

# Reactive uptake of NO2 on volcanic particles: A possible source of HONO in the atmosphere

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- 1 Reactive uptake of NO<sub>2</sub> on volcanic particles: a possible source of HONO in the
- 2 atmosphere

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4	Manolis N Romanias <sup>1,*</sup> , Yangang Ren <sup>2</sup> , Benoit Grosselin <sup>2</sup> , Véronique Daële <sup>2</sup> , Abdelwahid						
5	Mellouki <sup>2</sup> , Pavla Dagsson-Waldhauserova <sup>3,4</sup> , Frederic Thevenet <sup>1</sup>						
6	1. IMT Lille Douai, Univ. Lille, SAGE, F-59000 Lille, France						
7	2. CNRS-ICARE/OSUC, 45071, Orléans, France						
8	3. Agricultural University of Iceland, Keldnaholt, Reykjavik 112, Iceland						
9 10	4. Faculty of Environmental Sciences, Czech University of Life Sciences, Prague 165 21, Czech Republic						
11							
12							
13							
14	* Corresponding author. E-mail: <u>emmanouil.romanias@imt-lille-douai.fr</u> (Manolis N						

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Romanias)

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## Abstract

The heterogeneous degradation of nitrogen dioxide (NO<sub>2</sub>) on five samples of natural Icelandic 18 19 volcanic particles has been investigated. Laboratory experiments were carried out under simulated 20 atmospheric conditions using a coated wall flow tube (CWFT). The CWFT reactor was coupled to a 21 blue light nitrogen oxides analyzer (NO<sub>x</sub> analyzer), and a long path absorption photometer (LOPAP) to 22 monitor in real time the concentrations of NO<sub>2</sub>, NO and HONO respectively. Under dark and ambient relative humidity conditions, the steady state uptake coefficients of NO<sub>2</sub> varied significantly between 23 24 the volcanic samples probably due to differences in magma composition and morphological variation 25 related with the density of surface OH groups. The irradiation of the surface with simulated sunlight enhanced the uptake coefficients by a factor of three indicating that photo-induced processes on the 26 27 surface of the dust occur. Furthermore, the product yields of NO and HONO were determined under 28 both dark and simulated sunlight conditions. The relative humidity was found to influence the distribution of gaseous products, promoting the formation of gaseous HONO. A detailed reaction 29 mechanism is proposed that supports our experimental observations. Regarding the atmospheric 30 implications, our results suggest that the NO<sub>2</sub> degradation on volcanic particles and the corresponding 31 32 formation of HONO is expected to be significant during volcanic dust storms or after a volcanic 33 eruption.

- 34 Keywords:
- 35 volcanic dust
- 36 HONO
- 37 Uptake coefficients
- 38 Product yields
- 39 Simulated atmospheric conditions

## 40 **1. Introduction**

41 Nitrogen dioxide (NO<sub>2</sub>) is an important atmospheric pollutant that contributes to the formation of photochemical smog, (von Schneidemesser et al., 2015). NO<sub>2</sub> acts as an intermediate species in several 42 43 gas phase photochemical cycles occurring in the troposphere affecting the radical budgets for 44 hydroperoxyl radicals (HO<sub>x</sub>) and Peroxy radicals (RO<sub>x</sub>), as well as that of ozone (O<sub>3</sub>). In addition, the chemistry of  $NO_2$  is involved in the formation of organic nitrates (Nault et al., 2016; Perring et al., 45 2013; von Schneidemesser et al., 2015). Besides gas phase chemistry, the heterogeneous reactivity of 46 47 NO<sub>2</sub> with atmospheric particles has been shown to play an important role in atmospheric processes. NO<sub>2</sub> reacts with aerosols forming nitrous acid (HONO), a key atmospheric species, precursor of the 48 major atmospheric oxidant: OH radicals (Mellouki et al., 2015). So far, atmospheric simulation 49 models fail to reproduce the unexpectedly high daytime concentrations of HONO reported in field 50 51 campaigns, suggesting the existence of new, yet unknown, sources of HONO (Kleffmann, 2007; 52 Romer et al., 2018). Therefore, it has been suggested that the heterogeneous conversion of  $NO_2$  to 53 HONO could possibly explain the discrepancies between atmospheric models and field observations 54 (Romer et al., 2018).

The heterogeneous reaction of  $NO_2$  on various type of atmospheric particles, e.g. salts, soot, mineral 55 56 dust and proxies, has been already reported in literature (Crowley et al., 2010; George et al., 2015; Tang et al., 2017). However, the interaction of NO<sub>2</sub> with volcanic particles under simulated 57 58 atmospheric conditions has not been studied. Actually, the role of volcanic aerosols on tropospheric 59 chemistry has not been thoroughly evaluated yet. Volcanic aerosols may act as platforms for the scavenging or conversion of key atmospheric species (e.g. NO<sub>2</sub>,  $O_3$ , SO<sub>2</sub> etc.), and thus, impact the 60 tropospheric oxidation capacity regionally, or in a larger scale such as during massive volcanic 61 62 eruptions or during extreme dust storms of volcanic dust as reported for instance in Iceland (Arnalds et 63 al., 2016; Arnalds et al., 2013; Moroni et al., 2018; Ovadnevaite et al., 2009).

64 Considering that nearly 10% of the world population is living in the vicinity of active volcanos (Auker 65 et al., 2013), it is expected that pollutants emitted from anthropogenic sources, such as nitrogen oxide 66 species ( $NO_x$ ),  $O_3$  etc. can interact or react with naturally emitted volcanic gases or volcanic aerosols. 67 A prime example is Mexico City, known as one of the most polluted city in the world, situated 70 km 68 away from the frequently erupting Popocatépetl volcano. Besides, distal urban environments can also 69 be impacted by volcanic pollution. Resuspension of volcanic ash, defined as volcanic dust, or v-dust, 70 for the following of the paper, can prolong the effects of volcanic eruptions several months after the 71 eruption is finished (Arnalds et al., 2013; Butwin et al., 2019; Wilson et al., 2011). The recent 72 eruptions of Icelandic volcanoes Eyjafjallajökull (2010) and Grimsvötn (2011) resulted in the dispersal of massive loads of ashes and aerosols over a large part of Europe and Northern Atlantic 73 74 (Gudmundsson et al., 2012; Ilyinskaya et al., 2017). Holuhraun (2014-2015) was an effusive eruption 75 occurring in the center of volcanic dust hot spot, but volcanic dust was frequently suspended during the eruption (Pfeffer et al., 2018; Vignelles et al., 2016). Therefore, volcanic aerosols are regularly 76 77 emitted in the atmosphere and recent studies have shown that their lifetime is significantly longer than 78 it has been estimated (Boichu et al., 2016).

Consequently, the objective of the current study is to determine experimentally how efficiently  $NO_2$  is 79 80 scavenged on volcanic particles and to evaluate whether this interaction can be a significant source of 81 HONO in the atmosphere - on local or regional scale - during severe volcanic dust storms or volcanic 82 eruptions. To that end, laboratory experiments have been carried out under simulated atmospheric 83 conditions (temperature, relative humidity, NO<sub>2</sub> concentrations and simulated sunlight radiation) 84 employing a flow tube coupled with state of the art instrumentation for the simultaneous detection of 85  $NO_2$ , NO and HONO. The reactive uptake coefficients of  $NO_2$  are measured and the yields of the 86 products formed, and the mechanism of the reaction are addressed.

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# 2. Materials and methods

**2.1 Materials** 89

90 2.1.1 Origin of the volcanic samples

In the current paper, Iceland was chosen as a case study. Iceland is a very active geological site of 91 92 Earth with more than 30 active volcanoes and large eruptions occurring every 3 to 5 years. Hence it is 93 considered as the largest volcanic desert on Earth (Arnalds et al., 2016) with estimated dust deposition

of 31 - 40 million tons per year (Arnalds et al., 2014), that could represents up to 7% of total global 94 95 dust emissions and up to 21% of the Saharan dust emissions (Arnalds et al., 2014; Engelstaedter et al., 96 2006). Remote sensing studies have evidenced that volcanic plumes formed after a volcanic eruption 97 in Iceland can travel thousands of kilometers impacting urban populated areas of northern Europe (Boichu et al., 2016). Moreover, Iceland is one of the most active aeolian areas on the planet, where 98 99 unstable sandy surfaces are widespread and are subjected to frequent high-velocity winds, resulting in 100 numerous wind erosion events and massive volcanic dust storms (Arnalds et al., 2016). In Iceland, 101 around 100 dust events are reported each year (Arnalds et al., 2014); up to 300,000 tons of dust can be 102 emitted in a single storm. Particulate matter 10 micrometers or less in diameter (i.e.  $PM_{10}$ ) mass concentrations during dust storms can exceed 7000  $\mu$ g/m<sup>3</sup> with a 1000  $\mu$ g/m<sup>3</sup> 24-h average 103 104 concentration (Dagsson-Waldhauserova et al., 2015; Thorsteinsson et al., 2011) striking populated areas such as Reykjavik (Dagsson-Waldhauserova et al., 2016). Satellite images confirm that during a 105 dust storm, v-dust particles can travel distances longer than 1000 km impacting the air quality of 106 107 central Greenland and northern Europe (Arnalds et al., 2016). In a recent study (Dorđević et al., 2019) it has been shown that volcanic dust from Iceland can be transported to even central Balkans, and 108 109 authors pointed at the fact that Icelandic dust can contribute to impaired air quality in mainland 110 Europe. Icelandic dust is also transported towards Arctic (Groot Zwaaftink et al., 2017; Moroni et al., 2018). 111

The volcanic samples investigated in the present work were collected from different desert regions of Iceland, so called dust hot spots Hagavatn, Mýrdalssandur, Maelifellssandur and Dyngjusandur, while the fifth sample of Eyjafjallajökull volcanic ash, named typical ash in the following of the article, was collected several hours after the eruption of the Eyjafjallajökull in 2010 (Gislason et al., 2011). Details about the characterization of the physical and chemical properties of the samples can be found elsewhere (Baratoux et al., 2011; Dagsson-Waldhauserova et al., 2014b; Moroni et al., 2018; Urupina et al., 2019) and for clarity are briefly discussed in Appendix A. Supplementary data.

#### 119 2.1.2 Gases

The laboratory experiments were carried out using dry zero air as bath gas (moisture level ca. 2 ppmV) 120 121 inside a photochemical flow tube. For experiments requiring humid air, an extra flow of zero air 122 passing through a bubbler containing ultrapure water (milli-Q, resistivity 18.2 M $\Omega$ ·cm) was mixed 123 with the dry air flow, in proportions required to reach the targeted relative humidity (RH). The exact 124 RH in the gas flow is determined and monitored using a temperature and RH probe (KIMO HQ 210, 125 KIMO instruments, United Kingdom). Certified gas cylinders were used as NO<sub>2</sub> source. For low concentration experiments, aiming to determine uptake coefficients in the flow tube, a NO<sub>2</sub> cylinder of 126 200 ppbV in air (Air Liquide, France) was used. As far as concerns the experiments carried out to 127 determine the irradiance intensity inside the flow tube reactor, rather high concentrations of  $NO_2$  were 128 used supplied from a NO<sub>2</sub> cylinder of 50 ppmV in N<sub>2</sub> (Praxair, Belgium). 129

#### 130 2.1.3 Surface preparation.

A few mg of the volcanic sample was placed inside a Pyrex tube (20 cm length, 0.98 cm internal 131 132 diameter) and, depending on the volcanic sample mass, diluted in 0.5 to 2 mL of water. Then the tube 133 was hand-shaken, so that the water/dust suspension homogeneously spreads on the walls of the tube 134 without any trace of the suspension to be removed after the deposition. Thereafter, the tube was heated 135 at 120°C with a heat-gun for several minutes and further dried overnight in an oven at 100°C. It should be noted that potential modifications of the volcanic dust upon its dissolution in water in the process of 136 137 slurry preparation and subsequent drying process are expected to be of limited extend (Urupina et al., 2019), and not to influence the uptake coefficients measured in the current study. 138

#### 139 **2.2 Experimental set up**

The uptake experiments of  $NO_2$  on the volcanic samples were performed using a horizontal double wall Pyrex atmospheric pressure coated wall flow tube reactor (CWFT) (**Fig. 1**). Further details of the experimental set up have been described in previous publications (Lasne et al., 2018; Urupina et al., 2019). The Pyrex tube with the inner coated surface was introduced into the main reactor along its axis. Two Viton O-rings were placed in both ends of the Pyrex tube to set its position.  $NO_2$  was introduced in the reactor by means of a movable injector with internal diameter 0.3 cm. The injector can be moved along the main axis of the coated Pyrex tube, allowing the variation of the v-dust

147	sample length exposed to $NO_2$ and consequently the contact time ( <i>t</i> ). The total flow rate was regulated
148	between 200 and 400 mL/min, ensuring laminar flow conditions (Reynold number, Re, < 140).

149 The outlet gas flow was then mixed with a dry zero air flow (to compensate the sampling flow of the 150 analytical instruments) and was either sent to a  $NO_x$  analyzer (Teledyne T 200UP, Teledyne API, United States) equipped with a blue light converter for the detection of  $NO_2$  (detection limit 50 pptV), 151 152 and nitrogen oxide (NO that appeared as reaction product), or to a long path absorption photometer 153 (LOPAP, QUMA, Germany) to monitor HONO (reaction product) with a limit of detection of 1-2 154 pptV. Teledyne T 200UP NO<sub>x</sub> analyzer uses photolysis in its conversion process; the gas flow passes through a conversion chamber and it is exposed to a blue light from two high powered ultraviolet light 155 emitting diodes (LEDs) characterized by narrow emission bands centered at 395 nm as reported in Fig. 156 157 2 (bell curve area). The emission band of the LED includes  $NO_2$  absorption band with negligible 158 interferences expected from other gases such as HONO or  $NO_3$  radicals. Consequently, unlike  $NO_x$ analyzers relying molybdenum converters, the selective detection of  $NO_2$  is achieved. Based on the 159 LED output intensity and wavelength range as well as the absorption cross section of HONO given in 160 Fig. 2, we estimated that the HONO interference to  $NO_2$  concentration to be lower than 3%. 161

162 To further evaluate possible interferences to  $NO_2$  concentration measured along the uptake 163 experiments, in a dedicated set of experiments a cavity attenuated phase shift (CAPS) NO<sub>2</sub> analyzer 164 (AS32M, ENVEA S.A, France) was additionally coupled with the reactor. The NO<sub>2</sub> CAPS (detection 165 limit 100 pptV) provides a direct absorption measurement of nitrogen dioxide at 450 nm and does not 166 require any conversion of  $NO_2$  to another species. Thus, no interference from other N-containing 167 species are possible. Therefore, comparing the concentrations of  $NO_2$  monitored with the  $NO_x$  analyzer 168 and the CAPS we concluded that the possible interference of HONO is around 2% and thus included 169 in our experimental uncertainties.

#### 170 **2.3 Light source characterization.**

The CWFT reactor was surrounded by 3 Ultraviolet A lamps (UVA, Philips lighting 18 Watt; 315–400
nm with maximum emission at 352 nm) (Romanias et al., 2017). In order to characterize the irradiance
intensity in the flow tube, the NO<sub>2</sub> photolysis frequency was measured as a function of the number of

174	lamps turned on. These experiments were carried out with rather high $NO_2$ concentrations of ca. $10^{15}$
175	molecule/cm <sup>3</sup> , i.e. 50 ppmV, using dry $N_2$ as bath gas. NO <sub>2</sub> concentration was monitored with a
176	chemical ionization mass spectrometer (SIFT-MS, Syft Technology, new Zealand) (Romanías et al.,
177	2016; Romanias et al., 2016; Zeineddine et al., 2017). The values of NO <sub>2</sub> photolysis frequency, $j_{NO2}$
178	(sec <sup>-1</sup> ), determined from the exponential decays of $NO_2$ for different numbers of lamps switched on
179	were in the range of 4 to 15 $10^{-3}$ sec <sup>-1</sup> (Appendix A <b>Fig. S2</b> ). Thus it overlaps with the values of J <sub>NO2</sub>
180	measured in the atmosphere from cloudy to clear sky conditions (Barnard et al., 2004; Bohn et al.,
181	2005; Topaloglou et al., 2005). Further information regarding the light source characterization
182	experiments are reported in Appendix A. Supplementary data.

183 It has to be noted that under the experimental conditions where the uptake experiments were carried 184 out, the gas phase loss of NO<sub>2</sub> due to its photodissociation is negligible. Indeed, for only 1 UV lamp, i.e.  $j_{NO2} = 4.5 \times 10^{-3} \text{ sec}^{-1}$ , denoted as simulated sunlight conditions in the following of the article, and 185 186 with a residence time of 2 sec, the gas phase loss of  $NO_2$  due to its photolysis was lower than 0.8%. In 187 case more than one UV lamp was used, the flow in the reactor was increased to maintain gas phase consumption due to photolysis below 1.5% to avoid any impact on the measured uptake coefficient 188 values. To conclude, our experimental measurements of the uptake coefficients and the product yields 189 190 discussed below are not expected to be biased from interferences or experimental artifacts.

#### **191 2.4 Uptake of NO<sub>2</sub> on volcanic samples**

#### **192 2.4.1 Experimental procedure**

A typical uptake profile of NO<sub>2</sub> on Hagavatn sample under 30% of RH and 293 K is presented in Fig. 193 194 **3.** It was selected to present breakthrough profiles with all instruments coupled with the flow tube to 195 analyze the outlet flow. In these experiments, the gas phase was continuously monitored with the  $NO_x$ 196 analyzer, while either the NO<sub>2</sub> CAPS or the LOPAP were coupled to the flow tube (Fig. 3). First, the 197 injector is placed at the downstream position of the reactor, ensuring no contact of the gas flow with 198 the v-dust coating. At this stage the background concentration levels for all instruments is recorded. 199 Afterwards,  $NO_2$  is introduced in the reactor and a steady state concentration is achieved within a few 200 minutes. The NO levels are close to zero while a very low formation of HONO was observed,

201	probably due to hydrolysis of $NO_2$ on the wet Teflon lines (Finlayson-Pitts et al., 2003). This
202	contribution was taken into account for further calculations of the product yields. Thereafter, the
203	LOPAP was bypassed and the gas stream was driven through the NO <sub>2</sub> CAPS; as displayed in Fig. 3,
204	the $NO_2$ concentrations measured with the analyzer and the CAPS are in excellent agreement.
205	Subsequently, the injector is pulled out ( $t = 0$ in <b>Fig. 3</b> ) and the whole dust coating is exposed to the
206	gas flow. A fast consumption of $NO_2$ was observed due to $NO_2$ uptake on the volcanic sample, and a
207	gradual recovery to higher outlet concentrations was noticed. Simultaneously, NO was formed as a
208	product of the heterogeneous reaction. After achieving a steady state consumption of $NO_2$ and
209	formation of NO, the CAPS was bypassed to monitor the concomitant concentration of HONO with
210	the LOPAP for ca. 25 min. Then, the flow was sent back to the CAPS. The next step in the
211	experimental protocol was the irradiation of the surface with simulated sunlight; the initiated photo-
212	induced processes on the surface of the volcanic particles lead to an extra consumption of $NO_2$ and an
213	additional formation of NO. During the sample irradiation, the gas flow was sent to the LOPAP to
214	determine the HONO concentrations formed and then switched back to the CAPS. In the last step, the
215	injector is pushed back in, to control the $NO_2$ level. The steady state consumption of $NO_2$ and the
216	formation products such as NO and HONO under dark and simulated sunlight conditions, were used to
217	determine the corresponding uptake coefficients ( $\gamma_{ss}$ ) and product yields.

#### 218 2.4.2 Uptake coefficients and data analysis

219 The uptake coefficient of NO<sub>2</sub> on the v-dust surface can be derived from the following expression:

220 
$$\gamma_{\text{geom}} = \frac{2 \times k_{\text{kin}} \times r_{\text{tube}}}{c_{\text{NO2}}} \tag{1}$$

where  $\gamma_{geom}$  is the geometric uptake coefficient,  $k_{kin}$  is the first order reaction rate loss of gas-phase NO<sub>2</sub> (s<sup>-1</sup>) in the kinetic regime,  $r_{tube}$  is the radius of the tube, and  $C_{NO2}$  is the average molecular velocity (cm/sec). Both the radius of the tube (0.49 cm) and the average molecular speed at 293 K are known parameters, therefore only  $k_{kin}$  needs to be determined. Further information about the calculations of uptake coefficients and diffusion corrections can be retrieved elsewhere (Romanias et al., 2013, 2012; Tang et al., 2014) and are also given in Appendix A. Supplementary data. The determination of the uptake coefficient using the geometric surface area ( $S_{geom}$ ) can be considered as an upper limit (Crowley et al., 2010). An alternative way for the calculation of the uptake coefficient is using the specific surface area ( $A_s$ ) of the solid, considering that all surface sites are accessible to the gas phase molecules. For the determination of the Specific surface area (SSA), the Brunauer-Emmett-Teller (BET) isotherm approach is applied and the corresponding uptake coefficient ( $\gamma_{BET}$ ) is given by the following expression:

233 
$$\gamma_{\text{BET}} = \frac{\gamma_{\text{geom}} \times S_{\text{geom}}}{S_{\text{BET}}}$$
(2)

where  $S_{\text{BET}}$  (m<sup>2</sup>/g) is the product of the specific surface area ( $A_s$ ) of the solid with the deposited mass (g).

## 236 **3. Results and discussion**

#### 237 3.1 Determination of NO<sub>2</sub> uptake coefficients on v-dust samples

This section includes the results of the kinetic measurements performed under simulated atmospheric 238 239 conditions aiming to determine the uptake coefficients of  $NO_2$  on the volcanic samples. Initially, the 240 impact of experimental conditions on the uptake is discussed using dust from Hagavatn collection spot 241 as a model volcanic sample; this choice is based on two criteria. First, Hagavath dust hot spot is in 242 close vicinity to Reykjavik - the most densely populated area in Iceland - and secondly, the air quality 243 in Reykjavik is very frequently impacted from massive loads of v-dust particles originating from 244 Hagavatndust hot spot. Therefore, it was considered as the most representative sample to study with 245 the ultimate objective to evaluate the corresponding impacts of  $NO_2$  aged volcanic dust to the 246 chemistry of the local/regional atmosphere. Finally, at the end of this section a thorough comparison 247 between the uptake coefficients determined on the five volcanic samples under the same experimental 248 conditions is performed.

#### **3.1.1 Dependence of NO<sub>2</sub> uptake coefficient on the sample mass**

In this series of experiments, the steady state uptake coefficient was investigated as a function of Hagavatn v-dust mass exposed to the  $NO_2$  flow. The objective was to determine the v-dust surface area involved in the interaction with  $NO_2$  molecules. The experiments were performed under both dark

and simulated sunlight irradiation conditions ( $j_{NO2} = 4.5 \times 10^{-3} \text{ sec}^{-1}$ ) at temperature (T) = 293 K, 100 253 254 ppbV of  $NO_2$  and dry conditions. The obtained results are presented in Appendix A Fig. S3. A linear increase of the geometric uptake coefficient ( $\gamma_{ss,geom}$ ) was observed as a function of the mass of 255 256 Hagavatn coated on the tube. This trend points out that the entire surface area of the solid sample is accessible to NO<sub>2</sub>, and consequently, the BET specific surface area should be used for the 257 determination of the true uptake coefficient (Crowley et al., 2010). Furthermore, under UV light 258 259 conditions, the observed linear increase also indicates that the sample surface is homogeneously 260 irradiated, irrespectively of the sample mass (Herrmann, 2005). Similar trends were observed for the 261 remaining volcanic samples and have been also reported in literature for mineral oxides and natural mineral dusts (El Zein et al., 2013a, 2013b; Romanias et al., 2016; Zein et al., 2014). 262

**263 3.1.2 Dependence of uptake coefficients on the irradiance intensity** 

To evidence potential photocatalytic or photoinduced processes on the surface of the volcanic samples, 264 265 the irradiance intensity of the incident light was varied inside the reactor. The experiments were carried out at 30% of RH, 293 K using 20 ppbV of NO<sub>2</sub> as inlet concentration and the steady-state 266 267 uptake coefficient was measured as a function of the number of UV lamps switched on, varying from 268 0 to 3 lamps. Appendix A Fig. S4 displays the results on Hagavatn sample. A linear increase of the uptake coefficient with the irradiance intensity was observed. It varied almost by a factor of 10 269 comparing the values under dark and 3 UV lamps, pointing out that NO<sub>2</sub> degradation rate is linearly 270 271 driven by the radian flux (Herrmann, 2005). However, if photons clearly induce specific 272 photoactivated phenomena on the v-dust surfaces it cannot be determined at this point whether they 273 are photocatalytic or photochemical. Nevertheless, similar trends have been noted in literature for 274 mineral oxides and natural mineral dusts (El Zein and Bedjanian, 2012; Tang et al., 2017; Zein et al., 275 2014).

276 **3.1.3 Dependence of uptake coefficients on NO<sub>2</sub> concentration** 

The dependence of the steady state uptake coefficient on the inlet concentration of NO<sub>2</sub> was studied in the range of 20 to 140 ppbV. In this set of experiments Hagavatn volcanic sample was exposed to NO<sub>2</sub> at T = 293 K, 30% of RH under both dark and UV irradiation conditions and the results are displayed in **Fig. 4**. The simulated sunlight irradiation of the surface has promoted the uptake of NO<sub>2</sub> by a factor of three in the entire concentration range. Furthermore, under both dark and light conditions, the uptake coefficient was found to be almost constant for concentrations below 40 ppbV, however, above that limit a slight decrease was observed; the latter may be attributed to the surface saturation by the adsorbed intermediates. The solid lines of **Fig. 4** correspond to the empirical power fit of results according to the following expressions:

286 
$$\gamma_{ss,BET}(dark) = (3.9 \pm 1.2) \times 10^{-8} [NO_2]^{-0.53}$$
 (3)

287 
$$\gamma_{ss,BET}(simulated sunlight) = (2.5 \pm 0.8) \times 10^{-7} [NO_2]^{-0.73}$$
 (4)

where  $\gamma_{ss,BET}$  (dark) and  $\gamma_{ss,BET}$  (simulated sunlight) refer to dark and simulated sunlight conditions, respectively, and [NO<sub>2</sub>] (ppbV) is the initial concentration of NO<sub>2</sub>. The quoted uncertainties correspond to the two sigma (2 $\sigma$ ) precision of the fit of experimental results. Therefore, one can extrapolate the uptake coefficient of NO<sub>2</sub> at any tropospheric concentration range.

#### **292 3.1.4 Dependence of NO<sub>2</sub> uptake coefficient on RH**

The uptake coefficients of  $NO_2$  on Hagavatn sample were determined as a function of relative humidity from dry conditions to 75% RH, the results are displayed in **Fig. 5**. Aiming to isolate the effect of RH on the kinetic parameter, in this series of experiments the volcanic sample was always exposed to 60 ppbV of  $NO_2$  at 293 K. Under dark conditions, the steady state uptake coefficient was found to be independent of RH in the whole investigated RH range. The solid black line in **Fig. 5**, corresponds to the mean value of the steady state coefficient in the entire RH range:

299  $\gamma_{ss,BET}(dark) = (3.6 \pm 0.2) \times 10^{-9}$  (5)

300 the quoted uncertainties correspond to the  $2\sigma$  precision of the fit of experimental results.

- As shown in **Fig. 5**, the UV irradiation of the sample surface promoted the consumption of NO<sub>2</sub> in the entire RH range compared to dark conditions; the  $\gamma_{ss,BET}$  under simulated sunlight was promoted by a factor of 3 compared to dark and was found to be independent of RH under ambient relative humidity conditions i.e. RH higher than 10%. The dash black line in **Fig. 5**, corresponds to the mean value of
- 305 the steady state coefficient for ambient RH conditions:

306  $\gamma_{ss,BET}$ (simulated sunlight) = (9.4 ± 0.5) × 10<sup>-9</sup> (6)

307 Interestingly, under dry conditions, the enhancement of  $\gamma_{ss,BET}$  upon surface irradiation was by a factor 308 of 6.

309 The independent nature of  $\gamma_{ss,BET}$  on RH does not preclude that water molecules do not affect the reaction mechanism. More precisely, under UV light irradiation, H<sub>2</sub>O is a well-known source of 310 311 radicals (i.e. 'OH,  $O_2$ ') on mineral dust surfaces that contain photocatalysts (Chen et al., 2012; 312 Herrmann, 2005; Romanias et al., 2012; Schneider et al., 2014) and thus enhance the reactivity of the 313 surface towards NO<sub>2</sub> (Herrmann, 2005; Romanias et al., 2012; Zein et al., 2014). On the other hand, 314 adsorbed water may also block some sorptive and reactive sites and induce a decrease to  $\gamma_{ss,BET}$ 315 (Romanias et al., 2012, 2017; Zein et al., 2014). It seems that under ambient RH conditions (RH > 316 30%), the effect of sites blocking by water is probably compensated by its role as a source of radical 317 species on the surface of the dust. Under dark conditions, radical species are not expected to be 318 formed. However, the disproportionation of  $NO_2$  can occur on the surface of the volcanic sample 319 leading to a continuous consumption of  $NO_2$  (Finlayson-Pitts et al., 2003). Further details regarding 320 the reaction mechanism under both dark and light conditions versus relative humidity are given in 321 reaction mechanism section.

322 3.1.5 Comparison of NO<sub>2</sub> uptake coefficients on various volcanic samples under defined

323 experimental conditions

324 Aiming to compare the uptake coefficients of NO<sub>2</sub> on the different volcanic samples a series of 325 experiments were carried out under defined and atmospheric experimental conditions: 30% RH, NO<sub>2</sub> 326 concentration 60 ppbV, under both dark and simulated sunlight. The results obtained are displayed in 327 Fig. 6, where for comparison purposes the international union of pure and applied chemistry, IUPAC 328 recommended value on mineral oxides is also displayed. A significant variation of the  $\gamma_{ss,BET}$  as a 329 function of the origin of the samples can be noted. In addition, the simulated sunlight radiation of the 330 volcanic samples induced an increase to the  $\gamma_{ss,BET}$  for all samples indicating the existence of photo-331 induced processes on the surface of all dusts.

332 The dependence noted on the  $\gamma_{ss,BET}$  values from one volcanic sample to another is significant. Despite the fact that samples originate from close geographical regions, the chemical composition of the 333 334 magma that produced them is different (see SI materials) and thus variations in the composition of the 335 natural samples and contrasted surface properties are expected. Furthermore, taking into account that their crystalline fraction is relatively low, ca. 20% for all samples beside Hagavatn, it makes any 336 337 correlation between mineralogical composition and reactivity challenging. The relative abundance of 338 elements could have been a base for comparison, however, it is quite similar for all samples, and no 339 relevant correlation trends were noticed.

340 Interestingly, a correlation between the uptake coefficients and the specific surface area was noticed 341 for both dark and light conditions (Appendix A Fig. 5); the highest the specific surface area, the 342 lowest the uptake coefficient. The latter could indicate that morphological parameters can play an 343 important role and affect the surface properties of the samples studied. It has been shown that the specific surface area is inversely correlated with the particles size in literature (Ibrahim et al., 2018) – 344 345 and that small particles are more irregular. Moreover, it has been evidenced that the water monolayer 346 is formed at lower RH for natural mineral samples with small particle sizes due to higher OH group 347 surface density (Ibrahim et al., 2018; Tang et al., 2016). Therefore, a possible explanation could be 348 that the  $NO_2$  uptake or/and the abundance of intermediate surface species formed are influenced by the 349 presence of surface OH group density.

The  $\gamma_{ss, BET}$  values measured for the volcanic samples are in the same order of magnitude but relatively lower than the IUPAC recommendation. However, it should be stressed out that the IUPAC recommended value is based on studies performed with mineral oxides and not natural samples under low pressure and dry conditions, although RH is a governing factor for NO<sub>2</sub> uptake. Recent studies performed on oxides and natural or synthetic mineral dusts report uptake coefficients of NO<sub>2</sub> in the order of 10<sup>-9</sup> (Crowley et al., 2010; Li et al., 2010; Ndour et al., 2009), and thus similar or even lower than those determine in the current study.

#### **357 3.2 Investigation of products formed along with NO<sub>2</sub> uptake**

358 The results presented in this section concern the determination of the gas-phase products of NO<sub>2</sub> 359 interaction with the volcanic surfaces. The yield of the detected products (PY) was determined as the 360 ratio of the product concentration formed ( $\Delta$ [product]) to the concentration of NO<sub>2</sub> consumed 361 ( $\Delta$ [NO<sub>2</sub>]) after achieving steady state conditions:

362  $PY = \Delta[product]/\Delta[NO_2]$  (7)

To investigate the impact of experimental conditions, Hagavatn volcanic sample was used as model volcanic sample. Finally, at the end of the section the product yields determined for all volcanic samples under defined experimental conditions are presented and discussed.

#### **366 3.2.1** Product yield as a function of NO<sub>2</sub> initial concentration and relative humidity

The impact of experimental conditions, such as NO<sub>2</sub> initial concentration (in the range of 40 to 140 367 ppbV) and relative humidity (from dry to 75% RH) on the yield of the products formed was 368 369 investigated at room temperature, under both dark and simulated sunlight, using Hagavatn as model 370 surface sample. In all cases, NO and HONO were measured as the only nitrogen containing gas phase 371 products. Regarding concentration dependence, the experiments were carried out at 30% of RH and 372 the yields of NO and HONO were  $40\% \pm 10\%$  and  $60\% \pm 10\%$ , respectively, independent on NO<sub>2</sub> 373 initial concentration. The product yield measurements as a function of RH were carried out at 60 ppbV 374 of initial NO<sub>2</sub>. As displayed in Fig. 7, NO and HONO were detected as primary gas phase products, 375 with almost identical values under both dark and simulated sunlight conditions, however, their 376 formation was highly dependent on relative humidity. As relative humidity increases, the formation of 377 HONO is favored while that of NO is diminished. The latter is in accordance with the literature where 378 similar trends have been reported for mineral oxides (Baltrusaitis et al., 2009; Bedjanian and El Zein, 379 2012) and results seems to justify the mechanism proposed about enhanced formation of HONO on 380 humid surfaces (Finlayson-Pitts et al., 2003; Rodriguez et al., 2001). In addition, for RH above 20% a nitrogen mass balance evidenced the complete conversion of adsorbed  $NO_2$  to the gas phase products. 381 However, for RH below 10%, the nitrogen mass balance was not completed only considering the gas 382 phase products, thus the presence of nitrogen containing species on the surface, i.e. adsorbed  $NO_2$  or 383

surface reaction products, is expected. The latter indicates that water plays a key role on the sorption and/or reaction mechanism facilitating the transformation of  $NO_2$  to gas phase products. Similar formation yields and trend were noticed for the other volcanic samples and have been also reported in literature on the surface of mineral oxides and natural desert dusts (Bedjanian and El Zein, 2012; Tang et al., 2017).

389

390 3.2.2 NO and HONO product yields on various volcanic samples under defined experimental conditions 391 392 In this series of experiments the volcanic samples were exposed to 60 ppbV concentration of NO<sub>2</sub>, 393 30% of RH under both dark and simulated sunlight irradiation conditions. The results are presented in 394 Table 1. Considering the experimental uncertainties, no significant variation was noticed to the 395 formation yield of NO between the 5 volcanic samples under dark and simulated sunlight conditions 396 although significant differences were observed for the uptake coefficients. On the contrary, the HONO 397 formation yields for Eyjafjallajökull typical ash, Mýrdalssandur and Hagavatn, were systematically 398 higher than those measured for Dyngjusandur and Maelifellssandur v-dusts. Interestingly, it seems that HONO formation in the gas phase is promoted on volcanic dusts with higher NO<sub>2</sub> uptake coefficients. 399

400 In addition, the irradiation of the surface increased the formation of NO for all samples while the 401 HONO yield was not influenced. The latter observation suggests that (i) either surface adsorbed 402 nitrogen species are photolysed or (ii) the surface is photoactivated and photocatalytic processes occur, 403 leading to the formation of NO. It should be noted that HONO photolysis in the gas phase is not 404 expected to occur in our system since the photolysis frequency of HONO is almost one order of 405 magnitude slower than NO<sub>2</sub>, thus it is not expected to occur in our reactor, and HONO yields didn't 406 decrease upon surface irradiation. Concerning the total formation yields of gas phase products, under 407 surface irradiation a systematic increase was observed (except for typical ash sample) compared with 408 dark conditions, leading to an almost completion of the nitrogen mass balance for all samples within 409 the experimental uncertainties.

#### 410 **3.3 Proposed reaction mechanism**

#### 411 **3.3.1 Dark conditions**

412 The heterogeneous degradation of  $NO_2$  on the surface of volcanic samples resulted in the formation of 413 both NO and HONO in the gas phase. Several pathways have been proposed to occur for the 414 transformation of NO<sub>2</sub> on mineral surfaces. In particular, NO<sub>2</sub> can be initially physisorbed on metal 415 sites  $(SS^n \text{ stands for surface site})$  of the mineral surface (Eq. (8)) and then disproportionate (Eq. (9)), 416 leading to the formation of surface nitrates (Eq. (9)) and gaseous NO (Eq. (10)) (Finlayson-Pitts et al., 417 2003; Rodriguez et al., 2001). Furthermore, NO2 can directly react with surface O atoms, SS-O, and 418 form surface nitrates (Eq. (11)) (Baltrusaitis et al., 2009). Alternatively NO<sub>2</sub> can be chemisorbed and reduced to  $NO_2^-$  by metal sites, i.e. ferrous iron sites  $Fe^{2+}$ , expected to be relatively abundant in the 419 420 volcanic samples, in an electron transfer reaction that is facilitated by adsorbed water (Eqs. (12) and (13)) leading to HONO formation (Kebede et al., 2016). In the presence of water, the abundance of 421 422 surface OH groups increases, SS-OH (Ibrahim et al., 2018; Joshi et al., 2017; Tang et al., 2016), and 423 according to literature, NO<sub>2</sub> can form a hydrogen bond stabilized ONOOH species that could either lead to the formation of gas HONO or isomerize and form adsorbed nitric acid (Eqs. (14) and (15)) 424 425 (Baltrusaitis et al., 2009).

427	$SS^n + NO_{2(g)} \leftrightarrow SS^n \text{-}NO_{2(ads)}$	(8)
428	$2 [SS^{n}-NO_{2(ads)}] \rightarrow SS^{n}-NO_{3(ads)}^{-} + SS^{n}-NO_{(ads)}^{+}$	(9)
429	$SS^{n}-NO^{+}_{(ads)} \rightarrow SS^{n}+NO_{(g)}$	(10)
430	$SS-O + NO_{2(g)} \rightarrow SS^{n+1}-NO_{3(ads)}$	(11)
431	$SS^{n} + NO_{2(g)} \rightarrow SS^{n+1} - NO_{2(ads)}$	(12)
432	$SS^{n+1}-NO_2^{-}_{(ads)} + H^+ \rightarrow SS^n-HONO_{(ads)}$	(13)
433	$SS-OH + NO_{2(g)} \rightarrow SS-ONOOH_{(ads)} \rightarrow SS-O + HONO_{(gas)}$	(14)

#### 434 SS-ONOOH $\rightarrow$ SS-HNO<sub>3</sub> (isomerization)

(15)

435 The proposed reaction mechanism is in agreement with our experimental observations. Under dry 436 conditions, the dominant reaction pathways are: the disproportionation of NO<sub>2</sub>, leading to high 437 formation yields of NO<sub>(g)</sub> (Eqs. (8)-(10)); and reactions (11) and (12) respectively forming surface 438 nitrates and nitrites. Reactions (13)-(15) are still active pathways even under dry conditions, since the 439 removal of strongly adsorbed surface water requires a stronger thermal pretreatment method (> 500 K) 440 than the one followed in our study (Joshi et al., 2017). Consequently, the sequence of reactions proposed seems to explain the high formation yields of NO, and the corresponding low HONO yields 441 observed for Hagavatn sample under dry conditions, as well as the absence of nitrogen mass balance 442 443 of the gas products due to the formation of surface adsorbed nitrogen containing species.

By increasing the relative humidity, the surface density of OH groups also increases considerably (especially after water monolayer formation that is expected at RH greater than 25%-30% RH, (Joshi et al., 2017), and thus reactions (13)-(15) which are competing to reactions (8), (9), (11) and (12) , become the dominant pathways leading to an enhanced formation of HONO. Consequently, the formation of NO is diminished similar with the concentration of surface adsorbed species and a nitrogen mass balance is observed.

450

#### 451 3.2.2 UV light conditions

452 Under simulated sunlight irradiation of the surface, and comparing the results obtained under fixed experimental conditions (i.e. fixed NO<sub>2</sub> concentration and 30% RH - see Table 1) the uptake of NO<sub>2</sub> 453 454 was enhanced, however unlike HONO the yield of NO was significantly influenced. These observations could possibly point to the presence of photo-induced processes on the surface of 455 456 volcanic samples. Note that photocatalytic processes could also take place but considering that the 457 detailed mineralogical composition of the volcanic samples is not known and that samples are mostly 458 amorphous, the consideration of these processes is speculative and that they are not considered in the 459 proposed reaction mechanism.

460 The surface adsorbed nitrite species formed under dark conditions through reaction (12) could be

461 photolysed in presence of simulated sunlight (hv) producing gaseous NO and regenerating a SS-O

462 surface site according to reaction:

463 
$$SS^{n+1}-NO_{2(ads)}^{-} + hv \rightarrow SS-O + NO_{(g)}$$
 (16)

464 Similarly, nitrates formed through reaction (11) can be the source of gaseous NO and NO<sub>2</sub> through
465 reactions (17) and (18):

466 
$$SS^{n+1}-NO_{3(ads)} + hv \rightarrow SS^{n} + NO_{(g)} + O_{2}$$
 (17)

467  $SS^{n+1}-NO_{3(ads)} + hv \rightarrow SS-O + NO_{2(g)}$  (18)

Reaction pathways (16)-(18) corroborate the photoinduced consumption of NO<sub>2</sub> under simulated
sunlight conditions, as well as the slight but systematic photoenhanced formation of NO observed on
volcanic samples under 30% of RH.

#### 471 **4.** Conclusions and atmospheric implications

472 Within the framework of the current study, the reactivity of Icelandic volcanic samples towards  $NO_2$ was studied as a function of several atmospheric parameters. Despite the relative close regions where 473 474 the samples were collected, the uptake coefficients varied significantly from one sample to another. It 475 is suggested that the origin of the magma and morphological differences between the samples could 476 possibly explain the variations noted. Besides the kinetics, the formation yields of NO and HONO 477 were also determined. The relative humidity was found to influence the balance between adsorbed and gas phase products as well as the distribution of NO and HONO yields. Interestingly, under 478 479 atmospheric conditions (i.e. > 30% RH, ambient temperature) higher HONO formation yields were noticed for the volcanic samples. Finally, based on the experimental observations of the current study, 480 a reaction mechanism was proposed. 481

482 Regarding the atmospheric implications, we estimate that the corresponding impacts of  $NO_2$ 483 transformation to HONO is expected to be significant in cases of volcanic eruptions or during more 484 frequent events such as volcanic dust storms. Indeed, using Iceland as a case study, it has been evidenced that volcanic dust storms are much more frequent than volcanic eruptions transporting
volcanic aerosols to central Europe or Arctic (Dagsson-Waldhauserova et al., 2014a). Interestingly,
about half of these events occur in winter (dark season) when NO<sub>2</sub> concentrations in Reykjavik can
exceed 60 ppbV.

489 In an attempt to estimate the atmospheric significance of NO<sub>2</sub> heterogeneous degradation on volcanic 490 particles, the rate coefficient of NO<sub>2</sub> heterogeneous loss ( $k_{het}$  (hr<sup>-1</sup>)) was calculated from the following 491 equation:

492 
$$k_{\text{het}} = \frac{\gamma \times c \times A}{4} \tag{19}$$

493

494 where,  $\gamma$  is the NO<sub>2</sub> uptake coefficient, c (cm/hr) is the mean molecular velocity, and A the aerosol surface area density  $(cm^2/cm^3)$ . Using Iceland as case study, under typical PM<sub>10</sub> concentrations during 495 a v-dust storm, i.e. ca. 300 µg/m<sup>3</sup> (Arnalds et al., 2016, 2014; Dagsson-Waldhauserova et al., 2015; 496 497 Thorsteinsson et al., 2011), RH greater than 70%, NO<sub>2</sub> concentrations of ca 8 ppbV (annual average 498 values) (Hazenkamp-von Arx et al., 2004) and extrapolating the uptake coefficient of NO<sub>2</sub> under dark 499 conditions corresponding to winter season using Eq. (3), the heterogeneous loss of NO<sub>2</sub> is relatively low (in the order of  $10^{-6}$  hr<sup>-1</sup>) and the estimated HONO formation rate is not expected to exceed 0.1 500 501 pptV/hr. Even considering the peak NO<sub>2</sub> concentration of 60 ppbV the HONO formation rate is below 0.3 pptV/hr. However, during severe dust events (7000  $\mu$ g/m<sup>3</sup>), the heterogeneous loss of NO<sub>2</sub> could 502 be substantially higher (up to  $10^{-4}$  hr<sup>-1</sup>) and HONO formation rate could reach 4 pptV/hr. Considering 503 504 the severe dust event scenario during the summer season up to 10 pptV/hr of HONO can be expected 505 and thus a significant impact to the oxidative capacity of the regional atmosphere.

To conclude, the significance of  $NO_2$  scavenging on the surface of volcanic particles is mainly dependent on the v-dust atmospheric loads. Nevertheless, this study clearly evidences that volcanic dust particles can efficiently uptake, scavenge and transform  $NO_2$ . Similar trends could be expected for other pollutants. Indeed, in the recent study of Urupina et al. (2019) it was evidenced an efficient uptake and transformation of sulfur dioxide to sulfates on these Icelandic volcanic dusts. Therefore, 511 Icelandic v-dust aerosols should be considered as reactive components of the atmosphere; In the 512 context of their long range transportation to Europe and especially Arctic it is expected to transport, 513 scavenge and transform pollutants and thus to alter the corresponding air quality and chemistry of the 514 atmosphere.

#### 515 Appendix A. Supplementary data

516 Supplementary data associated with this article can be found in the online version at xxxxxx.

517

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#### 528 **References**

- 529 Arnalds, O., Dagsson-Waldhauserova, P., Olafsson, H., 2016. The Icelandic volcanic aeolian
- environment: Processes and impacts A review. Aeolian Res. 20, 176-195.
- 531 Arnalds, O, Olafsson, H, Dagsson-Waldhauserova, P, 2014. Quantification of iron-rich volcanogenic
- dust emissions and deposition over the ocean from Icelandic dust sources. Biogeosciences 11, 6623-6632.
- Arnalds, O, Thorarinsdottir, E F, Thorsson, J, Waldhauserova, P D, Agustsdottir, A M, 2013. An
- extreme wind erosion event of the fresh Eyjafjallajökull 2010 volcanic ash. Sci. Rep. 3, 1257.
- 536 Auker, M R, Sparks, R S J, Siebert, L, Crosweller, H S, Ewert, J, 2013. A statistical analysis of the
- 537 global historical volcanic fatalities record. J. Appl. Volcanol. 2, 2.
- 538 Baltrusaitis, J, Jayaweera, PM, Grassian, VH, 2009. XPS study of nitrogen dioxide adsorption on
- metal oxide particle surfaces under different environmental conditions. Phys. Chem. Chem. Phys. 11,8295-8305.
- 541 Baratoux, D, Mangold, N, Arnalds, O, Bardintzeff, J-M, Platevoët, B, Grégoire, M, et al., 2011.
- 542 Volcanic sands of Iceland Diverse origins of aeolian sand deposits revealed at Dyngjusandur and
- 543 Lambahraun. Earth Surf. Process. 36,1789-1808.
- 544 Barnard, J C, Chapman, E G, Fast, J D, Schmelzer, J R, Slusser, J R, Shetter, R E, 2004. An evaluation
- of the FAST-J photolysis algorithm for predicting nitrogen dioxide photolysis rates under clear and
- cloudy sky conditions. Atmos. Environ. 38, 3393-3403.
- Bedjanian, Y, El Zein, A, 2012. Interaction of NO<sub>2</sub> with TiO<sub>2</sub> surface under UV irradiation: Products
  study. J. Phys. Chem. A 116, 1758-1764.
- Bohn, B, Rohrer, F, Brauers, T, Wahner, A, 2005. Actinometric measurements of NO<sub>2</sub> photolysis
- frequencies in the atmosphere simulation chamber SAPHIR. Atmos. Chem. Phys. 5, 493-503.
- 551 Boichu, M, Chiapello, I, Brogniez, C, Pere, J C, Thieuleux, F, Torres, B, et al., 2016. Current
- challenges in modelling far-range air pollution induced by the 2014-2015 Baroarbunga fissure
- eruption (Iceland). Atmos. Chem. Phys. 16, 10831-10845.
- 554 Butwin, M K, von Löwis, S, Pfeffer, M A, Thorsteinsson, T, 2019. The effects of volcanic eruptions
- on the frequency of particulate matter suspension events in Iceland. J. Aerosol Sci. 128, 99-113.
- 556 Chen, H, Nanayakkara, C E, Grassian, V H, 2012. Titanium dioxide photocatalysis in atmospheric
- 557 chemistry. Chem. Rev. 112, 5919-5948.
- 558 Crowley, J N, Ammann, M, Cox, R A, Hynes, R G, Jenkin, M E, Mellouki, A, et al., 2010. Evaluated
- 559 kinetic and photochemical data for atmospheric chemistry: Volume V heterogeneous reactions on
- solid substrates. Atmos. Chem. Phys. 10, 9059-9223.
- Dagsson-Waldhauserova, P, Arnalds, O, Olafsson, H, 2014a. Long-term variability of dust events in
   Iceland (1949–2011). Atmos. Chem. Phys. 14, 13411-13422.
- 563 Dagsson-Waldhauserova, P, Arnalds, O, Olafsson, H, Skrabalova, L, Sigurdardottir, G M, Branis, M,
- et al., 2014b. Physical properties of suspended dust during moist and low wind conditions in Iceland.
  Icel. Agric. Sci. 27, 25-39.
- 566 Dagsson-Waldhauserova, P, Arnalds, O, Olafsson, H, Hladil, J, Skala, R, Navratil, T, et al., 2015.
- 567 Snow–Dust Storm: Unique case study from Iceland, March 6–7, 2013. Aeolian Res. 16, 69-74.
- 568 Dagsson-Waldhauserova, P, Magnusdottir, A O, Olafsson, H, Arnalds, O, 2016. The Spatial Variation
- of Dust Particulate Matter Concentrations during Two Icelandic Dust Storms in 2015. Atmosphere 7.
- 570 Đorđević, D, Tošić, I, Sakan, S, Petrović, S, Đuričić-Milanković, J, Finger, D C, et al., 2019. Can
- volcanic dust suspended from surface soil and deserts of Iceland be transferred to Central Balkan
- similarly to African dust (Sahara)? Front. Earth Sci. 7.
- 573 El Zein, A, Bedjanian, Y, 2012. Interaction of NO<sub>2</sub> with TiO<sub>2</sub> surface under UV irradiation:
- measurements of the uptake coefficient. Atmos. Chem. Phys. 12, 1013-1020.
- 575 El Zein, A, Bedjanian, Y, Romanias, M N, 2013a. Kinetics and products of HONO interaction with
- 576 TiO<sub>2</sub> surface under UV irradiation. Atmos. Environ. 67, 203-210.
- 577 El Zein, A, Romanias, M N, Bedjanian, Y, 2013b. Kinetics and products of heterogeneous reaction of
- 578 HONO with  $Fe_2O_3$  and Arizona test dust. Environ. Sci. Technol. 47, 6325-6331.
- 579 Engelstaedter, S, Tegen, I, Washington, R, 2006. North African dust emissions and transport. Earth-
- 580 Science Rev. 79, 73-100.

- 581 Finlayson-Pitts, B J, Wingen, L M, Sumner, A L, Syomin, D, Ramazan, K A, 2003. The
- 582 heterogeneous hydrolysis of NO2 in laboratory systems and in outdoor and indoor atmospheres: An
- integrated mechanism. Phys. Chem. Chem. Phys. 5, 223-242. 583
- 584 George, C, Ammann, M, D'Anna, B, Donaldson, D J, Nizkorodov, S A, 2015. Heterogeneous
- 585 Photochemistry in the Atmosphere. Chem. Rev. 115, 4218-4258.
- Gislason, S R, Hassenkam, T, Nedel, S, Bovet, N, Eiriksdottir, E S, Alfredsson, H A, et al., 2011. 586
- Characterization of Eyjafjallajökull volcanic ash particles and a protocol for rapid risk assessment. 587 588 Proc. Nat. Acad. Sci. 108, 7307.
- Groot Zwaaftink, C D, Arnalds, Ó, Dagsson-Waldhauserova, P, Eckhardt, S, Prospero, J M, Stohl, A, 589
- 590 2017. Temporal and spatial variability of Icelandic dust emissions and atmospheric transport. Atmos. Chem. Phys. 17, 10865-10878. 591
- 592 Gudmundsson, M T, Thordarson, T, Höskuldsson, Á, Larsen, G, Björnsson, H, Prata, F J, et al., 2012.
- 593 Ash generation and distribution from the April-May 2010 eruption of Eyjafjallajökull, Iceland. Sci. 594 Rep. 2, 572.
- 595 Hazenkamp-von Arx, M E, Götschi, T, Ackermann-Liebrich, U, Bono, R, Burney, P, Cyrys, J, et al.,
- 596 2004. PM<sub>2.5</sub> and NO<sub>2</sub> assessment in 21 European study centres of ECRHS II: annual means and
- 597 seasonal differences. Atmos. Environ. 38, 1943-1953.
- 598 Herrmann, J M, 2005. Heterogeneous photocatalysis: state of the art and present applications. Top. 599 Catal. 34, 49-65.
- 600 Ibrahim, S, Romanias, M N, Alleman, L Y, Zeineddine, M N, Angeli, G K, Trikalitis, P N, et al.,
- 2018. Water interaction with mineral dust aerosol: Particle size and hygroscopic properties of dust. 601
- 602 ACS Earth Space Chem. 2, 376-386.
- Ilyinskaya, E, Schmidt, A, Mather, T A, Pope, F D, Witham, C, Baxter, P, et al., 2017. Understanding 603
- the environmental impacts of large fissure eruptions: Aerosol and gas emissions from the 2014–2015 604
- 605 Holuhraun eruption (Iceland). Earth Planet. Sci. Lett. 472, 309-322.
- 606 Joshi, N, Romanias, M N, Riffault, V, Thevenet, F, 2017. Investigating water adsorption onto natural
- mineral dust particles: Linking DRIFTS experiments and BET theory. Aeolian Res. 27, 35-45. 607
- Kebede, M A, Bish, D L, Losovyj, Y, Engelhard, M H, Raff, J D, 2016. The role of iron-bearing 608
- 609 minerals in NO2 to HONO conversion on soil surfaces. Environ. Sci. Technol. 50, 8649-8660.
- Kleffmann, J. 2007. Daytime Sources of Nitrous Acid (HONO) in the Atmospheric Boundary Layer. 610 ChemPhysChem 8, 1137-1144. 611
- Lasne, J, Romanias, M N, Thevenet, F, 2018. Ozone Uptake by Clay Dusts under Environmental 612
- Conditions. ACS Earth Space Chem. 2, 904-914. 613
- 614 Li, H J, Zhu, T, Zhao, D F, Zhang, Z F, Chen, Z M, 2010. Kinetics and mechanisms of heterogeneous
- 615 reaction of NO<sub>2</sub> on CaCO<sub>3</sub> surfaces under dry and wet conditions. Atmos. Chem. Phys. 10, 463-474.
- Mellouki, A, Wallington, T J, Chen, J, 2015. Atmospheric chemistry of oxygenated volatile organic 616
- compounds: Impacts on air quality and climate. Chem. Rev. 115, 3984-4014. 617
- Moroni, B, Arnalds, O, Dagsson-Waldhauserová, P, Crocchianti, S, Vivani, R, Cappelletti, D, 2018. 618
- 619 Mineralogical and chemical records of icelandic dust sources upon Ny-Ålesund (Svalbard Islands). Front. Earth Sci. 6. 620
- 621 Nault, B A, Garland, C, Wooldridge, P J, Brune, W H, Campuzano-Jost, P, Crounse, J D, et al., 2016.
- 622 Observational constraints on the oxidation of NOx in the upper troposphere. J. Phys. Chem. A 120, 1468-1478. 623
- Ndour, M, Nicolas, M, D'Anna, B, Ka, O, George, C, 2009. Photoreactivity of NO2 on mineral dusts 624
- 625 originating from different locations of the Sahara desert. Phys. Chem. Chem. Phys. 11, 1312-1319.
- 626 Ovadnevaite, J, Ceburnis, D, Plauskaite-Sukiene, K, Modini, R, Dupuy, R, Rimselyte, I, et al., 2009.
- 627 Volcanic sulphate and arctic dust plumes over the North Atlantic Ocean. Atmos. Environ. 43, 4968-628 4974.
- 629 Perring, A E. Pusede, S E. Cohen, R C. 2013. An observational perspective on the atmospheric
- 630 impacts of alkyl and multifunctional nitrates on ozone and secondary organic aerosol. Chem. Rev.
- 113, 5848-5870. 631
- 632 Pfeffer, M A, Bergsson, B, Barsotti, S, Stefansdottir, G, Galle, B, Arellano, S, et al., 2018. Ground-
- based measurements of the 2014-2015 Holuhraun volcanic cloud (Iceland). Geosciences 8. 633

- Rodriguez, J A, Jirsak, T, Liu, G, Hrbek, J, Dvorak, J, Maiti, A, 2001. Chemistry of NO<sub>2</sub> on oxide
- surfaces: Formation of NO<sub>3</sub> on TiO<sub>2</sub>(110) and NO<sub>2</sub> $\leftrightarrow$ O vacancy interactions. J. Am. Chem. Soc. 123, 9597-9605.
- 637 Romanias, M N, Bedjanian, Y, Zaras, A M, Andrade-Eiroa, A, Shahla, R, Dagaut, P, et al., 2013.
- 638 Mineral oxides change the atmospheric reactivity of soot: NO<sub>2</sub> uptake under dark and UV irradiation
- 639 conditions. J. Phys. Chem. A 117, 12897-12911.
- 640 Romanias, M N, El Zein, A, Bedjanian, Y, 2012. Heterogeneous interaction of H<sub>2</sub>O<sub>2</sub> with TiO<sub>2</sub> surface
- under dark and UV light irradiation conditions. J. Phys. Chem. A 116, 8191-8200.
- 642 Romanias, M N, Zeineddine, M N, Gaudion, V, Lun, X, Thevenet, F, Riffault, V, 2016.
- Heterogeneous interaction of isopropanol with natural Gobi dust. Environ. Sci. Technol. 50, 11714-11722.
- Romanias, M N, Zeineddine, M N, Riffault, V, Thevenet, F, 2017. Isoprene heterogeneous uptake and
- reactivity on  $TiO_2$ : A kinetic and product study. Int. J. Chem. Kinet. 49, 773-788.
- 647 Romanías, M N, Ourrad, H, Thévenet, F, Riffault, V, 2016. Investigating the heterogeneous
- 648 interaction of VOCs with natural atmospheric particles: Adsorption of limonene and toluene on
- 649 Saharan mineral dusts. J. Phys. Chem. A 120, 1197-1212.
- Romer, P S, Wooldridge, P J, Crounse, J D, Kim, M J, Wennberg, P O, Dibb, J E, et al., 2018.
- 651 Constraints on aerosol nitrate photolysis as a potential source of HONO and NOx. Environ. Sci.
- 652 Technol. 52, 13738-13746.
- 653 Schneider, J, Matsuoka, M, Takeuchi, M, Zhang, J, Horiuchi, Y, Anpo, M, et al., 2014. Understanding
- $TiO_2$  photocatalysis: Mechanisms and materials. Chem. Rev. 114, 9919-9986.
- Tang, M, Cziczo, D J, Grassian, V H, 2016. Interactions of sater with mineral dust aerosol: Water
- adsorption, hygroscopicity, cloud condensation, and ice nucleation. Chem. Rev. 116, 4205-4259.
- Tang, M J, Cox, R A, Kalberer, M, 2014. Compilation and evaluation of gas phase diffusion
- coefficients of reactive trace gases in the atmosphere: volume 1. Inorganic compounds. Atmos. Chem.Phys. 14, 9233-9247.
- Tang, M J, Huang, X, Lu, K D, Ge, M F, Li, Y J, Cheng, P, et al., 2017. Heterogeneous reactions of
- mineral dust aerosol: implications for tropospheric oxidation capacity. Atmos. Chem. Phys. 17, 11727-11777.
- Thorsteinsson, T, Gísladóttir, G, Bullard, J, McTainsh, G, 2011. Dust storm contributions to airborne
   particulate matter in Reykjavík, Iceland. Atmos. Environ. 45, 5924-5933.
- Topaloglou, C, Kazadzis, S, Bais, A F, Blumthaler, M, Schallhart, B, Balis, D, 2005. NO<sub>2</sub> and HCHO
- photolysis frequencies from irradiance measurements in Thessaloniki, Greece. Atmos. Chem. Phys. 5,1645-1653.
- 668 Urupina, D, Lasne, J, Romanias, M N, Thiery, V, Dagsson-Waldhauserova, P, Thevenet, F, 2019.
- 669 Uptake and surface chemistry of  $SO_2$  on natural volcanic dusts. Atmos. Environ. 217, 116942.
- 670 Vignelles, D, Roberts, T J, Carboni, E, Ilyinskaya, E, Pfeffer, M, Dagsson Waldhauserova, P, et al.,
- 671 2016. Balloon-borne measurement of the aerosol size distribution from an Icelandic flood basalt
- eruption. Earth Planet. Sci. Lett. 453, 252-259.
- von Schneidemesser, E, Monks, P S, Allan, J D, Bruhwiler, L, Forster, P, Fowler, D, et al., 2015.
- 674 Chemistry and the linkages between air quality and climate change. Chem. Rev. 115, 3856-3897.
- 675 Wilson, T M, Cole, J W, Stewart, C, Cronin, S J, Johnston, D M, 2011. Ash storms: impacts of wind-
- remobilised volcanic ash on rural communities and agriculture following the 1991 Hudson eruption,
- 677 southern Patagonia, Chile. Bull. Volcanol. 73, 223-239.
- 678 Zein, A E, Romanias, M N, Bedjanian, Y, 2014. Heterogeneous interaction of  $H_2O_2$  with Arizona test 679 dust. J. Phys. Chem. A 118, 441-448.
- 680 Zeineddine, M N, Romanias, M N, Gaudion, V, Riffault, V, Thevenet, F, 2017. Heterogeneous
- 681 interaction of isoprene with natural Gobi dust. ACS Earth Space Chem. 1, 236-243.
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# **Tables.**

**Table 1:** Product yields of NO and HONO determined for the five volcanic samples under fixed

687 expe	erimental condition	$s [NO_2] = 60 ppl$	bV and RH = 30	)% under dark and	simulated sunlight radiation.
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Volconia comple	Yield NO		Yield HONO		Total yield	
voicanic sample	Dark	Light	Dark	Light	Dark	Light
Eyjafjallajökull typical ash	28%±4%	48%±7%	71%±11%	53%±8%	99%±15%	101%±15%
Mýrdalssandur	22%±3%	44%±7%	52%±8%	56%±8%	74%±11%	100%±15%
Hagavatn	37%±6%	45%±7%	60%±9%	62%±9%	97%±15%	107%±16%
Dyngjusandur	39%±6%	50%±8%	37%±6%	38%±6%	76%±11%	88%±13%
Maelifellssandur	28%±4%	37%±6%	48%±7%	$47\% \pm 7\%$	76%±11%	84%±13%

# 689 **Figures**

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**Fig. 1:** Schematic representation of the coated-wall flow-tube (CWFT) reactor used in this study. The

space filled with the coolant in between the two walls is shaded in blue. The volcanic sample coating

the inner surface of the Pyrex tube is shown in gray. The gas phase was monitored in real time with a

Nitrogen oxide (NO<sub>x</sub>) analyzer, a long path absorption photometer (LOPAP) and occasionally with a (C + DC) = 0

695 cavity attenuated phase shift (CAPS)  $NO_2$  analyzer.



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Fig. 2: Absorption spectra of NO<sub>2</sub> and possible interferences (HONO, NO<sub>3</sub>) along with the spectral
 output of the Light-emitting diode (LED) centered at 395 nm.







**Fig. 3:** Typical experiment of NO<sub>2</sub> adsorption/reaction on Hagavatn volcanic sample where all instruments were coupled with the flow tube reactor. Left vertical axis corresponds to the concentration profile of NO<sub>2</sub>, [NO<sub>2</sub>]. The concentrations profiles of NO and HONO, denoted as [NO] and [HONO], respectively, are also displayed (right vertical axis).  $\gamma_{ss}$ (dark) and  $\gamma_{ss}$ (sunlight) refer to the data averaged to determine the uptake coefficients under dark and simulated sunlight conditions, respectively.





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**Fig. 4:** Uptake coefficients of NO<sub>2</sub> ( $\gamma_{ss,BET}$ ) on Hagavatn volcanic sample as a function of initial concentration under dark ( $\gamma_{ss,BET}$  (dark)) and simulated sunlight irradiation ( $\gamma_{ss,BET}$  (simulated sunlight)). Error bars denote the overall uncertainty on  $\gamma_{ss,BET}$  determination (ca. 25%) that arise mostly from the determination of Specific surface area (SSA). The solid and dashed lines corresponds to the fit of experimental data using an empirical power function, under dark and simulated sunlight

radiation respectively. Temperature was set at T = 293 K, and relative humidity, RH = 30%



723 Fig. 5: Uptake coefficients of NO<sub>2</sub> on Hagavatn volcanic sample as a function of relative humidity

<sup>724</sup> under darkand simulated sunlight irradiation. Error bars reflects the overall uncertainty on  $\gamma_{ss,BET}$ <sup>725</sup> determination (ca. 25%) that arise mostly from the determination of SSA. The solid and dashed lines

726 corresponds to the average value of the  $\gamma_{ss,BET}$ .



RH (%)

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- 730 Fig. 6: Uptake coefficients of NO<sub>2</sub> as a function of volcanic samples origin under dark and simulated
- 731 sunlight irradiation. For comparison purposes, the recommended uptake coefficient value proposed by
- 732 the international union of pure and applied chemistry (IUPAC) panel for mineral oxides under dark 733
- and dry conditions is given.



Fig. 7: Product yields of NO (cycles) and HONO (squares) formation determined as a function of relative humidity upon reaction of NO2 with Hagavatn volcanic sample under dark and simulated sunlight radiation. Error bars reflects the overall uncertainty on the product yields (ca. 15%). The solid and dashed lines are the empirical fit of experimental results (both dark and simulated sunlight) to demonstrate the corresponding trends of NO and HONO yields versus RH. 

