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1 **Photodegradation of benzisothiazolinone: identification and biological** 2 **activity of degradation products**

3 Zsuzsanna Varga, Edith Nicol and Stéphane Bouchonnet*

4 Laboratoire de Chimie Moléculaire – CNRS / Ecole polytechnique, Institut Polytechnique de

5 Paris – 91128 Palaiseau, France

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7 **1. Introduction**

8 1,2-benzisothiazol-3(2H)-one (referred to as benzisothiazolinone or BIT, is an antimicrobial 9 agent used as a preservative in mainly aqueous-based chemical chemicals. Banned from 10 cosmetic products, BIT can be found in paints, lacquers, polishes, printing inks, emulsions, 11 cleaning agents, disposable gloves; it is also widely used as a preservative for pesticides 12 (Schwensen et al., 2015; Rafoth et al. 2007; Flyvholm, 2005; Aalto-Korte et al., 2007). Due to 13 environmental considerations, water-based paints recently replaced solvent-based paints, 14 which require the addition of biocides; isothiazolinones are widely used for this purpose 15 (Amsler et al., 2017). BIT is also used in paper-based jointing but it degrades during sunlight 16 exposure losing its biocidal efficiency (Lugg, 2001).

17 Isothiazolinone-type biocides act against various fungi and bacteria by the active sulfur 18 moiety able to oxidize thiol functional groups to form disulfides (e.g. with cysteine) (Collier 19 et al., 1990). Besides the biocidal effects on unwanted compounds by reacting with 20 intracellular sulfur-containing proteins (Collier et al., 1990; Morris et al., 1984), 21 isothiazolinones have also lethal effects on living organisms, the median lethal concentration 22 (LC50) of BIT for zebrafish embryos being 1.03 mg/L (Li et al., 2016). Growth inhibition 23 effect on Escherichia coli and Schizosaccharomyces pombe was measured; this inhibitory 24 activity was quenched by thiol-containing materials (glutathione, cysteine), confirming the 25 mechanism of interaction (Collier et al., 1990). According to the European Chemicals 26 Agency, BIT is very toxic to aquatic life, causes eye damage and skin irritation (ECHA, 27 2019). Toxicity tests on zebrafish embryos were performed during ozonation processes and a 28 decrease was found in the overall toxicity of the mixture, due to oxidation of the reductive 29 sulfur in the parent compound $(Li$ et al., 2016). BIT is a developmental toxicant in rats but 30 shows negative mutagenicity USEPA, 2005). Isothiazolinones were shown to cause severe 31 contact dermatitis and allergy in humans mainly with a high level of occupational exposure 32 (Bregnbak et al., 2013; Bang Pedersen, 1976).

33 It has been shown that isothiazolinones degrade rapidly in soil (Bollmann et al.., 2017). 34 Indeed, BIT is easily degraded in the environment but there is no extensive study on the 35 structure of its transformation products, their persistence, and their toxicity. Conventional 36 wastewater treatment plant efficiently removes isothiazolinones at low concentrations (Rafoth 37 et al., 2007). In reverse-osmosis (RO) wastewater treatment plants, BIT is used thanks to its 38 compatibility and non-oxidizing nature and it prevents biofilm growth on membranes (Tang et 39 al., 2012). Nevertheless, studies demonstrated that some isothiazolinones increase the 40 genotoxicity of the RO concentrates (Tang et al., 2013). BIT is added at high concentrations 41 (around 160 mg/l) for which the removal efficiencies of existing approaches are low; 42 ozonation was proposed as an efficient way to degrade BIT (Li et al., 2016). Considering the 43 photochemical degradation, the kinetics was studied under irradiation with a low-pressure 44 UV-C lamp. The process was described to follow pseudo-first order kinetics. Radical species 45 quenching studies were carried out, where the photodegradation of BIT was concluded to be 46 both direct, due to UV absorbance of BIT and indirect, induced by radical oxygen species 47 formation, a widespread mechanism of micropollutant photodegradation (Wang et al., 2017).

48 Different analytical methods were developed for the quantitation of BIT in various matrices. 49 Ultrasonic assisted extraction combined with LC-MS was used to determine BIT at 9.9 mg/kg 50 in hygienic consumer products, specifically in liquid detergents (Heo et al., 2019). BIT was 51 analyzed by matrix solid-phase dispersion followed by LC-MS/MS in household products 52 where it is used as a substitute to chlorinated isothiazolinone; it was found at quite high 53 concentrations in laundry detergent (0.0255% w/w), surface cleaner cream (0.0116 %w/w) 54 and liquid surface cleaner (0.00377% w/w) (Alvarez-Rivera et al., 2012). BIT was found in 55 94% of 47 paints analyzed by LC-MS at concentrations ranging from 28.6 to 1110.74 ppm. 56 These analyses also showed that these products were sometimes labeled incorrectly, e.g. one 57 paint labeled "preservative-free" contained BIT at 71.5 ppm (Goodier et al., 2018). 58 Isothiazolinones were monitored and not detected in environmental waters using 59 derivatization of BIT followed by pre-concentration and GC-MS analysis, with limits of 60 detection ranging from 0.01 to 0.1 µg/L (Rafoth et al., 2007).

61 This work aimed at performing accurate structural elucidation of the photodegradation 62 products of BIT using chromatography and tandem mass spectrometry approaches. 63 Benzisothiazolinone was degraded using two different types of photoreactors in water. 64 Structural elucidation of BIT photoproducts allowed to suggest a coherent degradation 65 pathway on one hand, and to carry out *in silico* toxicity tests to estimate the toxicity of the 66 photoproducts on another hand.

67

68 **2. Materials and methods**

69 *2.1. Chemicals, reagents and sample preparation*

3

70 1,2-benzisothiazol-3(2*H*)-one (CAS: 2634-33-5), benzamide, salicylamide, 4- 71 hydroxybenzamide and chromatography grade solvents: acetonitrile (ACN), methanol, 72 dichloromethane and formic acid (FA) were purchased from Sigma Aldrich (Saint-Quentin-73 Fallavier, France). A Purelab Chorus 1 water purification system (Veolia, Wissous, France) 74 was used to produce ultra-pure water with a conductance of 18.2 S/cm. BIT solutions were 75 prepared at different concentrations: 5, 10, 25, 50 and 100 ppm in water (solubility 1.1 g/l at 76 20 °C (SCCS, 2012) and in acetonitrile. Some experiments were carried out with degassing of 77 the initial solution using nitrogen bubbling for 30 minutes. Different irradiation times were 78 tested to determine the appearance order of photoproducts. Solutions were irradiated for 6 79 hours. Sampling was carried out every 10 minutes during the first hour and then every hour 80 until the end. For GC-MS analysis, all samples (ACN solutions and aqueous samples) were 81 dried using an Xcelvap automated evaporation/concentration system (Horizon Technologies, 82 Bullion, France) and 1 ml of dichloromethane was added. For LC-MS measurements, the 83 samples of different BIT starting concentrations were diluted, in a mixture made of 50% 84 ACN, 50% H₂O and 0.1% formic acid to reach a final concentration of 5 ppm. For HRMS 85 analysis in the infusion mode, samples were diluted in a 90% water / 10% ACN / 0.1% FA 86 mixture, for a starting concentration of BIT of 5 ppm. All the samples were stored away from 87 light at 4 °C; HPLC-MS data showed that degradation was negligible under these conditions, 88 in agreement with literature data where the hydrolytic stability of the compound is mentioned 89 (USEPA, 2005).

90 *2.2. Irradiation experiments*

91 The UV-Vis absorption spectrum of BIT was recorded in the range of 190 to 500 nm, using a 92 6800 UV-Vis Jenway spectrophotometer (Cole-Parmer, Villepinte, France), with a 1 cm 93 length quartz cell. Based on this spectrum, two irradiation setups were used for

94 photodegradation experiments. The photodegradation of BIT was mainly studied by a Philips 95 HPL-N 125W/542 E27 SC high-pressure mercury lamp (France-Lampes, Saint-Cirq, France) 96 with a wavelength range of 200-650 nm, a maximum irradiation wavelength at 254 nm, and a 97 radiation flux of 6200 lm. The setup consisted of one lamp, circularly surrounded by 6 quartz 98 tubes of 120 ml, all of the elements being in a water bath with water circulation so that the 99 reaction temperature was kept below 28 °C. The samples were also irradiated using a Q-sun 100 test chamber Xe-1-B/S (Q-Lab Saarbücken, Germany) equipped with a xenon arc lamp; a 101 natural light filter X-7640 from the same manufacturer has been used for the reproduction of 102 full sunlight spectrum. The lamp power was 1800 W and the irradiation 0.68 W.m⁻² at 103 a black-standard temperature of 55 °C.

104 *2.3. GC-MS analysis*

105 Gas chromatographic separations were carried out on an Agilent 450-GC instrument equipped 106 with a 60 m "FactorFour VF-Xms" (10% phenyl, 90% methylpolysiloxane) capillary column, 107 with an internal diameter of 0.25 mm and film thickness of 0.25 µm (Agilent Technologies, 108 Les Ulis, France), coupled with an Agilent 240-MS ion trap mass spectrometer. Samples were 109 injected in splitless mode using an Agilent CP-8400 autosampler, the injection volume was 110 1µl and injector temperature was 280 °C. A gradient temperature program was used with an 111 initial temperature of 50 °C, held for 0.5 min then increased to 280 °C with a ramp of 10 112 °C/min. The carrier gas was high purity helium, with a flow of 1.4 ml/min. Measurements 113 were carried out in electron ionization and positive chemical ionization using methanol as the 114 reagent gas. The filament emission current was set to 10 µA and automatic gain control was 115 used for all experiments. The electron multiplier voltage was automatically optimized at 2200 116 V for a gain value of 10^5 . In the full scan mode, ions were scanned on the range of m/z 50 to 117 m/z 500. Tandem experiment measurements were performed with an ionization storage level 118 of m/z 35 and an isolation window of 3 m/z, using a resonant waveform type with an 119 excitation storage level of m/z 50 and an excitation amplitude of 0.8 V.

120 *2.4. LC-MS analysis*

121 LC separations were carried out on an Acquity HPLC system (Waters Technologies, 122 Guyancourt, France) coupled with a SolarixXR FT-ICR mass spectrometer equipped with a 123 9.4 T superconducting magnet (Bruker Daltonics, Bremen, Germany). A Pursuit XRs^{Ultra} C18 124 column (2.8 µm x 50 mm x 2 mm) (Agilent Technologies, Les Ulis France) was used in 125 gradient mode. Water+0.1% FA (solvent A) and acetonitrile+0.1% FA (solvent B) were used 126 as solvents with a flow rate of 0.2 ml/min. The gradient started at 95% of solvent A for 3 min, 127 then changed linearly to 50% of solvent A in 9 min and then to 5% of solvent A after 12 min 128 of LC run. It was kept at 5% of solvent A for 5 minutes and then set back to the initial 129 percentage of solvent A (95%) for 5 min of equilibrium. An electrospray ionization source 130 was used in both positive and negative modes with a sample flow of 200 µL/min. for both 131 positive and negative mode the nebulizer and drying gas was nitrogen with a flow of 8 L/min 132 at 250 °C for drying gas and 1 bar for nebulizer gas. The capillary voltage was set at - 4500 V 133 and 4000 V in both positive and negative modes, respectively; the spray shield was set at – 134 500 V. The detection mode was broadband with a 4 Mpt resolution in the range of 57.5 to 700 135 m/z with a data reduction of noise of 95%. For tandem MS experiments, precursor ions were 136 isolated with an isolation window of 1 m/z and dissociated with collision energies of 10, 15, 137 20 and 25 V to study the fragmentation mechanisms. In direct infusion mode, the flow rate 138 was 120 µL/h. The spectra were acquired with 4 M data points on a range of 57.7 to 1000 139 m/z. Accumulation time was set to 0.02 s. 100 scans were recorded for each spectrum. The 140 Bruker DataAnalysis software was used to process the chromatograms and spectra.

141 *2.5. In silico bioassays*

142 Preliminary toxicity assessments were performed using the T.E.S.T. software developed by 143 the US Environmental Protection Agency (EPA). This software incorporates different 144 endpoints and approaches, with various individual predictions. The different individual 145 approaches include hierarchical clustering, the Food and Drug Administration (FDA) method, 146 single model, group contribution and nearest neighbor. (Martin, 2016) The reliability of the 147 results in the T.E.S.T. software should be assessed based on the capability of each individual 148 prediction to calculate the toxicity values and then study concordance of the results. A 149 complete description of T.E.S.T. and the included methods can be found on the EPA website 150 (Martin, 2016). Different QSAR methodologies are available to model mutagenicity. The 151 biological testing background of most of them, including T.E.S.T., is the Ames test, which is 152 based on a bacterial mutation of Salmonella typhimurium, reverting the bacteria from 153 histidine dependence to histidine independence (Mortelmans and Zeiger, 2000). The 154 correlation between carcinogenicity in animals and positive mutagenicity in the Ames test 155 was found to be high (Hengstler and Oesch, 2001). The QSAR models for mutagenicity of the 156 compounds of interest were found to be generally of high reliability.

157

158 **3. Results and discussion**

159 *3.1. Physicochemical analysis and photoproducts structural elucidation*

160 The UV/Vis absorption spectrum of BIT shows a high absorbance in the UV region and an 161 additional absorption band with a maximum at 319 nm (see Supplementary material, Fig. S1). 162 The photodegradation was evidenced by the decrease of the molecule's absorption bands 163 according to time (Fig. 1). A strong decrease in concentration can be observed during the first 164 15 minutes of irradiation, whereas the degradation slows down later on. This is in good 165 agreement with the exponential decrease evidenced by LC-MS experiments (see below).

166 *Fig. 1*

167 Whatever the chromatograph, using chromatography introduces some subjectivity in the 168 detection process. Actually, the choice of stationary and mobile phases induces some 169 selectivity regarding elution and thus detection of molecules. Therefore, direct infusion MS 170 was used prior to LC-MS coupling, to carry out non-targeted investigation of transformation 171 products in ESI-HRMS. The irradiated solutions were analyzed in both positive and negative 172 modes for the same reasons. The ions detected in direct ESI-HRMS were the same as those 173 detected in LC-MS. Retention times in LC-MS and GC-MS couplings, appearance times of 174 photoproducts during the irradiation process, exact m/z values of [M+H]⁺ and [M-H]⁻ ions 175 and corresponding formulae are listed in Table 1. Exact m/z values and corresponding 176 formulae of collision-induced product ions from $[M+H]^+$ or $[M-H]^-$ ions are given in 177 Supplementary material, Table S1.

178 *Table 1*

179 PP1 is among the first appearing photoproducts. It is eluted at 19.3 min in GC-MS; it has the 180 same mass as the parent compound (eluted at 17.8 min) and its EI mass spectrum shows a 181 very similar fragmentation pattern. Its exact formula (C7H5NOS) has been determined by HR-182 MS and confirmed that PP1 corresponds to an isomer of BIT. 1,3-benzothiazol-2(3H)-one 183 was postulated on the basis of PP1 mass spectrum. It was synthesized (the synthesis protocol 184 is reported in Supplementary material S1) and analyzed using GC-MS; both its retention time 185 and EI mass spectrum confirmed that PP1 corresponds to 1,3-benzothiazol-2(3H)-one (Table 186 1). Light-induced isomerization of BIT has been reported in an old study in which three 187 possible isomerization mechanisms were postulated but none was demonstrated (Darlage, 188 1971). Considering the rapid isomerization of BIT into PP1, consequent oxidation processes 189 will have to be considered from both isomers.

190 Three photoproducts (referred as PP2a, PP2b, and PP2c) were detected with retention times of 191 5.4, 6.9 and 8.3 min in LC-MS coupling. They all have a molecular weight of 167 (confirmed 192 by both EI and CI mass spectra in GC-MS) and have been assumed to result from oxygen 193 addition onto the parent compound. For the first one, the concerted losses of CO and SO (m/z 194 168.01196 \rightarrow m/z 96.04471 in CID measurements, Table S1) from the protonated molecule 195 allowed to conclude that the oxygen atom was added onto the sulfur one. The fact that the 196 nitrogen atom was retained in the m/z 96.04471 product ion permitted to conclude in favor of 197 the structure depicted in Table 1, which results from oxidation of PP1 (and not from BIT). 198 Both other PP2 isomers appear after 3 h of irradiation; they display exactly the same CID 199 dissociation pathways. Eliminations of HNCO and HS[.] from protonated molecules show that 200 the oxygen atom has been added on the aromatic ring. Hydroxylation of aromatic rings under 201 UV-Vis irradiation has been widely reported in the literature (Matsuura et al., 1974; Kinani et 202 al., 2016). Because it can play an important role regarding toxicity, the exact position of the 203 hydroxyl group has been determined based on CID experiments as depicted in Fig. S2. All the 204 possible structures resulting from hydroxyl addition on the aromatic ring either from BIT or 205 PP1 have been considered in a systematic way: only two of them can undergo two 206 consecutive eliminations of carbon monoxide from the protonated species in positive ESI; 207 their formulae are presented in Table 1.

208 The photoproduct referred as PP3 appears after only 10 min of irradiation. Its exact formula 209 corresponds to the addition of two oxygen atoms on BIT or PP1. In CID experiments, the 210 elimination of SO₂ from the deprotonated species (m/z 181.99191 \rightarrow m/z 118.02998, Table 211 S₁) leaves no doubt that both oxygen atoms have been added to the sulfur atom. Three 212 mechanisms could be responsible for the formation of PP3: straight addition of a dioxygen 213 molecule onto the sulfur atom from either BIT or PP1, or oxidation of PP2a. As this 214 photoproduct is no more observed when irradiating a degassed solution, it can be concluded 215 in favor of direct dioxygen addition on the sulfur atom. CID results do not permit to 216 discriminate between oxidation of either BIT or PP1 but the structures elucidated for PP4 to 217 PP8 photoproducts are strongly in favor of BIT oxidation, as discussed in the next section. It 218 is to be noted that the structure suggested for PP3 corresponds to saccharin, which has been 219 proved to be an oxidation product of BIT in a study dedicated to BIT ozonation (Li et al., 220 2016).

221 PP4 has been detected in negative ESI, at a very short retention time (1.2 min) in reverse-222 phase chromatography suggesting a highly polar structure. Its mass spectrum exhibits a major 223 ion resulting from deprotonation of the molecule. For these reasons, the structure suggested 224 for PP4 in Table 1 includes a SOOH group, which allows easy removal of a proton in 225 negative ESI. The $SO₂$ elimination in CID experiments is in good agreement with the 226 proposed structure.

227 PP5 has been detected in both positive and negative ESI. Its elemental composition 228 corresponds to that of PP4 plus a hydroxyl group. In the negative mode, observation of the 229 deprotonated molecule indicates an acidic function (in agreement with the short eluting time 230 in LC-MS). In the positive mode, the consecutive losses of ammonia and SO_2 imply that the 231 amido function has been kept and that the sulfur atom carries two oxygen atoms. This implies 232 that the additional hydroxyl group is bound to the aromatic ring. It has been located in ortho 233 position with regard to the amido function since the formation of both PP2b and PP2c 234 photoproducts shows that hydroxylation in this position is favored over additions onto other

235 positions of the ring. The consecutive loses of SO_2 and H_2O from [PP5 - H] in CID 236 experiments (m/z 200.00250 \rightarrow m/z 118.02997) confirm this hypothesis (see the dissociation 237 pathways of $[PP5 - H]$ ⁻ in Fig. S3).

238 Three photoproducts referred as PP6a to PP6c were detected at short retention times (between 239 0.9 and 2.3 min) in negative ESI. Their exact formulae correspond to the addition of an 240 oxygen atom onto PP5. In CID experiments, the isomer eluting at 0.9 min undergoes losses of 241 NH3 and O=C=NH indicating that it still has the amido function; it demonstrates that the 242 aromatic ring has been hydroxylated. The ammonia elimination, barely observed in negative 243 ESI, has to be charge-induced. By analogy with other photoproducts and to rationalize the 244 NH3 loss (see Fig. S4), the hydroxyl group was assumed to be in ortho position relative to the 245 amido group. In the negative mode, the losses of HCNO and $SO₂$ from [PP6b - H] indicate 246 that two hydroxyl groups are carried by the aromatic ring in the PP6b structure. 247 Unfortunately, CID experiments provided two few ions to allow the location of these 248 hydroxyl groups. Both losses of $HCNO₂$ and $SO₃$ from $[PP6c - H]$ ⁻ allowed the easy 249 determination of the PP6c structure, as depicted in Fig. S4.

250 PP7 has been easily identified as benzamide based on GC-MS and LC-MS analysis of the 251 corresponding standard. As for PP7, PP8 isomers (PP8a to PP8c, see Table 1) were easily 252 identified since salicylamide and 4-hydroxybenzamide were commercially available and were 253 used to check mass spectra and retention times in GC-MS.

254 *3.2. Mechanistic approach of BIT photodegradation*

255 The photodegradation of BIT and the appearance of photoproducts were monitored using LC-256 MS in both positive and negative modes, at various concentrations ranging from 5 to 100 257 ppm. The detected photoproducts were the same regardless the initial BIT concentration, 258 although the reaction kinetics slightly differed, the photodegradation being slower at higher 259 concentrations, as displayed in Supplementary material, Fig. S5 and S6. At 5 ppm, BIT is 260 undectable after 30 minutes of irradiation while the initial amount of BIT has been reduced by 261 60% at 100 ppm for the same irradiation time. For each photoproduct, Table 1 gives the time 262 at which it is first detected and the time at which it is no longer detected. Based on the 263 kinetics of the appearance and disappearance of photoproducts, stepwise photodegradation 264 mechanisms were proposed (Fig. 2). Structures between brackets were not observed, likely 265 because of too short lifetimes. It is to be noted that the formation of sulfite ions, reported by a 266 previous study, was not evidenced in the present work but is in good agreement with the 267 observed decrease of the mixture pH (Li et al., 2016).

268 Fig. 2

269 Isomerization of BIT into PP1 is very fast; it is not possible to state if this isomerization is 270 partially reversible. Isomerization of other compounds including five-membered ring 271 structures (PP2 isomers) could also not be established but could explain the formation of PP4 272 from PP2a (see Fig. 2). From BIT or PP1, the formation of all the photoproducts can be 273 rationalized by successive or competitive mechanisms including direct oxygen addition onto 274 the sulfur atom (so that it reaches its highest oxidation state), hydroxyl addition, hydrolysis, 275 and SO_3 elimination. The comparison of results between degassed and non-degassed solutions 276 showed that the same photoproducts are produced in both cases. They are less abundant in 277 degassed solutions, for which BIT dimers are also observed. This is likely because a reduced 278 concentration in oxygen disfavors mechanisms involving dissolved oxygen and thus favors 279 intermolecular reactions between BIT excited species.

280 *3.3. In silico toxicity estimations*

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281 For each photoproduct, QSAR results on mutagenicity, oral rat and Fathead minnow toxicity 282 were considered to be compared to that of the parent compound. The results of *in silico* 283 *toxicity* estimations are summarized in Table 2. Since the structure of PP6b was not accurately 284 determined (the position of the second hydroxyl group on the aromatic ring could not be 285 stated, see Table 1), toxicity estimation could not be performed on this compound. Regarding 286 mutagenicity, T.E.S.T. works with a dataset of 5743 chemicals and provides a binary result 287 (positive/negative). Experimental values were provided by the software for BIT, PP3, PP7, 288 and PP8a; they were all in accordance with the calculated results. *In silico* estimations showed 289 non-mutagenicity for all the photoproducts except PP4. For this compound, only 2 out of 3 290 individual predictions provided any result and both results were contradictory, therefore the 291 mutagenicity positive result for PP4 is of low reliability; it might be a false positive result and 292 should be treated accordingly. Fathead minnow LC_{50} (96 h), represents the concentration that 293 kills half of the fish population (*Pimephales promelas*) in 4 days. The training set was 294 obtained from the ECOTOX database, with a dataset containing 823 chemicals. The parent 295 compound showed to be the most toxic but LC50 values could not be estimated for PP1, PP2b, 296 PP3, and PP6c: none of the individual models provided a result since the requirements for 297 reliable toxicity assessment were not fulfilled for these compounds. For PP8a (salicylamide) 298 the proposed experimental toxicity value was in good agreement with the calculated one but 299 for PP7, for which an experimental result was also available, the experimental value was 3.5 300 times higher than the calculated one. The oral rat LD_{50} endpoint gives the amount of 301 chemical, in terms of the mass of the chemical over weight of the rat, which kills half of the 302 rats when orally ingested. In this case, the ChemIDplus database was used to calculate 303 toxicity data of interest with a final dataset of 7413 chemicals. For this endpoint, there was no 304 prediction for photoproduct PP6a but the others proved to be of high reliability. Experimental 305 values showed good accordance with the calculated ones for the parent compound.

306 Photoproducts PP2a, PP2c, PP4, PP5, PP8b, and PP8c exhibited slightly higher or similar 307 toxicity as the parent compound. Compound PP5 fell into the moderately toxic category (50- 308 500 mg/kg) on the Hodge and Sterner scale (Hodge and Sterner, 1949). All the other 309 compounds were categorized as slightly toxic. For PP8a different oral rat toxicity values were 310 reported in the literature but generally, the experimental values were approximately 2 times 311 lower, closer to the toxicity values determined for PP8b and PP8c (Table 2) (Goldenthal, 312 1971; Boxill et al., 1958). It could explain the discrepancy among the PP8 compound family, 313 in which photoproducts display isomeric structures. Given that oral rat toxicity calculations 314 proved to be of adequate reliability, toxicophores of the critical compounds are further 315 discussed. The potential toxicity of PP2a cannot be compared to those of other compounds 316 since its structure includes a very specific and uncommon –SO-CO-NH- structure (Table 1). It 317 is interesting to note that, at the exception of PP2a, all the photoproducts exhibiting potential 318 toxicity include either a phenolic (PP2c, PP8b and PP8c) or sulfino (-SO₂H) group (PP4). 319 Furthermore, the most toxic compound (PP5) is the one including both functions. The toxicity 320 of phenolic compounds has been widely reported but that of the sulfino group is not discussed 321 in the literature, therefore no conclusion could be drawn but it might be interesting to 322 investigate further this class of compounds.

323 *Table 2*

324 **4. Conclusion**

325 The UV-VIS irradiation of an aqueous solution of benzisothiazolinone led to the formation of 326 fourteen photoproducts, their chemical structures were elucidated based on GC-MS, LC-327 MS/MS, and FT-ICR-MS experiments. Based on the chemical structures determined and their 328 appearance order, a photo induced-degradation mechanism of benzisothiazolinone was 329 proposed, which combines isomerization, oxidation, hydroxylation, hydrolysis, and 330 elimination processes. Regarding to toxicity estimations, the absence of mutagenicity of 331 photoproducts and the moderate oral rat LD₅₀ values determined for some of them (while 332 most of them are less toxic than the parent compound) leads to consider direct 333 photodegradation as a potential powerful tool to remove benzisothiazolinone, even when it is 334 used in large amounts in some wastewater treatment plants. This is even more interesting 335 considering that these photoproducts were not persistent after a few hours of UV-light 336 irradiation.

337

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345

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Figure 2. Suggested pathways for the photodegradation of BIT

Table 2. Toxicity values estimated by the T.E.S.T. software for BIT and its photoproducts except PP6b

NA^a: Non applicable