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### ► To cite this version:

Julien Morin, Adrien Gandolfo, Brice Temime-Roussel, Gregory Brochard, Virginie Bergé, et al.. Key parameters influencing the uptake of m-xylene on photocatalytic paints. *Building and Environment*, 2020, 179, pp.106979. 10.1016/j.buildenv.2020.106979 . hal-02733482

**HAL Id: hal-02733482**

**<https://hal.science/hal-02733482>**

Submitted on 2 Jun 2020

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1 **Key parameters influencing the uptake of m-xylene on photocatalytic paints**

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13

## 1 **ABSTRACT**

2 Given the toxic nature of many volatile organic compounds (VOCs) present within indoor  
3 environments, it is necessary to develop new materials to eliminate them. Photocatalytic  
4 paints represent a promising technology to remove the indoor pollutants but they are not  
5 optimised yet. Here we evaluate the capacities of one conventional white-wall paint, three  
6 photocatalytic paints containing 3.5, 5.25 and 7% of nano-TiO<sub>2</sub> PC 500, and two  
7 photocatalytic paints with a hybrid cellulose nanocrystals-nano-TiO<sub>2</sub> (CNC-nano-TiO<sub>2</sub>) to  
8 eliminate *m*-xylene from the indoor environment. In particular, we assess the impact of nano-  
9 TiO<sub>2</sub> content, light intensity, relative humidity (RH), surface temperature (288-320 K) and  
10 pigment volume concentration (PVC) on uptake of *m*-xylene. Uptake coefficients of *m*-xylene  
11 increase by a factor 3, 5 and 3, respectively, as a function of nano-TiO<sub>2</sub> content, irradiance  
12 and PVC. In contrast, uptakes of *m*-xylene decrease for a factor of 2 and 3, as a function of  
13 RH, and surface, suppressing the photocatalytic activity. The addition of CNC-nano-TiO<sub>2</sub> in  
14 the binder induces limited binder degradation. These are very promising results which may  
15 help to improve the performances of indoor photocatalytic paints aimed to eliminate indoor  
16 airborne pollutants.

17 **Keywords:** Indoor Air; Photocatalysis; Uptake coefficient; Cellulose Nanocrystals; Flow  
18 Tube; PTR-MS

19

## 1. INTRODUCTION

Volatile organic compounds (VOCs) represent ubiquitous indoor air pollutants [1,2]. Some VOCs are known to cause human health effects and VOC compounds, such as formaldehyde, benzene, dichloromethane, have been identified as carcinogens [3]. The indoor sources of VOCs are numerous and include emissions from building itself, decorative material (paints, vanishes...), furniture, inhabitants' activities (cooking, smoking...) and domestic cleaning products, among others. Recent studies carried out in France [4] and in Europe [5] shown a VOCs indoor-to-outdoor environment concentration ratios up to 14.3. In occidental countries, humans spend on average over 90% of their lifetime within indoor settings [6]. Based on these two observations, it is necessary to develop methods to reduce VOCs levels in order to improve indoor air quality. This approach include different techniques such as biofiltration, adsorption, condensation, incineration and photocatalytic oxidation (PCO) [7–9]. The photocatalytic system allows to oxidize, as well as reduce, air pollutants adsorbed on semiconductor surface. PCO is intensively studied and considered as a promising technology [10] to improve indoor air quality.

Benzene, Toluene, Ethylbenzene and *o-m-p* Xylenes also called BTEX have been extensively studied in indoor air because of their toxicity [11–14] and their high measurements frequency. For instance, Hazrati *et al* .[15] summarized house BTEX concentrations in different countries. Their concentrations range as 1-5 ppb, 3-10 ppb, 1-5 ppb and 2-15 ppb for benzene, toluene, ethylbenzene and xylenes, respectively. The main sources of xylenes are vehicles exhaust [16].

The degradation of *m*-xylene has been studied on different supports as a function of Relative Humidity (RH) [17–20], irradiance [21] and nano-TiO<sub>2</sub> content [21]. To the best of our knowledge, there is no previous studies focused on the influence of surface temperature, aging and Pigment Volume Concentration (PVC) of photocatalytic materials on the *m*-xylene

1 degradation. The study of Strini *et al.* [21] has shown the increase of the photocatalytic  
2 activity with increase of light intensity. The improved photocatalytic ability was also observed  
3 with increase of nano-TiO<sub>2</sub> content [21]. Jo *et al.* [19] observed no impact of RH on  
4 photocatalytic activity whereas two other studies showed an influence of RH on  
5 photocatalytic efficiency. At low concentration of *m*-xylene (46 ppb) [18] a decrease of  
6 photocatalytic activity has been observed with increase of RH. At high concentration  
7 (460 ppb) of *m*-xylene [18] an increase between 0 and 45% of RH and a decrease above 45%  
8 of RH; of photocatalytic activity have been observed. These different studies have been  
9 performed on photocatalytic thin films [19], glass fibers [17] and cement mortars [18,21] but  
10 never on photocatalytic paints. Tryba *et al.* [22] have studied the degradation of  
11 benzo[a]pyrene by photocatalytic paints but they have not assessed the impact of different  
12 environmental and physical parameters on photocatalytic efficiency to eliminate indoor  
13 VOCs. Additional studies are recommended to improve the knowledge o, this matter.

14 In this study we have chosen *m*-xylene as a model compound because: 1) it is commonly  
15 observed in indoor environment [15], 2) it is not emitted by the studied paints [23] thus  
16 avoiding experimental artefacts, 3) it presents hazardous effects on human health (neurologic  
17 effects, digestive disorder, irritation...) [14] and 4) it is not an easily degradable compound  
18 due to the relative stability of aromatic rings with respect to its oxidation and reduction  
19 process.

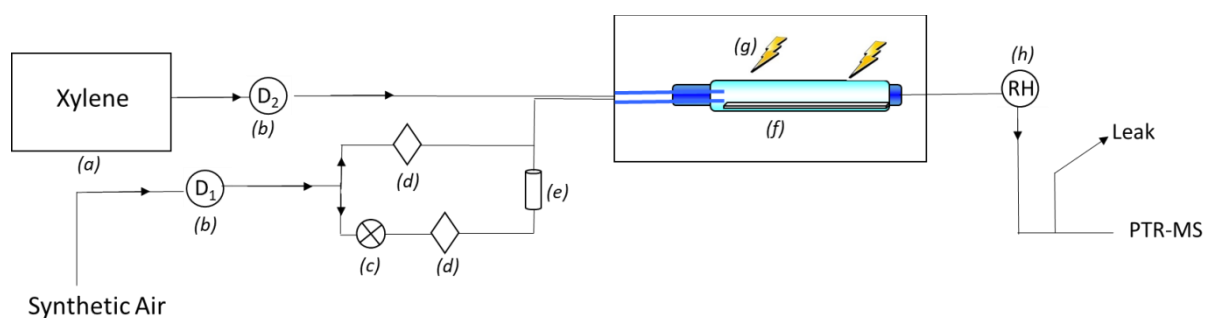
20 In the present work, the uptake coefficients of *m*-xylene on various paints (non-photocatalytic,  
21 hereafter, reference paint and photocatalytic paints) were determined under different  
22 experimental conditions in presence of UV irradiation. The influence of key parameters  
23 driving the reaction was evaluated by the comparative approach. The influence of  
24 environmental parameters (surface temperature, RH, irradiance intensity) as well as paint

1 physical parameters (photocatalyst type and content, paint aging and paint porosity) on *m*-  
2 xylene degradation have been examined.

## 3 2. MATERIALS AND METHODS

### 4 2.1. Experimental study

5 A double wall flow tube photoreactor was connected to High Sensitivity-Proton Transfer  
6 Reaction-Mass Spectrometry (HS-PTR-MS) (IONICON Analytik, Austria). The operational  
7 procedure of flow tube photoreactor has been described in detail by Gandolfo *et al.* [24] and  
8 Morin *et al.* [23]. The schematic diagram of the experimental setup is presented in Figure 1.  
9 Experimental details that are particularly relevant to this work are given below.



11 Figure 1: Schematic illustration of the experimental setup. Permeation oven (a), mass flow  
12 controller (b), on/off valve (c), needle valve (d), relative humidity system (e), reactor + glass  
13 paint (f), UV-lamps (g), RH probe + hygrometer (h).

14

15 The flow tube is made of borosilicate glass with an internal volume of 131 cm<sup>3</sup> and was  
16 coupled to a temperature-controlled circulating bath (RC6 LAUDA, Germany) which allows  
17 to perform the experiments at different surfaces temperatures between 288 and 320 K ( $\pm 1$  K).  
18 Experiments were carried out at 1 atm total pressure of synthetic air (Linde Gas, France,  
19 >99.999 stated purity). The glass plate recovered by the paint was introduced into the flow  
20 tube reactor with the painted side upward oriented. Inside the flow tube, a glass made

1 movable injector was inserted into the flow tube to inject *m*-xylene and vary the residence  
2 time and thus assess the kinetics of the heterogeneous reactions.

3 The flow tube reactor was placed in a stainless steel box with four UV fluorescent lamps  
4 (Philips TL-D 18W Actinic BL, 340-400 nm,  $\lambda_{\text{max}} = 368$  nm, length = 60 cm) mounted side by  
5 side on the upper part of the box above the flow tube photoreactor. The UV-lamps simulated  
6 UV solar irradiation and could be operated individually. The spectral irradiance of the UV  
7 lamps used in this work had been determined previously [25]. The integrated spectral  
8 irradiance ( $340 < \lambda < 400$  nm) was estimated to be 4.9, 8.8, 13.5 and 18.6 W m<sup>-2</sup> for 1, 2, 3 and 4  
9 lamps, respectively.

10 Gaseous *m*-xylene was generated using a gas-phase generator (PUL 200, Saint Chamas,  
11 France). A sealed permeation cell consisting of a Teflon tube (perfluoroalkoxy, 4.0×6.4 mm,  
12 8 cm long) filled with pure *m*-xylene (Sigma-Aldrich,  $\geq 99.5\%$ ) was used to generate gas-  
13 phase *m*-xylene. The permeation cell was placed in the oven chamber of the gas-phase  
14 generator. At the exit of the oven, a constant flow of 10 sccm (Brooks SLA, Series mass flow  
15 controller, in the range 0–10 sccm,  $\pm 1\%$  stated accuracy) of gaseous *m*-xylene was allowed to  
16 enter the injector. A sheath synthetic air flow of 190 sccm (Brooks SLA, Series mass flow  
17 controller, in the range 0–500 sccm,  $\pm 1\%$  stated accuracy) was allowed to enter the flow tube  
18 to dilute the *m*-xylene flow. To better simulate realistic indoor concentration, all experiments  
19 were carried out at  $\sim 50$  ppb of *m*-xylene.

20 Relative humidity was adjusted by splitting the sheath synthetic air flow into a dry flow and a  
21 humidified flow. For all the experiments, the RH was measured online at the exit of reactor by  
22 a hygrometer “Hygrolog NT2” (Rotronic) with “HygroClip SC04” probe and ranged from 2  
23 to 70% with accuracy of  $\pm 1.5\%$ .

24 The *m*-xylene mixing ratio was measured in real time by a HS-PTR-MS by following its

1 protonated molecular pic ( $m/z$  107). Formaldehyde ( $m/z$  31), acetaldehyde ( $m/z$  45) and  
2 acetone ( $m/z$  59) were also monitored as intermediates by-products. The HS-PTR-MS high  
3 time resolution (10 seconds) allows us to measure rapid changes in gas phase concentrations.  
4 The drift temperature, drift voltage and a pressure were fixed at  $333\pm 1$  K,  $500\pm 1$  V and  
5  $2.02\pm 0.01$  mbar respectively, that corresponded to the E/N value of 124 Townsend  
6 ( $1 \text{ Townsend} = 10^{-17} \text{ V cm}^2$ ), where E is the electric field ( $\text{V cm}^{-1}$ ) and N is the ambient air  
7 number density within the drift tube ( $\text{cm}^{-3}$ ). The calculation of VOCs mixing ratio and surface  
8 emission fluxes is described in detail by Morin *et al.* [23].

## 9 **2.2. Preparation of paints**

10 Paints were formulated and produced by the paint manufacturer ALLIOS (Marseille, France).  
11 TITANE PC 500 in anatase form with 85% of  $\text{TiO}_2$  nanoparticle was used as a photocatalyst.  
12 TITANE PC 500 represents ultrafine  $\text{TiO}_2$  powder with specific area of  $350 \text{ m}^2 \text{ g}^{-1}$  and  
13 particle size between 3 and 5 nm. A mixture of ground additives (slurry) was prepared with  
14 35% (w/w) active nanosized  $\text{TiO}_2$ . Then, other paint constituents (architectural constituent of  
15 paints, i.e.,  $\text{CaCO}_3$  and micrometric  $\text{TiO}_2$ ) were mixed with 10, 15 and 20% of the slurry to  
16 reach 3.5, 5.25 and 7% (w/w) of photocatalytic active  $\text{TiO}_2$  nanoparticles, respectively. The  
17  $\text{TiO}_2$  nanoparticles are not present in the reference paint. The prepared paint is applied on the  
18 glass plate by dint of a spiral film applicator. This methodology allows to obtain controlled  
19 and reproducible wet paint thickness of  $100 \mu\text{m}$ . Then, the paints were dried in an oven  
20 ventilated by clean air during 21 days at 296 K and 55% of RH resulting in a uniform  
21 thickness film of  $40 \mu\text{m}$ .

22 To investigate the impact of porosity on the degradation rates, three paints with different  
23 porosities were developed. To assess the paint porosity, it was necessary to introduce the  
24 Critical Pigment Volume Concentration (CPVC). CPVC is a specific Pigment Volume  
25 Concentration (PVC) in which the voids between the solid particles (structural agents and



1 pigments) are just filled with the binder [26]. The PVC and CPVC are linked to the paint  
2 porosity as follows:

$$3 \quad \text{porosity (\%)} = 1 - \frac{CPVC}{PVC} \quad (\text{Eq.1})$$

4 The value of CPVC is not clearly known but it can be estimated between 50-60%  
5 (manufacturer information). For this study, three paints with PVC of 53, 72 and 80% were  
6 considered. The calculation of PVC were shown previously by Gandolfo *et al.* [26].

7 In another series of paints, the standard nano-TiO<sub>2</sub> (PC 500) was replaced by a hybrid  
8 cellulose nanocrystals-nano-TiO<sub>2</sub> (CNC-TiO<sub>2</sub>) in the paint preparation to increase nano-TiO<sub>2</sub>  
9 dispersion in the slurry. This addition allowed to form hybrid CNC-TiO<sub>2</sub> nanoparticles which  
10 may improve the photocatalytic activity of TiO<sub>2</sub>. The aim of this hybrid nanoparticle was to  
11 reduce the quantity of nano-TiO<sub>2</sub> embedded in the paint and keeps high photocatalytic  
12 efficiency. In this work, two photocatalytic paints with 0.16 and 0.23% (w/w) of nano-TiO<sub>2</sub>  
13 have been studied. Photocatalytic paints with hybrid CNC-nano-TiO<sub>2</sub> and the photocatalytic  
14 paint with 3.5% of nano-TiO<sub>2</sub> PC 500 have been aged in a accelerated weathering chamber  
15 for 500 hours and 1000 hours according to the ISO norm 16474-3 [27,28]. Aging of 500 hours  
16 correspond to an important aging of paint and an aging of 1000 hours correspond to a paint of  
17 end-of-life.

18 The paints, except CNC paints, were applied on one side of glass plates that were 29 cm long  
19 and 1.9 cm wide according to a standard operating procedure developed by the manufacturer  
20 ALLIOS [29]. The CNC paints were applied on Leneta supports that were 29 cm long and  
21 1.9 cm wide. To assure proper fit of the Leneta surface into the reactor, the later were fixed on  
22 a glass plate with wire to both ends prior to kinetic experiments.

### 23 **2.3.Kinetic measurements**

24 The aim of this study was to evaluate the uptake coefficients ( $\gamma$ ) of *m*-xylene on reference

1 paint and on photocatalytic paints. The uptake coefficients were determined using the  
2 following equation [30]:

$$3 \quad \gamma = \frac{4k_{1st}}{\omega} \frac{V}{S} \quad (\text{Eq.2})$$

4 where  $k_{1st}$  is the pseudo-first order rate constant of *m*-xylene loss ( $s^{-1}$ ),  $\omega$  is the mean  
5 molecular velocity ( $cm\ s^{-1}$ ) and  $V/S$  represents the ratio of the volume of the reactor to surface  
6 of the glass plate (cm).

7  $k_{1st}$  was determined by the slope of  $\ln(C_0/C_x)$  plotted as a function of the residence time ( $t$ ).  $C_0$   
8 is the initial concentration of *m*-xylene in position  $P_0$ . In this case, gaseous *m*-xylene is not  
9 exposed to the paint.  $C_x$  is the concentration of *m*-xylene at different positions from  $P_1$  to  $P_4$   
10 (Table 1). The difference of concentrations between  $C_0$  and  $C_x$  corresponds to amount of *m*-  
11 xylene consumed by its reaction with the paint surface. Each position was hold during about  
12 30 min to get stabilized signal and reliable measuring points. After each position change, the  
13 injector was pushed in position  $P_0$  to go back to the initial conditions. The variation of the  
14 concentration in function of the injector position enabled to calculate the pseudo-first order  
15 rate constant [25,26,31,32].

16 **Table 1:** Distance to the end of reactor (cm), exposed paint surface ( $cm^2$ ) and contact time (s)  
17 between gaseous *m*-xylene and paint as a function of movable injector position.

Position	Distance (cm)	Exposed surface ( $cm^2$ )	Contact time (s)
$P_0$	0	0	0
$P_1$	4	7.6	5.4
$P_2$	8	15.2	10.9
$P_3$	15	28.5	20.4
$P_4$	22	41.8	29.9

18

19 A typical experiment and the assessment of pseudo-first order rate constant of *m*-xylene are  
20 presented in SI (Figure S1-S2). Blank experiments have been performed to ensure that no wall  
21 loss occurs during experiments and corrections have been applied to the obtained results.

1 The Cooney-Kim-Davis (CKD) method was used to correct the uptake coefficients from  
2 diffusion limitation in the flow tube [33,34]. Under the applied experimental conditions, the  
3 estimated correction by CKD was less than 10%.

4 The experimental conditions are given in Table 2.

5 **Table 2:** Experimental conditions

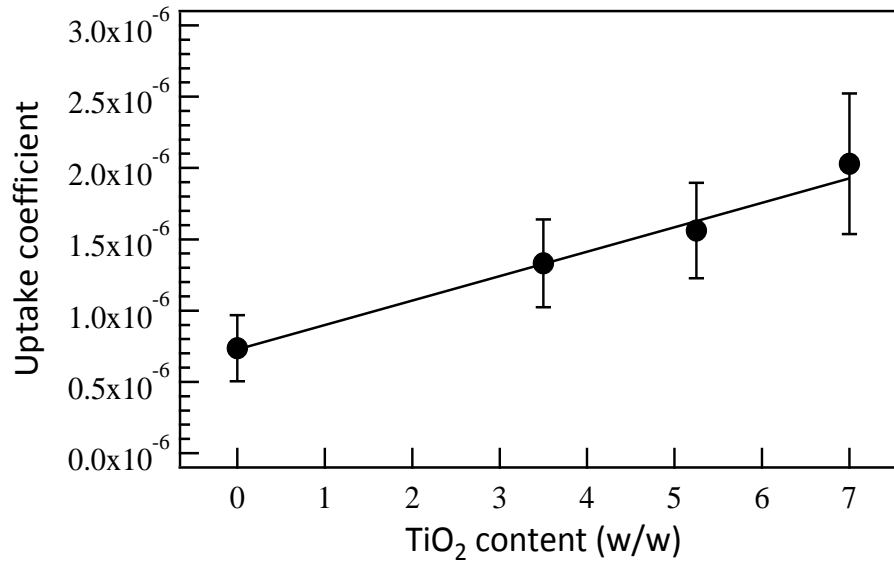
Experiments	Nano-TiO <sub>2</sub> content w/w (%)	Surface temperature (K)	RH (%)	Irradiance (W m <sup>-2</sup> )	PVC (%)
1	0, 3.5, 5.25 and 7	296	40	8.8	72
2	0 and 7	298	40	0-18.6	72
3	7	295	0-70	8.8	72
4	7	288-320	37	8.8	72
5	7	296	40	8.8	53-80
6	NCC : 0.16, 0.23 PC 500 : 3.5	298	40	8.8	NA*

6 \* Not available

### 7 **3. RESULTS AND DISCUSSION**

#### 8 **3.1. Nano-TiO<sub>2</sub> content**

9 The influence of the quantity of nano-TiO<sub>2</sub> on the uptake coefficients of *m*-xylene is presented  
10 in Figure 2.



1

2 **Figure 2:** The effect of nano-TiO<sub>2</sub> content on uptake coefficients of *m*-xylene. *m*-xylene  
 3 mixing ratio: 50 ppb, temperature: 296 K, RH: 40%, light intensity: 8.8 W m<sup>-2</sup>, PVC: 72%,  
 4 aging: 0 hour, nano-TiO<sub>2</sub> PC 500. The line represents the linear regression. The errors bars are  
 5 derived from uncertainties associated to uptake coefficient.

6

7 The uptake coefficient for the reference paint is  $(7.3 \pm 2.3) \cdot 10^{-7}$  and increases up to  
 8  $(2.0 \pm 0.5) \cdot 10^{-6}$  for the photocatalytic paint at 7% of nano-TiO<sub>2</sub>. The linear relation between  
 9 uptake coefficient and nano-TiO<sub>2</sub> content is described as follows:

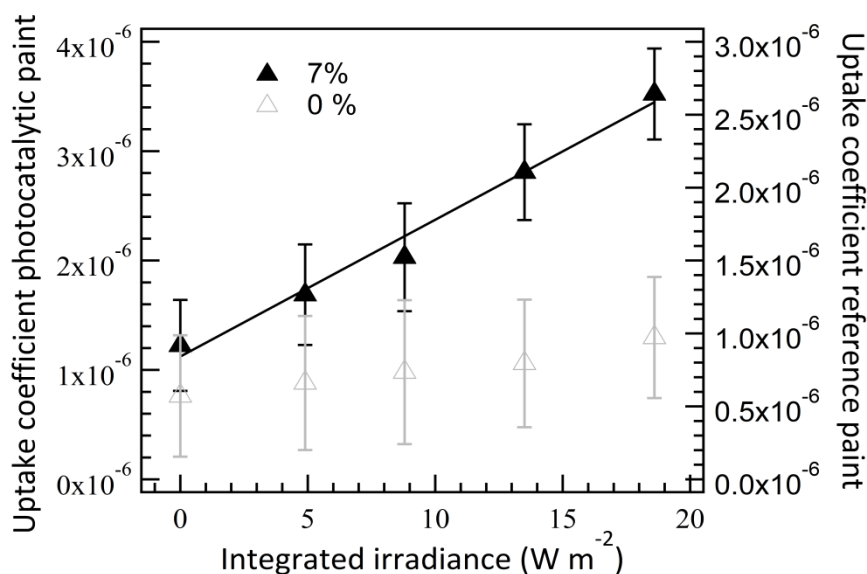
10 
$$\gamma = (1.8 \pm 0.2) \times 10^{-7} \times nanoTiO_2content + (7.1 \pm 0.7) \times 10^{-7} \quad (\text{Eq.3})$$

11 When the nano-TiO<sub>2</sub> quantity increases, the quantity of OH radicals produced increases [35]  
 12 on the paint surface which accelerates the degradation of *m*-xylene. For this reason, the  
 13 increase of quantity of TiO<sub>2</sub> nanoparticles from 0 to 7%, improves the removal of *m*-xylene  
 14 by a factor of 3 which is in agreement with the reported factor of 4 in a previous study  
 15 focused on removal of *o*-xylene [21] on cementitious photocatalytic materials.

16 Due to improved removal efficiency towards *m*-xylene, the photocatalytic paint containing  
 17 7% of nano-TiO<sub>2</sub> was used to assess the impact of RH, surface temperature, light intensity and  
 18 pigment volume concentration.

19 **3.2.UV irradiation intensity**

- 1 The effect of light intensity on the uptake coefficients of *m*-xylene on reference paint and on  
 2 photocatalytic paint containing 7% of nano-TiO<sub>2</sub> has been evaluated (Figure 3).



3  
 4 **Figure 3:** The effect of light intensity on uptake coefficients of *m*-xylene. *m*-xylene mixing  
 5 ratio: 50 ppb, temperature: 298 K, RH: 40%, PVC: 72%, aging: 0 hour. Photocatalytic paint:  
 6 7% nano-TiO<sub>2</sub> PC 500 (filled triangle) and reference paint (empty triangle). The line  
 7 represents the linear regression of the plotted data. The errors bars are derived from  
 8 uncertainties associated to uptake coefficient.

9  
 10 The uptakes of *m*-xylene on a reference paint only slightly increase with the variation of  
 11 spectral irradiance from  $(5.7 \pm 1.0) \cdot 10^{-7}$  in dark up to  $(9.7 \pm 1.4) \cdot 10^{-7}$  at  $18.6 \text{ W m}^{-2}$ . In contrast,  
 12 the uptake of *m*-xylene on a photocatalytic paint with 7% TiO<sub>2</sub> nanoparticles increases for a  
 13 factor of 3, from  $(1.2 \pm 0.4) \cdot 10^{-6}$  in dark up to  $(3.5 \pm 0.4) \cdot 10^{-6}$  under maximum light irradiation  
 14 of  $18.6 \text{ W m}^{-2}$ . The linear relation depicted in plot is described by the equation (Eq.4):

$$15 \quad \gamma = (1.3 \pm 0.1) \times 10^{-7} \times \text{Irradiance (W m}^{-2}) + (1.1 \pm 0.1) \times 10^{-6} \quad (\text{Eq.4})$$

16 The increase of light intensity leads to enhanced formation of OH radicals formed on the paint  
 17 surface which in turn leads to accelerated degradation of *m*-xylene [36].

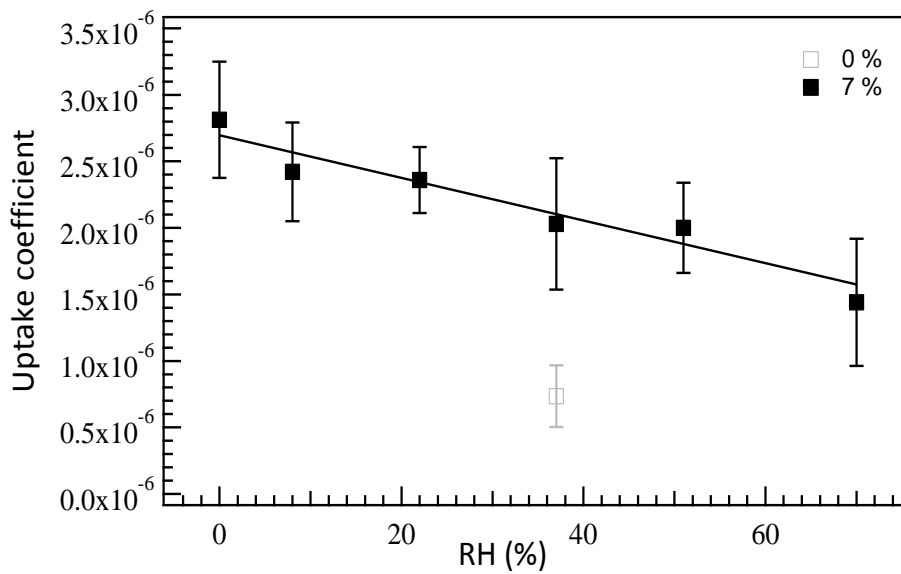
18 A similar trend of linear increase of the degradation of toluene on glass fiber as a function of  
 19 light intensity was observed by Palau *et al.* [17] and the degradation of *o*-xylene on

1 cementitious material by Strini *et al.* [21]. In their work, an increase by a factor 2.5 was  
 2 observed that is in agreement with the factor 2 measured here.

3 These results show that the experiments were performed under light unsaturation and that the  
 4 surface chemical reaction is the rate-determining step.

### 5 3.3.Relative humidity

6 The impact of the RH on the uptake coefficients is shown in Figure 4.



7  
 8 **Figure 4** : The effect of RH on uptake coefficients of *m*-xylene. *m*-xylene mixing ratio:  
 9 50 ppb. Photocatalytic paint: 7% nano-TiO<sub>2</sub> PC 500 (filled square) and reference paint (empty  
 10 square), light intensity: 8.8 W m<sup>-2</sup>, temperature: 295 K, PVC: 72%, aging: 0 hour. The line  
 11 represents the linear regression. The errors bars are derived from uncertainties associated to  
 12 uptake coefficient.

13  
 14 The uptake coefficient of *m*-xylene is (2.8±0.4)·10<sup>-6</sup> at 2% RH and decreased up to  
 15 (1.4±0.5)·10<sup>-6</sup> at 70% RH. The linear relation is described by the following mathematical  
 16 equation (Eq.5):

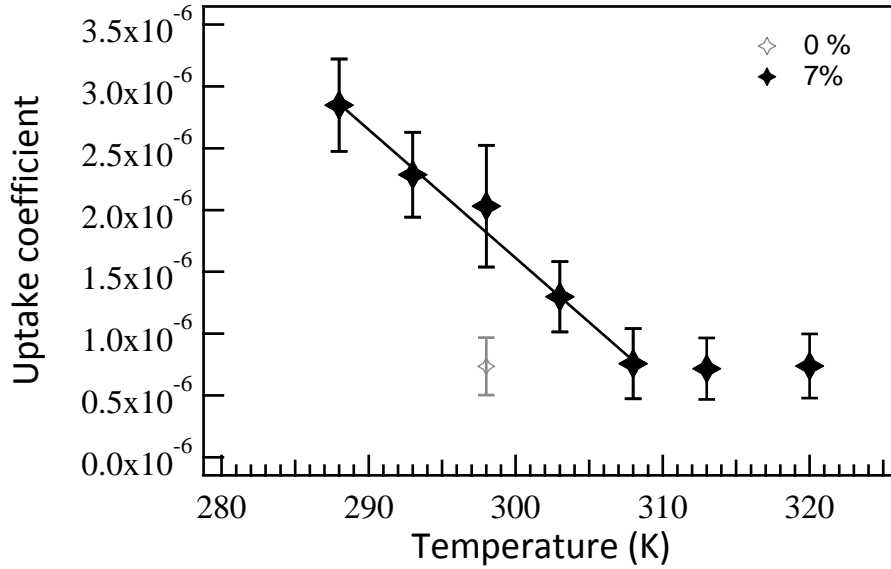
$$17 \quad \gamma = -(1.7 \pm 0.2) \times 10^{-8} \times RH (\%) + (2.7 \pm 0.1) \times 10^{-6} \quad (\text{Eq.5})$$

1 The decrease of the uptake by a factor 2 is in good agreement with the observed decreasing  
2 conversion ratio (46 ppb of *m*-xylene) by a factor of 1.5 as the RH increased from 0 to 70%  
3 [18]. This phenomenon can be ascribed to a competitive adsorption between water and *m*-  
4 xylene molecules on the paint surface. At low humidity, the adsorption competition on  
5 semiconductor surface is limited which allowed the more efficient degradation of *m*-xylene.  
6 As the humidity increases, absorbed water molecules on nano-TiO<sub>2</sub> competes with the *m*-  
7 xylene adsorption. Consequently, the photocatalytic heterogeneous reactivity is reduced at  
8 high relative humidity [8].

9 At very high mixing ratio of *m*-xylene (460 ppb), Martinez *et al.* [18] observed an increased  
10 removal efficiency of *m*-xylene at 45% RH which further decreased for RH between 45 and  
11 80%. Such increase of removal efficiency of xylene (230 ppb) on photocatalytic materials has  
12 also been observed by Palau *et al.* [17] for RH comprised between 4 and 70%.

### 13 **3.4.Surface temperature**

14 To the best of our knowledge, the impact of surface temperature on the uptake coefficients of  
15 *m*-xylene has not been studied previously. The surface temperature varied between 288 and  
16 320 K (Figure 5).



1

2 **Figure 5** : The effect of surface temperature on uptake coefficients of *m*-xylene. *m*-xylene  
 3 mixing ratio: 50 ppb. Photocatalytic paint: 7% nano-TiO<sub>2</sub> PC 500 (filled lozenge) and  
 4 reference paint (empty lozenge), light intensity: 8.8 W m<sup>-2</sup>, RH: 37%, PVC: 72%, aging:  
 5 0 hour. The line represents the linear regression. The errors bars are derived from  
 6 uncertainties associated to uptake coefficient.

7

8 Figure 5 shows that the uptake coefficients of *m*-xylene decreased significantly as the  
 9 temperature increases followed by a saturation effect. At 288 K; the uptake coefficient is  
 10  $(2.8 \pm 0.4) \cdot 10^{-6}$  and decreased up to  $(7.6 \pm 2.9) \cdot 10^{-7}$  at 308 K. At higher temperatures than  
 11 308 K, the uptakes of *m*-xylene remain stable. For the range of temperatures between 288 and  
 12 308 K, the solid line in Figure 5 represents the regression line which corresponds to the  
 13 following equation (Eq.6):

$$14 \quad \gamma = -(1.0 \pm 0.1) \times 10^{-7} \times T (K) + (3.3 \pm 0.2) \times 10^{-5} \quad (\text{Eq.6})$$

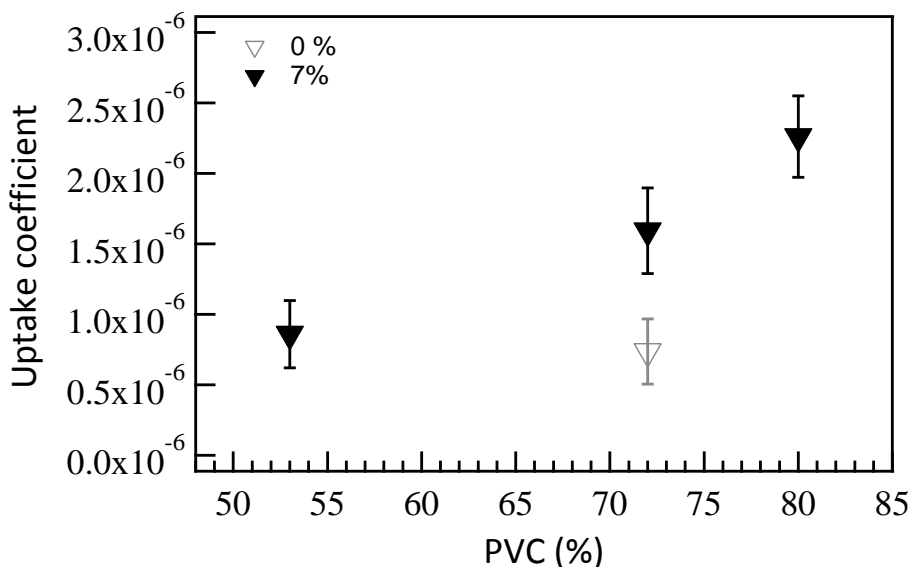
15 The reduced removal efficiency of photocatalytic paint with 7% TiO<sub>2</sub> nanoparticles at higher  
 16 temperatures can be ascribed to the reduced contact time between gaseous *m*-xylene and the  
 17 paint surface. Namely, as the surface temperature increases *m*-xylene desorb more rapidly and  
 18 thus decrease the contact time with the paint surface. In that way, the contact time becomes  
 19 not sufficient to achieve the degradation process [37].



1 The uptakes of *m*-xylene for temperature range between 308 and 320 K,  $(7.6\pm 2.9)\cdot 10^{-7}$ , fall in  
2 the same range as the uptake of *m*-xylene on a reference paint,  $(7.3\pm 2.3)\cdot 10^{-7}$ . Hence, the  
3 observed plateau in Figure 5 indicates that at higher temperature than 308 K, the removal of  
4 *m*-xylene is not driven by the photocatalytic process.

### 5 3.5.Effect of pigment volume concentration (PVC)

6 The removal of *m*-xylene was also evaluated on photocatalytic paints containing 7% of nano-  
7 TiO<sub>2</sub> with PVC of 53%, 72% and 80% (Figure 6).



8  
9 **Figure 6** : The effect of PVC on uptake coefficients of *m*-xylene. *m*-xylene mixing ratio:  
10 50 ppb. Photocatalytic paint: 7% nano-TiO<sub>2</sub> PC 500 (filled triangle) and reference paint  
11 (empty triangle), temperature: 296 K, light intensity: 8.8 W m<sup>-2</sup>, RH: 37%, aging: 0 hour. The  
12 errors bars are derived from uncertainties associated to uptake coefficient.

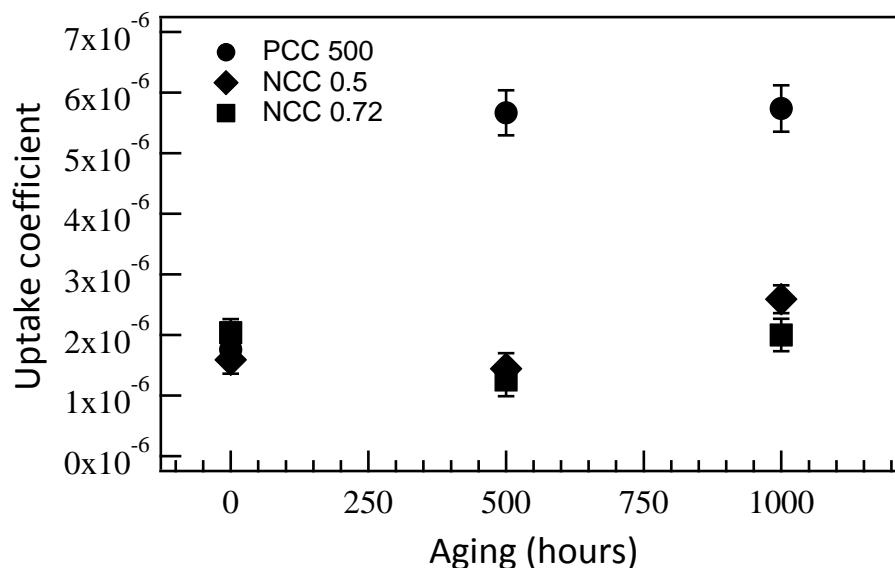
13  
14 The uptake coefficients of *m*-xylene increase from  $(8.6\pm 2.4)\cdot 10^{-7}$  to  $(2.3\pm 0.3)\cdot 10^{-6}$  as the PVC  
15 increases from 53 to 80%.

16 This increase of photocatalytic efficiency towards the removal of *m*-xylene can be ascribed to  
17 the fact that the increase of the porosity allows gaseous *m*-xylene to penetrate more easily  
18 through the paint and increase the contact time with the OH radicals which are generated by  
19 the irradiation of TiO<sub>2</sub> nanoparticles [35]. Similarly, the increase of porosity on a commercial

1 paint induced faster degradation of methyl ethyl ketone [38]. Jansson *et al.* [39] observed  
2 accelerated degradation of formaldehyde by zeolite/TiO<sub>2</sub>.

### 3 3.6.CNC and aging

4 The aging of three photocatalytic paints with various photocatalysts namely PC 500, CNC 0.5  
5 and CNC 0.72 have been examined (Figure 7). The PC 500 represents the photocatalytic paint  
6 with 3.5% of nano-TiO<sub>2</sub> and CNC 0.5 and CNC 0.72 paints contain 0.5 and 0.72% of hybrid  
7 cellulose nanocrystals (CNC) embedded in that paint that correspond to 0.16% and 0.23% of  
8 nano-TiO<sub>2</sub>, respectively. At first sight, the three paintings are identical in term of color and  
9 appearance regardless of the aging.



10

11 **Figure 7:** The effect of aging and photocatalyst on uptake coefficients of *m*-xylene. *m*-xylene:  
12 mixing ratio: 50 ppb, temperature: 296 K; light intensity: 8.8 W m<sup>-2</sup>, RH: 37%, paint PC 500  
13 (circle), paint CNC 0.5 (lozenge) and paint CNC 0.72 (square). The errors bars are derived  
14 from uncertainties associated to uptake coefficient.

15

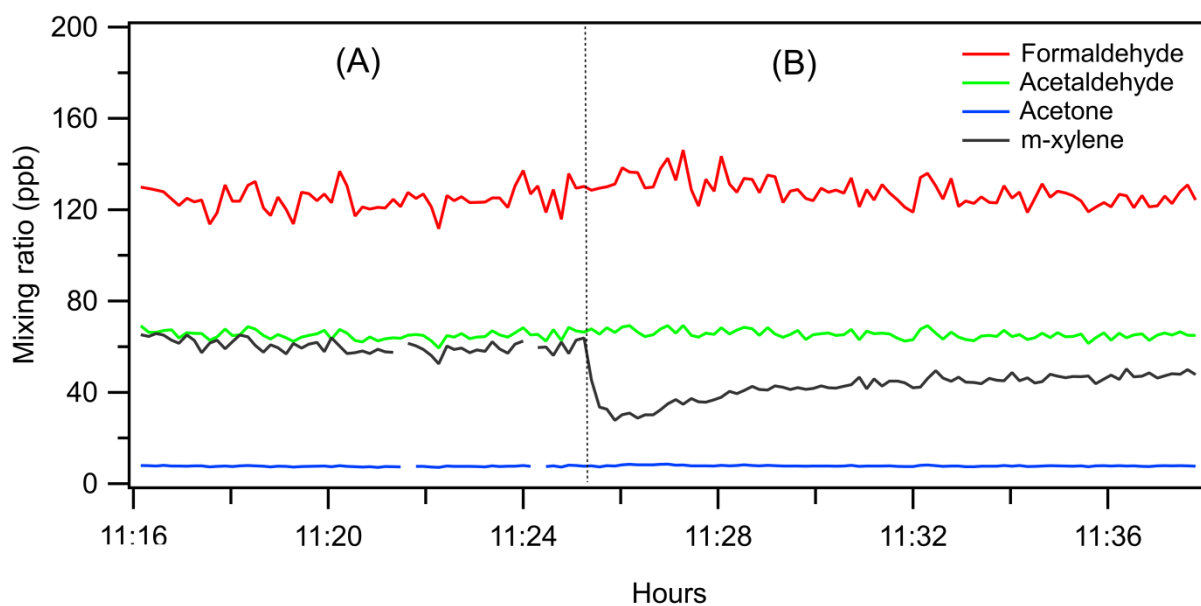
16 As shown on Figure 7, the uptake coefficients of *m*-xylene prior to aging of paints are similar  
17 for all the considered photocatalytic paints,  $(1.7\pm 0.3)-(2.0\pm 0.3)\cdot 10^{-6}$ . The addition of CNC in  
18 the paint formulation allowed to obtain similar effect on *m*-xylene degradation with 15 and 22  
19 times lower nano-TiO<sub>2</sub> quantity.

1 The uptake coefficient of *m*-xylene on the paint PC 500 increased with the aging. After 500 or  
2 1000 hours of aging, uptake coefficients are 3 times higher than for the fresh dry paint. On the  
3 other hand, the aging has no effect on paint containing CNC-TiO<sub>2</sub> hybrid nanoparticles. The  
4 effect of aging on the binder of the paint was previously observed by Truffier-Boutry *et al.*  
5 [28]. Once the binder was degraded, nano-TiO<sub>2</sub> became more accessible to interact with *m*-  
6 xylene; hence, increased the removal efficiency. This hypothesis is confirmed by the results  
7 presented in Figure 7 for the paint PC 500. However, the consumption of *m*-xylene is not  
8 altered anymore after 500 hours of aging. This can be explain by the fact that after 500 hours  
9 the paint binder is completely degraded and cannot further influence the photocatalytic  
10 process.

11 For CNC paints, the uptake coefficients of *m*-xylene are independent of the aging process.  
12 Thus, the hybrid CNC-nano-TiO<sub>2</sub> plays a protective role of the binder, limiting the contact  
13 between the OH radicals and *m*-xylene.

### 14 **3.7. By-products formation**

15 In theory, the photocatalytic oxidation (PCO) should lead to the total mineralization of  
16 pollutants in CO<sub>2</sub> and H<sub>2</sub>O [35]. Unfortunately, the PCO is known to produce volatile  
17 aldehydes and other partially oxidized by-products that can be released back to indoor air  
18 [24,40]. Formaldehyde, acetaldehyde, acetone, formic acid et acetic acid were produced due  
19 to the incomplete mineralization [41]. Formaldehyde and acetaldehyde have been observed as  
20 by-products during *p*-xylene degradation [42], whereas formaldehyde, acetaldehyde and  
21 acetone during toluene degradation [43]. The mixing ratios of *m*-xylene, formaldehyde,  
22 acetaldehyde and acetone in position P<sub>0</sub> and P<sub>4</sub> are shown in Figure 8.



1  
 2 **Figure 8:** Temporal profile of formaldehyde (red line), acetaldehyde (green line), acetone  
 3 (blue line) and *m*-xylene (dark line) mixing ratio (ppb) at temperature: 296 K, RH: 37%, light  
 4 intensity: 18.6 W m<sup>-2</sup>, aging: 0 hour, photocatalytic paint: 7% of nano TiO<sub>2</sub> PC500. In the part  
 5 (A), injector is in position P<sub>0</sub> and in part (B) in position P<sub>4</sub>.

6  
 7 In this experiment approx. 13 ppb of *m*-xylene has been degraded in position P<sub>4</sub>. If 100% of  
 8 this compound is converted, a 2.55·10<sup>10</sup> molecules cm<sup>-2</sup> s<sup>-1</sup> of VOCs would have been formed.  
 9 In order to compare *m*-xylene consumption and by-products formation, we normalized  
 10 observed mixing ratio to exposed paint surface area and reaction time to obtain the surface  
 11 emission fluxes. For the mixing ratio and surface emission fluxes calculation, we used the  
 12 method presented in detail by Morin *et al.* [23]. The surface emission fluxes of formaldehyde,  
 13 acetaldehyde and acetone due to the reaction between molecules present in the binder and  
 14 radical species produced by the photocatalyst (nano-TiO<sub>2</sub>) is 1.77·10<sup>11</sup>, 9.33·10<sup>10</sup>,  
 15 1.07·10<sup>10</sup> molecules cm<sup>-2</sup> s<sup>-1</sup>, respectively. The sum of surface emissions fluxes for these three  
 16 compounds is 2.81·10<sup>11</sup> molecules cm<sup>-2</sup> s<sup>-1</sup>. The additional emission of VOCs due to *m*-xylene  
 17 degradation would correspond to an increase of about 9% of the surface emission flux.  
 18 Moreover, a part of degraded *m*-xylene is mineralized to CO<sub>2</sub> and H<sub>2</sub>O [43,44]. Debono *et*  
 19 *al.* [43] observed a mineralization up to 76% for toluene that decreased the quantity of VOCs

1 potentially formed at 2% or  $5.81 \cdot 10^9$  molecules  $\text{cm}^{-2} \text{s}^{-1}$  which represents the uncertainty of  
2 the measurements. For this reason, no difference has been observed in the mixing ratios of  
3 this by-product between the position  $P_0$  and  $P_4$  (Figure 8); hence, the formed by-products  
4 cannot be determined in this study.

5 To identify and quantify the formed products due to the degradation of *m*-xylene, the same  
6 approach as Debono *et al.* [43] should be applied based on pure nano-TiO<sub>2</sub>. This method will  
7 allow to determine the formed by-product but it will not be relevant from environmental point  
8 of view. Surely, many efforts are needed to optimize the photocatalytic paints and establish as  
9 a safe option for removal of indoor pollutants.

#### 10 **4. CONCLUSION**

11 The influence of nano-TiO<sub>2</sub> content, relative humidity, light intensity, surface temperature,  
12 pigment volume concentration, type of nanoparticles and aging, on the degradation of *m*-  
13 xylene on photocatalytic paints have been assessed under UV irradiation. The obtained results  
14 contribute to improve the knowledge on a poorly documented area that is the removal of  
15 VOCs by photocatalytic paints in indoor air. The results are crucial for the development of an  
16 optimal formula for photocatalytic paint as a function of environmental and physical  
17 parameters.

18 The increase of nano-TiO<sub>2</sub> content and the PVC improve the removal of *m*-xylene by a factor  
19 of 3. The application of innovative CNC nanoparticles do not impact significantly the  
20 photocatalytic activity of the paint in comparison of widely used TiO<sub>2</sub> nanoparticles. Paint  
21 aging has no influence on removal of *m*-xylene by paint containing CNC photocatalyst but the  
22 removal efficiency increased for the paints containing PC 500 nano-TiO<sub>2</sub>. However, the aging  
23 of the paint implies environmental problem because of the emission of nanoparticles in the

1 air. This subject is out of the scope in this study but deserves an attention from a healthcare  
2 point of view.

3 The increase of light intensity improved the removal efficiency of *m*-xylene by a factor 5. In  
4 contrast, high surface temperature and RH slowed down the degradation of *m*-xylene.

5 In order to assess the impact of photocatalytic paints on indoor air quality, we suggest a  
6 simple calculation to compare the lifetime of *m*-xylene through the major removal pathways  
7 in an indoor environment [45]. The consumption of *m*-xylene by photocatalytic paint is  
8 considered for a model room of 50 m<sup>3</sup> with a constant irradiated wall surface of 3 m<sup>2</sup>. From  
9 the experimental results from this study, the removal rate of *m*-xylene is estimated to be  
10  $7.3 \cdot 10^{-6} \text{ s}^{-1}$  for a non-aged photocatalytic paint containing 7% of nano-TiO<sub>2</sub> with a PVC of  
11 72% at 40% of RH under a spectral irradiation intensity of  $8.8 \text{ W m}^{-2}$  and a surface  
12 temperature of 298 K. The lifetime associated to this removal pathway is approximatively  
13 38 hours while the lifetime due to the oxidation by gas-phase OH radicals (assuming a mean  
14 OH concentration of  $5 \cdot 10^5 \text{ radicals cm}^{-3}$  [46] and bimolecular rate constant of  $k_{\text{OH}+m\text{-xylene}} = 2.3 \cdot 10^{-11} \text{ molecules cm}^{-3} \text{ s}^{-1}$  [47]) is 24 hours, and 2 hours for a typical air exchange  
15 rate of  $0.5 \text{ h}^{-1}$  [48]. Under these conditions, the main removal pathway of *m*-xylene is air  
16 exchange rate which account for more than 99% of its elimination. The photocatalytic  
17 degradation and OH oxidation account only minor contribution to the removal of *m*-xylene  
18 from the indoor air.

19  
20 The future formulation of an optimized photocatalytic paint should consider the influence of  
21 the physical (quantity of TiO<sub>2</sub> nanoparticles, type of photocatalyst, and PVC), and  
22 environmental (RH, light intensity, surface temperature) parameters on the degradation of  
23 indoor air pollutants. In the future, photocatalytic oxidation must be studied under visible  
24 irradiation because 45% of solar irradiation is in visible region (400-700 nm) against only 4%  
25 for in the UV region of wavelengths.

## 1   **ACKNOWLEDGMENTS**

2   The authors acknowledge gratefully LABEX SERENADE (no.ANR-11-LABX-0064) funded  
3   by the French National Research Agency (ANR) through the PIA (Programme Investissement  
4   d’Avenir). S.G. acknowledges the funding support by National Natural Science Foundation of  
5   China (No. 41773131, No. 41977187).

## 6   **ASSOCIATED CONTENT**

### 7   **Supporting Information**

8   Typical experiment of *m*-xylene degradation (Figure S1), loss of gas phase *m*-xylene as a  
9   function of exposure time (Figure S2).

## 10   **Notes**

11   The authors declare no competing financial interest.

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