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Chemical composition and *in vitro* aryl hydrocarbon receptor-mediated activity of atmospheric particulate matter at an urban, agricultural and industrial site in North Africa (Bizerte, Tunisia)

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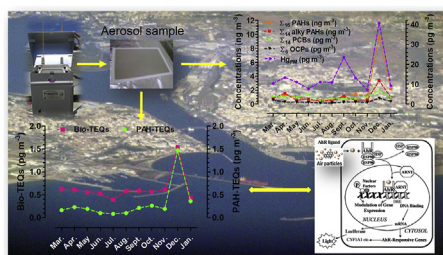
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HIGHLIGHTS

- The first investigation of atmospheric particulate Hg (Hg_{PM}) in whole Africa.
- PCA analysis showed that Hg_{PM} in Bizerte city was mainly produced by fossil fuel combustion.
- Ambient temperature, precipitation and wind speed were identified as the major factors influenced Hg_{PM} concentrations.
- All atmospheric PM organic extracts from Bizerte city showed dioxin-like activity.
- PAHs are responsible for 40% of dioxin-like activity of atmospheric PM.

GRAPHICAL ABSTRACT



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ABSTRACT

As recognized risk factor to pose a health threat to humans and wildlife globally, atmospheric particulate matter (PM) were collected from a North African coastal city (Bizerte, Tunisia) for one year, and were characterized for their chemical compositions, including mercury (Hg_{PM}), as well as organic contaminants (polycyclic aromatic hydrocarbons (PAHs), polychlorinated biphenyls (PCBs), organochlorine pesticides (OCPs)), organic carbon (OC) and organic nitrogen (ON), determined in a previous study. Then, we applied an *in vitro* reporter gene assay (DR-CALUX) to detect and quantify the dioxin-like activity of PM-associated organic contaminants. Results showed that average Hg_{PM} concentration over the entire sampling period was found to be 13.4 ± 12 $\mu g m^{-3}$. Seasonal variation in the Hg_{PM} concentration was observed with lower values in spring and summer and higher values in winter and autumn due to the variation of meteorological conditions together with the emission sources. Principal component analysis suggested that fossil fuel combustion and a nearby cement factory were the dominant anthropogenic Hg_{PM} sources. Aryl hydrocarbon receptor (AhR)-mediated activities were observed in all organic extracts of atmospheric PM from Bizerte city (388.3 – 1543.6 $fg m^{-3}$), and shows significant positive correlations

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with all PM-associated organic contaminants. A significant proportion of dioxin-like activity of PM was related to PAHs. The dioxin-like activity followed the same trend as PM-associated organic contaminants, with higher dioxin-like activity in the cold season than in the warm season, indicating the advantage and utility of the use of bioassays in risk assessment of complex environmental samples.

1. Introduction

Atmospheric particulate matter (PM) is a highly complex mixture, whose toxicity depends on its composition, particle size, emission sources, geography, demography and weather conditions (Seinfeld and Pankow, 2003; Brown et al., 2005; Wenger et al., 2009). It can induce various pathologies including bronchitis, asthma, lung cancer, preterm birth, inflammatory and cardiovascular problems, and chronic diseases (Hoek et al., 2002; Bernstein et al., 2004; Pope and Dockery, 2006; Reche et al., 2012; Breysse et al., 2013; Mesquita et al., 2015). Among the PM constituents, pollutants such as polycyclic aromatic hydrocarbons (PAHs), organochlorine pesticides (OCPs), polychlorinated biphenyls (PCBs) and mercury (Hg) have attracted much interest due to their mutagenic, carcinogenic, teratogenic and bioaccumulative effects, as well as their long-range transport (Wania and Mackay, 1996; Schroeder and Munthe, 1998; Kim et al., 2013; AMAP/UN Environment, 2019).

Hg in the atmosphere originated from both natural and anthropogenic sources (Zhang et al., 2015; Outridge et al., 2018; AMAP/UN Environment, 2019), and exists under three physicochemical forms: gaseous elemental Hg (GEM), gaseous oxidized Hg (GOM) and particulate-bound Hg (Hg_{PM}) (Lindqvist and Rodhe, 1985). Although atmospheric Hg_{PM} accounts for only <5% of total atmospheric Hg, it plays an important role in the global Hg biogeochemical cycle due to its high solubility, deposition velocity and scavenging coefficients (Schroeder and Munthe, 1998; Fang et al., 2001; Driscoll et al., 2013). Incomplete combustion of organic matter, diffuse sources, and agricultural applications were identified as primary emission sources of PAHs, PCBs and OCPs in the atmosphere, respectively (Breivik et al., 2002; Dyke et al., 2003; Zhang and Tao, 2009; Yilmaz et al., 2016; Barhoumi et al., 2018). Depending on their physicochemical properties and emission sources, as well as meteorological conditions, PAHs, PCBs, OCPs and Hg_{PM} can exhibit strong temporal and regional distribution patterns in atmospheric PM.

Over past decades, both developed and developing countries have established new directives/guidelines to reduce PM exposures, and therefore protect human health (Krzyzanowski, 2008). Traditionally, air pollution monitoring is made using very expensive instrumental tools for analysis of some classes of contaminants such as PAHs, PCBs, OCPs, Hg, inorganic oxides, and ozone (Klanova et al., 2009; Novák et al., 2013; Huang et al., 2016). However, the determination of contaminant concentrations alone is often insufficient to derive accurate conclusions on the potential toxicity of such very complex matrix like atmosphere, as it is almost impossible to identify and quantify all potential toxicants in atmospheric PM. In addition, the levels of contamination can not take into consideration synergistic, antagonistic, and additive interactions of toxicants. Thus, it seems important to use bioassays for screening of toxicity because, unlike chemical analyses, they can integrate the effects of various chemicals in a complex mixture and their potential interactions (Novák et al., 2013). An essential aspect in determining the toxic effects of organic contaminants, such as PAHs, PCBs and OCPs, is their ability to bind and activate the aryl

hydrocarbon receptor (AhR), a key regulator of phase I and II metabolic enzymes (Nebert et al., 1993; Gonzalez and Fernández-Salguero, 1998). This biological activity is commonly known as AhR-mediated or dioxin-like activity. This activity can be monitored by the DR-CALUX (Dioxin Responsive Chemical-Activated Luciferase gene expression) assay, in which the AhR is challenged with extracts of PM samples to determine their toxic potential (Olivares et al., 2011; Khedidji et al., 2017). Due to its simplicity, efficiency, rapidity and ability to screen a large number of samples (Hodge et al., 2003; Kroese et al., 2015), several studies worldwide have used this bioassay to assess the toxic effect of ambient air (Clemons et al., 1998; Arrieta et al., 2003; Novák et al., 2009; Wenger et al., 2009).

In Africa, population growth, urbanization, traffic emissions, transported dust and open burning have strongly contributed to the current air pollution (Fayiga et al., 2018). It has been shown that on this continent, where air quality monitoring stations are almost non-existent and studies on atmospheric contaminants few, air pollution causes more premature deaths per year than either unsafe drinking water or malnutrition (Roy, 2016; Bahino et al., 2018). This highlights the need of carrying out atmospheric pollution studies over the African continent. A year-long sampling campaign was conducted in a North African city (Bizerte, Tunisia) which was affected by a considerable anthropogenic pressure in last decade due to the increase of urbanization, traffic density and economic activity, including agriculture, fishing and heavy industries. The chemical composition of atmospheric PM and its sources in Tunisia have been described previously (Castro-Jiménez et al., 2017; Castro-Jiménez and Sempéré, 2018; Barhoumi et al., 2018, 2020); however, studies on atmospheric Hg_{PM} and the toxicological effects of atmospheric PM are not available in Tunisia yet. Therefore, the objective of this study is to (1) investigate the concentration levels, seasonal variability and sources of Hg_{PM} in Bizerte city, and (2) to assess the dioxin-like potency of atmospheric PM using an *in vitro* reporter gene assay (DR-CALUX). The PM-associated organic contaminants (PAHs, alkylated PAHs, PCBs and OCPs) determined in a previous study were appealed here, in order to identify their role in the dioxin-like activity of atmospheric PM. To the best of our knowledge, this is the first long-term monitoring of Hg_{PM} in whole Africa.

2. Materials and methods

2.1. Site description and PM sampling

Fifty-four atmospheric PM samples were collected from March 2015 to January 2016 on the roof of the Faculty of Sciences of Bizerte (37° 16' 0.5802" N, 9° 52' 49.875" E), located directly adjacent to the Mediterranean Sea shoreline and the Bizerte lagoon and ~1 km from the Bizerte city center (Fig. S1 in the supplementary information (SI)). More detailed on sampling procedures including descriptions of the sampling site have been reported in Barhoumi et al. (2018). Briefly, Bizerte is a medium-sized city, with economic activity focused mainly on agriculture (7800 ha of cereal crops, 3400 ha of vegetables and 500 ha of tree crops), fishing (1253

fishing units, 3 fishing ports and 2 international shipping ports) and operations of light and heavy industries (cementery, plastic, textile, mechanic and electronic manufacturing, iron and steel metallurgy, petroleum refining and lubricants) (Barhoumi et al., 2018; Mhadhbi et al., 2019). In addition to these activities, the high urbanization and the high traffic density around the sampling site have induced a considerable anthropogenic pressure.

For sampling, a high volume air sampler equipped with a pre-combusted quartz fiber filters (QFFs, 20.3 x 25.4 cm) was used to collect the samples for 48 h at an average flow of $0.66 \text{ m}^3 \text{ min}^{-1}$. At regular intervals, field blanks were collected to provide an environmental background blank filter for biological and chemical assessment (Castro-Jiménez and Sempéré, 2018). Then, QFFs were wrapped in aluminum foil, transported to the laboratory and stored in a freezer at $-20 \text{ }^\circ\text{C}$ before processing. Sampling details and meteorological data are given in the SI (Table S1).

2.2. Sample preparation for chemical analysis and bioassay

Each QFF was conditioned in a desiccator ($25 \text{ }^\circ\text{C}$) for 24 h, weighted using a microbalance, and then cut into two equal parts. One part was used for PAHs, alkylated PAHs, PCBs and OCPs analysis and the other part was divided into subsamples for the determination of Hg_{PM} , organic carbon (OC), organic nitrogen (ON), total suspended particles (TSP), and preparation of the organic extracts for *in vitro* tests.

To obtain enough PM mass for organic extracts preparation for the *in vitro* tests, filter subsamples (3–7) of each month were pooled together. The organic extracts were obtained by sonication with a mixture of hexane and acetone (1:1 v/v; 3 x 20 mL; 20 min). The extracts were then concentrated using a Turbopap to 1 mL, filtered in a glass vial using a filter syringe containing a $0.45 \text{ }\mu\text{m}$ PTFE membrane, and further concentrated under a gentle nitrogen gas stream. At the point of dryness, dimethyl sulfoxide (DMSO) was added to obtain an equivalent concentration of 20 m^3 of sampled air per 1 mL of solution. A dilution series were prepared by 4-fold dilution (0.04, 0.16, 0.63, 2.5 and $10 \text{ m}^3 \text{ mL}^{-1}$) from the final extract to generate a dose–response curve. Samples were stored at $4 \text{ }^\circ\text{C}$ and tested using *in vitro* bioassays within a few days after sample preparation. Details on chemicals and reagents used in present study are presented in the SI (Text S1).

2.3. PM chemical characterization

Details on analytical procedures for organic compounds analysis are reported elsewhere (Barhoumi et al., 2018). The analyzed compounds which included 16 US Environmental Protection Agency (EPA) PAHs, 14 alkylated PAHs, 14 PCBs and 6 OCPs are summarized in Table S2. Particulate Hg was analyzed using standard cold vapor atomic absorption spectrometry (CV-AAS; LECO AMA 254) equipped with a low-level optical cell (Diez et al., 2017). A subsample of 25 mm diameter was punched out of the QFF filter, loaded into a nickel boat and placed inside the combustion tube. The subsample was thermally decomposed at $750 \text{ }^\circ\text{C}$ during 5 min and the evolved gases passed through catalytic tube removing possible interferants. The Hg(0) vapor was then retained by the amalgamator. The amalgamator was heated to $900 \text{ }^\circ\text{C}$ to release the accumulated Hg at once into the optical cell for detection. The concentrations of PM, OC and ON on QFFs were determined according to Barhoumi et al. (2018) and Raimbault et al. (2008).

2.4. *In vitro* DR-CALUX bioassay

The DR-CALUX (Dioxin Responsive Chemical-Activated LUCiferase gene eXpression) assay, consisting of rat hepatocarcinoma

H4IIE cells, stably transfected with a luciferase reporter gene under the control of dioxin responsive elements, was applied to detect AhR-mediated activity (dioxin-like activity) in PM extracts as described previously (Scippo et al., 2005; Van der Heiden et al., 2009; Van Langenhove et al., 2011). Briefly, cells were cultivated in MEM α medium (supplemented with 10% v/v fetal bovine serum, 50 IU/mL penicillin and 50 mg/mL streptomycin) at $37 \text{ }^\circ\text{C}/5\% \text{ CO}_2$. When the confluency was about 80–90%, $100 \text{ }\mu\text{L}$ of cells suspension were transferred into wells of the 96-well microplate and incubated overnight at $37 \text{ }^\circ\text{C}/5\% \text{ CO}_2$ to attach the cells in wells. Afterwards, the cells were exposed in triplicates to the dilution series of organic extract samples, solvent controls, field blanks, filter blanks and dilution series of calibration reference TCDD ($0.025\text{--}25 \text{ pg } \mu\text{L}^{-1}$), with a final DMSO concentration of 0.4% (v/v) in the exposure media. Five concentration levels of PM extracts were prepared by 4-fold dilution to provide a dose-response curve for each sample. The final volume per well was $200 \text{ }\mu\text{L}$. At least three independent assays were conducted for each PM sample. After 24 h exposure, the cell viability was confirmed under a microscope. Subsequently, the medium was removed and the cells were washed with non-sterile phosphate buffer and lysed using a solution containing Triton X100. Next, a glow-mix solution containing the substrates of luciferase, i.e. ATP and luciferin, was added and luciferase activity (luminescence production) was determined under a luminometer (ORION II, Berthold Detection System, Pforzheim, Germany) and reported as relative light units (RLUs). The RLUs were converted to relative responses, calculated as a percentage of the mean maximum responses induced by the reference TCDD. The mean solvent control response (DMSO) was subtracted from both the sample and the TCDD standard responses prior to conversion into percentages. The calculation of the sample extracts was based on Effective Concentration (EC_{50}), which was directly generated from the Hill regression of both the TCDD standards and sample extract dose–response curves (Khinkis et al., 2003). Therefore, AhR-mediated activity of each PM extract was reported as toxic equivalents (Bio-TEQ) expressed in femtograms of equivalent TCDD per cubic meter of air (fg m^{-3}) and assessed as the ratio of $\text{EC}_{50, \text{TCDD}}$ (fg/mL) over $\text{EC}_{50, \text{Sample}}$ (m^3/mL).

2.5. Quality assurance/quality control (QA/QC)

QA/QC for organic compounds analysis are detailed elsewhere (Barhoumi et al., 2018). The method detection limits (MDLs) of Hg, estimated as three times standard deviation of the blank samples, was 1.2 pg , equivalent to 0.01 pg m^{-3} , considering an average sampled air volume of 84 m^3 . The certified reference material (NRC MESS-4: $90 \pm 4 \text{ ng g}^{-1}$, 1σ) was run several times per analytical batch and constantly before starting the measurements, to check the accuracy of the measurements. The measured values were on average, within $\pm 5\%$ of the recommended values.

For *in vitro* bioassays, to eliminate the interference of solvent traces in DMSO solutions, and extraction methods on luminescence intensity, extraction solvents (hexane, acetone, dichloromethane, and mixture of hexane:acetone (1/1, v/v)) and extraction methods (ultrasound bath, ASE) were optimized before sample extractions. In addition, the performance and stability of the H4IIE cells response were checked before each bioassay (Text S3 and Figs. S2 and S3).

2.6. Data analysis

Results of the DR-CALUX assay were expressed as toxic equivalents (Bio-TEQ) (femtograms of equivalent TCDD per cubic meter of air (fg m^{-3})). The Bio-TEQs determined in this study were calculated as the means of three independent assays for pooled PM samples of each month. To estimate the degree of contribution of

analyzed PAHs to the AhR-mediated activities (dioxin-like activity) of the PM extracts, PAH-TEQs were calculated on the basis of PAH-potencies relative to TCDD (Bols et al., 1999), as the sum of the product of the concentration of each of the 7 carcinogenic PAHs (BaA, Chr, BbF, BkF, BaP, IcdP and DahA) multiplied by its respective bioassay-specific relative potency factor (REP) (Table S2) from Machala et al. (2001) for H4IIE-luc cell-based assays.

$$PAH - TEQs (pg / m^3) = \sum_{i=1}^7 \text{concentration PAHi} \times REP$$

In order to statistically compare datasets, the Mann-Whitney nonparametric *U* test was used (Stat-View 5.0). To identify whether there was a significant correlation between two sets of variables, a Pearson's correlation coefficient was used (graph prism 5 software). The Pearson's correlation coefficients were calculated based on the average of each variable (PAHs, alkylated PAHs, PCBs, OCPs, Hg_{PM}, Bio-TEQs, PAH-TEQs, OC, ON, TSP, AT, RH, WS, P) for 11 months. Box-and-whisker plots and Pearson principal component analysis (PCA) were performed using graph prism 5 and XLSTAT 2013.5.01, respectively.

3. Results and discussion

3.1. Characterization of aerosol samples according to chemical analyses

A detailed description of the distribution of organic contaminants (PAHs, PCBs, OCPs) in Bizerte atmospheric samples for the sampling period (March 2015–January 2016) is given in Barhoumi et al. (2018). Briefly, the analyzed compounds were detected in most of the samples (Table S2). Overall, all contaminants showed higher levels in winter. Average concentration of Σ_{16} PAHs (1.61 ng m⁻³) was 2, 447 and 1464 fold higher than that of Σ_{14} alkylated PAHs (0.95 ng m⁻³), Σ_{14} PCBs (3.60 pg m⁻³) and Σ_6 OCPs (1.10 pg m⁻³), respectively. Source analysis showed that PAHs/alkylated PAHs originated from vehicular traffic emissions (diesel combustion), while cement, electronic, iron and lubricant factories were identified as dominant sources for PCBs. The dominant OCP congeners were *p,p'*-DDT and *p,p'*-DDE, reflecting a current or past use in agriculture. Wind sector and back-trajectories of air mass showed predominance of local sources relative to long-range atmospheric transport for these pollutants.

Fig. 1A displays the time series of daily atmospheric Hg_{PM} and TSP concentrations, and the Hg_{PM} mass content in TSP (Hg_{PM}/TSP) in Bizerte city during the study period. The Hg_{PM} concentrations varied from 2.5 to 63.9 pg m⁻³ with an annual average of 13.4 ± 12 pg m⁻³ and a median of 9.8 pg m⁻³. The median value, which was smaller than the mean value, indicated the occurrence of episodes of higher levels of Hg_{PM} during the study period. To our knowledge, this is the first report of temporal distribution of Hg_{PM} across Africa (AMAP/UN Environment, 2019; Zhang et al., 2019a, 2019b). Although attention should be paid to the comparisons due to the uncertainties caused by the characteristics of sampling sites (source strength, size of the particles collected, sampling period, different methods of extraction used ...), our values were compared with those reported in other locations worldwide (Table 1) to fully understand the pollution levels of Hg_{PM} in Bizerte city. As shown in Table 1, Hg_{PM} concentrations in Bizerte city are comparable to those measured in the cities of Taichung (China), Elora (Canada), Göteborg (Sweden), Waldhof (Germany), and Marais Vernier (France), but far lower than those recorded in urban megacities, such as Mumbai (India), Beijing (China), São Paulo State (Brazil) and Mexico City. At those sites, the mean Hg_{PM} concentrations are

approx. 12 to 70-fold higher than those we measured in Bizerte (Table 1). Compared to Hg_{PM} levels observed over open ocean and coastal areas such as the southern Indian Ocean, north Atlantic Ocean, Adriatic Sea, Thompson Farm (USA) and Cape Hedo (Japan), the Hg_{PM} levels in Bizerte city are about 12-fold higher (Table 1), suggesting that the atmosphere of Bizerte was contaminated with Hg_{PM} during the study period. Globally, our values are in the same order of magnitude as Hg_{PM} concentrations in the Southern Hemisphere (53.5 ± 47.4 pg m⁻³), but much lower than in the Northern Hemisphere (113.8 ± 102.1 pg m⁻³) (Zhang et al., 2019a, 2019b).

Moreover, as shown in Fig. 1A, the Hg_{PM} to TSP mass ratio (Hg_{PM}/TSP) ranged from 53.7 to 1179 ng g⁻¹ with an average of 233.3 ± 187.9 ng g⁻¹, which was similar to Okinawa Island, Japan (200 ng g⁻¹) (Chand et al., 2008), higher than Chuncheon, Korea (60 ng g⁻¹) (Kim et al., 2012), but lower than Shanghai, China (1050–2850 ng g⁻¹) (Xiu et al., 2005) and Nepal, South Asia (2586.0 ng g⁻¹) (Guo et al., 2017). The positive correlation between Hg_{PM} and TSP (*r* = 0.256; *p* = 0.06) and the strong significant positive correlation between Hg_{PM} and Hg_{PM}/TSP (*r* = 0.860; *p* < 0.0001) may indicate that TSP and Hg_{PM} had either same or similar sources or gaseous Hg (e.g., GOM) in the same air mass got adsorbed on to existing atmospheric particles before they arrived to the sampling site. It is important to keep in mind that only Hg_{PM} was investigated within this study but GEM and GOM were not considered. Therefore, more studies are required to verify this assumption in the future. The previous studies in other locations have reported that the Hg_{PM} was mainly associated with atmospheric particles such as dust or was likely produced by adsorption of Hg²⁺ species onto atmospheric particulates (Schroeder and Munthe, 1998).

Temporal variations of monthly/seasonality average Hg_{PM} concentrations in the atmosphere of Bizerte during the study period are illustrated in Fig. 1. The seasonal mean Hg_{PM} concentrations were higher in winter (27.2 ± 25.2 pg m⁻³) and autumn (13.8 ± 10.1 pg m⁻³) than in spring (11.6 ± 5.4 pg m⁻³) and summer (9.8 ± 6.3 pg m⁻³), with the highest monthly mean value of 40.6 ± 26.4 pg m⁻³ in December and the lowest of 7.2 ± 3.7 pg m⁻³ in November (Fig. 1B and C). Statistical analysis (*U* tests) showed that Hg_{PM} concentrations were significantly higher (*p* < 0.05) in winter than in summer (Fig. 1C). A similar seasonal behavior has also been observed in other studies, including Seoul (Korea), Beijing (China), Michigan (US) and Ostrava (Czech Republic) (Lynam and Keeler, 2005; Kim et al., 2009; Topinka et al., 2015; Tang et al., 2019), which was in consistence with the severe particulate pollution during cold season as well as the unfavorable meteorological conditions. It has been reported that anthropogenic emissions from industrial facilities such as coal-fired power plants, cement factories, as well as coal burning for heating purpose (especially in winter) and liquid fossil fuel combustion (petroleum refinery, fuel including gasoline and diesel) were the main contributors of Hg into atmosphere (Wang et al., 2000; Jiang et al., 2006; Xiu et al., 2009; Zhang et al., 2015). In view of the fact that no industry using coal as fuel source is present in the study area and that the Bizerte cement factory, the petroleum refinery and the highway are located close to our sampling site (<1 km), the high levels of atmospheric Hg_{PM} during the winter season could be greatly related to emissions from these sources. Besides emissions by local anthropogenic sources, the changes of the meteorological conditions (e.g., precipitation (P), ambient temperature (AT), wind speed (WS) and direction, relative humidity (RH)), chemical transformations through interactions with other chemical species (O₃, SO₂, CO, radicals), emissions from natural sources, and long-range transportation might also be responsible for the seasonal variation of atmospheric Hg (WIOŚ, 2013; Huang et al., 2016;

Table 1Comparison of Hg_{PM} concentrations (in pg m⁻³) at Bizerte city with those reported from other locations worldwide.

Locations	Sampling time	Site type	Size	Concentrations	References
Bizerte city, Tunisia	2015/03–2016/01	Mixed area	TSP	13.40 (2.51–63.95)	<i>This study</i>
Taichung city, Taiwan, China	2010/03–2011/02	Suburban	TSP	71.1 ± 46.1	Huang et al. (2012)
Mt. Gongga, China	2005/05–2006/04	Remote	TSP	30.7 (5.2–135.7)	Fu et al. (2008)
Taiwan, China	2015/01–2015/04	Mixed area	PM _{2.5}	30.5	Fang et al. (2017)
Baltic sea, Poland	2007/12–2008/12	Coastal	TSP	20 ± 18	Beldowska et al. (2012)
Elora, Ontario, Canada	2006/11–2007/08	Rural	TSP	16.40 ± 9.54	Baya and Van Heyst (2010)
Göteborg, Sweden	2005/02	Urban	PM _{2.5}	12.5 (3.59–20.26)	Li et al. (2008)
Marais Vernier, France	2010/10–2012/03	Rural	TSP	10 ± 10	Connan et al. (2013)
Rochester, NY, USA	2007/12–2009/11	Suburban	PM _{2.5}	8.7 (<DL–271.2)	Choi et al. (2013)
Waldhof, Germany	2009/01–2011/12	Rural	PM _{2.5}	6.3 (<0.4–262)	Weigelt et al. (2013)
Seoul, Korea	2009/12–2010/07	Urban	TSP	6.8 ± 6.5	Kim et al. (2012)
Holland	2007/07–2007/11	Urban	PM _{2.5}	6 ± 6	Gratz et al. (2013)
Southern Indian Ocean	2012/01–2012/04	Ocean	TSP	0.34	Barret et al. (2013)
Thompson Farm, USA	2009/02–2010/08	Coastal	PM _{2.5}	0.19–1.14	Mao et al. (2012)
North Atlantic Ocean	2004/06–2004/08	Ocean	TSP	2.4 ± 1.1	Aspmo et al. (2006)
Nova Scotia, Canada	2009/01–2010/12	Urban	PM _{2.5}	2.3 ± 3.1	Cheng et al. (2013)
Cape Hedo, Japan	2004/03–2004/05	Remote	PM _{2.5}	3.0 ± 2.5	Chand et al. (2008)
Adriatic Sea	2004/10–2004/11	Ocean	TSP	4.5 ± 8	Sprovieri and Pirrone (2008)
Mumbai, India	1991/02–1992/04	Mixed area	TSP	1100–2400	Kumar et al. (2001)
Beijing, China	2003/01–2004/09	Urban	TSP	1180 (180–3510)	Wang et al. (2006)
Santander, Northern Spain	2008	Urban	PM ₁₀	1400 (<DL–25400)	Arruti et al. (2010)
Lichwin, Poland	2004/01–2004/02	Suburban	TSP	1050 ± 180	Zielonka et al. (2005)
Kathmandu, Nepal	2013/04–2014/04	Suburban	TSP	850.5 ± 926.8	Guo et al. (2017)
Changchun city, China	1999/07–2000/01	Urban	TSP	22–1984	Fang et al. (2001)
São Paulo State, Brazil	2002–2003	Urban	TSP	400 ± 300	Fostier and Michelazzo (2006)
Lhasa, China	2013/04–2014/08	Urban	TSP	224 (61.2–831)	Huang et al. (2016)
Mexico City	2006/03	Urban	PM _{2.5}	187 ± 300	Rutter et al. (2009)

ND: no data, <DL: below detection limit, TSP: total suspended particles, PM: particulate matter.

Kumari and Kulshrestha, 2018).

Meteorological datas and other pollutants (OC and ON) are applied to further understand the influencing factors on Hg_{PM} concentrations over Bizerte city (Table S1 and Table S3). As shown in Table S1, AT was higher in summer/autumn than in spring/winter with a significant difference between all seasons, while RH and WS showed little variation among the seasons. Compared to previous years, the period during which the study took place (March 2015–January 2016) was characterized by an exceptionally dry year, with very little rainfall (annual mean of 0.3 mm). Gas phase-particulate phase partitioning of atmospheric Hg is highly dependent on AT, with high Hg adsorption onto the particles at low temperature and vice versa (Pankow, 1987). The Pearson correlation coefficient between Hg_{PM} concentrations and AT (Table S3) showed that Hg_{PM} was negatively correlated with AT based on the whole yearly observation in Bizerte city, indicating that the adsorption of Hg_{PM} onto particles surface might be the key process for the formation of Hg_{PM} (Xiu et al., 2009). This is in accordance with the above results, showing that the lowest and the highest Hg_{PM} concentrations were observed in summer (with highest temperature) and winter (with lowest temperature), respectively (Fig. 1C, Table S1). RH can influence Hg concentrations in particles through its impact on Hg gas-particle partitioning. The high RH can increase the aerosol liquid water content, which promotes more Hg partitioning into the condensed phase and adsorption onto particles (Zhang et al., 2019a, 2019b). In this study, although RH was slightly higher in winter than in spring/summer, the influence of this parameter on seasonal variation was probably minor since Bizerte city is generally characterized by a high RH over the whole year (mean of 69.4 ± 7.6%) (Table S1). As shown in Fig. S4, mean Hg_{PM} concentration on the days with precipitation was lower by 50% relative to that on the days without precipitation. Therefore, precipitation was one of the main meteorological factors affecting Hg_{PM} concentration in the atmosphere of Bizerte city. Similarly, previous studies (Kim et al., 2012; Guo et al., 2017) have also documented that Hg_{PM} concentrations decreased dramatically after

heavy precipitation events. The latter can remove the Hg_{PM} in the atmosphere *via* wet scavenging processes (Schleicher et al., 2015). WS plays a key role in the dispersion of air pollution, and generally calm wind could restrict the diffusion and dispersion of atmospheric pollutants (Guo et al., 2017). As shown in Fig. 1 and Table S1, the Hg_{PM} concentrations under calm wind condition (winter season) were remarkably higher than those under high speed wind which promoted air masses mixing. The calm wind speed combined with lower temperatures might enhance adsorption of GOM onto particles (Kim et al., 2012; Lee et al., 2016) and therefore, increase the Hg_{PM} levels in winter season (particularly in December). This result was also confirmed by the significant negative correlation between WS and Hg_{PM} (Table S3). In addition, OC, which is often used as tracer of both primary aerosols formed from combustion processes and secondary aerosols generated from atmospheric oxidation and gas-to-particle conversion (Han et al., 2018), was positively correlated with Hg_{PM} (Table S3), indicating that combustion processes and gas-to-particle conversion are important mechanisms of Hg_{PM} formation in Bizerte city. Lin and Pehkonen (1999) found that in the presence of OC, iron oxides can catalyze GEM oxidization to Hg_{PM}. Despite their clear influences on Hg_{PM} seasonal variation, some meteorological parameters listed above (AT and WS) have shown a minor or not significant correlations with Hg_{PM}. The lack of significant correlations may due, in part, to the anthropogenic source emissions, which very likely contributed more to Hg_{PM} levels in the ambient air than the meteorological parameters did (Li et al., 2017).

As mentioned by several authors (Huang et al., 2016; Siudek et al., 2016), Hg_{PM} concentration in atmospheric PM may be influenced not only by anthropogenic inputs and meteorological parameters, but also by wind direction and long-range transport. In a previous study, Barhoumi et al. (2018) found through the wind sector analysis that in the same area and same sampling period, the most predominant wind directions were from the west-north-west (WNW) in spring, the north-west (NW) in summer and autumn, and the south-east (SE) in winter, suggesting that local emissions

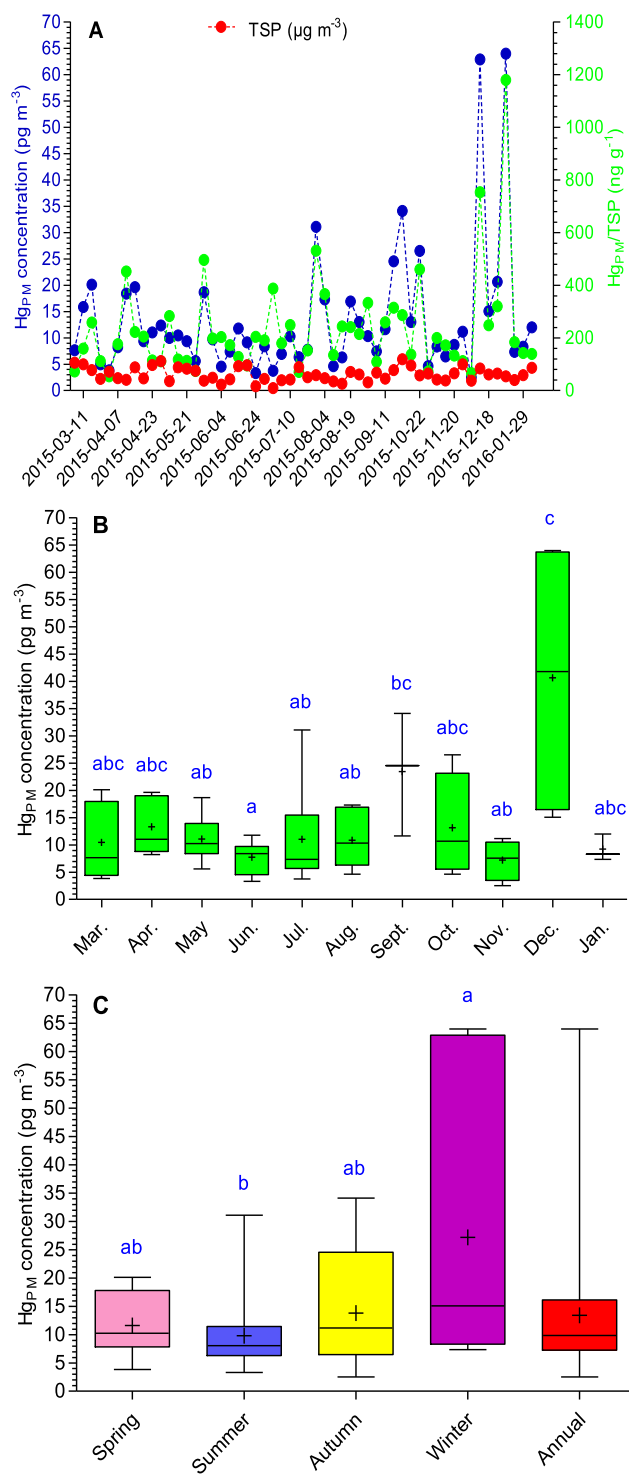


Fig. 1. Time series plots of (A) daily, (B) monthly and (C) seasonally atmospheric Hg_{PM} concentrations in Bizerte city during the study period (March 2015 to January 2016). Boxes represent the interquartile range, bound by the 75th and 25th percentile. Line markers, cross markers and error bars indicate the medians, averages and minimum/maximum Hg_{PM} concentrations, respectively. Lowercase letters on the top correspond to statistically different groups of data, calculated by Mann-Whitney nonparametric tests (U tests). Groups sharing the same letter are not statistically different at $p < 0.05$.

from downtown Bizerte (vehicular exhaust), Bizerte harbor and the cement factory, located NW of the site, and from Zarzouna area (petroleum refinery, the Tunisian Company of Lubricants (SOTULUB)) and Menzel Jemil city (electronic and plastic industries), located SE of the site, are the most likely to influence the pollutant levels observed in atmospheric PM. The highest concentrations occurred in winter when winds were from the SE, indicating the presence of a local source in this direction. By using the NOAA HYSPLIT Backward Trajectory Model, Barhoumi et al. (2018) also found that the influence of long-range atmospheric transport was minor, since the air masses reaching the sampling site were of oceanic origin with limited passage through the European continent. Indeed, the clean nature of the oceanic air masses and high wind speeds above the Mediterranean Sea can disperse pollutants before they reach the sampling site. The negative correlations observed here between WS and Hg_{PM} concentrations (Table S3) can support the previous observations and confirmed the predominance of Hg_{PM} local sources, as strong winds flush pollutants out of the study area whereas weak winds allow pollutants to accumulate over time.

To identify the potential sources of Hg_{PM} and infer the possible influence factors on their concentrations, principal component analysis (PCA) was applied (Fig. 2). Two factors with eigenvalues > 1 accounted for 56.4% of the dataset total variance. Factor 1, which explained 41.3% of the variance, showed high loading of \sum_{16} PAHs (0.91), \sum_{14} alkylated PAHs (0.77), \sum_6 OCPs (0.79), Hg_{PM} (0.70), OC (0.81) and ON (0.82), and most samples collected in winter season (0.72–0.90). This may suggest on one hand that Hg_{PM} , \sum_{16} PAHs, \sum_{14} alkylated PAHs and \sum_6 OCPs were dominant in winter and were associated with OC and ON, and on the other hand that Hg_{PM} had similar sources as \sum_{16} PAHs, \sum_{14} alkylated PAHs, \sum_6 OCPs, OC and ON. This result is in agreement with that of the Pearson's correlation coefficients (Table S3) and aforementioned results, which indicates that liquid fossil fuel combustion mixed with biomass burning are important sources of Hg_{PM} in Bizerte city. These observations are in good agreement with data obtained by Yu et al. (2019) in Shanghai and Beijing, China. In view of urban, residential, industrial and agricultural activities around the sampling site, Hg_{PM} was likely to have originated from the extensive vehicular emissions of heavy and light cars from the highway next the sampling station (400 m), the petroleum refinery, and from the

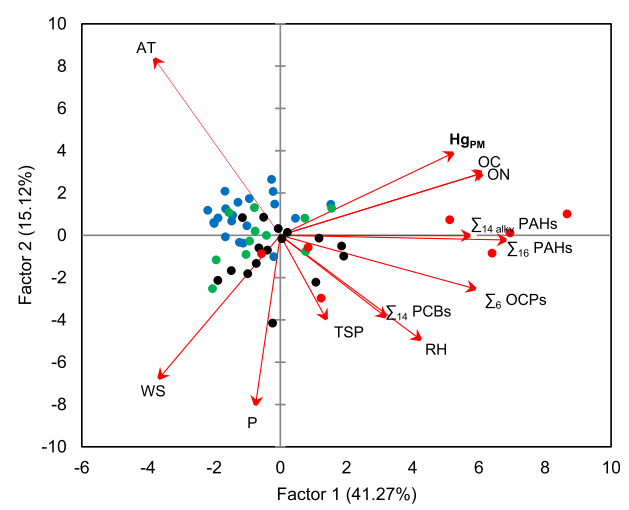


Fig. 2. PCA loading (arrows and Black text) and score plots (colored points) of TSP, OC, ON, chemicals and meteorological variables.

open burning of biomass for crop/forest residue disposal and land preparation (Barhoumi et al., 2018). Although the combustion of gasoline and diesel had been regarded as important sources of Hg in elemental form (Hg^0), it can also significantly contribute to the formation of Hg_{PM} (Conaway et al., 2005; Landis et al., 2007; Won et al., 2007). Indeed, a lot of reactive compounds from exhaust emission can foster strong oxidative environment (higher OH radicals and O_3) near traffic road to spur Hg oxidation resulting in the formation of Hg_{PM} (Xiu et al., 2009). Similarly, Xiu et al. (2009) measured significantly higher Hg concentrations in traffic originated particulate matter at the measurement site near a road ($20 \cdot 10^3$ vehicles hr^{-1} in rush hour) in Shanghai, China, suggesting a large contribution from tailpipe exhaust, wear dust from tires or brake linings, and the resuspension of road dust. The correlation between Hg_{PM} and \sum_6 OCPs may also indicated that Hg_{PM} has agricultural sources, through the intense use of fertilizers. Factor 2 of PCA, which explained 15.1% of the total variance, had significant positive loading of AT (0.68), and significant negative loadings of WS (-0.55) and precipitation (-0.65), suggesting that AT, WS and precipitation were probably the main meteorological parameters affecting Hg_{PM} variations over the year. Therefore, this factor was thought to describe the climatological change in Bizerte city.

To summarize, liquid fossil fuel combustion, including petroleum refinery and traffic emissions, and the Bizerte cement factory were considered as the dominant local sources of Hg_{PM} in Bizerte city, with a contribution from biomass burning and agricultural activities. Moreover, the local residential sources, the mechanic, electronic and iron factories (located SW of the site), the incineration of municipal waste, and the natural sources (evasion from surficial soils, water bodies, vegetation surfaces, forest fires) should not be neglected.

3.2. Aryl hydrocarbon receptor mediated activity of atmospheric PM extracts

It has been reported that atmospheric particle-bound PAHs, PCBs and OCPs can induce dioxin-like activity through their ability to bind and activate the AhR (Nebert et al., 1993; Olivares et al., 2011; Khedidji et al., 2017). In this regard, AhR-mediated activities of atmospheric PM collected from Bizerte city were analyzed using a DR-CALUX assay, and were assessed in relation to their PAHs, alkylated PAHs, PCBs and OCPs contents determined in a previous study (Barhoumi et al., 2018). The DR-CALUX assay is widely used for screening of AhR-mediated (dioxin-like) activity of pure substances, as well as environmental samples including atmospheric particles from urban, industrial, rural and semi-rural environments (Nebert et al., 1993; Olivares et al., 2011; Mesquita et al., 2014, 2015; Khedidji et al., 2017). There have been several studies worldwide describing AhR-mediated activities of PM1 (Wenger et al., 2009), PM10 (Clemons et al., 1998; Khedidji et al., 2017) or total particulate matter (Hamers et al., 2000; Klein et al., 2006). However, such data remain very limited for African countries (Khedidji et al., 2017). In this study, AhR-mediated activities, expressed as Bio-TEQs, were observed in all PM extract samples (Fig. 3A), suggesting the existence of AhR ligands such as dioxins and dioxin-like compounds in the atmospheric PM of Bizerte city. Bio-TEQs were between 388.3 and 1543.6 fg m^{-3} , with a mean value of 637.0 fg m^{-3} . Compared with other studies, Bio-TEQs in atmospheric PM of Bizerte city were comparable to those measured in Czech Republic (Novák et al., 2009), but significantly lower than those measured in Switzerland (Wenger et al., 2009) and Barcelona, Spain (Mesquita et al., 2014, 2017). Although they were not statistically significant, both monthly/seasonally average Bio-TEQs were slightly higher in the cold season than in the warm season (probably due to the low degradation of pollutants and more pronounced combustion

sources in cold season) (Fig. 3A and Fig. S5). This trend coincided with high levels of \sum_{16} PAHs, \sum_{14} alkylated PAHs, \sum_{14} PCBs and \sum_6 OCPs in the cold season (Fig. 3B), suggesting a key role of these compounds in the dioxin-like activity of PM organic extracts. This has been confirmed by the significant positive correlations between the dioxin-like activity values and the concentrations of \sum_{16} PAHs, \sum_{14} alkylated PAHs, \sum_{14} PCBs and \sum_6 OCPs, as well as the individual concentrations of different congeners of PAHs, alkylated PAHs, PCBs and OCPs (Table S3). Using Bio-TEQs and PAH-TEQs, results showed that the 16 EPA priority PAHs were responsible for 40% (annual mean) of total AhR-mediated activity of PM (Fig. 3A). This suggests that analyzed PAHs represented a substantial fraction of the total dioxin-like activity present in the samples, which is confirmed by a strong positive correlation between Bio-TEQs and PAH-TEQs (Table S3). This contribution are similar to those measured by Olivares et al. (2011), but higher than those determined in other studies (Brown et al., 2005; Cavanagh et al., 2009; Wenger et al., 2009; Novák et al., 2014). The emissions related to traffic activities (fossil fuel combustion) are probably the main causes of dioxin-like activity of PM extracts observed here. The PAH isomeric ratios calculated by Barhoumi et al. (2018) using the same PM samples were consistent with a dominance of traffic emission sources. As indicated above, total amount of PAHs cannot fully account for the total AhR-mediated activity observed in

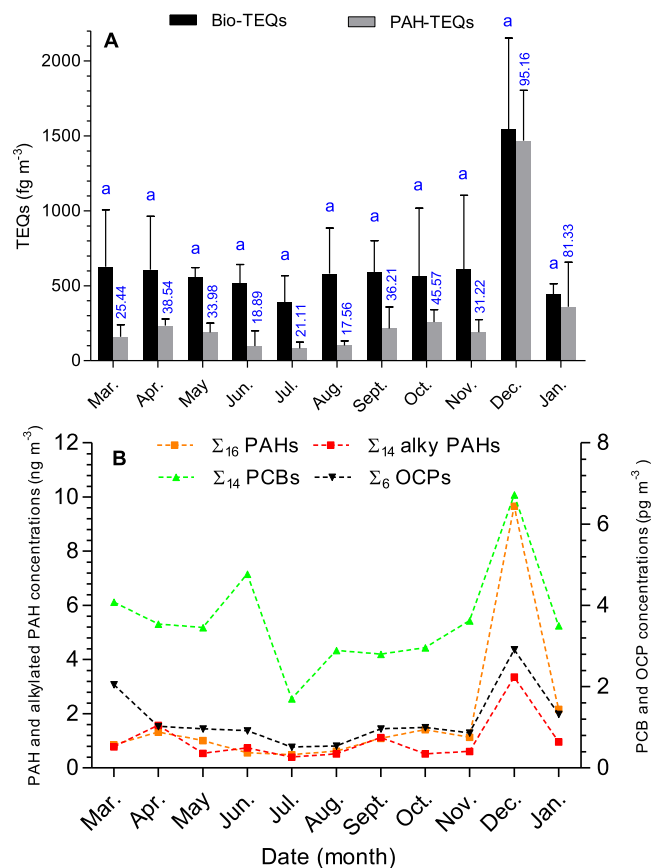


Fig. 3. Aryl hydrocarbon receptor-mediated activity (Bio-TEQs), PAH-TEQs and the calculated percentage of contribution of these PAHs to the assessed Bio-TEQs (A), and concentrations of \sum_{16} PAHs, \sum_{14} alkylated PAHs, \sum_{14} PCBs and \sum_6 OCPs (B) in each PM extract collected from Bizerte city in the study period (March 2015 to January 2016). Lowercase letters on the top correspond to statistically different groups of data, calculated by Mann-Whitney nonparametric tests (U tests) (No significant differences were observed between Bio-TEQs for the different months). PAH, alkylated PAH, PCB and OCP concentrations are taken from Barhoumi et al. (2018).

the present study. The unexplained remaining 60% dioxin-like activity could be attributed on one hand to the \sum_{14} alkylated PAHs, \sum_{14} PCBs and \sum_6 OCPs, which show significant correlations with Bio-TEQs, and on the other hand to other priority and non-priority pollutants not analyzed in this study that showed to possess dioxin-like potency, such as polycyclic aromatic ketones and quinones (Misaki et al., 2007), polychlorinated naphthalenes (PCNs) (Villeneuve et al., 2001), benzofuran and 2,3-dimethylbenzofuran (Hinger et al., 2011; Misaki et al., 2007; Vondracek et al., 2007). The interaction between contaminants (e.g. antagonism, synergism and additive effects) may also have an effect on dioxin-like potency.

AhR-mediated activities of PM extracts were observed in previous works (Škarek et al., 2007; Novák et al., 2009; Olivares et al., 2011), where organic contaminants concentrations including PAHs, were 2–6 fold higher than those measured in this study, which indicates that DR-CALUX assay is able to detect dioxin-like activity even with low concentrations. Therefore, this bioassay can be considered as an early-warning tool for routine screening and control of particle-bound contaminants emissions.

4. Conclusions

This study investigated chemical composition and biological activity of atmospheric PM collected from March 2015 to January 2016 in Bizerte city, a mixed urban, agricultural, and industrialized site bordering the Southwestern Mediterranean Sea (North Tunisia). Physicochemical characterization revealed that atmospheric PM is a complex mixture with inorganic (Hg_{PM}) and organic compounds (PAHs, alkylated PAHs, PCBs, OCPs, OC and ON) deriving from anthropogenic as well as natural sources. Contributions of several sources such as traffic emissions, heavy industries (Bizerte cement, iron factory and petroleum refinery), biomass burning, mechanic and electronic factories, and agricultural activities have been evidenced in the PM composition. All of the PM components analyzed exhibited higher concentrations during the cold season than in the warm season, which could be largely attributed to the local anthropogenic sources and meteorological conditions (AT, WS and precipitation). The DR-CALUX bioassay allowed showing that the organic extracts of atmospheric PM from Bizerte city were biologically active and could elicit AhR-mediated activity. The biological response followed the same temporal variation as PM-associated organic contaminants, with higher dioxin-like activity in the cold season, highlighting the relevance of this *in vitro* bioassay for the evaluation of toxic potential of environmental samples. To our knowledge, this is the first study to provide atmospheric Hg_{PM} data and to examine the dioxin-like activity of atmospheric PM in North Africa. Further investigations using fine and ultrafine PM (PM_{2.5} and PM₁₀) are needed, since they are more hazardous than the TSP. More attention should be also given to the gaseous phase due to its high content of atmospheric pollutants.

Credit author statement

Barhouni, B., Conceptualization, Writing-Original draft preparation; Tedetti, M., Supervision, Writing-Reviewing and Editing; Heimbürger-Boavida L.E., Methodology, Writing Reviewing and Editing; Dufour, A., Tesan Onrubia, J.A., Doan, Q.T., Boutaleb, S., Initial data analysis, SoftwareMethodology; Touil, S., Supervision; Scippo, M.L., Supervision, Funding acquisition, Writing-Reviewing and Editing.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have

appeared to influence the work reported in this paper.

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