

Pioneer trees of Betula pendula at a red gypsum landfill harbour specific structure and composition of root-associated microbial communities

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1	Pioneer tress of <i>Betula pendula</i> at a red gypsum landfill harbour specific							
2	structure and composition of root-associated microbial communities							
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11								
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13	Betula pendula, Populus tremula, Populus nigra, Salix purpurea, native trees,Illumina							
14	MiSeq,							
15								
16	Abstract							
17	The study of root-associated microbial communities is important to understand the natural							
18	processes involved in plant colonisation at degraded areas. Root associated bacterial and							
19	fungal communities of woody species colonising a red gypsum landfill (a metal-enriched							

- 20 environment) were characterised through metabarcoding. Among trees naturally growing
- 21 on the landfill, *Betula pendula* is the only tree species in the centre of the area, whereas

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companion tree species such as Populus nigra, P. tremula and Salix purpurea were 22 23 present on the edges. The bacterial community was dominated by *Proteobacteria* (38%), Actinobacteria (35%) and Bacteroidetes (20%) and the most abundant bacterial OTU 24 belonged to the family Streptomycetaceae. The fungal community was dominated by 25 Ascomycota (60%) and Basidiomycota (30%) and the most abundant family was 26 Pyronemataceae. Analysis of similarities, heatmap and hierarchical cluster analysis 27 showed that B. pendula grown in the centre of the landfill harboured a specific microbial 28 29 community, which was unique and different, not only from other tree species (Populus or Salix spp.), but also from other *B. pendula* growing at the edges. Our findings on relevant 30 indicator OTUs associated to the birches located in the centre of the landfill (such as 31 32 Otu00716 Catellatospora sp (family Micromonosporaceae, phylum Actinobacteria) or Otu4 35502 Russula sp. (family Russulaceae, phylum Basidiomycota)) may have 33 34 important implications for the successful revegetation of these harsh environments using 35 microbial-based phytostabilisation approaches.

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38 **1. Introduction**

Gentle Remediation Options (GRO) focus on the use of plants in combination with 39 associated microorganisms for the remediation and revitalisation of contaminated sites 40 (Kidd et al., 2015). Metal(loid) toxicity can severely limit the performance and 41 establishment of plants in polluted areas and therefore, the selection of metal-tolerant 42 plant species able to successfully establish and grow at each specific site is of vital 43 44 importance for the success of GRO. Sites contaminated with potentially toxic elements (PTE); such as mine tailings, spoils or landfills (made-up of by-products of metal 45 46 extraction) are important sources of metal-tolerant plant genotypes. Red gypsum landfills 47 are made of by-products of titanium refining; mainly used for the TiO₂ pigment industry 48 and enriched in S, Fe and Mn (Zappelini et al., 2018, Zapata-Carbonell et al., 2019). Despite the unfavourable growth conditions at these sites, several metallophyte plants 49 have evolved mechanisms to tolerate these conditions and are able to colonise these types 50 of substrates (Whiting et al., 2004; Batty, 2005; Mendez and Maier, 2008). Fast-growing 51 52 trees are ideal plants for the revegetation of these sites due to their extensive root systems,

rapid growth, large biomass production and easy harvesting with subsequent re-sprouting
(Peuke and Rennenberg, 2005).

In the last years, an increasing number of studies including the use of native 55 pioneer trees for the revegetation of degraded lands is found in the (Fernandez et al., 56 2017; Mleczek et al., 2017; Asension et al., 2018) A greater understanding of the 57 mechanisms, which influence the response of plants to the local conditions frequently 58 59 found at these sites, could help to improve the efficiency of GRO techniques. Although 60 an increasing number of studies have addressed the influence of the associated microbial community in plant metal tolerance, its role in harsh environments remains unravelled. 61 Studies approaching the description of plant-associated microbial communities are useful 62 as a first approximation to understand these interactions. Plants interact closely with 63 microbes, and these can enhance plant growth and health by increasing nutrient uptake 64 65 and improving plant resistance to pathogens and stress (Weyens et al., 2009; Kidd et al., 2017). Plant growth-promoting microorganismsinclude nitrogen-fixing bacteria, 66 67 biocontrol microorganisms or mycorrhizal fungi, among others (Coninx et al., 2017; Kidd et al., 2017). Plant associated microorganisms can also improve nutrient availability 68 69 through the production of organic acids and siderophores or phosphate solubilisation. 70 Bacteria can also produce phytohormones, such as auxins, cytokinins and gibberellins, which stimulate plant growth and development (Tanimoto, 2005; Taghavi et al., 2009). 71 72 Plant associated bacteria can also reduce plant stress through the synthesis of 1aminocyclopropane-1-carboxylic acid (ACC) deaminase which reduces the level of 73 ethylene in stressed plants (Glick, 2014). Other well-known mechanisms of PGPBs 74 include increasing plant water uptake, alteration of root morphology, production of 75 76 antibiotics and the induction of plant defence mechanisms (van Loon and Bakker, 2003; 77 Kidd et al., 2009).

78 Ectomycorrhizal fungal (ECM) are the dominant group of root symbionts of pioneer trees that attempt to colonise heavily contaminated sites (Colpaert, 2008). Fungal 79 80 endophytes also play a key role on primary colonisation and subsequently on phytoremediation. It has been observed on contaminated sites that pioneer populations of 81 82 mycorrhizal birches, willows, poplars or pines are able to successfully establish (Vrålstad 83 et al., 2002; Adriaensen et al., 2004; Colpaert et al., 2004). Root-associated arbuscular mycorrhizal fungi (AMF) usually improve the growth of plants in non-fertile soils by 84 85 increasing the exchange surface between plant and soil (Smith and Read, 1996). AMF 86 have been mostly studied for their role in plant growth, mainly due to their role in plant

P uptake from soils (Smith and Read, 2002). Moreover, AMF are able to alleviate metal 87 toxicity in plants (Hildebrandt et al., 2007). Other key soil microorganisms such as 88 saprotrophic fungi are central in soils due to their capacity to degrade recalcitrant organic 89 compounds and may strongly interact with soil bacteria (Baldrian, 2008). Saprotrophic 90 fungi may also have a potential in mobilising metals from contaminated soils, due to their 91 ability to exudate chelating agents (Arwidsson et al., 2009). 92

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Further interactions between bacteria, fungi and plants have been shown to 94 contribute to microbial community stability (Bonfante and Anca, 2009; Bell et al., 2014) 95 and many bacterial strains have indeed been reported to promote mycorrhizal symbioses (Frey-Klett et al., 2007; Hrynkiewicz et al., 2010). 96

97 Here, we characterised the root microbial communities associated to Betula pendula, which is the pioneer colonising vegetation at a red gypsum landfill in the region 98 99 of Thann in NE France (Zapata-Carbonell et al., 2019). At this site, B. pendula is the only 100 woody species that had successfully established across the whole landfill while other 101 woody species showed only sparse growth at some specific spots (Zapata-Carbonell et al., 2019). Moreover, in previous studies B. pendula trees growing at this site have been 102 103 shown to efficiently accumulate Mn and therefore this species is a promising candidate 104 to be used as biocatalyst. The present study hypothesised that B. pendula hosts a specific root-associated microbial community responsible for its high adaptability to the soil 105 conditions at this site. Therefore, the objective of this research was to analyse the structure 106 107 of these communities using high-throughput sequencing technologies (Illumina next-108 generation sequencing).

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110 2. Material and methods

111 2.1. Study site

The study site is located in the southern part of the Alsace region (47°47'47.7"N 112 7°08'18.5"E) near the city of Thann (NE of France). Here, Cristal France produces TiO₂ 113 (an important pigment) and TiCl₄ (an excellent Lewis acid catalyst). The Thann plant 114 produces TiO₂ through a sulphate-based technology and TiCl₄ through a chlorine process 115 116 leading to the production of acidic effluents with a high load in trace metals. The 117 extractive industry led to the production of by-products, such as red gypsum that are stored on large and open landfills. Red gypsum is produced by the neutralisation of waste-118 waters with limestone and is characterised by high Mn and Fe concentrations and elevated 119

pH (Asad *et al.*, 2017). *Betula pendula* was found to be the dominant (> 80%) plant
species that had naturally recolonised the D1 landfill area after the deposit of red gypsum
ceased in 2006 (Zapata-Carbonell et al., 2019). According to the "*National methodology for the management of polluted sites and soils*", (2017) developed by the French Ministry
of Environment, Energy and Sea, and based on the data recorded by Zapata-Carbonell *et al.*, (2019), the D1 area presents values higher than those found in natural soils for Cd,
Co, Cr, Cu, Pb, Zn and (especially) for Ti.

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128 2.2. Sampling procedure

All soil and plant samplings were performed in October 2016 in the D1 landfill area 129 (described previously by Zapata-Carbonell et al., 2019). Roots and shoots of B. pendula 130 were collected across the D1 area. Ten individual plants of B. pendula were sampled at 131 the centre of the landfill where *B. pendula* is the only tree growing. One soil sample was 132 taken from the root system of each plant. For comparative purposes, 10 individual trees 133 134 of other woody species growing at the edges (Populus tremula, P. nigra, Salix purpurea) 135 were sampled at three specific spots. Moreover, 10 individual plants of *B. pendula*, which were found growing in close proximity to these species at the edges, were also sampled. 136 137 These plant combinations were always growing at the same point and only one soil sample was taken for each pair of woody species. Soil samples (although not rhizospheric 138 139 soils) were collected in the close proximity of the roots. To clarify, we named B. pendula associated to P. tremula as B. pendula-(Pt), those associated to P. nigra as B. pendula-140 141 (Pn) and those associated to S. purpurea as B. pendula-(Sp). Birch plants sampled at the 142 centre were named as B. pendula (Fig. 1).

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144 2.3. Soil physicochemical properties and elemental composition of leaves

Soil pH and extractable-element concentrations were determined in the < 2mm fraction
of the dried soils. Soil pH was measured in H₂O using a 1:2.5 soil:solution ratio after 30
min agitation. To determine the extractable fractions of soil macro- and micro-elements,
5 g of 2 mm sieved soil were incubated with 50 mL of 10 mM CaCl₂ under agitation for
3 h at room temperature. The mixture was filtered with ash-free filters and acidified at
2% (v/v) with HNO₃ before inductively coupled plasma atomic emission spectrometry
analysis (ICP-AES, Thermo Fischer Scientific, Inc., Pittsburgh, PA, United States).

Plants were thoroughly washed with pressurised tap water to remove any adhering soil particles, and rinsed with deionised water. Leaves were dried for 48 h at 40 °C and 0.1 g of dry material was digested in a 4:1 HNO₃:H₂O₂ mixture. Shoot nutrients and trace elements were measured also by ICP-AES.

156 2.4. Molecular methods

157 The roots were carefully washed with tap water to remove visible soil and rinsed with sterile deionised water. The fine roots were then cut off and transferred to Eppendorf 158 tubes and frozen at -20 °C. DNA was extracted from roots of 10 single individual trees 159 160 from each plant species at each sampling point. To isolate DNA from the roots, plant cell lysis was first performed using 50 mg of frozen root samples that were pulverised with 3-161 162 mm tungsten carbide beads (Qiagen S.A.S., Courtaboeuf, France) in a Mixer Mill for 3 min at 30 Hz (modelMM400; Retsch Inc., Newtown, PA). Root DNA was then extracted 163 with a PowerSoil® DNA Isolation Kit (MO-BIO Laboratories, Inc., Carlsbad, CA USA). 164 To improve DNA extraction, an additional step of 10 min at 60°C was realised after 165 166 addition of C1 solution. The DNA quality and quantity were assessed by agarose gel electrophoresis and with the Quant-iTTM PicoGreen® dsDNA Assay Kit (Invitrogen, 167 Carlsbad, CA, USA) using an FLX-Xenius spectrofluorometer (SAFAS, Monaco). Two-168 169 step PCR amplifications were performed with the bacterial primers 341F (5'- CCT ACG GGR SGC AGC AG -3') and 802R (5'- GAC TAC HVG GGT ATC TAA TCC -3') (Baker 170 171 et al., 2003; Beckers et al., 2016) and the fungal primers ITS1F (5'- CTT GGT CAT TTA GAG GAA GTA A -3') and ITS2 (5'- GCT GCG TTC TTC ATC GAT GC -3') (White 172 173 et al., 1990; Gardes, 1993). The cycling conditions for 16S rRNA gene amplification involved an initial 3 min denaturing step at 94°C followed by 30 cycles of 25 s at 94°C, 174 175 25 s at 43.6°C, and 1:30 min s at 72°C. The cycling conditions for ITS genes included an initial 3 min denaturing step at 94°C followed by 30 cycles of 25 s at 94°C, 25 s at 45.4°C, 176 and 1:30 min s at 72°C. The sequencing of the V3-V4 domains of 16S rRNA genes and 177 the fungal ITS1 was performed with an Illumina MiSeq platform (Microsynth AG, 178 Switzerland). 179

180 2.5. Data analysis

Bioinformatics processing of the bacterial datasets was carried out using the MOTHUR pipeline (Schloss *et al.*, 2009). Raw reads were filtered by length and quality and 16S rRNA gene reads were pre-clustered using sumaclust (Mercier *et al.*, 2013) at 0.99

identity. Sequences with less than 8 reads were discarded, then remaining sequences were 184 185 aligned with those present in the Silva database to remove non-16S rRNA gene sequences, and plant DNA contamination was removed by suppressing reads identified 186 as k Bacteria (100), p Cyanobacteria (100), c Chloroplast (100), o Streptophyta 187 (100), and unclassified (100). Taxonomic assignments were made using a Bayesian 188 approach (Wang et al., 2007) with the Greengenes database (DeSantis et al., 2006). 189 Finally, OTUs classification was derived using the Needleman distance and average 190 191 neighbour clustering at a distance of 0.03.

Fungal Sequence de-multiplexing and bioinformatics processing of the datasets were performed using the PIPITS pipeline (Gweon *et al.*, 2015). This pipeline uses the most recent RDP Classifier to classify sequences against the curated UNITE fungal dataset. As for bacteria, fungal sequences were filtered by length and quality and clustered into OTUs, defined at 97% similarity. Singleton OTUs were excluded (Dickie, 2010; Tedersoo *et al.*, 2010) and the taxonomic assignment of OTUs was performed using the UNITE (Kõljalg *et al.*, 2013) database at a 97% similarity threshold.

199 2.6. Statistical analysis

All statistical analyses were performed using R software v. 3.0.2 (R Core Team, 2013). The Shapiro test was used to verify the normality of the data. Analysis of variance (ANOVA) was used to detect significant differences in shoot and soil measured parameters.

Richness and diversity indices were calculated based on a data set in which the 204 205 numbers of reads per sample were rarefied to 6972 and 28129 reads per sample for 206 bacteria and fungi, respectively. Rarefaction curves were created with the "rarecurve" 207 function in the Vegan package. NMDS was calculated using the Bray Curtis method (k =208 3) using the "metaMDS" function in the Vegan package. Results of relative abundance of the most abundant OTUs were paired with a heatmap of Spearman's correlations 209 created with the "heatmap.2" function using the gplot package. The numbers of OTUs 210 that were shared between plants were visualised using Venn diagrams implemented by 211 "VennDiagram" in the Vegan package. Indicator species analysis was performed using 212 the "multipat" function of the indicspecies package (De Caceres and Legendre, 2009). An 213 214 ANalysis Of SIMilarities (ANOSIM) was carried out to obtain P (significance levels) and R (the strength of the factors on the samples) values using the "anosim" function with the 215 vegan package. A spearman correlation was carried out to determine the relationships 216

between soil/plant physicochemical characteristics and the abundant OTUs and betweenfungal and bacterial OTUs. Differences in family and OTUs abundance between plants

219 were analysed using the Kruskal-Wallis test.

220 **3. Results**

3.1. Soil physicochemical properties and elemental composition of leaves

Supporting Information Table S1 and Supporting Information Table S2 present the main 222 soil properties as well as the leaf macro- and micro-nutrients used in subsequent statistical 223 224 analysis as environmental variables. In the case of trees growing on the edges, only one 225 soil sample for each pair of plant species has been studied. Briefly, soils present a strongly 226 alkaline pH (between 7.8 up to 10). The main CaCl₂-extractable elements found at the landfill were Mg > K > Fe > Mn > P > Cu and a relation between soil pH and CaCl₂-227 extractable Mn and Fe was clearly observed. Supporting Information Table S2 evidences 228 229 the Mn accumulation capacity of B. pendula, showing the highest levels of Mn leaf concentration (up to 570 mg kg⁻¹). Moreover, this accumulation was not related to the site 230 of growth (neither to the pH nor to the CaCl₂-extractable fraction of Mn) (Table S1 and 231 Table S2).Plants of birch also tended to accumulate a higher P concentration in their 232 233 leaves than the other plant species growing at the same sampling points although this difference was not significant (Table S2). 234

3.2. Richness and diversity of root bacterial and fungal communities for the 4 woody species

Following total genomic DNA extraction from root samples, amplicons of the V3-V4 16S 237 238 rRNA gene region were generated and a total of 5,262,333 paired-end reads were obtained through Illumina MiSeq sequencing (Supporting Information Table S3). After carrying 239 240 out a quality, chimera, singleton and contaminant filter, a total of 907,006 filtered and non-chimeric bacterial sequences constituted our final processed dataset, spread among 241 1,325 non-singleton OTUs defined by representative DNA sequences with sizes of 231 242 to 257 bp (mean = 237 bp). After subsampling, the dataset was rarefied to 6,972 reads per 243 sample (among the samples from each tree, those exhibiting a low sequence count were 244 245 eliminated from the rest of the analysis), which were distributed among 1,322 nonsingletons OTUs. Amplicons of the ITS region were generated and a total of 5,581,697 246 paired-end reads were obtained (Supporting Information Table S4). A total of 5,148,296 247

filtered and non-chimeric fungal sequences constituted the final processed dataset, spread
among 613 non-singleton OTUs defined DNA sequences with sizes above to 100 bp.
After subsampling, the dataset contained 28,129 reads per sample, distributed among 577
non-singletons OTUs.

Rarefaction curve analysis showed that all samples approached an asymptote in both the bacterial and fungal dataset, revealing that the overall diversity was well represented (Supporting Information Fig. S1). Moreover, the measured Good's coverage values (an estimator of completeness of sampling) were greater than 98% for each sample type in the case of the bacterial dataset, and 100% for the fungal dataset (Supporting Information Table S3 and S4 respectively).

258 Richness and diversity estimations were calculated for each dataset (Table 1 and Table 2). Both bacterial richness (based on Chao1 estimator and number of observed 259 260 OTUs) and diversity indices (estimated through the Shannon Index, Inverse Simpson Index and Evenness) were generally lower in the roots of the *B. pendula* plants collected 261 262 at the centre of the landfill than in the roots of all other trees collected at the edges. The highest bacterial richness and diversity indices were found to be associated with P. 263 264 tremula (Table 1). Fungal richness and diversity (Table 2) followed the same trend as for 265 bacteria and were generally lower in the roots of the *B. pendula* growing at the centre of the landfill. In this case, both richness indices (Chao1 and number of observed OTUs) 266 were highest in the roots of all birches growing at the edges. 267

3.3. Effects of plant species and growth location in microbial structure and composition

270 The bacterial dataset comprised a total of 17 phyla, 151 families, and 246 genera (17, 79 271 and 100 taxonomically identified respectively). The bacterial community was dominated 272 by Proteobacteria (38%), Actinobacteria (35%) and Bacteroidetes (20%). Only 26 (20 identified) families and 31 (14 identified) genera presented an abundance of higher than 273 274 0.5%. The fungal dataset was composed of a total of 7 phyla, 138 families, and 211 genera (5, 98 and 128 taxonomically identified respectively). The fungal community was 275 dominated by Ascomycota (60%) followed by Basidiomycota (30%). Only 18 (4 276 277 identified) families and 19 (5 identified) genera presented an abundance higher than 0.5%. Figure 2 shows a barplot for the 20 most abundant bacterial (Fig. 2a) and fungal 278 (Fig. 2b) families for each studied tree. The three most abundant identified families were 279 Cytophagaceae (19%), Streptomycetaceae (16%) and Hyphomicrobiaceae (15%) (Fig. 280

281 2a) and the three most abundant identified genera were *Lentzea* (7.11%), *Devosia* (5.52%) and Rhodoplanes (3.33%) (data not shown). The roots of B. pendula showed a slight 282 enrichment in members belonging to the genera Kribella, Rhodoplanes and Methylibium 283 compared to the other trees (data not shown). The most abundant identified fungal family 284 corresponded to Pyronemataceae (33%), followed by Russulaceae (15.5%). The next 285 identified family was Thelephoraceae, although present in a much lower abundance 286 (2.0%) (Fig. 2b). The most abundant genera were Geopora (33.1%), Lactarius (15.1%) 287 288 and Tomentella (2.0%). Roots of B. pendula showed a slight increase in members of the genera Oidiodendron and Geopora (data not shown). 289

290 Figure 2 shows a clear pattern in the distribution of families between each pair of 291 neighbouring trees (P. tremula - B. pendula (Pt), P. nigra - B. pendula (Pn) and S. purpurea – B. pendula (Sp)). However, due to the high data scattering, few of these results 292 293 were significant. In the case of the bacterial dataset, the statistical differences were mainly 294 associated to P. nigra and the neighbour birch (named as B. pendula (Pn)). These trees 295 harboured a higher number of members of the families Actinosynnemataceae (Kruskal Wallis $\chi 2=37.9$, $p=1.2\times10^{-6}$), *Glycomycetaceae* (Kruskal Wallis $\chi 2=34.61$, $p=5.2\times10^{-6}$) 296 297 and *Comamonadaceae* (Kruskal Wallis $\chi 2=20.6$, $p=2.1\times10^{-3}$) and lower members of the $\chi^{2=32.4}$, $p=1.4\times10^{-5}$) 298 families Pseudonocardiaceae (Kruskal Wallis and Methylocystaceae (Kruskal Wallis $\chi 2=39.2$, $p=6.6\times10^{-7}$) than all the other trees. In the 299 case of the fungal dataset the statistical differences were observed only for two families 300 *Pyronemataceae* (Kruskal Wallis $\chi 2=24.4$, $p=4.4\times10^{-4}$) and *Sebacinaceae* (Kruskal 301 Wallis $\gamma 2=20.5$, $p=2.3\times 10^{-3}$), which showed respectively a lower and higher number in 302 roots associated to *P. tremula* and the nearby growing birch (*B. pendula* (Pt)) 303

304 Table 3 shows the result of the ANOSIM analysis for the dissimilarity values of microbial communities between samples. A significant dissimilarity (p < 0.05) was 305 306 observed when all the tree species growing at the landfill were analysed together. Interestingly, when comparing *B. pendula* from the centre of the landfill with the other 307 308 tree species (Salix and Populus) or with the other B. pendula growing at the edges, a strong and significant dissimilarity in their root-associated bacterial communities was 309 310 observed. Similarly, trees growing at the different sampling points at the edges showed a 311 significant dissimilarity amongst their associated microbial communities. However, trees 312 of B. pendula growing at the edges of the landfill (B. pendula (Pt), B. pendula (Pn) and B. pendula (Sp)) did not show a significant dissimilarity between their root bacterial 313

communities and those of their neighbour trees (*P. tremula*, *P. nigra* and *S. purpurea*respectively).

The dissimilarity analysis between different root fungal communities, (ANOSIM, 316 Table 3) showed similar patterns as for bacteria although less marked. The highest 317 dissimilarity was found between *B. pendula* in the centre with either *B. pendula* (Pt) 318 (R²=0.621, p=0.001) or with P. tremula (R²=0.804, p=0.001). A high and significant 319 dissimilarity was found also when comparing together (a) all the plants of B. pendula 320 growing at the different locations ($R^2=0.327$, p=0.001); (b) the *B. pendula* at the centre 321 and the other tree species ($R^2=0.471$ p=0.001), (c) the other tree species growing at the 322 edges (either when comparing only *B. pendula* (R²=0.327, p=0.001) or (d) all species 323 $(R^2=0.472 p=0.001))$. No significant dissimilarity was observed when comparing B. 324 pendula either with B. pendula (Pn) or with P. nigra. 325

326 Supporting Information Table S5 shows the number of total reads of the most abundant OTUs (>0.5%) found in the bacterial collection. Only 33 OTUs (from the 1,322) 327 328 presented abundances higher than 0.5%, highlighting the presence of very rare OTUs dominating the community. The most abundant OTU belongs to the family 329 330 Streptomycetaceae (OTU00001, 14.9%). This OTU presented the highest abundance in 331 roots of B. pendula and S. purpurea, although it was also uniformly distributed in the roots of all the plant species and locations. Within the other abundant OTUs, 332 OTU00005 f:Micromonosporaceae, OTU00022 g:Kribella, 333 OTU00030 f:Methylocystaceae and OTU00034 g:Methylibium tended to be more 334 abundant in roots of B. pendula growing at the centre of the landfill (although significant 335 336 differences were not observed).

337 Total nº of reads of the most abundant OTUs (>0.5%) in the fungal dataset are 338 shown in Supporting Information Table S6. Similar to what was observed for bacteria, 339 only few OTUs presented abundance >0.5% (24 from a total of 577). The fungal community was clearly dominated by the OTU 25 92637 belonging to Geopora 340 341 arenicola species (31.8% of the total dataset). This OTU was present in high abundance in roots of all the studied trees, however, its abundance was slightly higher in the roots of 342 343 B. pendula at the centre. An extreme variability between replicates of the same plant species or growth location (which is reflected by the high standard errors in Supporting 344 345 Information Table S6) was observed, leading to the lack of important statistical 346 differences. The other most abundant OTUs belonged to the genera *Lactarius* (15.7%) 347 and the order *Pezizales* (7.2%), *Agaricales* (5.7%) and *Helotiales* (3.6%).

Supporting Information Figure S2 shows the distribution of unique and shared 348 349 bacterial (Supporting Information Fig. S2a and S2b) and fungal OTUs (Supporting Information Fig. S2c and S2d) amongst *B. pendula* growing at the centre and the other 350 351 trees growing at the edge. B. pendula from the centre presented 44 specific OTUs compared to other tree species and 60 specific OTUs when compared to other birch trees 352 growing at the edges. The number of specific OTUs in the B. pendula growing at the 353 edges (compared to the B. pendula at the centre) was similar to that found in their 354 355 neighbour trees. A similar pattern in the fungal dataset was found: B. pendula from the 356 centre presented 5 specific OTUs when compared to the other *B. pendula* but also when 357 compared to the other tree species.

358 A heatmap was carried out using the most abundant bacterial and fungal OTUs (>0.5% abundance) (Fig. 3). B. pendula growing at the centre clustered separately from 359 360 all other trees. Trees of different species, which were growing together at the edges were clustered in couples, depending on their growth location. The birches at the centre of the 361 362 landfill were more closely associated to trees growing at the locations of *P. nigra*. When 363 a heatmap was performed separately for either bacterial or fungal communities (data not 364 shown) the same pattern was found in the case of bacteria. Similarly, birch trees also 365 presented a specific fungal OTU distribution but in this case, plants of B. pendula were associated to the cluster made up of P. tremula-B.pendula (Pt) and S. purpurea-B. pendula 366 (Sp) (data not shown). 367

An analysis of indicator OTUs for each tree species and growth locations was also 368 369 carried out. A summary for B. pendula indicator OTUs is presented in Table 4. Fourteen indicator bacterial OTUs and 4 indicator fungal OTUs were observed for this tree species. 370 371 The bacterial OTUs include one belonging to the class Gammaproteobacteria, one to the 372 family Cytophagales, 3 to the family Rhizobiales, one to the family Streptomycetaceae 373 and one to the family Hyphomicrobiaceae, one to the genus Rhodoplanes and one to the genus Catellatospora. In the case of fungi, OTUs belonged to the species Oidiodendron 374 375 maius, the genus Russula, and the order Agaricales and Hypocreales.

376 3.4. Relationships between microbial communities, physicochemical soil properties 377 and elemental composition of leaves.

Considering an interspecies analysis based on the bacterial and fungal community composition using the most abundant OTUs (>0.5%), (NMDS plots (Fig. 4)) it was observed that roots of *B. pendula*, *P. tremula*, *P. nigra* and *S. purpurea* were clustered separately for each tree. The distribution of the samples of *B. pendula* was related to the
abundance of 5 specific bacterial OTUs (Otu00001 (f:*Streptomycetaceae*), Otu00017
(g:*Rhodoplanes*), Otu00009 (f:*Cytophagaceae*), Otu00030 (f:*Methylocystaceae*) and
Otu00014 (f:*Haliangiaceae*)) and to leaf Mn- concentration.

Supporting Information Table S7 shows the one-by-one correlations found 385 between the most abundant bacterial OTUs (>0.5%), soil physicochemical parameters 386 387 and leaf metal and nutrient concentrations. On the whole, correlations of several OTUs 388 with both, CaCl₂-extractable Fe and Mn concentrations were observed (either positive in the case of OTU00003 g:Lentzea, OTU00006 g:Devosia, OTU00012 g:Glycomyces, 389 390 and OTU00029 g:Catellatospora or negative in OTU00001 and OTU00026 both 391 belonging to the Streptomycetaceae family). Also strong positive correlations were found with some soil nutrients, such as CaCl₂-extractable P (OTU00002 f:Cytophagaceae, 392 393 OTU00026 f:Streptomycetaceae and OTU00028 g:Arenimonas) or CaCl2-extractable K (OTU00005 f:*Micromonosporaceae*, OTU00016 o:Rhizobiales 394 or 395 OTU00037 g:Dyadobacter). However, correlations with the corresponding leaf nutrient 396 concentrations were not observed (except in the case of OTU00002 f:Cytophagaceae 397 where a correlation was observed with soil CaCl₂-extractable P and leaf P concentration). 398 A positive correlation with both leaf K and P concentrations was found with OTU00003 g:Lentzea OTU00010 f:*Cytophagaceae*. 399 and Finally, OTU00012 g: Glycomyces showed a positive correlation with leaf P and OTU00029 with 400 leaf K concentrations. Similarly, in the case of fungal community, a strong correlation 401 402 with both, CaCl₂-extractable Fe and Mn was observed for the OTUs 403 24 66088 o: Pezizales, 23 37852 o: Agaricales and 51 62345 Unclassified. The OTU 24 66088 also showed a correlation with leaf K and P concentrations while the OTU 404 67 38353 o:Hypocreales was correlated with Mn leaf concentrations and OTU 405 406 51 62345 Unclassified with leaf P concentrations (Supporting Information Table S8).

Finally, a spearman correlation was carried out to study significant co-occurrence 407 408 in the community of the most abundant bacterial and fungal OTUs. A positive cooccurrence may be considered as co-presence while a negative co-occurrence as mutual 409 410 exclusion (Supporting Information Table S9). The bacterial **OTUs** OTU00003_g:Lentzea, OTU00004 f:*Hyphomicrobiaceae* 411 and 412 OTU00021 f:Methylocystaceae presented the highest numbers of significant cooccurrence with fungal OTUs (either positive or negative). The fungal OTUs 413 414 44 70433 g:Lactarius, 24 66088 o:Pezizales and 18 28512 p:Ascomycota showed the

highest number of significant co-occurrence with bacterial OTUs. On the contrary, there 415 416 were 4 bacterial OTUs (Otu00005 f: Micromonosporaceae, Otu00009 f: Cytophagaceae, Otu00030 f:Methylocystaceae and Otu00033 g:Rhodoplanes) and 3 fungal OTUs 417 (13 15910 o:Agaricales, 25 11792 s:Sphaerosporella 418 brunnea and 51 62345:Unclassified) which did not show any significant correlation with the 419 occurrence of either fungal or bacterial OTUs, respectively. The most abundant fungal 420 421 OTU 25 92637 s: Geopora arenicola presented a significant correlation with 5 bacterial OTUs and 4 of these were negative. The unique positive correlation was found with 422 423 Otu00003 g:Lentzea. Similarly, the major bacterial OTU, Otu00001 f:Streptomycetaceae was found to be significantly related to three fungal 424 425 OTUs (from those positively correlated with 37 4836 c:Sordariomycetes and 426 3 31227 o: Agaricales). Overall, the strongest correlations (rho>0.5) were found between 427 fungal OTU 24 66088 o:Pezizales and bacterial Otu00012 g:Glycomyces (rho=0.55), the fungal OTU 10 6322 Unclassified and the bacterial Otu00034 g:Methylibium 428 429 (rho=0.53), the fungal OTU 37 4836 c:Sordariomycetes and the bacterial Otu00039 f:Cytophagaceae (rho=0.58) and finally between the 12 355 s:Tomentella 430 431 *ellisii* and the bacterial Otu00015_c:TM7_1 (rho=0.64).

432 **4. Discussion**

433 This study aimed to describe the root-associated microbiota of *B. pendula* which naturally 434 colonises a red-gypsum landfill in the NW of France. It is well known that birch is a pioneering plant species, able to establish in nutrient deficient and degraded soils 435 436 (Kopponen et al., 2001). Part of this success is related to their well-developed root 437 system(Birch is also a high biomass producer and fast-growing tree which can provide 438 high economic return with end-use applications such as the production of biofuels, pulp, 439 and paper and other bio-based products such as biocatalysts (Kidd et al., 2015; Asad et al., 2017). Furthermore, birch can be grown on marginal land thus evading the food versus 440 fuel debate (Searchinger et al., 2008). On the other hand, plants harbour multiple 441 microbial taxa, which affect several plant traits such as biomass production, nutrient 442 443 availability and transport from the soil (Sugiyama et al., 2012), metabolite production (Badri et al., 2013), drought tolerance (Lau and Lennon, 2012) and flowering time 444 445 (Wagner et al., 2014). In the last years, several studies have also highlighted the influence of mycorrhizas on the alleviation on metal toxicity to plants through extracellular 446 inactivation, metal binding in fungal wall, enhanced efflux pumps activity, intracellular 447

inactivation and adaptive response to oxidative stress (Amir *et al.* 2014). In addition, the
composition of the plant microbiota can enhance host resistance to pathogens (Mendes *et al.*, 2011; Busby *et al.*, 2016).

Little information about root microbial communities associated with *B. pendula* growing at contaminated sites using metabarcoding approaches is available in the literature. Some studies such as Mesa *et al.* (2017) studied the rhizospheric and endospheric bacterial communities associated with *Betula celtiberica* growing at a contaminated abandoned fertiliser industrial site in NW Spain. Data of the number of observed OTUs and Chao index found in roots of *B. celtiberica* by these authors are comparable with those found in the present study.

458 Here, B. pendula growing at the centre of the landfill showed, generally, lower richness and diversity (either in bacterial and fungal communities) than the other trees 459 460 growing at the edges. One possible explanation could be the presence of harsher physicochemical soil properties at the centre of the landfill, which would negatively affect 461 462 root microbial communities. However, we did not find any clear difference in the 463 measured parameters (pH and CaCl₂-extractable elements) between the different studied 464 areas. Similarly, in a deeper study of the area, Zapata-Carbonell et al. (2019) did not 465 observe any pattern of distribution on measured soil nutritional parameters (based on organic matter and CaCl₂-extractable elements) between the centre and the edges of the 466 landfill. However, it is also important to note that other factors (such as water, N content 467 or CEC) could also be affecting the distribution and composition of plant and microbial 468 469 communities. It has been reported by several authors that beneficial microorganisms, increase plant tolerance to drought, (Timmusk et al., 2014). Moreover, the incidence and 470 471 virulence of diseases or pests can be higher in plants growing in stressed environments 472 (such as degraded soils). Associated microorganisms can also protect plants against 473 diseases through for example the induction of plant defence mechanisms or the production of antibiotics (Compant et al., 2010). Related to that point, it is worth 474 475 mentioning the importance of members of the order Actinomycetales (notably the Streptomyces genus), as the richest natural source of antibiotics or antimetabolites (Bérdy, 476 477 2005; Olano et al., 2009; Lemanceau et al., 2006).

In the present study, the bacterial community was dominated by *Proteobacteria* and *Actinobacteria*. It is important to note that here, no distinction between endophytes and epiphytes was made, and therefore only one habitat was considered. However, roots were exhaustively washed and we can consider that only strongly root-associated microorganisms are being discussed. Interestingly, *Proteobacteria* (families such as *Pseudomonadaceae*, *Burkholderiaceae* or *Xanthomonadales*) and several OTUs
belonging to *Actinobacteria* have been shown to be the most dynamic taxa associated
with disease suppression in the rhizosphere of sugar beet plants in a study using DNA
metagenomics with Phylochip (Mendes *et al.*, 2011).

Mesa et al. (2017) also found that in the endosphere of B. celtiberica, 487 488 Proteobacteria was the most abundant phylum (accounting for 64% of the total reads), 489 followed by Bacteroidetes (17%), Actinobacteria (7.9%), Firmicutes (1.8%), and Chlamydiae (1.3%). It is generally accepted that the Alpha- and Betaproteobacteria are 490 491 copiotrophic members (Fierer et al., 2007; Zhao et al., 2014), since they are usually 492 associated with habitats having relatively low soil C/N ratio, high NO-3-N (Nugroho et 493 al., 2005), and enriched nutrients (Leff et al., 2015; Li et al., 2016). Roots constitute a 494 nutrient-rich niche since they receive high amounts of available C and N as exudates, 495 which could promote the growth of these bacterial inside the root and at the rhizoplane. 496 Mesa et al. (2017) observed in roots of B. celtiberica, that the phylum Proteobacteria 497 was dominated by taxa related to Burkholderiales, and Pseudomonadales. However, in 498 our study, the phylum Proteobacteria was mostly represented by the family 499 Hyphomicrobiaceae (order Rhizobiales) and the genera Devosia and Rhodoplanes, whereas *Pseudomonas*, (which are frequently described in contaminated soils) were very 500 rare (only three OTUs belonging to the Pseudomonas genera were observed (Otu00824, 501 502 Otu01204 and Otu00214)) with an abundance ranging from 0.001% to 0.064%.

503 Moreover, OTUs belonging to the order Rhizobiales (including members of the 504 family *Hyphomicrobiaceae* and the genera *Rhodoplanes*) were found to be indicators of 505 the roots of *B. pendula*. There is little information about the role of *Rhodoplanes* in the 506 environment, and several studies related their preference to inhabiting rhizospheric soil 507 rather than endospheric habitats (Touceda-González et al., 2015; Gkarmiri et al., 2017). These bacteria were also abundant in the root DNA-based community of Arabidopsis 508 509 thaliana (Haichar et al., 2012) and Panax notoginseng (Tan et al., 2017). A study which used ¹⁵N-DNA-SIP to investigate soil microorganisms responsible for N fixation, 510 511 identified, among others, bacteria of the genus Rhodoplanes as being potential N fixers (Buckley et al., 2007). The abundance of microorganism related to N-cycling together 512 513 with the low values of soil N reported in previous studies (Zapellini et al. 2018) seem to indicate that these bacteria could be important for the growth of birch in the landfill. 514 515 Several members of rhizobiales (such as Ensifer, Rhizobium, Phyllobacterium) have been

shown to reduce plant stress response and improve plant growth in different hosts and
soils (Larcher *et al.*, 2008; Guo and Chi, 2014; Lafuente *et al.*, 2015; Mesa *et al.*, 2017).
Mesa *et al.* (2017) have also observed members of the *Rhizobiales* present in both the
rhizosphere and endosphere of *B. celtiberica*. Moreover, in the same study, most of the
culturable strains from the endosphere belonged to the genus *Rhizobium* (11%).

521 Actinobacteria was the second most represented phylum and the main OTU 522 (Otu00001) of the dataset belonged to the Streptomycetaceae family. The predominance 523 of this phylum was also found in several other studies involving either soil or plant 524 compartments (Lopez et al., 2017). For example, Ghosh and Das (2018) studied the soil 525 bacterial community in Mn-enriched mine tailings by using a shotgun-metagenomic 526 approach and found that Proteobacteria, and Actinobacteria dominated the soil 527 community (and specifically the Streptomycetaceae family). Several studies have 528 confirmed the high adaptability of Actinobacteria to toxic concentrations of metals (DeGrood et al., 2005; Schmidt et al., 2005) and it has been shown that their strong 529 530 secondary metabolism enables them to cope with stress factors (So et al., 2001). In one of our recent study (Zappelini et al. 2018), we found that Actinobacteria (and specifically 531 532 the genera Streptomyces) was the predominant phylum in the soil culturable bacterial 533 collection isolated from this same soil. The high proportion of the Actinobacteria phylum, which was found amongst all tree roots (independently of the species and growth 534 location), could suggest the importance of the soil bacterial community in determining 535 the structure and composition of root-associated communities. Similarly, Yang et al. 536 (2017) found that after growing barley seeds in soils, an enrichment in Actinobacteria 537 was observed in roots of plantlets compared to seeds growing in axenic substrate, 538 suggesting that the soil had a major impact driving the root bacterial communities. 539

The biological control activity of these endophytic actinobacteria is also of 540 541 interest given their ability to colonise healthy plant tissues and produce antibiotics in situ (Shimizu, 2011). For example, the endophytic Streptomycetes isolated from healthy 542 543 banana plants (Musa sp.) were studied for the ability to produce antifungal molecules that inhibited the growth of the pathogen Fusarium oxysporum (Cao et al., 2005). An 544 545 enrichment of Streptomyces in roots compared to rhizospheric soil is frequently described in the literature (Foulon et al., 2016; Durand et al., 2018). Moreover, Gkarmiri et al. 546 547 (2017) suggested that within the active bacterial community that Streptomyces were highly active in the root compartment but not in the rhizosphere. 548

In the present study, Otu00716 belonging to Catellatospora sp. (family 549 Micromonosporaceae, phylum Actinobacteria) was found to be an indicator species of 550 the birch trees at the central location. The genus *Catellatospora* is considered one of the 551 rarely occurring genera of actinomycetes: only a few papers mentioning these 552 microorganisms can be found and most of them concern descriptions of the few species 553 554 and their taxonomic position (Saracchi et al., 2004). During a study of the endophytic 555 root actinomycete population from 252 leaf litter and 205 plant root samples (Petrolini, 556 1992; Sardi et al., 1992) found that most of the 2950 isolates obtained belonged to Streptomyces (Sardi et al., 1992) and 215 to the genus Micromonospora (Williams et al., 557 1993). However, representatives of the "rare" genera were also found (Petrolini, 1992; 558 559 Petrolini et al., 1995), and among these, 53 strains of Catellatospora sp. However, in the present study 6 OTUs belonging to this genus were observed. Moreover, one of these 560 561 (Otu00029) was one of the abundant OTU (abundance >0.5%), although this was not the OTU with highest abundance in the birch roots. 562

563 Members of the plant microbiota can be transmitted either horizontally (acquired from the surrounding environment) or vertically (acquired directly from the parent) 564 565 (Gundel et al., 2011). The results showed that B. pendula harboured a specific microbial 566 community which is unique and different, not only from other plant species growing at the red gypsum landfill (Populus or Salix spp.) but also from other B. pendula growing at 567 the edges (analysis of similarities ANOSIM, heatmap and hierarchical cluster analysis of 568 the relative abundance of bacterial and fungal OTUs), suggesting that horizontally 569 transmission is an important via of microbial acquisition in this case. Moreover, this plant-570 location specific community was more noticeable in the case of the bacterial community. 571 572 However, concerning the distribution of unique and shared OTUs, it was found that, even though a high and consistent number of OTUs were unique for each plant species, most 573 574 of them were shared among different plant species or sites of growth. These results also suggest that a generalised plant colonisation by certain indigenous soil bacterial species 575 576 is occurring and that the specificity of microbial communities can be related mainly to the differences between OTU abundances. Yang et al. (2017) found that the differences 577 578 in root active microbiomes associated with different barley cultivars were mainly quantitative and not as much qualitative. For example, the variation between the 579 580 genotypes was manifested in the abundance of many OTUs from diverse taxa, rather than 581 by the presence/absence of single OTUs in the given genotypes.

582 Data obtained from fungal sequencing showed a much higher replicate variability 583 than those obtained in bacterial sequencing. The same pattern was found in several other 584 studies when comparing fungal communities from soil and root habitats (Beckers *et al.*, 585 2017; Durand *et al.*, 2017; Kolaříková *et al.*, 2017). This result suggests that fungal 586 endophytic colonisation and formation of stable communities appears to be a more 587 variable process than bacterial colonisation (Beckers *et al.*, 2017).

588 Although in our study most taxa (of both bacterial and fungal communities but 589 especially in the case of fungi sequencing) were not sequenced at a high enough read 590 depth to enable powerful comparisons, an enrichment of the genus Geopora (specifically 591 G. arenicola) in the roots of all tree species was observed and tended to be slightly higher 592 in roots of B. pendula at the centre. Kolaříková et al. (2017) investigated the root-593 associated fungal communities of two dominant trees Salix caprea and B. pendula along 594 a primary successional chronosequence on a mine spoil bank in the Czech Republic and 595 found that G. arenicola) was specifically associated to the youngest trees in the 596 chronosequence independently of the tree species. Geopora arenicola is an 597 ectomycorrhizal fungus, which, among others has the characteristic to promote plant 598 growth in stressful environments (Hrynkiewicz et al., 2009). Within the fungal indicator 599 OTU, we found the genera Oidiodendron and Russula (OTU 3 5515 s:Oidiodendron maius|SH217755.06FU and OTU 4 35502 g:Russula). Russula is an 600 601 ECM fungus frequently found in soils and also associated with birch cultures (Bent et al., 2011; Parker et al., 2017). Moreover, Parker et al. (2017) screened plant/soil fungal 602 communities associating with Picea mariana, Populus tremuloides and Betula papyrifera 603 604 after a wildfire and found that each plant species seemed to be affected by different groups 605 of fungi. A higher shoot biomass of birch was associated with those associated with 606 ribotype 732 to 734 bp belonging to *Russula* spp. It is interesting to note, however, that 607 the Russulaceae family was not abundant in roots of the central B. pendula. It seems that 608 the specificity of the root-associated microbial community of the central birch plants is 609 not related to the most abundant taxa. Interestingly, Durand et al. (2017) found that the less abundant fungal OTUs present a higher co-occurrence network connectivity than the 610 611 most abundant OTUs. In the present study it was also observed that the most abundant 612 OTUs (both fungal and bacterial) showed a low number of bacterial and fungal 613 correlations. Fungi and bacteria share the same habitats and are therefore almost certain to frequently interact in soil and plants (Boer et al., 2005; Zhao et al., 2014). Interactions 614 615 between co-occurring bacteria and fungi play a considerable part in determining their

respective roles, and investigating these interactions will lead to a more completeunderstanding of microbial ecology (de Menezes *et al.*, 2017).

618 Surprisingly, the second indicator OTU belonged to the ericoid mycorrhizal group 619 *Oidiodendron maius* (Chambers *et al.*, 2008; Lukešová *et al.*, 2015); although no 620 *Ericaceae* plants are described at this landfill. Although less frequently, this mycorrhizal 621 type has indeed been found in association with other plant species (Chambers *et al.*, 2008)

622 Similarly to what was observed in the bacterial dataset (although not clearly 623 reflected by the ANOSIM analysis), B. pendula growing at the centre hosted a specific 624 fungal community while trees growing at the edges shared an important part of both root 625 community structure and composition. Bogar and Kennedy (2013) studied the relative 626 importance of host and neighbourhood effects on the ECM fungal communities 627 associated with Alnus rhombifolia (a host of specific ECM fungi) and Betula occidentalis 628 (a host of generalist ECM fungi). The authors found that while the host identity acts as a 629 primary filter on the composition and diversity of ECM fungal communities, proximity to a closely related host can mediate significant changes in community structure. In the 630 same way, Hausmann and Hawkes (2009) found that the structure of arbuscular 631 632 mycorrhizal (AM) fungal communities on neighbouring grass plants depended upon 633 which pairs of host species were near each other. Neighbourhood effects can also be driven by changes in abiotic conditions that favour the growth of certain ECM fungal 634 species, as demonstrated by Meinhardt and Gehring (2012); in that study, the authors 635 examined the influence of invasive Tamarix spp. on the mycorrhizal associations of 636 native Populus fremontii in an early-invasion field site and in the greenhouse and 637 laboratory. They found that the ECM fungal community structure on P. fremontii shifted 638 639 significantly in the presence of *Tamarix* spp., likely as a result of increases in soil nitrate 640 concentrations and electrical conductivity associated with soils affected by Tamarix. 641 Some ECM fungi, Russula luteolus, Suillus bovinus, and Hebeloma crustuliniforme, form highly branched mycelial strands up to 40 cm (skinner and Bowen, 1974; Finlay and 642 643 Read, 1986a, b). In addition to the beneficial effects on the plant nutritive status, during the scavenging process fungi will simultaneously colonise more than one plant found at 644 the site. 645

As previously described, abundances of both fungal and bacterial phyla were highly variable. Ghosh and Das (2018) also found a high variability in the abundance of different phyla and classes in the soil bacterial communities of a Mn-enriched tailing by using a shotgun-based approach. These authors suggest that this could be the response of

heterogeneity in the environmental conditions and they highlight the importance of 650 651 studying and integrating results of bacterial diversity and biotic and abiotic related factors for its correct interpretation. Although significant correlations were found between 652 bacterial and fungi OTUs and soil physicochemical properties or plant nutrients, these 653 were generally very weak (rho value usually below than 0.6). This could be due to the 654 655 lack of a strong gradient in soil properties and rather a random heterogeneity found in these constructed waste landfills. Moreover, although this study involved the 656 657 characterisation of epispheric root microorganisms (which would be directly influenced by soil properties) most of the analysed community belonged to the endophytic 658 659 community and therefore are not directly influenced by the soil physicochemical 660 properties. However, some interesting results were found, for example the correlation 661 between some bacterial and fungal OTUs and both soil CaCl₂-extractable Fe and Mn. 662 This could indicate the co-dissolution of Mn- and Fe-oxides from this soil leading to the release of both elements. A deeper study involving the isolation of these bacterial strains 663 664 and characterisation of the biological pathways involved in this process would be 665 interesting for further exploitation of these bacterial strains in Mn-biomining. Das et al. 666 (2011) review the role of microorganisms in Mn-bioleaching operations for the recovery 667 of metals from low-grade ores that would be otherwise unfeasible using current metallurgical technologies by their prohibitively high energy and capital costs. In general, 668 669 dissolution of Mn-rich minerals occurs through enzymatic reduction which is often carried out by bacteria in the form of respiration where oxidised Mn acts as a terminal 670 671 electron acceptor or by the indirect release of reductive compounds resulting from 672 bacterial metabolism (Das et al., 2011; Ghosh et al., 2016). In both cases the biological 673 process occurs in the presence of organic carbon and energy sources (Das et al., 2011; 674 Ghosh et al., 2016). As it is well known, roots are a C-rich environment since up to 50 % 675 of the net photosynthetic carbon is released from plant roots into the rhizosphere as rhizodeposits (Curl and Truelove, 1986). At the same time, it is known that the molecular 676 677 mechanisms behind Fe and Mn tolerance are frequently the same in bacteria (Chandrangsu et al., 2017). 678

Finally, several OTUs were correlated with either soil nutrient availability and/or shoot nutrient concentrations. All trees growing at the area present shoot P and K deficiency (according to the values reported by Marschner (1995)), and therefore plantassociated microorganisms involved in the uptake and absorption of these elements can play a key role in the growth of the vegetation at the site. At the same time, Zapata684 Carbonell *et al.* (2019) found a correlation between the distribution of *B. pendula* at the 685 site and CaCl₂-extractable P and organic matter, highlighting again the importance of 686 nutrients in the growth of the plants at this site.

687

688 5. Conclusions

To successfully implement a phytomanagement strategy it is vitally important to 689 690 characterise the plant-soil-microbial system. In this study we characterised the microbial 691 community associated with the roots of B. pendula, the most successful plant species 692 colonising a red gypsum landfill. The bacterial community associated with birch growing 693 in the centre of the landfill was very different from that of other tree species studied at 694 the mine but also that of B. pendula growing at the edges of the landfill. The most abundant bacterial OTU (Otu 00001) belonged to the family Streptomycetaceae and was 695 696 widespread through all the studied roots. At the same time, according to the NMDS 697 analysis, this OTU seemed also to be related to the presence of birch at the centre of the 698 landfill. Other bacterial OTUs associated to birch roots (such as members of the order 699 *Rhizobiales*) are known to be involved in the N cycling, suggesting that this may be a 700 critical factor for effective growth of the plants at the site.

Data obtained from fungal sequencing showed a much higher replicate variability than those obtained in bacterial sequencing. In the case of fungi, a less clear pattern was observed although it seems also that birch selected some specific OTUs (e.g OTUs belonging to the genera *Oidiodendron* and *Russula*, and the order *Agaricales* and *Hypocreales* were found to the indicators of the presence of *B. pendula*).

Several OTUs specifically associated to the birch growing at the centre of the area
are known to be involved in nutrient cycling (e.g. members of *Rhizobiales* or the fungal
species *Geopora*) and can therefore play a role in the high capacity of this plant to tolerate
the growth limiting conditions in the landfill.

710 Differences between microbial communities (bacterial and fungal) were based on 711 microbial structure but also on microbial composition and distribution: e.g. the Venn diagram showed that the unique bacterial and fungal OTUs associated to each plant were 712 713 always present in a very low proportion in the total community (< 0.5% of abundance). However, an analysis of the most abundant OTUs, generally shared by all plants (e.g. 714 715 heatmap or NMDS) also showed a specificity of the communities associated with B. pendula. Finally, it is also important to remember that assemblages of host-associated 716 717 microbial communities are often function-based rather than structure-based (Burke et al.,

2011; Mendes *et al.*, 2014). Therefore, additional studies on the active part of microbial
communities (e.g. rDNA sequencing) or microbial ecosystem functionality (e.g., based
on commercial chips (e.g. GeoChip or Shotgun) or the *in vitro* characterisation of the
culturable community) can give a more complete view of these plant-soil-microorganism
systems.

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732 7. Conflict of interest

- 733 Authors have no conflict of interest to declare
- 734

735 8. Bibliography

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1056 Figure legends

Figure 1 Map of sampling points for each plant species. *Betula pendula* sampled at the
center of the landfill and named as *B. pendula*. Plants of *P. tremula*, *P. nigra* and *S. purpurea* were sampled at the edges. Plants of *B. pendula* growing in association with
these species at the edges were also sampled and named as *B. pendula*- (Pt), *B. pendula*(Pn) and *B. pendula*-(Sp). One soil sampling point was carried out for each plant pair.

Figure 2 Composition of the root-associated bacterial (a) and fungal (b) communities.
Only the 20 most abundant identified families are presented.

1064 Figure 3 Heatmap and hierarchical cluster analysis of the most abundant bacterial and 1065 fungal OTUs (>0.5%). The dendrogram represents linkage clustering using Euclidean 1066 distance measures. OTU delineation was based on a threshold of <97% sequence 1067 similarity. Assignments between brackets show the lowest taxonomic level associated 1068 with the OTU: k kingdom, p phylum, o order, c class, f family, g genus, s genus_species

Figure 4 Non-parametric multidimensional scaling (NMDS) plot of fungal and bacterial abundant OTUs (>0.5%) communities associated with the roots of *B. pendula*, *P. tremula*, *P. nigra* and *S. purpurea*. Each point represents the bacterial community of a given root.
Each color represents one of the four trees.

	B. pendula	P. tremula	B. pendula (P.t)	P. nigra	B. pendula (P.n)	S. purpurea	<i>B. pendula</i> (S.p)
Mean number of sequences per sample	13,317	9,912	13,024	13,907	12,163	11,973	16,405
Subsample size	6,972	6,972	6,972	6,972	6,972	6,972	6,972
N° observed OTUs	257 ± 42 d	345 ± 31 a	$310 \pm 32 \text{ bc}$	$329\pm26~ab$	$292 \pm 36 \text{ cd}$	316 ± 59 abc	$287 \pm 59 \text{ cd}$
Chao estimation	$377\pm70~d$	505 ± 46 a	464 ± 62 abc	$483\pm42 \; ab$	$422 \pm 48 \text{ cd}$	$450 \pm 94 \ bcd$	$443 \pm 107 \text{ bcd}$
Shannon Index (H)	$3.19\pm0.31~\text{c}$	3.73 ± 0.33 a	$3.47\pm0.38~abc$	$3.54\pm0.24\ ab$	3.44 ± 0.31 abc	$3.46\pm0.49~abc$	$3.36\pm0.42\ bc$
Inverse Simpson Index (1/D)	$10.5\pm2.8~b$	$16.2 \pm 7.3 \text{ a}$	$12.2\pm5.0 \text{ ab}$	$14.1 \pm 3.6 \text{ ab}$	$13.5 \pm 5.3 \text{ ab}$	12.6 ± 5.4 ab	$11.9 \pm 4.5 \text{ ab}$
Shannon Index Evenness E	$0.58\pm0.04~\text{b}$	$0.64\pm0.05~a$	$0.61\pm0.06~ab$	$0.61\pm0.04~ab$	$0.61\pm0.04\;ab$	$0.60 \pm 0.07 \text{ ab}$	0.59 ± 0.06 ab

Table 1. Richness and diversity indices for the root-associated bacterial communities of the woody species sampled at the Thann site

All diversity indices were calculated using an OTU threshold of \geq 97% sequence similarity on randomly sub-sampled data at the lower sample size (6,972 reads). Richness was calculated using the number of OTUs and Chaol estimators. Diversity was estimated using the Shannon-Wiener (H), Inverse Simpson (1/D), and Shannon Index Evenness (E) indices. Mean values and standard deviations (mean \pm SD) are provided for the root samples. Values designated with the same letters were not significantly different (Kruskal Wallis test, p<0,05).

	B. pendula	P. tremula	B. pendula (P.t)	P. nigra	B. pendula (P.n)	S. purpurea	B. pendula (S.p)
Mean number of sequences per sample	81,989	65,652	64,469	74,027	74,192	70,200	84,301
Subsample size	28,129	28,129	28,129	28,129	28,129	28,129	28,129
N° observed OTUs	52 ± 13 d	$79 \pm 16 \text{ ab}$	84 ± 14 a	58 ± 14 cd	$70\pm20~bc$	$70 \pm 18 \text{ bc}$	$79 \pm 14 \text{ ab}$
Chao estimation	$72 \pm 24 b$	103 ± 18 a	105 ± 23 a	$81 \pm 25 \text{ b}$	$89 \pm 25 ab$	$94 \pm 30 \text{ ab}$	102 ± 18 a
Shannon Index (H)	$1.33 \pm 0.37 \ c$	$1.80\pm0.30\ ab$	1.97 ± 0.36 a	$1.56\pm0.30~bc$	1.56 ± 0.56 bc	1.52 ± 0.34 bc	$1.66 \pm 0.55 \text{ abc}$
Inverse Simpson Index (1/D)	$2.96\pm1.05~\text{b}$	$3.78\pm1.15 \text{ ab}$	4.33 ± 1.93 a	$3.31\pm0.93 \text{ ab}$	$3.74\pm2.50\ ab$	$3.18\pm0.98 \text{ ab}$	$3.88\pm2.16\ ab$
Shannon Index Evenness E	$0.34\pm0.09~b$	0.41 ± 0.06 ab	0.44 ± 0.07 a	$0.39\pm0.07~ab$	$0.37\pm0.11~\text{b}$	$0.36\pm0.07~b$	0.38 ± 0.11 ab

Table 2. Richness and diversity indices for the root-associated fungal communities of the woody species sampled at the Thann site

All diversity indices were calculated using an OTU threshold of \geq 97% sequence similarity on randomly sub-sampled data at the lower sample size (28,129reads). Richness was calculated using the number of OTUs and Chao1 estimators. Diversity was estimated using the Shannon-Wiener (H), Inverse Simpson (1/D), and Shannon Index Evenness (E) indices. Mean values and standard deviations (mean \pm SD) are provided for the root samples. Values designated with the same letters were not significantly different (Kruskal Wallis test, p<0,05).

	Ba	acteria	Fungi		
Habitats	R ²	p-value	R ²	p-value	
Global	0.384	0.001**	0.311	0.001**	
Intra-species					
<i>B. pendula- B. pendula</i> (Pt)- <i>B. pendula</i> (Pn)- <i>B. pendula</i> (Sp)	0.379	0.001**	0.327	0.001**	
B. pendula (Pt)- B. pendula (Pn)- B. pendula (Sp)	0.405	0.001**	0.327	0.001**	
B. pendula - B. pendula (Pt)	0.284	0.001**	0.621	0.001**	
<i>B. pendula - B. pendula</i> (Pn)	0.617	0.001**	0.058	0.134	
B. pendula - B. pendula (Sp)	0.174	0.028*	0.206	0.015*	
Inter-species					
B. pendula - P. tremula - P. nigra - S. purpurea	0.48	0.001**	0.471	0.001**	
B. pendula –P. tremula	0.539	0.001**	0.804	0.001**	
B. pendula –P. nigra	0.607	0.001**	0.145	0.038*	
B. pendula –S. purpurea	0.178	0.028*	0.12	0.041*	
B. pendula (Pt)- P. tremula	-0.028	0.607	0.133	0.02*	
B. pendula (Pn)- P. nigra	-0.031	0.613	-0.018	0.552	
B. pendula (Sp)- S. purpurea	0.014	0.338	0.122	0.04*	
<i>B. pendula</i> (Pt). <i>P. temula- B. pendula</i> (Pn). <i>P.nigra-</i> <i>B. pendula</i> (Sp). <i>S. purpurea</i>	0.472	0.001**	0.472	0.001**	
0 **** 0.001 *** 0.05 **					

Table 3. Analysis of similarities (ANOSIM) of the bacterial and fungal communities associated with the different trees and their interactions. An ANOSIM R value of 1 indicates complete dissimilarity between groups. Significant levels were estimated (p-value < 0.05)

		stat	p.value	Phyllum	Class	Order	Family	Genus / Species
	Otu01003	0.775	0.001 ***	Proteobacteria	Gammaproteobacteria	unclass	unclass	unclass
	Otu00390	0.721	0.004 **	Bacteroidetes	Cytophagia	Cytophagales	unclass	unclass
	Otu00260	0.655	0.001 ***	Proteobacteria	Alphaproteobacteria	Rhizobiales	unclass	unclass
	Otu01224	0.636	0.006 **	Proteobacteria	Alphaproteobacteria	Rhizobiales	unclass	unclass
_	Otu00475	0.618	0.004 **	Proteobacteria	unclass	unclass	unclass	unclass
S S	Otu00331	0.612	0.005 **	Actinobacteria	Actinobacteria	Actinomycetales	Streptomycetaceae	unclass
E	Otu00799	0.604	0.006 **	Proteobacteria	Alphaproteobacteria	Rhizobiales	Hyphomicrobiaceae	Rhodoplanes
5	Otu00429	0.566	0.01 **	Proteobacteria	Alphaproteobacteria	Rhizobiales	Hyphomicrobiaceae	unclass
3A	Otu00480	0.566	0.012 *	unclass	unclass	unclass	unclass	unclass
-	Otu00416	0.555	0.008 **	Proteobacteria	Alphaproteobacteria	Rhizobiales	unclass	unclass
	Otu00574	0.53	0.017 *	Proteobacteria	Alphaproteobacteria	unclass	unclass	unclass
	Otu01267	0.516	0.036 *	Proteobacteria	Alphaproteobacteria	unclass	unclass	unclass
	Otu00716	0.5	0.038 *	Actinobacteria	Actinobacteria	Actinomycetales	Micromonosporaceae	Catellatospora
	Otu01261	0.474	0.041 *	unclass	unclass	unclass	unclass	unclass
F	3_5515	0.993	0.001 ***	Ascomycota	Dothideomycetes	Incertae_sedis_8	Myxotrichaceae	Oidiodendron_maius SH217755.06FU
ğ	41_79123	0.735	0.005 **	Basidiomycota	Agaricomycetes	Agaricales	Unclass	Unclass
5	1_99723	0.577	0.002 **	Ascomycota	Sordariomycetes	Hypocreales	Unclass	Unclass
<u> </u>	4_35502	0.572	0.007 **	Basidiomycota	Agaricomycetes	Russulales	Russulaceae	Russula

Table 4. Indicator fungal and bacterial OTUs of *B. pendula* plants

Associations were calculated with the Dufrene–Legendre indicator species analysis routine in R. Significance levels: $*P \le 0.05$; $**P \le 0.01$.



Figure 1 Map of sampling points for each plant species. *Betula pendula* sampled at the center of the dump and named as *B. pendula*. Plants of *P. tremula*, *P. nigra* and *S. purpurea* were sampled at the edges. Plants of B. pendula growing in association with these species at the edges were also sampled and named as *B. pendula*- (Pt), *B. pendula*- (Pn) and *B. pendula*-(Sp). Only one soil sampling point was carried out for each plant pair.



Figure 2 (a) Composition of the bacterial communities from the different roots at the family level and (b) composition of the fungal communities from the different roots at the family level. Only the 20 most abundant identified families are presented.



Figure 3 Heat map and hierarchical cluster analysis of the most abundant bacterial and fungal OTUs (>0.5%). The dendrogram represents linkage clustering using Euclidean distance measures. OTU delineation was based on a threshold of <97% sequence similarity. Assignments between brackets show the lowest taxonomic level associated with the OUT: k kingdom, p phylum, o order, c class, f family, g genus, s genus_species



Figure 4 Non-parametric multidimensional scaling (NMDS) plot of fungal and bacterial abundant OTUs (>0.5%) communities associated with the roots of the *B. pendula*, *P. tremula*, *P. nigra* and *S. purpurea*. Each point represents the bacterial community of a given root. Each color represents one of the four trees.