

Interspecific competition between two partridges in farmland landscapes

Tony Rinaud, Clément Harmange, Olivier Pays, Mathieu Sarasa, Maxime Saillard, Vincent V. Bretagnolle

▶ To cite this version:

Tony Rinaud, Clément Harmange, Olivier Pays, Mathieu Sarasa, Maxime Saillard, et al.. Interspecific competition between two partridges in farmland landscapes. Animal Behaviour, 2020, 165, pp.23-34. 10.1016/j.anbehav.2020.04.018. hal-02768595

HAL Id: hal-02768595 https://hal.science/hal-02768595v1

Submitted on 3 Jun 2022

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



1 Interspecific competition between two partridges in farmland landscapes 2 Tony Rinaud*1, Clément Harmange*1,2, Olivier Pays2, Mathieu Sarasa3,4, 3 4 Maxime Saillard¹ and Vincent Bretagnolle^{1,5} 5 6 ¹Centre d'Etudes Biologiques de Chizé, UMR 7372, CNRS and Université La Rochelle, F-79360 7 Beauvoir-sur-Niort, France 8 ²LETG-Angers, UMR 6554 CNRS, Université d'Angers, 49000 Angers, France 9 ³BEOPS, 1 Esplanade Compans Caffarelli, 31000 Toulouse, France (current address) 10 ⁴Fédération Nationale des Chasseurs, 13 rue du Général Leclerc, 92136 Issy-les-Moulineaux 11 cedex, France (former address) 12 ⁵ LTSER « Zone Atelier Plaine & Val de Sèvre », CNRS, 79360, Beauvoir sur Niort, France 13 *Co-first authors 14 Corresponding author: 15 Clément Harmange, LETG-Angers, UMR 6554 CNRS, Université d'Angers, 49000 Angers, 16 France, +33671918570, clement.harmange@orange.fr 17 18 **Acknowledgements** 19 We thank all the fieldworkers who contributed to part of the data collection. We also thank 20 Nicholas Aebischer, Elisabeth Bro, Jérome Moreau, Romain Julliard and Mathieu Boos for 21 their valuable comments on previous drafts. This work was supported by the Fédération 22 Nationale des Chasseurs (grant number FNC-PSN-PR15-2014) and the "Ressources alimentaires" project funded by the LISEA foundation. 23 24 Word count: 6800 words from abstract to references (excluding supplementary material)

3

4

5

6

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

22

23

1

ABSTRACT

Interspecific competition is expected to occur between phylogenetically closely related species when sharing resources. In birds, interspecific competition often occurs by songmediated interference, and frequently implies asymmetrical outcomes between the speciespairs involved. Habitat loss resulting from agricultural intensification is expected to have aggregated bird species in the remaining suitable habitats, thus increasing the likelihood of interspecific competition. However, this process has rarely been considered as a potential factor limiting population recovery in farmland birds. We investigated whether interspecific competition occurs between grey (Perdix perdix L. 1758) and red-legged (Alectoris rufa L. 1758) partridges, two phylogenetically related species. Originally parapatric, they have suffered an artificial increase in their contact zone due to huge human-mediated gamebird releases. We analyzed territorial behaviour through a playback stimuli experiment and investigated shifts in habitat niche in absence and presence of a hypothetical competitor. Results showed that the grey partridge appeared less territorial when co-occurring with the red-legged partridge and shifts its habitat niche away from that occupied by the red-legged partridge, while no such change was detected for the red-legged partridge. These asymmetrical patterns in behaviour and ecology are predicted under an interspecific competition scenario beneficial to the red-legged partridge, and therefore suggest that redlegged partridges are competitively dominant to grey partridges where they co-occur. This result has potentially strong implications for the management of grey partridges as gamebirds, and for their conservation in areas where they are almost extirpated.

25	Key words

- 26 Asymmetric competition, Grey partridge, Habitat niche displacement, Playback experiment,
- 27 Red-legged partridge, Territorial behaviour.

INTRODUCTION

30

31

32

33

34

35

36

37

38

39

40

41

42

43

44

45

46

47

48

49

50

51

52

53

Interspecific competition is a major driver of trophic interactions (MacArthur & Levins, 1964), and may define ecological niches (Diamond, 1978; Martin & Martin, 2001b), determine biogeographic patterns (Bull, 1991; Jankowski et al., 2010) and affect evolutionary processes (Svärdson, 1949; Wiens, 1992). Interspecific competition is expected to occur between phylogenetically close species sharing limiting resources (Wiens, 1992), either indirectly through resource depletion (Dhondt, 2012; Schoener, 1983), or directly through agonistic interactions (Persson, 1985; Wiens, 1992). Agonistic interactions are in most cases asymmetric in vertebrates (Martin et al., 2017; Pasch et al., 2013; Persson, 1985; Schoener, 1983): in an analysis of phylogenetically close bird species, 224 out of 270 speciespairs showed asymmetric agonistic interactions, benefiting the larger species in 87% of the cases (Martin et al., 2017). Typically, larger and dominant species show little difference in territorial response (vocal behaviour and approaches) to heterospecific versus conspecific songs, while smaller (subordinate) species show lower response rates to the heterospecific versus conspecific song (Martin & Martin, 2001a; McEntee, 2014). This may result in a shift in the song schedule to avoid territorial interaction (Brumm, 2006; Cody & Brown, 1969) and in the displacement and/or contraction of the ecological niche of the subordinate species (Jankowski et al., 2010; Martin & Martin, 2001b; Svärdson, 1949; Tarjuelo et al., 2017). Such asymmetric outcomes may impact population dynamics of subordinate species, by constraining its distribution to lower quality habitats, resulting in lower breeding success and/or survival (Martin et al., 2017; Martin & Martin, 2001b).

Recent declines in rural landscape habitat quality, the result of post-war changes in agricultural practices, has not only triggered a decline in European farmland birds (Benton et al., 2003; Newton, 2004; Robinson & Sutherland, 2002), but has also aggregated surviving

bird populations into the remaining local patches of suitable habitat (Aebischer & Ewald, 2012). While an increasing likelihood of interspecific competition should be expected in those habitats (Newton, 2004; Robillard et al., 2013), the implications for farmland bird decline or population recovery limitation has remained largely unexplored (though see Robillard et al., 2013). The grey partridge Perdix perdix L. 1758 and red-legged partridge Alectoris rufa L. 1758 are phylogenetically related and belong to the Phasianidae (Wang et al., 2013). They were originally parapatric, showing separated but contiguous distributions (Cramp & Simmons, 1980; see also Watson, 1962; Bull, 1991), with the grey originating from central Europe and Asia (Liukkonen-Anttila et al., 2002) and the red-legged from Iberia (Ferrero et al., 2011). Originally, the distributional contact zone was primarily restricted to central and southern France and north-west Italy (Cramp & Simmons, 1980). However, huge releases of red-legged partridges as gamebirds in the UK and France (Aebischer, 2019; Cramp & Simmons, 1980) resulted in the current situation where both species are sympatric over vast areas of western Europe. Both species select open agricultural landscapes (Aebischer & Kavanagh, 1997; Aebischer & Lucio, 1997; Potts, 1980), have similar breeding habitats (Bro et al., 2013; Green et al., 1987; Rands, 1988), though slight differences in micro-habitat exist (Meriggi et al., 1991). They also overlap in their breeding phenology (egg laying in May-June, peak of territorial activity in February-April; Rotella & Ratti, 1988). In addition, chicks feed on the same arthropod taxa, while adults shift to leaves and seeds (Bro & Ponce-Boutin, 2004; Green, 1984; Green et al., 1987). Finally, partridges are highly territorial, with both mates contributing to territory defense (Casas et al., 2016; Potts, 1980). The original narrow contact zone between both species was suggestive of the presence of competition interactions (Bull, 1991; Watson, 1962). These can now expected to be present

54

55

56

57

58

59

60

61

62

63

64

65

66

67

68

69

70

71

72

73

74

75

over the entire, vast, contact zone (Jankowski et al., 2010; Martin & Martin, 2001a; McEntee, 2014).

77

78

79

80

81

82

83

84

85

86

87

88

89

90

91

92

93

94

95

96

97

98

99

100

The possible increased competition pressure from the introduction of the larger and bulkier red-legged partridge in the human-created contact zone (Carpio et al., 2017; Robertson, 1996; Sokos et al., 2008), has rarely been studied as a possible cause of grey partridge decline or as a limiting factor in population recovery (Bro, 2016; Meriggi et al., 1991). However, while the grey partridge is one the most iconic farmland bird in Europe (Sotherton et al., 2014), it has declined strongly over the last few decades (Aebischer & Kavanagh, 1997; Aebischer & Potts, 1994) and is now threatened at national or sub-national scales (Charra & Sarasa, 2018; Eaton et al., 2015). In this study, we investigated behaviours that might potentially indicate interspecific competition between the two partridges, in study areas located within the original area of sympatry (central western France). We quantified territorial behaviour through playback stimuli, analyzing the patterns of territorial response in single- and two-species situations (Cody & Brown, 1969; Martin & Martin, 2001a), and further investigated behavioural (territorial daily activity) and habitat niche shifts in the area of co-occurrence. Based on the asymmetric interspecific competition hypothesis, we predicted that grey partridge (smaller species) should respond less to redlegged partridge than to the conspecific call both in the absence and co-occurrence of redlegged partridge, and should show lower levels of agonistic behaviour in co-occurrence areas. Conversely, red-legged partridge should increase its territorial response to grey partridge in co-occurrence situations, while maintaining or increasing its agonistic behaviour. We also predicted that grey partridge should restrict their daily activity pattern when cooccurring with red-legged partridge to minimize inter-specific encounters. Finally, habitat niche of the putatively subordinate grey partridge should shift when in co-occurrence with

red-legged partridges, while the putatively dominant red-legged partridge should be unaffected by the presence of grey partridge.

MATERIAL & METHODS

Study area

The study was conducted in the Long-Term Socio-Ecological Research platform (LTSER) "Zone Atelier Plaine & Val de Sèvre" (hereafter, ZAPVS), Central Western France, (Fig. 1; Bretagnolle et al., 2018b), in 2016 and 2017. This is a 435 km² zone of intensive agricultural cultivation, comprising winter cereals (41.5%), sunflower (10.4%), maize (9.6%), rape (8.3%), meadows (13.5%), woodlands (2.9%) and built-up areas (9.8%; average values 2009-2016 in Bretagnolle et al., 2018b). Within the ZAPVS, a restricted study site of 34 km² (2016) to 68 km² (2017) was chosen where both partridge species were historically the most observed during the past two decades (Fig. 1). In 2017, a second study site, with higher partridge densities than the ZAPVS, was added to increase the sightings dataset (Table A1, Appendix A). To maintain homogeneity in birds' phenology, this second site is located close to the ZAPVS and has similar habitat types. This 18 km² site is located in the south of Maine-et-Loire (METL; Fig. 1).

Survey protocol and playback tests

We first established a grid of playback stations with a cell size of 500 m (excluding built-up areas) to locate territorial pairs accurately (the size of a single home range is about 20 ha in spring-summer: Birkan et al., 1992). Sample points were placed as close as possible to the centroid of the grid cells, but still on a path or a road for ease of access and minimize disturbance. Each sample point was surveyed four times per year, with two sessions in the

morning and two in the evening, as patterns of daily vocal activity peak at dawn and dusk (Pépin & Fouquet, 1992). Sample points were slowly approached by car, the observer then went out quietly and waited for 1 minute before starting a session. A session consisted of broadcasting, at fixed volume (calibrated to be heard up to 250 meters away to avoid overlap between neighboring points) with a NEWONLINE N74® speaker, of an audio sequence containing territorial calls of males of both study species (155 seconds for each species), and of the common quail Coturnix coturnix L. 1758 call as control. For each species, bouts were separated by 1-min silence (see Appendix B for more details about sequence preparation and call origins). Since calls of three species were broadcast, six audio sequences containing the six possible broadcasting combinations were prepared and one of them was randomly played at each session. In 2016, surveyed were conducted at 140 playback stations at ZAPVS (April 6 to April 25), while in 2017, 275 stations were surveyed at ZAPVS and 75 at METL, from March 14 to April 28. Surveys involved tracks of 14-17 sampling points along a road/path, with each track corresponding to a sequence of points made in the morning or evening session. Tracks were made alternately in a way, then in the other, to avoid hourly biases in the sequence of points.

125

126

127

128

129

130

131

132

133

134

135

136

137

138

139

140

141

142

143

144

145

146

147

148

For each point in a given year, we determined whether none, one or the two species were present, if contacted at least once over the four sampling sessions. Points were then classified into four distinct categories as to whether none, only one or both partridges were contacted. The "no contact" category was discarded, so that only points with at least one occurrence were kept for analyses, assuming that a species was absent of a point if no sighting was recorded there across the four sampling sessions. We checked whether discarded points might actually contain partridges using rarefaction curve analyses (see Appendix C): only 7.3% and 12.5% of points for grey and red-legged partridges, respectively,

may have been removed erroneously. For each individual contacted at a point, we noted the playback call that elicited the response, the type of response (0: no response, 1: response to one of the partridge species calls, 2: response to both partridge stimuli), the latency of response (elapsed time between the beginning of the playback and any territorial response) and territorial behaviour (walking approach, flight approach, calling).

Ecological habitat description

Analysis was based on three datasets of grey and red-legged partridge occurrence collected by fieldworkers at ZAPVS between 2013 and 2017 (see Harmange et al., 2019). The first dataset contained occurrences from systematic bird point counts without playback (see Brodier et al., 2014 for protocols), and sampled some 450 points spread throughout the ZAPVS each year during the passerine breeding period (Bretagnolle et al., 2018a). The second dataset comprised of opportunistic sightings made when fieldworkers saw partridges while studying other bird species (Bretagnolle et al., 2018b), or during the systematic land cover monitoring performed twice a year over the entire study site (see Harmange et al., 2019). Each year, opportunistic monitoring occurred daily from late March to late July. The third dataset contained partridge sightings from the playback experiment (see above). Collating the three datasets, a total of 286 and 244 locations had grey and red-legged partridges, respectively.

From the annually collected land cover data, we then created a grid of $600 \times 600 \text{ m}^2$ cells, using R (R Core Team, 2017) and QGIS (Quantum GIS Development Team, 2017), excluding all pixels that had less than half their area (18 ha) lying within the study site. Pixel size was chosen as the minimal size that contained sufficient numbers of pixels with (i) only grey partridge occurrences over the period 2013-2017 (n = 115), (ii) only red-legged

partridge occurrences (n = 137), or (iii) both partridge species (n = 32). Analyzed landscape metrics were based on those identified as being selected (either positively or negatively) in grey and/or red-legged partridges (Birkan et al., 1992; Bro et al., 2013; Harmange et al., 2019; Meriggi et al., 1991; Rands, 1987): cereals, rape, meadows, spring crops (corn, sunflower etc.), hedgerows, roads/tracks, woodlands and buildings. Landscape metrics used in analyses corresponded to the surface area (or total length in m) of each habitat had in the pixel, divided by the pixel surface area (to account for edge effect). Each parameter was calculated annually from 2013 to 2017, except for roads and hedgerows, which remained stable and for which we set only one layer (Harmange et al., 2019).

Ethical note

This study is restricted to behavioural observations of partridges and, therefore, excludes any animal handling or invasive experiments. Field studies did not involve endangered or protected species. Farmers were made aware of the study, which did not involve any penetration to private properties. The study thus adheres to the "Guidelines for the Use of Animals in Research", and to the legal requirements of the country in which the work has been carried out. Our project was conducted with clearance from the authority *Préfecture Départementale des Deux-Sèvres* (Number: 2017/1).

- Statistical analyses
- 193 Territorial behaviour

GLMMs (binomial response variable, logit link) were used to investigate presence/absence of each species' territorial response (vocalization and/or approach) using the *lme4* R package (Bates et al., 2018), as a function of the interaction between playback stimuli (*broadcast* species) and the interspecific situation (target species alone or in co-occurrence with the other species). The playback sequence (6 levels) and the rank of partridge call (1st or 2nd position) were considered as random factors. When needed, Tukey's multiple comparison tests were performed using with the R package emmeans (Lenth et al., 2019). For each species, we further analyzed the magnitude of territorial response, described by four parameters calculated for each point count station. These included territorial response rate (number of responding individuals within pairs, 0/1/2), response latency (in seconds), calling behaviour (0: no response, 1: response to one of the two partridge species calls, 2: response to both partridge stimuli), and a territoriality score in the form of a ranking of increasing territorial behaviour intensity (0: no response, 1: approach, 2: vocalization, 3: approach with vocalization, 4: flying approach with vocalization; see Table A2 in Appendix A for more details about variables calculation). Then, for each species, we carried out linear discriminant analyses (LDA) (James & McCulloch, 1990) to investigate whether territorial behaviour (described from the four behavioural variables) allowed discrimination between the two interspecific situations, i.e. when a single species was observed and when both species cooccurred. LDA allowed the difference between two situations (alone or co-occurrence) to be maximized statistically for each species using a discrimant function assessed from four observed territorial behaviours. This analysis was performed using the MASS R package (Ripley, 2017). To handle the differences in sample sizes between the two situations (alone or co-occurrence) given the low sample size of the "co-occurrence" category, we produced 95% confidence intervals (CI) for linear discriminant scores following Luttrell & Lohr (2018). CI was assessed by generating bootstrap samples with replacement and simulating 1000 sets for the "alone" category (highest sample size) with reduced sample size equal to the smaller "co-occurrence" set. The mean of the linear discriminant scores +/- SD for each of the 1000

197

198

199

200

201

202

203

204

205

206

207

208

209

210

211

212

213

214

215

216

217

218

219

bootstrapped samples was calculated and a 95% CI was computed for the "alone" category. Finally, we carried out analyses of variance (ANOVA) to test whether each territorial behaviour (see above) contributes significantly to discriminate the two interspecific situations (species alone vs co-occurrence).

To investigate the daily pattern of territorial behaviour for each species, we first summed the number of territorial responses (vocal behaviour and approaches) for a 30-minute period, using four categories: grey partridge alone and grey co-occurring with red-legged partridge, and red-legged partridge alone and red-legged co-occurring with grey partridge. The sum (per half-hour) was then standardised (divided) by the number of trials performed on that given 30-minute period. Seasonal effects were controlled for by considering time (GMT) relative to sunrise (for sampling points in the morning) or sunset (evening). To compare the distribution of territorial responses across the four modalities, we performed Cramér-von Mises distribution tests (Anderson, 1962; Conover, 1999), using the *RVAidememoire* R package (Hervé, 2018).

Habitat niche modelling

Within-species habitat niches were analyzed with data from 2013-2017, averaging environmental layers across years for each pixel. Three types of occurrences (2013-2017) were considered: (1) occurrence of grey partridge only (at least once) in the pixel over the 5 years (n = 222); (2) occurrence of red-legged partridge only (n = 192); (3) grey (n = 64) and red-legged (n = 52) partridges co-occurring, each being recorded at least once (but not necessarily simultaneously). To compare species niches in presence of a single species versus co-occurrence of the two species, we used Broennimann et al. (2012)'s niche modelling framework implemented in the *ecospat* R package (Broennimann et al., 2016). The first step

consists of describing the environmental space available in the study area, using the first two axes of a principal component analysis (PCA-env) performed on all pixels, with each environmental layer being a different variable in the PCA. Then, for each species in each situation (species alone or in co-occurrence), a smoothed density of occurrences in the PCAenv space is obtained using a kernel density function. This allows niche space to be graphically delineated, for each situation, as a relative abundance index ranging from 0 (environmental conditions in which the species was not observed) to 1 (environmental conditions in which the species was most commonly observed), thus handling betweengroup differences in the number of occurrences. Niche overlap can then be calculated (e.g. between grey alone and grey co-occurring with red-legged partridge), using Schoener's D index (Warren et al., 2008) which ranges from 0 (no overlap) to 1 (full overlap). The niche equivalency test, a commonly-used niche randomization test, was performed (Broennimann et al., 2012; Warren et al., 2008). In this test, the two occurrence datasets are mixed, and two new datasets are reconstituted randomly (1000 times), conserving the sizes of the two original datasets. The distribution of the 1000 D niche overlap values from the paired simulated datasets are then compared with the observed value of D: if the latter is outside the 95% range for the simulated values, the null hypothesis is rejected, i.e. the observed niches are less equivalent than expected by chance (Broennimann et al., 2012; Warren et al., 2008). All statistical analyzes were performed using R software (R Core Team, 2017).

264

265

266

245

246

247

248

249

250

251

252

253

254

255

256

257

258

259

260

261

262

263

RESULTS

- Territorial behaviour
- Among the surveyed points (combining both study sites), there were 21 and 106 points (respectively 2016 and 2017) with at least one species in one of the four sampling sessions.

Assuming that the presence of an individual in at least one sampling session for a given point count reflected the presence of a territorial pair, 37 points (7.5%) with grey partridge, 74 (15.1%) with red-legged partridge and 16 (3.3%) with both species were recorded. We found that grey partridge territorial responses were significantly influenced both by the broadcast species and whether red-legged partridge co-occurred (Table 1): they responded more to the conspecific than heterospecific call or to the control call (Fig; 2a, Table A3), but co-occurrence with red-legged partridge significantly reduced grey partridge response (Fig. 2a, Table A3). Similarly, territorial responses of red-legged partridge, when alone, were higher to the conspecific call than the grey partridge or control call (Fig. 2b, Tables 1, A3). Unlike the grey partridge, the red-legged demonstrated a balanced response whichever the species when co-occurring with grey partridge (Fig. 2b, Table A3).

Scores of territorial behaviour in response to playback (from the four behavioural variables) allowed discrimination between situations when grey partridges were alone or co-occurring with red-legged partridges. Territorial response rate, calling behaviour and territoriality score were the three highest contributing variables discriminating territorial behaviour between the two situations (Fig. 3b). In situations of co-occurrence the grey partridge reduced its response rate, number of calls and intensity of territorial response (i.e. territorial score) significantly (ANOVA, Table A4). These results suggest a shift in the territorial behaviour of the grey partridge in co-occurrence situations (Fig. 3a). Conversely, behavioural variables were very poor contributors in discriminating situations where red-legged were alone or co-occurring with grey partridges (Fig. 3d). Indeed, no significant differences appeared for any of the four behavioural variables in the red-legged partridge responses (ANOVA, Table A4), indicating no shift in behaviour similar to that found for the grey partridge (Fig. 3c).

Daily pattern of grey partridge territorial behaviour was significantly affected by cooccurrence with red-legged partridge (Fig. 4, Table A5, Appendix A): under co-occurrence,
grey partridge territorial response became concentrated around the sunset in the evening,
while no changes were detected in the morning (Fig. 4). A similar pattern was detected for
the red-legged partridge where alone versus in co-occurrence with the grey partridge (Fig. 4,
Table A5).

Habitat niche

Analysis showed that the grey partridge's habitat niches when alone or when co-occurring with the red-legged partridge overlapped by only 53%. The equivalency test was significant (p = 0.02), indicating that the grey partridge habitat niche significantly shifted in the presence of the red-legged partridge. Although the Figure 5 suggests a moderate shift (Fig. 5a compared to 5b), the core of the grey partridge's habitat niche changed in the opposite direction of the red-legged partridge's habitat niche, towards greater proportions of spring crop, cereal and rape habitats, and lower proportions of meadows and hedgerows (Fig. 5e). Overall, the red-legged partridge's habitat niche was wider, comprised of its preferred open arable land (e.g. spring crops, cereals and rape) associated with semi-natural permanent covers (meadows and hedgerows), but the species also avoided closed habitats (woodlands and built-up areas, Fig. 5e). In contrast to the grey partridge, the red-legged partridge's habitat niches when alone or when co-occurring with the grey partridge showed more overlap (67%, Fig. 5c, d). When co-occurring with the grey partridge, the red-legged's habitat niche did not shift, so remaining equivalent in both situations (equivalency test, p = 0.55).

DISCUSSION

As expected, our results suggest an asymmetric pattern of territorial behaviour between grey and red-legged partridges. Grey partridges reduced their territorial response, and shifted both their hourly territorial behaviour and their habitat niche when co-occurring with the red-legged partridge.

317

318

319

320

321

322

323

324

325

326

327

328

329

330

331

332

333

334

335

336

337

338

339

340

Such shifts in territorial behaviour intensity towards a more discreet behaviour when facing a competitive species have been documented previously in birds (e.g. Martin & Martin, 2001a; McEntee, 2014) and other vertebrates, such as mammals (e.g. Pasch et al., 2013). Indeed, in such cases, the typical response of the subordinate species is to avoid unfriendly interactions with a dominant and larger competitor. Conversely, the red-legged partridge territorial response rate showed much higher response rate to its own species when alone, but shifted to balanced response rates to both species in situations of cooccurrence. Such convergence of territorial behaviour has already been documented for dominant species involved in asymmetric interspecific interactions (Jankowski et al., 2010; Pasch et al., 2013) and reject the hypothesis of Murray (1971), which suggests that interspecific territoriality results from misdirected intraspecific interactions due to incorrect signal recognition. However, huge releases of captive-reared partridges for game management could alternatively explain responses to heterospecific call, as both partridge species are often reared in the same farms and may have been sensitized or conditioned by a precocious exposure to the heterospecific call. We suggest that this alternative hypothesis is however unlikely because under the hypothesis of a precocious sensitization in rearing farms, both species should have responded to heterospecific calls (not just to the red-legged partridge), and additionally, each species should have responded to the heterospecific stimuli, whether in absence of competitors or in co-occurring situations. In addition, we found that both species adjusted their daily territorial activity under co-occurrence, though

territorial activity of both species remained synchronized, while a temporal segregation was expected. This may be the result of a too coarse temporal resolution in our analyses, making temporal shifts in territorial response hard to detect. For instance, Brumm (2006) showed the temporal segregation in territorial activity occurs at a finer scale in Nightingale *Luscinia megarhynchos*, which adjusts and focuses its vocal activity within the silent intervals between two heterospecific broadcast songs (see also Cody & Brown, 1969). An alternative explanation would be that given the calls of both partridges are acoustically different, this might have enabled vocal activity to remain synchronized, while ensuring a low interspecific interference (Dhondt, 2012; Doutrelant & Lambrechts, 2001).

Finally, we detected a slight habitat niche shift in the less competitive species in situations of co-occurrence with the dominant one, while there was no such shift for the dominant species. This pattern is typical of asymmetric competition for habitat likely resulting from the agonistic behaviour of the dominant species (Jankowski et al., 2010; Martin & Martin, 2001a, 2001b). While the grey partridge (subordinate species) selects preferentially grass along hedgerows to nest (Potts & Aebischer, 1991), its habitat niche shifted to more open habitats (more annual crops and less hedgerows) in co-occurrence situations with the red-legged. These habitats that are more subject to human disturbance, may therefore be of poorer-quality for breeding (Burel et al., 1998), likely resulting in lower fitness and/or survival (see Martin et al., 2017; Martin & Martin, 2001a, 2001b). Such habitat niche shift, likely resulting from interspecific competition, could have also driven the microhabitat segregation observed by Meriggi et al. (1991), with the grey partridge being generally more associated with crops and open lands, while the red-legged is more linked to semi-natural elements like hedgerows.

Asymmetric competition occurring in the contact zone of parapatric species is wellknown to reduce fitness of the subordinate species (Jankowski et al., 2010; Martin & Martin, 2001a). Such behavioural outcomes of competition may have strong implications for the conservation and management of partridges. Although releasing non-native species has long been used as a strategy to limit hunting pressure on wild native populations (Aebischer & Ewald, 2004; Bro, 2016; Carpio et al., 2017), the huge releases of red-legged partridges in the core area of the grey partridge raise concerns not only for the grey partridge decline, but also for its future possible recovery. The presence of morphologically similar species has long been identified as interfering with the re-establishment success in translocations (Griffith et al., 1989). The grey partridge is currently a focus of interest for many reinforcement projects, that are identified as the only tool for restoring self-sustaining populations in many areas (Buner et al., 2011; Buner & Aebischer, 2008). However, this has proved inefficient or even counterproductive (Reitz, 2003; Sokos et al., 2008), because of many unintentional but adverse effects on population dynamics (inadvertent shots: in Aebischer & Ewald, 2004; parasite-mediated competition: in Tompkins, Greenman, et al., 2000; Tompkins, Draycott, et al., 2000). Our results suggest further impacts on territorial behaviour and access to resources for the native species, though the effects on demography remain unexplored. Our study thus suggests that the presence of red-legged partridges may limit the chances of success in such projects. Therefore, to quantify the strength of competitive interaction effects on population dynamics, further field experiments are required to formally conclude that interspecific competition occurs between these species. These should include (i) controlling for the origin of birds (natural or captive birds reared in monospecific farms preventing any early contact between both species), (ii) treatments (deliberately releasing grey partridge only, red-legged only or both species together, in different sites offering the

364

365

366

367

368

369

370

371

372

373

374

375

376

377

378

379

380

381

382

383

384

385

386

same availability in environmental conditions) and (iii) assessing the influence of species cooccurrence on demographic parameters (adult survival, nest predation, chick survival, see Martin & Martin, 2001b).

391

392

393

394

395

396

397

398

399

400

401

402

403

404

405

406

407

408

409

410

BirdLife International.

388

389

390

REFERENCES

Aebischer, N. J. (2019). Fifty-year trends in UK hunting bags of birds and mammals, and calibrated estimation of national bag size, using GWCT's National Gamebag Census. European Journal of Wildlife Research, 65(4), 64. https://doi.org/10.1007/s10344-019-1299-x Aebischer, N. J., & Ewald, J. A. (2004). Managing the UK Grey Partridge *Perdix perdix* recovery: population change, reproduction, habitat and shooting. Ibis, 146(s2), 181-191. Aebischer, N. J., & Ewald, J. A. (2012). The grey partridge in the UK: population status, research, policy and prospects. *Animal Biodiversity and Conservation*, 35(2), 353–362. Aebischer, N. J., & Kavanagh, B. (1997). Grey partridge. In W. J. M. Hagemeijer, & M. J. Blair (Eds.), The EBCC Atlas of European breeding birds, their distribution and abundance (pp. 212-213). London, UK: T&AD Poyser. Aebischer, N. J., & Lucio, A. (1997). Red-legged partridge. In W. J. M. Hagemeijer, & M. J. Blair (Eds.), The EBCC Atlas of European breeding birds, their distribution and abundance (pp. 209–208). London, UK: T&AD Poyser. Aebischer, N. J., & Potts, G. R. (1994). Partridge Perdix perdix. In M. F. Tucker, & M. F. Heath

(Eds.), Birds in Europe, their conservation status (Vol. 3, pp. 220–221). Cambridge, UK:

- 411 Anderson, T. W. (1962). On the Distribution of the Two-Sample Cramér-von Mises Criterion.
- The Annals of Mathematical Statistics, 33(3), 1148-1159. JSTOR.
- Bates, D., Maechler, M., Bolker, B., Walker, S., & Christensen, R. H. B. (2018). Ime4: Linear
- 414 mixed-effects models using Eigen and S4. *R package version*.
- 415 Benton, T. G., Vickery, J. A., & Wilson, J. D. (2003). Farmland biodiversity: is habitat
- heterogeneity the key? *Trends in Ecology & Evolution, 18*(4), 182-188.
- 417 https://doi.org/10.1016/S0169-5347(03)00011-9
- 418 Birkan, M., Serre, D., Skibnienski, S., & Pelard, E. (1992). Spring-summer home range, habitat
- 419 use and survival of grey partridge (*Perdix perdix*) in a semi-open habitat. *Gibier Faune*
- 420 *Sauvage*, *9*, 431-442.
- 421 Bretagnolle, V., Berthet, E., Gross, N., Gauffre, B., Plumejeaud, C., Houte, S., Badenhausser,
- 422 I., Monceau, K., Allier, F., & Monestiez, P. (2018a). Description of long-term
- 423 monitoring of farmland biodiversity in a LTSER. *Data in Brief*, 19, 1310-1313.
- 424 Bretagnolle, V., Berthet, E., Gross, N., Gauffre, B., Plumejeaud, C., Houte, S., Badenhausser,
- 425 I., Monceau, K., Allier, F., & Monestiez, P. (2018b). Towards sustainable and
- 426 multifunctional agriculture in farmland landscapes: Lessons from the integrative
- 427 approach of a French LTSER platform. Science of The Total Environment, 627, 822–
- 428 834.
- 429 Bro, E. (2016). La Perdrix grise. Biologie, écologie, gestion et conservation (p. 304). Mèze,
- 430 France: Biotope.
- 431 Bro, E., Millot, F., Delorme, R., Polvé, C., Mangin, E., Godard, A., Tardif, F., Gouache, C., Sion,
- 432 I., & Brault, X. (2013). PeGASE, bilan synthétique d'une étude perdrix grise
- 433 «population-environnement». Faune Sauvage, 298, 17–48.

434	Bro, E., & Ponce-Boutin, F. (2004). Régime alimentaire des Phasianidés en plaine de grandes
435	cultures. Faune Sauvage, 263, 5-13.
436	Brodier, S., Augiron, S., Cornulier, T., & Bretagnolle, V. (2014). Local improvement of skylark
437	and corn bunting population trends on intensive arable landscape: a case study of
438	the conservation tool Natura 2000. Animal Conservation, 17(3), 204–216.
439	Broennimann, O., Di Cola, V., & Guisan, A. (2016). ecospat: Spatial ecology miscellaneous
440	methods. R package version 2.1. 1.
441	Broennimann, O., Fitzpatrick, M. C., Pearman, P. B., Petitpierre, B., Pellissier, L., Yoccoz, N.
442	G., Thuiller, W., Fortin, MJ., Randin, C., & Zimmermann, N. E. (2012). Measuring
443	ecological niche overlap from occurrence and spatial environmental data. Global
444	ecology and biogeography, 21(4), 481–497.
445	Brumm, H. (2006). Signalling through acoustic windows: nightingales avoid interspecific
446	competition by short-term adjustment of song timing. Journal of Comparative
447	Physiology A, 192(12), 1279-1285. https://doi.org/10.1007/s00359-006-0158-x
448	Bull, C. M. (1991). Ecology of parapatric distributions. Annual review of ecology and
449	systematics, 22(1), 19–36.
450	Buner, F., & Aebischer, N. J. (2008). Guidelines for re-establishing grey partridges through
451	releasing. Fordingbridge, UK: Game and Wildlife Conservation Trust.
452	Buner, F., Browne, S. J., & Aebischer, N. J. (2011). Experimental assessment of release
453	methods for the re-establishment of a red-listed galliform, the grey partridge (Perdix
454	perdix). Biological Conservation, 144(1), 593-601.
455	https://doi.org/10.1016/j.biocon.2010.10.017
456	Burel, F., Baudry, J., Butet, A., Clergeau, P., Delettre, Y., Le Coeur, D., Dubs, F., Morvan, N.,
457	Paillat, G., Petit, S., Thenail, C., Brunel, E., & Lefeuvre, JC. (1998). Comparative

458	biodiversity along a gradient of agricultural landscapes. Acta oecologica, 19(1), 47-
459	60.
460	Carpio, A. J., Guerrero-Casado, J., Barasona, J. A., Tortosa, F. S., Vicente, J., Hillström, L., &
461	Delibes-Mateos, M. (2017). Hunting as a source of alien species: a European review.
462	Biological Invasions, 19(4), 1197–1211.
463	Casas, F., Arroyo, B., Viñuela, J., Guzmán, J. L., & Mougeot, F. (2016). Are farm-reared red-
464	legged partridge releases increasing hunting pressure on wild breeding partridges in
465	central Spain? European journal of wildlife research, 62(1), 79–84.
466	Charra, M., & Sarasa, M. (2018). Applying IUCN Red List criteria to birds at different
467	geographical scales: similarities and differences. Animal Biodiversity and
468	Conservation, 41(1), 75–95.
469	Cody, M. L., & Brown, J. H. (1969). Song asynchrony in neighbouring bird species. <i>Nature</i> ,
470	<i>222</i> (5195), 778-780.
471	Conover, W. J. (1999). Practical Nonparametric Statistics. New-York, US: John Wiley & Sons,
472	Inc.
473	Cramp, S., & Simmons, K. E. L. (1980). Birds of Europe, the Middle East and North Africa (Vol.
474	2). Oxford, UK: Oxford University Press.
475	Dhondt, A. A. (2012). Interspecific competition in birds (Vol. 2). Oxford, UK: Oxford University
476	Press.
477	Diamond, J. M. (1978). Niche Shifts and the Rediscovery of Interspecific Competition: Why
478	did field biologists so long overlook the widespread evidence for interspecific
479	competition that had already impressed Darwin? American scientist, 66(3), 322–331.
480	Doutrelant, C., & Lambrechts, M. M. (2001). Macrogeographic variation in song—a test of
481	competition and habitat effects in blue tits. Ethology, 107(6), 533-544.

482	Eaton, M. A., Aebischer, N. J., Brown, A., Hearn, R., Lock, L., Musgrove, A., Noble, D., Stroud
483	D., & Gregory, R. (2015). Birds of Conservation Concern 4: the population status of
484	birds in the UK, Channel Islands and Isle of Man. British Birds, 108, 708–746.
485	Ferrero, M. E., Blanco-Aguiar, J. A., Lougheed, S. C., Sánchez-Barbudo, I., Nova, P. J. G. D.,
486	Villafuerte, R., & Dávila, J. A. (2011). Phylogeography and genetic structure of the
487	red-legged partridge (Alectoris rufa): more evidence for refugia within the Iberian
488	glacial refugium. Molecular Ecology, 20(12), 2628-2642.
489	https://doi.org/10.1111/j.1365-294X.2011.05111.x
490	Green, R. E. (1984). The Feeding Ecology and Survival of Partridge Chicks (Alectoris rufa and
491	Perdix perdix) on Arable Farmland in East Anglia. Journal of Applied Ecology, 21(3),
492	817-830. https://doi.org/10.2307/2405049
493	Green, R. E., Rands, M. R. W., & Moreby, S. J. (1987). Species differences in diet and the
494	development of seed digestion in partridge chicks Perdix perdix and Alectoris rufa.
495	<i>Ibis, 129</i> (s2), 511–514.
496	Griffith, B., Scott, J. M., Carpenter, J. W., & Reed, C. (1989). Translocation as a species
497	conservation tool: status and strategy. Science, 245(4917), 477–480.
498	Hammer, Ø., Harper, D. A. T., & Ryan, P. D. (2001). PAST: Paleontological statistics software
499	package for education and data analysis. Palaeontologia Electronica, 4(1), 9.
500	Harmange, C., Bretagnolle, V., Sarasa, M., & Pays, O. (2019). Changes in habitat selection
501	patterns of the gray partridge Perdix perdix in relation to agricultural landscape
502	dynamics over the past two decades. Ecology and evolution, 9(9), 5236-5247.
503	https://doi.org/10.1002/ece3.5114
504	Hervé, M. (2018). RVAideMemoire: Testing and plotting procedures for biostatistics. R
505	package version 0.9-69-3.

506	James, F. C., & McCulloch, C. E. (1990). Multivariate analysis in ecology and systematics:
507	panacea or Pandora's box? Annual review of Ecology and Systematics, 21(1), 129–
508	166.
509	Jankowski, J. E., Robinson, S. K., & Levey, D. J. (2010). Squeezed at the top: interspecific
510	aggression may constrain elevational ranges in tropical birds. Ecology, 91(7), 1877–
511	1884.
512	Lenth, R., Singmann, H., Love, J., Buerkner, P., & Herve, M. (2019). Package « emmeans ». R
513	package version, 1(1).
514	Liukkonen-Anttila, T., Uimaniemi, L., Orell, M., & Lumme, J. (2002). Mitochondrial DNA
515	variation and the phylogeography of the grey partridge (Perdix perdix) in Europe:
516	from Pleistocene history to present day populations. Journal of Evolutionary Biology
517	15(6), 971-982. https://doi.org/10.1046/j.1420-9101.2002.00460.x
518	Luttrell, S. A., & Lohr, B. (2018). Geographic variation in call structure, likelihood, and call-
519	song associations across subspecies boundaries, migratory patterns, and habitat
520	types in the Marsh Wren (Cistothorus palustris). The Auk: Ornithological Advances,
521	<i>135</i> (1), 127–151.
522	MacArthur, R. H., & Levins, R. (1964). Competition, habitat selection, and character
523	displacement in a patchy environment. Proceedings of the National Academy of
524	Sciences of the United States of America, 51(6), 1207.
525	Martin, P. R., Freshwater, C., & Ghalambor, C. K. (2017). The outcomes of most aggressive
526	interactions among closely related bird species are asymmetric. <i>PeerJ</i> , 5, e2847.
527	Martin, P. R., & Martin, T. E. (2001a). Behavioral interactions between coexisting species:
528	song playback experiments with wood warblers. <i>Ecology</i> , 82(1), 207–218.

529	Martin, P. R., & Martin, T. E. (2001b). Ecological and Fitness Consequences of Species
530	Coexistence: A Removal Experiment with Wood Warblers. <i>Ecology</i> , 82(1), 189-206.
531	https://doi.org/10.2307/2680096
532	McEntee, J. P. (2014). Reciprocal territorial responses of parapatric African sunbirds: species-
533	level asymmetry and intraspecific geographic variation. Behavioral Ecology, 25(6),
534	1380–1394.
535	Meriggi, A., Montagna, D., & Zacchetti, D. (1991). Habitat use by partridges (<i>Perdix perdix</i>
536	and Alectoris rufa) in an area of northern Apennines, Italy. Italian Journal of Zoology,
537	<i>58</i> (1), 85–89.
538	Murray, B. G. (1971). The ecological consequences of interspecific territorial behavior in
539	birds. <i>Ecology</i> , <i>52</i> (3), 414–423.
540	Newton, I. (2004). The recent declines of farmland bird populations in Britain: an appraisal of
541	causal factors and conservation actions. <i>Ibis</i> , 146(4), 579–600.
542	Novoa, C. (1992). Validation of a spring density index for Pyrenean grey partridge, <i>Perdix</i>
543	perdix hispaniensis obtained with playbacks of recorded calls. Gibier Faune Sauvage,
544	9, 105-118.
545	Pasch, B., Bolker, B. M., & Phelps, S. M. (2013). Interspecific dominance via vocal interactions
546	mediates altitudinal zonation in Neotropical singing mice. The American Naturalist,
547	182(5), E161–E173.
548	Pépin, D., & Fouquet, M. (1992). Factors affecting the incidence of dawn calling in red-legged
549	and grey partridges. Behavioural processes, 26, 167–176.
550	Persson, L. (1985). Asymmetrical competition: are larger animals competitively superior? <i>The</i>
551	American Naturalist, 126(2), 261–266.

552	Potts, G. R. (1980). The effects of modern agriculture, nest predation and game management
553	on the population ecology of partridges (Perdix perdix and Alectoris rufa). Advances
554	in ecological research, 11, 1–79.
555	Potts, G. R., & Aebischer, N. J. (1991). Modelling the population dynamics of the grey
556	partridge: Conservation and management. In C. M. Perrins, J. D. Lebreton, & G. J. M.
557	Hirons (Eds.), Bird Population Studies: Relevance to Conservation and Management
558	(pp. 373-390). Oxford, UK: Oxford University Press.
559	Quantum GIS Development Team. (2017). Quantum GIS Geographic Information System.
560	Open Source Geospatial Foundation Project. http://qgis.osgeo.org
561	R Core Team. (2017). R: A language and environment for statistical computing. R Foundation
562	for Statistical Computing. https://www.R-project.org/.
563	Rands, M. R. W. (1987). Hedgerow management for the conservation of partridges <i>Perdix</i>
564	perdix and Alectoris rufa. Biological Conservation, 40(2), 127–139.
565	Rands, M. R. W. (1988). The effect of nest site selection on nest predation in grey partridge
566	Perdix perdix and red-legged partridge Alectoris rufa. Ornis Scandinavica, 19(1), 35-
567	40.
568	Reitz, F. (2003). Le statut communal de la perdrix grise et de la perdrix rouge en France:
569	résultats d'une enquête. Faune Sauvage, 258, 25–33.
570	Ripley, B. (2017). MASS: support functions and datasets for Venables and Ripley's MASS. R
571	package version.
572	Robertson, P. A. (1996). Naturalised introduced gamebirds in Britain. In J. S. Holmes, & J. R.
573	Simons (Eds.), The introduction and naturalisation of birds (pp. 63-69). London, UK:
574	HMSO.

575	Robillard, A., Garant, D., & Bélisle, M. (2013). The Swallow and the Sparrow: how agricultural				
576	intensification affects abundance, nest site selection and competitive interactions.				
577	Landscape Ecology, 28(2), 201-215. https://doi.org/10.1007/s10980-012-9828-y				
578	Robinson, R. A., & Sutherland, W. J. (2002). Post-war changes in arable farming and				
579	biodiversity in Great Britain. Journal of Applied Ecology, 39(1), 157-176.				
580	https://doi.org/10.1046/j.1365-2664.2002.00695.x				
581	Rotella, J. J., & Ratti, J. T. (1988). Seasonal variation in gray partridge vocal behavior. <i>Condor</i> ,				
582	90, 304–310.				
583	Schoener, T. W. (1983). Field experiments on interspecific competition. <i>The American</i>				
584	Naturalist, 122(2), 240–285.				
585	Sokos, C. K., Birtsas, P. K., & Tsachalidis, E. P. (2008). The aims of galliforms release and				
586	choice of techniques. Wildlife Biology, 14(4), 412–422.				
587	Sotherton, N. W., Aebischer, N. J., & Ewald, J. A. (2014). Research into action: grey partridge				
588	conservation as a case study. Journal of Applied Ecology, 51(1), 1-5.				
589	https://doi.org/10.1111/1365-2664.12162				
590	Svärdson, G. (1949). Competition and habitat selection in birds. <i>Oikos</i> , <i>1</i> (2), 157–174.				
591	Tarjuelo, R., Traba, J., Morales, M. B., & Morris, D. W. (2017). Isodars unveil asymmetric				
592	effects on habitat use caused by competition between two endangered species.				
593	Oikos, 126(1), 73–81.				
594	Tompkins, D. M., Draycott, R. a. H., & Hudson, P. J. (2000). Field evidence for apparent				
595	competition mediated via the shared parasites of two gamebird species. Ecology				
596	Letters, 3(1), 10-14. https://doi.org/10.1046/j.1461-0248.2000.00117.x				

597	Tompkins, D. M., Greenman, J. V., Robertson, P. A., & Hudson, P. J. (2000). The role of shared
598	parasites in the exclusion of wildlife hosts: Heterakis gallinarum in the ring-necked
599	pheasant and the grey partridge. Journal of Animal Ecology, 69(5), 829–840.
600	Wang, N., Kimball, R. T., Braun, E. L., Liang, B., & Zhang, Z. (2013). Assessing Phylogenetic
601	Relationships among Galliformes: A Multigene Phylogeny with Expanded Taxon
602	Sampling in Phasianidae. PLoS ONE, 8(5), e64312.
603	https://doi.org/10.1371/journal.pone.0064312
604	Warren, D. L., Glor, R. E., Turelli, M., & Funk, D. (2008). Environmental Niche Equivalency
605	versus Conservatism: Quantitative Approaches to Niche Evolution. Evolution, 62(11),
606	2868-2883. https://doi.org/10.1111/j.1558-5646.2008.00482.x
607	Watson, G. E. (1962). Sympatry in palearctic <i>Alectoris</i> partridges. <i>Evolution</i> , <i>16</i> (1), 11–19.
608	Wiens, J. A. (1992). The ecology of bird communities (Vol. 2). Cambridge, UK: Cambridge
609	University Press.

612	
613	

611

Species	Variable	χ²- value	df	P-value	Estimate±SE	Variance of random effects
Grey					(intercept: -2.867 ± 0.413)	Sequence: <0.001
partridge						Rank: 0.146
	Broadcast species x interspecific situation	1.455	2	0.483		
	Broadcast species	38.023	2	<0.001	(see Table A3 for multiple comparisons)	
	Interspecific situation	6.942	1	0.008	Co-occurrence: -0.691 ± 0.27	72
Red-legged partridge					(intercept: -2.736 ± 0.476)	Sequence: 0.041 Rank: 0.085
	Broadcast species x interspecific situation	8.060	2	0.018	(see Table A3)	
	Broadcast species	19.352	2	< 0.001	(see Table A3)	
	Interspecific situation	0.258	1	0.611		

Explanatory variables are the broadcast species (control, grey partridge, red-legged partridge) and the interspecific situation (Alone, Co-occurrence). Modalities in italics are considered as references. All models include the playback sequence (n=6) and the rank (n=2) of the partridge broadcast calls (grey partridge or red-legged partridge broadcast first) as random effect.

Table A1 Grey and red-legged partridge density estimations (mean ± SD) in the Long-Term Socio-Ecological Research site "Zone Atelier Plaine & Val de Sèvre" (LTSER ZAPVS) and the Maine-et-Loire study area (METL) in 2017.

Species	Density in the LTSER ZAPVS	Density in the METL study area
Grey partridge	0.25 (± 0.07)	0.49 (± 0.20)
Red-legged partridge	0.15 (± 0.09)	2.62 (± 0.54)

Estimations were derived from playback experiments of 2016 and 2017.

Table A2 Description of behavioural variables considered in the analysis of territorial behaviour, and predictions under a hypothetic asymmetric competition for the dominant and subordinate species.

Behavioural metric	Description	Prediction
Territorial response intensity	The sum of responding individuals (by approach and or vocalization) controlled by the number of expected responses (number of individuals seen on this point)	In co-occurrence areas, we expected a higher territorial response intensity for the dominant species, and a lower response intensity for the subordinate one (see Jankowski et al., 2010; Martin & Martin, 2001a)
Response latency	The mean response latency (i.e. the time spent between the beginning of the broadcast call and the vocal response) recorded over the 4 sampling sessions	In co-occurrence area, the response latency is expected to be lower for the dominant species, higher for the subordinate (see Jankowski et al., 2010; Martin & Martin, 2001a)
Calling behaviour	The mean number of calls responded to by individuals, between the conspecific and heterospecific partridge calls. Ranges from 0 (no individual responded to partridge calls over the four sampling sessions), to 2 (all individuals responded to the two partridge calls during each of the four sessions)	Expected to be high for a dominant species responding as well as conspecific than heterospecific call, while it is expected to be low for a subordinate species not responding to the heterospecific call (Martin & Martin, 2001a)
Territoriality score	The mean of the territoriality score obtained by each individual across the four sampling sessions, following the ranking of increasing territoriality: 0 – no response, 1 – approach, 2 – vocalization, 3 – approach with vocalization, 4 – flying approach with vocalization	expected for the subordinate species

Metrics were calculated for each species, from each point count.

Table A3 Tukey's multicomparison post-hoc tests computed after GLMMs (Link = binomial, logit) performed on the territorial response of grey and red-legged partridge species.

Species	Variable	Estimate	SE	z-ratio	p-value
Grey					
partridge					
	Grey partridge – Red-legged partridge	1.140	0.279	4.084	< 0.001
	Grey partridge – Common quail	1.748	0.337	5.192	< 0.001
	Red-legged partridge – Common quail	0.608	0.376	1.620	0.234
Red-legged					
partridge					
	Red-legged partridge, Alone – Grey partridge, Alone	1.073	0.253	4.245	< 0.001
	Red-legged partridge, Alone – Common quail, Alone	0.946	0.244	3.883	0.001
	Red-legged partridge, Alone – Grey partridge, Co-occurrence	0.327	0.339	0.967	0.922
	Red-legged partridge, Alone – Common quail, Co-occurrence	1.091	0.443	2.461	0.125
	Red-legged partridge, Alone – Red-legged partridge, Co-occurrence	0.780	0.394	1.979	0.334
	Red-legged partridge, Co-occurrence – Grey partridge, Alone	0.293	0.425	0.689	0.982
	Red-legged partridge, Co-occurrence – Common quail, Alone	0.166	0.420	0.395	0.999
	Red-legged partridge, Co-occurrence – Grey partridge, Co-occurrence	-0.453	0.480	-0.944	0.929
	Red-legged partridge, Co-occurrence – Common quail, Co-occurrence	0.311	0.559	0.556	0.993
	Grey partridge, Alone – Common quail, Alone	-0.127	0.290	-0.438	0.998
	Grey partridge, Alone – Common quail, Co-occurrence	0.018	0.470	0.038	1.000
	Grey partridge, Alone – Grey partridge, Co-occurrence	-0.746	0.374	-1.994	0.327

	Grey partridge, Co-occurrence – Common quail, Alone	0.618	0.368 1.681	0.523
	Grey partridge, Co-occurrence – Common quail, Co-occurrence	0.764	0.521 1.466	0.668
	Common quail, Alone – Common quail, Co-occurrence	0.145	0.465 0.311	1.000
638	Variables considered are the broadcast species call (grey partridge, red-legg	ged partridge	and common	

Variables considered are the broadcast species call (grey partridge, red-legged partridge and common quail as control) and the interspecific situation (Alone, Co-occurrence).

Species	Dependent variables	Independent variables	t- value	df	P-value	Estimate ± SE
Grey	Territoriality score	(intercept)	8.013	1	<0.001	(4.270 ± 0.533)
partridge		Interspecific situation	-2.212	1	0.032	Co-occurrence: -2.145 ± 0.970
	Response latency	(intercept)	7.337	1	<0.001	(194.720 ± 26.540)
		Interspecific situation	-0.733	1	0.467	
	Calling behaviour	(intercept)	12.118	1	<0.001	(1.284 ± 0.106)
		Interspecific situation	-2.741	1	0.008	Co-occurrence: -0.529 ± 0.193
	Territorial response	(intercept)	14.044	1	<0.001	(0.255 ± 0.018)
	rate	Interspecific situation	-3.098	1	0.003	Co-occurrence: -0.102 ± 0.033
Red-	Territoriality score	(intercept)	10.179	1	<0.001	(2.095 ± 0.206)
legged partridge		Interspecific situation	-0.781	1	0.437	
	Response latency	(intercept)	9.042	1	<0.001	(194.550 ± 21.520)
		Interspecific situation	0.794	1	0.429	
	Calling behaviour	(intercept)	8.966	1	<0.001	(0.718 ± 0.080)
		Interspecific situation	-0.543	1	0.589	
	Territorial response	(intercept)	10.900	1	<0.001	(0.174 ± 0.016)
	rate	Interspecific situation	-0.267	1	0.790	

Territorial behaviour was considered as four dependent variables: territoriality score, response latency, calling behaviour and territorial response rate (see Table A2 for more details). The level in italic was considered as the reference in models.

Table A5 Cramer – von Mises' distribution tests comparing the territorial response rates of grey and red-legged partridges in single-species versus co-occurring situations.

			654
Species	Period	Т	P-value $\frac{655}{655}$
Grey partridge	Morning	0.108	0.274
	Evening	0.445	0.019
Red-legged partridge	Morning	0.116	0.165 658
	Evening	0.306	0.008 659
			660

Different periods of the day (morning, evening) were analyzed separately.

Table A6 Six orders of arrangement of broadcast calls according to the sequence prepared.

Sequence number	1 st species	2 nd species	3 rd species
1	Grey partridge	Red-legged partridge	Common quail
2	Grey partridge	Common quail	Red-legged partridge
3	Common quail	Grey partridge	Red-legged partridge
4	Common quail	Red-legged partridge	Grey partridge
5	Red-legged partridge	Common quail	Grey partridge
6	Red-legged partridge	Grey partridge	Common quail

Table A7 Origin of calls used to prepare audio sequences for the grey partridge, the red-legged partridge and the common quail.

Species	Call position	Author	Source	File name	Date	Time	Country
Grey partridge	1 st and 2 nd calls	Jean C. Roché	Oiseaux d'Europe et du Maghreb (CD)¹	Perdrix grise.mp3	-	-	-
	3 rd call	Krzysztof Deoniziak	xeno-canto ²	XC247485	01/06/2015	21:00	Poland
	4 th call	Peter Boesman	xeno-canto ²	XC280950	28/05/2008	06:50	Belgium
	5 th call	David M.	xeno-canto ²	XC288385-T05	08/01/2012	07:19	UK
Red-legged partridge	1 st call	Jean C. Roché	Oiseaux d'Europe et du Maghreb (CD)¹	Perdrix rouge.mp3	-	-	-
	2 nd call	David M.	xeno-canto ²	XC284092	25/04/2013	21:03	UK
	3 rd call	David M.	xeno-canto ²	XC197092-T02	03/07/2013	20:53	UK
Common quail	1 st call	Jean C. Roché	Oiseaux d'Europe et du Maghreb (CD)¹	Caille des blés.mp3	-	-	-
	2 nd call	Cedric Mrozko	xeno-canto ²	XC256801	03/07/2015	15:00	Portugal
	3 rd call	Piotr Szczypinski	xeno-canto ²	XC252505	12/06/2015	07:30	Poland
	4 th call	Peter Boesman	xeno-canto ²	XC280955	26/05/2007	11:15	Belgium
	5 th call	Tomas Belka	xeno-canto ²	XC80610	27/05/2015	14:00	Portugal
	6 th call	vprannila	soundcloud ³	Viiriäinen	25/05/2015	-	Finland

¹ Roché, J.C. (2001). On *Oiseaux d'Europe et du Maghreb* [CD]. Sittelle-Ceba.

² Xeno-canto Foundation, & Naturalis Biodiversity Center. (2005). Xeno-canto Partager les sons d'oiseaux du monde entier. Retrieved February, 2016, from http://www.xenocanto.org/

³ SoundCloud Limited. (2016). SoundCloud. Retrieved February, 2016, from https://soundcloud.com/

Table A8 Estimation of the expected number of point counts hosting grey and red-legged partridges

Richness estimator	Grey partridge	Red-legged partridge
Chao 2	98.313	156.375
Jackknife 1	81.250	135.000
Jackknife 2	97.083	158.667
Bootstrap	65.031	110.211
Mean of estimations	85.419	140.063

according to the richness estimator.

Table A9 Sampling error calculated as the percentage of false negative point counts (i.e. the proportion of points considered as free of partridges while the species is present but not detected).

_	Grey partridge	Red-legged partridge
Total number of sampled points	490	490
Observed number of points with a species occurrence	53	90
Estimated number of points with species occurrence	85	140
Proportion of false negative points	7.3%	12.5%

Figure legends

Figure 1 (a) Location of the two study sites in the Maine-et-Loire (1b) and Deux-Sèvres (1c) departments, central western France. (b) Distribution of the point count stations in Maine-et-Loire. (c) Distribution of the point count stations on the LTSER "Zone Atelier Plaine & Val de Sèvre". White dots refer to points sampled from 2016 to 2017, black dots refer to points sampled in 2017 only.

Figure 2 Rates of territorial response for grey (a) and red-legged (b) partridges according to the broadcast stimuli (conspecific, heterospecific, or control calls), assessed from a playback stimuli experiment carried out in two interspecific situations (Alone: point count where only the focus species was observed; co-occurrence: both partridge species were observed; see Methods for more details).

Figure 3 Comparison of territorial behaviours of grey (top) and red-legged (bottom) partridges, alone (light-grey) or in co-occurrence areas (dark-grey). Four behavioural metrics were considered: *territorial response rate, calling behaviour, response latency* and *territoriality score* (see Table A2 in Appendix A for more details). (a) and (c) show the distribution of the Linear Discriminant Function for grey and red-legged partridges, respectively. Horizontal bars represent 95% confidence intervals of bootstrapped "alone" linear discriminant scores. (b) and (d) show the correlations in the first dimension (first axis) of the four variables.

Figure 4 Daily distribution of the territorial response rate of grey (top) and red-legged (bottom) partridges according to the morning time deviation from sunrise (left) and evening time deviation from sunset (right), assessed by playback experiment in 2016 and 2017. Solid lines represent the response rate of each species where alone, dashed lines the response rate where both species cooccurred.

Figure 5 Habitat niche of grey (a) and red-legged (c) partridges when alone and in co-occurrence with the other species (b, d respectively). The solid and dashed lines illustrate respectively 100% and 50% of the available environment. A white arrow indicates a significant shift from the niche centroid in absence to the niche centroid in presence of the second species. (e) represents the correlation circle of the PCA describing the environmental space available in the study area. PC1 and PC2 show the inertia of the first two axes of the PCA.

Figure A1 Cumulated number of point counts with occurrence of (a) grey, and (b) red-legged partridges, as a function of the number of the sampling session. Dashed lines correspond to 95% confidence intervals.

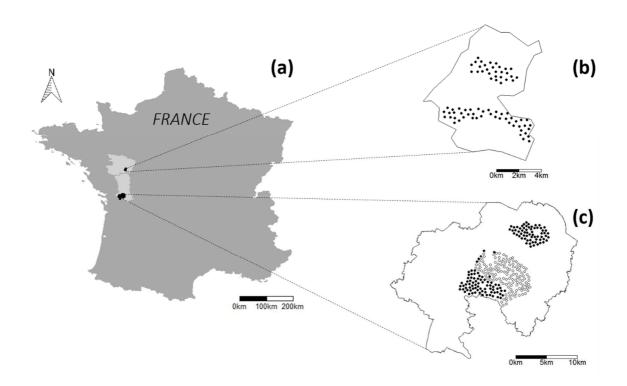


Figure 1

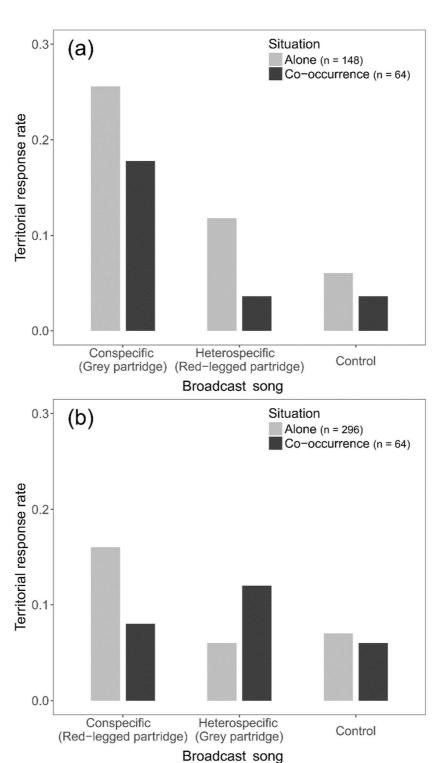


Figure 2

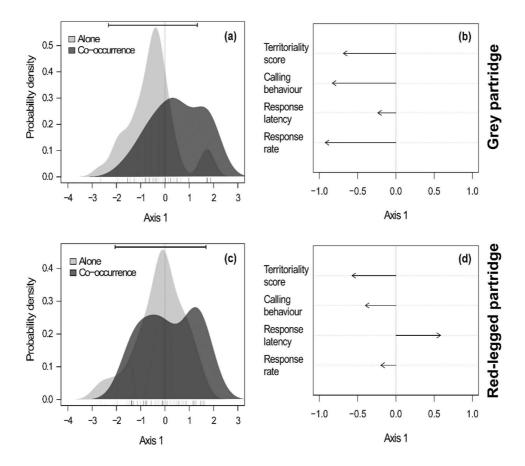
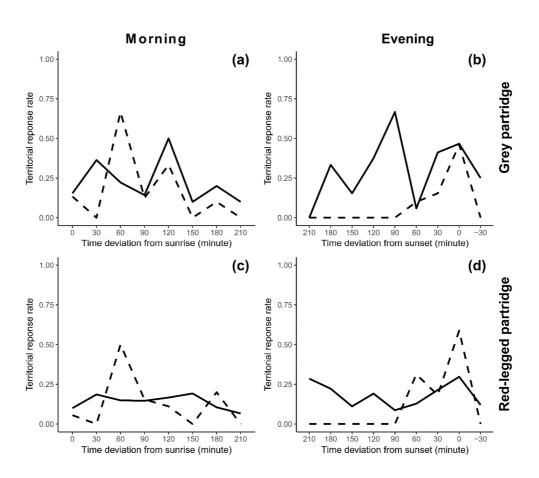
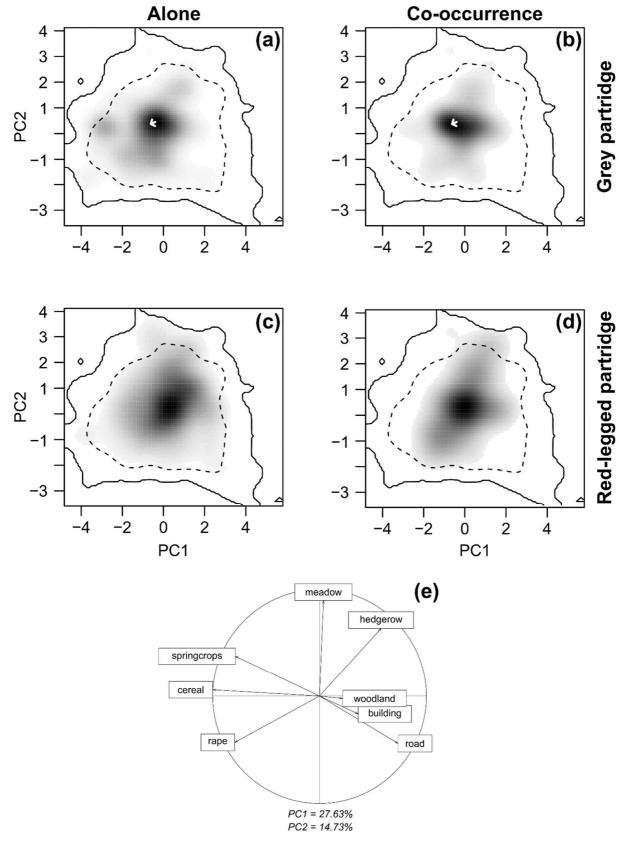


Figure 3 736





741 Figure 5



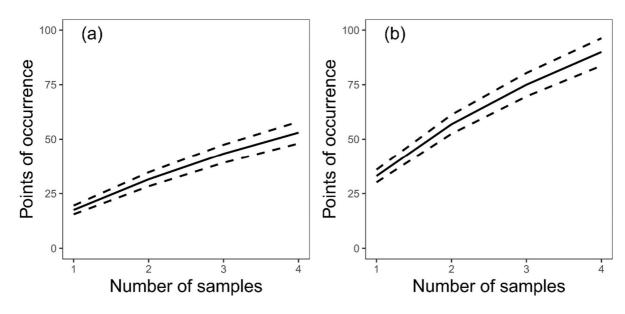


Figure A1