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1 Drought offsets the positive effect of summer heat waves on the canopy greenness of mountain

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5 M.C. Corona-Lozada¹, S. Morin² & P. Choler^{1,3}

- 6 ¹ Univ. Grenoble Alpes, Univ. Savoie Mont Blanc, CNRS, LECA, F-38000 Grenoble, France
- 7 ² Univ. Grenoble Alpes, Université de Toulouse, Météo-France, CNRS, CNRM, Centre d'Études de
- 8 la Neige, F-38000 Grenoble, France
- 9 ³ Univ. Grenoble Alpes, Univ. Savoie Mont Blanc, CNRS, LTSER Zone Atelier Alpes, F-38000
- 10 Grenoble, France
- 11
- 12 Corresponding author: Philippe CHOLER
- 13 philippe.choler@univ-grenoble-alpes.fr
- 14 Tel: +33 4 76 51 48 96
- 15 ORCID : 0000-0002-9062-2721
- 16

17 Keywords

18 Climatic water balance; European Alps; heat wave; MODIS; mountain grasslands; phenology.

19 Abstract (195 words)

20 Heat waves are becoming more frequent in a warming climate and understanding the impacts 21 of these extreme events on terrestrial ecosystems remains a major challenge. Seasonally snow-22 covered mountain grasslands are temperature-limited ecosystems and one may question whether the 23 summer heat waves there have the same negative effect on primary productivity as they have on 24 lowland ecosystems. Here, we examined the remotely-sensed phenology of mountain grasslands in 25 the French Alps with a particular focus on the four unusually warm summers that occurred in 2003, 26 2012, 2015, and 2016. Our results showed that an early and rapid senescence and a lack of late-27 season regrowth were the most significant responses in 2003 and 2012 when heat waves coincided 28 with climatic water deficit, i.e. negative anomalies of precipitation minus reference 29 evapotranspiration. Mountain grasslands located below 2350m were the most affected. In contrast, 30 we found low senescence and moderate to high regrowth in 2015 and 2016 which were years 31 unaffected by severe water deficit. When water was non-limiting, warm and even very warm 32 summers have a positive effect on the canopy greenness of the highest grasslands, particularly by 33 extending the green period in late summer and early fall. Our study highlights the pivotal role that 34 summer water balance has in the phenological response of mountain grasslands exposed to heat 35 waves.

37 Introduction

38 There has been a marked increase in air temperature in the European Alps from the 1980s 39 onwards and this has caused warmer and longer snow-free periods for mountain vegetation (Durand et al., 2009a; Gobiet et al., 2014; Beniston et al., 2018). A growing body of evidence indicates that 40 41 above treeline ecosystems have undergone significant changes over the last decades, including 42 changes in community composition (Gottfried et al., 2012; Cannone and Pignatti, 2014) and 43 ecosystem functioning (Choler, 2015). Many of these documented changes, such as the upward shift 44 of species (Lenoir et al., 2008; Steinbauer et al., 2018), the thermophilization of high summit flora 45 (Gottfried et al., 2012), or the long-term greening of vegetation at high elevation (Carlson et al., 46 2017), are consistent with the expected positive effects of rising temperatures in temperature-limited 47 ecosystems. Nonetheless, recent reports indicate that earlier frost exposure due to the earlier snow 48 melt-out date (Wheeler et al., 2014; Liu et al., 2018), more frequent heat waves (Cremonese et al., 49 2017) and droughts (Marchin et al., 2018) may be additional effects of on-going climate change 50 which could attenuate this positive response. These studies highlighted the need to consider both 51 trends and extreme events when assessing mountain vegetation response to climate change. 52 In a warming climate, the duration, intensity, and frequency of heat waves is expected to 53 markedly increase (Stocker et al., 2013). The few reports that examine the response of mountain 54 grasslands to summer heat waves have yielded contrasting results. Using both remote sensing and in 55 situ measurements, Jolly et al. (2005) found a positive effect of the 2003 heat wave on the summer 56 greenness of high-elevation ecosystems in the Swiss Alps, and this was in stark contrast with what 57 was happening in lowland ecosystems. Cremonese et al. (2017) used flux data and time-lapse 58 cameras to characterize the effect of the 2015 heat wave on a subalpine Nardus-rich community and 59 noted a reduction of canopy development in addition to advanced senescence. Experimental 60 approaches using monolith transplantation have highlighted that the negative effect of an increased 61 air temperature only occurred when it coincided with reduced water supply (De Boeck et al., 2016).

This calls for a comparative study that includes a broader range of temperature-precipitation
combinations to better capture the interplay of heat waves and drought on mountain grassland
phenology.

65 The remote sensing of vegetation phenology provides a unique opportunity to track ecosystem 66 responses to climate change at a regional scale and over decades (Pettorelli et al., 2005; Richardson 67 et al., 2013). In mountainous environments, several studies have successfully used high or medium 68 resolution satellite images to examine the relationships between climate and the growth response of 69 mountain grasslands (Fontana et al., 2008; Dunn and de Beurs, 2011; Zhang et al., 2013b; Choler, 70 2015). For seasonally snow-covered ecosystems, the use of snow and vegetation indices may provide 71 several phenological metrics including the onset of growth, the date and magnitude of maximum 72 canopy greenness, the dynamics of senescence, and the date of the first snowfalls. However, many 73 comparative studies (either inter-annual or inter-spatial comparisons) have focused on the maximum 74 or averaged summer greenness and, by doing so, have largely overlooked the late and end of season 75 phenology (Gallinat et al., 2015; Garonna et al., 2018). Yet extreme events occurring after the date of 76 peak standing biomass, such as heat wave or drought, may accelerate the senescence or diminish the 77 possibility of regrowth and this has important consequences for the management of mountain 78 pastures (Nettier et al., 2017). We believe that remotely-sensed phenological metrics that capture the 79 kinetics of both growth and senescence phases will provide a more comprehensive assessment of 80 mountain grassland responses to heat waves.

In this paper, we used time series of the Normalized Difference Vegetation Index (NDVI) for the period 2000 - 2016 to examine how the phenology of mountain grasslands has been impacted by the summer heat waves in 2003, 2012, 2015 and 2016. We performed our comparative analysis by using the NDVI from Moderate-Resolution Imaging Spectroradiometer (MODIS) and meteorological variables derived from the SAFRAN (Système d'Analyse Fournissant des Renseignements Atmosphériques à la Neige) reanalysis system (Durand et al., 2009a, 2009b). We

focused on the French Alps for which the availability of high-resolution land cover data allowed us to select MODIS pixels with predominant coverage of mountain grasslands. More specifically, we addressed the following three sets of questions: (1) what was the intensity, duration, and timing of summer heat waves in the French Alps from 2000 to 2016 and to what extent did these events coincide with climatic water deficit, (2) what are the remotely-sensed indicators that best capture the impact of extreme meteorological events on the phenology of mountain grasslands, and (3) how do heat wave and water balance interfere on the growth and senescence of mountain grasslands?

94

95 Materials and Methods

96 The distribution of mountain grasslands in the French Alps was derived from a 10 m resolution 97 land cover map of France produced by the Centre d'Expertise Scientifique Occupation des sols, 98 downloadable at http://osr-cesbio.ups-tlse.fr/~oso/ (Inglada et al., 2017). To obtain a binary map of 99 mountain grasslands, we aggregated the three land cover classes corresponding to meadows 100 ("pelouses", code 34), heathlands ("landes ligneuses", code 36) and grasslands ("prairies", code 101 211). We noticed that the class "landes ligneuses" included a limited number of pixels and was not 102 very reliable due to the difficulty in distinguishing mountain low shrubs from nearby grasslands and 103 meadows. We calculated the percent cover of the aggregated class at a 250 m resolution to align with 104 the MODIS data, selecting those pixels with >90% cover and a mean elevation ranging from 1350 m 105 to 2550 m m a.s.l.. This yielded a total of 44,295 pixels which fall into four elevation bands (1350 m 106 - 1650 m, 1650 m - 1950 m, 1950 m - 2250 m, and 2250 m - 2550 m) and 23 SAFRAN massifs (Fig. 107 1 and Table S1). Pixel-based estimates were averaged by elevation band and massif. This resulted in 108 a table containing 87 combinations of elevation band and massif and 17 years (2000 - 2016). Annual 109 standardized anomalies were calculated from this table.

For meteorological data, we used the SAFRAN reanalysis that simulates surface conditions as
a function of elevation in 300 m increments and for 23 massifs of French Alps. These massifs were

112 defined for their climatological homogeneity in a previous spatial analysis of all weather data 113 available for the French Alps (Durand et al., 2009a; 2009b). The SAFRAN reanalysis was evaluated 114 against in-situ observations by Durand et al. (2009a) and is used for many real-time and 115 climatological applications in the French mountain areas (e.g. Verfaillie et al., 2018). The period 116 from 1981 to 2010 was set as a reference period to estimate standardized anomalies. For each 117 MODIS pixel, we extracted all the required meteorological data to compute a Heat Wave Magnitude 118 index, hereafter HWI, and a Reference Crop Evapotranspiration, hereafter ET_0 as described below. 119 We simplified the methodology proposed by Russo et al. (2014) to calculate a HWI that 120 accounts for both magnitude and duration of heat waves. First, we identified days exhibiting a 121 maximum temperature above a threshold value. For each day of year (DOY) i, we assembled a 122 dataset (DS_i) comprising the daily Tmax values of the n days before and after day i for the 30 year-123 long reference period (1981-2010) according to Eq. 1

$$DS_{i} = \bigcup_{y=1981}^{2010} \bigcup_{i=n}^{i+n} Tmax_{y,i}$$
(1) 124
125

where the symbol U denotes the union of values. The 80th percentile of DS_i defined the 126 127 threshold value for day i. As in Russo et al. (2014), we chose n = 15 days to estimate the threshold. 128 We examined the sensitivity of HWI to varying values of n. The relative change in HWI was 129 predominantly less than 5% when comparing n=15 with n=13, and less than 10% when comparing 130 n=15 with n=11 (Fig. S1). In all cases, the ranking of years was the same and the main findings of 131 the study did not change. Second, we selected all episodes of at least three consecutive days during 132 for which the maximum air temperature exceeded the threshold and summed the daily maximum 133 temperature of each episode to obtain an unscaled magnitude of the heat wave. Finally, we summed 134 the unscaled magnitudes of all heat waves occurring during the three periods of interest (see below)... 135 Compared to Russo et al. (2014), we did not consider sub-heat waves and did not rescale heat wave 136 magnitude into probability values.

ET₀ is a Penman–Monteith estimate of potential evapotranspiration of a reference crop surface that quantifies the evaporating demand of the atmosphere when water is not limiting. We followed the recommendations of the Food and Agricultural Organization (Allen et al., 1998) to compute ET_0 (in mm day⁻¹) from weather data (Eq. 2).

$$ET_0 = \frac{0.408 \,\Delta \left(R_n - G\right) + \gamma \frac{900}{(Tm + 273)} \,u_2(e_s - e_a)}{\Delta + \gamma (1 + 0.34u_2)} \tag{2}$$

where R_n is the net radiation at crop surface (MJ m⁻² day⁻¹), G is the soil heat flux density (MJ 142 $m^{-2} day^{-1}$), Tm is mean daily air temperature at 2 m (°C), u₂ is the wind speed at 2 m (m s⁻¹), e_s is 143 144 saturation vapour pressure (kPa), e_a is the actual vapour pressure (kPa), Δ is the slope of vapour pressure curve (kPa $^{\circ}C^{-1}$), and γ is the psychrometric constant (kPa $^{\circ}C^{-1}$). The use of SAFRAN 145 146 reanalysis to compute ET₀ according to Eq. 2 is described in Vannier and Braud (2012). The cliamtic 147 water balance corresponding to the difference between precipitation (P) and reference crop 148 evapotranspiration, hereafter P-ET₀, was taken as a proxy to assess the dry–wet conditions of a given 149 period.

150 For each year, HWI and P - ET_0 were computed over three periods of interest: (i) the 151 climatological summer (June, July and August), hereafter JJA, (ii) the growth phase defined as the 30 152 days preceding the date of maximum observed NDVI, hereafter B30, and (iii) the senescent phase 153 defined as the 60 days following the date of maximum observed NDVI, hereafter A60 (Fig. 2). We 154 examined the sensitivity of the results to the length of the latter two periods and did not find any 155 significant effect if periods were increased or decreased by 10 to 15 days. 156 Our remote sensing analysis was based on the surface reflectance acquired by MODIS, a 157 sensor onboard Terra satellite. We downloaded the collection 6 of the MOD09Q1 products, corresponding to tile h18.v4 (40–50° N, 0–15.6° E), from the Land Processes Distributed Active 158 159 Archive Center (https://e4ftl01.cr.usgs.gov/). The 8-day composite at 250 m resolution, covering the 160 period from 18 February 2000 to 27 December 2016, was processed following Choler (2015). To

161 summarize, re-projected surface reflectance values (ρ) in the red and near-infrared (NIR) bandwidths 162 were used to calculate a NDVI according to Eq. 3, keeping only reflectance values of optimum 163 quality (according to the MOD09Q1 Quality Control flag). Following Choler & al. (2010), we first 164 removed outliers defined as NDVI values that were two times larger or smaller than the average 165 value of the two preceding and the two following observations. Such sudden changes in NDVI are 166 inconsistent with the growth or senescence dynamics of mountain grasslands and may be due to 167 quality control issues, ephemeral snowfall during summer etc. Only 0.2% of the total NDVI values 168 were discarded at this step. Missing values and removed outliers were gap-filled using a cubic spline 169 interpolation. We then applied a Savitzky and Golay low pass filter in which the original data value 170 at time t_i was replaced by the value of a cubic polynomial fitted for 2n+1 points, with n being the 171 number of time steps preceding or following t_i (Savitzky and Golay, 1964). Consistent with our 172 previous studies (Choler et al., 2010; 2015), we empirically found that a window size of seven days 173 (i.e. n=3) was able to track ecologically significant changes in NDVI while still able to reduce noise 174 in the NDVI time series efficiently. Finally, we used cubic splines to daily interpolate the NDVI time 175 series and to derive the following phenological metrics (Fig. 2): (i) the growth onset, hereafter Onset, 176 defined as the Julian day on which NDVI reached half its maximum, (ii) the NDVI maximum, 177 hereafter NDVImax, as a proxy for the maximum canopy greenness, (iii) the Julian day of 178 NDVImax, hereafter TNDVImax, as the date of the peak vegetation (iv) the absolute NDVI decrease 179 during the 60 days following NDVImax, hereafter SEN (A60), see Eq. 4 as a proxy for the canopy 180 senescence, and (v) the absolute NDVI increase during the 60 days following NDVImax, hereafter 181 REG (A60), see Eq. 5 as a proxy for the end-of-season regrowth...

$$NDVI = \frac{\rho_{NIR} - \rho_{Red}}{\rho_{NIR} + \rho_{Red}}$$
(3)

$$SEN_A60 = \sum_{i=TNDVImax}^{TNDVImax+59} \operatorname{abs}(\min(0, NDVI_{i+1} - NDVI_i)) \quad (4)$$

$$REG_A60 = \sum_{i=TNDVImax}^{TNDVImax+59} \max(0, NDVI_{i+1} - NDVI_i)$$
(5)

Pixel-based estimates of phenology were averaged by elevation band and massif in order to align with meteorological variables. For REG (A60) we transformed the raw values into binary data (zero or positive values) and we calculated the percent of pixels exhibiting regrowth for each combination of elevation band and massif.

187 Standard Major Axis (SMA) regression analysis was used to analyse the bivariate relationships 188 between anomalies of phenology and to test the statistical significance of any deviation of the slope 189 from unity and of the intercept from zero (Warton and Ormerod, 2007). We modelled anomalies of 190 NDVImax, SEN (A60) and REG (A60) as a linear function of anomalies of HWI and P-ET₀. The 191 model was written as follows (Eq. 6)

$$anom(Y_z) = \beta_{0,z} + \beta_{1,z} anom(P - ET_0)_z + \beta_{2,z}anom(HWI)_z + \varepsilon_z$$
(6)

192

193 where *Y* is the dependent variable (phenology), β are the regression coefficients, *anom* means 194 standardized anomaly and ε is an error term. The β parameters were solved separately for each 195 elevation band (z) using ordinary least squares regression.

All statistical analyses were performed in *R* (R Core Team, 2017), using the *signal* library
(signal developers, 2013) for the time series analyses and the *raster* library (Hijmans, 2017) for the

- 198 spatial analysis. All figures were produced using the *ggplot2* (Wickham, 2016) and the *rasterVis*
- 199 (Perpinan Lamigueiro and Hijmans, 2018) libraries.

201 **Results**

202 Since the 1980s, warm to very warm summers with above normal HWI have become more 203 frequent in the French Alps (Fig. 3A). The four summers with the highest HWI all occurred in the 204 last 15 years, i.e. during the years 2003, 2012, 2015 and 2016 (Fig. 3A). The intensity of summer 205 heat waves reached a maximum in 2003 with an HWI (JJA) median value of 184 for the 1950 m -206 2250 m elevation band (Fig. S2A). During the same period, we found a wide range of water balance 207 conditions during the summer period, with years when high values of HWI and negative $P-ET_0$ 208 values co-occurred and other years when they occurred independently (Fig. 3, Fig. S2B). Summer 209 droughts coincided with the 2003 and 2012 heat waves, while summer water balance was close to 210 normal in 2015 and 2016 (Fig. 3B). The French Alps also experienced a severe drought in 2009, but 211 this was not accompanied by a significant heat wave (Fig. 3B). 212 Figure 4 showed that the 2012 and 2016 heat waves mainly concerned the late season while 213 those of 2003 and 2015 occurred throughout the growing season. Noticeably, there was no negative 214 anomaly of P-ET₀ before the peak of biomass in 2015 (Fig. 4A) and only a moderate one after the 215 peak (Fig. 4B). By comparison, the 2003, 2012 and 2016 late season heat waves co-occurred with 216 dry conditions, with the most adverse situations taking place in 2003 and 2012 (Fig. 4B). 217 Plant growth onset in mountain grasslands typically occurred from the second half of April in 218 the 1350 m 1650 m elevation band (mean DOY of 112) to the first half of June in the 2250 m 2550 219 m elevation band (mean DOY of 157) (Table 1). The date of maximum canopy greenness followed 220 around 1.5 months later, with the length of the growing period (TNDVImax-ONSET) decreasing 221 with elevation (Table 1). Late season phenology demonstrated a much larger range of inter-annual 222 variability - with a coefficient of variation ranging from 18% to 31% for senescence and regrowth -223 as compared to those of Onset, TNDVImax and NDVImax (Table 1, Fig. 5), and this was observed 224 across all elevation bands (Table 1). Overall, year-to-year variations in phenology were much

smaller than those of HWI and $P-ET_0$ (Table 1).

The last four years (2013-2016) were among the five greenest summers of the study period, while the four strongest negative anomalies of NDVImax occurred before 2006 (Fig. 5D). The summers of 2003 and 2012 were characterized by an above normal senescence and a lack of regrowth, although the NDVImax was close to normal (Fig. 5D, E, F). A lack of regrowth was also observed in 2016, although senescence was close to normal and NDVImax was particularly high (Fig. 5D, E, F). By comparison, the year 2015 exhibited a below normal senescence and above normal regrowth and NDVImax (Fig. 5D, E, F).

233 Figure 6 shows the bivariate relationships among phenological anomalies and illustrates the 234 particularity of the four years marked by heat wave occurrence. We found a strong positive 235 relationship between anomalies of Onset and of TNDVImax (Table 2, Fig. 6A). The SMA slope was 236 significantly below one indicating that maximum canopy greenness was reached in a shorter time in 237 case of a delayed onset (Table 2). In addition, the two years with early season drought (e.g., 2003 238 and 2012) exhibited an earlier peak of NDVI than the one predicted by the SMA (Fig. 6A). The 239 relationship between TNDVImax and NDVImax showed that a delayed phenology tend to be 240 associated with higher values of NDVImax (Table 2, Fig. 6B). Remarkably, the two years with a 241 pronounced early season heat wave (e.g., 2003 and 2015) exhibited a higher anomaly of NDVImax 242 than predicted (Fig. 6B). Finally, we found a significant negative relationship between SEN (A60) 243 and REG (A60) (Table 2, Fig. 6C). The years 2003, 2012 and 2016 exhibited a late-season 244 phenology that was marked by normal to rapid senescence and limited regrowth while the year 2015 245 clearly showed an opposite pattern (Fig. 6C).

Finally, we examined the effects of HWI and $P-ET_0$ on phenology using multiple regression models. For this quantitative analysis, we distinguished between the two periods, i.e. before and after the peak, and among different elevation bands. Irrespective of elevation, there was a significant effect of $P-ET_0$ (B30) anomalies on NDVImax while no effect was detected for HWI (B30) (Table 3, Fig. 7A, B). Wetter conditions in the early season significantly enhanced the growth of grassland

251 while heat waves did not. P-ET₀ after the peak was also a significant driver of SEN (A60) and REG 252 (A60) and the magnitude of this effect markedly differed among elevation bands. Wet summers 253 significantly slowed down senescence and enhanced regrowth at low elevation, with up to 47% of 254 variance explained in SEN (A60) and 24% of variance explained in REG (A60) in the 1350 m 1650 255 m elevation band (Table 3, Fig. 7C, E). These effects vanished higher up and became non-significant 256 (SEN) or marginally significant (REG) above 2250 m (Table 3). Regarding HWI, we only found a 257 significant and negative effect on senescence above 2250 m (Table 3). Noticeably, the effect of late-258 season heat waves on senescence depended on elevation with an increased senescence at low 259 elevation (although not significantly) and a delayed senescence at high elevation (Fig. 7D). 260 Overall, these results indicated that maximum canopy greenness, senescence and regrowth 261 were much more responsive to summer water balance than to heat waves. This is consistent with our 262 initial hypotheses concerning the combined effects of heat wave and climatic water deficit on plant 263 phenology. An early heat wave that did not coincide with a severe drought - as in 2015 and 2016 -264 did not significantly impact NDVImax. A late heat wave coinciding with a severe drought - as in 265 2003, 2012 and 2016 - resulted in an increased senescence and a low regrowth. A late heat wave not 266 accompanied by a pronounced drought – as in 2015 - induced a delayed senescence, especially at 267 high elevation, and a substantial regrowth.

268

269 **Discussion**

Our study of mountain grassland phenology in the French Alps for the period from 2000 to 2016 has provided the framework to analyse more thoroughly how these ecosystems respond to heat waves. Using data from the past two decades, we could show that (i) warmer summers generally have a positive effect on mountain canopy greenness especially at high elevation, (ii) water deficit hinders the positive temperature effect of heat waves on canopy greenness, and (iii) faster senescence and weaker late-season regrowth are the main negative effects of heat waves that co-occur with water deficit. Our findings emphasize the need to consider both the growth and senescent phases to

capture the effect of climate extremes on phenology. We also highlighted the pivotal role of the
magnitude of summer water balance and its within-summer distribution for understanding whether
mountain grasslands can benefit or not from higher summer temperatures.

280 We found a general increase in canopy greenness (NDVImax) in the French Alps during the study 281 period (2000-2016) which is consistent with our previous remote sensing study that showed a 282 greening trend for high-elevation ecosystems in the Ecrins Massif (Carlson et al., 2017). This 283 greening has also been documented in other temperate mountains (Xu et al., 2011) as well as in the 284 arctic and sub-arctic (Jia et al., 2003; Forbes et al., 2010). The combination of earlier snow melt-out 285 date and higher summer temperatures has been put forward as the main driver of changes in primary 286 productivity (Jonas et al., 2008; Zhang et al., 2013a; Choler, 2015; Dedieu et al., 2016; Choler, 287 2018). Besides temperature, other factors such as the increased deposition of atmospheric nitrogen 288 possibly led to some eutrophication of upland ecosystems and further contributed to greening 289 (Bowman et al., 2018). In alpine plants, nitrogen addition may induce stronger phenological changes 290 than warming alone (Smith et al., 2012). Our study indicated that sparsely vegetated meadows in the 291 French Alps are the most prone to greening, and we hypothesized that the prolonged growing season 292 in high-elevation contexts has created new thermally suitable niches for early plant colonizers which 293 could lead to an increase in primary productivity (Carlson et al., 2017). These conclusions are in 294 agreement with ground-based observations of community dynamics that emphasise the densification 295 of plant cover in open grasslands and/or the upward/northward shift of more productive functional 296 groups (Kullman, 2010; Myers-Smith et al., 2011; Steinbauer et al., 2018).

Our results also showed that the frequency of summer heat waves has increased over the last decades in the French Alps. This is consistent with reports highlighting the increased probability of heat waves in a warming climate (Stocker et al., 2013; Russo et al., 2014). We pointed out that the four main heat waves that hit the Alps over the last 20 years did not have the same effect on mountain grassland phenology. Remarkably, the 2015 heat wave had opposite late-season effects to

302 those of 2003, 2012, and 2016. This discrepancy is due to the combined effects of temperature and 303 climatic water balance, both in magnitude and timing. Water deficit coincided with heat wave in 304 2003 and 2012, and to a lesser extent in 2016. We hypothesize that a threshold in the water balance 305 was reached for these three years leading to the diminished canopy greenness during the late-season. 306 Other reports have suggested that the disproportionate effect of the 2003 summer heat wave may not 307 only be due to the unusually warm temperatures, but also to a negative soil water anomaly at the 308 beginning of the growth period (Loew et al., 2009). Compared to the lowland ecosystems, mountain 309 grasslands did not exhibit a strong reduction of NDVImax in 2003 and 2012, probably because the 310 snowmelt provided sufficient water for the initial growth period, at least in the Northern Alps. 311 Subsequently, the soil water content has likely decreased during the summer due to a climatic water 312 deficit causing a significant reduction of canopy greenness during the late-season. For the year 2015, 313 we hypothesize that the above normal water balance in the early season and the limited water deficit 314 in the late-season have favoured the positive effect of the late-season heat wave, especially for the 315 highest, i.e. the most temperature-limited, mountain grasslands. In their study of the 2003 heat wave, 316 Jolly et al. (2005) reported marked regional differences in the response of high elevation ecosystems. 317 In particular, it was striking that mountain ecosystems of the Southern Alps showed a reduced 318 canopy greenness compared to the Northern French or Swiss Alps, which was likely related to the 319 interplay of heat wave and water availability. Similarly, Cremonese et al. (2017) pointed out the co-320 limitation of high temperature and soil water content to impede the greening of a *Nardus*-rich 321 grassland in 2015. Recent experimental studies conducted on temperate mountain grasslands showed 322 that the direct effects of air temperature on photosynthesis and aboveground phytomass production 323 strongly depend on soil water content (De Boeck et al., 2016). These findings should prompt us to 324 improve our understanding and regional modelling of the soil water balance of mountain grasslands. 325 Unfortunately, this remains notoriously difficult due to the many unknowns (e.g., the soil water 326 capacity and lateral run-offs). However, with further temperature increases looming in the not so

distant future, there is an urgent need to formulate eco-hydrological models of seasonal leaf
dynamics and to identify the different thresholds that drive ecosystem response. This is particularly
true for grasslands of the subalpine belt that are the most sensitive to drought and are a key resource
for summer livestock farming (Nettier et al., 2017)

331 Numerous studies on arctic and alpine ecosystems have underlined that the response to 332 warming varies across functional groups (Arft et al., 1999; Jonas et al., 2008; Jung et al., 2014; 333 Livensperger et al., 2016). In relation to the recent observed greening it is plausible that the 334 expansion of more productive habitats has changed the responsiveness of mountain grasslands to 335 climate extremes. The coexistence of a richer set of plant life forms may arise, e.g., through the 336 migration of low shrubs onto the grasslands, or of tall perennial grasses into otherwise low-stature 337 habitats. There are several lines of evidence that an increased functional diversity may provide some 338 ecosystem insurance by buffering the impact of environmental changes (Diaz and Cabido, 2001). 339 However, further studies are needed to examine a possible relationship between plant functional 340 diversity and the phenological responses of mountain grasslands to heat waves and drought. For 341 example, the combined use of high-resolution satellite imagery in combination with a regional-scale 342 database of vegetation could provide new insights into this line of enquiry.

343 Our comparative analysis of mountain grassland phenology over the last two decades 344 provides a knowledge base to help anticipate the long-term response of grasslands under future 345 climate conditions. Regional climate models used to compute climate projections are operating at 346 coarse resolutions (10-50 km) (Jacob et al., 2014) and are insufficient to provide ecologically-347 relevant surface parameters in mountainous regions (Ford et al., 2013). Although downscaling 348 methods can refine projections of meteorological variables and snow cover dynamics (Verfaillie et 349 al., 2018), further work is needed to translate these model outputs into bioclimatic indicators and 350 assess to what extent they are driving canopy greenness. In the longer term, efforts have to be 351 conducted towards prognostic models of mountain grassland phenology that capture both

temperature and soil water constraints. The dynamics of leaf growth and senescence remains poorly represented in land surface models (Arora and Boer, 2005; Richardson et al., 2013), and this holds particularly true for seasonally snow-covered mountain grasslands. Our study has focused on interannual fluctuations to provide a broad picture over the French Alps. Further work examining between-site phenological variation along elevational and bioclimatic gradients will provide a complementary view and will be helpful to improve our representation of the feedback between phenology and climate within process-oriented models.

360 **Conclusions**

361 Our study highlighted the combined effect of summer temperature and water balance on the 362 phenology of temperate mountain grasslands. In the European Alps, the increased or prolonged 363 canopy greenness of mountain grasslands that result from warmer conditions becomes less 364 pronounced when water is limiting. While there has been a significant decadal-scale increase in both 365 the summer temperature and the frequency of heat waves, summer precipitation and reference 366 evapotranspiration did not show a similar trend and exhibited large inter-annual variability. In 367 addition, there is a growing body of evidence to support long-term changes of mountain vegetation in 368 relation to climate warming and land use change. The extent to which these changes modify the 369 sensitivity of mountain grassland to extreme events will also be pivotal to make possible the 370 prediction of canopy greenness under future climate.

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- 381 Tables
- 382

Table 1. Annual means and inter-annual coefficients of variation (CV in %) of the meteorological

384 and phenological variables used in the study.

			Mean (CV)			
			1350 m -	1650 m -	1950 m -	2250 m -
Metric	Period	Abbreviation	1650 m	1950 m	2250 m	2550 m
Meteorology						
	Summer	HWI (JJA)	55 (79)	65 (71)	58 (79)	45 (89)
Heat Wave magnitude Index	Preceding the peak	HWI (B30)	18 (76)	21 (77)	20 (107)	14 (127)
	Following the peak	HWI (A60)	35 (111)	43 (98)	34 (116)	22 (130)
Precipitation –	Summer	P-ET ₀ (JJA)	61 (79)	58 (110)	51 (24)	16 (22)
Reference Crop	Preceding the peak	P-ET ₀ (B30)	35 (150)	24 (135)	14 (98)	5 (-285)
Evapotranspiration (mm)	Following the peak	P-ET ₀ (A60)	32 (25)	36 (-41)	39 (11)	25 (103)
Phenology						
Onset of growth (DOY)		ONSET	112 (8)	130 (8)	145 (8)	157 (7)
Date of maximum NDVI (DOY)		TNDVImax	175 (7)	184 (5)	192 (5)	200 (6)
Offset of growth (DOY)		OFFSET	327 (5)	315 (5)	306 (4)	300 (3)
Maximum NDVI		NDVImax	0.79 (2)	0.78 (2)	0.75 (3)	0.69 (4)
Senescence	Following the peak	SEN (A60)	0.13 (22)	0.14 (18)	0.16 (18)	0.19 (21)
Regrowth	Following the peak	REG (A60)	69 (18)	62 (19)	54 (24)	48 (31)

386	Table 2. Summary statistics for Standardized Major Axis (SMA) regression analysis of phenological
387	variables. Shown are correlation coefficients (r), estimates of slopes (a) and intercepts (b) with 95%
388	confidence intervals, Mean Square Error (MSE) and Mean Absolute Error (MAE). For the intercept,
389	(NS) indicates a non-significant offset. For the slope, (NS) indicates an estimate statistically
390	indistinguishable from one.

			SMA	H ₀ : b=0	SMA	H ₀ : a=1		
Relation	n	r	intercept (b)	P value	slope (a)	P value	MSE	MAE
ONSET - TNDVImax	1479	0.69	-0.11	***	0.73	***	0.39	0.50
			(-0.13, -0.08)		(0.7, 0.76)			
TNDVImax - NDVImax	1479	0.31	0.04	*	0.97	NS	0.51	0.55
			(0.01, 0.08)		(0.93, 1.02)			
SEN (A60) - REG (A60)	1478	-0.45	-0.1	***	-1.82	***	1.70	1.06
			(-0.16, -0.05)		(-1.74, -1.91)			

- **Table 3.** Regression coefficients and adjusted R^2 for the linear models testing the effect of P-ET₀ and
- 393 HWI on NDVImax, SEN (A60) and REG (A60). (x) denotes the interaction between the two effects.
- Plain text is for highly significant effect (P < 0.001), italics is for significant effect (P < 0.01). Empty
- 395 cells indicate non-significant effects.

	NDVImax					
	1350 m - 1650 m - 1950 m - 2250 m					
Effect	1650 m	1950 m	2250 m	2550 m		
Intercept	-0.002	0.002	-0.043	-0.051		
P-ET ₀ (B30)	0.27	0.33	0.25	0.14		
HWI (B30)						
P-ET ₀ (B30) x HWI (B30)	0.16	0.15				
Adjusted R^2	0.14	0.16	0.11	0.04		

		SEN (A60)			REG (A60)			
	1350 m -	1650 m -	1950 m -	2250 m -	1350 m -	1650 m -	1950 m -	2250 m -
Effect	1650 m	1950 m	2250 m	2550 m	1650 m	1950 m	2250 m	2550 m
Intercept	-0.191	-0.147	-0.088	-0.042	0.108	0.016	-0.067	0.082
P-ET ₀ (A60)	-0.37	-0.27	-0.20		0.57	0.59	0.51	0.26
HWI (A60)				-0.18				
P-ET ₀ (A60) x HWI (A60)	-0.14	-0.08				-0.06	-0.21	
Adjusted R^2	0.47	0.33	0.08	0.05	0.24	0.24	0.20	0.05

		Elevation	n band		
	1350 m -	1650 m -	1950 m -	2250 m -	
Massif	1650 m	1950 m	2250 m	2550 m	Total
1. Chablais	553	787	119	0	1459
2. Aravis	878	679	94	0	1651
3. Mont-Blanc	174	728	714	49	1665
4. Bauges	268	233	8	0	509
5. Beaufortin	373	1425	1562	234	3594
6. Haute-Tarentaise	68	277	580	404	1329
7. Chartreuse	71	132	1	0	204
8. Belledonne	97	788	365	4	1254
9. Maurienne	296	1054	1179	430	2959
10. Vanoise	162	457	1572	899	3090
11. Haute-Maurienne	64	215	979	1181	2439
12. Grandes Rousses	415	1258	1529	622	3824
13. Thabor	75	274	717	636	1702
14. Vercors	561	324	4	0	889
15. Oisans	499	967	818	55	2339
16. Pelvoux	194	251	564	583	1592
17. Queyras	63	254	679	1489	2485
18. Devoluy	644	460	28	1	1133
19. Champsaur	601	651	918	419	2589
20. Embrunnais Parpaillon	113	165	523	454	1255
21. Ubaye	50	185	416	355	1006
22. Haut-Var Haut-Verdon	356	707	794	185	2042
23. Mercantour	342	946	1481	517	3286
Total	6917	13217	15644	8517	44295

Supplementary Table 1. Number of MODIS pixels by massif and elevation band.

399 Figure legends

400

401 Figure 1. (A) Study area showing the 23 SAFRAN massifs in the French Alps and the distribution of 402 the 250 m-resolution MODIS pixels (n = 44,295) dominated by mountain grasslands. (B) Pixel distribution by massif and elevation band. Names of massifs are: 1. Chablais; 2. Aravis; 3. Mont-403 404 Blanc; 4. Bauges; 5. Beaufortin; 6. Haute-Tarentaise; 7. Chartreuse; 8. Belledonne; 9. Maurienne; 10. 405 Vanoise; 11. Haute-Maurienne; 12. Grandes Rousses; 13. Thabor; 14. Vercors; 15. Oisans; 16. 406 Pelvoux; 17. Queyras; 18. Devoluy; 19. Champsaur; 20. Embrunnais Parpaillon; 21. Ubaye; 22. 407 Haut-Var Haut-Verdon; 23. Mercantour. Account of pixels by massif and elevation band is given in 408 Fig. S2.

409

410 Figure 2. An idealized yearly course of vegetation index showing the different variables and periods411 that were used to capture the phenology of the mountain grasslands.

412

Figure 3. Annual standardized anomalies of (A) the Heat Wave magnitude Index (HWI (JJA)) and (B) the difference between precipitation and reference crop evapotranspiration (P-ET₀ (JJA)) during summer in mountain grasslands located between 1950 m and 2250 m. Meteorological variables were estimated using the SAFRAN reanalysis (see Materials and Methods for details). The study period of mountain grassland phenology started in year 2000 (dashed line). Boxplots represent the median (central bar), and the first and third quartiles, with whiskers showing the 95% confidence intervals. Grey boxplots represent the four main summer heat waves (2003, 2012, 2015, and 2016).

420

Figure 4. Bivariate plots of standardized anomalies of HWI and $P-ET_0$ during (A) the 30 days before (B30), and (B) the 60 days after (A60) the date of maximum NDVI (TNDVImax). Large dots represent the four main summer heat waves (2003, 2012, 2015, and 2016).

425	Figure 5. Annual standardized anomalies of mountain grassland phenology: (A) onset of growth, (B)
426	date of maximum NDVI (TNDVImax), (C) offset of growth, (D) maximum canopy greenness
427	(NDVImax), (E) senescence during the 60 days following NDVImax (SEN (A60)), and (F) regrowth
428	during the 60 days following NDVImax (REG (A60)). Boxplots represent the median (central bar),
429	and the first and third quartiles, with whiskers showing the 95% confidence intervals. Grey boxplots
430	represent the four main summer heat waves (2003, 2012, 2015, and 2016).
431	
432	Figure 6. Bivariate relationships between annual standardized anomalies of TNDVImax, NDVImax,
433	SEN (A60) and REG (A60). Standard Major Axis (SMA) regression lines are shown. Large dots
434	represent the four main summer heat waves (2003, 2012, 2015, and 2016).
435	
436	Figure 7. Effect of HWI and P-ET ₀ on: NDVImax (A, B), SEN (A60) (C, D) and REG (A60) (E, F).
437	Dependent and independent variables used in the linear models are standardized anomalies.
438	Anomalies of HWI have been log-transformed. Regression lines are shown for each elevation band
439	(see Materials and Methods for details).
440	
441	Supplementary Figure 1. Sensitivity of Heat Wave magnitude Index (HWI) to the length of the
442	window (parameter n) used to estimate the threshold temperature value (Eq. 1 in Materials and
443	Methods). Shown are the relative changes in HWI estimates when n varies from 15 days to 13 days
444	(A), or from 15 days to 11 days (B).
445	
446	Supplementary Figure 2. Annual time series of summer HWI and P-ET ₀ for mountain grasslands
447	located between 1950m and 2250m. Boxplots represent the median (central bar), and the first and

third quartiles, with whiskers showing the 95% confidence intervals. Grey boxplots represent the
four main summer heat waves (2003, 2012, 2015, and 2016).

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Figure 6





Figure S1

