

# Characterization of a chemical amplifier for peroxy radical measurements in the atmosphere

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- 2 atmosphere
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### 12 Abstract

13 Peroxy radicals (HO<sub>2</sub> and RO<sub>2</sub>) are key species in atmospheric chemistry. They are produced during the oxidation of volatile organic compounds (VOCs) and are involved in the formation of photochemical 14 pollutants such as ozone (O<sub>3</sub>) and secondary organic aerosols (SOA). However, ambient measurements of 15 these reactive species are still challenging and only a few techniques can achieve both a good selectivity 16 and a detection limit that is low enough for ambient measurements. In this publication we present the 17 characterization of a Chemical Amplifier (CA) using two different approaches for ambient measurements 18 of peroxy radicals, including the PEroxy Radical Chemical Amplifier (PERCA) and the Ethane based 19 CHemical AMPlification (ECHAMP). At 50% Relative Humidity (RH), the experimental CL for PERCA 20 was found to be higher by approximately a factor of 3.7 compared to ECHAMP. The RH-dependence of 21 the CL was also found to be larger for PERCA by a factor of 1.12. Box modeling of the chemistry taking 22 place in the instrument highlighted that the formation of HNO<sub>3</sub> from the HO<sub>2</sub>+NO reaction has a strong 23 24 impact on the CL for both approaches. In addition, experiments conducted to quantify the  $RO_2$ -to-HO<sub>2</sub> 25 conversion efficiency for a large range of organic peroxy radicals confirmed that it mainly depends on the organic nitrate (RONO<sub>2</sub>) formation yield, while the alkyl nitrite (RONO) yield is not limiting the CL in 26 27 most cases. Ambient measurements using the PERCA approach are shown to illustrate the performances 28 of this new instrument.

29 Keywords: Instrument development; peroxy radical chemical amplifier; PERCA; ECHAMP; atmosphere

#### 30 1 Introduction

Due to their high reactivity, and therefore short lifetime, peroxy radicals (HO<sub>2</sub>, and RO<sub>2</sub>,  $R = C_x H_y O_z$ ) are key species in atmospheric chemistry. These radicals are mainly produced when the OH radical reacts with CO, the abundant greenhouse gas methane and volatile organic compounds (VOCs). They are also formed through the photolysis of carbonyl compounds, reactions of alkenes with ozone, and some nighttime reactions involving the NO<sub>3</sub> radical and VOCs (Finlayson-Pitts and Pitts, 2000).

36 In the atmosphere, there is a rapid propagation chemistry between peroxy radicals, alkoxy radicals (RO) and the hydroxyl radical (OH), forming a group of species known as "ROx" (ROx  $\equiv$  HO<sub>2</sub> + RO<sub>2</sub> + RO + 37 OH). Of particular interest, reactions of peroxy radicals with nitric oxide (NO) play a central role in 38 tropospheric chemistry. Indeed, peroxy radicals propagate to OH, which sustains the oxidizing capacity of 39 the atmosphere, and which converts NO into nitrogen dioxide (NO<sub>2</sub>), whose photolysis leads to ozone 40 formation. Organic peroxy radicals also play a key role as intermediates in the formation of secondary 41 42 organic aerosols (SOA) (Kroll and Seinfeld, 2008). Both of these secondary pollutants, O<sub>3</sub> and SOA, are known to cause severe health effects. Together with OH, peroxy radicals therefore control the global 43 44 oxidizing capacity of the Earth's atmosphere, and as a consequence, the concentration and distribution of greenhouse gases and secondary pollutants. Understanding the spatial and temporal variability of these 45 radicals and being able to model this variability is key to evaluate future changes in the chemical 46 47 composition of the atmosphere, with implications for both air quality and climate change.

Despite their important role in the atmosphere, the chemistry of peroxy radicals has yet to be fully 48 understood under a wide range of atmospheric conditions, including remote locations (marine boundary 49 50 layer) (Berresheim et al., 2002; Creasey et al., 2003), continental low-NO<sub>x</sub> regions influenced by biogenic emissions (Archibald et al., 2011; Griffith et al., 2013; Wolfe et al., 2011), polluted urban areas (Dusanter 51 52 et al., 2009; Volkamer et al., 2010), and polar regions (Mao et al., 2010; Read et al., 2008). Some of these studies have reported significant differences between measured and modeled concentrations of RO<sub>x</sub> 53 species likely due to an incomplete characterization of ambient trace gases and the use of incomplete 54 55 chemical mechanisms in models. For instance, peroxy radical measurements can be significantly lower (Griffith et al., 2013) or higher (Wolfe et al., 2014) than model predictions in forested areas and 56 57 significant missing OH sinks, likely leading to an underestimation of the OH-to-peroxy radical

propagation rate, have been reported for forested (Griffith et al., 2013; Heard and Pilling, 2003; Stone et al., 2012) and urbanized (Dusanter et al., 2009; Griffith et al., 2016; Volkamer et al., 2010) areas.

Measuring peroxy radicals is particularly difficult due to their reactive nature and their low ambient 60 concentrations, which requires highly sensitive techniques with inlets designed to avoid the loss of the 61 targeted radicals. Various instruments using different principles have been developed during the last 62 decades to perform these measurements, including Matrix Isolation Electron Spin Resonance spectroscopy 63 64 (MIESR: HO<sub>2</sub>, RO<sub>2</sub>) (Mihelcic et al., 2003, 1985), Laser-Induced Fluorescence (LIF-FAGE: HO<sub>2</sub>, HO<sub>2</sub>\*; RO<sub>x</sub>LIF: HO<sub>2</sub>, HO<sub>2</sub>+RO<sub>2</sub>) (Stevens et al., 1994; Fuchs et al., 2008; Dusanter et al., 2009; Faloona et al., 65 2004), Chemical Ionisation Mass Spectrometry using PERCIMS (HO<sub>2</sub>, HO<sub>2</sub>+RO<sub>2</sub>) (Edwards et al., 2003; 66 Hornbrook et al., 2011; Kukui et al., 2008) and Br-CIMS (HO<sub>2</sub>) (Sanchez et al., 2016), PEroxy Radical 67 Chemical Amplification (PERCA: HO<sub>2</sub>+RO<sub>2</sub>) (Cantrell et al., 1984; Hernández et al., 2001; Kartal et al., 68 69 2010; Liu et al., 2009; Wood and Charest, 2014), and the more recent Ethane based CHemical 70 AMPlification (ECHAMP) technique (Wood et al., 2017).

71 PERCA is an indirect measurement method pioneered by Cantrell and Stedman (Cantrell et al., 1984), which measures the sum of peroxy radicals continuously with high sensitivity. Ambient air is sampled in a 72 reactor where each peroxy radical will lead to the formation of several NO<sub>2</sub> molecules via chain reactions 73 74 after the addition of high concentrations of NO and CO. Once all the peroxy radicals have been consumed, NO<sub>2</sub> is measured by a suitable detector based on luminol chemiluminescence (Clemitshaw et al., 1997), 75 LIF (Sadanaga et al., 2004), cavity ring-down spectroscopy (CRDS) (Horstjann et al., 2014; Liu et al., 76 77 2009; Liu and Zhang, 2014) or cavity attenuated phase-shift spectroscopy (CAPS) (Wood and Charest, 2014). 78

79 The amplification chemistry resulting from the addition of NO and CO is shown below:

80	$RO_2 + NO \rightarrow RO + NO_2$	R1a
81	$RO_2 + NO + M \rightarrow RONO_2 + M$	R1b
82	$RO + O_2 \rightarrow R_{-H}O + HO_2$	R2
83	$RO + NO + M \rightarrow RONO + M$	R3
84	$HO_2 + NO \rightarrow OH + NO_2$	R4a
85	$HO_2 + NO + M \rightarrow HNO_3 + M$	R4b

86	$HO_2 + wall \rightarrow loss$	R5
87	$OH + CO (+O_2) \rightarrow \cdots \rightarrow CO_2 + HO_2$	R6

88 
$$OH + NO + M \rightarrow HONO + M$$
 R7

89  $OH + wall \rightarrow loss$  R8

90 With the addition of a large amount of NO in the sampling reactor, RO<sub>2</sub> radicals are converted into HO<sub>2</sub> through the formation of an alkoxyl radical (RO) (R1a & R2). This HO<sub>2</sub>, together with the sampled HO<sub>2</sub>, 91 92 are converted into OH (R4a). During this process, several molecules of  $NO_2$  are produced, depending on 93 the competition between propagation (R1a, R2, R4a) and termination (R1b, R3, R4b, R5) reactions. Then, 94 several fast interconversions between OH and HO<sub>2</sub> occurs through the reaction of OH with CO (R6) and 95 the reaction of HO<sub>2</sub> with NO (R4a), leading to the formation of additional molecules of NO<sub>2</sub>. The number of NO<sub>2</sub> molecules formed during these interconversion steps is called chain length (CL) and is limited by 96 the competition occurring between propagation (R4a, R6) and termination (R4b, R5, R7, R8) reactions of 97 both OH and HO<sub>2</sub>. The CL is typically in the range of 100–200 under dry conditions (Clemitshaw et al., 98 99 1997; Sadanaga et al., 2004; Wood and Charest, 2014).

100 The most problematic issue concerning PERCA instruments is the chain length dependence on Relative 101 Humidity (RH), which translates into a RH-dependent sensitivity (Mihele and Hastie, 2000, 1998; Sadanaga et al., 2004). The chain length decreases monotonically with RH, likely due to increased losses 102 of radicals on wet surfaces (Mihele et al., 1999; Miyazaki et al., 2010) and an increase in the HNO3 yield 103 in reaction R4b. The latter may involve H<sub>2</sub>O dimers or the formation of a HO<sub>2</sub>-H<sub>2</sub>O complex (Reichert et 104 al., 2003). The effect of water vapor on the branching ratio of R4b was investigated by Butkovskaya et al. 105 106 (Butkovskaya et al., 2009) who reported a linear increase of the HNO<sub>3</sub> yield with RH, with an enhancement factor of about 8 at 50% RH. A rate constant for the reaction between HO<sub>2</sub>-H<sub>2</sub>O and NO 107 108 producing HNO<sub>3</sub> was estimated from this work and was found to be 40 times faster than the rate constant for HO<sub>2</sub>+NO $\rightarrow$ HNO<sub>3</sub>. This water vapor enhancement of the gas-phase termination rate of HO<sub>2</sub> may play 109 an important role in the reduction of the chain length. 110

The amplification chemistry used in ECHAMP relies on the addition of ethane  $(C_2H_6)$  instead of CO as proposed by Wood et al. (Wood et al., 2017). The advantages reported by the authors are a lower sensitivity of the CL to relative humidity and a safer use of the instrument in confined areas since ethane is less toxic than CO. The chemical reactions describing the radical cycling when ethane and NO are added in the reactor are the same as above (R1-R8), with the exception than R6 is replaced by the set of reactionsshown below:

117 
$$0H + C_2H_6(+O_2) \to \dots \to H_2O + C_2H_5O_2$$
 R9

118 
$$C_2H_5O_2 + NO \rightarrow C_2H_5O + NO_2$$
 R10a

119 
$$C_2H_5O_2 + NO \rightarrow C_2H_5ONO_2$$
 R10b

120 
$$C_2H_5O + O_2 \rightarrow CH_3CHO + HO_2$$
 R11

121 
$$C_2H_5O + NO \rightarrow C_2H_5ONO$$
 R12

Ethane reacts with OH to produce an ethylperoxyl radical (C<sub>2</sub>H<sub>5</sub>O<sub>2</sub>) (R9), which then (i) propagates to 122 123 HO<sub>2</sub> (R10a, R11) through the formation of an alkoxyl radical (C<sub>2</sub>H<sub>5</sub>O) or (ii) terminates through the formation of organic nitrate (R10b) or nitrite (R12) compounds. Since the interconversion steps between 124 OH and HO<sub>2</sub> go through the formation of  $C_2H_5O_2$  and  $C_2H_5O$ , the additional loss of radicals (R10b, R12) 125 126 will lead to a lower CL than for the PERCA approach. It has been shown that this amplification chemistry leads to chain lengths ranging from 25-30 (RH of 1-10%) (Wood et al., 2017). However, if heterogeneous 127 128 losses of peroxy radicals on the reactor wall contribute significantly to the total termination rate of radicals in the reactor, the CL for ECHAMP should be less sensitive to RH compared to PERCA since C<sub>2</sub>H<sub>5</sub>O<sub>2</sub> 129 exhibits a lower loss rate on wet surfaces than HO<sub>2</sub> (Mihele et al., 1999). For example, Mihele et al. 130 (1999) reported a wall loss rate on Teflon (PFA) of  $2.8 \pm 0.2$  s<sup>-1</sup> for HO<sub>2</sub> and  $0.8 \pm 0.1$  s<sup>-1</sup> for C<sub>2</sub>H<sub>5</sub>O<sub>2</sub> 131 under dry conditions and showed that the HO<sub>2</sub> loss rate increases with RH in contrast to organic peroxy 132 radicals. This observation was also confirmed by Miyazaki et al. (2010) who reported a 6-fold higher 133 134 removal efficiency of HO<sub>2</sub> on Teflon (PFA) surfaces compared to organic peroxy radicals.

PERCA and ECHAMP instruments used in the field are based on dual-channel sampling systems (Liu et 135 al., 2009; Horstjann et al., 2014; Wood et al., 2017; Kartal et al., 2010) to simultaneously quantify NO<sub>2</sub> 136 under two measurement modes (background and amplified modes). In the background mode, NO and N2 137 are injected at the entrance of the sampling reactor and CO or ethane (same volumetric flow rate than N<sub>2</sub>) 138 is injected at the exit of the reactor once all the radicals have been terminated. Since CO or ethane are not 139 added together with NO, there is no amplification chemistry. The NO<sub>2</sub> mixing ratio exiting the reactor is 140 the sum of ambient NO<sub>2</sub>, a fraction of O<sub>3</sub> titrated by NO, and the amount of NO<sub>2</sub> produced during the first 141 peroxy radical-to-OH conversion step (R1a, R2, R4a). For the amplification mode, CO or ethane is 142

injected at the entrance of the reactor while N<sub>2</sub> is injected at the exit, allowing the amplification chemistry 143 of peroxy radicals to take place. The mixing ratio of NO<sub>2</sub> at the exit of the reactor is the sum of that 144 145 observed during the background mode and the NO<sub>2</sub> produced during amplification. The amount of NO<sub>2</sub> generated during amplification ( $\Delta NO_2$ ) is inferred from the difference between the amplification and 146 background modes. 147

148 It is important to note that these instruments will amplify all radicals that can propagate to HO<sub>2</sub>, including 149 RO<sub>2</sub> but also OH and RO. While the measurements of a chemical amplifier represent the sum of RO<sub>x</sub>, tropospheric concentrations of OH and RO are ~200-1000 times lower than those of HO2 and RO2 and the 150 measurements are thus considered as being the sum of HO<sub>2</sub> and RO<sub>2</sub>. 151

The total mixing ratio of peroxy radicals calculated from the amount of NO<sub>2</sub> produced during the 152 153 amplification stage ( $\Delta NO_2$ ) and the CL is retrieved as shown in Eq. 1.

154

$$[HO_2] + \sum [RO_2] = \Delta NO_2 / CL$$
Eq. 1

Quantifying the sum of peroxy radicals from Eq. 1 requires calibrating the CL by generating a known 155 concentration of HO<sub>2</sub> or another peroxy radical. The CL is usually only calibrated for a few radicals by 156 157 generating a known concentration of methyl peroxy (CH<sub>3</sub>O<sub>2</sub>) and peroxyacetyl (CH<sub>3</sub>C(O)O<sub>2</sub>), both being simultaneously produced from the photolysis of acetone at 254 nm (Miyazaki et al., 2010; Wood and 158 Charest, 2014). Another method used to calibrate the chain length is based on the generation of  $CH_3O_2$ 159 160 from the photolysis of methyl iodide (CH<sub>3</sub>I) in air at 254 nm (Clemitshaw et al., 1997; Green et al., 2006).

161 This publication reports the construction of a dual-channel chemical amplifier for ambient measurements of peroxy radicals using both the PERCA (CO/NO) and ECHAMP (ethane/NO) approaches. The main 162 objective was to improve our understanding of these two approaches, which will ultimately provide a 163 164 better assessment of uncertainties associated to chemical amplifiers' measurements. In this context, optimal operating conditions were identified for each approach and experimental observations were 165 compared to box-model simulations to gauge our understanding of the radical amplification chemistry. 166 The instrument was calibrated using a large range of peroxy radicals, including HO<sub>2</sub> and organic peroxy 167 radicals formed during the OH-oxidation of alkanes, alkenes, aromatics and oxygenated VOCs. These 168 experiments allowed investigating whether the change in CL observed experimentally is consistent with 169 170 the formation yield of organic nitrate (RONO<sub>2</sub>) or nitrite (RONO) species reported in the literature for

each tested radical. Finally, ambient measurements are reported to illustrate the performances of thisinstrument.

### **2 Experimental Section**

#### **174 2.1 Description of the chemical amplifier**

A schematic of the apparatus built at IMT Lille Douai is shown in Figure 1. It consists of two sampling reactors operated in amplification and background modes, two monitors to simultaneously measure NO<sub>2</sub> at the exit of the reactors, and two sets of solenoid valves to switch (i) between background and amplification modes in each reactor (SV1a, SV1b, SV2a, SV2b) and (ii) the sampling of the NO<sub>2</sub> monitors between the two reactors (SV3, SV4) as further described below for the measurement sequence.

Ambient air is sampled into two PerFluoroAlkoxy (PFA) reactors (0.635 cm o.d. × 60 cm length, 180 thickness 0.08 cm) using two PFA inlets (0.635 cm o.d. × 1.5 cm length) at a total flow rate of 181 approximately 800 cm<sup>3</sup> min<sup>-1</sup>. The inlets and the reactors are connected together using home-made 3D 182 183 printed nylon couplers (Figure S1), allowing to add and mix the reagent gases with the sampled air. The reagent gases are mixed with ambient air through two circular channels (2 mm apart) between the 184 185 sampling inlet and the reactor, each channel being characterized by 5 radial injection holes. This design 186 was adopted to improve the mixing between ambient air and the reagents. All flow rates are regulated by mass flow controllers. 187

For the amplification channel, NO (84 SCCM-PERCA, 14 SCCM-ECHAMP, 50 ppm in N<sub>2</sub>) and CO (80 188 SCCM, 100%) or ethane (50 SCCM, 30% in N<sub>2</sub>) are added upstream the reactor via the nylon injector, 189 190 while a flow of nitrogen (equivalent to the CO or ethane flow rate) is added downstream the reactor, approximately 60 cm (0.9 s residence time) after the initial injection of CO (or ethane). For the 191 192 background mode, N<sub>2</sub> is added upstream while CO (or ethane) is added downstream. Four solenoid valves (SV1a, SV1b, SV2a, SV2b) allow switching the addition of CO (or ethane) and nitrogen between the 193 downstream and upstream injection ports, to change the measurement mode in each reactor. Flow rates of 194 195 CO (or ethane), NO and N<sub>2</sub> have been optimized to get the highest CL for both amplification chemistries and are discussed in the results section (Table 2). At the exit of the reactors, air exempt of RO<sub>x</sub> radicals is 196 transported through additional 6-m long pieces of PFA tubing (0.635 cm o.d.) to the NO<sub>2</sub> monitors at a 197 flow rate of 800 cm<sup>3</sup> min<sup>-1</sup>. 198

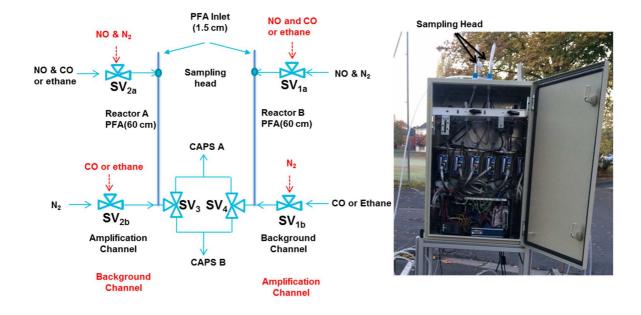






Figure 1: Schematic of the IMT Lille Douai PERCA instrument

High sensitivity CAPS (Cavity Attenuated Phase Shift) monitors from AERODYNE Research Inc. were coupled to the reactors to quantify NO<sub>2</sub>. For the monitors used in this work, air is sampled through a nafion dryer and enters an 82 cm<sup>3</sup> stainless steel absorption cell held at 27 °C, consisting of two high reflectivity mirrors (R ~ 99.99%). The photons provided to the absorption cell are emitted by a blue lightemitting diode (LED). The light exiting the cell passes through a bandpass filter centered at 450 nm and is detected by a vacuum photodiode. The NO<sub>2</sub> concentration is derived from the phase-shift observed in the detected signal, which is proportional to the amount of NO<sub>2</sub>.

209 All reagent gases were found to generate a spurious signal in the CAPS monitors. The addition of 1-6 ppm of NO (99.995% purity) in the reactors leads to several ppb of NO<sub>2</sub> due to contamination of the NO 210 mixture by NO<sub>2</sub> and a potential conversion of NO into NO<sub>2</sub> in the lines and mass flow controllers. To 211 reduce this spurious signal, the NO reagent gas was purified using an inline chemical filter made of 212 iron(II) sulfate heptahydrate (FeSO<sub>4</sub>.7H<sub>2</sub>O, Fisher scientific). This filter was found to be very efficient 213 with less than 0.2 ppb of NO<sub>2</sub> remaining in the reactor after the addition of 1 ppm of NO. The addition of 214 10% CO and 2% ethane in the reactors led to a spurious signal equivalent to approximately 1.5 and 3 ppb 215 of NO<sub>2</sub>. This signal was found to be proportional to the reagent concentration and is thought to be due to 216 light scattering since these species do not absorb at 450 nm. A lowest signal observed for CO seems 217

consistent with lower scattering cross-sections reported in the literature for this species compared to
alkanes (Sneep and Ubachs, 2005; Sutton and Driscoll, 2004).

A sequence of 4 steps of 90-s measurements spanning a 6-min period is required to measure a 220 concentration of peroxy radicals (Figure S2): (1) NO<sub>2</sub> is measured from each reactor, a reactor being 221 operated in background mode and the other reactor in amplification mode; (2) the CAPS sampling inlets 222 223 are switched between the two reactors; (3) both the measurement mode and the CAPS sampling are 224 switched between the 2 reactors; (4) the CAPS sampling inlets are switched. All NO<sub>2</sub> measurements performed under amplified conditions are averaged together and all measurements performed under 225 background conditions are also averaged together. Only the last 70-s of each NO<sub>2</sub> measurement step are 226 used in the averaging process to calculate  $\Delta NO_2$  (see S2). This sequence of steps is necessary to cancel out 227 228 any bias in  $\Delta NO_2$  arising from (i) a drift in monitors' zeros and (ii) slightly different behaviors of the 2 reactors (spurious NO<sub>2</sub> signals from reagents,  $O_3$  titration by NO, wall losses of  $O_3...$ ). 229

All compressed gas cylinders used in this study were provided commercially: CO (100%, Air Liquide, France), ethane ( $30\pm2\%$  in N<sub>2</sub>, Linde Co.), NO (50 ppm in N<sub>2</sub>, Air Liquide), N<sub>2</sub> (Messer, purity >99.9999%). Volatile organic compounds were purchased from Sigma Aldrich with a purity better than 99%.

#### 234 **2.2** Calibration of the Chain Length (CL)

As discussed in the introduction section, the CL can be quantified from Eq. 1 when the instrument samples air containing a known concentration of peroxy radicals. Peroxy radicals were generated in this study through the photolysis of water-vapor at 184.9 nm in a turbulent flow tube, creating an equimolar mixture of OH and HO<sub>2</sub> (Reactions R13-R14). This method is commonly used to calibrate field measurements of OH and HO<sub>2</sub> performed by LIF-FAGE instruments (Dusanter et al., 2008).

240 
$$H_2O + h\nu (\lambda = 184.9nm) \rightarrow OH + H$$
 R13

241 
$$H + O_2 + M \rightarrow HO_2 + M$$
 R14

In this study, OH was quantitatively converted into  $HO_2$  by adding CO in the calibrator as shown in R6 or by adding a volatile organic compound (VOC) to generate a mixture of  $HO_2$  and  $RO_2$  radicals (Lew et al., 2018). For instance, isoprene (C<sub>5</sub>H<sub>8</sub>) is added inside the calibrator to convert OH into an organic peroxy radical mixture of  $C_5H_8(OH)O_2$  isomers. The CO and VOC concentrations were set to convert approximately 99% of OH in less than 2 ms, ensuring a negligible OH loss on the calibrator inner surface. Diluted mixtures of each targeted VOC (Table 1) were prepared by injecting the pure compound into 18 liters of zero air (6-L stainless steel canisters at 3 bars). A flow rate of 2.5 SCCM (Standard Cubic Centimeter per Minute) of the diluted CO or VOC mixtures was mixed with 35 SLPM (Standard Liter per Minute) of dry or humid zero air in the calibrator using mass flow controllers. RH was varied in the range 10-90% (T=24±2°C) to produce peroxy radical concentrations within the range 2×10<sup>9</sup>-3×10<sup>10</sup> cm<sup>-3</sup>.

Series of calibration experiments were performed using CO and all the VOCs reported in Table 1. When CO is used, only HO<sub>2</sub> exits the calibrator and the measured CL is characteristic of only one peroxy radical. This CL is defined as  $CL_{(HO2)}$  in the following. However, when a VOC is added in the calibrator, an apparent CL is measured for a mixture of HO<sub>2</sub> and RO<sub>2</sub> radicals. Rearranging Eq. 1 leads to Eq. 2 for the calculation of this CL, which will depend on the generated RO<sub>2</sub> radical.

257 
$$CL = \Delta NO_2 / ([HO_2] + [RO_2])$$
 Eq.2

258 The radical concentrations at the exit of the calibrator are inferred from Eqs 3-4.

259 
$$[HO_2] = (1+X) \times [H_2O] \times \sigma_{water} \times \phi_{OH+H} \times (F \times t)$$
 Eq. 3

260 
$$[RO_2] = (1 - X) \times [H_2O] \times \sigma_{water} \times \phi_{OH+H} \times (F \times t)$$
Eq. 4

Here, HO<sub>2</sub> and RO<sub>2</sub> concentrations are calculated from the water concentration inside the calibrator, 261 [H<sub>2</sub>O], its absorption cross-section at 184.9 nm,  $\sigma_{water}$ , the photo dissociation quantum yield,  $\Phi_{OH+H}$ , the 262 photon flux, F, and the photolysis duration, t. The terms (1+X) and (1-X) account for the non-unity RO<sub>2</sub> 263 yield observed for some VOCs (Table 1), X being the prompt yield of HO<sub>2</sub>, which allows accounting for 264 its direct formation together with the formation of RO<sub>2</sub> radicals from RH+OH reactions. The absorption 265 cross section for H<sub>2</sub>O and quantum yield for OH+H reported in several studies are 7.14×10<sup>-20</sup> cm<sup>2</sup> 266 molecule<sup>-1</sup> (Cantrell et al., 1997; Creasey et al., 2000) and unity, respectively. The product (F×t) is 267 measured experimentally by ozone actinometry using Eq. 5, where  $[O_2]$  is the oxygen concentration,  $\sigma_{O2}$ 268 the oxygen absorption cross section at 184.9 nm  $(1.21 \times 10^{-20} \text{ cm}^{-2})$  (Dusanter et al., 2008), and [O<sub>3</sub>] the 269 ozone concentration produced in the calibration cell. The latter has to be measured during a calibration 270 271 experiment.

272 
$$(F \times t) = [0_3]/(2 \times [0_2] \times \sigma_{0_2})$$
 Eq.5

During a calibration experiment [O<sub>3</sub>] was measured as NO<sub>2</sub> by CAPS after titration by NO in one of the 273 reactors operated in background mode, keeping NO at the same mixing ratio than used for PERCA or 274 ECHAMP measurements of peroxy radicals. For this measurement, dry zero air was flown in the 275 calibrator and the mercury lamp was switched ON and OFF several times to determine [O<sub>3</sub>] by 276 subtraction. Ozone concentrations were in the range of 2.5-3.0 ppb. [H<sub>2</sub>O] was measured by a LI-COR 277 840A based on infrared absorption and calibrated versus a dew-point hygrometer (Mitchell S8000). The 278 CAPS instruments were calibrated with an NO<sub>2</sub> standard mixture at 190  $\pm$  3 ppb (2 $\sigma$ ) certified by LNE 279 280 (French National Metrology Institute).

The accuracy of the calibrated CL (Eq. 2) depends on the uncertainty of both the measured  $\Delta NO_2$  and the generated peroxy radical concentration. The precision on  $\Delta NO_2$  can be neglected since the use of large radical concentrations during calibration experiments leads to a negligible random error associated to the quantification of the NO<sub>2</sub> signal. If CO is added in the calibrator to only produce HO<sub>2</sub>, the accuracy (2 $\sigma$ ) on  $\Delta NO_2$  (1.5%) is much lower than the accuracy on the generated peroxy radical concentration of 31% (2 $\sigma$ ) as reported by Dusanter et al. (2008) for the same calibrator ran under similar operating conditions. A propagation in quadrature leads to a total accuracy of 31% (2 $\sigma$ ) for the CL.

The apparent chain length for individual  $RO_2$  radicals,  $CL_{(RO2)}$ , was computed from Eq. 6.  $\Delta NO_2$  is the amount of  $NO_2$  produced when both  $HO_2$  and  $RO_2$  radicals are sampled by the instrument,  $CL_{(HO2)}$  is determined during a calibration experiment using CO, and both  $HO_2$  and  $RO_2$  concentrations are calculated from Eqs. 3-5.

292 
$$CL_{(RO_2)} = (\Delta NO_2 - CL_{(HO_2)} \times [HO_2])/[RO_2]$$
 Eq. 6

Direct photolysis of the parent VOCs was investigated during these calibration experiments using dry conditions (RH< 1%). The fraction of  $\Delta NO_2$  due to radicals produced by VOC photolysis was quantified for each experiment and was subtracted from  $\Delta NO_2$  to calculate  $CL_{(RO2)}$  from Eq. 6. Table 1 reports the fraction of radicals produced by direct photolysis for each VOC. The amount of NO<sub>2</sub> produced by the radicals generated through the photolysis of VOCs was always lower than 7% of  $\Delta NO_2$  (photolysis of VOCs + water photolysis), with the exception of vinylacetate (14%). The difference between  $CL_{(HO2)}$  and the apparent  $CL_{(RO2)}$  should only depends on RO<sub>2</sub>-to-HO<sub>2</sub> propagation reactions (R1-R3), which can be factored into a RO<sub>2</sub>-to-HO<sub>2</sub> transmission term as shown in Eq. 7,  $T_{(RO2)}$ being the fraction of RO<sub>2</sub> radicals propagated to HO<sub>2</sub>.

302 
$$CL_{(RO_2)} = CL_{(HO_2)} \times T_{(RO_2)}$$
 Eq. 7

# **303 2.3 Box modeling of the amplification chemistry**

The Master Chemical Mechanism (MCM) v3.2 and the F0AM model (Framework for 0-D Modeling) 304 305 (Wolfe et al., 2016) were used to simulate both the amplification and background modes under different operating conditions of reagents and humidity. A radical wall loss for HO<sub>2</sub> was added in the chemical 306 307 mechanism with a first order loss rate parameterized from experimental measurements of radical losses in the PFA reactors (supplementary material S3). In addition, since MCM does not include RO+NO 308 reactions, which are not important for atmospheric chemistry but essential for the chemistry occurring in 309 310 the chemical amplifier, the reaction of  $C_2H_5O$  with NO leading to the formation of  $C_2H_5ONO$  was added using a rate constant of 4.4×10<sup>-11</sup> cm<sup>3</sup> molecule<sup>-1</sup> s<sup>-1</sup> (Atkinson et al. 1997). MCM does not include the 311 formation of HNO<sub>3</sub> from the HO<sub>2</sub>+NO reaction (Butkovskaya et al., 2009, 2007) which is thought to occur 312 313 through the formation of the HO<sub>2</sub>.H<sub>2</sub>O complex. A branching ratio for the formation of HNO<sub>3</sub> was included in the MCM to perform sensitivity tests. The branching ratio for this reaction ( 314  $k_{HO_2+NO \rightarrow HNO_3}/k_{HO_2+NO \rightarrow OH+NO_2}$ ) was considered to be 0.5% under dry conditions (Butkovskaya et al., 315 2007) and dependent on RH with an amplification factor of 8 at 50% RH (Butkovskaya et al., 2009). 316 317 Sensitivity tests were performed using (1) a water-independent branching ratio of 0.5% (HNO<sub>3</sub>/dry in Figures 2-4), (2) a linear parameterization of the ratio on RH leading to a value of 0.5% under dry 318 conditions and a value 8 time higher at 50% RH (HNO<sub>3</sub>/RH\_dep in Figures 2-4), and (3) a 319 parameterization leading to half the water-dependence reported by Butkovskaya et al. (2009) 320 321 (HNO<sub>3</sub>/adj\_RH\_dep in Figures 2-4). Uncertainties on the later simulations (dotted lines in Figures 2-4) were derived as reported in Wood et al. (2007) assuming an error of 25% on the branching ratio under dry 322 conditions and an additional error of 28% associated to the humidity dependence of this ratio. 323

The model was used to calculate the  $NO_2$  concentration at the exit of each reactor for an initial concentration of  $HO_2$  at the inlet similar to those generated during calibration experiments of the CL. The simulations were run for 5-s of reaction time and the model output of  $NO_2$  at 2-s of reaction time was used to calculate the modeled CL using Eq. 2. These simulations have shown that the amount of  $NO_2$  produced from the radical amplification chemistry levels off at approximately 0.5-s of reaction time, e.g. NO<sub>2</sub>(0.5s)/NO<sub>2</sub>(5-s)>97.7%.

VOC	Concentration in calibrator (molecules cm <sup>-3</sup> )	<b>OH rate constant*</b> (cm <sup>3</sup> molecule <sup>-1</sup> s <sup>-1</sup> )	HO₂ yield <sup>‡</sup> VOC+OH reactions (X) (%)	ΔNO <sub>2</sub> (VOC photolysis)/ΔNO <sub>2</sub> (VOC photolysis+water photolysis) (%)
Isoprene	6.2×10 <sup>12</sup>	1.0×10 <sup>-10</sup>	6	1.7
Limonene	3.7×10 <sup>12</sup>	1.7×10 <sup>-10 ¢</sup>	-	1.9
m-xylene	$2.5 \times 10^{13}$	2.3×10 <sup>-11</sup> **	45	0.5
<b>β</b> -pinene	7.7×10 <sup>12</sup>	7.8×10 <sup>-11</sup>	-	3.3
Methylvinylketone	3.5×10 <sup>13</sup>	2.0×10 <sup>-11</sup>	-	3.2
Acetaldehyde	3.8×10 <sup>13</sup>	1.5×10 <sup>-11</sup>	-	0.1
Pentane	1.6×10 <sup>14</sup>	3.8×10 <sup>−12 ∞</sup>	-	0.3
3-methylpentane	1.2×10 <sup>14</sup>	5.2×10 <sup>-12 ∞</sup>	-	-
Cyclohexane	8.8×10 <sup>13</sup>	7.0×10 <sup>−12 ∞</sup>	-	-
Toluene	1.0×10 <sup>14</sup>	5.9×10 <sup>-12</sup> **	28	0.7
Vinylacetate	4.8×10 <sup>13</sup>	1.3×10 <sup>-11</sup> ***	-	14
1-pentene	2.1×10 <sup>13</sup>	3.1×10 <sup>-11</sup> "	-	5.6
Cyclopentene	1.1×10 <sup>13</sup>	6.7×10 <sup>-11</sup> "	-	6.6

**330** Table 1 Summary table of targeted VOCs.

\*(Atkinson et al., 2006); \*\*(Calvert et al., 2011)CAC \*\*\*(Teruel et al., 2006); (Atkinson, 1986); (Atkinson, 2003); (Braure et al., 2014); \* HO<sub>2</sub> yields from MCM V3.2

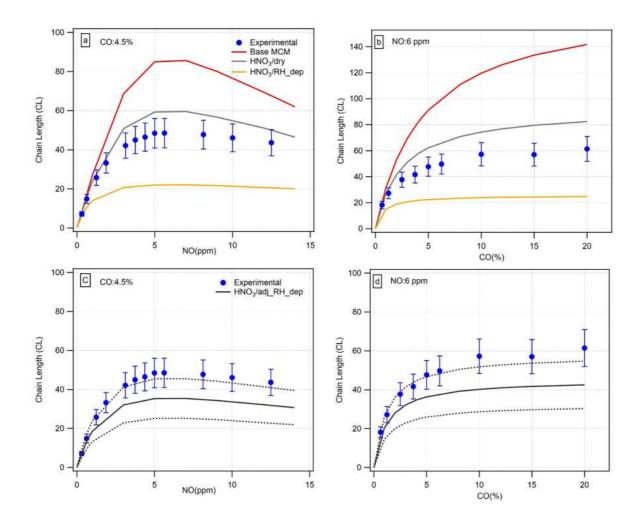
## 333 **3 Results & Discussion**

This section reports a series of characterization and optimization experiments using both PERCA (CO/NO) and ECHAMP (ethane/NO) approaches. The impact of relative humidity on the chain length is investigated for each approach using operating conditions leading to the highest sensitivity. In addition, a comparison of measured CL to values inferred from box modeling is used to test our understanding of each amplification chemistry. The differences observed in CL when calibrating the chemical amplifier with different RO<sub>2</sub> radicals are investigated and contrasted to differences expected from the formation of organic nitrate (R1b) and nitrite (R3) species. Finally, field measurements are presented to illustrate the performances of this instrument.

### 342 3.1 Gas Reagents Optimization

343 A mixture of HO<sub>2</sub> and organic peroxy radicals was generated by adding isoprene in the calibrator as described in the experimental section. The dependence of the measured (markers) and modeled (lines) 344 chain lengths on reagent gases at 50% RH are shown in Figures 2 and 3. For PERCA (Figure 2), CO and 345 NO were varied in the range 0.6-20% and 0.3-12.5 ppm, respectively. Figures 2a shows that both 346 347 experimental and modeled CL increase with the addition of NO up to 5-6 ppm at a constant CO mixing ratio of 4.5%. When NO is increased over 6 ppm, the CL starts decreasing due to a faster increase of the 348 349 OH + NO (R7) reaction rate compared to  $HO_2 + NO (R4a)$ . Figure 2b shows that when CO is varied at a constant NO mixing ratio of 6 ppm, a plateau seems to be reached at CO mixing ratios higher than 10%. 350 351 Other studies have shown similar results with a plateau reached at approximately 7% (Clemitshaw et al., 1997; Kartal et al., 2010; Sadanaga et al., 2004), 10% (Sadanaga et al., 2004) and 8.7% CO (Wood et al., 352 2014) for NO mixing ratios of 3, 3 and 3.3 ppm, respectively. The difference between these operating 353 354 conditions that are optimized to get the largest CL may be due to the use of different materials to build the reactors and different reactor designs, which led to different wall losses of radicals. 355

While the relative NO- and CO-dependencies of the CL are well described by the different models, Figure 356 2b shows that experimental CL values are lower than modelled values by approximately a factor of 2 357 when the base MCM simulation is considered (see section 2.3, red line), with maximum values around 55 358 359 and 120 at 10% CO, respectively. The simulation accounting for the formation of HNO<sub>3</sub> from HO<sub>2</sub>+NO 360 with a water-independent branching ratio of 0.5% (HNO<sub>3</sub>/dry, grey line) is in better agreement with the 361 measured CL values, with a small overestimation of the model by approximately 15-25%. Implementing the water-enhancement of the HNO<sub>3</sub> yield reported in the literature (see section 2.3) leads to a severe 362 363 model underestimation by a factor of approximately 2. This issue is further discussed for the 2 amplification chemistries below. 364

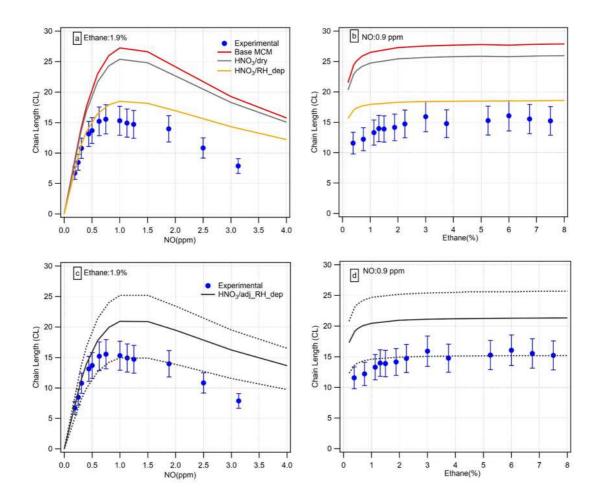


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Figures 2: Chain length dependences on reagent gases for the PERCA (CO/NO) approach. Experimental and
 modeled values are shown as markers and lines, respectively. Experiments were performed at ambient temperature
 (23°C) and 50% RH. Uncertainty on experimental values are 1σ. Panels c,d: dotted lines are uncertainty on modelled
 values. See section 2.3 for the modeling scenarios.

For the ECHAMP approach, the 2 reagents were varied on a smaller range of concentrations, c.a. 0.4-6.0% for ethane and 0.2-3.1 ppm for NO. Figure 3 displays the dependence of both experimental and modeled CL on NO (panel a) and ethane (panel b). The relative dependence on both reagents (curve shape) is similar to that observed for PERCA with optimum CL values for NO and ethane mixing ratios of 0.9 ppm and 2.1%, respectively. These optimum CL values are similar to values reported by Wood et al. 378 (Wood et al., 2017) using 2.3% ethane and 1 ppm NO, respectively. The highest CL is approximately 3 times lower than for the PERCA approach, due to additional losses of radicals through the formation of 379 nitrate (R1b) and nitrite (R3) compounds during the radical amplification stage. Similarly to the PERCA 380 381 approach, CL values simulated by the base MCM model are systematically higher than experimental values by a factor of approximately 2. In addition, as clearly seen in Figure 3c, the shape of the 382 experimental trace is not well reproduced by the model with a peak in CL at higher NO for the model. 383 Implementing the formation of HNO<sub>3</sub> from HO<sub>2</sub>+NO using the water-independent branching ratio has a 384 385 much lower impact on the modelled CL values than for PERCA with a decrease of the CL by only 5-7%. 386 This lower impact is due to a smaller contribution of this additional loss process to the total loss of radicals in the reactor (larger total loss rate compared to PERCA due to organic nitrite and nitrate 387 formation from R3 and R1b, respectively). Using the water-dependent parameterization of the yield brings 388 389 the modeled values in much better agreement with the measurements, which contrasts to that observed for 390 the PERCA approach.

391 Additional simulations were performed for both PERCA and ECHAMP in order to see whether a smaller water-enhancement of the HNO<sub>3</sub> formation yield could reconcile the model/measurement comparison for 392 the two amplification chemistries when uncertainties on modeled and measured CL values are considered. 393 394 As shown in Figures 2 and 3 (panels c and d), simulated CL values considering an enhancement factor of 395 4 at 50% RH (HNO<sub>3</sub>/adj\_RH\_dep) would lead to modeled CL values that are within uncertainties with experimental measurements. Looking at each amplification chemistry separately would lead to the 396 397 conclusion that the measured CL values and dependences on reagent gases could be relatively well reproduced if the water-dependence reported by only one study in the literature is overestimated by a 398 factor 2. However, a systematic error on the parameterization of the HNO<sub>3</sub> yield in the model should lead 399 400 to a systematic deviation from the measurements (either positive or negative) for both comparisons 401 (PERCA and ECHAMP). It is clear that while the additional loss of radicals due to HNO<sub>3</sub> formation is 402 important for both PERCA and ECHAMP, the uncertainty related to its parameterization cannot fully 403 account for the model/experiment disagreement observed in this study (model underestimation for 404 PERCA and overestimation for ECHAMP) and additional work is needed to improve our understanding of 405 these amplification chemistries.



Figures3: Chain length dependences on reagent gases for the ECHAMP (ethane/NO) approach. Experimental and
 modeled values are shown as markers and lines, respectively. Experiments were performed at ambient temperature
 (23°C) and 50% RH. Uncertainty on experimental values are 1σ. Panels c,d: dotted lines are uncertainty on modelled
 values. See section 2.3 for the modeling scenarios.

413

It is interesting to note that a large impact of radical losses in the inlet of the chemical amplifier on this model/measurement comparison was ruled out. As mentioned in the experimental section, first order wall loss rates of HO<sub>2</sub> and CH<sub>3</sub>CH<sub>2</sub>O<sub>2</sub> were measured experimentally at different RH and were found to be 7.1 s<sup>-1</sup> and lower than 1 s<sup>-1</sup>, respectively, at 50% RH. Based on the residence time in the PFA inlet (22 ms) and the first order loss rate observed for HO<sub>2</sub>, a conservative upper limit of 11% is estimated for the amount of peroxy radicals lost on the wall. In addition, increasing the total flow rate inside the inlet by a factor of 2 420 did not significantly impact the measured CL values. It is also interesting to note that setting the wall loss

421 of  $HO_2$  to zero in the base MCM model leads to an increase of the modeled chain length by a factor of 2,

422 indicating that a modification of the reactor geometry to reduce the surface-to-volume ratio and the use of

423 a more hydrophobic material to build the reactors may help increasing the experimental CL and hence the

424 sensitivity of the instrument.

	426
Parameter	Value
PFA reactor	Length: 60 cm, Outer diameter: 0.635 cm
PFA inlet	Length:1.5 cm, Outer diameter: 0.635 cm
Sampling flow rate	Approx. 800 SCCM
Reactor residence time	0.9 s
Inlet residence time	22 ms
ECHAM	P approach (Ethane/NO) 0.9 ppm
Ethane	2.1%
Experimental CL (50% RH)	15
Modeled CL (50% RH)	28
PERC	'A approach (CO/NO)
NO	6 ppm
СО	10%
Experimental CL (50% RH)	55
Modeled CL (50% RH)	105

### 425 Table.2 Operating conditions for the PERCA and ECHAMP approaches

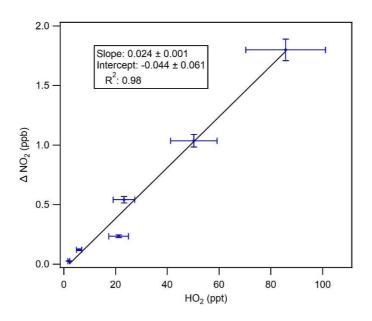
## 427 **3.2** Linearity of the chemical amplifier response ( $\Delta NO_2$ ) with HO<sub>2</sub>

The linearity was investigated using the ECHAMP approach by adding CO in the calibrator to only produce HO<sub>2</sub>. This experiment was performed using a calibrator equipped with a chemical N<sub>2</sub>O filter located between the Hg lamp housing and the turbulent flow tube. The photon flux at 184.9 nm was varied over a factor of 40 by adjusting the concentration of N<sub>2</sub>O inside the chemical filter, allowing to vary the HO<sub>2</sub> mixing ratios in the range 2-85 ppt ( $4.9 \times 10^7$ - $2.1 \times 10^9$  cm<sup>-3</sup>) at a constant RH of 23%. These values encompass ambient concentrations usually observed during field campaigns.

434 Plotting ΔNO<sub>2</sub> versus HO<sub>2</sub> in Figure 4 indicates a good linearity over the tested range of concentrations. A 435 linear regression exhibits a coefficient of determination ( $\mathbb{R}^2$ ) of 0.98 and a negligible intercept (not 436 statistically significant at 1 $\sigma$ ). The CL determined at 23% RH from the slope of this regression line is 24 ± 437 1 (1 $\sigma$ ), which is larger than the CL reported in Figure 2d (approx. 15) under similar conditions of reagents 438 at 50% RH, indicating a strong dependence of the CL with RH.

439





441

Figure 4: Scatter plot of ΔNO<sub>2</sub> with HO<sub>2</sub> for the ECHAMP approach. Experiments performed at a temperature of 23 °C
 and (23±2) % RH. The solid line represents the linear regression line; Error bars are 1σ uncertainty on NO<sub>2</sub>
 measurements and HO<sub>2</sub> generation. Errors on slope and intercept are 1σ.

445

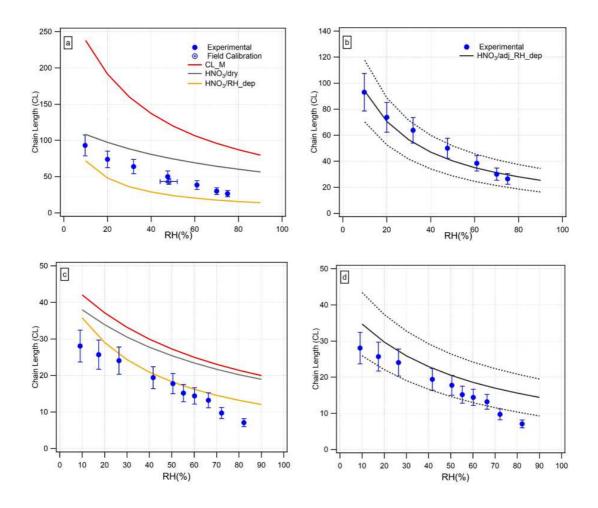
### 446 **3.3 Dependence of the CL on RH**

447 The CL dependence on RH is considered as a major issue for chemical amplification techniques and was investigated for both ECHAMP and PERCA by varying RH in the range 10-85% during calibration 448 449 experiments (isoprene added in the calibrator). As discussed in the introduction section, the CL decreases 450 when humidity increases due to enhanced radical losses on wet surfaces and the potential impact of water 451 dimers and HO<sub>2</sub>.H<sub>2</sub>O complex on the gas-phase chemistry. To quantify this effect, the RH-dependence of both the experimental and modeled CL was investigated. Experimental (markers) and modeled (lines) CL 452 for both amplification approaches are shown in Figure 5 as a function of RH. Operating conditions used 453 454 during these experiments are reported in table 2 and experiments were performed at a temperature of 23±2°C. 455

In this figure, the experimental CL decreases with RH for both approaches, with a drop of about a factor of 3-4 between 10 and 70% RH. At 10% RH the CL for PERCA and ECHAMP are different by approximately a factor of 4. However, a larger RH-dependence of the CL is observed for PERCA, leading to a factor 3 of difference between the 2 approaches at 70% RH. These results are consistent with the work of Wood et al. (Wood et al., 2017) who reported a lower impact of RH for the ECHAMP technique.

The RH-dependence of the PERCA CL reported in this study exhibits a similar behavior to what was observed by Sadanaga et al. (Sadanaga et al., 2004) up to 60% RH, but seems steeper for RH values larger than 60%. This dependence is also consistent with a decrease by a factor of 2 at 40% RH compared to dry conditions reported in Wood et al. (2017). The modelled CL values displayed in panels a and b show that the simulations assuming a water-dependence of the HNO<sub>3</sub> formation yield adjusted at half the reported value can describe the RH-dependence of the measured CL within uncertainty, while the other simulations are in disagreement for both the absolute CL value and the RH-dependence.

For the ECHAMP approach, the RH-dependence is similar to observations reported in Wood et al. (2017), where the CL is lower by a factor of 1.5 at 40 % RH compared to dry conditions. The modelled CL displayed in panels c and d lead to similar conclusions than for the PERCA approach, where simulations based on a 2-fold lower-than-reported water-dependence for the HNO<sub>3</sub> formation yield are in agreement with the experimental observations within uncertainty up to 70% RH. These results are consistent with the simulation presented in the supplementary material of Wood et al. (2007).



474

Figure 5. Dependences of experimental (markers) and modeled (lines) CL on RH for the PERCA (panels a and b) and
ECHAMP (panels c and d) approaches (T = 23±2°C). The empty circle represents calibration experiments performed in
the field (see section Field Testing). Uncertainty on experimental values are 1σ. Panels b,d: dotted lines are uncertainty
on modelled values. See section 2.3 for the modeling scenarios.

## 479 **3.4** Quantification of T<sub>(RO2)</sub> for several RO<sub>2</sub> radicals

As mentioned in the introduction section, PERCA and ECHAMP techniques rely on the conversion of RO<sub>2</sub> radicals into HO<sub>2</sub> in order to initiate the amplification chemistry. The propagation reactions involved in this process will only transmit a certain fraction of each RO<sub>2</sub>, depending on their organic nitrate (R1b) and nitrite (R2-R3) yields. As a consequence, PERCA and ECHAMP techniques will be blind to the fraction of RO<sub>2</sub> radicals that is not propagated to HO<sub>2</sub>. It is important to ensure that PERCA or ECHAMP measurements can be compared to model simulations when each modeled peroxy radical concentration is weighted by a transmission factor derived from known organic nitrate and nitrite yields. Experiments were performed to (i) quantify  $T_{(RO2)}$  (Eq. 6-7) for a large suite of peroxy radicals and to (ii) contrast measured  $T_{(RO2)}$  values to calculations based on organic nitrate and nitrite yields used in atmospheric models for each RO<sub>2</sub>. HO<sub>2</sub> and thirteen organic peroxy radicals were produced from the oxidation of CO and VOCs reported in Table 1 as described in the experimental section. These experiments were performed at 50% RH and ambient temperature (23±2°C) using the ECHAMP approach and operating conditions set to their optimum values as reported in Table 2.

493 Due to the large concentration of  $O_2$  in ambient air, nitrite formation is negligible (< 3%) for all RO 494 radicals generated in these experiments, with the exception of methoxyl (CH<sub>3</sub>O) which exhibits a 5-fold 495 lower rate constant with  $O_2$ . Under normal conditions of temperature and pressure (298 K and 1 atm) and 496 using operating conditions reported in Table 2, the nitrite formation yield is estimated to be 8.7% for 497 CH<sub>3</sub>O. This higher nitrite yield only concerns the oxidation of acetaldehyde due to the formation of an 498 alkoxyl radical which decomposes to form CH<sub>3</sub>O<sub>2</sub>.

499 The measured  $T_{(RO2)}$  values are displayed in Table 3 and Figure 6 as a function of an average yield for nitrate formation. The nitrate yields used in these calculations are from MCM V3.2. This figure clearly 500 indicates that there is a relationship between the two variables, with a decrease of  $T_{(RO2)}$  when the yield 501 502 increases, limonene and  $\beta$ -pinene derived peroxy radicals exhibiting the lowest T<sub>(RO2)</sub> and highest nitrate yields of approximately 77% and 23%, respectively. These results are consistent with observations 503 504 reported by Ashbourn et al. (1998), showing that R1b becomes more significant for large alkyl peroxy 505 radicals, which in turn leads to T<sub>(RO2)</sub> values lower than unity. While T<sub>(RO2)</sub> values are scattered within  $\pm 10\%$  around the 1:1 line, it is surprising that 9 of the 12 experiments are lower than the 1:1 line. These 506 507 lower-than-expected values are likely due to the combination of R1b removal reactions and the injection of large concentrations of VOCs in the calibrator to convert OH into peroxy radicals, which, when 508 509 introduced in the reactor ran under background mode, can lead to a small amplification of the radicals. Tests performed by introducing different concentrations of isoprene in the calibrator showed that the CL is 510 511 underestimated by less than 10% for this compound under experimental conditions reported in this study. This would bring T<sub>(RO2)</sub> for isoprene in very good agreement with the known nitrate yield. A similar 512 513 impact is expected for the other  $T_{(RO2)}$  values measured in this study. In addition, it is interesting to note 514 that lower-than-expected T<sub>(RO2)</sub> values could also be partly due to NO in homogeneities when ambient air 515 is mixed with the reagents, leading to local concentrations of NO larger than 1 ppm, which in turn can lead 516 to a higher nitrite yield for each RO radical. Indeed, experimental determinations of T<sub>(RO2)</sub> provide values

which are integrated over the reaction time when the reagents are getting mixed with ambient air (NO getting diluted down to 1 ppm). Overall, these results indicate that known nitrate and nitrite yields can be used to scale modelled concentrations of peroxy radicals for comparison with ECHAMP measurements with an uncertainty better than 10%.

It is important to note that these scaling factors will be different for the PERCA approach since a higher NO mixing ratio is used to generate the amplification chemistry. The RO<sub>2</sub>-to-HO<sub>2</sub> conversion step will be less efficient due to a higher formation rate of organic nitrites. The nitrite yields are expected to increase from 2-3% to 12-18% for most of the RO radicals used in this study, with a higher increase for CH<sub>3</sub>O whose nitrite yield will be close to 36%.

VOC	RONO <sub>2</sub> yield	T <sub>(RO2)</sub>
	MCM V3.2 (%)	
Isoprene	10.7	0.84
Limonene	22.8	0.65
m-xylene	7.3	0.91
β-pinene	23.3	0.66
Acetaldehyde	0.1	0.95
Pentane	14.1	0.92
3-methylpentane	12.4	0.74
Cyclohexane	15.0	0.87
Toluene	8.7	0.72
1-pentene	5.9	0.88
Cyclopentene	4.5*	0.84

Table 3: RO<sub>2</sub> to HO<sub>2</sub> transmission – T<sub>(RO2</sub>

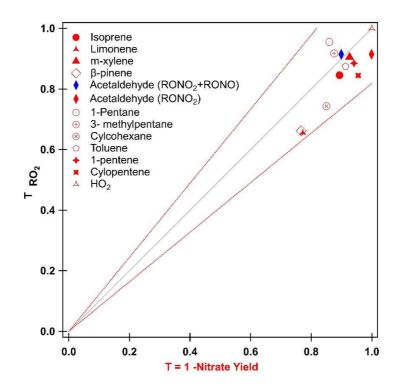
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\* Reference: (Perring et al., 2013)



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Figure 6: Dependence of  $T_{(RO2)}$  (ECHAMP approach) on average organic nitrate yields. The blue symbol for acetaldehyde accounts for nitrate & nitrite formation (see text). Plain grey line is 1:1 and dashed red lines are 1.10:1 and 0.90:1.

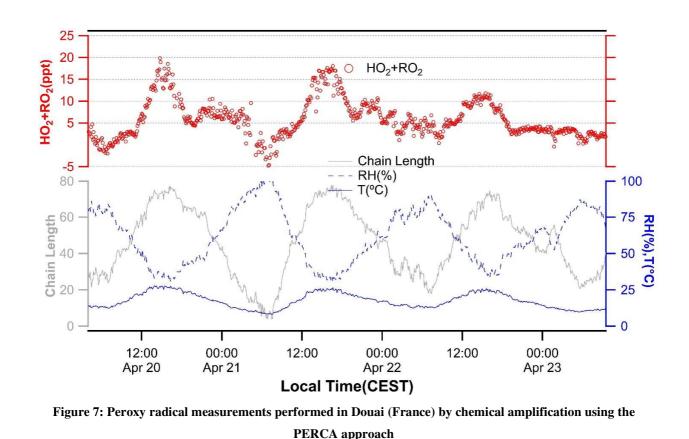
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#### 535 **3.5** Field deployment of the chemical amplifier

The chemical amplifier was first tested in an urban area, close to our laboratory, in the city of Douai, 536 France. These measurements were performed with the PERCA approach, i.e. using CO and NO as 537 538 reagents (operating conditions from Table 2). The measurements were conducted over 3 consecutive days in April 2018 using the sequence shown in Figure S2 (supplementary material), which in turn led to a time 539 resolution of 6-min. The overall meteorological conditions allowed to run the instrument under optimal 540 541 conditions since these 3 days were characterized by clear skies, ambient temperatures between 8 and 25°C, and RH values ranging from 35 to 90%. The only exception was a light rain in the early morning of 542 21 April, during which RH reached 100%. The peroxy radical measurements are shown in Figure 7 543 together with the CL calculated at the measured RH values. The latter varied from approximately 10 up to 544 75.  $RO_2$ +HO<sub>2</sub> mixing ratios ranged from the detection limit up to 20 ppt around 2 pm local time (UTC+2), 545 with sustained mixing ratios of 5-10 ppt during the late afternoon and nighttime. The magnitude of the 546

547 measured RO<sub>2</sub>+HO<sub>2</sub> mixing ratios and their diurnal variations are similar to trends reported in studies 548 performed at measurements sites under urban influences (Tan et al., 2018; Whalley et al., 2018; Wood and 549 Charest, 2014).

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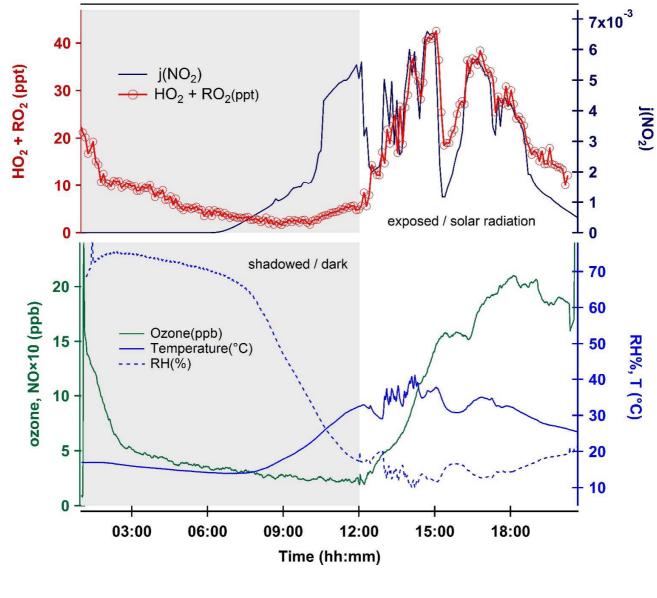
556 This instrument was then deployed in the Landes forest (France) from 2 to 26 July 2018 as part of the CERVOLAND (Characterization of emissions and reactivity of volatile organic compounds in the Landes 557 forest) field campaign. This forest is mainly composed of pine trees (Pinus Pinaster), which are strong 558 emitters of monoterpenes such as  $\alpha$ - and  $\beta$ -pinene. The instrument was coupled to a newly designed 9-m<sup>3</sup> 559 Teflon photo-reactor to investigate in-situ the oxidation of biogenic VOCs (BVOCs). Several experiments 560 were conducted by flushing the photo-reactor with ambient air at different times of the day (nighttime, 561 daytime) to trap air masses characterized by different chemical compositions. Oxidation experiments were 562 then performed either in the dark or under solar irradiation. Here, we present peroxy radical measurements 563 564 from one experiment when the photo-reactor was flushed with ambient air during nighttime for 90

minutes. The air mass trapped in the photo-reactor was loaded with large mixing ratios of monoterpenes (14 ppb) and sesquiterpenes (50 ppt). The ambient ozone mixing ratio was low (12-14 ppb) during the flushing period. The air mass was kept in the dark until 12 pm, BVOCs slowly reacting  $O_3$  away (2-4 ppb around 11 am), and was then exposed to solar radiations until 8 pm by uncovering the photo-reactor. Figure 8 displays the peroxy radical measurements during these two periods.

570 Peroxy radicals slowly decreased together with  $O_3$  (O3 42E, Environmement SA) from 1 to 9 am. The 571 mixing ratios dropped from 10 to 2 ppt, and 14 to 2 ppb for peroxy radicals and O<sub>3</sub>, respectively. O<sub>3</sub>-BVOC reactions were likely the main source of radicals during this time period. Peroxy radicals slowly 572 increased from 9 am to 12 am together with the temperature, indicating that thermo-labile precursors of 573 radicals may have been present in the photo-reactor. Once the photo-reactor was uncovered, peroxy 574 radicals rapidly increased from 6 ppt up to 45 ppt around 3 pm. The latter is similar to previously reported 575 peroxy radicals mixing ratios in forested areas (Stone et al., 2012; Vaughan et al., 2012; Wolfe et al., 576 577 2014). During this time period peroxy radical measurements are well correlated with j(NO<sub>2</sub>) measured 578 above the photo-reactor (fast CCD spectroradiometer from METCON meteorologieconsult gmbh). Interestingly, O<sub>3</sub> increased at a rate of 3-4 ppb h<sup>-1</sup> from 12 to 3 pm, reaching a plateau at approximately 20 579 ppb. These results will be presented in detail in a forthcoming publication. 580

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Figure 8: Peroxy radical measurements performed in a mobile photo-reactor during the CERVOLAND 2018 field
 campaign

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# 590 **3.6 Estimation of the detection limit**

Lower bounds for the detection limit can be derived from the measurement noise of the 2 CAPS monitors and the RH-dependent CL, assuming that no other sources of noise are present. Measurements of NO<sub>2</sub> by the two CAPS analyzers used in this work exhibit precisions of 10 and 16 ppt (1 $\sigma$ ) when zero air is sampled, which translate into detection limits (3 $\sigma$ ) of 30 and 48 ppt for 70-seconds of integration (70 occurrences of 1-s NO<sub>2</sub> measurements averaged within each measurement step of 90-s, see supplementary material). This leads to detection limits of 1.3, 2.0, and 3.3 ppt for the ECHAMP approach at RH values of 10, 50, and 80%, respectively, for a total measurement time of 6 min (see measurement sequence in the experimental section). For the PERCA approach, the same procedure leads to detection limits of 0.3, 0.6, and 1.4 ppt at 10, 50, and 80% RH, respectively.

600 However, detection limits for ambient air measurements are expected to be significantly higher due additional sources of noise such as small fluctuations in reagent flow rates, switches of solenoid valves, 601 and changes in ambient concentrations of NO<sub>2</sub> and O<sub>3</sub>, which will degrade the precision on  $\Delta$ NO<sub>2</sub>. In 602 order to better estimate limits of detection, both reactors were run in background mode for long time 603 604 periods of 3-10 hours during the CERVOLAND field campaign. An experiment of this type is shown in the supplementary material (S4), where  $\Delta NO_2/CL$  is plotted as a function of time for RH values ranging 605 from 20-30%. A statistical analysis of this series indicate that the mean value of  $0.03 \pm 0.04$  ppt is not 606 607 statistically significant. The standard deviation of 0.3 ppt translates into a detection limit  $(3\sigma)$  of 0.9 ppt. Assuming that  $\Delta NO_2$  is not dependent on RH, since peroxy radicals are not amplified, this would translate 608 into detection limits of 0.7, 1.3, and 3.0 ppt at RH values of 10, 50, and 80 %, respectively. These values 609 610 are approximately 2 times higher than values calculated from the noise of the CAPS monitors. While these values may still underestimate the detection limit in ambient air, since the O<sub>3</sub> and NO<sub>2</sub> variability is 611 reduced in the photo-reactor, it is clear that detection limits on the order of a few ppt are achievable using 612 613 the PERCA approach. The ECHAMP approach, which has yet to be tested in the field on this instrument, should exhibit detection limits that are approximately 3 times higher. 614

### 615 **4** Conclusion

This publication presents the development of a chemical amplifier for ground-based measurements of peroxy radicals in the troposphere. This instrument was used in the laboratory to compare two different approaches regarding the radical amplification chemistry, including the use of CO/NO (PERCA) or ethane/NO (ECHAMP) as reagent gases. Tests performed to identify optimum conditions leading to the highest sensitivity, i.e. the largest chain length, indicate that 10% CO and 6 ppm NO should be used for the PERCA approach, while 2.1% ethane and 0.9 ppm NO are better for ECHAMP. These optimum conditions lead to chain lengths values of approximately 55 and 15 at 50% RH for the PERCA and ECHAMP approaches, respectively. The RH-dependence of the CL was also investigated for both approaches. It was found that the CL decreases with RH, in agreement with previously published studies. The magnitude of the decrease is slightly lower for the ECHAMP technique, with a decrease by a factor 3 and 2.7 for the PERCA and ECHAMP approaches, respectively, when RH is increased from 10 to 70% at  $23\pm2^{\circ}$ C.

628 The amplification chemistry was modelled using MCM v3.2 for comparison to experimental observations. The model/measurement comparison indicates that modeled CL values overestimate experimental 629 630 observations by a factor of approximately 2. Additional simulations conducted to assess whether the formation of HNO<sub>3</sub> from HO<sub>2</sub>+ NO could impact the CL of chemical amplifiers highlighted that using the 631 632 yield reported under dry conditions, and a 2-fold lower-than-reported water-dependence for this yield, provides a reasonable description of the CL-dependence on reagent gases and RH for both PERCA and 633 ECHAMP. However, the model overestimation of the ECHAMP CL and underestimation of the PERCA 634 635 CL seems to indicate that our understanding of the amplification chemistry is still incomplete and deserves 636 addition scrutiny.

The sensitivity of the chemical amplifier to a large range of RO<sub>2</sub> radicals was quantified to determine whether PERCA and ECHAMP measurements can be compared to model simulations when the concentration of each modelled peroxy radical is weighted by a transmission factor derived from known organic nitrate and nitrite yields. This work showed that transmission factors can be calculated from known kinetic parameters and ambient conditions for the pool of organic peroxy radicals tested in this study.

Finally, ambient testing of this chemical amplifier using the PERCA approach showed that this instrument is capable of measuring ambient concentrations of  $HO_2+RO_2$  at levels higher than 1-4 ppt under RH conditions up to 90% once the RH-dependence of the CL has been correctly characterized. Comparisons of the PERCA and ECHAMP techniques to other techniques capable of measuring peroxy radicals would be useful to provide more insight into the accuracy of chemical amplifiers.

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