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New route to toxic nitro and nitroso products upon irradiation of micropollutants mixtures containing imidacloprid: role of NOx and effect of natural organic matter

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6 **ABSTRACT**

7 In this study we reveal the capacity of imidacloprid (a neonicotinoid insecticide) to 8 photoinduce nitration and nitrosation of three aromatic probes (phenol, resorcinol, tryptophan) 9 in water. Using a gas-flow reactor and NOx analyzer, the production of gaseous NO/NO₂ was demonstrated during irradiation (300-450 nm) of imidacloprid (10⁻⁴ M). Quantum calculations 10 showed that formation of NOx proceeds via homolytic cleavage of the RN-NO₂ bond in the 11 triplet state. In addition to gaseous NO/NO₂, nitrite and nitrate were also detected in water, 12 with the following mass balance: 40±8% for NO₂, 2±0.5% for NO, 52±5% for NO₃⁻ and 13 $16\pm 2\%$ for NO₂. The formation of nitro/nitroso probe derivatives was evidenced by high 14 resolution mass spectrometry and their yields were found be ranging between 0.08% and 15 5.1%. The contribution of NO_3^{-}/NO_2^{-} to the nitration and nitrosation processes was found 16 minor under our experimental conditions. In contrast, the addition of natural organic matter 17 (NOM) enhanced significantly the yields of nitro/nitroso derivatives, likely via production of 18

triplet excited states (³NOM*) and HO⁻. These findings reveal the importance of investigating
the photochemical reactivity of water contaminants in mixture to better understand the
cocktail effects on their fate and toxicity.

22 Graphical abstract



23 INTRODUCTION

Imidacloprid (1-[(6-chloropyridin-3-yl)methyl]-N-nitro-imidazolidin-2-imine, IMD, Scheme 24 1) is a widely used and effective neonicotinoid insecticide.¹ IMD, like other neonicotenoids, is 25 used in a variety of crops and its applications have been recently extended to domestic sector 26 and animal health.¹ Once released in the environment, it can reach the aquatic, atmospheric 27 and terrestrial compartments and be in contact with a lot of living organisms.²⁻⁴ IMD affects 28 the central nervous system of insects⁵ and there are increasing concerns about its deleterious 29 effects with confirmed toxicity for pollinators, especially bees.⁶ Toxicity or adverse effects in 30 other species has been also reported⁷⁻⁹. 31

32



IMD

desnitro-IMD

urea-IMD

Scheme 1: Molecular structure of IMD and its main photoproducts

Several studies have been performed to investigate the fate of IMD under solar light, mainly 33 in aqueous solutions^{1, 10-13} on plant leaves and cuticles¹⁴⁻¹⁶ and more recently on solid thin 34 films¹⁷. Desnitro -IMD (1-[(6-chloropyridin-3-yl)methyl]imidazolidin-2-imine) was reported 35 be the main photoproduct and the urea-IMD (1-[(6-chloropyridin-3-36 to yl)methyl]imidazolidin-2-one), a minor photoproduct (Scheme 1).¹⁶ The formation of these 37 by-products seems to imply the cleavage of the N-NO₂ bond; however, the fate of NO₂ is not 38 yet clear. Recently, the formation of gaseous nitrogen oxides during the irradiation of IMD at 39 the surface of a germanium attenuated total reflectance (ATR) crystal was investigated.¹⁷ The 40

authors detected nitrous oxide (N₂O) which was proposed to be formed by recombination of
detached NO₂ with the IMD fragments in the electronic ground state.

In previous studies, we showed that the fungicides chlorothalonil and thiophanate-methyl can 43 photoinduce the degradation of other pesticides during irradiation.^{18,19} These reactions can 44 take place in surface waters where fungicides are present together with a wide range of other 45 chemicals and on solid supports such as the surface of leaves where pesticides are often 46 applied in combination. However, the effect that a pollutant can have on another pollutant 47 remains largely overlooked in photochemical studies and little is known on the transformation 48 of photostable pollutants via light induced reactive intermediates of other co-pollutants. The 49 ability of IMD to release NO2 upon irradiation makes this compound important to study in 50 this context since nitration of chemicals have been reported in a lot of systems generating NO₂ 51 20-28 52

53 The goal of this study was thus to explore the formation of reactive NOx via IMD photolysis and their reactivity towards surrogates of water contaminants and typical moieties of natural 54 organic matter (phenol, resorcinol and tryptophan). In particular, we investigated the 55 formation of nitro and nitroso-derivatives, two potentially toxic categories of compounds, 56 from the three selected probes under polychromatic irradiation (300-450 nm) in the presence 57 58 of IMD. Experiments were carried out using a flow-tube reactor equipped with NOx analyzer whereas the formation of by-products was characterized using high resolution UHPLC-ESI-59 MS. Theoretical calculations were also performed to elucidate the mechanisms of NOx 60 61 formation. To the best of our knowledge, this is the first study reporting on the measurement of NOx during irradiation of IMD in water and on their role in nitrosation/nitration of other 62 water pollutants. Environmental implications of this light-induced indirect degradation are 63 discussed. 64



phenol

tryptophan

resorcinol

65 **Scheme 2**: Selected probes

66 MATERIALS AND METHODS

Chemicals and materials. Imidacloprid (Pestanal[®], analytical standard), resorcinol (purity 67 99%), phenol (purity \geq 99.5%), L-tryptophan (purity \geq 98%) and 2,4-dinitrophenylhydrazine 68 (DNPH, purity 97%) were purchased from Sigma-Aldrich and used as received. Sodium 69 70 nitrite (Rectapur 98%), sodium nitrate (Fluka, purity 99%), 2 and 4-nitrophenol (Fluka, purity \geq 99%) and 4-nitrosophenol (Aldrich-Chemie, purity 60% due to a content of water of 40%) 71 were also used without further purification. Suwannee River NOM (2R101N) was purchased 72 from IHSS. Water was purified using a reverse osmosis RIOS 5 and Synergy (Millipore) 73 device (resistivity 18 M Ω cm, DOC < 0.1 mg.L⁻¹). All solvents or other reactants were of the 74 highest grade available. 75

Irradiations. Two different irradiation devices were used. To monitor NOx formation in the 76 gas phase, device 1, was designed with a cylindrical Pyrex gas flow-reactor (0.65 L, length 27 77 cm and diameter 5.7 cm) containing 200 mL of IMD solution (10⁻⁴ M or 25.6 mg.L⁻¹) and 78 79 irradiated from the top with two fluorescent tubes (Sylvania F15W/BL 368, 438 mm×26 mm, 300-450 nm, λ_{max} at 365 nm, see Figure SI-1) placed at a distance of 5 cm. The rate of 80 incident light entering the solution was measured using a QE65000 (Ocean optics) radiometer, it 81 was equal to 1.3×10^{16} Photon.cm⁻².s⁻¹. The gas flow-reactor was connected to a NOx 82 83 chemiluminescence analyzer (Thermo Scientific i-42 NOx analyzer) as shown in Figure 1. Experiments were carried out under continuous flow of clean air, to reduce the residence time 84 of NOx and minimize NO₂ photolysis. Gaseous inlet of the reactor was connected to flow 85

controllers for N₂ and O₂ (model Brooks 4800 series) allowing to select the atmosphere under 86 which the irradiations were performed (pure N₂, pure O₂ or a 80/20 O₂/N₂ mixture) and to 87 adjust the flow rate to 0.7 L min⁻¹. Background levels of NOx in the gas inlet were negligible 88 (<0.4 ppbv). Before irradiation, the levels of NOx in the gas phase of the reactor were 89 monitored continuously every 10 seconds until reaching a steady state within few minutes 90 (Figure SI-2). Despite that NO₂ concentrations up to 40 ppbv were detected initially (first 91 30s), a background level of NOx (<5 ppbv) was quickly reached. This production of NO₂ 92 might possibly be due to IMD photolysis under indoor lighting during preparation and 93 handling of the solutions. During IMD irradiation, we have chosen to monitor NOx levels in 94 95 the gas phase of the reactor at selected times and not continuously. After selected irradiation times, the light was turned off and the outlet of the reactor connected to the NOx analyzer. 96 NOx were measured every 10 s for a period of 1-2 min at a flow rate of 0.7 L min⁻¹. The time 97 98 profile of NOx concentrations shown in Figure SI-2 corresponds to the decay of NOx level in the reactor due to the dilution with clean air. The level of NOx measured was relatively 99 100 constant at 1h, 2h, 4h and 6h of IMD irradiation, in accordance with a steady state regime. 101 The levels of NOx generated were then obtained by integrating the time profiles of NOx concentrations. For calculating the yields of NOx, the ratio of NOx concentrations (ppbv) 102 103 were converted into molar by taking into account the gas phase volume of the reactor (450 104 mL). By measuring the converted quantity (in moles) of IMD during the irradiation period, it was possible to calculate the molar yield for NO and NO₂. 105



106 Figure 1. Schematic representation of device 1 used to monitor NOx formation upon107 irradiation of IMD solutions.

108 A second device, device 2, was used to monitor the formation of nitro and nitroso-derivatives. 15 mL of solutions were placed in a cylindrical Pyrex reactor sealed with an air-tight silicon 109 cap and surrounded by six fluorescent tubes (Sylvania F15W/BL 368), installed inside a 110 111 custom-made cylindrical irradiation device. Solutions were not buffered, but pH was controlled during the reactions. The initial pH was 6.5±0.3 and in the course of the reactions, 112 pH laid above 5. In the case of IMD+resorcinol mixtures, aliquots were sampled after several 113 selected irradiation times and the consumption profile of IMD and nitro- and nitroso-114 resorcinol were obtained. For IMD+phenol and IMD+tryptophane mixtures and experiments 115 with nitrate/nitrite and NOM, aliquots were only sampled after 16 h of irradiation. Samples 116 were analyzed by HPLC to monitor the loss of IMD and the probes and by UHPLC-HRMS to 117 characterize photoproducts and estimate their levels. For the experiment at 10⁻⁵ M of IMD and 118 10⁻⁴ M of resorcinol, 75 ml were irradiated by portions of 25 ml, and water was evaporated up 119 to 6 ml using a rotavapor before further analyses. Experiments were done in duplicate or 120 triplicate. Analytical methods. Absorption spectra were recorded using a Varian Cary 3 121 122 spectrophotometer. IMD, phenol, resorcinol and L-tryptophan concentrations were monitored by HPLC (Waters Alliance 2695) equipped with a photodiode array detector (Waters 2996) 123 and a EC150/4.6 Nucleodur 100-5 C₈ endcapped column. The HPLC conditions are presented 124 in Table SI-1. was quantified using the same instrument after derivatization with DNPH.²⁹ 15 125

mL of derivatizing solution were prepared by mixing concentrated HCl, ultrapure water and
acetonitrile in the ratio 2:5:1. 50 µL of this solution were added to 5 mL of sample solution.
Derivatization reaction was completed in 5 min and HPLC analyses were run shortly after.
Calibration used derivatized NO₂⁻ solutions in the concentration range 10⁻⁶-10⁻⁵ M. All the
HPLC analyses were done in triplicate and presented data are the mean values obtained.

NO₃⁻ was measured by ionic chromatography on a Dionex ICS-5000⁺ using a column model 131 Dionex Ion Pack AS11 2mm×250mm; the flow was 0.25 mL.min⁻¹ and the mobile phase was 132 an aqueous solution of KOH at the concentration of 0.43 mM for the first 4.5 min followed by 133 a linear gradient up to 18 min of runtime by rising KOH concentration up to 11.70 mM. The 134 injection volume was 750 µL and the temperature of the column oven was set at 30°C. 135 External standards were prepared using NaNO₃ in the concentration range 50-200 μ g.L⁻¹. 136 Nitro and nitroso-derivatives of phenol, resorcinol and L-tryptophan were characterized and 137 138 their concentrations estimated by high resolution mass spectrometry (HRMS) performed on an Orbitrap Q-Exactive (Thermo Scientific) coupled to an ultra-high performance liquid 139 140 chromatography system (UHPLC) Ultimate 3000 RSLC (Thermo Scientific). Due to the lack available standards, desnitro-IMD and minor photoproducts were tentatively identified by 141 HRMS but quantification was not possible. Analyses were carried out both in negative (ESI) 142 and positive (ESI⁺) electrospray modes. The column was the same column as for HPLC-UV. 143 The binary solvent system was composed of acetonitrile and acidified water using formic acid 144 at 40% and 60% respectively with a flow of 1.0 mL min⁻¹. Nitro and nitroso-phenols were 145 quantified by injection of external standards of 4-nitro and 4-nitroso-phenols at concentrations 146 varying from 5×10^{-7} to 4×10^{-6} M (Figure SI-3). Due to commercial unavailability of nitrated 147 derivatives, the concentrations of nitro and nitroso-resorcinols were estimated using the 148 calibration curves of 4-nitro and 4-nitroso-phenols. The concentrations of nitro-tryptophan 149 and nitroso-tryptophan were estimated using L-tryptophan at the concentration of 10^{-6} M as a 150

reference. Yields were obtained by dividing the photoproducts concentrations by the loss ofIMD in concentration and by multiplying these ratios by 100.

Computational Method. Potential Energy Surfaces (PES) were investigated with the 153 Gaussian series of programs.³⁰ Density Functional Theory (DFT) calculations (optimizations, 154 frequency calculations, identification of transition states and IRC calculations) were 155 performed at the MN12SX/6-311++G(d,p)//MN12SX/6-31+G(d,p) level of calculation. 156 Among the large set of hybrid range-separated functionals that we tested, MN12SX³¹ gave the 157 best agreement between experimental and theoretical absorption spectra. Finally, the time 158 dependent-DFT (TD-DFT) method was used to compute singlet excited states. Triplet state 159 was explored with conventional self-consistent field calculations. Solvent was modeled 160 implicitly with the Polarizable Continuum Model (PCM). Electronic energies are reported 161 throughout (Gibb's free energies are not provided here because low vibrational modes may 162 163 introduce large uncertainties in this calculation).

164

165 **RESULTS AND DISCUSSION**

Photolysis of IMD. The degradation profile of neutral aqueous IMD (10⁻⁴ M) irradiated in device 1 is shown in Figure 2. HPLC-UV and UPLC-HRMS analyses indicated that the main photoproduct was desnitro-IMD (Figure SI-4). Three minor photoproducts were also found: a compound produced by addition of an oxygen atom on IMD (IMD+O), another one produced by addition of two oxygen atoms on desnitro-IMD (desnitro-IMD+2O) and the nitroso derivative of IMD (IMD-O) as traces (Table SI-2).





Figure 2. Consumption profile of aqueous IMD (10^{-4} M) irradiated in device 1 (A) and yields of NO₃⁻ (white bars) and NO₂⁻ (shaded bars) (B).). Error bars are standard deviations.

175

176 NO₃⁻ and NO₂⁻ were detected in the aqueous phase (Figure SI-4). The yield of NO₃⁻ increased 177 constantly to reach 72±5% after 50 h of irradiation (Figure 2B, white bars). In contrast, the 178 yield of NO₂⁻ decreased from 27% after 7 h of irradiation to 6% after 50 h (Figure 2B, shaded 179 bars). The higher accumulation of NO₃⁻ compared to that of NO₂⁻ is consistent with the 180 spectral properties and the known photoreactivity of these ions. NO₃⁻ (λ_{max} = 300 nm and ϵ = 8 181 M⁻¹.cm⁻¹) absorbs much less solar light than NO₂⁻ (λ_{max} = 355 nm and ϵ = 22 M⁻¹.cm⁻¹) and its 182 quantum yield of photolysis is also lower (reactions 1-4) leading to a smaller photolysis rate³²:

183
$$\operatorname{NO}_3^-$$
 + hv \rightarrow NO_2 + $\operatorname{O}^ \Phi \approx 0.01$ (1)

184	NO ₃ ⁻	+	hν	\rightarrow	NO_2^-	+	$O(^{3}P)$	$\Phi\approx~0.001$	(2)
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185
$$NO_2^- + hv + H^+ \rightarrow NO + HO^- \Phi = 0.025$$
 (3)

186 $NO_2^- + h\nu \rightarrow NO_2 + e_{aq}^- \Phi < 0.0019$ (4)

- 188 The detection of IMD+O and of desnitro-IMD+2O appear to be consistent with the formation189 of hydroxyl radicals through reactions (1) and (3).
- 190

NOx formation upon irradiation of IMD. We also measured the NOx formation in the 191 gaseous phase above the solutions. NO and NO₂ were successfully detected upon irradiation 192 of IMD (10⁻⁴ M) in device 1. Examples of collected data are shown in Figure SI-2. During the 193 first two hours of reaction of irradiation, 24 µg of NO₂ and 0.88 µg of NO were produced 194 whereas IMD loss was around 380 µg (See Table 2). After integrating nitrite and nitrate levels 195 produced and considering the volume of gas (200 mL) and of water (450 mL), an overall mass 196 balance of IMD conversion into inorganic N-containing products was determined: 40±8% for 197 NO₂, $2\pm0.5\%$ for NO, $52\pm5\%$ for nitrate and $16\pm2\%$ for nitrite. 198

Table 2. Formation of NOx after 2 h of irradiation of IMD (10^{-4} M) or mixtures NO₃⁻/NO₂ in device 1.

Conditions	Quantity of IMD	Quantity of NO ₂	Quantity of NO
	converted	formed	formed
		(molar yield)	(molar yield)
$IMD (10^{-4} M)$	380±20 μg	24±5 µg (40±8%)	0.88±0.20 µg (2±0.5%)
NO ₃ ⁻ (1.8×10 ⁻⁶ M)			
+		1.6±0.3 µg	3.9±0.8 µg
NO_2^{-1} (1.7×10 ⁻⁶ M)			
$NO_3^- (2 \times 10^{-5} \text{ M})$			
+		$4.8{\pm}1.0~\mu g$	11.6±2.3 µg
NO_2^{-} (5×10 ⁻⁶ M)			

As the presence of NO_3^{-}/NO_2^{-} in the aqueous solution could potentially generate NO_x^{-33} , we 201 irradiated these ions independently to estimate their contribution to the pool of NO_x detected 202 upon irradiation of IMD. We chose to perform these comparative experiments with NO_3^- = 203 1.8×10^{-6} M and NO₂⁻ = 1.7×10^{-6} M corresponding to the concentrations estimated after 2 h of 204 irradiation in device 1 (Figure SI-4). The irradiation of this mixture for 2 h yielded 15-fold 205 less NO₂ and 4.4-fold more NO than the irradiation of IMD (10⁻⁴ M) (Table 2). The amount of 206 NO_x formed above the IMD solution was 4.5-fold higher than that above the NO_3^{-}/NO_2^{-} 207 solution showing that IMD was the main contributor. Moreover, the higher yield of NO in the 208 209 latter system confirmed that the NO_x precursor in the NO₃⁻/NO₂⁻ mixture was NO₂⁻ through reaction (3). At higher concentrations of NO₃⁻/NO₂⁻ (2.0×10^{-5} M and 5×10^{-6} M, respectively) 210 not far from those reached after 16h of IMD irradiation, NO and NO₂ levels were 3-fold 211 higher than in the previous case, respectively, showing that, in this concentration range, the 212 NO and NO₂ amounts accumulated proportionally to NO₂⁻ concentration. Again, the amount 213 of NO_x generated from NO₂⁻ (2.0×10⁻⁵ M) was still below the one formed upon irradiation of 214 215 IMD solution.

As previously shown, the bond N-NO₂ was cleaved when IMD was irradiated in the solid phase at 305 nm and 254 nm.¹⁷ However, some questions remain unanswered: would the proposed mechanism in the solid phase and at short wavelength be still valid in solution and at longer wavelength, and thus at lower excitation energy? How is NO formed and how the oxygen atom is eliminated? In an attempt to answer these questions, we performed quantum calculations.

Theoretical Calculations. Several conformers were found for the solvated imidacloprid system. The structure of the global minimum was similar to that found from recent firstprinciples^{17,34} and crystallographic investigations³⁴ (see SI section for all structures reported in this work). Moreover, the DFT method used here was able to reproduce accurately the experimental absorption spectrum (see Figure 3).

228



229

230 Figure 3. Experimental (Abs, left axis) and theoretical (oscillator strengths, right axis) absorption spectra of imidacloprid (theoretical spectrum for the most stable conformer (Fig. 231 SI-5a) at the TD-MN12SX/6-311++G(d,p),PCM//MN12SX/6-31+G(d,p),PCM level. The 232 oscillator strength is the probability of absorption. It is related to the dipole strength of the 233 transition and is proportional to the molar absorptivity maximum for this transition. The S_1 - S_0 234 transition is optically active (284 nm). The oscillator strength of this transition is 0.02 which 235 is significant (an ε of about 6000 L.mol⁻¹.cm⁻¹ can be estimated from this value when a UV-236 Vis Half-Width at half-height of 0.333 eV is used to model the absorption spectrum). 237 238

Detailed investigation of the ground state (S_0) , first triplet state (T_1) and first singlet excited state (S_1) was performed. Some of the key results are shown in Figure 4 and discussed below. Minima and transition state structures are given in the SI section. The dissociation pathways of interest are depicted in this Figure: NO_2 is obtained in the left-hand side of Figure 4, and triplet atomic oxygen (a reaction that is precursor to the production of NO) is shown on the right-hand side of Figure 4. The Minimum Energy Path was followed to investigate the formation of NO_2 in S₁ by performing partial optimizations of imidacloprid for fixed RN- NO_2 bond distances. The MEP maximum is indicated by a dashed level in Figure 4.





Figure 4. The electronic energies of stationary points in the ground state (S_0 , red), first triplet (T₁, green) and first singlet (S_1 , blue) excited states of imidacloprid (¹RNNO₂). Calculations were performed with the MN12SX/6-311++G(d,p),PCM//MN12SX/6-31+G(d,p),PCM method. Transition states and the MEP maximum are represented with dashed levels. Dashed bonds represent long-range separations. The energies of separated fragments **RN**+**NO**₂ and

¹**RNNO**+³**O** are also given in this Figure. The energy of the heterolytic dissociation products (left-hand side, black) and the triplet state of **RNNO** (right-hand side, purple) are also indicated. The separated fragments **RN** and **NO**₂ in S₁ are respectively in the first doublet excited state (**RN**) and doublet ground state (**NO**₂).

258

 S_0 and S_1 states. Our results show that departure of atomic triplet oxygen is not favored in the 259 ground state because the predissociated intermediate ¹RNNO-O (R(O-O)=1.523 Å; 56.4 260 kcal.mol⁻¹) cannot be reached due to very large activation energy (106.1 kcal.mol⁻¹). 261 Similarly, the formation of NO₂ is not expected in the ground state because the respective 262 electronic dissociation energy in S_0 is large (64.4 kcal.mol⁻¹; **RN**+**NO**₂). A transition state 263 could not be found between imidacloprid (${}^{1}RNNO_{2}$) and separated radicals in S₀. This is 264 expected for homolytic dissociations in the ground state. In S₁, the dissociation of the RN-265 NO₂ bond required 14.0 kcal.mol⁻¹ (MEP maximum). In addition, the in-cage dissociation 266 complex ¹**RN--NO**₂ is less stable (85.3 kcal.mol⁻¹) than the respective reactant (79.3 kcal.mol⁻¹) 267 ¹). Thus, dissociation in S_1 is less favored than that in the triplet state (an activation energy of 268 6.1 kcal.mol⁻¹ was found in T_1 and the reaction is now exothermic; see Figure 4). Moreover, 269 our results show that the heterolytic dissociation mechanism would lead to species RN^+/NO_2^- 270 that are slightly less stable (see also SI) than the respective separated radicals (RN/NO₂). 271 Thus, it seems reasonable to consider that it is the radicals that are eventually obtained in the 272 liquid phase. However, considering the small energy difference between heterolytic and 273 homolytic dissociation energies, a mixture of homolytic and heterolytic dissociation products 274 might not be completely ruled out. Implicit solvation models underestimate the solvation free 275 energy of small ions, the cases of the proton and hydroxide being emblematic of this issue. In 276 order to address this point, we computed $\Delta G^*_{r,calc}$ the dissociation Gibbs free energy for both 277 heterolytic and homolytic NO₂ cleavage. In this calculation, RT×ln(24.46) was added to 278

 $\Delta G^*_{r,calc}$ to account for the free energy change of 1 mol of an ideal gas from 1 atm to 1M.³⁵ $\Delta G^0_{r,calc} = \Delta G^*_{r,calc} + RT \times ln(24.46)$ was 58.6 and 49.2 kcal.mol⁻¹ for, respectively, the heterolytic and homolytic dissociation reactions, thus a difference of 9.4 kcal.mol⁻¹ (recall that the respective difference of electronic energies was smaller, i.e. 4.6 kcal.mol⁻¹). Considering the implicit solvation model may underestimate the solvation free energy for NO₂⁻ by as much as 10 kcal.mol⁻¹, the radical dissociated species are still more stable than the heterolytic disociation fragments, in agreement with our results.

286

Triplet State T_1 . Second, S_1/T_1 intersystem-crossing (isc) is favored because minimum 287 energy geometries in S_1 and T_1 are very similar for **RNNO**₂ and the energy gap is small (less 288 than 0.1 kcal.mol⁻¹ at the Franck-Condon geometry). In addition, in the triplet state, the barrier 289 for NO_2 dissociation (6.1 kcal.mol⁻¹) is significantly smaller than that for the dissociation into 290 triplet atomic oxygen and singlet fragment (${}^{1}\mathbf{RNNO} + {}^{3}\mathbf{O}$; 11.4 kcal.mol⁻¹), in agreement with 291 our experimental findings. Moreover, the in-cage dissociation complex ³RNNO--O is 292 thermodynamically less stable than the respective reactant ${}^{3}RNNO_{2}$ and this reaction is 293 294 displaced toward the reactant. Nevertheless, this pathway was investigated below to see if this reaction could play a role, even minor, in the formation of NO radicals. 295

296

NO formation, a minor pathway. The mechanism for production of nitric oxide from 1 RNNO is also shown in Figure 4. NO bond dissociation energy was 55.7 kcal.mol⁻¹ for 1 RNNO in the ground state, whereas cleavage in the triplet state (reached through S₁ followed by *isc*) required only 14.8 kcal.mol⁻¹ (**RN** is obtained). A transition state could not be found in T₁, and the energy of activation may be identified with the dissociation energy in that case. Thus, the formation of NO as a primary product from RNNO is probably a minor pathway. Thus, our quantum calculations provide strong evidence that photolysis of IMD mainly leads to NO_2 formation in the overall triplet state potential energy surface while NO formation is a minor pathway.

306

Evidence of nitro/nitroso derivatives formation. Nitrate and nitrite ions which generate HO[•] 307 radicals simultaneously with NO or NO₂ have been reported to photoinduce the nitration 308 and/or the nitrosation of phenolic derivatives.²⁰⁻²⁴ Whether the photolysis of IMD could also 309 contribute to the formation of such toxic products was a hypothesis that we aimed to verify in 310 this study. Thus, IMD was irradiated with each of the three probes shown in Scheme 2 in a 311 small reactor (device 2) for 16 h. In these conditions, the loss of IMD irradiated alone reached 312 60% (Figure SI-7). We first compared the probes consumption (10^{-4} M) in the absence and in 313 the presence of IMD (10^{-4} M) to quantify the photoinductive effect (Table SI-4). The 314 315 disappearance of phenol and resorcinol were drastically faster in the presence of IMD, evidencing a strong effect. Only in the case of tryptophan, the acceleration effect was small 316 317 due to the fast photolysis of the probe under the studied conditions.

318 Then, we monitored the formation of nitro/nitroso-derivatives by UHPLC-HRMS analyses. Figure SI-8 shows the formation profile of nitroso and nitro-resorcinol upon irradiation of 319 IMD (10^{-4} M) and resorcinol (10^{-4} M). The concentration of the two photoproducts increased 320 linearly up to 7h of irradiation before reaching a plateau value after 16 h. Nitro and nitroso-321 derivatives were also detected with phenol and tryptophan. Detected levels after 16h of 322 irradiation are given in Figure 5 (white bars) and Table SI-4. The estimated yields of nitro-323 resorcinols reached 3.3% of converted IMD and that of nitroso-resorcinols 5.1%. The yield of 324 nitro-phenols $(ortho+para \text{ derivatives})^{21}$ was equal to 2.6% and that of nitroso-phenols 325 (ortho+para derivatives) smaller (0.5%). For tryptophan, both yields were very small, 326

reaching only 0.08% (nitro-derivatives) and 0.14% (nitroso-derivatives) of tryptophan
converted, due to the fast direct photolysis of tryptophan.



Figure 5. Concentrations of nitro and nitroso-derivatives of phenol, resorcinol and tryptophan detected after 16 h of irradiation of the probes (10^{-4} M) in device 2 in the presence of IMD (10^{-4} M) (white bars) and in presence of the mixture of NO₃⁻ (2×10⁻⁵ M) and NO₂⁻ (5×10⁻⁶ M) (shaded bars). Error bars are standard deviations.

Experiments were also performed at lower reactants concentrations to determine the impact of 334 this parameter on the rates. A 10-fold decrease of the concentration of IMD for a 335 concentration of resorcinol kept constant at 10⁻⁴ M, reduced the amount of nitro-resorcinols by 336 18-fold and that of nitroso-resorcinols by 9-fold while the sum of the two photoproducts was 337 reduced by about 10-fold (Table SI-4). A 10-fold decrease of the concentration of resorcinol 338 for a concentration of IMD kept constant at 10⁻⁴ M, decreased 3-fold the amount of nitro-339 340 resorcinols and 100-fold that of nitroso-resorcinols and the sum (nitro+nitroso) was reduced by about 5-fold (Table SI-4). These results indicate that the reaction rate was proportional to 341 the concentrations of reactants. The decay rate of resorcinol followed a first order kinetic with 342 a rate constant k that can be expressed as : $k = 5.9 \times 10^{-3}$ [IMD] s⁻¹. 343

We also studied the nitro/nitrosation capacity of the NO_3^-/NO_2^- mixture to quantify their 344 contributions in these reactions (shaded bars in Figure 5 and Table SI-4). For these 345 comparisons, we fixed NO_3^-/NO_2^- concentrations to those measured in the IMD solution after 346 16h of irradiation in device 2 (2×10^{-5} M and 5×10^{-6} M, respectively) and the mixtures NO₃⁻¹ 347 $/NO_2^-$ + probes were irradiated for 16 h in device 2. Nitrosation was 121-fold lower in the 348 case of phenol, 4.1-fold in the case of resorcinol and 2.6-fold in the case of tryptophan than in 349 the presence of IMD while nitration was 5.9-fold lower for phenol, 4.1-fold for resorcinol and 350 1.5-fold for tryptophan. One concludes that the contributions of NO_3^-/NO_2^- to the 351 nitration/nitrosation processes observed in irradiated IMD solutions were minor especially for 352 phenol and resorcinol even though we acknowledge that the conditions chosen for the 353 comparative experiments (fixed NO_3^-/NO_2^- concentrations) cannot perfectly reproduce the 354 dynamic and complex evolution of NO_3^-/NO_2^- during IMD irradiation. 355

356

Effect of NOM on nitration/nitrosation of probes. Another important aspect we attempted 357 to explore is the effect of NOM on these photochemical nitration/nitrosation. Experimental 358 conditions and results are reported in Table 3. NOM (11 mg.L⁻¹) increased the percentage of 359 resorcinol loss in the presence of IMD by a factor of 2.6, after correction for the light 360 screening effects (10%). The formation of nitroso and nitro-resorcinols in IMD solutions were 361 increased by a factor of 1.3 and 2.0, respectively, and the formation of nitro and nitroso-362 resorcinols in NO₂⁻ solutions by factors of 35 and 2.8, respectively. The very low formation of 363 nitro-resorcinols in the system NO_2^- +resorcinol was already shown.²⁰ These data highlight 364 the significant effect of NOM on the nitro/nitroso-derivatives formation. 365

Table 3. Effect of NOM (11 mg.L^{-1}) on the photo-nitration/photo-nitrosation of resorcinol

368 (10^{-4} M) after 2 h of irradiation in device 2 in the presence of IMD (10^{-4} M) or NO₂⁻ $(5 \times 10^{-6} \text{ M})$

369 M). The light screening effect of NOM (10%) is taken into account.

Conditions%Nitroso-Nitro-resorcinolderivativesderivativesconvertedformedformed(μg.L ⁻¹)(μg.L ⁻¹)IMD +resorcinol7.0±0.383±8				
resorcinolderivativesderivativesconvertedformedformed(μg.L ⁻¹)(μg.L ⁻¹)(μg.L ⁻¹)IMD +resorcinol7.0±0.383±844±4	Conditions	%	Nitroso-	Nitro-
converted formed formed (μg.L ⁻¹) (μg.L ⁻¹) (μg.L ⁻¹) IMD +resorcinol 7.0±0.3 83±8 44±4		resorcinol	derivatives	derivatives
(μg.L ⁻¹) (μg.L ⁻¹) IMD +resorcinol 7.0±0.3 83±8 44±4		converted	formed	formed
IMD +resorcinol 7.0±0.3 83±8 44±4			$(\mu g.L^{-1})$	$(\mu g.L^{-1})$
	IMD +resorcinol	7.0±0.3	83±8	44±4
IMD +resorcinol+NOM 18±1 110±10 88±9	IMD +resorcinol+NOM	18±1	110±10	88±9
NO_2 +resorcinol <1 20±2 1±0.1	NO ₂ ⁻ +resorcinol	< 1	20±2	1±0.1
NO_2 +resorcinol+NOM <1 55±5 35±3	NO ₂ ⁻ +resorcinol+NOM	< 1	55±5	35±3

370

Reaction mechanism. Many reactions are expected to take place in this complex system. The 371 main ones are summarized below. The photolysis of IMD in water generates NO₂ in the liquid 372 phase (NO₂₍₁₎) together with the radical RN as shown by the quantum calculations and this 373 latter gives rise mainly to desnitro-IMD (reaction 5). The radicals $O(^{3}P)$ and RNNO are also 374 produced but in much lower amounts (reaction 6). The reactivity of NO₂ in water is very high. 375 It is reported in the literature that NO_3^- and NO_2^- are generated according reactions (7a and 7b) 376 through the intermediary dimerisation of $NO_{2(l)}$. ^{36,37} NO_3^- and NO (NO_(l)) can also be 377 produced through reaction (8).^{37,38} In our experiments, $N_2O_{4(g)}$ cannot be distinguished from 378

NO_{2(g)}. Reaction (8) is probably the major NO formation pathway, but the photolysis of RNNO (reaction 11), of NO_{2(g)} (reaction 12) and of NO₂⁻ (reaction 3) are alternative pathways for NO formation. NO is stable in water and volatilizes in the gaseous phase (reaction 9).³⁷ It generates NO_{2(g)} by reaction with oxygen (reaction 10).

In this scheme, we did not take into account the oxidation of IMD by HO⁻formed in reaction (3) because the formation of oxidation products (IMD+O) was small. NO_2^- and NO_3^- could be also oxidized by HO^{-, 39} However, due to the high concentration of IMD compared to those of NO_2^- and NO_3^- in the first half of the reaction, IMD and its main photoproduct desnitro-IMD are expected to be the main sinks of HO⁻.

388

389	$IMD + h\nu \rightarrow NO_{2(l)} + RN$ (5)
390	$IMD + hv \rightarrow O(^{3}P) + RNNO$ (6)
391	$2NO_{2(l)} \leftrightarrow N_2O_{4(l)} \tag{7a}$
392	$N_2O_{4(1)} + H_2O \rightarrow NO_3^- + 2H^+ + NO_2^-$ (7b)
393	$3NO_{2(l)} + H_2O \leftrightarrow 2NO_3^- + 2H^+ + NO_{(l)}$ (8)
394	$NO_{(l)} \leftrightarrow NO_{(g)}$ (9)
395	$2NO_{(g)} + O_2 \rightarrow 2NO_{2(g)} $ (10)
396	RNNO + $h\nu \rightarrow NO_{(l)}$ + radical (11)
397	$NO_{2(g)} + h\nu \rightarrow NO_{(g)} + O(^{3}P)$ (12)

398

The nitration/nitrosation of phenolic/aromatic compounds in the presence of NO_2^- or NO_3^- has been already studied. ^{20-25,39-41} NO₂ is recognized as having a key role in these reactions, even though the mechanisms are not fully understood. Several studies attributed the formation of phenoxyl radicals to the reaction of phenols with $NO_2^{-20,25,39,41}$ while others to their reaction

with HO^{. 40}. Once formed phenoxyl radicals further react NO₂ to yield the ortho and para 403 derivatives of nitro-phenols. Similarly, the reaction of phenoxyl radicals with NO are 404 expected to generate nitroso-phenols. The effect of NOM on these reactions is difficult to 405 predict due to the existence of opposite reactions. Indeed, on the first hand, NOM can 406 potentially favor nitro/nitrosation reactions by generating photooxidants such as triplet excited 407 states (³NOM^{*}), singlet oxygen or HO[•] that are expected to induce the formation of phenoxyl 408 radicals as shown in the literature.³⁹ In addition, ³NOM* or HO[•] could be able to oxidize NO₂⁻ 409 and NO_3^{-1} into NOx^{39} (see process 13 related to NO_2^{-1}): 410

411

412
$$NO_2^- + {}^3NOM^*/HO^- \rightarrow NO_2^- + NOM^-/OH^-$$
 (13)

413

On the other hand, NOM could be also a sink for NOx due to the presence of aromatic 414 moieties in its structure and therefore could inhibit the nitration/nitrosation reaction to some 415 extent.^{26,42} The overall effect of NOM will thus depend on the experimental conditions and in 416 417 particular on the relative concentrations of reactants. In our set-up, the presence of NOM (11 mg.L⁻¹) in IMD or NO₂⁻ solutions containing resorcinol (10^{-4} M) increased the amounts of 418 nitro and nitroso-resorcinols showing that in this case NOM favors the formation of the 419 phenoxyl radical and of NO2. Since Suwannee River NOM contains 68% of fulvic acids and 420 fulvic acids contain \sim 3 mmol oxidizable phenol per g of carbon⁴³, there are about 20 μ M of 421 oxidizable phenols originating from NOM in a solution containing 11 mg/l. Under these 422 conditions where the quenching molecule is in large excess over phenolic moieties in NOM, it 423 can be expected that the majority of NOM-derived photooxidants will be quenched by the 424 probe compounds to form phenoxyl radicals, which can then further react with NOx to 425 produce nitro/nitroso-phenol/resorcinol, enhancing the yields of these products as it is 426 observed. 427

428

429 Environmental significance. Our study showed that NOx (NO and NO₂) could be generated by irradiation of IMD and that these NOx were capable of reacting with phenolic probes to 430 produce nitro and nitroso-derivatives. This suggests that in surface water, IMD could induce 431 the formation of nitro and nitroso derivatives of other contaminants that could be toxicants in 432 the aquatic environment. The importance of the phenomenon will depend on the level of IMD 433 present in surface waters. The frequency of IMD detection in surface waters is high and 434 concentration of IMD varies in a large range reaching 1 μ g.L⁻¹ in agricultural areas.^{3,4} Based 435 436 on this work and on the light absorption capacities of IMD and NO₂, one can calculate that at 1 μ g.L⁻¹, IMD could generate as much as nitroso/nitro-derivatives as NO₂⁻¹ at 0.1 μ g.L⁻¹. The 437 relative contributions of NO_2^- and IMD to nitroso/nitro-derivatives formation depend 438 therefore on their levels in water. NO_2^- is present in surface waters at a maximal concentration 439 of 0.1 mg·L⁻¹. ⁴⁴ In many cases, NO₂⁻ concentration will be high enough to make negligible 440 the contribution of IMD in nitro/nitrosation reactions. However, in waters containing very low 441 levels of NO₂⁻ and high levels of IMD, IMD might play a role in the nitro/nitrosation 442 reactions. 443

On the other hand, we observed that NOM $(11 \text{ mg} \cdot \text{L}^{-1})$ enhanced nitro and nitrosation 444 reactions of phenols (10⁻⁴ M). The yield of nitro-resorcinol was multiplied by 2 whereas that 445 of nitroso-resorcinol by 1.3 and when NO2⁻ was used instead of IMD, the yields of nitro-446 resorcinol and of nitroso-resorcinol were 35-fold and 2.8-fold increased respectively. This can 447 be explained by the enhanced formation of phenoxyl radicals in the quenching of NOM 448 deriving photooxidants by phenols. However, this effect might be concentration-dependent. 449 The level of NOM used in this study falls in the range of typical NOM levels found in rivers 450 and eutrophic lakes⁴⁵, but those of probes are high. At environmentally relevant 451 concentrations of the probe compounds (<1 µM), phenolic moieties in NOM are in excess and 452

it might be possible that the NOM-bound phenoxyl radicals, in higher concentration than 453 probe -derived phenoxyl radicals might outcompete the latter for reaction with NOx. In other 454 words, at environmentally relevant conditions, a shift towards nitration/nitrosation of NOM 455 456 seems possible and the yield of nitro/nitroso-derivatives of probes could then be lower compared to the ones in the absence of NOM. The incorporation of N inorganic nitrogen (N) 457 into NOM via photolysis of nitrate/nitrite and/or in advanced oxidation processes treatments 458 was already reported and shown to generate also potentially toxic compounds.²⁶⁻²⁸ Therefore 459 such reactions might have also environmentally negative consequences. We could not detect 460 any significant structural changes in NOM using HRMS analyses. Nevertheless, further 461 462 experiments using N-labelled IMD or nitrite/nitrate and other hyphenated techniques may provide evidence of NOM nitration/nitrosation. Therefore, a consistent monitoring of 463 nitro/nitroso byproducts is highly recommended to verify their potential formation. 464

Our study confirms the capacity of water contaminants to interact with each other under 465 irradiation and to induce mutual degradation. Mutual effects can involve reactions between 466 467 excited and ground states contaminants or the intermediary formation of reactive species like NOx in the case of IMD. Up to now, they have been poorly investigated in photochemical 468 studies and would deserve more attention. Thus, future studies should further consider the 469 investigation of the "cocktail effect" on the environmental fate of contaminants and more 470 specifically the reactivity of intermediates in order to enable a more reliable monitoring of 471 non-target pollutants and assessment of potential risks. 472

473

474 ASSOCIATED CONTENTS

475 Supporting information:

476 15 pages, 8 Figures and 4 Tables

Spectrum of light received by the solutions in devices 1 and 2. NOx formation in device 1 477 upon photolysis of IMD (10⁻⁴ M). Calibration curve for nitroso-phenol and for nitro-phenol. 478 Profile of NOx, NO₃, NO₂, nitroso-resorcinol and nitro-resorcinol formations and of IMD 479 consumption. Structure and cartesian coordinates of the minimum energy structures of IMD in 480 S_0 (a), S_1 (b) and T_1 (c), the TS (d), intermediate RNNO-O (e) in S_0 , TS_1 in T_1 (f), TS_2 in T_1 481 (g), ³RN-NO₂ (h), ³RN-NO₂ in T₁ (i), ¹RNNO (j) and ³RNNO (k) at the B3LYP/6-482 311++G(d,p)//B3LYP/6-31+G(d,p) level. Minima and transition state structures with 483 cartesian coordinates. Photo-nitration and photo-nitrosation of the probes after 16 h of 484 irradiation in presence of IMD (10^{-5} or 10^{-4} M) or a mixture of NO₃⁻ (2×10⁻⁵ M) and NO₂⁻ 485 $(5 \times 10^{-6} \text{ M})$. Effect of NOM (11 mg L⁻¹) on the photo-nitration/photo-nitrosation of resorcinol 486 (10^{-4} M) after 2 h of irradiation in device 2 in the presence of IMD (10^{-4} M) or NO₂⁻ $(5 \times 10^{-6} \text{ M})$ 487 M). 488

489

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