)
- $Cd^{2+} + L^{3-} + H_2O \leftrightarrows CdOHL^{2-} + H^+ (logK = -3.4)$ 197 (4)

Cadmium competes with other cations, such as Ca²⁺for complexation with soluble organic 198 ligands (Lin et al., 2016b). 199

In most soils, the amount of dissolved Cd is much lower than that bound to the solid phase. 200 201 Cadmium can precipitate as carbonate or phosphate, but soils are generally under-saturated 202 with these minerals (Tye et al., 2003). Even if super-saturated, dissolved organic matter has been reported to inhibit CdCO₃ precipitation (Holm et al., 1996). Therefore, rather than 203 precipitation, adsorption on mineral surfaces and binding to humic substances are the 204 processes which control Cd partitioning between solid and liquid phases (Lofts and Tipping, 205 1998; Tipping, 1998; Ren et al., 2015b). Indeed, Cd can bind with particulate fulvic or humic 206 acid (R) with a net charge Z, exchanging against protons according to the following equation: 207 $RH^{Z} + Cd^{2+} \leftrightarrows RCd^{Z-1+2} + H^{+}$

(5)

208

209

hydroxides and clay minerals. Surface complexation reactions on variable charged sites 210

The metal also interacts with the surface sites on minerals, particularly on Fe and Mn

(≡SOH), such as those at the edges of clay minerals or of iron hydroxides can be formalized
by:

$$= \text{SOH}^0 + \text{Cd}^{2+} \leftrightarrows = \text{SOCd}^+ + \text{H}^+ [\log K = -3.23 \text{ for kaolinite, (Gu and Evans, 2008)}]$$
(6)

while adsorption on permanent negatively charged sites ($\equiv X^{-}$) is

215
$$\equiv 2X^{-}H^{+} + Cd^{2+} \Leftrightarrow \equiv X_{2}^{-}Cd^{2+} + 2H^{+} [logK = 1.22 \text{ for kaolinite, (Gu and Evans, 2008)]. (7)}$$

However, as the soil matrix is very complex, with numerous constituents having each different types of surface reactive sites, Cd adsorption is generally modelled using the values of the distribution coefficient, K_d (L. kg⁻¹), which is the ratio of Cd reversibly adsorbed on the solid phase divided by metal in solution. K_d ranges from several tens to several thousands (Sauvé *et al.*, 2000a; Sauvé *et al.*, 2000b). K_d is generally well predicted by a function of soil organic carbon and pH (Sauvé *et al.*, 2000a; de Meeûs *et al.*, 2002; Degryse *et al.*, 2009), such that

223
$$\log(K_d) = -1.04 + 0.55 p H_{CaCl2} + 0.70 \log(C_{org}) \ (n = 123, R^2 = 0.72)$$
 (8)

proposed by Degryse et al. (2009), C_{org} being the organic carbon content (%) and pH_{CaCl2} the 224 pH of a soil suspension in a 0.01 M CaCl₂ solution. This relationship supports the idea that 225 organic carbon is the most important reactive fraction of the soil solid phase (Tye et al., 2003; 226 Ren et al., 2015b) and that Cd sorption increases with organic carbon content of the solid 227 phase. In addition, the solubility of Cd associated to Fe oxides is affected by the soil redox 228 potential (Yu et al., 2016). Moreover, in carbonated soils, Cd can sorb at the surface of 229 calcite, with Cd²⁺ replacing Ca²⁺. This could lead to an irreversible absorption, as Cd diffuses 230 231 through the calcite and makes a solid solution of otavite-calcite (Martin-Garin et al., 2003). As with complexation in solution, Cd sorption is a reversible and competitive process, which 232

233 depends on the various cations in solution together with Cd^{2+} . Eqs. 5-8 indicate that protons

compete with Cd²⁺ both for adsorption on the soil minerals and for binding with humic 234 substances phase. Various works have evidenced that Cd sorption increases with pH 235 (Christensen, 1984; Sauvé *et al.*, 2000a). Eq. 8 shows that sorption at pH = 5.0 is about ten 236 times lower than at pH = 7.0. Competition with other cations, in particular with Ca^{2+} , also 237 inhibits Cd sorption (Christensen, 1984; Alloway, 1995). 238 Although mostly sorbed on the solid phase, its physico-chemical characteristics make Cd 239 more mobile from solid to liquid and therefore more available for root absorption than most 240 other trace metals. When applied to variably contaminated surface horizons, isotopic dilution 241 showed that on average, 61% (9% to 92%) of Cd was isotopically exchangeable, i.e. in 242 equilibrium with the solution content and likely to dissolve in case of the depletion of the 243 solution (Degryse et al., 2004). In another study, it was shown that 46% of Cd in cultivated 244 245 soils was isotopically exchangeable (Sterckeman et al., 2009).

As a consequence of its chemistry, Cd availability for absorption by roots increases with the 246 soil Cd content and acidity, as well as with the Ca²⁺ in soil solution, and decreases when soil 247 sorption capacity, in particular soil organic matter, increases (Lin et al., 2016b). 248

III. 249

Cadmium species absorbed by the roots

Most studies suggest or consider that the main absorbed Cd species is Cd^{2+} , i.e. the free ion 250 (Welch and Norvell, 1999). Custos et al. (2014) showed that when reducing the concentration 251 252 of Cd²⁺ in nutrient solution by chelating it by EDTA, the absorption of the metal by maize in hydroponics was strongly decreased. However, Cd complexes (or chelates) can contribute to 253 root uptake in two ways. Firstly, it has been reported that organo-metallic complexes can 254 dissociate in the vicinity of the root surface, as the solution is depleted in free ions due to 255 absorption by the root (Degryse et al., 2006; Panfili et al., 2009; Wang et al., 2009). 256 Therefore, the pool of complexed Cd contributes to the buffering of the free Cd pool. 257

258	However, this contribution was shown to be very small. Through modelling, Lin et al.
259	(2016b) estimated that in most cases, the dissociation of Cd complexes is slower than
260	desorption from the solid phase and not fast enough to supply the roots with Cd ²⁺ .
261	The other contribution of the complexed Cd would be through direct absorption. However,
262	this seems variable according to the experimental conditions and has never been directly
263	demonstrated. According to Panfili et al. (2009), durum wheat (Triticum turgidum L. ssp.
264	durum) is not able to absorb Cd-citrate complexes, in opposition to what Berkelaar and Hale
265	(2003a, 2003b) suggested. However, different experiments have shown the presence of Cd-,
266	Pb- and Zn-EDTA complexes in plant tissues, suggesting that they may be absorbed by roots
267	(Sarret et al., 2001; Collins et al., 2002; Schaider et al., 2006). The uptake of Cd by maize in
268	hydroponics being greater than that predicted by a model based on the sole absorption of
269	Cd ²⁺ , Custos et al. (2014) also suggested that Cd-EDTA was absorbed from the nutrient
270	solution. Blaylock et al. (1997) showed that an addition of EDTA and other
271	aminopolycarboxylates like EGTA or DTPA to soil, increases the accumulation of Cd in
272	Indian mustard, although it is not excluded that this could be the consequence of the
273	damaging of the endodermis by the chelates (Nowack et al., 2006). Cadmium complexes
274	would enter the plant through the apoplast at sites lacking apoplastic barriers, i.e. at the root
275	apex (Redjala et al., 2011) and at the emergence of lateral roots. The driver of the complex
276	uptake would be the water flow caused by the plant transpiration (Nowack et al., 2006; Custos
277	et al., 2014), although it is not clear how the complexes would cross cell membranes.
278	Some studies have shown a positive effect of chloride on Cd uptake, in both soil and
279	hydroponics (Smolders and McLaughlin, 1996a, 1996b; McLaughlin et al., 1997; Weggler et
280	<i>al.</i> , 2004). Two mechanisms could account for this effect: (i) the formation of $CdCl_n^{2-n}$
281	complexes could favour the desorption of the metal from the soil solid phase, (ii) the
282	complexes $(CdCl_n^{2-n})$ could diffuse faster than the free ion in the rhizosphere or in the

apoplast. Absorption of the complexes themselves is not excluded, as $CdCl^+$ could cross the plasma membrane through selective ion channels for monovalent cations such as K^+ . In addition, the neutral form $CdCl_2^0$ could passively cross the lipid bilayer (Welch and Norvell, 1999).

Analogous effects, but less intense than with Cl^- , have been obtained with SO_4^{2-} . The Cd concentration in the soil solution increases upon addition of sulphate, which slightly increases Cd absorption (McLaughlin *et al.*, 1998b). An experiment in hydroponics suggests that the CdSO₄⁰ complex is taken up by plants as efficiently as free Cd²⁺ ion (McLaughlin *et al.*, 1998a). The electrically neutral CdSO₄⁰ complex would passively diffuse across the plasma membrane.

293 IV. Influence of the plant on soil Cd availability

In theory, roots can influence the concentration and speciation of Cd in the rhizosphere and 294 consequently their ability to take up the metal. Indeed, according to the charge balance 295 resulting from the absorption of the nutritive ions, the roots compensate the imbalance by 296 excreting protons (H⁺) or hydroxyls (OH⁻) (Hinsinger et al., 2003). In the first case, the 297 rhizosphere acidification would favour the desorption of Cd²⁺ from the soil solid phase while 298 in the second, the alkalization would increase the sorption of the metal (Eq. 8). Acidification 299 of the rhizosphere should therefore increase root Cd uptake, in contrast to alkalization. 300 However, this effect could only rarely be demonstrated in the case of Cd. Wu (1989) found 301 that nitrogen fertilization of ryegrass using NH₄⁺ increased plant Cd uptake, compared with 302 fertilization using nitrate. He attributed this observation to the increase in soil acidity 303 304 following the higher proton release by the roots absorbing ammonium.

- The excretion of organic ligands such as organic acids (Jones *et al.*, 2009) or
- 306 phytosiderophores could also cause changes in Cd availability in the rhizosphere. Using

307 isotopic dilution, Sterckeman et al. (2005) found an increase in the availability of Cd in the rhizosphere (measured by L value) compared to bulk soil (represented by the E value). This 308 was the case in an alkaline soil but not in a slightly acidic one. In the alkaline soil, the pool of 309 310 isotopically exchangeable Cd (i.e. Cd in solution and on the solid phase in equilibrium with that in solution) was greater in the rhizosphere than that in the bulk soil. Some reactions, due 311 to changes in the chemical composition of the rhizosphere soil might therefore have modified 312 the Cd speciation. However, these reactions as well as their origin (roots or associated 313 microflora) are unknown. The increase in Cd availability varied according to the six plant 314 315 species tested. It was maximal for ryegrass and nil for the hyperaccumulator N. caerulescens, in agreement with the results reported by Hutchinson et al. (2000) and Zhao et al. (2001) who 316 also showed that root exudates of N. caerulescens did not increase desorption of Cd and Zn. 317 318 Organic acids can increase the solubility of soil Cd, displacing ions adsorbed on the solid phase into the solution as they complex the dissolved free ions (Naidu and Harter, 1998). 319 Whether they favour Cd uptake (Cieslinski et al., 1998) or decrease it through the formation 320 of Cd complexes which are not absorbable by the roots (Pinto et al., 2008) is a matter of 321 debate (as discussed above). Oxalate excreted by tomato roots in response to Cd exposure has 322 323 been shown to reduce the metal toxicity, probably because Cd oxalate complexes are excluded from the roots (Zhu et al., 2011). In contrast, Tao et al. (2015) suggested that the 324 325 strong excretion of oxalate at the root apex of the Cd hyperaccumulating accession of S. 326 alfredii may favour Cd uptake and accumulation. The mechanisms underlying this 327 observation are however, not clearly identified. They could involve the dissociation of the complex at the root surface, a direct uptake of the chelate through specific transport systems 328 329 or an apoplastic absorption of the chelates.

330 Phytosiderophores, such as mugineic and deoxy mugineic acids (MA, DMA), are iron-

331 complexing agents synthetized from nicotianamine and secreted by grasses suffering from a

332 Fe deficiency (Shojima et al., 1990). Experiments with rice (Oryza sativa) and maize suggest that phytosiderophores may favour the availability and uptake of Zn (Zhang et al., 1989; von 333 Wirén et al., 1996; Arnold et al., 2010; Ptashnyk et al., 2011). However, their effect on Cd 334 335 uptake has not been completely clarified. Puschenreiter et al. (2017) observed enhanced Zn, Cu, Mn and Ni concentrations in wheat shoots on some soils associated with 336 phytosiderophore exudation in Fe-deficient wheat (Triticum aestivum). However, plant Cd 337 concentration did not increase. Accordingly, an addition of DMA to a soil suspension did not 338 increase the Cd concentration in the solution. In contrast, the work of Shenker et al. (2001) 339 indicated that phytosiderophores are able to increase Cd desorption from the soil solid phase. 340 However, as in the work of Puschenreiter et al. (2017), they did not cause an increase in metal 341 uptake by wheat or barley (Hordeum vulgare). With maize cultivated in hydroponics, Hill et 342 al. (2002) showed that Cd stress, as with Fe deficiency, causes greater DMA production by 343 roots, while DMA appears to limit Cd accumulation. These results suggest that 344 phytosiderophores can form complexes with Cd which may contribute to the desorption of Cd 345 346 from the solid phase of the soil, but prevent its internalization by roots. Meda et al. (2007) partly contradicted this model, showing that Cd-induced phytosiderophore release cannot 347 protect maize plants from Cd toxicity. Instead, phytosiderophore-mediated Fe acquisition 348 would provide an advantage under Cd stress by improving Fe uptake. Accordingly, Astolfi et 349 al. (2014) showed that most of the genes involved in phytosiderophore biosynthesis and 350 351 secretion induced by Fe deprivation were also significantly upregulated in the presence of Cd under Fe sufficient conditions in barley. Interestingly, the enhanced expression of 352 phytosiderophore biosynthesis genes in roots was associated with the downregulation of Cd 353 uptake transporter genes (see below). The downregulation of Cd uptake transporter expression 354 would constitute a protective mechanism against further Cd uptake by these transporters. 355

356

V. Absorption of Cd by the roots

357

A. Location of Cd uptake on the root

Laporte et al. (2013) showed that in sunflower first order lateral roots, Cd²⁺ influx was 2.9 358 times higher at root tip than at the root base. These results are consistent with those of Pineros 359 et al. (1998), who observed significantly higher Cd influx at positions 1 to 1.5 mm from the 360 root apex of wheat and N. caerulescens seedlings. Similar profiles were observed in various 361 young roots, in durum wheat (T. turgidum L. ssp. durum) (Farrell et al., 2005), in poplar 362 (Populus × canescens plantlets) (He et al., 2011), in rice (Chen et al., 2018b) and in the 363 halophyte Suaeda salsa (Li et al., 2012a). These profiles can be attributed to a higher activity 364 of the transport systems close to the root tip, where they could be expressed at higher levels 365 366 (Chen et al., 2018b) and/or benefit from more energy (Laporte et al., 2013). Higher Cd influx at the root apex could also result from the incomplete development of apoplastic barriers, 367 thereby favouring apoplastic Cd uptake (Redjala et al., 2011) in this part of the root. The 368 Casparian strips and suberin lamellae in the endodermis and exodermis are formed away from 369 the root apex in more differentiated parts of the roots. However, according to the modelling 370 371 carried out by Laporte et al. (2013), the higher influx at the root tip would not significantly influence the overall Cd root uptake, whereas the total root surface area is indeed a major 372 parameter. 373

374

B. Mechanisms of Cd root absorption

There is little evidence of a biological role for cadmium. Its involvement in the enzymatic system of a diatom was mentioned above (Lane *et al.*, 2005). As in the case of Zn and Ni, Cd accumulation has been proposed as a defence against herbivorsy of hyperaccumulating plants (Jiang *et al.*, 2005; Plaza *et al.*, 2015). It is therefore very unlikely that transport systems with specificity for a non-essential and even highly toxic metal were selected during evolution. Instead, Cd is probably taken up through the plasma membrane of root cells by transporters for essential elements with insufficient selectivity with respect to their substrate (Clemens,2001).

383	Cadmium transport into roots was first studied through absorption kinetics, formalized using
384	the Michaelis-Menten model (Cataldo et al., 1983; Costa and Morel, 1994; Cohen et al.,
385	1998; Hart et al., 1998; Redjala et al., 2010a). Although considered as not Cd specific, the
386	transport systems involved show a high affinity to the metal, with Km falling in the range of
387	0.02 to 1 μ M. For instance, the K_m values for Cd were 0.059 μ M and 0.067 μ M in bread
388	wheat and durum wheat, respectively, whereas for Zn, a micro-nutrient, they were 2.3 μ M and
389	3.3 μ M (Hart <i>et al.</i> , 2002). The maximum influx (V_{max}) values ranged from 12 to 188 nmol
390	(gFW) ⁻¹ h ⁻¹ (Lux et al., 2011). The highest values were found for the Ganges populations of
391	the metal hyperaccumulator N. caerulescens under Fe deficient conditions (Lombi et al.,
392	2002). Cohen et al. (1998) also found that Fe deficiency stimulated high affinity Cd uptake in
393	<i>Pisum sativum</i> roots. Lombi <i>et al.</i> (2001) measured the K_m and V_{max} of the Prayon (low Cd
394	accumulator) and Ganges (high Cd accumulators) populations of N. caerulescens. They found
395	identical K_m values (around 0.2 μ M), but the V_{max} of the Ganges population (160 nmol (gFW) ⁻
396	¹ h ⁻¹) was significantly higher than that of the Prayon population (33 nmol (gFW) ⁻¹ h ⁻¹).
397	Similar V_{max} differences were found in the comparison between Calamine (Viviez) and
398	Serpentine (Vosges) populations of the same species (Redjala <i>et al.</i> , 2009). Since the V_{max}
399	value is proportional to the density of membrane transporters, the difference in Cd uptake
400	ability by the Prayon and Ganges populations could be due to a difference in the expression of
401	a certain transporter gene with high Cd affinity.

402 Absorption kinetics are often biphasic in appearance, resulting from the combination of

403 Michaelis-Menten kinetics and linear kinetics. The linear component has often been attributed

to adsorption of Cd^{2+} on the apoplasm and therefore to incomplete desorption before the

determination of Cd in the root (Hart et al., 1998; Lombi et al., 2001). However, this

406 interpretation was invalidated using Cd measurement in the symplasm and apoplasm [Redjala 407 (2009); Redjala *et al.* (2010a)]. The linear component would rather correspond to the entry of 408 Cd into the symplasm *via* low affinity transport systems, such as ion channels for Ca²⁺ 409 (Perfus-Barbeoch *et al.*, 2002; Rodriguez-Hernandez *et al.*, 2015; Chen *et al.*, 2018b). In any 410 case, this linear component significantly contributes to the root uptake only at high 411 concentrations of Cd²⁺ in the medium, which do not occur in agricultural soil and are rare in 412 contaminated soils.

413 In mouse-ear cress (*Arabidopsis thaliana*) (Figure 2, Table 1), Eide *et al.* (1996) showed that

414 Fe deficiency induces the expression of an iron regulated transporter IRT1 (from the zinc-

415 regulated transporter, iron-regulated transporter protein - ZIP family), which allows the

416 transport of divalent metals such as Cd^{2+} , Co^{2+} , Mn^{2+} and Zn^{2+} , in addition to Fe^{2+} , when

417 expressed in yeast (Eide et al., 1996; Korshunova et al., 1999). AtIRT1 is mainly expressed in

418 root epidermis and to a lesser extent in root cortex (Vert *et al.*, 2002). Cd accumulation is

419 strongly reduced in an *irt1* knockout mutant under Fe deficient conditions (Vert *et al.*, 2002).

420 Cohen *et al.* (1998) proposed that the up-regulation of an IRT1 orthologue accounts for

421 elevated Cd uptake under Fe deficiency in *Pisum sativum*. Even though the IRT1 homologue

422 in barley is also able to transport Fe^{2+} , Zn^{2+} , and Cd^{2+} when expressed in yeast (Pedas *et al.*

423 (2008); (Wu et al., 2016a), in planta, HvIRT1 mainly controls Mn²⁺ uptake (Long et al.,

424 2018). In rice (Figure 3, Table 1), OsIRT1, as well as its homologue OsIRT2, are

425 predominantly expressed in roots in Fe deficient conditions and are able to transport Cd^{2+}

426 when expressed in yeast. Iron deficiency both enhanced their expression as well as Cd uptake

427 and translocation (Ishimaru et al., 2006; Nakanishi et al., 2006). IRT1 may also be involved

- 428 in Cd uptake by the hyperaccumulator *N. caerulescens* (Lombi *et al.*, 2002; Plaza *et al.*,
- 429 2007), but whether its activity accounts for the particularly high ability of the Ganges
- 430 population to accumulate Cd remains to be formally demonstrated. In this species (Figure 4),

NcZNT1, the orthologue of AtZIP4, was presented as mediating high-affinity Zn^{2+} uptake as 431 well as low affinity Cd²⁺ uptake (Pence et al., 2000)(Table1). However, this finding was 432 contradicted by Milner et al. (2012) who reported that NcZNT1 can transport Zn, but not Cd 433 when expressed in yeast or A. thaliana. More recently, the results of Lin et al. (2016a) 434 confirmed those of Pence et al. (2000). They explained the discrepancy by the fact that the 435 NcZNT1 cDNA sequence used by Milner et al. (2012) lacked the original ATG start codon, 436 which would have affected the normal function and cellular localization of the aberrant 437 NcZNT1 protein. 438

439 Transport proteins of the NRAMP (natural resistance-associated macrophage protein) family

440 are also involved in the Cd uptake and more generally in its transport in plants. AtNRAMP

441 1/3/4 transport Fe²⁺, Mn²⁺ and Cd²⁺ when expressed in yeast (Thomine *et al.*, 2000).

442 AtNRAMP3/4 are involved in Fe, Mn and Cd efflux from the vacuole (Lanquar *et al.*, 2005;

Lanquar et al., 2010; Pottier et al., 2015). AtNRAMP1 is responsible for high-affinity Mn

444 uptake by *A. thaliana* roots (Cailliatte *et al.*, 2010). Its role in Cd uptake *in planta* has yet to

be investigated. In rice, OsNRAMP1, which transports Fe, was suggested to contribute to Cd

446 uptake and transport within rice plants, but did not transport Mn (Takahashi et al., 2011)

447 (Table 1), while NRAMP1 from *Malus baccata* may be involved in Mn, as well as for Fe and

448 Cd transport (Xiao *et al.*, 2008). OsNRAMP5 turned out to be the major transporter for the

449 uptake of Mn and Cd in rice roots (Ishikawa et al., 2012; Ishimaru et al., 2012; Sasaki et al.,

450 2012; Tang *et al.*, 2017) (Table 1). Ishimaru *et al.* (2012) reported that OsNRAMP5 is

451 expressed in the epidermis, exodermis, and outer layers of the cortex as well as in tissues

452 around the xylem. Sasaki *et al.* (2012) detected the protein only in exodermis and endodermis.

453 OsNRAMP5 loss of function mutants are defective when it comes to high affinity Cd uptake

454 into the roots (Sasaki et al., 2012) and in Cd translocation from roots to shoots (Yang *et al.*,

455 2014). Consequently, these mutants accumulate Cd at extremely low levels in all organs

including grain. In *Triticum polonicum*, TpNRAMP5 is mainly expressed in roots and basal
stems, and the protein is localized in the plasma membrane. Expression of TpNRAMP5 in *A*. *thaliana* enhanced the accumulation of Cd, Co and Mn, but not that of Zn and Fe (Peng *et al.*,
2018). The NRAMP5 transporter is also present in maize and in bread wheat, but displays a
lower level of expression and a lower activity than in rice (Sui *et al.*, 2018).

461 Several root metal uptake transporters display a polar localization. OsNRAMP5 and AtIRT1,

that have been show to mediate Cd uptake are concentrated on the membrane facing the soil

463 solution (Sasaki *et al.*, 2012; Barberon *et al.*, 2014). This raises the question of how metals are

transferred towards the stele, from the other side of the cell facing the central cylinder. In

465 plants, members of the cation diffusion facilitator (CDF) family are known as metal tolerance

466 proteins (MTP). They are involved in metal efflux from the cytosol to other compartments:

the apoplasm, the vacuole and the Trans Golgi Network (TGN). In rice, Ueno *et al.* (2015)

showed that OsMTP9 is polarized on the side of the plasma membrane facing the stele and is

469 required for efficient Mn uptake (Table 1). Working on cucumber (*Cucumis sativus*),

470 Migocka *et al.* (2015) found that the CsMTP9 gene encodes a plasma membrane H^+ -coupled 471 Mn^{2+} and a Cd²⁺ antiporter expressed exclusively in roots. CsMTP9 was shown to be involved 472 in the efflux of Mn and Cd from root cells, thereby allowing the transport of both metals from 473 the endodermis into the vascular cylinder for further translocation to the shoot.

474 Some authors have suggested that Cd could also enter cells through Ca^{2+} channels (Welch and

Norvell, 1999; Perfus-Barbeoch *et al.*, 2002; Li *et al.*, 2012b). This is supported both by the

476 finding that Cd influx is affected by Ca^{2+} channels blockers (Cohen *et al.*, 1998), as well as by

477 the correlation between Cd and Ca accumulation in numerous non-metallicolous populations

478 of *N. caerulescens* grown (Sterckeman *et al.*, 2017). The work of Chen *et al.* (2018b) on rice

479 indicates that the Ca channels OsANN4 and OsGLR3.4 can mediate Cd uptake in roots (Table

1). However, their results were obtained after exposing the roots to 20 µM Cd, which is about
1000 times higher than the concentration plants are exposed to in field conditions.

The existence of a root efflux system for Cd was suggested quite a long time ago, in particular 482 by Costa and Morel (1993) from the absorption kinetics of white lupin (Lupinus albus). This 483 hypothesis was further reinforced by studies showing that a small portion of the radioactive 484 Cd deposited on lettuce, radish and bean leaves was transferred to the rhizosphere soil 485 (Fismes et al., 2005). More recently, the PDR8 protein (pleiotropic drug resistance), 486 expressed in all cells of A. thaliana, but more intensely in the root hairs and epidermis, has 487 488 been shown to be responsible for Cd efflux (Kim et al., 2007) (Table 1). This protein belongs to the large family of ATP-binding cassette (ABC) transporters, known for transporting a 489 broad range of substrates, including ions, carbohydrates, lipids, xenobiotics, drugs and toxic 490 491 metals. In addition, a small Cys-rich membrane protein, AtPCR1 (Arabidopsis plant cadmium resistance), is also able to reduce Cd accumulation when expressed either in yeast or A. 492 thaliana protoplasts (Song et al., 2004). This protein is localized in the plasma membrane and 493 expressed in both roots and shoot. In contrast to AtPCR2, which is a Zn efflux transporter 494 (Song et al., 2010), AtPCR1does not seems to act as an efflux transporter and the mechanism 495 496 by which it confers Cd resistance to the plant remains unclear.

497 VI. Complexation of Cd by organic ligands in the plant

Cadmium can form complexes with organic ligands both in the cytosol and the vacuole. This reaction is generally considered to be a detoxification process occurring in the roots as well as in the shoots (Vögeli-Lange and Wagner, 1990; Cobbett and Goldsbrough, 2002; Clemens, 2006; Verbruggen *et al.*, 2009a). In the articles reporting research into this process, it is often not easy to distinguish the plant organs in which it takes place. This is the reason why, unless indicated, this part of the review considers Cd complexation in both roots and shoots. 504

A. Glutathione and phytochelatins

Glutathione (GSH) and phytochelatins (PCs), are non-proteinogenic peptides involved in Cd
detoxification. PCs have been extensively studied since they were discovered (Grill *et al.*,
1985). In roots of non-hypertolerant plants, most Cd is chelated by PCs (Verbruggen *et al.*,
2009a), yet GSH and PCs also bind Cd in the shoot.

- 509 GSH (Glu-Cys-Gly), which is essential for the cellular redox balance can form complexes
- 510 with various metals. Upon metal or metalloid excess, PCs, with the structure: (γ -Glu-
- 511 Cys)_nGly, where n = 2-11) are synthetized from GSH by the phytochelatin synthase (PCS)
- 512 (Grill et al., 1989; Clemens et al., 1999; Ha et al., 1999; Vatamaniuk et al., 1999; Vatamaniuk
- 513 *et al.*, 2000). The gene encoding PCS was shown to be essential for Cd and As tolerance (Ha
- 514 et al., 1999). In numerous plant species, the synthesis of PCs is induced by exposure to Cd
- 515 (Salt *et al.*, 1997; Schat *et al.*, 2002; Hernandez-Allica *et al.*, 2006; Vazquez *et al.*, 2006;
- Lindberg *et al.*, 2007). The relatively soft acid character of Cd^{2+} leads to a strong affinity for
- 517 soft sulphur donor atoms, especially for thiolates, for instance in cysteine residues (Mah and
- 518 Jalilehvand, 2010). The Cd-GSH and Cd-PC complexes have been determined and
- 519 characterized *in vitro* (Cruz *et al.*, 2005; Chekmeneva *et al.*, 2007; Leverrier *et al.*, 2007;
- 520 Chekmeneva et al., 2008; Mah and Jalilehvand, 2010; Jacquart et al., 2017). In vivo, the
- 521 existence, the distribution and the formation kinetics of Cd-PC complexes was supported by
- analyses using X-ray absorption spectroscopy (Salt *et al.*, 1995; Salt *et al.*, 1997), small-scale
- 523 gel filtration (Rauser, 2003) or fractionation techniques coupled to mass spectrometry
- 524 (Persson *et al.*, 2006).
- 525 The first step of the detoxification process is the binding of Cd^{2+} with GSH to form
- 526 bisglutathionate-Cd complex, GS₂-Cd(II). GS₂-Cd(II) interacts with the phytochelatin
- 527 synthase, a constitutively expressed enzyme, which triggers the phytochelatin synthesis and
- 528 the formation of PC-Cd(II) low molecular weight complexes (LMW PC-Cd) (Vatamaniuk et

23

al., 2000). LMW PC-Cd complexes are then transported into the vacuole *via* an ABC-type
transporter (see below) to form high molecular weight, sulphide-containing complexes (HMW
PC-Cd S)(Rauser, 1995).

Phytochelatin production can increase plant tolerance to Cd (Persson et al., 2006; Gasic and 532 Korban, 2007; Brunetti et al., 2015), but this cannot be generalized. Comparing a Cd 533 accumulating and tolerant cultivar of black oat (Avena sativa) to a Cd-sensitive one, Uraguchi 534 535 et al. (2009a) observed that Cd exposure increased PC concentrations in both cultivars, thus indicating that the differences between the cultivars could not be accounted for by difference 536 537 in PC synthesis. In this case, increased storage in the cell wall would appear to account for enhanced tolerance and accumulation of Cd (see below). Overexpression of AtPCS1 in 538 Arabidopsis was even found in some instances to increase Cd sensitivity (Lee et al., 2003). 539 540 This was attributed to a detrimental depletion of GSH in the overexpression lines. The role of PCs in adaptation to Cd-contaminated soils has been questioned by the work of 541 Leopold et al. (1999), who showed that Cd-complexes disappeared after 7 to 14 days of Cd 542 exposure in the metal tolerant Silene vulgaris. This result indicated that synthesis of metal-543 PC-complexes was transient and not necessarily important for long-term metal tolerance by 544 plants. Wójcik et al. (2006) also observed that PC concentration increases with the increasing 545 Cd concentration to which N. caerulescens is exposed, but decreases with exposure time. PCs 546 seem to play a minor role in Cd tolerance by this hyperaccumulator (Wojcik et al., 2005b; 547 Wójcik et al., 2006). This is consistent with the fact that in young leaves of N. caerulescens, 548 around 25% of the Cd ligands were S ligands (e.g. phytochelatins, and other Cys-rich 549 peptides), and the proportion of S ligands decreased with increasing leaf age. In senescent 550 551 leaves, almost exclusively O ligands were found (Küpper et al., 2004). Schat et al. (2002); Wójcik et al. (2006) suggested that adaptive Cd tolerance was not dependent on PC-mediated 552 sequestration. For Sun et al. (2007a), PCs are not involved in Cd transport, 553

554 hyperaccumulation or tolerance in the metal-hypertolerant population of S. alfredii, while Zhang et al. (2010) showed that PCs do not detoxify Cd in the roots of this species, contrarily 555 to what might occur in the leaves. Comparing different species and organs, Marentes and 556 557 Rauser (2007) showed that phytochelatins were not necessarily the major Cd ligands. Sun et al. (2007b) concluded that PC synthesis was not the primary reason for Cd tolerance in 558 Solanum nigrum. Moreover, strikingly, genes involved in phytochelatin synthesis or transport 559 have not been identified in transcriptomic or genetic analyses of Cd hyperaccumulation or 560 hypertolerance. 561

In addition, PCs have been implicated in long distance Cd transport. Expressing wheat PCS
specifically in *A. thaliana* roots, Gong *et al.* (2003) found that PCs synthesized in roots are
transported to the shoots and that this correlates with enhanced Cd transfer from root to shoot.
However, later studies indicated that this is not the main pathway for Cd root to shoot transfer
(Wong and Cobbett, 2009). Using grafts between a phytochelatin-deficient Arabidopsis
mutant and the wild type, Chen *et al.* (2006) demonstrated that phytochelatins are also
transported from shoot to roots.

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B. Metallothioneins and other cysteine-rich polypeptides

Metallothioneins (MTs) are cysteine rich, low molecular weight, gene-encoded, metal-binding 570 polypeptides that are ubiquitously found in eukaryotes and cyanobacteria (Cobbett and 571 Goldsbrough, 2002). It has been shown that MTs play an important role in metal homeostasis 572 and tolerance in plants (Zimeri et al., 2005; Guo et al., 2008; Ren et al., 2012; Benatti et al., 573 2014). Elevated expression of MT genes (MT1, MT2, MT3) has frequently been observed in 574 metal-hyperaccumulating and hypertolerant species or populations (van Hoof *et al.*, 2001; 575 576 Mengoni et al., 2003; Roosens et al., 2005; Chiang et al., 2006; van de Mortel et al., 2006; Hassinen et al., 2007; Jack et al., 2007). However, no positive correlation between these 577 expression levels and metal tolerance/accumulation was observed. Recently, Peng et al. 578

(2017a) proposed that the cysteine-rich MT-like protein SpMTL could act as a cytosolic
chelator to detoxify Cd in the hyperaccumulator *S. plumbizincicola* (Table 1). However, this
role has to be confirmed in other species.

In *Digitaria ciliaris*, Kuramata *et al.* (2009) identified a novel gene, DcCDT1 (*D. ciliaris* cadmium tolerance), encoding a 55 amino acid peptide, containing 15 (27%) cysteine residues, as well as its rice homolog, OsCDT1 (Table 1). When constitutively expressed in yeast cells and *A. thaliana*, this gene enhanced their Cd tolerance by limiting cellular Cd accumulation. The authors proposed that the peptide is secreted outside the cell, where it partly associates with the cell wall. It then prevents Cd internalization by chelating the metal at the cellular surface.

589

Organic acids

С.

590 Carboxylic acids also contribute to Cd complexation. Salt et al. (1997) showed that during the initial stages of Cd uptake (6-12 h) by Brassica juncea seedlings, most of the intracellular Cd 591 was bound to oxygen atoms, possibly from organic acids. The spectra of pure Cd citrate fitted 592 relatively well with the spectra of B. juncea seedlings. Later, Ueno et al. (2005) showed that 593 Cd is mainly associated to malate in the leaves of N. caerulescens. In roots and shoots of the 594 hyperaccumulator Thlaspi praecox, Vogel-Mikuš et al. (2010) found that up to 80% of the Cd 595 ligands are oxygen ligands provided by the cell walls (see below) and by organic acids stored 596 in the vacuoles. In the stems and leaves of Gomphrena claussenii, a Cd tolerant species, Cd 597 598 was mainly bound to oxalate in crystals also containing Ca. The plant was however, exposed to very high Cd concentrations in hydroponics (100 µM CdSO₄). In the roots, there was no 599 oxalate crystals and Cd was predominantly associated to S ligands (Pongrac et al., 2018). The 600 601 complexation of Cd by organic acids is particularly important in the epidermal cells, where the vacuoles represent 99% of the symplast volume. In the stems and leaves of the Cd 602 hyperaccumulating ecotype of S. alfredii, Cd is mainly associated with oxygen ligands and the 603

majority (about 70% in leaves and 47% in stems) is bound to malic acid, i.e. the major
organic acid in the shoots (Tian *et al.*, 2011).

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607 VII. Storage of Cd in vacuoles of roots and shoots

The vacuole is an important storage organelle for many ions. As with complexation with 608 609 organic ligands, transport from the cytosol into the vacuole is regarded as a process increasing plant tolerance and plant ability to accumulate high levels of toxic metallic ions, occurring in 610 either the roots or the shoots. In the leaves of N. caerulescens, Wojcik et al. (2005a) found 611 that Cd accumulates in vacuoles and suggested these are the main compartments for storage 612 and detoxification of this metal. In the literature, it is sometimes difficult to distinguish 613 614 between root and aerial parts with respect to the nature of the Cd transporters present in the vacuole membrane, the tonoplast. This is why this part of the review concerns storage in 615 616 vacuoles in both below- and above-ground organs.

Several transport pathways mediate Cd sequestration in the vacuole either as free Cd or PC-617 Cd complexes. They use either ATP hydrolysis or the proton gradient across the vacuolar 618 membrane as sources of energy to drive substrate accumulation inside the vacuole. Cadmium 619 can be transported into the vacuole by means of a H⁺/cation-antiporters designated as CAX 620 (CAtion eXchanger), that use the proton gradient to mediate divalent cation storage in the 621 622 vacuole (Salt and Wagner, 1993; Hirschi et al., 2000; Pittman et al., 2004; Shigaki and Hirschi, 2006; Shen et al., 2012). Among the members of the CAX family that have been 623 characterized as calcium transporters, CAX1, CAX2 and CAX4 have the ability to transport 624 625 Cd as well (Table 1). In A. thaliana, CAX2 is strongly expressed in the leaf vascular tissue, in vegetative meristematic tissue, as well as in the root vascular tissue and in the root tip 626 (Pittman et al., 2004). Koren'kov et al. (2007) compared seven different CAX genes from A. 627

628 thaliana in transformed tobacco. They showed that all the genes are involved in the transport of Cd²⁺, Ca²⁺, Zn²⁺, and Mn²⁺ to varying degrees, but that AtCAX4 and AtCAX2 have a 629 higher Cd²⁺ selectivity. AtCAX4 is expressed at higher levels in the roots. Korenkov et al. 630 631 (2009) showed that expression of AtCAX2 or AtCAX4 in transformed tobacco roots reduces Cd translocation to the aerial parts by promoting sequestration of the metal in the roots. In 632 addition, CAX1a is expressed in vacuoles of barley (H. vulgare) exposed to Cd, thus 633 suggesting that this protein plays a role in Cd vacuolar storage (Schneider et al., 2009). 634 Expression of AtCAX1 in petunia enhanced Cd tolerance and accumulation (Wu et al., 2011). 635 636 Moreover, Baliardini et al. (2015) suggested that the role of CAX1 in Cd tolerance is not linked to Cd transport but in maintaining cytosolic Ca levels to avoid uncontrolled ROS 637 accumulation. The evidence that CAX transporters are involved in Cd transport in planta 638 639 come from overexpression experiments, whereas loss of function would be necessary to determine their actual contribution to Cd storage and Cd tolerance. Membrane transporters of 640 the CDF family do not seem to be involved into vacuolar Cd sequestration. Indeed, neither 641 642 vacuolar MTP from A. thaliana nor its hyperaccumulator relative A. halleri displayed Cd transport activity (Dräger et al., 2004; Desbrosses-Fonrouge et al., 2005; Arrivault et al., 643 2006). Accordingly, MTP1 was identified as a QTL for Zn tolerance, but not for Cd tolerance 644 in A. halleri (Courbot et al., 2007; Willems et al., 2007). 645

The P_{1B}-ATPase subfamily HMA (for Heavy Metal Associated) also plays a role in Cd sequestration in vacuoles. Morel *et al.* (2009) showed that AtHMA3 is located in the vacuolar membrane of *A. thaliana* (Table 1). In shoots, AtHMA3 has a high expression level in guard cells, hydathodes, vascular tissues, while in roots, it is expressed in the stellar cells and at the apex. AtHMA3 participates in the vacuolar storage of Cd, improving plant tolerance to the metal (Morel *et al.*, 2009). A homologous gene, OsHMA3, is a major determinant of the variation in Cd accumulation in shoots and grains among rice cultivars (*O. sativa*) (Ueno *et*

al., 2010; Miyadate et al., 2011; Sui et al., 2019). OsHMA3 is mainly expressed in the 653 tonoplast of root cells. OsHMA3 was found to limit Cd translocation from roots to shoot by 654 selectively sequestrating Cd into root vacuoles (Ueno et al., 2010). Mutations leading to a loss 655 656 of functionality of the OsHMA3 transporter in the roots explain the elevated accumulation of Cd in the aerial parts of several indica rice cultivars, Jarjan (Ueno et al., 2011a; Yan et al., 657 2016), Hitomebore (Tanaka et al., 2016) and Cho-Ko-Koku (Miyadate et al., 2011), 658 compared to japonica cultivars. Sun et al. (2019) showed that a Gly to Arg substitution at 659 position 512 in OsHMA3 causes some *japonica* rice cultivars to accumulate larger amounts of 660 661 Cd in the shoots and grains. Interestingly, variations in the coding sequence of AtHMA3 also account for the variations in shoot Cd accumulation among A. thaliana accessions (Chao et 662 al., 2012), in Brassica rapa (Zhang et al., 2019) and in Durum wheat (Maccaferri et al., 663 664 2019). Overexpression of a functional OsHMA3 in a japonica rice cultivar further decreased Cd accumulation in shoots and grains (Ueno et al., 2010). Shao et al. (2018) fused a promotor 665 of OsHMA2 (highly expressed in the node) to OsHMA3. The consequence was an enhanced 666 667 expression of OsHMA3 and a modification in the tissue localization of the transporter, which was also expressed in the nodes. This resulted in a reduced Cd accumulation in grain due to 668 the sequestration of Cd in the vacuoles of various tissues. The genome of the metal 669 hyperaccumulator N. caerulescens also encodes one or several HMA3 homologues that are 670 671 localized in the vacuolar membrane and allow Cd sequestration (Ueno et al., 2011b). In 672 contrast with rice and A. thaliana, NcHMA3 is also strongly expressed in shoot tissues and is responsible for Cd sequestration in leaf cell vacuoles (Ueno et al., 2011b; Mishra et al., 673 2017). The latter suggested that greater expression of NcHMA3 would explain the 674 hyperaccumulation of Cd by certain populations (e.g. Ganges) of this species (Ueno et al., 675 2011b). In the hyperaccumulator S. plumbizincicola (Crassulaceae), SpHMA3 is also strongly 676 expressed in shoots and is localized in the tonoplast. It plays an important role in Cd 677

2017). Working on HMA3 from soybean (*Glycine max*), Wang *et al.* (2018) proposed that GmHMA3 limits the translocation of Cd from the roots to the stems, by sequestrating Cd into the root endoplasmic reticulum (ER). This is surprising, as Cd toxicity would be expected to be very high in the ER, probably interfering with disulphide bound formation. It has recently been proposed that the homeodomain-containing protein, OsHB4a was responsible for Cd sensitivity and accumulation in the leaves and grains of rice plants through

detoxification in young leaves and stems by sequestrating the metal in the vacuole (Liu et al.,

the regulation of other genes, such as OsHMA2 and OsHMA3 (Ding *et al.*, 2018). OsHB4a is

686 up-regulated upon Cd exposure and favours xylem loading and Cd translocation. Moreover,

miR166 a well-conserved miRNA present in all plants, targets OsHB4a for degradation. The

expression of miR166 decreases upon Cd exposure. Cd exposure thus favours its

accumulation and toxicity in rice by simultaneously up-regulating OsHB4a and down-

690 regulating its inhibitor, miR166.

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Based on their up-regulation upon Cd exposure and their localization at the vacuolar 691 membrane, the MRP3 transporters of the ABC family from mouse-ear cress and barley have 692 been suggested as driving the influx of PC-Cd complexes into the vacuole (Tommasini et al., 693 694 1998; Schneider et al., 2009) (Table 1). A similar function has been demonstrated for the AtMRP7 transporter (Wojas et al., 2009). In the case of AtMRP7 or AtMRP3, whether Cd is 695 696 transported as a complex with PC rather than as a free ion, was not determined. However, 697 Park et al. (2012) showed that transporters of the same subfamily, AtABCC1/AtMRP1 and 698 AtABCC2, are indeed responsible for the sequestration of phytochelatin-Cd complexes into the root vacuoles of A. thaliana (Table 1). Song et al. (2014) demonstrated that a vacuolar 699 700 transporter for PC-Cd was also present in barley and proposed that it could be an ABC-type transporter, such as AtABCC1 or AtABCC2. In addition, AtABCC3 increases vacuolar Cd 701

702	through the transport of Cd-phytochelatin complexes. Its activity is regulated by Cd and
703	coordinated with that of AtABCC1 and AtABCC2 (Brunetti et al., 2015).

In contrast with CAX, MTPs, HMA and ABCCs, transporters belonging to the NRAMP 704

705 family participate in Cd efflux from the vacuole (Table 1). In A. thaliana, AtNRAMP3 and

- AtNRAMP4 both reside on the vacuolar membrane (Thomine et al., 2003; Languar et al., 706
- 2005) and can transport Cd when expressed in yeast (Thomine et al., 2000). Consistent with 707
- 708 its redundant function in Cd efflux from root vacuoles, *nramp3nramp4* double mutant in the

WS accession, which harbours a functional HMA3 gene, accumulates greater Cd 709

710 concentrations in roots than the wild type (Chao et al., 2012; Pottier et al., 2015). Conversely,

overexpression of AtNRAMP4 decreases root Cd concentration. In the hyperaccumulator N. 711

712 caerulescens, NcNRAMP3 and NcNRAMP4 have the same protein function as their A.

713 thaliana homologues, but are expressed at higher levels (Oomen et al., 2009). The efflux of

Cd and Zn from root vacuoles driven by NcNRAMP3 and 4 may be required for the efficient 714

translocation of these metals from roots to shoots in *N. caerulescens*. 715

716

VIII. Interactions of Cd with the apoplast

717

A. Transport through the apoplast

The root apoplast is a compulsory pathway for Cd^{2+} and for Cd complexes before entering the 718 719 symplast. The solute is considered to move radially in the root apoplast by diffusion and advection, before reaching the plasma membrane and crossing it through transport proteins 720 (see above). Cd needs to use the symplastic route to cross the exodermis and/or the 721 722 endodermis cell layers, where apoplastic transport is blocked by the apoplastic barriers (Casparian strip, suberine lamellae, secondary thickening). It has however been suggested that 723 Cd could reach the xylem *via* the sole apoplastic pathway in the parts of the root where the 724 apoplastic barriers are lacking or incomplete, namely the root tip, stage I endodermis and sites 725

of lateral root emergence (White, 2001). This hypothesis is supported by studies showing agreater Cd influx along the few millimetres at the root tip (see above).

The relative contributions of the symplastic and apoplastic pathways to Cd loading in the 728 729 xylem have yet to be determined. The work of Redjala et al. (2009); Redjala et al. (2010b) showed that apoplastic adsorption had a minor contribution to root Cd uptake, when 730 compared to symplastic uptake at low concentration exposure, such as that in soil solution. 731 Nevertheless, Feng et al. (2018) showed that, under Cd stress, endodermal apoplastic barriers 732 were much thicker in a low Cd accumulating sorghum genotype than in a highly accumulating 733 734 one. The weaker apoplastic barriers were considered to favour xylem loading via the apoplastic pathway in the highly accumulating genotype. Tao et al. (2017) highlighted the 735 significance of the apoplastic pathway in Cd accumulation using a Cd hyperaccumulating 736 737 ecotype of S. alfredii. The contribution of apoplast to Cd xylem loading depended on the Cd exposure level and represented up to 37% of total uptake. The main entry sites for apoplastic 738 uptake were in the root apices and at the base of lateral roots, where suberin lamellae are not 739 well-developed. 740

In the aerial parts, it is difficult to quantify the contribution of the apoplastic transport of Cd.
The cell walls of leaves (see below) contain Cd that may migrate into the leaves towards the
epidermis, by advection in the network of cell walls driven by the transpiration flow.

744

B. Storage in the apoplast

Electronegative sites of cellulose, hemicellulose and pectin chains enable the sorption of
cations in the cell walls (Haynes, 1980; Grignon and Sentenac, 1991). Sorption in the apoplast
is considered, together with chelation, vacuolar sequestration and translocation, as a means of
Cd detoxification (Vazquez *et al.*, 2006; Nocito *et al.*, 2011). However, the fraction of the
metal stored in this compartment appears to vary widely. When maize and pea (*Pisum sativum*) were cultivated and exposed to Cd in hydroponics, higher Cd levels were found in

751 the cell-wall fraction of leaves and roots of maize than in those of pea. This correlated with more severe toxicity symptoms in pea (Lozano-Rodriguez et al., 1997). Based on Energy 752 Dispersive X-ray micro-analysis in the root cortex of A. thaliana, Cd accumulated in the cell 753 wall together with phosphate whereas, in the central cylinder, Cd was present in the cell wall 754 as Cd/S granular deposits in the middle lamella of the pericycle (Van Belleghem et al., 2007). 755 In the other parts of the central cylinder, the Cd/S granular deposits were located near the 756 plasma membrane and not in the middle lamella. In leaves of oilseed rape (Brassica napus), 757 Cd in cell walls represented only 11% of the metal content (Carrier et al., 2003). In tea 758 seedlings, most of the Cd accumulated in roots, with about 12% to 30% in the cell walls, 759 while cell wall represented 43.6-83.4% of the very low Cd content in leaves (Cao et al., 760 2018a). Cosio et al. (2005) showed that 33-35% of Cd in the aerial parts of N. caerulescens is 761 762 in the cell walls. In the case of willow (Salix viminalis) exposed to high Cd levels, the main sink of the metal was the pectin layer of collenchyme cell walls, close to the leaf veins 763 (Vollenweider et al., 2006). Xiong et al. (2009) showed that exogenous NO increases Cd 764 tolerance in rice plantlets by enhancing pectin and hemicelluloses content in the root cell wall, 765 which is associated with an increase in Cd accumulation in the root cell walls and a decrease 766 in Cd concentration the soluble fraction of leave cells. Xu et al. (2015) showed that, in A. 767 thaliana, Cd itself increases the polysaccharide contents and the binding capacity of root cell 768 walls. Among oat varieties, the highest Cd-resistance coincided with greatest Cd storage in 769 770 the cell walls of aerial parts (Uraguchi et al., 2009a). This was also suggested by the work of Sterckeman et al. (2011), which showed an increase in Cd apoplastic fraction in maize and N. 771 caerulescens roots after exposure to high Cd concentrations. In the hyperaccumulating species 772 A. halleri, Meyer et al. (2015) correlated Cd induced modification in the cell wall with 773 intraspecific variations in Cd tolerance and accumulation. The apoplast also appeared to be an 774 important storage compartment for Cd in the Cd hyperaccumulator S. alfredii (Zhang et al., 775

2010). Li *et al.* (2015) showed that more Cd was retained in the root cell wall of the nonhyperaccumulating ecotype (NHE) of *S. alfredii* than in the hyperaccumulating one (HE).
Transcriptomic analyses also indicated an important role played by the cell wall, in the Cd
hyperaccumulation in *S. plumbizincicola* (Peng *et al.*, 2017b). As in *A. thaliana*, Cd treatment
induced an increase in cell-wall polysaccharides (pectins and hemicelluloses) in the NHE of *S. alfredii*, but not in the HE. Consequently, more Cd was available for xylem loading in the
HE ecotype.

It should be noted that in several of the above-cited works (Lozano-Rodriguez *et al.*, 1997;

784 Carrier *et al.*, 2003; Cao *et al.*, 2018a), as well as in others (e.g. (Fu *et al.*, 2011; Wang *et al.*,

2015a; Wang *et al.*, 2015b), the method used to determine Cd (or other elements) in cell walls is questionable. Indeed, it is based on a fine grinding of the fresh tissues in a buffer solution to destroy the cells. The homogenate is generally then sieved to collect the cell walls. The filtrate is centrifuged to separate chloroplasts from other organelles and from cytosolic and vacuole solution. This procedure in liquid phase does not exclude the diffusion of Cd from one fraction to another during the tissue grinding and fraction separation. The content measured in each fraction might not therefore, correspond to the distribution of the metal *in vivo*.

792

IX. Translocation and reallocation of Cd

793 Cadmium concentration is often higher in the roots than in the shoots (Jarvis *et al.*, 1976;

Lozano-Rodriguez et al., 1997; Harris and Taylor, 2013; Cornu et al., 2016), and generally

lower in the seeds than in the stem and leaves (Hinesly *et al.*, 1978; Wiggenhauser *et al.*,

2016; Chen et al., 2018a). In some species, such as lettuce (Lactuca sativa), spinach (Spinacia

797 *oleracea*), cauliflower (*Brassica oleracea*) and oat, concentrations are higher in the shoot than

in the roots when plants are grown on low Cd contaminated soils. However, when the plants

are grown on more contaminated soil, the root concentration becomes higher than that of the

shoots (John, 1973). In maize, Perriguey et al. (2008) observed increasingly greater

801 concentrations in roots than in shoots and an increase in the Cd root-to-shoot ratio, as that of the Cd concentration in the nutrient solution rose. The ratio sharply decreased after reaching a 802 peak, as if the root retention capacities became overloaded at this point. The sequestration of 803 804 Cd in the roots up to a certain limit corresponds to the excluder strategy proposed by Baker (1981) for certain species. This author also defined an accumulator strategy, "where metals 805 are concentrated in above-ground plant parts from low or high soil levels". This is the case for 806 N. caerulescens, S. nigrum or S. alfredii (Yang et al., 2004; Wei et al., 2006; Lovy et al., 807 2013), which can be regarded as Cd hyperaccumulators as the Cd concentration in their shoots 808

809 can reach 100 mg Cd (kg DW)⁻¹ (van der Ent *et al.*, 2013).

810

A. Cd loading into the xylem

HMA transporters play an important role in the loading of Cd in xylem vessels. As ATPases, 811 812 they can transport the cation against its electrochemical gradient, from the cytosol into the xylem sap. AtHMA4 is expressed at its highest level in tissues surrounding the root vascular 813 vessels in A. thaliana (Mills et al., 2003; Verret et al., 2004). Located in the plasma 814 membrane and acting as an efflux pump, HMA4 plays a role in metal loading into the xylem, 815 thereby enhancing the root-to-shoot translocation of Zn and Cd in A. thaliana (Verret et al., 816 817 2004; Mills et al., 2005). Together with AtHMA4, its paralogue AtHMA2 was shown to be involved in the root-to-shoot transfer of Cd (Wong and Cobbett, 2009; Wong et al., 2009) 818 819 (Table 1). Moreover HMA4 is also important for Cd translocation in the hyperaccumulator N. 820 caerulescens (Papoyan and Kochian, 2004). Accordingly, in A. halleri, enhanced expression 821 and a triplication of HMA4 was shown to be responsible for the hypertolerance to Cd (Courbot et al., 2007; Hanikenne et al., 2008). In N. caerulescens, the HMA4 gene is also 822 823 present as multiple tandem copies (Ó Lochlainn et al., 2011) and differences in HMA4 expression correspond to the differences in Cd translocation and accumulation among 824 different populations of N. caerulescens (Craciun et al., 2012). 825

826 In rice, a single gene, OsHMA2, which is orthologous to AtHMA2 and AtHMA4, encodes an efflux transporter involved in Cd xylem loading. OsHMA2 is constitutively expressed in the 827 vascular bundles of roots at the vegetative stage, and knocking-out this gene results in lower 828 829 root-to-shoot translocation and grain content of Cd (Nocito et al., 2011; Satoh-Nagasawa et al., 2012; Takahashi et al., 2012; Fontanili et al., 2016). OsHMA2 is also expressed in the 830 phloem tissue of nodes (Yamaji et al., 2013). To account for the role of OsHMA2 in the 831 control of Cd transfer between phloem and xylem in nodes, Yamaji et al. (2013) proposed that 832 OsHMA2 could function as an influx transporter for both Cd and Zn. Such reverse transport is 833 834 extremely unlikely, as the function of HMAs in metal efflux has been conserved across kingdoms and the transport mechanism involves ATP hydrolysis at the cytosolic side of the 835 protein. Based on a Cd/Zn reciprocal competition study, Fontanili et al. (2016) suggested that, 836 837 beside OsHMA2, another transport system which is not affected by the presence of Zn^{2+} , yet to be identified, may mediate root-to-shoot Cd translocation in rice. Recently, Hao et al. 838 (2018) suggested that OsCCX2, a putative cation/calcium (Ca) exchanger mainly expressed in 839 xylem region of the nodes, could be an efflux transporter, controlling Cd loading into xylem 840 (Table 1). 841

842 Defensins are small cysteine-rich proteins, widely distributed in invertebrates, plants, and mammals and known for their involvement in the ancestral non-specific immune response and 843 antifungal activity. Their role in metal homeostasis was first established in a work reporting 844 845 that this conferred enhanced Zn tolerance in yeast and plant (Mirouze et al., 2006). More 846 recently, CAL1, a defensin-like protein from rice, was found to bind Cd and to drive its efflux from xylem parenchyma cells to extracellular spaces (the apoplast), thus contributing to Cd 847 848 loading into the xylem sap (Luo et al., 2018) (Table 1). Therefore, the expression of CAL1 leads to an increased Cd concentration in leaves, through increased translocation by the 849
xylem. Whether the secretion of OsCAL1–Cd complexes indeed involves transporters or
vesicular trafficking, is however still unknown.

852

B. Transport in the xylem

Once loaded into the root xylem, Cd migrates to the aerial parts with the xylem sap flow, 853 driven by plant transpiration, *i.e.* by the gradient of water potential between the soil and the 854 atmosphere. Numerous works show that Cd concentration in aerial plant parts increases with 855 the concentration in the root exposure solution (see for instance Jarvis et al. (1976); Florijn 856 and Van Beusichem (1993a); Lovy et al. (2013)). Moreover, various investigations have 857 858 shown that Cd content in the xylem sap increases with Cd content in the exposure solution (Salt et al., 1995; Wei et al., 2007; Ueno et al., 2008; Nakamura et al., 2013; Hazama et al., 859 2015; Cornu et al., 2016). 860

Measuring Cd concentration in 69 rice accessions, Uraguchi et al. (2009b) found strong 861 correlations between Cd levels in the xylem sap and in shoots and grains. Zhang et al. (2018) 862 obtained similar results in the case of rape (B. napus). Harris and Taylor (2013) demonstrated 863 the importance of root-to-shoot translocation in the accumulation of Cd in durum wheat grain. 864 Together, these results demonstrate the importance of xylem transport for shoot Cd 865 accumulation across species. However, the transpiration rate generally only slightly affects 866 the root-to-shoot translocation of Cd (Van der Vliet et al., 2007; Liñero et al., 2016), probably 867 because it is always sufficient to transport the Cd absorbed by the roots to the shoots. 868

869

C. Cadmium speciation in the xylem sap

870 Theoretically, based on the pH and the composition of the xylem sap, Cd could be present as a

871 free ion or involved in various complexes with sulfhydryl (SH), or N- or O-based groups.

872 Cadmium could be linked to cysteine-containing molecules, such as glutathione or

873 phytochelatins, organic and amino acids (Welch and Norvell, 1999).

874 White et al. (1981) have shown that Cd forms negatively-charged complexes in soybean and tomato xylem sap. In soybean, Cataldo et al. (1988) found Cd as a single, slightly anionic 875 species and as a cationic component, which appeared to be inorganic Cd. Citric acid promotes 876 877 the translocation of Cd in tomato, which may be due to the formation of citrate Cd complexes in the xylem (Senden et al., 1995). In contrast, in the xylem sap of B. juncea, Salt et al. (1995) 878 879 found that Cd is coordinated either with oxygen or with nitrogen, which excludes a complexation by PCs or MTs. Based on the analysis of low molecular weight compounds in 880 the xylem sap of *B. juncea*, Wei *et al.* (2007) calculated that the Cd speciation was 33% 881 882 organic acid-Cd and 67% inorganic anion-Cd in plants exposed to a low Cd concentration (10^{-8} M) . In the xylem of plants exposed to 10 μ M Cd, the calculated Cd speciation was 1% 883 PC-Cd, 35% organic acid-Cd, and 64% inorganic anion-Cd. 884

885 According to Ueno et al. (2008), based on nuclear magnetic resonance measurement and modelling of chemical speciation, Cd is essentially present as a free ion in the xylem sap of 886 the hyperaccumulator A. halleri. In castor bean plants (Ricinus communis) exposed to 1 µM 887 Cd for 3 weeks, 92% of the Cd in xylem was found to be present as a free cation (Hazama et 888 al., 2015). In sunflower exposed to low and realistic Cd concentrations (2.5 nM or 20 nM), 889 890 Cornu et al. (2016) determined that the fraction of free Cd ion in the xylem sap was around 50%. These results are difficult to reconcile, as they suggest that depending on the species and 891 892 on the Cd concentration in the medium, Cd may be transported within the xylem both as 893 complexes with organic and inorganic acids, but also as a hydrated free cation.

894

D. Xylem unloading and Cd distribution in shoots

After their long-distance transport in the xylem sap towards the aerial parts of the plant,
mineral ions are unloaded from the xylem vessels to be distributed in the other leaf tissues
(Figure 5). Xylem unloading of Cd is expected to follow the same process as the nutrient ions.
However, even for nutrients, this process has yet to be completely described (Shabala, 2007).

899 Because of water evaporation, element concentration in the xylem sap of the leaves can 900 potentially increase strongly. To prevent toxic salt excess, plants have evolved mechanisms of xylem unloading in leaves. These mechanisms take place in the highly-branched network of 901 902 major and minor veins crossing the leaf blade. The ions may then be transported from the xylem parenchyma towards leaf cells via symplastic pathways. There is very little information 903 about the transport systems responsible for the unloading of Cd from the xylem. Cd²⁺ 904 probably enters the leaf cells using transport proteins dedicated to nutrients (Fe, Mn, Zn), as 905 occurs during root absorption. In the shoot of *N. caerulescens*, HMA4 and HMA3 expression 906 is highest in the bundle sheath of the vein, while for A. halleri, these proteins are expressed in 907 greater amounts in the mesophyll. However, their role in the distribution and storage of Cd 908 909 (and Zn) in the shoot is unclear and the transport pathway to the epidermis, in the case of N. 910 caerulescens, remains unknown (Mishra et al., 2017).

Cosio et al. (2004) measured the Michaelis-Menten kinetic parameters (K_m, V_{max}) of Cd 911 uptake by mesophyll protoplasts of N. caerulescens (Ganges and Prayon populations) and of 912 A. halleri. They observed no significant difference between the three plants, although N. 913 caerulescens Ganges and A. halleri were both Cd hyperaccumulators while N. caerulescens 914 915 Prayon accumulates Cd at a lower level. The authors suggested the existence of regulation mechanisms upstream of the plasma membrane of leaf mesophyll cells to explain the 916 917 differences in shoot accumulation between populations. Alternatively, Cd accumulation may 918 be driven downstream by enhanced sequestration into vacuoles. Pre-exposure of the plants to 919 Cd induced an increase in Cd accumulation in protoplasts of "Ganges", whereas it reduced Cd accumulation in A. halleri. Cosio et al. (2004) suggested that several parallel transport 920 921 systems could be responsible for Cd internalization. These could be the Zn and Ca uptake pathways for the Prayon accession, whilst other as yet to be described transporters would be 922 used in the Ganges accession. 923

Cosio et al. (2005) showed that at very low Cd exposure, the metal was preferentially located 924 in the young leaves of *N. caerulescens*, while at higher exposures, the metal accumulated 925 more in the older leaves. In both Ganges and Prayon populations, Cd was present mainly at 926 927 the edges of the leaves, where it was much more concentrated in the epidermal cells, particularly in the large ones, than in the mesophyll cells. The relationship between cell size 928 and Cd concentration in the epidermis suggests that vacuolar sequestration may be a driving 929 force for Cd distribution in leaves. Cadmium hotspots spread across the entire limb surface 930 were also observed. These correspond to the necrotic spots visible on the leaves of the Prayon 931 932 population. Cadmium is not only concentrated in vacuoles, but also in cell walls, which together account for up to 35% of the metal in the leaves (Cosio et al., 2005). 933 Measuring Cd uptake kinetics from leaf sections on *N. caerulescens*, Ebbs *et al.* (2009) 934 935 observed higher K_m and higher V_{max} for the leaf mesophyll cells in the Prayon accession than for that of the Ganges. Using influx measurements in various types of leaf protoplasts from 936 the Ganges accession, Leitenmaier and Küpper (2011) confirmed that sequestration in the 937 vacuole indeed drives Cd uptake into cells. These authors also showed that the influx into 938 specific so-called epidermal storage cells was stronger than in the other protoplasts of the 939 940 epidermis and mesophyll. This indicates that the storage of Cd in epidermal tissues is due to active Cd uptake rather than to transpiration, in contrast to what had been suggested for Indian 941 942 mustard (B. juncea) (Salt et al., 1995).

943 These results are in line with previous findings in *Silene vulgaris*. In this species, Cd

944 preferentially accumulates in the lower epidermis, independently of the transpiration flow

945 (Chardonnens et al., 1998). In Noccaea praecox, another Cd and Zn hyperaccumulator, the

946 Cd concentration is twice as high in the upper and lower leaf epidermis than in the mesophyll

947 (Vogel-Mikus *et al.*, 2008). However, because of the larger volume of the mesophyll

948 compared with that of the epidermis, mesophyll should still be considered as a location where

significant amounts of Cd are detoxified. In the embryo, Cd is also preferentially stored in the
epidermis of the cotyledons of *N. praecox* (Vogel-Mikus *et al.*, 2007). Studies on *N. praecox*and on *A. halleri* confirmed that vacuolar sequestration as a complex with organic acids, and
binding to the cell walls are the main mechanisms allowing Cd accumulation in shoots. They
also indicated that Cd-thiolate complexes would be the species transported through phloem
(Vogel-Mikuš *et al.*, 2010; Huguet *et al.*, 2012).

Ueno *et al.* (2005) found that Cd formed complexes with malate, probably located in leaf
vacuoles of *N. caerulescens* (Ganges accession). In *S. alfredii*, Cd also appeared to be stored
as malate in vacuoles of parenchyma tissues (pith, cortex, and mesophyll) thanks to efficient
Cd transport across the tonoplast and a constitutively high concentration of malate in the
vacuoles (Tian *et al.*, 2011; Tian *et al.*, 2017).

960 In the case of willow (*Salix viminalis*) exposed to various Cd concentrations, Cd is located at the tips and edges of the young leaves, as well as at the base of the older leaves. This 961 localization coincides with necrotic spots (Cosio et al., 2006). Vollenweider et al. (2006) 962 noted that in willow leaves, the pectin-rich layers of the collenchyma cell walls of the veins 963 are the main Cd sink. They also observed lesions in the phloem, thus suggesting that 964 retranslocation of Cd occurs through this pathway. In the leaves of *A. thaliana*, Van 965 Belleghem et al. (2007) detected Cd in tracheids, but not in the mesophyll tissue. Based on the 966 high symplastic and apoplastic sequestration of Cd in the root parenchyma, combined with 967 968 retranslocation from the leaves via the phloem, the authors suggested that A. thaliana has an excluding strategy. This strategy is probably similar in Salix viminalis and in all other 969 excluders, while the strategy used by hyperaccumulators is to translocate most of the metal to 970 the shoots, where it is stored in various tissues and cell compartments. 971

E. Allocation through phloem

Phloem plays a key role in the distribution of Cd in the aerial parts of the plant, particularly in 973 sink organs such fruits and seeds. For instance, Kato et al. (2010) showed that the Cd 974 975 concentrations in the grains of three rice varieties correlates with the Cd concentration in the phloem sap, but not with that in the xylem sap. Tanaka et al. (2007) evaluated that 91% to 976 100% of Cd accumulated in rice grains is transported via the phloem. Many studies of phloem 977 flow have tried to determine the extent to which it contributes to Cd reallocation from shoots 978 to fruits, when compared to a direct translocation from the roots. 979 980 Several works have showed that Cd is reallocated from leaves to fruits, grain or even roots

through the phloem flow (Herren and Feller, 1997; Fismes *et al.*, 2005; Yoneyama *et al.*,

982 2010; Wei *et al.*, 2014). Using Ca as a reference of phloem supply, Liñero *et al.* (2018)

983 concluded that phloem is the main route for transporting Cd to sunflower seeds. They

calculated that the contribution of the net remobilization from the shoot accounted for 14% of
the total Cd amount in mature seeds. Using Cd stable isotopes to trace the metal fluxes in two
durum wheat cultivars, Yan *et al.* (2018) showed that 50 to 60% of the Cd accumulated in
mature grains is remobilized through phloem from pre-anthesis stores, i.e leaves and stem.
However, the metal is poorly remobilized from the leaves, even during senescence. This led
the authors to consider these organs as irreversible Cd sinks. In contrast, Cd from the stem is

990 efficiently remobilized towards the seeds.

In the hyperaccumulator *S. alfredii*, Cd storage and distribution has been shown to depend on
the Ca status. Calcium deficiency seems to trigger a phloem transport system that reallocates
Cd from storage tissues (old leaves) to newly grown tissues (Tian *et al.*, 2016). In durum
wheat, Cakmak *et al.* (2000) observed a decrease in Cd reallocation with increasing Zn

995 concentration in the nutrient solution. To account for this interaction, they suggested a

996 competitive inhibition of Cd loading into phloem sap by Zn.

997 In rice, the rise in Cd concentration in the leaves is much slower than the flow of xylem sap. Indeed under direct observation by ¹⁰⁷Cd positron emission, the metal failed to reach leaf 998 blades after 36 h even though it had already reached the panicles after 7 hours of exposure 999 1000 (Fujimaki et al., 2010). This suggests that Cd transport is decoupled from the xylem sap flow. In addition, Cd was singularly accumulated in the nodes and redistributed to the non-exposed 1001 1002 roots, thereby indicating a major role for phloem in the transport and distribution of Cd in the plant (Fujimaki et al., 2010). Kobayashi et al. (2013) provided evidence that Cd can be loaded 1003 1004 into the rice phloem at the level of first node immediately after the metal has been transported 1005 from the root. Phloem is also involved in Cd translocation from root to leaves in Solanum melongena and Tagetes erecta (Qin et al., 2013; Qin et al., 2015). 1006

1007 OsIRT1 is expressed in the rice phloem, in stem and leaves and has been suggested to 1008 transport Fe into phloem cells (Ishimaru et al., 2006). This transporter may therefore also contribute to the loading of Cd into the phloem. OsLCT1, the rice homologue of the low 1009 affinity cation transporter LCT1, identified in a wheat cDNA library (Clemens et al., 1998) is 1010 expressed in the nodes and leaf blades at rice reproductive stages (Table 1). This phloem Cd 1011 transporter would appear to be responsible for Cd delivery to rice grains, probably by 1012 1013 mediating xylem to phloem Cd transfer in the nodes and/or re-translocation of Cd from leaves to grains (Uraguchi et al., 2011; Uraguchi et al., 2014). In addition, Yamaji and Ma (2017) 1014 1015 suggested that OsHMA2, which is highly and constitutively expressed in the nodes, close to 1016 phloem in vascular bundles, could also be involved in mediating Cd (and Zn) transfer from 1017 the xylem sap to the phloem.

1018

F. Cadmium speciation in the phloem sap

Mendoza-Cozatl *et al.* (2008) found high ratios between PCs and Cd and between glutathione
(GSH) and Cd in the phloem sap of *B. napus*. They hypothesized that the metal could be
complexed with these compounds, which would then act as long-distance Cd carriers. Kato *et*

al. (2010) showed that, in rice phloem sap, Cd may bind to a 13 kDa protein distinct from
MT. Cd is also associated to SH compounds, probably PCs and GSH. Hazama *et al.* (2015)
studied the speciation of Cd in the phloem sap of the castor bean (*R. communis*). Based on
size-exclusion HPLC, they found that about 50% of the metal is bound to Cd-containing
proteins or peptides, mixtures of proteins and PCs (such as PC4 and PC3) as well as
complexes with GSH and PC2.

1028 X. Interaction of Cd with other elements

1029 Cadmium absorption by roots can be affected by the other elements in the soil solution, in 1030 three ways. First, as mentioned above, some anions such as Cl⁻, NO₃⁻, or SO₄⁻, can form complexes with Cd and modify the diffusion and the absorption of the metal. Moreover, 1031 major ions such as H^+ , Ca^{2+} and Mg^{2+} , present at concentrations several orders of magnitude 1032 1033 higher than those of Cd, decrease the plasma membrane surface potential, by compensating the negatively-charged lipid heads present at the external side of the plasma membrane. In the 1034 absence of negative charges, which would attract them, the local concentrations of trace 1035 cations, such as Cd²⁺, at the vicinity of the plasma membrane are reduced. The lower local 1036 concentration around the uptake transporters results in lower uptake (Kinraide, 1998; Wang et 1037 al., 2011; Wang et al., 2012). On the other hand, Cd^{2+} and other cations directly compete for 1038 transport. Using soybean in hydroponics, Cataldo et al. (1983) showed that root absorption of 1039 Cd^{2+} fell in the presence of Cu^{2+} , Fe^{2+} , Mn^{2+} , and Zn^{2+} , thus suggesting that these cations 1040 compete with Cd^{2+} to enter the root through the same transport systems. Hart *et al.* (2002) 1041 also observed that trace amounts of Zn in the exposure solution inhibit Cd uptake by roots of 1042 durum wheat (and vice versa), thereby indicating that the two ions compete for root uptake 1043 via the same transport systems. In contrast, Salt et al. (1997) found no competition of Mn²⁺ 1044 and Zn^{2+} with Cd^{2+} for root absorption by *B. juncea*. Sterckeman *et al.* (2011) exposed maize 1045

1046	root plants to a 1 μ M Cd solution with 5 μ M of another cation (Ca ²⁺ , Co ²⁺ , Cu ²⁺ , Fe ²⁺ , Mg ²⁺ ,
1047	Mn^{2+} , Na^+ , NH_4^+ , Ni^{2+} , Pb^{2+} or Zn^{2+}) and found a reduction in Cd uptake by Cu^{2+} only.

1048 The expression of many transporters of the ZIP family, such as ZNT1, is only detectable

1049 under Zn deficiency, except in Zn hyperaccumulating plants (Grotz *et al.*, 1998; Verbruggen

1050 *et al.*, 2009b). This up-regulation may account for the increase in Cd absorption observed

under Zn deficiency and the lower Cd accumulation when Zn is applied as a soil or foliar

1052 fertilizer (Saifullah et al., 2016; Qaswar et al., 2017). Zinc can also affect Cd translocation, as

these metals compete with each other for xylem loading by the metal pumping ATPase

1054 OsHMA2 in rice (Fontanili *et al.*, 2016). Such competition is likely to occur in many

1055 transporters, which are not Cd specific, such as HMA4 or CAX.

1056 There is a relationship between the Fe nutritional status of the plant and Cd accumulation

1057 (Chen *et al.*, 2019). As mentioned above, Fe deficiency induces the expression of IRT1,

1058 which is responsible for Cd absorption. The positive effect of Fe deficiency on Cd uptake was

1059 observed in *N. caerulescens* and in *A. thaliana* (Lombi *et al.*, 2002; Vert *et al.*, 2002).

1060 Conversely, Cd stress induces symptoms similar to those of Fe deficiency, in particular the

1061 biosynthesis and secretion of phytosiderophores, which improve Fe uptake and provides an

advantage under Cd stress relative to Fe acquisition (Astolfi et al., 2014). He et al. (2017)

showed that supplying A. *thaliana* with Fe^{2+} prevents Cd uptake by inhibiting IRT1

1064 expression and favouring Fe^{2+} in its completion with Cd^{2+} . The Fe transporter, NRAMP also

transports Cd (Thomine et al., 2000; Takahashi et al., 2011), which suggest a possible

1066 competition between the two elements at this level (Chen *et al.*, 2019).

1067 Cadmium also interacts with the Fe plaque that is deposited at the root surface under certain

1068 conditions. An Fe plaque is commonly formed in various aquatic and terrestrial plant species,

such as rice (*O. sativa*). Radial loss of oxygen from the root and oxidation by microorganisms

1070 cause the oxidation of ferrous to ferric iron, which precipitates at the root surface as iron 1071 hydroxide (Liu *et al.*, 2007). The iron plaque is expected to sequester trace metals through adsorption or co-precipitation, thereby affecting the availability and uptake of these elements 1072 1073 by the root. However, the effect of the iron plaque on Cd uptake is controversial. Liu et al. (2007); Liu et al. (2008) found that the iron plaque is of little significance for Cd uptake and 1074 1075 accumulation by rice plants and suggested instead that root tissues control Cd uptake in rice. In contrast, Cheng et al. (2014) and Cao et al. (2018b) showed that the iron plaque indeed 1076 1077 promotes Cd deposition onto root surface and limits metal accumulation in rice plants. 1078 Sebastian and Prasad (2016) also suggested that iron plaque would act as a barrier against Cd exposure, but also as an Fe source preventing Cd stress and Cd-induced Fe deficiency. 1079 1080 However, Zhou et al. (2018) found a positive correlation between Cd in iron plaque and Cd in 1081 rice plant tissues, suggesting that the plaque does not prevent Cd accumulation by the plant. However, the same group of authors (Zhou et al., 2015) previously reported a negative 1082 1083 correlation between Cd in the iron plaque and that in the roots of 32 rice cultivars, suggesting 1084 a protective role of the plaque.

Nitrogen speciation also influences Cd root uptake independently of its effect on rhizosphere 1085 1086 or root apoplast pH (see above). Indeed, in hydroponics, supplying N. caerulescens with NO₃⁻ resulted in a doubling of the Cd concentration in the shoots compared with NH₄⁺ supply. 1087 1088 Moreover, when cultivating the same species in soil, Cd uptake was highest in the NO₃⁻ 1089 treatment, despite the fact that a higher pH decreases Cd and Zn bioavailability in the 1090 rhizosphere under these conditions (Xie et al., 2009). Nitrate also enhances Cd uptake in tomato (Solanum lycopersicum) (Luo et al., 2012), potato (S. tuberosum) (Larsson Jönsson 1091 1092 and Asp, 2013) and S. alfredii (Hu et al., 2013) grown in hydroponics. However, the mechanisms accounting for this positive effect of NO₃⁻ on Cd absorption are still unclear. It 1093 has been proposed that the uptake of NO₃⁻ would favour Cd uptake by increasing membrane 1094

potential, while that of NH_4^+ would depolarize the membrane (Monsant *et al.*, 2010). More recently, based on results obtained with rice, Yang *et al.* (2016) proposed that NO_3^- enhances Cd uptake by up-regulating the expression of IRT1.However, in the Cd accumulators *S. nigrum* and *Carpobrotus rossii*, shoot Cd accumulation was 30% higher with NH_4^+ than with NO_3^- in the nutrient solution (Cheng *et al.*, 2016).

Some of the works cited above suggest that N forms can affect Cd absorption not only by the root, but also its translocation (Hu *et al.*, 2013; Larsson Jönsson and Asp, 2013; Cheng *et al.*, 2016), through as yet unknown mechanisms. Li *et al.* (2010) reported that Cd regulates the low affinity nitrate transporter AtNRT1.8, which controls nitrate concentration in the xylem sap. Moreover, a loss of the AtNRT1.8 function increased Cd sensitivity, thus suggesting cross-talk between Cd and nitrate translocation.

1106 Nocito et al. (2002) showed that the formation of PC-Cd complexes (see above) induces an increase in sulphate uptake by maize, in order to sustain the higher sulphur demand during PC 1107 synthesis. Moreover, Cd stimulates the expression of genes involved in sulphur assimilation 1108 pathways (Harada et al., 2002; Herbette et al., 2006; Ernst et al., 2008; Liang et al., 2014; 1109 Liang et al., 2016). Wu et al. (2018) found that an excess supply of sulphate increases Cd 1110 1111 accumulation in the leaves of Vicia faba. In rice, Yamaguchi et al. (2016) showed that Cd exposure up-regulated the transcription of the sulphate transporter SULTR1/2, as well as 1112 sulphate uptake and translocation to shoot. 1113

1114 Silicon (Si) is known to reduce Cd toxicity to plants (Nwugo and Huerta, 2008; Vieira da

1115 Cunha et al., 2008; Vaculík et al., 2009; Wang et al., 2015c; Rahman et al., 2017). It was

- reported to decrease the availability of Cd in soil (Liang *et al.*, 2005; Vieira da Cunha *et al.*,
- 1117 2008) and decrease Cd absorption and/or translocation (Liang *et al.*, 2005; Wang *et al.*,
- 1118 2015c; Wu *et al.*, 2016b; Rahman *et al.*, 2017; Shao *et al.*, 2017), although not systematically
- 1119 (Vaculík et al., 2009). Silicon improves Cd sequestration and complexation, restores the

1120 redox balance and strengthens the antioxidant capacity of the plant (Farooq *et al.*, 2016).

1121 Alleviation of Cd toxicity by Si has also been attributed to enhanced binding of Cd to the root

apoplast (Vaculík *et al.*, 2012; Wu *et al.*, 2016b). More recently, the work of Greger *et al.*

1123 (2016) on wheat (*T. aestivum*) showed that SiO₃ down-regulated LCT1 and HMA2 and up-

regulated PCS1. IRT1 was down-regulated by Cd and up-regulated by Si in roots and shoot.

1125 Silicon had no effect on the expression of NRAMP1. Shao et al. (2017) showed that

1126 OsNRAMP5 and OsHMA2 were down-regulated by Si, which could explain the reduction in

1127 Cd uptake and translocation in rice exposed to silicic acid.

1128 XI. Conclusions

Knowledge of Cd accumulation mechanisms in plants has considerably progressed over recent decades. However, many uncertainties remain and this knowledge is still incomplete at various levels. Cadmium speciation in the rhizosphere solution, which controls the root uptake, is poorly understood, because of the difficulties in identifying and quantifying the species, particularly the organometallic complexes. There is also a need to better characterize the reactive sites in the soil solid phase in order to accurately predict Cd solubility.

Concerning the processes responsible for the distribution of the metal in the plant tissues and organs, much is known of the mechanisms and the factors (such as the transport systems) controlling root absorption and storage, and xylem loading. In contrast, xylem unloading, transport and storage in the stem and leaf tissues, as well as reallocation through phloem loading and transport towards the reproductive organs are still poorly understood. It is likely that future work will shed new light on the mechanisms of Cd distribution in aerial parts and especially to fruits and seeds.

1142 The diversity of the transport systems involved in Cd uptake and translocation is consistent1143 with the idea that Cd, as a non-essential element, highjacks the pathways of certain nutrients,

such as calcium, zinc, manganese or iron to accumulate in plants. Moreover, the pathways used by Cd differ according to plant species. Investigations in the model species *A. thaliana* and in rice have revealed striking differences in the uptake mechanisms, as well as commonalities in the mechanisms that allow Cd sequestration in roots. There will be a need to extend this knowledge to other species, including species with contrasting accumulations and crops such as lettuce, cocoa and tobacco, for which Cd accumulation is an important public health issue. Moreover, the precise location of root Cd carriers at the tissue scale, will need to

1151 be determined in the future in order to better describe Cd uptake paths into plants.

Among crop species, rice has been the subject of recent major advances. Whilst rice produces the staple diet of hundreds of millions of people in the world, it is the cereal that accumulates Cd at the highest level in its grains. These scientific advances are about to lead to the release of rice cultivars that accumulate much less Cd. In contrast, few results have been obtained on bread wheat (*T. aestivum*) and durum wheat, whereas these cereals are responsible for the chronic exposure to Cd of a very large portion of the world population (EFSA, 2012). In the future, research should focus on the biomolecular mechanisms of Cd accumulation in wheat.

Recent works have suggested the involvement of new actors in Cd transport in plants, such as 1159 1160 CAL1, a defensin-like protein, which would appear to chelate the metal and favour its xylem loading, or again the homeodomain-containing protein gene, OsHB4a, which regulates several 1161 genes involved in Cd uptake and translocation. The regulation of OsHB4a by the miR166 also 1162 indicates that other mechanisms involving micro-RNA may be important for the movement of 1163 Cd in a plant. The effects of Si or of N forms on Cd accumulation are relatively new 1164 discoveries. These correspond to complex processes, which are as yet only partly explained, 1165 but may help to improve the management of Cd-contaminated land. All these topics represent 1166 avenues of research leading to the possibility of reducing Cd accumulation in crops. 1167

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2309 Figure captions

2310	Figure 1. Principal reactions and processes controlling cadmium availability at the root-soil
2311	interface. PM: Plasma membrane; L _S : Organic ligand originating from the soil; Lp: organic
2312	ligand originating from rhizodeposition. For simplification, the cell wall is not represented.
2313	Figure 2. Schematic representation of Arabidopsis thaliana root with the gene/proteins
2314	involved in Cd transport. For simplification, the cell wall is not represented.
2315	Figure 3. Schematic representation of a rice (Oryza sativa) root with the gene/proteins
2316	involved in Cd transport. For simplification, the cell wall is not represented. Question mark
2317	after a protein name means it might not be expressed in the tissue, although it has been found
2318	in rice roots.
2319	Figure 4. Schematic representation of <i>Noccaea caerulescens</i> root with the gene/proteins
2320	involved in Cd transport. For simplification, the cell wall is not represented.
2321	Figure 5. Schematic representation of leaf tissues with the transporters involved in Cd
2322	distribution. The prefix before the gene name indicates the plant species in which it was
2323	observed. Nc: Noccaea caerulescens, Os: Oryza sativa. No prefix means that the
2324	gene/transport protein has been observed in several plant species. For simplification, the cell
2325	wall is not represented.
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2331	Table 1.	Genes	invol	ved	in (Cd	accumu	lation	in	plants.
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Function	Name	Annotation	Organ of expression	Plant species	Co- transported	Factor of	References
Uptake into cell	IRT1	ZIP family	Roots	Pisum sativum, Arabidopsis thaliana, Hordeum vulgare, Oryza sativa	Co ²⁺ , Fe ²⁺ , Fe ³⁺ , Mn ²⁺ , Zn ²⁺	Fe deficiency	Cohen <i>et al.</i> (1998); Korshunova <i>et al.</i> (1999); Nakanishi <i>et al.</i> (2006); Pedas <i>et al.</i> (2008); Barberon <i>et al.</i> (2014)
	IRT2		Roots	O. sativa	Fe ²⁺	Fe deficiency	Nakanishi et al. (2006)
	ZNT1		Roots, shoot	Noccaea caerulescens	Zn ²⁺	Zn deficiency	Pence <i>et al.</i> (2000); Milner <i>et al.</i> (2012); Lin <i>et al.</i> (2016a)
	NRAMP1	natural resistance- associated macrophage protein	Roots(+), shoot(-)	A. thaliana ¹ , O. sativa, Malus baccata	Fe, Mn	Fe and Mn deficiency	Thomine et al. (2000); Xiao et al. (2008); Cailliatte et al. (2010); Takahashi et al. (2011); Ishimaru et al. (2012)
	NRAMP5		Roots	O. sativa, Triticum polonicum, T. aestivum, Zea mays	Mn, Fe, Co		Ishimaru <i>et al.</i> (2012); Sasaki <i>et al.</i> (2012); Tang <i>et al.</i> (2017); Sui <i>et al.</i> (2018)
	GLR3.4	glutamate receptors	Roots	O. sativa	Са		Chen <i>et al.</i> (2018b)
	ANN4	annexin	Roots	O. sativa	Са		Chen <i>et al.</i> (2018b)
Efflux out of cell	MTP9	cation diffusion facilitator (CDF)	Root endodermis	Cucumis sativus	Mn	Mn and Cd excess	Migocka <i>et al.</i> (2015)
	PDR8	ATP binding cassettes (ABC)	Roots(++) (hair and	A. thaliana	Pb ²⁺		Kim <i>et al.</i> (2007)

			epidermis), shoot				
Chelation	MTL	MT-like protein	Roots	Sedum plumbizincicola			Peng et al. (2017a)
	CDT1	Cys-rich peptide	Roots, shoots (cell surface, cell wall)	Digitaria ciliaris, O. sativa		Cd exposure	Kuramata <i>et al</i> . (2009)
	CAL1	Defensin like protein	Roots	O. sativa			Luo <i>et al.</i> (2018)
Uptake into the vacuole	HMA3	P type ATPase family, subfamily HMA (for Heavy Metal Associated)	Roots, shoot	A. thaliana, O. sativa, N. caerulescens, S. plumbizincicola, Brassica rapa, Triticum durum	Co, Pb, Zn		Morel <i>et al.</i> (2009); Miyadate <i>et al.</i> (2011); Ueno <i>et al.</i> (2011a); Ueno <i>et al.</i> (2011b); Yan <i>et al.</i> (2016); Liu <i>et al.</i> (2017); Shao <i>et</i> <i>al.</i> (2018); Maccaferri <i>et al.</i> (2019); Sui <i>et al.</i> (2019); Zhang <i>et al.</i> (2019)
	MRP3	ATP binding cassettes (ABC)	Roots, shoot	A. thaliana ¹ , H. vulgare			Tommasini <i>et al.</i> (1998); Schneider <i>et al.</i> (2009)
	MRP7		Roots, shoot	A. thalian a^2			Wojas <i>et al.</i> (2009)
	ABCC1		Roots, shoot	A. thaliana, H. vulgare	As		Park <i>et al.</i> (2012); Song <i>et al.</i> (2014)
	ABCC2		Roots, shoot	A. thaliana, H. vulgare	As		Park <i>et al.</i> (2012); Song <i>et al.</i> (2014)
	ABCC3		Roots, shoot	A. thaliana		Cd exposure	Brunetti et al. (2015)
	CAX2	CAtion eXchanger	Roots, shoot	A. thaliana ²	$Ca^{2+}, Zn^{2+}, Mn^{2+}$		Pittman <i>et al.</i> (2004); Koren'kov <i>et al.</i> (2007)
	CAX4		Roots, shoot	A. thaliana ²	Ca ²⁺ , Zn ²⁺ , Mn ²⁺		Koren'kov <i>et al.</i> (2007)

Efflux out of the vacuole	NRAMP3	natural resistance- associated macrophage protein	Roots, shoot	A. thaliana, N. caerulescens	Fe, Mn	Fe deficiency	Oomen <i>et al.</i> (2009) Thomine et al. (2000)
	NRAMP4		Roots, shoot	A. thaliana, N. caerulescens	Fe, Mn, Zn	Fe deficiency	Oomen <i>et al.</i> (2009) Thomine et al. (2000)
Xylem loading	HMA2	P type ATPase family, subfamily HMA (for Heavy Metal Associated)	Roots(++), shoot	A. thaliana, O. sativa	Zn		Wong and Cobbett (2009); Wong <i>et al.</i> (2009); Nocito <i>et al.</i> (2011); Satoh- Nagasawa <i>et al.</i> (2012); Takahashi <i>et al.</i> (2012); Yamaji <i>et al.</i> (2013)
	HMA4		Roots(++), shoot	A. thaliana, A. halleri, N. caerulescens	Zn, Pb, Cu	Zn and Mn exposure, downregulation by Cd exposure	Papoyan and Kochian (2004); Verret et al. (2004); Mills et al. (2005); Courbot et al. (2007); Hanikenne et al. al. (2008); Ó Lochlainn et al. (2011); Craciun et al. (2012)
	CCX2	cation/calcium (Ca) exchanger	Root, Shoot (node++)	O. sativa	Са		Hao <i>et al.</i> (2018)
Phloem loading	LCT1	low affinity cation transporter	Shoot	O. sativa	K, Mg, Ca, Mn		Uraguchi <i>et al.</i> (2011); Uraguchi <i>et al.</i> (2014)
	HMA2	P type ATPase family, subfamily HMA (for Heavy Metal Associated)	Shoot (node)	O. sativa			Yamaji and Ma (2017)

Figure 1. Principal reactions and processes controlling cadmium availability at the root-soil interface. PM: Plasma membrane; L_s: Organic
ligand originating from the soil; Lp: organic ligand originating from rhizodeposition. For simplification, the cell wall is not represented.

CYTOSOL PM SOIL SOLUTION SOIL SOLID PHASE 2338 Ls **CHELATION** 2339 DISSOCIATION $Cd^{2+} + L_s \leftarrow$ Cd²⁺ Cd²⁺ € $Cd-L_s \leftarrow$ 2340 Negative charges Cd²⁺ \rightarrow L_P + Cd²⁺ → Cd-L_P Cd²⁺ 2341 ABSORPTION DIFFUSION ADSORPTION 2342 Cd²⁺ € Cd²⁺ ← **⇒ Cd**²⁺ **ADVECTION** 2343 [09] Cd concentration gradient **CLAY MINERALS** 2344 **IRON HYDROXIDES** 2345 **ORGANIC MATTER** 2346 Distance to the root surface 3 mm

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Figure 2. Schematic representation of *Arabidopsis thaliana* root with the gene/proteins involved in Cd transport. For simplification, the cell wall







Figure 3. Schematic representation of a rice (*Oryza sativa*) root with the gene/proteins involved in Cd transport. For simplification, the cell wall is not represented. A question mark after a protein name means it might not be expressed in the tissue, although it has been found in rice roots.

Figure 4. Schematic representation of the hyperaccumulator *Noccaea caerulescens* root with the gene/proteins involved in Cd transport. For



2373 simplification, the cell wall is not represented.

Figure 5. Schematic representation of leaf tissues with the transporters involved in Cd distribution. The prefix before the gene name indicates the

2384 plant species in which it was observed. Nc: Noccaea caerulescens, Os: Oryza sativa. No prefix means that the gene/transport protein has been



2385 observed in several plant species. For simplification, the cell wall is not represented.