

Passive larval transport explains recent gene flow in a Mediterranean gorgonian

Mariana Padrón, Federica Costantini, Sandra Baksay, Lorenzo Bramanti, Katell Guizien

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Larval transport explains contemporary gene flow in a Mediterranean gorgonian Mariana Padrón¹, Federica Costantini², Sandra Baksay³, Lorenzo Bramanti¹, Katell Guizien^{1*} ¹ Sorbonne Universités, UPMC Univ Paris 06, CNRS, Laboratoire d'Ecogéochimie des Environnements Benthiques (LECOB), Observatoire Océanologique, 66650, Banyuls sur Mer, France. ² Dipartimento di Scienze Biologiche, Geologiche ed Ambientali (BiGeA) & Centro Interdipartimentale di Ricerca per le Scienze Ambientali (CIRSA), Università di Bologna, ULR CoNISMa, Via S. Alberto 163, I-48123 Ravenna, Italy. ³ Laboratoire Evolution et Diversité Biologique EDB, CNRS, Université Toulouse III Paul Sabatier, F-31062 Toulouse, France. *Corresponding author: tel: +33 468887319; e-mail: guizien@obs-banyuls.fr; ORCID: 0000-0001-9884-7506 Keywords: marine connectivity, Pelagic Larval Duration, genetic structure, Marine Protected Areas, hydrodynamic model, Gulf of Lion.

Abstract

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Understanding the patterns of connectivity is required by the Strategic Plan for Biodiversity 2011-2020 and will be used to guide the extension of marine protection measures. Despite the increasing accuracy of ocean circulation modeling, the capacity to forecast the population connectivity of sessile benthic species with dispersal larval stages can be limited due to the potential effect of demographic filters acting before or after dispersal, which modulate offspring release or settlement, respectively. We applied an interdisciplinary approach that combined demographic surveys, genetic methods (assignment tests and coalescent-based analyses) and larval transport simulations to test the relative importance of demographics and ocean currents in shaping the contemporary patterns of gene flow among populations of a Mediterranean gorgonian (Eunicella singularis) in a fragmented rocky habitat (Gulf of Lion, NW Mediterranean Sea). We show that larval transport is a dominant driver of gene flow among the populations, and significant correlations were found between the contemporary gene flow and recent larval transport when the pelagic larval durations (PLDs) ranged from 7 to 14 days. Our results suggest that PLDs that efficiently connect populations distributed over a fragmented habitat are filtered by the habitat layout within the species competency period. Moreover, a PLD ranging from 7 to 14 days is sufficient to connect the fragmented rocky substrate of the Gulf of Lion. The rocky areas located in the center of the Gulf of Lion, which are currently not protected, were identified as essential hubs for the distribution of migrants in the region. We encourage the use of a range of PLDs instead of a single value when estimating larval transport with biophysical models to identify potential connectivity patterns among a network of marine protected areas or even solely a seascape.

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Introduction

The Strategic Plan for Biodiversity 2011-2020 aims to conserve 10% of the ocean through well-connected systems of marine protected areas (MPAs) by 2020 (Aichi Target 11,

https://www.cbd.int/sp/targets/rationale/target-11/). Hence, understanding the degree and patterns of connectivity among existing MPAs and areas where protection could be extended becomes urgent. Connectivity is also known as the mechanism through which populations of sessile marine species, and particularly coral species, can recover after local and global disturbances (Roberts 1997; Hanski 1998; Cowen et al. 2000). In the context of global change, relating contemporary connectivity to the seascape processes that shape the spatial distribution of marine species is essential for the design of effective biodiversity conservation strategies.

Seascape processes that drive population connectivity of sessile species with a larval dispersal stage, such as corals or gorgonians, should include both seabed geomorphology and dynamic hydrography (Manderson 2016). Correlations at the local scale between demographic population descriptors and the environment may, indeed, not be sufficient, and the successful exchange of individuals between distant locations (i.e., connectivity) should be included. Recent studies on seascape connectivity have focused on evaluating how larval transport drives genetic patterns across the seascape using a combination of biophysical and genetic approaches (see Liggins et al. 2013 for a review). Nonetheless, Selkoe et al. (2016) reported that only 31% of the analyzed studies found ocean currents to be a good predictor of gene flow patterns and ascribed this poor predictability to inappropriate scales in genetic sampling. These findings suggest that demographic and genetic connectivity of benthic sessile species, for which connectivity among populations only occurs during the larval dispersive stage, are affected by several processes acting as successive biological filters at different temporal and spatial scales along their life cycles from spawning to reproduction by the settled individuals (Pineda 2007).

Prior to dispersal, population size and fecundity are among the key factors that determine the number of emigrant offspring and have been shown to influence connectivity patterns in marine species (Treml et al. 2012; Dawson et al. 2014). After dispersal, gene flow among areas that are well connected through larval transport can be limited by settlement regulation due to space limitations (Roughgarden et al. 1985). However, the primary filter of connectivity is larval

transport, which is notably driven by the amount of time that larvae spend being transported by ocean currents, otherwise known as the pelagic larval duration (PLD). Hence, the PLD has been the primary factor considered in biophysical modeling when attempting to differentiate dispersal distance capabilities among different species (Shanks et al. 2003).

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Genetic studies have tested the relationship between realized connectivity (gene flow) and the PLD following the same reasoning, and the results have led to contrasting conclusions (Kinlan and Gaines 2003; Weersing and Toonen 2009; Riginos et al. 2011, D'Aloia et al. 2015). Several arguments have been invoked to explain the absence of a relationship between PLD and realized dispersal distance and migration rates, with some arguments referring to the way in which realized connectivity is inferred from measures of genetic differentiation among populations. For instance, the inverse relationship between the most common metric of genetic differentiation, F_{ST}, and the migration rate (Wright's Island Model 1931), may be blurred by the uncertainty of the estimation of the effective size of the population (Faurby and Barber 2012). However, the failure to identify a relationship between PLD and realized connectivity is more likely explained by the lack of unequivocal correlation between PLD and dispersal distance. In fact, the variability of ocean currents (Cowen et al. 2006, Guizien et al. 2014), larval behavior (Guizien et al. 2006, Paris et al. 2007) and larval mortality (Cowen et al. 2000) can lead to very different dispersal patterns given the same PLD. Moreover, for any given species, PLD does not have a single value, but it varies along a competency window (Arnold and Steneck 2011), within which successful recruitment can occur at any time once a suitable habitat is found along the dispersal pathway.

To determine which biological filters transform larval release into gene flow (potential into realized connectivity), we used one of the most abundant and conspicuous species of the rocky sublittoral zone in the northwestern Mediterranean Sea, the white gorgonian *Eunicella singularis*.

In the hydrodynamic province of the Gulf of Lion (NW Mediterranean, Rossi et al. 2014),

the spatial distribution of *E. singularis* is discontinuous (Gori et al. 2011) due to fragmentation of the rocky habitat (Aloisi et al. 1973). *E. singularis* is believed to play an important ecological role due to its erect structure, which provides habitat for the local epifauna and increases the biomass and diversity of the rocky community (Ballesteros 2006; Mitchell et al. 1992). Furthermore, its indirect economic value due to its contributions to shaping the seascape and attracting SCUBA diving and fishing activities makes *E. singularis* a species of interest in the monitoring plans of all MPAs in the Gulf of Lion. *E. singularis* is a long-lived (up to 15 years) internal brooding sessile species. Lecitotrophic larvae (planulae) are released once a year in June-July (Théodor 1967; Weinberg and Weinberg 1979). Larvae can survive up to 122 days in *ex situ* experiments in the absence of predation (Théodor 1967), with 50% to 80% survival after 40 days (K. Guizien, *Personal communication*). Competency for recruitment was observed after a few days and lasted up to a few weeks (L. Bramanti, *Personal communication*). Such larval traits suggest a wide dispersive potential of the species.

In the present study, we propose a methodological approach that combines demographic, biophysical and genetic methods to test the relationship between the dispersive potential of a sessile benthic species and the effective gene flow in a fragmented habitat layout. The study aims to (1) assess the genetic structure of *Eunicella singularis* within the fragmented habitat of the Gulf of Lion, (2) test the role of demographic filters (i.e., reproductive output and intra-specific recruitment limitations) in shaping gene flow, and (3) test the significance of various PLD-driven larval transport modes in shaping the spatial pattern of gene flow.

MATERIAL AND METHODS

Study area and regional hydrodynamics

The Gulf of Lion (hereafter GoL) is a large (~100 km), microtidal continental shelf with a 100-km radius located in the northwestern Mediterranean Sea (Figure 1). The first 30 meters of the bathymetry is mainly covered by soft sandy sediment; however, fragmented patches of hard

substrate can be found in the eastern (Côte Bleue), central (Plateau des Aresquiers and Cap d'Agde), and western (Cap Leucate and Côte Vermeille) parts of the gulf (Aloisi et al. 1973). Circulation in the GoL is driven by three main forces: 1) strong continental winds from the north (Mistral) and the northwest (Tramontane) that frequently generate upwelling and downwelling along the coast, as well as transient currents and eddies; 2) the Northern Current (NC), which usually flows westward along the shelf break of the GoL from the Ligurian Sea to the Catalan Sea; 3) river inputs, particularly from the Rhône, that deliver freshwater and high nutrient and organic matter loads (Millot 1990; Hu et al. 2009; Campbell et al. 2013).

Numerical simulations at the basin scale have identified that the GoL is an isolated hydrodynamic province, with the Northern Current acting as a hydrodynamic barrier that limits the transport of particles across the shelf break (Rossi et al. 2014). Large-scale flow limits entries into the Gulf of Lion up to a 60-day PLD from the east and up to a 30-day PLD from the west. Moreover, despite an overall westward drift leading to large sediment exports through western canyons (Ulses et al. 2008), persistent anticyclonic eddies are observed in the southwestern part of the GoL during the summer (Hu et al. 2009), and connectivity patterns within the Gulf of Lion stabilize after 3 weeks.

Sampling design

From July 2013 to October 2014, detectable colonies (taller than 1 cm) of *Eunicella singularis* were counted and sampled at 17 geo-referenced stations distributed along the GoL, from the Côte Bleue to the Côte Vermeille (Table 1) (Figure 1). The stations were distributed along four rocky habitat patches separated by large areas of sandy bottom (Côte Bleue, Plateau des Aresquiers, Cap d'Agde and Côte Vermeille) at depths between 15 and 35 meters. The distance between stations varied from less than 1 km to 180 km. The population size structures were similar in the four rocky habitat patches, with 25% of the colonies smaller than 5 cm, 55% of the colonies between 5 and 20 cm and 20% of the colonies taller than 20 cm (L. Bramanti, *Personal communication*). The number of stations sampled in each habitat patch varied according

to the along-shore extension of *E. singularis* coverage of the rocky bottom (L. Bramanti, *Personal communication*). In Côte Bleue, where the smallest coverage of *E. singularis* was found, only three stations were sampled; in Plateau des Aresquiers and Cap d'Agde, four stations were sampled, while six stations were sampled in the Côte Vermeille.

The mean population density at each station was calculated from colony counts in 4 quadrats (1 m² each). For each station, the 4 quadrats were located 5 m from the geo-referenced station, one in each of the four cardinal directions. Small apical fragments (<3 cm) of all colonies of *E. singularis* that were counted at each station were collected, preserved in 95% ethanol and kept at 6 °C until DNA extraction. For the present study, 13 to 36 colonies per station were randomly selected and genotyped, for a total of 434 colonies.

Genetic diversity and population structure

All colonies were genotyped at eight microsatellite loci specifically developed for *E. singularis*: C21, C30 and C40 (Molecular Ecology Resources Primer Development Consortium 2010) and *E. verrucosa*: Ever003, Ever004, Ever009, Ever013 and Ever014 (Holland et al. 2013) after DNA extraction (Appendix S1). The scoring of the alleles was carried out using GeneMapper v.3.7 software (Applied Biosystems). We calculated the summary statistics of the genetic diversity (allelic richness, observed heterozygosity, expected heterozygosity), and tested for departure from the Hardy-Weinberg equilibrium, linkage disequilibrium and the presence of null alleles (Appendix S1). The statistical power to detect the level of genetic variation among populations was estimated using Powsim v.4.1 (Ryman and Palm 2006). The spatial genetic structure of *E. singularis* was analyzed by testing for the presence of a pattern using the isolation by distance (IBD) and the Loiselle kinship coefficient (Appendix S1). The patterns of population genetic structure were analyzed using F_{ST}, a Bayesian clustering method that identifies clusters of individuals who share similar patterns of variation (Structure v.2.3.4; Pritchard et al. 2000) and a hierarchical analysis of molecular variance (AMOVA). For details, see the Methods in Appendix S1.

Contemporary gene flow

The genotyped individuals from the 17 stations were clustered into 4 groups defined by habitat discontinuity, corresponding to the four largest rocky habitat patches (Côte Vermeille, Cap d'Agde, Plateau des Aresquiers and Côte Bleue). Contemporary gene flow over the last 2-3 generations was estimated in the form of matrices containing migration rates including a proxy of self-recruitment among these 4 clusters using BayesAss 3.0 applied on multilocus genotypes (Wilson and Rannala 2013). It is important to note that the BayesAss 3.0 output is the proportion of individuals in the destination population that is descendant from a source population.

The confidence intervals and the uncertainty values around the average migration rates were estimated (Appendix S1).

Recent larval transport

Using the 3D ocean model Symphonie (S-2010 release 26, at http://sirocco2.omp.obs-mip.fr/outils/Symphonie/Accueil/SymphoAccueil.htm), high-resolution flow simulations were carried out and used for larval dispersal simulations (Appendix S1). The latter were upscaled to the larval transport matrices for single reproductive events. Larval transport matrices contain transfer rate values defined as the proportion of larvae (neutrally buoyant particles) released from a source habitat patch (row) during a given release period that reach a destination habitat patch (column) after a given PLD. Fifteen variations of larval transport matrices were established per PLD, with each variant corresponding to a non-overlapping weeklong release period in June (five variants per year) over three years (2010, 2011, 2012, Appendix S1). Six weeks of dispersal were simulated and eight PLD values ranging from 3.5 to 42 days were tested. The PLD range that was tested was greater than the period during which *Eunicella singularis* larvae competency for recruitment as observed in laboratory experiments (L. Bramanti, *Personal Communication*). Moreover, the maximum PLD tested in the present study (42 days) was longer than the time required to observe stable connectivity patterns within the Gulf of Lion (Briton et al., submitted).

Comparison between larval transport and gene flow

With gene flow being estimated as the proportion of individuals in the destination population that is a descendant from a source population, and larval transport being the proportion of the larvae released from a source population that arrives in a destination population, comparing gene flow and larval transport requires a series of data transformations (see details in Appendix S1). Briefly, as the actual number of larvae released from a source population is unknown, the transformation of larval transport into the number of larvae reaching a destination population is not possible. Hence, gene flow was transformed into the proportion of migrants that leaves a source population and arrives at each destination population. For each PLD, recent larval transport was estimated as the average of the 15 larval transport matrix variants (one for each single reproductive event). Recent larval transport matrices were normalized so that the sum of the destination probabilities from the same source (i.e., all columns within each row) was equal to one, which removed the proportion of larvae dispersed out of the 4 habitat patches sampled for genetics as this proportion is not included in the gene flow assessment. It is worth noting that after these normalizations, for both the gene flow and larval transport matrices, only rows, not columns, are statistically independent. Hence, the significance of the Pearson product-moment correlations (excluding diagonal terms, i.e., retention rates) between the contemporary gene flow and recent larval transport matrices (4x4) was tested using a Mantel test (Mantel 1967) for all 24 possible permutations, after modifying the random QAP.m Matlab routine by Puck Rombach (2011). In addition, the determination coefficient between contemporary gene flow and recent larval transport was computed including diagonal terms, i.e., retention rates.

Relationship between population density and migration

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Finally, the effect of demographic filters on the number of migrants (see calculation in Appendix S1) was tested, namely, the modulation of emigrants by reproductive output from the source populations and the modulation of immigrants by settlement limitation in the destination populations. Two correlation analyses were performed to test the relationship between 1) the number of contemporary emigrants from a source habitat patch (sum of migrants number by row)

and the mean population density in that source (proxy of reproductive output given the same population size structure in all habitat patches), and 2) the number of contemporary immigrants to a destination habitat patch (sum of migrant numbers by column) and the mean population density in that destination patch (proxy of intra-specific competition). All statistical analyses were conducted in Matlab (R2012a).

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RESULTS

Genetic diversity and population structure

All sampled individuals were included in the analysis (no identical multilocus genotypes). All loci were polymorphic, with more than 4 alleles among stations on average and from 5 to 13 alleles over all loci, which is comparable to the degree of polymorphism obtained by Holland et al. (2013). Linkage disequilibrium was not observed (Appendix S2, Table A2). The observed and expected heterozygosities were similar among stations with no significant heterozygote excesses, but there were significant heterozygote deficits across two populations (CV1 and CA1) (Appendix S2, Table A2) and in two of the eight loci (Ever003 and C40) (Appendix S2, Table A3). The frequency of null alleles was negligible (Appendix S2, Table A3). The overall F_{ST} value was low (0.04), and almost all pairwise F_{ST} values among stations were significantly different (P-value < 0.001) (Appendix S2, Table A4). Nei's genetic distance values were closer among the stations within habitat patches than between them (Figure 2). While the stations in Côte Bleue and Cap d'Agde-Plateau des Aresquiers are close to each other, the stations in the Côte Vermeille (CV) appear more scattered along the first axis, with one station (CV5) falling near the stations of Cap d'Agde and Plateau des Aresquiers (CA and PA, respectively). The Bayesian clustering analysis identified three genetic clusters (K = 3) (mean LnP(K) =

-7532.99; Delta K = 25.09). The resulting clustering partially corresponded to the rocky habitat

fragmentation in the region (Figure 3): Côte Vermeille (6 stations), Cap d'Agde and Plateau des

Aresquiers (8 stations), and Côte Bleue (3 stations). Some individuals in Côte Vermeille depicted higher similarity to individuals from Cap d'Agde and Plateau des Aresquiers rather than to other individuals in Côte Vermeille. The results of the AMOVA confirmed the clustering structure (P < 0.001) (Appendix S2, Table A5).

The correlation between F_{ST} / (1- F_{ST}) and the logarithm of the geographic distance (R^2 = 0.28; P-value = 0.01) was significant but low among the 17 stations in the GoL and was not significant among the 4 patches defined according to rocky habitat fragmentation (R^2 = 0.81; P-value = 0.100) (Appendix S2, Figure A2). The estimated maximum size of the genetic patch based on Loiselle's kinship coefficient was 30 km, which is larger than the maximum size of any of the habitat patches in the Gulf of Lion (Appendix S2, Figure A3).

Contemporary gene flow

Figure 4 shows the spatial patterns of the number of contemporary migrants among the 4 separated habitat patches in the GoL. The uncertainty of the migration pattern, following an error metric adapted from the Euclidean matrix norm (see details in Appendix S1), was below 25%. Local retention dominated over emigration in all habitat patches. However, local retention was much lower in Côte Bleue (mean = 51 migrants, CI (95%) = 48 - 52 migrants) compared to all other habitat patches (from 90 to 121 migrants) and was comparable to the value of emigration from Cap d'Agde (45 ± 14 migrants). Cap d'Agde acted as a hub, distributing migrants to both Côte Vermeille (westward) (mean = 16 migrants, CI (95%) = 6-19 migrants) and Plateau des Aresquiers (eastward) (mean = 28 migrants, CI (95%) = 17-34 migrants).

Relationship between population density and migration

Neither the number of contemporary emigrants coming from a habitat patch (Figure 5A) nor the number of contemporary immigrants arriving to a habitat patch (Figure 5B) showed a significant correlation with the population density of the source ($R^2 = 0.203$, P > 0.05) or the destination ($R^2 = 0.787$, P > 0.05) habitat patch, respectively.

Although the lowest number of emigrants corresponded to the source habitat patch with the lowest population density, the opposite was not true. The highest number of emigrants did not come from the habitat patch with the highest population density. The number of immigrants did not decay with increasing population density in the destination habitat patch but conversely increased. However, a similar number of immigrants were found in the destination habitat patches with quite different population densities.

Comparison between recent larval transport and contemporary gene flow

In the following section, both larval transport and gene flow have been normalized for each source population and are subsequently no longer comparable between different source populations. Recent larval transport varied greatly among the PLDs ranging from 3.5 to 42 days: from a scenario of high local retention within all habitat patches to a scenario of high levels of larval transport among most habitat patches, except Côte Vermeille (Figure 6A to 6E). Larval transport was restricted to the exchange between Cap d'Agde and Plateau des Aresquiers for the shortest PLD tested (3.5 days, Figure 6A), and increased in a westward direction for the PLDs longer than 14 days (Figures 6D, 6E and data not shown). For the PLDs of 7 and 10.5 days (Figure 6B and 6C), larval transport was westward from Plateau des Aresquiers and Côte Bleue, but both westward and eastward from Cap d'Agde. Moreover, it is interesting to note that the local retention decayed more rapidly in Cap d'Agde compared to the other habitat patches due to efficient larval export towards the closest habitat patches of Plateau des Aresquiers and Côte Vermeille. This pattern matched the one observed for contemporary gene flow particularly well (Figure 6F).

The Mantel test showed a significant correlation (Mantel statistics <5%) between contemporary gene flow and recent larval transport (excluding retention rates) only for the PLDs ranging from 7 to 14 days with a determination coefficient (including retention rates) larger than 50% (Figure 7). However, the determination coefficient between contemporary gene flow and recent larval transport, which includes retention rates, was also high for a PLD of 3.5 days (R²)

>0.95). Interestingly, the determination coefficient did not decrease monotonically with PLD but rather peaked at a PLD of 7 days ($R^2 = 0.98$).

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DISCUSSION

Habitat fragmentation fails to explain regional genetic structure of Eunicella singularis in the 317 Gulf of Lion

The analysis of the genetic differentiation of groups of Eunicella singularis colonies among a wide range of spatial scales (from less than 1 km to 180 km) showed high levels of genetic variability at small scales with a deficit of heterozygotes at some stations, as reported previously for the same species (Costantini et al. 2016) and in other Mediterranean gorgonians such as Paramuricea clavata (Mokhtar-Jamaï et al. 2011) and Corallium rubrum (Costantini et al. 2007). In these studies, genetic patchiness was attributed to inbreeding, putatively due to the low dispersal ability of the larvae. However, the genetic clustering analysis of E. singularis populations in the GoL revealed strong similarity among the individuals sampled in Cap d'Agde and Plateau des Aresquiers, which are separated by 30 km of sandy bottom that is unsuitable for colonization by the species (Weinberg 1978). Moreover, one station in the Côte Vermeille cluster displayed higher similarity to the stations at Cap d'Agde and Plateau des Aresquiers than to the other stations in Côte Vermeille. These similarities to the genetic profiles of distant stations led to an estimated genetic patch size of 30 km, meaning that the individuals found within an area of 30 km would be more genetically related than those taken randomly from any further distance. These results, together with the absence of genetic structure among E. singularis populations found within 15 km (Costantini et al. 2016) in the neighboring hydrodynamic province (Balearic Sea, Rossi et al. 2014), suggest that the heterozygote deficit in the E. singularis populations in the GoL should be attributed to on-going migration among open populations rather than to inbreeding in closed populations. Nevertheless, migration was not explained by isolation by distance (Sexton et al. 2013; Nanninga et al. 2014; Thomas et al. 2015), which pinpoints the limitation of using the

IBD model to explain the chaotic genetic patchiness resulting from larval dispersal via ocean flow (Selkoe et al. 2010). Indeed, undirected pairwise F_{ST} is not a good measure of genetic distance when migration among populations is asymmetrical (Beerli 1998; Broquet and Petit 2004; van Strien et al. 2015). In such a case, assignment tests or coalescent methods should be used to describe gene flow patterns and determine the factors shaping them (Manel et al. 2005).

Gene flow is determined by larval transport

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We tested three major successive filters that potentially influence gene flow during larval dispersal among habitat patches. First, the number of emigrants of E. singularis within separated habitat patches in the GoL was not explained by reproductive output. Second, the number of immigrants was not limited by settlement regulation due to intra-specific competition, although this mechanism is often presented (Roughgarden et al., 1985; Padron and Guizien 2015). It is interesting to note that conversely, the population density in the destination habitat patch increased with the number of immigrants, suggesting that the population densities observed in the GoL were below habitat saturating capacity. Third, larval transport explained a large portion of the gene flow among populations. In the GoL, the determination coefficient between contemporary gene flow and recent larval flow was extremely high for a 7-day larval transport duration among four habitat patches. The combination of the absence of a density-dependent regulation of *E. singularis* migrants in the GoL and the fact that gene flow is driven by unsteady larval transport support the alternate dominance of recruits vs. adult colony populations hypothesis that was observed for the same species along the Costa Brava (NW Mediterranean, Spain) (Linares et al. 2008). The strong correlation between recent larval transport and contemporary gene flow found in this study contrast with other studies based on the correlation with multi-generational gene flow (White et al. 2010; Alberto et al. 2011). This underpins the importance of testing the relationship between larval transport and gene flow estimated for a single dispersive event rather than for multiple

generations. Moreover, estimations at the single dispersive event scale enabled the discrimination of the efficiency of different PLDs in connecting populations.

The environment filters the PLDs that connect populations

Ex situ observations of survival expectancy and competency duration of *E. singularis* larvae (up to 40 days, K. Guizien and L. Bramanti, *Personal Communication*) suggest that all tested PLDs were equally probable in the absence of predation. In such a case, one would expect gene flow to be a composite of larval transport for all PLDs. However, gene flow among populations was best explained by PLDs ranging from 7 to 14 days in the habitat configuration of the GoL, and led to a spatially structured genetic population despite high local chaotic genetic patchiness. Similar finding was obtained among lobster populations along the Californian coast and related to upwelling intensity (Iacchei et al. 2013). In the present case of a lecithotrophic larva with a wide competence window, we suggest that larval mortality significantly altered larval transport patterns for periods longer than 14 days (Cowen et al. 2000). Correlatively, it suggests that mortality, particularly by predation, on *Eunicella singularis* larvae (2-3 mm) can be neglected up to a 14-day PLD in the GoL.

Treml et al. (2012) found that longer maximum PLDs lead to a more connected seascape on an ocean basin scale. The study claimed that maximum PLD is a good predictor of larval transport at large geographic scales (200-500 km) but a poor predictor at a mesoscale (<200 km). Our study evaluated the larval transport patterns among populations at a regional scale (up to 180 km) and showed that the PLD-indexed larval transport could significantly determine gene flow at a mesoscale. Our findings suggest that several PLDs should be tested along the competency period of a species rather than the average or the maximum PLD to provide a more accurate assessment of the importance of larval transport in shaping gene flow among sessile benthic species (D'Aloia et al. 2015).

Implications for conservation

Our results have important implications for regional conservation and the spatial

management of benthic sessile species in the GoL. Empirical gene flow patterns allow for the identification of key habitat patches of *E. singularis*: the high levels of gene flow between Cap d'Agde and Plateau des Aresquiers, along with transfers to other habitat patches, suggest that these two areas play a pivotal role in the spread of *E. singularis* in the region. Furthermore, the significant correlation between contemporary gene flow and recent larval transport for a PLD ranging from 7 to 14 days suggests that Cap d'Agde and Plateau des Aresquiers could be important sites for any species with similar dispersive traits (PLD, larval behavior, or reproductive period) dwelling on the same rocky substrate, as they should follow the same potential connectivity patterns. Hence, we advocate for enforcing the protection of those sites already identified as important for the regional persistence of soft-bottom species (Guizien et al. 2014), which also appears to be functionally important for the life cycle of some rocky species.

In summary, the methodological approach presented in this study highlights the hidden determinism within the apparent genetic chaos by incorporating interdisciplinary data (demography and ocean currents) and applying accurate genetic tools (coalescent-based analyses) to evaluate the gene flow among populations.

Authors' Contributions

M.P., L.B., and K.G. designed the study. S.B. performed all the laboratory work. M.P., F.C. and K.G. analyzed all the data. All authors contributed to the writing of the manuscript.

Data Accessibility

410 Microsatellite genotypes will be made available in Dryad.

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spatial distributions induced by coastal eddy in the Gulf of Lion. Insights from a numerical

- model. Progress in Oceanography 109:47–69.
- 448 Costantini F, Fauvelot C, Abbiati M (2007) Genetic structuring of the temperate gorgonian coral
- (Corallium rubrum) across the western Mediterranean Sea revealed by microsatellites and
- nuclear sequences. Molecular Ecology 16:5168–5182.
- 451 Costantini F, Gori A, López-González P, Bramanti L, Rossi S, Gili J-M, Abbiati M (2016) Lim-
- ited genetic connectivity between gorgonian morphotypes along a depth gradient. PLoS ONE
- 453 11:e0160678.
- Cowen R (2000) Connectivity of marine populations: open or closed? Science 287:857–859.
- 455 Cowen R (2006) Scaling of connectivity in marine populations. Science 311:522–527.
- D'Aloia C, Bogdanowicz S, Francis R, Majoris J, Harrison R, Buston P (2015) Patterns, causes,
- and consequences of marine larval dispersal. Proceedings of the National Academy of Sci-
- 458 ences 112:13940–13945.
- 459 Dawson M, Hays C, Grosberg R, Raimondi P (2014) Dispersal potential and population genetic
- 460 structure in the marine intertidal of the eastern North Pacific. Ecological Monographs
- 461 84:435–456.
- 462 Faurby S, Barber P (2012) Theoretical limits to the correlation between pelagic larval duration
- and population genetic structure. Molecular Ecology 21:3419–3432.
- 464 Gori A, Rossi S, Berganzo E, Pretus J, Dale M, Gili JM (2011) Spatial distribution patterns of the
- gorgonians Eunicella singularis, Paramuricea clavata, and Leptogorgia sarmentosa (Cape
- of Creus, Northwestern Mediterranean Sea). Marine Biology 158:143–158.
- 467 Guizien K, Brochier T, Duchêne JC, Koh BS, Marsaleix P (2006) Dispersal of Owenia fusiformis

468 larvae by wind-driven currents: turbulence, swimming behaviour and mortality in a three-di-469 mensional stochastic model. Marine Ecology Progress Series 311:47–66. 470 Guizien K, Belharet M, Moritz C, Guarini JM (2014) Vulnerability of marine benthic metapopu-471

lations: implications of spatially structured connectivity for conservation practice in the Gulf

- 472 of Lions (NW Mediterranean Sea). Diversity and Distributions 1–11.
- 473 Hanski I (1998) Metapopulation dynamics. Nature 396:41–49.
- 474 Holland L, Dawson D, Horsburgh G, Krupa A, Stevens J (2013) Isolation and characterization of 475 fourteen microsatellite loci from the endangered octocoral Eunicella verrucosa (Pallas 1766).
- 476 Conservation Genet Resour 5:825–829.
- 477 Hu ZY, Doglioli AM, Petrenko AA, Marsaleix P, Dekeyser I (2009) Numerical simulations of 478 eddies in the Gulf of Lion. Ocean Modelling 28:203–208.
- 479 Iacchei M, Ben-Horin T, Selkoe KA, Bird CE, Garcia-Rodrigues FJ, Toonen RJ (2013) Com-

bined analyses of kinship and FST suggest potential drivers of chaotic genetic patchiness in

- 481 high gene-flow populations. Mol Ecol 22:3476-3494.
- 482 Kinlan B, Gaines S (2003) Propagule dispersal in marine and terrestrial environments: a commu-483 nity perspective. Ecology 84:2007–2020.
- 484 Liggins L, Treml E, Riginos C (2013) Taking the plunge: an introduction to undertaking seascape
- 485 genetic studies and using biophysical models. Geography Compass 7:173–196.
- 486 Linares C, Coma R, Garrabou J, Diaz D, Zabala M (2008) Size distribution, density and distur-
- 487 bance in two Mediterranean gorgonians: Paramuricea clavata and Eunicella singularis.
- 488 Journal of Applied Ecology 45:688–699.

489 Manderson J (2016) Seascapes are not landscapes: an analysis performed using Bernhard Rie-490 mann's rules. ICES J. Mar. Sci. 73:1831-1838. 491 Manel S, Gaggiotti O, Waples R (2005) Assignment methods: matching biological questions with 492 appropriate techniques. Trends in Ecology & Evolution 20:136–142. 493 Mantel N (1967) The Detection of Disease Clustering and a Generalized Regression Approach. 494 Cancer research 27(2):209 -220 495 Millot C (1990) The Gulf of Lions' hydrodynamics. Continental shelf research 10:885–894. 496 Mitchell N, Dardeau M, Schaeffer W, Benke A (1992) Secondary production of gorgonian corals 497 in the northern Gulf of Mexico. Marine Ecology Progress Series 87:275–281. 498 Mokhtar-Jamaï K, Pascual M, Ledoux JB, Coma R, Féral JP, Garrabou J, Aurelle D (2011) From 499 global to local genetic structuring in the red gorgonian *Paramuricea clavata*: the interplay 500 between oceanographic conditions and limited larval dispersal. Molecular Ecology 20:3291-501 3305. 502 Molecular Ecology Resources Primer Development Consortium (2010) Permanent genetic re-503 sources added to molecular ecology resources database 1 August 2009 - 30 September 2009. 504 Molecular Ecology Resources 10:232–236. 505 Nanninga G, Saenz-Agudelo P, Manica A, Berumen ML (2014) Environmental gradients predict 506 the genetic population structure of a coral reef fish in the Red Sea. Molecular Ecology 507 23:591-602. 508 Padrón M, Guizien K (2015) Modelling the effect of demographic traits and connectivity on the 509 genetic structuration of marine metapopulations of sedentary benthic invertebrates. ICES 510 Journal of Marine Science. doi: 10.1093/icesjms/fsv158. 511 Paris C, Cherubin L, Cowen R (2007) Surfing, spinning, or diving from reef to reef: effects on 512 population connectivity. Marine Ecology Progress Series 347:285–300. 513 Pineda J, Hare J, Sponaugle S (2007) Larval transport and dispersal in the coastal ocean and con-514 sequences for population connectivity. Oceanologica Acta 20:22–39. 515 Pritchard J, Stephens M, Donnelly P (2000) Inference of population structure using multilocus 516 genotype data. Genetics 155:945–959. 517 Riginos C, Douglas K, Jin Y, Shanahan D, Treml E (2011) Effects of geography and life history 518 traits on genetic differentiation in benthic marine fishes. Ecography 34:566–575. 519 Roberts C (1997) Connectivity and management of Caribbean coral reefs. Science 278:1454-520 1457. 521 Rossi V, Ser-Giacomi E, López C, Hernandez-García E (2014) Hydrodynamic provinces and 522 oceanic connectivity from a transport network help designing marine reserves. Geophysical 523 Research Letter 41:2883-2891. 524 Roughgarden J, Iwasa Y, Baxter C (1985) Demographic theory for an open marine population 525 with space-limited recruitment. Ecology 66:54–67. 526 Ryman N, Palm S (2006) POWSIM: a computer program for assessing statistical power when 527 testing for genetic differentiation. Molecular Ecology 6:600–602. 528 Selkoe K, Watson J, White C, Horin T, Iacchei M, Mitarai S, Siegel D, Gaines S, Toonen R 529 (2010) Taking the chaos out of genetic patchiness: seascape genetics reveals ecological and 530 oceanographic drivers of genetic patterns in three temperate reef species. Molecular Ecology 531 19:3708-3726. 532 Selkoe K, D'Aloia C, Crandall E, Iacchei M, Liggins L, Puritz J, Heyden von der S, Toonen R 533 (2016) A decade of seascape genetics: contributions to basic and applied marine connectiv-534 ity. Marine Ecology Progress Series 554:1–19. 535 Sexton J, Hangartner S, Hoffmann A (2013) Genetic isolation by environment or distance: which 536 pattern of gene flow is most common? Evolution 68:1–15. 537 Shanks A, Grantham B, Carr M (2003) Propagule dispersal distance and the size and spacing of 538 marine reserves. Ecological Applications 13:159–169. 539 Théodor J (1967) Contribution à l'étude des gorgones. VII. Ecologie et comportement de la plan-540 ula. Vie Mileu 18:291–301. 541 Thomas L, Kennington W, Stat M, Wilkinson S, Kool J, Kendrick G (2015) Isolation by resis-542 tance across a complex coral reef seascape. Proceedings of the Royal Society B: Biological 543 Sciences 282:20151217. 544 Treml E, Roberts J, Chao Y, Halpin P, Possingham H, Riginos C (2012) Reproductive output and 545 duration of the pelagic larval stage determine seascape-wide connectivity of marine popula-546 tions. Integrative and Comparative Biology 52:525–537. 547 Ulses C, Estournel C, Durrieu de Madron X, Palanques A (2008) Suspended sediment transport 548 in the Gulf of Lions (NW Mediterranean): Impact of extreme storms and floods. Continental 549 shelf research 28:2048-2070. 550 van Strien MJ, Holderegger R, Van Heck HJ (2015) Isolation-by-distance in landscapes: consid-551 erations for landscape genetics. 114:27–37.

552 Weersing K, Toonen R (2009) Population genetics, larval dispersal, and connectivity in marine 553 systems. Marine Ecology Progress Series 393:1–12. 554 Weinberg S (1978) Mediterranean octocorallian communities and the abiotic environment. Ma-555 rine Biology 49:41–57. 556 Weinberg S, Weinberg F (1979) The life cycle of a gorgonian: *Eunicella singularis*. Bijdragen tot 557 de Dierkunde 48:1–14. 558 White C, Selkoe K, Watson J, Siegel D, Zacherl D, Toonen R (2010) Ocean currents help explain 559 population genetic structure. Proceedings of the Royal Society B: Biological Sciences 560 277:1685-1694. 561 Wilson G, Rannala B (2003) Bayesian inference of recent migration rates using multilocus geno-562 types. Genetics 163:1177–1191. 563 Wright S (1931) Evolution in Mendelian populations. Genetics 16:97–159.

Figure 1.- Locations of sampling stations and habitat patches of *Eunicella singularis* in the Gulf of Lion, NW Mediterranean. A) Black rectangles highlight the sampled habitat patches. Red areas show the distribution of the rocky habitat in the region. B) Sampling stations of *E. singularis* and rocky habitat extent in the Côte Bleue. C) Sampling stations of *E. singularis* and rocky habitat extent in the Côte Vermeille. D) Sampling stations of *E. singularis* and rocky habitat extent in Cap d'Agde and Plateau des Aresquiers. Red squares correspond to the areas of larval release considered in the hydrological model. Sampling stations are represented by gray dots. Information for each station is provided in Table 1.

Figure 2.- Plot of the Principal Coordinate Analysis (PCoA) from the genetic distance matrix assessed with pairwise Nei's genetic distance estimates of *E. singularis* in the Gulf of Lion. The first two axes explain 61.51% of the variation.

Figure 3.- Population structure of *E. singularis* in the Gulf of Lion, as revealed by Structure software. Each individual is represented by a vertical color line, with each color representing the proportion of membership to each cluster. Black lines represent the division among the identified clusters (K = 3). Values between parentheses correspond to the overall proportion of membership of the individuals in each cluster.

Figure 4.- Patterns of contemporary migration among habitat patches of *Eunicella singularis* in the Gulf of Lion. The matrix shows the number of migrants coming from any potential source and reaching any destination habitat patch. CV = Côte Vermeille, CA = Cap d'Agde, PA = Plateau des Aresquiers, and CB = Côte Bleue. The number of migrants between any two habitat patches is represented by the grayscale intensity.

Figure 5.- A) Relationship between the number of emigrants (sum of number of migrants by row) from a source habitat patch and the population density in that source habitat patch. B) Relationship between the number of immigrants (sum of number of migrants by column) to a destination habitat patch and the population density in that destination habitat patch. X error bars correspond to the standard deviation around the mean population density values. Y error bars correspond to the confidence intervals around the number of migrants estimated by BayesAss.

Figure 6.- Patterns of recent larval transport and contemporary gene flow among habitat patches of *E. singularis* in the Gulf of Lion. A) Recent larval transport for a PLD of 3.5 days; B) recent larval transport for a PLD of 7 days; C) recent larval transport for a PLD of 10.5 days; D) recent larval transport for a PLD of 14 days; E) recent larval transport for a PLD of 21 days; F) contemporary gene flow. The Grayscale indicates the probability values.

Figure 7.- Determination coefficient between contemporary gene flow and recent larval transport (including retention rates) for a PLD ranging from 3.5 to 42 days. Filled circles depict the PLDs for which the Mantel test showed significant correlations between contemporary gene flow and recent larval transport matrices, excluding retention rates.

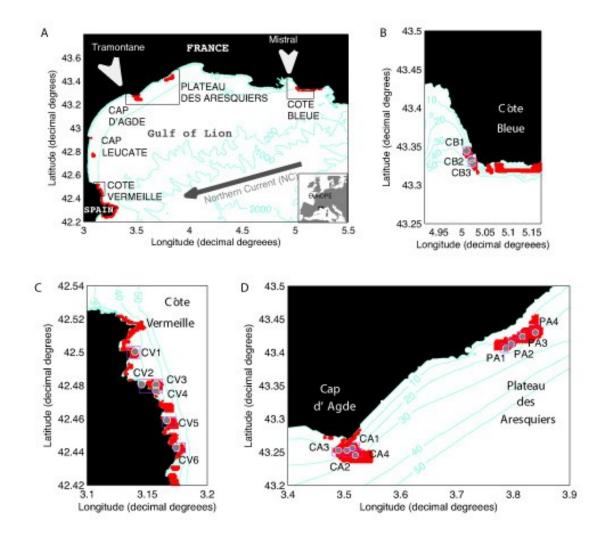


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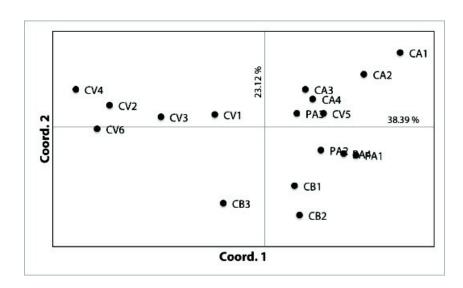


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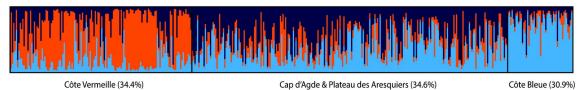


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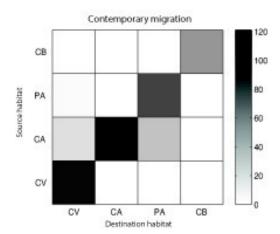


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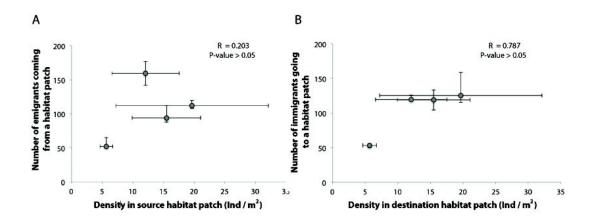


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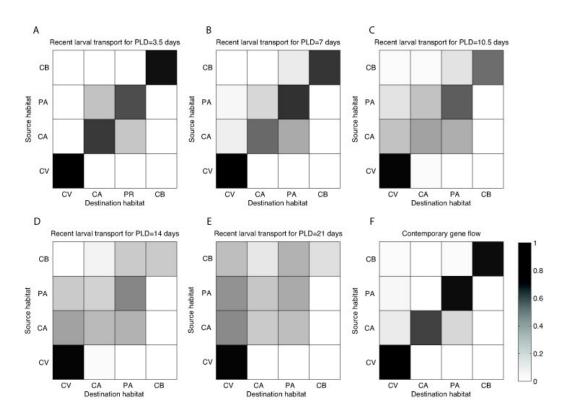


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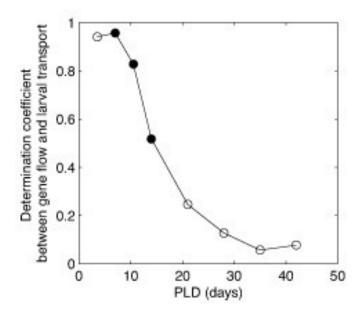


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