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David W Shanafelt, Michel Loreau. Stability trophic cascades in food chains. Royal Society Open Science, 2018, 5 (11), pp.180995. 10.1098/rsos.180995 . hal-02350567

HAL Id: hal-02350567

<https://hal.science/hal-02350567v1>

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Research



Cite this article: Shanafelt DW, Loreau M. 2018 Stability trophic cascades in food chains. *R. Soc. open sci.* **5**: 180995.
<http://dx.doi.org/10.1098/rsos.180995>

Received: 21 June 2018

Accepted: 8 October 2018

Subject Category:

Mathematics

Subject Areas:

computational biology/ecology/theoretical biology

Keywords:

cascade effects, food chain, invariability, stability

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Electronic supplementary material is available online at <https://dx.doi.org/10.6084/m9.figshare.c.4277321>.

Stability trophic cascades in food chains

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While previous studies have evaluated the change in stability for the addition or removal of individual species from trophic food chains and food webs, we know of no study that presents a general theory for how stability changes with the addition or removal of trophic levels. In this study, we present a simple model of a linear food chain and systematically evaluate how stability—measured as invariability—changes with the addition or removal of trophic levels. We identify the presence of trophic cascades in the stability of species. Owing to top-down control by predation and bottom-up regulation by prey, we find that stability of a species is highest when it is at the top of the food chain and lowest when it is just under the top of the food chain. Thus, stability shows patterns identical to those of mean biomass with the addition or removal of trophic levels in food chains. Our results provide a baseline towards a general theory of the effect of adding or removing trophic levels on stability, which can be used to inform empirical studies.

1. Introduction

The works of Elton [1], Lindeman [2], May [3], Pimm [4] and others sparked a plethora of research in the field of food webs [5,6], yet we know of no work that systematically evaluates the effect of food chain length on the stability of linear food chains. In this paper, we present a simple model of a linear food chain, illustrate how stability changes with the addition of trophic levels, and identify the presence of trophic cascades in the stability of species.

Since the initial contributions of Elton [1] and Lindeman [2], there have been numerous theoretical and empirical studies investigating the many aspects of food webs including the identification of the structural properties of natural food webs [7], classification of species into trophic levels [8,9], construction of artificial food webs with the properties of real ones [10,11] and quantification of food web structure and complexity

[12–15]. The literature on food webs is vast, and we make no attempt to review it in its entirety. Books by Pimm [4], DeAngelis [16], Loreau [5] and McCann [6] provide comprehensive perspectives of the literature, while papers by Paine [17], DeAngelis *et al.* [18], Morin & Lawler [19], Polis & Strong [8], Amarasekare [20], Hall [21], and Brose *et al.* [22] give more focused reviews.

The most relevant part of the food web literature to our work lies in the observations of direct and indirect effects of the change in a species' abundance on the overall food web, often broadly termed 'cascade effects' [23–25]. The effect of addition or removal of a species cascades through all levels of the food chain. How a particular species is affected depends on its level in the food chain and the length of the food chain. Growth of species at odd-numbered levels are limited by top-down control by predation; even-numbered species are limited by bottom-up regulation by prey [24,26,27]. This pattern in the distribution of biomass between trophic levels at equilibrium was later extended to consider more complex food web structures [28], convex and logistic growth functions [29] and biodiversity within trophic levels [30]. Empirical evidence suggests that cascade effects are common in nature [31–33], though the strength of the effect is likely to be weaker in terrestrial than in aquatic ecosystems [31].

In our study, we focus on how stability changes with the addition or removal of trophic levels in a food chain. Much of the initial analyses on the stability of food webs centred on the relationship between food web complexity and stability, and the types and strengths of species interactions necessary to achieve it [3,12,34]. Many of these studies considered the stability of a fixed point equilibrium or the probability of encountering such a point, and numerous papers have evaluated the equilibrium dynamics of specific food web structures (see Rosenzweig [35], Hastings & Powell [36], DeAngelis [16] and Post *et al.* [37] for examples of three-species food chains). Recently, scientists have shifted to general measures of stability such as persistence, resilience, resistance and variability [38,39]. While these measures are interrelated [40,41], they capture different perspectives of stability and differ in their applicability to empirical data. While we are aware of other studies that evaluate the change in stability for the addition or removal of an individual species [31,42,43], we know of no general theory for how stability changes with the addition or removal of trophic levels.

Beginning with a uni-trophic single species model, we systematically build up to a penta-trophic linear food chain. We evaluate how stability—defined as invariability, the inverse of temporal variability [44–46]—changes with the addition of trophic levels. The most similar studies to ours are the works of Hairston *et al.* [25], Fretwell [26] and Oksanen *et al.* [24], which were generalized by Loreau [5]. They found that equilibrium biomass for each species exhibits a distinct pattern with the addition of trophic levels, which depends on the level of top-down and bottom-up controls. A species possesses its highest equilibrium biomass if it is at the top of the food chain, the lowest when it is just under the top, and a consistent intermediate switching pattern for higher-order food chains (figure 1). However, the focus of these studies—and their subsequent extensions [28–30]—has been solely on patterns in the distribution of equilibrium biomass.

We extend this literature to consider patterns of stability across food chains and in doing so highlight the importance of trophic dynamics on stability. We hypothesize that we will find a similar pattern to equilibrium biomass. That is, top-down control by predation and bottom-up regulation will alter the mean (equilibrium) and temporal standard deviation of biomass to drive the stability of individual species in the food chain. We predict that top-down predation by the top predator will be the most destabilizing, while bottom-up regulation will stabilize the biomass of the top species. Indeed, we find that a species has the highest stability when it is at the top of the food chain (highest equilibrium biomass) and lowest when it is just under the top (lowest equilibrium biomass).

Our results provide a baseline for a general theory for the effect of adding or removing trophic levels on stability, which can be tested in empirical studies.

2. Methods

2.1. The model

We consider a general food chain model based on the ecosystem models of DeAngelis [16] and Loreau & Holt [47]. By varying the number of trophic levels we study how stability—defined as the ratio of the mean and temporal standard deviation of biomass—changes with the addition or removal of trophic levels.

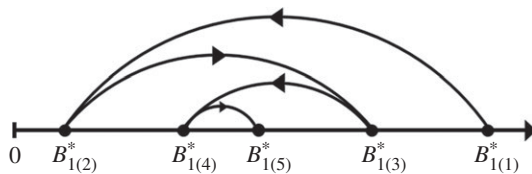


Figure 1. Visual representation of how equilibrium biomass changes with the addition of trophic levels [5]. The notation $B_{i(j)}^*$ represents the equilibrium biomass of species i for a trophic food chain of length j . Results are presented for the basal resource, although the same pattern holds for all species (electronic supplementary material, B).

For a trophic food chain of length n , the change in the biomass of each species is given by

$$\left. \begin{aligned} \frac{dB_1}{dt} &= I(1 + \varepsilon_1) - lB_1 - a_2B_1B_2 + B_1\varepsilon_1, \\ \frac{dB_2}{dt} &= a_2b_2B_1B_2 - a_3B_2B_3 - m_2B_2 + B_2\varepsilon_2, \\ \frac{dB_3}{dt} &= a_3b_3B_2B_3 - a_4B_3B_4 - m_3B_3 + B_3\varepsilon_3, \\ &\vdots \\ \frac{dB_n}{dt} &= a_nb_nB_{n-1}B_n - m_nB_n + B_n\varepsilon_n, \end{aligned} \right\} \quad (2.1)$$

and

where B_1 represents the basal (abiotic) resource, B_2 represents a primary producer (plant), B_3 represents a primary consumer (herbivore) and B_i for $i > 3$ represent secondary consumers (carnivores). The parameters I and ε_i govern the quantity of resource influx that flows into the community. The parameter I is the baseline quantity of resource influx. We introduce stochasticity in the rate of resource influx (growth rate of the resource), where ε_i is an independent, normally distributed variable with a mean of zero and variance of ζ_i .¹ Resources are lost at a constant proportion l . The predation rates, biomass conversion efficiencies and mortality rates of consumers are given by a_i , b_i and m_i , respectively.

The final group of terms in (2.1) represents the effect of environmental stochasticity on the growth rates of each species. White noise is applied to each species as an independent, normally distributed random variable ε_i with a mean of zero and variances of ξ_i . Each stochastic effect only affects a single trophic level and does not exhibit temporal autocorrelation. We motivate our choice of stochasticity for two reasons. Empirically, we would expect all species to exhibit some degree of temporal variation in their population biomass (*sensu* Schaffer *et al.* [48]), and our formulation is a base case of modelling this. Second, in the absence of noise the system of equations in (2.1) converges to a stable, non-fluctuating equilibrium. In order to measure stability without perturbing the system, we require some sort of fluctuation in biomass around equilibrium [41].

We vary the number of trophic levels in (2.1) to consider uni-trophic (only the basal resource), bi-trophic (basal resource and primary producer), tri-trophic, quadri-trophic and penta-trophic food chains. The last three food chains include one, two and three consumers, respectively.

2.2. Calculating stability

In order to study stability across trophic food chains, we assume an interior solution in which all species coexist—a reasonable assumption when studying linear food chains. We define stability as invariability, the inverse of temporal variability or the ratio of the mean and standard deviation of biomass [44–46]. While other stability measures exist in the literature, invariability is simple to interpret, straightforward to derive analytically and numerically, and can be readily applied to empirical data [44–46]. We derive invariability in two ways. First, we analytically calculate invariability via the Lyapunov equation [49,50]. This requires linearizing the system and studying the dynamics around the equilibrium in

¹In effect, stochasticity in the rate of resource influx acts as a bottom-up regulation in the growth rate of the resource. In the absence of stochasticity, the growth (influx) rate of the basal resource is taken as constant, which has implications for the pattern of stability of the basal resource across food chains (electronic supplementary material, B).

response to a slight perturbation. We then calculate invariability numerically by simulating the system of equations of each food chain. While we still restrict our analysis to cases in which all species coexist, the numerical method relaxes the assumption of linearization and allows the system to deviate from the region close to equilibrium.

To analytically calculate invariability, we derive the stationary covariance matrix in the vicinity of the equilibrium population dynamics of each food chain. This tells us how each species responds to stochasticity in its own biomass and the biomass of other species. The system of equations representing the equilibrium (linearized) dynamics of the food chain is given by

$$\frac{d\mathbf{X}}{dt} = \mathbf{A}\mathbf{X}(t) + \mathbf{E}(t), \quad (2.2)$$

where $\mathbf{X}(t)$ is a vector representing the difference between the states of the system at time t and their equilibrium values, \mathbf{A} is a matrix describing the deterministic dynamics of the system around the equilibria (the Jacobian evaluated at equilibrium) and $\mathbf{E}(t)$ is a vector capturing the effects of stochasticity on the dynamics around the equilibrium. Given the functional forms of stochasticity in (2.1), the vector $\mathbf{E}(t)$ is written explicitly as

$$E(t) = \begin{bmatrix} I & B_1^* & 0 & 0 & 0 \\ 0 & 0 & B_2^* & 0 & 0 \\ 0 & 0 & 0 & \ddots & \vdots \\ 0 & 0 & 0 & \cdots & B_n^* \end{bmatrix} \begin{bmatrix} \varepsilon_I \\ \varepsilon_1 \\ \vdots \\ \varepsilon_n \end{bmatrix}, \quad (2.3)$$

where B_i^* is the equilibrium biomass of species i . It is worth recalling that each stochastic event affects each species independently and is not correlated through time.

The state variables of equation (2.2) can be seen as stochastic variables. To calculate the stationary covariance matrix, we evaluate the variance of the left- and right-hand sides of equation (2.2). This collapses to the Lyapunov equation

$$AC + AC^T + MV_\varepsilon M^T = 0, \quad (2.4)$$

where A is defined above, and the matrices M and V_ε are defined as

$$M = \begin{bmatrix} I & B_1^* & 0 & 0 & 0 \\ 0 & 0 & B_2^* & 0 & 0 \\ 0 & 0 & 0 & \ddots & \vdots \\ 0 & 0 & 0 & \cdots & B_n^* \end{bmatrix} \quad (2.5)$$

and

$$V_\varepsilon = \begin{bmatrix} \xi_I & 0 & 0 & 0 \\ 0 & \xi_1 & 0 & 0 \\ 0 & 0 & \ddots & \vdots \\ 0 & 0 & \cdots & \xi_n \end{bmatrix}. \quad (2.6)$$

The matrix V_ε is the variance–covariance matrix of stochasticity. The diagonal entries of V_ε represent the variances of white noise imposed on the rate of resource influx and on each species (ξ_I and ξ_i , respectively).

The matrix C is the stationary covariance matrix that satisfies the Lyapunov equation [41,45,51]. The diagonals of C are the variances of the various species around their equilibria given environmental stochasticity. The off-diagonals are the covariances between species, e.g. how the equilibrium biomasses of two species covary around their equilibria in response to a stochastic effect in the other species.

We calculate invariability directly from the matrix C . We derive invariability per species as the inverse of the coefficient of variation of biomass for each species, e.g. the mean (equilibrium) biomass of each species divided by its standard deviation of biomass (the square root of the diagonal elements of C). Invariability of the entire system is defined as the inverse of the coefficient of variation in the total biomass of all species in the system. An illustration of the Lyapunov method for a bi-trophic system can be found in electronic supplementary material, A.

In addition, we evaluate invariability of each trophic food chain numerically. We simulate the system of nonlinear equations of each trophic food chain in (2.1) using a first-order Euler approximation with step size Δt . Initial conditions are set at the equilibrium values of each species in the absence of stochasticity, and parameter values held constant across food chains. We run each simulation for a time horizon T and check that no species went extinct during the simulation. We calculate

Table 1. Model parameters. The subscript i indicates the species or trophic level: basal resource (1), primary producer (2), primary consumer (3), secondary consumer (4) and tertiary consumer (5). For numerical simulations, initial conditions were taken as the equilibrium values in the absence of stochasticity. Parameter values were chosen to guarantee the presence of a stable, interior equilibrium and were held constant across food chains (e.g. $a_2 = 0.2$ for bi- to penta-trophic food chains). Variances of white noise were equal for all species and were chosen such that stochastic effects neither perturbed the system out of the coexistence basin of attraction nor caused the stochastic additive terms to exceed mortality.^a

parameter	interpretation	value
I	baseline rate of resource influx	variable [5, 150]
l	resource loss rate	0.1
a_i	predator predation rate	$a_2 = a_3 = a_4 = 0.2$ $a_5 = 0.5$
b_i	biomass conversion efficiency	$b_2 = b_3 = 0.3$ $b_4 = b_5 = 0.4$
m_i	mortality rate	0.1
ε_l	white noise (rate of resource influx)	$\sim N(0, \zeta_l)$
ε_i	white noise (per species)	$\sim N(0, \xi_i)$
ζ_l	variance of white noise (rate of resource influx)	0.0100
ξ_i	variance of white noise (per species)	0.0025
T	simulation time horizon	1500
Δt	simulation step size	0.1

^aAs pointed out by a reviewer, in order to be a truly self-contained food chain the effect of stochasticity on population biomass should come in the form of mortality only. Indeed, if a positive stochastic effect were to exceed mortality then it would result in a net input or subsidy to biomass.

invariability for each species and total biomass from the final 1000 time steps of each simulation. A full list of parameters is found in table 1.²

3. Results

3.1. Equilibrium biomass, temporal standard deviation in biomass and synchrony between species

We recover the well-established equilibrium biomass results characteristic of top-down predator control in linear food chains [5,24–26]. That is, a species has its highest equilibrium biomass when it is at the top of the food chain, and the lowest when it is at the level below the top (figure 1; electronic supplementary material, B). In the absence of top-down control, species growth is limited solely by the quantity of prey. By contrast, when a species is just below the top of the food chain, it experiences top-down control by an unbounded predator and bottom-up limitation by the availability of its prey. As we increased the number of trophic levels, there is a pattern of intermediate switching of equilibrium biomass for each species.

In addition, we recover the well-known pattern in the response of equilibrium biomass to nutrient enrichment (figures 2–5) [24,26]. For all food chains, total biomass increases with nutrient enrichment. However, the equilibrium biomass of each species depends on its place in the food chain. If the number of trophic levels is odd, then equilibrium biomass increases with the rate of resource influx in odd-numbered trophic levels and is constant in even-numbered ones. If the number of trophic levels is even, then equilibrium biomass increases with resource influx in even-numbered levels and is constant in odd-numbered ones. Consider, for example, a trophic food chain with n levels. The n th level consumer (unbounded top predator) exerts top-down control on the $n - 1$ level consumer, which

²Analytical solutions for invariability were calculated in Wolfram Mathematica 10.2. Numerical simulations were conducted in Matlab 2016. Source code for the analytical and numerical solutions can be found in the electronic supplementary material, C and D, or on the Open Science Framework (<https://osf.io/hxfd2/>).

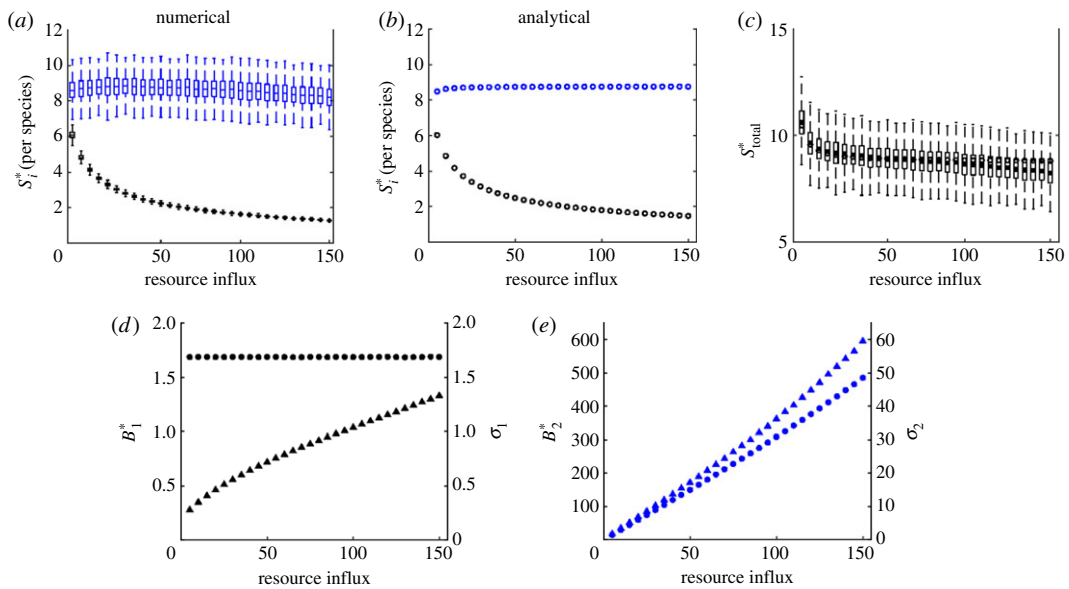


Figure 2. Stability in a *bi-trophic* food chain. (a–c) Numerical and analytical results for stability per species, and stability of total biomass. Invariability for each species is calculated as the ratio of equilibrium biomass and the standard deviation of biomass (d,e). Colour and subscript indicate species: basal resource (black, 1) and primary producer (blue, 2). In (a,c), solid bars indicate the mean values of 1000 simulations. Boxes represent the 25th and 75th percentiles of simulation results. Whiskers correspond to approximately 2.5 standard deviations from the mean. In (d,e), the marker shape denotes equilibrium biomass (circle) and standard deviations of biomass (triangle). Markers are the mean value of the simulation results.

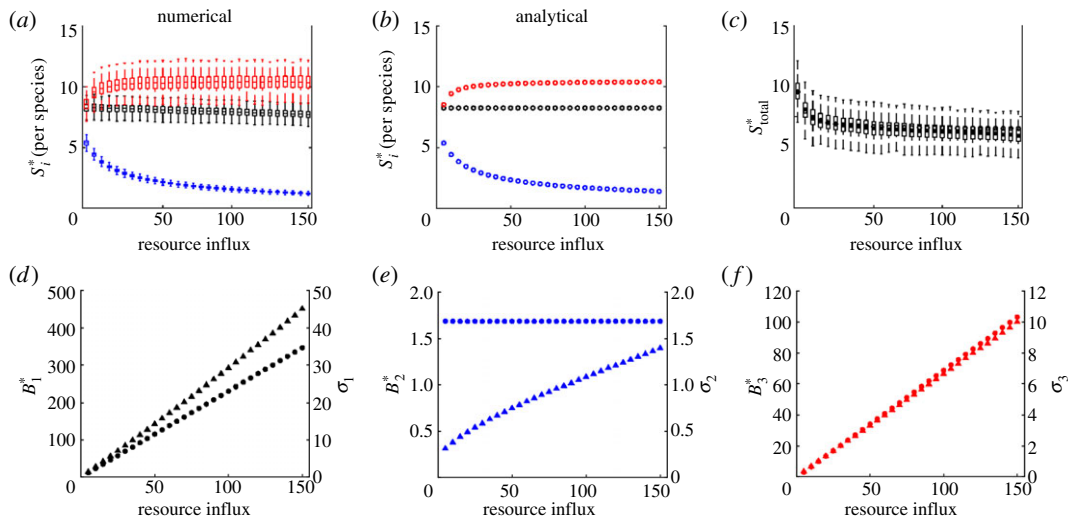


Figure 3. Stability in a *tri-trophic* food chain. (a–c) Numerical and analytical results for stability per species, and stability of total biomass. Invariability for each species is calculated as the ratio of equilibrium biomass and the standard deviation of biomass (d–f). Colour and subscript indicate species: basal resource (black, 1), primary producer (blue, 2) and primary consumer (red, 3). In (a,c), solid bars indicate the mean values of 1000 simulations. Boxes represent the 25th and 75th percentiles of simulation results. Whiskers correspond to approximately 2.5 standard deviations from the mean. In (d–f) the marker shape denotes equilibrium biomass (circle) and standard deviations of biomass (triangle). Markers are the mean value of the simulation results.

relieves predation pressure on the $n - 2$ consumer. The $n - 2$ consumer then exerts top-down control on the $n - 3$ level, which relieves predation pressure on the $n - 4$ level (and so on).

We find the same patterns in the temporal standard deviations in biomass for each species. A species has the highest standard deviation in biomass if it is at the top of the food chain, lowest if it is just under the top, and a pattern of switching between intermediate trophic levels (electronic supplementary material, B). For odd-numbered (even-numbered) species, the temporal standard deviation in biomass increases with the baseline rate of resource influx when the number of trophic levels is odd (even),

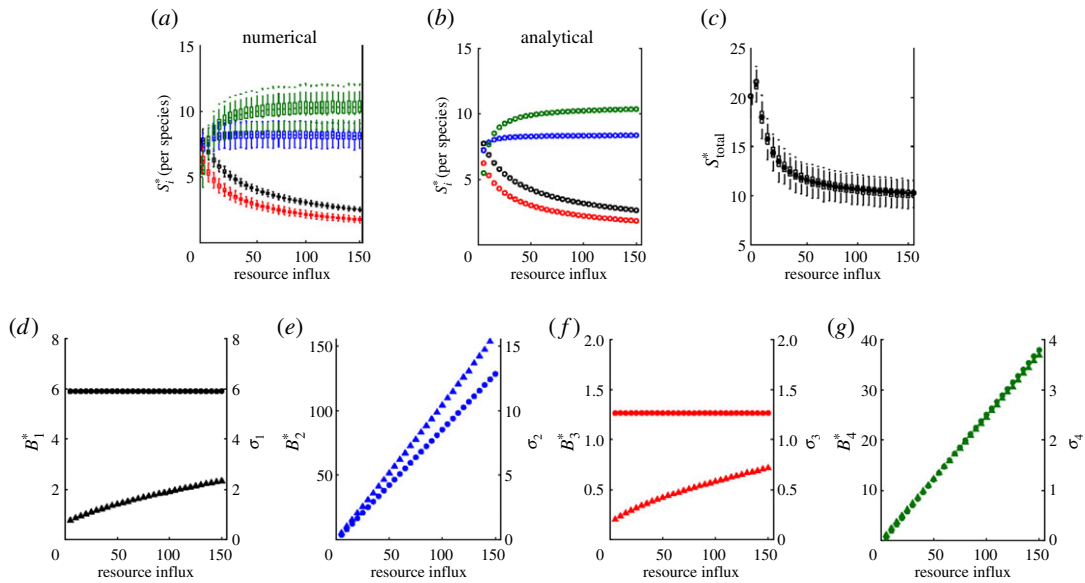


Figure 4. Stability in a *quadri-trophic* food chain. (a–c) Numerical and analytical results for stability per species, and stability of total biomass. Invariability for each species is calculated as the ratio of equilibrium biomass and the standard deviation of biomass (d–g). Colour and subscript indicate species: basal resource (black, 1), primary producer (blue, 2), primary consumer (red, 3) and secondary consumer (green, 4). In (a,c), solid bars indicate the mean values of 1000 simulations. Boxes represent the 25th and 75th percentiles of simulation results. Whiskers correspond to approximately 2.5 s.d. from the mean. In (d–g), the marker shape denotes equilibrium biomass (circle) and standard deviations of biomass (triangle). Markers are the mean value of the simulation results.

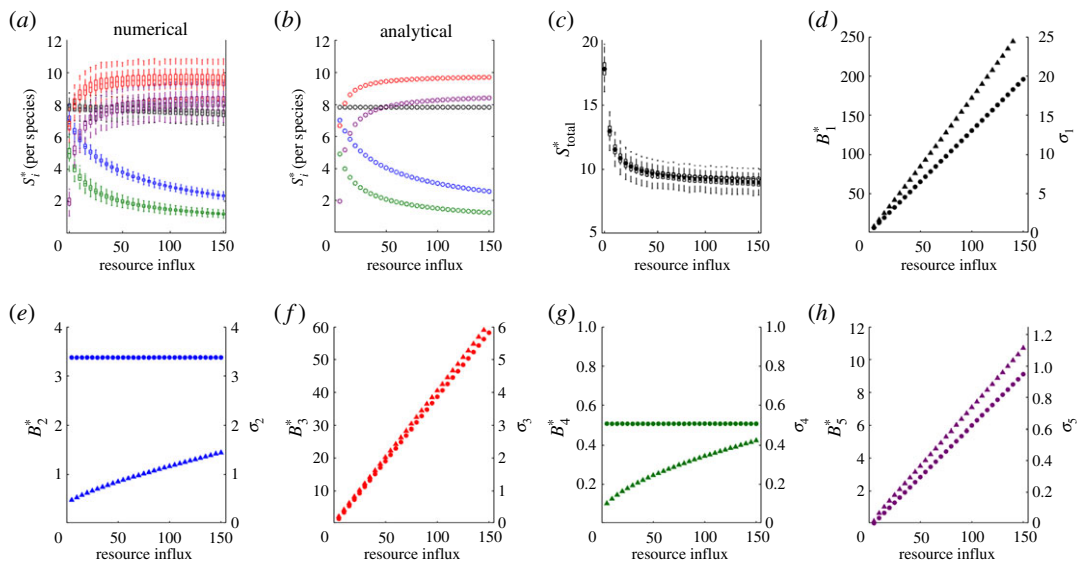


Figure 5. Stability in a *penta-trophic* food chain. (a–c) Numerical and analytical results for stability per species, and stability of total biomass. Invariability for each species is calculated as the ratio of equilibrium biomass and the standard deviation of biomass (d–h). Colour and subscript indicate species: basal resource (black, 1), primary producer (blue, 2), primary consumer (red, 3), secondary consumer (green, 4) and tertiary consumer (purple, 5). In (a,c), solid bars indicate the mean values of 1000 simulations. Boxes represent the 25th and 75th percentiles of simulation results. Whiskers correspond to approximately 2.5 s.d. from the mean. In (d–h), marker shape denotes equilibrium biomass (circle) and standard deviations of biomass (triangle). Markers are the mean value of the simulation results.

and increases at a lower rate when the number of trophic levels is even (odd). As we will see below, these results have implications on stability.

Species covariances—calculated from the off-diagonals of the stationary covariance matrix—are consistent with our findings of equilibrium biomass and the temporal standard deviations in biomass (electronic supplementary material, B). How the biomass of one species fluctuates around its

equilibrium in response to a perturbation in the biomass of another species depends on the trophic level of each species and the length of the food chain. The covariance between a predator and its prey is always negative. The cascade effects described above govern the covariances of the remaining species. In bi-, tri- and quadri-trophic food chains the covariances between two odd-numbered (or two even-numbered) species is positive, while the covariances between an odd-numbered and an even-numbered species are negative. The covariances of species in a penta-trophic food chain are generally consistent with this, though there is more nonlinearity in the response of the system to increases in the baseline rate of resource influx.

3.2. Stability within individual food chains

Our results from bi-trophic to penta-trophic food chains are presented in figures 2–5. We find a general agreement between analytical predictions derived from the linear approximation and numerical simulations of the full nonlinear system.

In a bi-trophic food chain, predation by the primary producer destabilizes the biomass of the basal resource (figure 2*a–c*). Equilibrium biomass of the basal resource remains constant as the baseline level of resource influx increases. Any new biomass is consumed by the primary producer, whose equilibrium biomass increases with the baseline level of resource influx. The temporal standard deviations in biomass of the two species increase with the baseline resource influx rate (figure 2*d,e*). However, the standard deviation in biomass for the primary producer increases at approximately the same rate as equilibrium biomass, causing stability to remain more or less constant.

For tri-, quadri- and penta-trophic food chains, stability patterns are more complicated. At low baseline levels of resource influx, all species coexist but the resource limits the biomass of the higher trophic levels. The top predator exists but at low biomass and low stability. Additional nutrients provide more energy to the system that allow additional biomass and stabilize the system. However, as the baseline rate of resource influx increases and the equilibrium biomass of each species follows, we observe greater exploitation of prey by the top predator. Stability of the species just under the top of the food chain declines, as does the stability of total biomass. Indeed, these results generalize the paradox of enrichment to dynamics in the vicinity of a stable equilibrium [52].

In all food chains, however, the species just below the top of the food chain is the least stable (figures 3–5). The species just below the top of the food chain and those at alternating levels below it experience both top-down control and bottom-up regulation. At equilibrium, these species experience pure top-down control; in the vicinity of its equilibrium, these species experience fluctuations due to bottom-up processes. But in contrast to other species that experience similar top-down control and bottom-up regulation, the species just below the top of the food chain is consumed by an unbounded predator. Biomass of the top predator is regulated solely by the availability of its prey, and the top predator consumes any new biomass of the species just below the top of the food chain.

The lack of top-down control on the biomass of the top predator is the reason that predation by the top predator is more destabilizing than the top-down control of another species. Focusing on the quadri- and penta-trophic food chains, as we increase the baseline rate of resource influx we find that the equilibrium biomass of the species just below the top of the food chain and that of the second trophic level below it remain constant while their standard deviations in biomass increase (figures 4 and 5; electronic supplementary material, B). While the latter species has a higher standard deviation in biomass, the equilibrium biomass of the species just under the top is lower due to predation by the unbounded top predator, resulting in a lower stability. Our results hold when varying other model parameters, such as the rate of predation by the primary producer (electronic supplementary material, B).

3.3. Stability across food chains

Like our results on equilibrium biomass, we find a consistent pattern in the stability of species across food chains (figure 6; electronic supplementary material, B). Species are most stable when they are at the top of the food chain, and least stable when they are just below it. When species are at intermediate positions within food chains, we find a switching pattern of stability. Bottom-up regulation of growth provides a stabilizing effect for the top predator. Top-down control by an unbounded top predator is most destabilizing. At intermediate positions within the food chain, the stability of a species depends on top-down control caused by the cascade effects of the top predator. In even-numbered (odd-numbered) food chains, the top predator relieves predation pressure on other even-numbered (odd-numbered) species and the equilibrium biomasses and standard deviations in biomass increase with

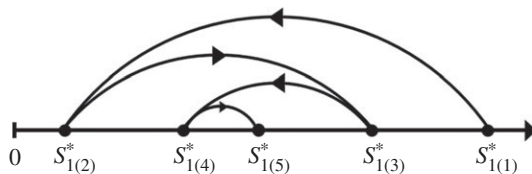


Figure 6. Visual representation of how stability changes with the addition of trophic levels. The notation $S_{i(j)}^*$ represents the stability of species i for a trophic food chain of length j . Results are presented for the basal resource, although we observe the same pattern for all species (electronic supplementary material, B).

the rate of resource influx. At the same time, any new biomass of odd-numbered (even-numbered) species is consumed by even-numbered (odd-numbered) species. Equilibrium biomasses of odd-numbered (even-numbered) species remain constant with the rate of resource influx while the standard deviations in biomass increase. Stability of odd-numbered (even-numbered) species declines, scaling with the level of equilibrium biomass.

Thus we can summarize our findings by the inequalities

$$\left. \begin{array}{ll} S_{1(2)}^* < S_{1(4)}^* < S_{1(5)}^* < S_{1(3)}^* < S_{1(1)}^* & \text{(basal resource)} \\ S_{2(3)}^* < S_{2(5)}^* < S_{2(4)}^* < S_{2(2)}^* & \text{(primary producer)} \\ S_{3(4)}^* < S_{3(5)}^* < S_{3(3)}^* & \text{(primary consumer)} \\ S_{4(5)}^* < S_{4(4)}^* & \text{(secondary consumer),} \end{array} \right\} \quad (3.1)$$

where $S_{i(j)}^*$ is the stability of species i in trophic network j (figure 6).

The observed pattern of stability is driven by the interplay between bottom-up regulation of prey and top-down control by predation, which affect the equilibrium biomass and temporal standard deviation in biomass of each species in our food chains. Stability (as invariability) is calculated as the ratio of equilibrium biomass to the temporal standard deviation in biomass. If equilibrium biomass scales at the same rate as the standard deviation in biomass, then stability is insensitive to the baseline rate of resource influx. If equilibrium biomass remains constant while the standard deviation in biomass increases, then stability declines.

4. Discussion

Our results for mean biomass and stability within single food chains are generally consistent with existing theory [4,16,35,36] and empirical data [32,53]. For mean biomass, we recover the consistent ‘alternation of regulation’ in equilibrium biomass characteristic of top-down control by predation in linear trophic food chains [24–26]. We also capture the effects of bottom-up regulation due to nutrient or prey limitation [8,16].

For stability, we identify cascade effects due to the addition of novel predators in the food chain. Like the pattern for mean biomass, we find a general pattern in the stability of each species as we add trophic levels. We show that the stability of a species is highest when it is at the top of the food chain, lowest when it is just below the top level, and exhibits a switching pattern at intermediate levels. While previous research has looked at the equilibrium biomass [5,24], the stability of specific food web structures [16,20] and secondary extinctions associated with the removal of species [42,54–56], we provide a first step towards a general theory on how the stability of species changes with the addition of trophic levels. Further, we generalize the paradox of enrichment to the vicinity of stable equilibria. The paradox of enrichment was established for locally unstable systems, e.g. the emergence of limit cycles in response to the addition of nutrients [52]. By analysing dynamics around stable equilibria, we demonstrate the destabilizing effect of nutrient addition and extend the paradox of enrichment to locally stable systems.

Our results have implications outside the food web literature. We find an overall destabilizing effect of resource enrichment, increasing the variability of total biomass in all food chains and of the species just below the top in the tri- to penta-trophic chains. This relates to the literature on the effects of nutrient addition on natural systems. In general, there is evidence that nutrient loading can be destabilizing [16,52,57,58] and the literature on allochthonous inputs suggests that additional resources can be stabilizing or destabilizing, depending on the species affected and preference for resources [8,16,59]. In aquatic systems, the effects of nutrient addition—primarily from terrestrial inflows—and eventual

eutrophication of aquatic ecosystems are well documented [60–62]. Empirical evidence suggests the presence of tipping points in the quantity of nutrient loads prior to the transition to a eutrophic state [63]. In our study, extinctions are not possible (by assumption), but our results indicate regions of the parameter space where further nutrient additions will push the system out of the coexistence basin of attraction (electronic supplementary material, B). We further observe differential effects of nutrient loading for each trophic level and nonlinear declines in the stability of total biomass with the addition of nutrients. This suggests that not only is nutrient loading destabilizing, but that this destabilization manifests itself differently depending on the trophic level and structure of the food chain. Nutrient additions enter the food chain via the basal resource, the effects of which then filter up the food chain based on the trophic dynamics in play at each trophic level. Depending on the trophic level and length of the food chain, the means (equilibrium) and temporal standard deviations in biomass can change in different ways in response to nutrient loading.

While we present a general theory for the addition of trophic levels on stability, we do so for simple linear food chains. Many extensions are beyond the scope of our analysis. For example, it would be useful to extend these results to more complex species interactions and food web structures. Though we present our results for type I functional responses, we would expect our conclusions to hold for more complex predator response functions. We present numerical results for a type II functional response in electronic supplementary material, B. Like the type I case, we restrict our analysis to scenarios in which all species coexist. In general, we find the same qualitative results, though there is a larger degree of variation in stability associated with the addition of trophic levels. Oscillatory behaviour destabilizes the system, with species more likely to go extinct due to environmental stochasticity during periods of low population biomass. Other types of interactions and food web structures could include mutualistic interactions [64,65], omnivory and multiple prey species [59,66], intraguild predation [67], parasitism [68] and cannibalism [6,16,20]. Similarly, topological properties of the trophic network such as looping [69] and modularity [70] affect the presence of locally stable equilibria, and spatial structure can alter the dynamics of food webs compared with single-patch systems [20,71]. Indeed, Jager & Gardner [28] and Abrams [30] demonstrated that the ‘alternation of regulation’ pattern in equilibrium biomass does not always hold in more complex food webs, and it would be useful to test how more complex dynamics alter patterns in stability.

We model stochasticity as environmental stochasticity. It would be interesting to model other types of stochasticity—such as demographic stochasticity—though we would expect this to matter most when species populations are small [72]. Similarly, we take stochasticity as a series of independent events uncorrelated across species and time. Other possible extensions include relaxing this assumption by varying the relative strength of stochastic effects across trophic levels or having stochastic effects be correlated between species and time. Environmental shocks—natural or anthropogenic—often affect multiple species regardless of their place in the food chain [73–75], and preliminary results suggest that varying the relative strength of stochastic effects across trophic levels can change stability, though in general the pattern is consistent with our observations [76]. For example, one could envision scenarios where top predators are more sensitive than others to disturbance or perturbation or have greater restrictions on their growth [77,78]. We test the robustness of our results to the top predator possessing higher levels of stochasticity (relative to other species) and density-dependent mortality (electronic supplementary material, B). In each case, we find that our pattern of stability generally holds, particularly for higher trophic levels. When patterns deviate, they do so in a way consistent with our intuition of how stability cascades down the food chain. We leave a detailed analysis of these extensions for future work.

It is possible to test the predictions of our model empirically, such as with micro- or mesocosm experimental set-ups [56,79,80]. In these cases, researchers will require model species with low enough generation times for the system to reach equilibrium and be able to exert enough control on the system to enforce a strict food chain interaction structure (e.g. no omnivory). To calculate stability as invariability, researchers would need to gather data on population abundances (biomass or density) over time [38,46,81].

Data accessibility. We include source code for the analytical and numerical solutions in the electronic supplementary material, C and D. Alternatively, all source code and model results can be retrieved from the Open Science Framework (<https://osf.io/hxfcd2/>).

Authors' contributions. D.W.S. analysed the model and drafted the manuscript. M.L. conceived the study and helped draft the manuscript. Both authors gave final approval for publication.

Competing interests. We have no competing interests.

Funding. The authors acknowledge financial support from the BIOSTASES Advanced Grant, funded by the European Research Council under the European Union's Horizon 2020 Research and Innovation Programme (grant agreement no. 666971). The authors were also supported by the TULIP Laboratory of Excellence (ANR-10-LABX-41). Acknowledgements. We thank Bart Haegeman for his guidance in solving the analytical solution and stimulating discussions.

References

- Elton CS. 1927 *Animal ecology*. London, UK: Sidgwick and Jackson.
- Lindeman RL. 1942 The trophic-dynamic aspect of ecology. *Ecology* **23**, 399–418. (doi:10.2307/1930126)
- May RM. 1973 *Stability and complexity in model ecosystems*. Princeton, NJ: Princeton University Press.
- Pimm SL. 1982 *Food webs*. London, UK: Chapman and Hall.
- Loreau M. 2010 *From populations to ecosystems: theoretical foundations for a new ecological synthesis*. Princeton, NJ: Princeton University Press.
- McCann KS. 2012 *Food webs*. Princeton, NJ: Princeton University Press.
- Paine RT. 1966 Food web complexity and species diversity. *Am. Nat.* **100**, 65–75. (doi:10.1086/282400)
- Polis GA, Strong DR. 1996 Food web complexity and community dynamics. *Am. Nat.* **147**, 813–846. (doi:10.1086/285880)
- Oksanen L. 1991 Trophic levels and trophic dynamics: a consensus emerging? *Trends Ecol. Evol.* **6**, 58–60. (doi:10.1016/0169-5347(91)90124-G)
- Allesina S, Alonso D, Pascual M. 2008 A general model for food web structure. *Science* **320**, 658–661. (doi:10.1126/science.1156269)
- Williams RJ, Martinez ND. 2000 Simple rules yield complex food webs. *Nature* **404**, 180–183. (doi:10.1038/35004572)
- May RM. 1972 Will a large complex system be stable? *Nature* **238**, 413–414. (doi:10.1038/238413a0)
- May RM. 1986 The search for patterns in the balance of nature: advances and retreats. *Ecology* **67**, 1115–1126. (doi:10.2307/1938668)
- Pimm SL, Lawton JH, Cohen JE. 1991 Food web patterns and their consequences. *Nature* **350**, 669–674. (doi:10.1038/350669a0)
- Solow AR, Costello C, Beet A. 1999 On an early result on stability and complexity. *Am. Nat.* **154**, 587–588. (doi:10.1086/303265)
- DeAngelis DL. 1992 *Dynamics of nutrient cycling and food webs*. London, UK: Chapman and Hall.
- Paine RT. 1980 Food webs: linkage, interaction strength and community infrastructure. *J. Anim. Ecol.* **49**, 667–685. (doi:10.2307/4220)
- DeAngelis DL, Mulholland PJ, Palumbo AV, Steinman AD, Huston MA, Elwood JW. 1989 Nutrient dynamics and food-web stability. *Annu. Rev. Ecol. Syst.* **20**, 71–95. (doi:10.1146/annurev.es.20.110189.000443)
- Morin PJ, Lawler SP. 1995 Food web architecture and population dynamics: theory and empirical evidence. *Annu. Rev. Ecol. Syst.* **26**, 505–529. (doi:10.1146/annurev.es.26.110195.002445)
- Amarasekare P. 2008 Spatial dynamics in foodwebs. *Annu. Rev. Ecol. Syst.* **39**, 479–500. (doi:10.1146/annurev.ecolsys.39.110707.173434)
- Hall SR. 2009 Stoichiometrically explicit food webs: feedbacks between resource supply, elemental constraints, and species diversity. *Annu. Rev. Ecol. Syst.* **40**, 503–528. (doi:10.1146/annurev.ecolsys.39.110707.173518)
- Brose U *et al.* 2017 Predicting the consequences of species loss using size-structured biodiversity approaches. *Biol. Rev.* **92**, 684–697. (doi:10.1111/brv.12250)
- Carpenter SR, Kitchell JF, Hodgson JR. 1985 Cascading trophic interactions and lake productivity. *Bioscience* **35**, 634–639. (doi:10.2307/1309989)
- Oksanen L, Fretwell SD, Arruda J, Niemela P. 1981 Exploitation ecosystems in gradients of primary productivity. *Am. Nat.* **118**, 240–261. (doi:10.1086/283817)
- Hairston NG, Smith FE, Slobodkin LB. 1960 Community structure, population control, and competition. *Am. Nat.* **94**, 421–425. (doi:10.1086/282146)
- Fretwell SD. 1977 The regulation of plant communities by the food chains exploiting them. *Perspect. Biol. Med.* **20**, 169–185. (doi:10.1353/pbm.1977.0087)
- Hairston NGJ, Hairston NGS. 1993 Cause-effect relationships in energy flow, trophic structure, and interspecific interactions. *Am. Nat.* **142**, 379–411. (doi:10.1086/285546)
- Jager HI, Gardner RH. 1988 A simulation experiment to investigate food web polarization. *Ecol. Modell.* **41**, 101–116. (doi:10.1016/0304-3800(88)90048-8)
- Schmitz OJ. 1992 Exploitation in model food chains with mechanistic consumer-resource dynamics. *Theor. Popul. Biol.* **41**, 161–183. (doi:10.1016/0040-5809(92)90042-R)
- Abrams PA. 1993 Effect of increased productivity on the abundances of trophic levels. *Am. Nat.* **141**, 351–371. (doi:10.1086/285478)
- Shurin JB, Borer ET, Seabloom EW, Anderson K, Blanchette CA, Broitman B, Cooper SD, Halpern BS. 2002 A cross-ecosystem comparison of the strength of trophic cascades. *Ecol. Lett.* **5**, 785–791. (doi:10.1046/j.1461-0248.2002.00381.x)
- Schmitz OJ, Hambäck PA, Beckerman AP. 2000 Trophic cascades in terrestrial systems: a review of the effects of carnivore removals on plants. *Am. Nat.* **155**, 141–153. (doi:10.1086/303311)
- Halaj J, Wise DH. 2001 Terrestrial trophic cascades: how much do they trickle? *Am. Nat.* **157**, 262–281. (doi:10.1086/319190)
- Pimm SL. 1979 The structure of food webs. *Theor. Popul. Biol.* **16**, 144–158. (doi:10.1016/0040-5809(79)90010-8)
- Rosenzweig ML. 1973 Exploitation in three trophic levels. *Am. Nat.* **107**, 275–294. (doi:10.1086/282830)
- Hastings A, Powell T. 1991 Chaos in a three-species food chain. *Ecology* **72**, 896–903. (doi:10.2307/1940591)
- Post DM, Conners ME, Goldberg DS. 2000 Prey preference by a top predator and the stability of linked food chains. *Ecology* **81**, 8–14. (doi:10.1890/0012-9658(2000)081[0008:PPBATP]2.0.CO;2)
- Pimm SL. 1984 The complexity and stability of ecosystems. *Nature* **307**, 321–326. (doi:10.1038/307321a0)
- Donohue I *et al.* 2016 Navigating the complexity of ecological stability. *Ecol. Lett.* **19**, 1172–1185. (doi:10.1111/ele.12648)
- Harrison GW. 1979 Stability under environmental stress: resistance, resilience, persistence, and variability. *Am. Nat.* **113**, 659–669. (doi:10.1086/283424)
- Arnoldi J-F, Loreau M, Haegeman B. 2016 Resilience, reactivity and variability: a mathematical comparison of ecological stability measures. *J. Theor. Biol.* **389**, 47–59. (doi:10.1016/j.jtbi.2015.10.012)
- Dunne JA, Williams RJ, Martinez ND. 2001 Network structure and biodiversity loss in food webs: robustness increases with connectance. *Ecol. Lett.* **5**, 558–567. (doi:10.1046/j.1461-0248.2002.00354.x)
- Solé RV, Montoya JM. 2001 Complexity and fragility in ecological networks. *Proc. Biol. Sci.* **268**, 2039–2045. (doi:10.1098/rspb.2001.1767)
- Wang S, Loreau M. 2014 Ecosystem stability in space: alpha, beta and gamma variability. *Ecol. Lett.* **17**, 891–901. (doi:10.1111/ele.12292)
- Wang S, Loreau M. 2016 Biodiversity and ecosystem stability across scales in metacommunities. *Ecol. Lett.* **19**, 510–518. (doi:10.1111/ele.12582)
- Wang S, Loreau M, Arnoldi J-F, Fang J, Rahman KA, Tao S, de Mazancourt C. 2017 An invariability-area relationship sheds new light on the spatial scaling of ecological stability. *Nat. Commun.* **8**, 15211. (doi:10.1038/ncomms15211)
- Loreau M, Holt RD. 2004 Spatial flows and the regulation of ecosystems. *Am. Nat.* **163**, 606–615. (doi:10.1086/382600)
- Schaffer WM, Ellner S, Kot M. 1986 Effects of noise on some dynamical models in ecology.

- J. Math. Biol.* **24**, 479–523. (doi:10.1007/BF00275681)
49. van Kampen NG. 1992 *Stochastic processes in physics and chemistry*. Amsterdam, The Netherlands: Elsevier BV.
50. Nisbet RM, Gurney WSC. 1982 *Modelling fluctuating populations*. Chichester, UK: John Wiley & Sons.
51. Wang S, Haegeman B, Loreau M. 2015 Dispersal and metapopulation stability. *PeerJ.* **3**, e1295. (doi:10.7717/peerj.1295)
52. Rosenzweig ML. 1971 Paradox of enrichment: destabilization of exploitation ecosystems in ecological time. *Science* **171**, 385–387. (doi:10.1126/science.171.3969.385)
53. Pace ML, Cole JJ, Carpenter SR, Kitchell JF. 1999 Trophic cascades revealed in diverse ecosystems. *Trends Ecol. Evol.* **14**, 483–488. (doi:10.1016/S0169-5347(99)01723-1)
54. Krause AE, Frank KA, Mason DM, Ulanowicz RE, Taylor WW. 2003 Compartments revealed in food-web structure. *Nature* **426**, 282–285. (doi:10.1038/nature02115)
55. Stouffer DB, Bascompte J. 2011 Compartmentalization increases food-web persistence. *Proc. Natl Acad. Sci. USA* **108**, 3648–3652. (doi:10.1073/pnas.1014353108)
56. O’Gorman EJ, Emmerson MC. 2009 Perturbations to trophic interactions and the stability of complex food webs. *Proc. Natl Acad. Sci. USA* **106**, 13 393–13 398. (doi:10.1073/pnas.0903682106)
57. Abrams PA, Roth JD. 1994 The effects of enrichment of three-species food chains with nonlinear functional responses. *Ecology* **75**, 1118–1130. (doi:10.2307/1939435)
58. Diehl S, Feiße M. 2000 Effects of enrichment on three-level food chains with omnivory. *Am. Nat.* **155**, 200–218.
59. McCann K, Hastings A. 1997 Re-evaluating the omnivory-stability relationship in food webs. *Proc. R. Soc. Lond. B* **264**, 1249–1254. (doi:10.1098/rspb.1997.0172)
60. Pinckney JL, Paerl HW, Tester P, Richardson TL. 2001 The role of nutrient loading and eutrophication in estuarine ecology. *Environ. Health Perspect.* **109**, 699–706.
61. Nixon SW. 1995 Coastal marine eutrophication: a definition, social causes, and future concerns. *Ophelia* **41**, 199–219. (doi:10.1080/00785236.1995.10422044)
62. Chapin III FS, Matson PA, Vitousek PM. 2012 *Principles of terrestrial ecosystem ecology*, 2nd edn. New York: NY: Springer.
63. Carpenter SR, Brock WA. 2006 Rising variance: a leading indicator of ecological transition. *Ecol. Lett.* **9**, 311–318. (doi:10.1111/j.1461-0248.2005.00877.x)
64. Gause GF, Wit AA. 1935 Behavior of mixed populations and the problem of natural selection. *Am. Nat.* **64**, 596–609. (doi:10.1086/280628)
65. May RM. 1981 *Models for two interacting populations. Theoretical ecology: principles and applications*. Sunderland, MA: Sinauer.
66. Pimm SL, Lawton JH. 1978 On feeding on more than one trophic level. *Nature* **275**, 542–544. (doi:10.1038/275542a0)
67. Holt RD, Polis GA. 1997 A theoretical framework for intraguild predation. *Am. Nat.* **149**, 745–764. (doi:10.1086/286018)
68. Lafferty KD, Dobson AP, Kuris AM. 2006 Parasites dominate food web links. *Proc. Natl Acad. Sci. USA* **103**, 11 211–11 216. (doi:10.1073/pnas.0604755103)
69. Neutel A, Heesterbeek JAP, de Ruiter PD. 2002 Stability in real food webs: weak links in long loops. *Science* **296**, 1120–1123. (doi:10.1126/science.1068326)
70. Grilli J, Rogers T, Allesina S. 2016 Modularity and stability in ecological communities. *Nat. Commun.* **7**, 12031. (doi:10.1038/ncomms12031)
71. Polis GA, Anderson WB, Holt RD. 1997 Toward an integration of landscape and food web ecology: the dynamics of spatially subsidized food webs. *Annu. Rev. Ecol. Syst.* **28**, 289–316. (doi:10.1146/annurev.ecolsys.28.1.289)
72. Lande R. 1993 Risks of population extinction from demographic and environmental stochasticity and random catastrophes. *Am. Nat.* **142**, 911–927. (doi:10.1086/285580)
73. Bellard C, Bertelsmeier C, Leadley P, Thuiller W, Courchamp F. 2012 Impacts of climate change on the future of biodiversity. *Ecol. Lett.* **15**, 365–377. (doi:10.1111/j.1461-0248.2011.01736.x)
74. Gilman SE, Urban MC, Tewksbury J, Gilchrist GW, Holt RD. 2010 A framework for community interactions under climate change. *Trends Ecol. Evol.* **25**, 325–331. (doi:10.1016/j.tree.2010.03.002)
75. Perrings C. 2014 *Our uncommon heritage: biodiversity, ecosystem services and human wellbeing*. Cambridge, UK: Cambridge University Press.
76. Barbier M, Loreau M. 2018 Pyramids and cascades: a synthesis of food chain functioning and stability. *bioRxiv* 361246. (doi:10.1101/361246)
77. Purvis A, Gittleman JL, Cowlishaw G, Mace GM. 2000 Predicting extinction risk in declining species. *Proc. R. Soc. Lond. B* **267**, 1947–1952. (doi:10.1098/rspb.2000.1234)
78. Cardillo M, Mace GM, Jones KE, Bielby J, Bininda-Emonds ORP, Sechrest W, Orme CD, Purvis A. 2005 Multiple causes of high extinction risk in large mammal species. *Science* **309**, 1239–1241. (doi:10.1126/science.1116030)
79. Gonzalez A, Charetton EJ. 2002 Heterotroph species extinction, abundance and biomass dynamics in an experimentally fragmented microecosystem. *J. Anim. Ecol.* **71**, 594–602. (doi:10.1046/j.1365-2656.2002.00625.x)
80. Thompson PL, Shurin JB. 2012 Regional zooplankton biodiversity provides limited buffering of pond ecosystems against climate change. *J. Anim. Ecol.* **81**, 251–259. (doi:10.1111/j.1365-2656.2011.01908.x)
81. Donohue I *et al.* 2013 On the dimensionality of ecological stability. *Ecol. Lett.* **16**, 421–429. (doi:10.1111/ele.12086)