

Characterization of the Active Bacterial Community Involved in Natural Attenuation Processes in Arsenic-Rich Creek Sediments

Odile Bruneel, Aurelie Volant, Sébastien Gallien, Bertrand Chaumande, Corinne Casiot, Christine Carapito, Amélie Bardil, Guillaume Morin, Gordon D Brown, Christian Personné, et al.

To cite this version:

Odile Bruneel, Aurelie Volant, Sébastien Gallien, Bertrand Chaumande, Corinne Casiot, et al.. Characterization of the Active Bacterial Community Involved in Natural Attenuation Processes in Arsenic-Rich Creek Sediments. Microbial ecology, 2011, 61 (4), pp.793-810. 10.1007/s00248-011-9808-9. hal-02341109

HAL Id: hal-02341109 <https://hal.science/hal-02341109v1>

Submitted on 9 Feb 2022

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

Characterization of the Active Bacterial Community Involved in Natural Attenuation Processes in Arsenic-Rich Creek Sediments

Odile Bruneel • Aurélie Volant • Sébastien Gallien • Bertrand Chaumande • Corinne Casiot • Christine Carapito • Amélie Bardil • Guillaume Morin • Gordon E. Brown Jr · Christian J. Personné · Denis Le Paslier · Christine Schaeffer · Alain Van Dorsselaer • Philippe N. Bertin • Françoise Elbaz-Poulichet • Florence Arsène-Ploetze

O. Bruneel $(\boxtimes) \cdot$ A. Volant \cdot C. Casiot \cdot A. Bardil \cdot C. J. Personné : F. Elbaz-Poulichet Laboratoire HydroSciences Montpellier, UMR5569 (CNRS-IRD-Universités Montpellier I et II), Université Montpellier II, CC MSE, Place Eugène Bataillon, 34095 Montpellier Cedex 05, France e-mail: bruneel@msem.univ-montp2.fr

B. Chaumande : P. N. Bertin : F. Arsène-Ploetze Génétique Moléculaire, Génomique Microbiologie, UMR7156, Université de Strasbourg/CNRS, 28 rue Goethe, 67083 Strasbourg Cedex, France

S. Gallien : C. Carapito : C. Schaeffer : A. Van Dorsselaer Laboratoire de Spectrométrie de Masse Bio-organique, Institut Pluridisciplinaire Hubert Curien, UMR7178 (CNRS-Université de Strasbourg), 25 rue Becquerel, 67087 Strasbourg, France

G. Morin Institut de Minéralogie et de Physique des Milieux Condensés (IMPMC), UMR7590 (CNRS - Universités Paris 6&7 - IPGP), 140, rue de Lourmel, 75015 Paris, France

G. E. Brown Jr Surface and Aqueous Geochemistry Group, Department of Geological and Environmental Sciences, Stanford University, Stanford, CA 94305-2115, USA

G. E. Brown Jr Stanford Synchrotron Radiation Laboratory, SLAC, 2575 23 Sand Hill Road, MS 69, Menlo Park, CA 94025, USA

D. Le Paslier Génomique Métabolique, UMR8030, CNRS, 2 rue Gaston Crémieux, 91057 Evry Cedex, France

D. Le Paslier Commissariat à l'Energie Atomique (CEA), Direction des Sciences du Vivant, Institut de Génomique, Genoscope, Laboratoire de Génomique Comparative, 2 rue Gaston Crémieux, 91057 Evry Cedex, France

Abstract Acid mine drainage of the Carnoulès mine (France) is characterized by acid waters containing high concentrations of arsenic and iron. In the first 30 m along the Reigous, a small creek draining the site, more than 38% of the dissolved arsenic was removed by co-precipitation with Fe(III), in agreement with previous studies, which suggest a role of microbial activities in the co-precipitation of As(III) and As(V) with Fe(III) and sulfate. To investigate how this particular ecosystem functions, the bacterial community was characterized in water and sediments by

16S rRNA encoding gene library analysis. Based on the results obtained using a metaproteomic approach on sediments combined with high-sensitivity HPLC-chip spectrometry, several GroEL orthologs expressed by the community were characterized, and the active members of the prokaryotic community inhabiting the creek sediments were identified. Many of these bacteria are β-proteobacteria such as Gallionella and Thiomonas, but γ-proteobacteria such as Acidithiobacillus ferrooxidans and α -proteobacteria such as Acidiphilium, Actinobacteria, and Firmicutes were also detected.

Introduction

Acid Mine Drainage (AMD) is one of the most serious forms of water pollution in industrial and post-industrial areas worldwide [38]. AMD is generated when the wastes from the mining and processing of sulfide ores (such as pyrite or arsenopyrite) come into contact with oxygenated water [5]. AMD is often characterized by pH values of 2–4. Such waters generally contain high levels of iron, toxic metals (such as aluminum, manganese, lead, cadmium, and zinc), and metalloids (arsenic) [5, 32, 48]. AMD can still occur hundreds of years after mine closure and tens of thousands of kilometers of groundwater, streams, lakes, and estuaries throughout the world have been directly impacted [40]. In several cases of AMD, natural remediation has been observed, as for example at the Carnoulès site in France and the Rio Tinto site in Spain [18, 53]. In such AMD, toxic compounds are accumulated in sediments consisting of a variety of iron (oxyhydr)oxides and hydroxysulfates such as jarosite, schwertmannite, and ferrihydrite [48]. Natural remediation of metal pollutants is generally due to the occurrence of abiotic reactions and/or microbial activities that make these toxic compounds insoluble and lead them to accumulate in sediments [32, 40]. This toxic compound precipitation processes mainly involve the oxidation and precipitation of iron, which is often the main soluble metal present in AMD, and the adsorption of other metals and metalloids by the ferric minerals formed [32, 51]. Indeed, many elements such as Sr, Cs, Pb, U, and As show a strong affinity for solid iron oxide [18, 27, 48]. Abiotic oxidation of Fe(II) proceeds very slowly in acidic (pH 3.5) waters [51]. In contrast, iron-oxidizing bacteria catalyze the reaction and thus accelerate the formation of solid iron oxide [32, 41, 51]. In addition, several bacteria contribute to the immobilization of arsenic via their ability to oxidize this metalloid [6, 14, 18, 22, 48], arsenate $(As(V))$ being adsorbed more strongly than arsenite (As(III)) by Fe(III) oxides and hydroxides at acidic pH levels [10]. Thiomonas strains show a high arsenite oxidation capacity, and these metabolic activities have been extensively analyzed under laboratory conditions [6, 14, 22].

However, to be able to develop remediation processes and/or optimize existing processes, further knowledge is required about how these bacteria function in situ. In particular, it is of crucial importance to determine which bacteria are viable and active in such ecosystems.

The AMD of Carnoulès mine in Southern France is a highly suitable site for analyzing how microorganisms contribute to the transformation of metals and metalloids in situ, since efficient natural remediation processes are known to occur at this site [12, 18]. This former mine generated around 1.2 Mt of tailings containing 0.7% Pb, 10% FeS₂, and 0.2% As. Water percolating through the tailings forms the head of the Reigous creek. This creek is acidic (pH around 3) and highly contaminated with As (100 to 350 mg L^{-1}). The behavior of As and Fe in the Reigous creek has been intensively studied [18, 23, 48]. In the creek spring, As(III) is the main As species present and Fe occurs in the form of Fe(II) [18]. Along the first 30 m of the creek (about 1 h residence time), the oxidation of Fe(II) leads to the co-precipitation of more than 38% of the dissolved As [18, 23]. Arsenic accounts for up to 22% of the total dry weight of the sediments formed along the first 10 m along the creek. In the wet season, approximately 30 m downstream of the spring, these sediments are mainly composed of As(III)–Fe(III) oxyhydroxysulfates, whereas As(V)–Fe (III) oxyhydroxysulfates compounds predominate during the dry season [48]. Because of the very high molar As/Fe ratio (up to 0.3) existing in the dissolved phase of the Carnoulès creek, the mineralogical content of the sediments differs significantly from that classically observed at most AMD [47], especially along the first 50 m of the creek. Several strains of Thiomonas and Acidithiobacillus ferrooxidans have been isolated from Reigous creek waters, and based on the results of laboratory experiments, it has been suggested that these bacteria may contribute to the oxidation of Fe(II) and the co-precipitation of As [14, 18, 21–23, 48]. Preliminary analyses have shown that the DNAs of both bacteria are present in the Reigous creek, as well as that of Gallionella sp., Thiobacillus sp., and some sulfate-reducing bacteria [12]. Archaea have also been found to occur in the Reigous creek (Ferroplasma acidiphilum and sequences affiliated to uncultured Thermoplasmatales archaeon) as well as a eukaryotic microorganism, Euglena mutabilis [12, 13, 15]. However, the bacterial population inhabiting the As-rich Reigous sediments has never been characterized so far. It, therefore, seemed to be necessary not only to identify the bacteria present in this creek but also to determine which members of this community are viable and, therefore, perform metabolic activities in situ.

The aim of this study was to describe the bacterial populations occurring in both the sediments and waters at the disused Carnoulès site and to identify the bacteria at work. For this purpose, three complementary approaches

were used. First, chemical and mineralogical studies were performed in order to determine the arsenic species present. A 16S rRNA encoding gene library was then analyzed in order to identify the bacterial population present in the creek sediments and waters. Lastly, based on the findings obtained using a metaproteomic approach combined with high-sensitivity mass spectrometry methods, the active species inhabiting the sediments were identified.

Methods

Sampling and Analysis

Samples were collected from Reigous creek in April 2006 at COWG station located 30 m downstream from the spring. This sampling was part of a long-term monitoring of the physicochemistry of the Reigous Creek water [23]. The main physicochemical parameters (pH, temperature, and dissolved oxygen concentrations) were measured in situ at this sampling point. The 5-cm deep sediments on the bottom of the creek and a thin column (less than 10 cm) of running water covering the sediments were sampled. Solid samples were removed with a sterile spatula from the surface of the sediments. Water samples (300 ml) were immediately filtered through 0.22 μm Millipore membranes fitted on Sartorius polycarbonate filter holders (for water chemical analysis) or through sterile 0.22-μm Nucleopore filters that were then transferred to a collection tube (Nunc), frozen in liquid nitrogen, and stored at −80°C until DNA extraction (for 16S rRNA encoding gene analysis). Sampling was repeated three times. For total Fe and As determination, filtered water was acidified to pH=1 with $HNO₃$ (14.5 M) and stored at 4 $\rm{°C}$ in polyethylene bottles until analysis. For As and Fe speciation, a 10 μl aliquot of filtered sample water was added to either 0.5 ml of 5% (v/v) 0.25 M EDTA solution for As speciation [7] or a mixture of 0.5 ml acetate buffer (pH 4.5) and 1 ml of 1,10 phenanthrolinium chloride solution for Fe speciation [50]. The vials were completed to 10 ml with deionized water. The samples used for arsenic speciation and Fe(II) and sulfate determination were stored in the dark and analyzed within 24 h.

Chemical Analysis

The determination of total dissolved As was performed by ICP-MS using Thermo X7 series with a conventional external calibration procedure. Indium was used as internal standard to correct for instrumental drift and possible matrix effects. It was not necessary to correct interference with chloride because of the extremely high As levels present. Certified reference material SLRS-4 (freshwater samples) was used to check analytical accuracy and precision. The results showed that the recovery rate obtained was within ±5%.

Analyses of inorganic arsenic species (As(III), As(V)) were carried out using anion-exchange chromatography (25 cm×4.1 mm i.d. Hamilton PRP-X100 column with Varian ProStar gradient solvent delivery system) coupled to a hydride generation (VGS 200, FISONS, France) with an atomic fluorescence spectrometry detector (Excalibur, PSAnalytical, GB) [17]. The detection limit obtained was 172 ng L⁻¹ for As(III) and 458 ng L⁻¹ for As(V), with a precision better than 5%. Total dissolved Fe was determined by flame atomic absorption spectrometry. Fe(II) concentration was determined using colorimetry at 510 nm after complexation with 1,10-phenanthrolinium chloride solution in buffered samples (pH 4.5) [50] (detection limit: 11 μg L^{-1} ; precision better than 5%). Sulfate concentration was determined after precipitation of $BaSO₄$ with $BaCl₂$ and spectrophotometric measurement at 650 nm [50].

Solid Sample Characterization

XAFS data were gathered on the laboratory samples and the sample taken at COWG on April 2006, at 10 K in transmission mode on a bending magnet D44 at the LURE synchrotron (Orsay, France), and in fluorescence mode on the 11-2 wiggler beamline at SSRL (Stanford, CA), respectively. Experiments and data reduction were previously reported [48, 49].

DNA Isolation, 16S rRNA Encoding Gene Cloning, Restriction Analysis, and Sequencing

Genomic DNA was extracted in triplicate from filtered water and sediments using the UltraClean Soil DNA Isolation Kit according to the manufacturer's recommendations (MoBio Laboratories Inc., USA). These triplicates were pooled before PCR amplification. All the genomic DNA samples extracted were stored at −20°C until further processing. Bacterial diversity was analyzed by cloning PCR amplified 16S rRNA encoding genes. Bacterial 16S rRNA encoding genes were amplified with 8F (5′-AGAGTTTGA TCCTGGCTCAG-3′) and 1489R primers (5′-TACCTTGT TACGACTTCA-3′) [43, 57], as previously described [12]. These PCR products were cloned into Escherichia coli TOP 10 strain using the pCR2.1 Topo TA cloning kit (Invitrogen, Inc.). Cloned 16S rRNA encoding gene fragments were reamplified using the primers TOP1 (5′-GTGTGCT GGAATTCGCCCTT-3′) and TOP2 (5′-TATCTGCAGA ATTCGCCCTT-3′) that anneal to the vector and surround the inserted PCR fragment and then digested with HaeIII or HinfI enzymes. Restriction profiles were analyzed using 2.5% agarose gel electrophoresis (small fragment resolution

agarose; QA agarose, QBiogène, Inc.). Around 200 clones from each library were analyzed and grouped according to the RFLP patterns obtained. PAST (Paleontological STatistics v 1.19) software from the website http://folk. uio.no/ohammer/past/ was used to calculate different diversity indices (rarefaction analysis, taxa, total clones, singletons, dominance, coverage, shannon, equitability, and simpson, Table 1) for each clone library. The total number of clones obtained compared with the number of clones representing each unique phylotype was used to produce the rarefaction curves. Coverage values were calculated to determine how efficiently the libraries described the complexity of a theoretical community such as an original bacterial community. The coverage [29] value is given as $C=$ $1-(n_1/N)$ where n_1 is the number of clones that occurred only once in the library. Rarefaction analysis showed that the curves generated were near saturation (data not shown) and consistent with the high coverage values of the two clone libraries (97.8 for the sediment and 98.6 for the water). This indicated that the clone libraries were sufficiently sampled.

Partial sequences of the clones from dominant groups were determined by the dideoxy nucleotide chaintermination method using the BigDye 3.1 kit (Applied Biosystems) on an ABI PRISM 3730XL Genetic analyzer (Applied Biosystems). The PINTAIL program [4] was used to check the presence of chimerae. Sequences were also examined manually for chimerae, which were excluded from further analyses. These sequences were compared with known sequences (NCBI database) using the BLAST, ALIGN, and CLUSTALW programs [1, 26, 55]. All sequence data obtained were submitted to the EMBL databases under accession numbers (FR676963-FR677013; AM988784-AM988794; AM988796; AM988798; AM988801-AM988805; AM988807-AM988809).

Preparation of Protein Extracts and Gel Electrophoresis Analysis

Prokaryotes were separated from sediments and eukaryotes using a Nycodenz density gradient. It should be noted that the main population determined from the DNA directly extracted from sediments was similar to that identified after Nycodenz treatment (data not shown), which suggests that this treatment did not result in enrichment of the sample by any particular microorganisms. Ten grams of sediments

were washed in 10 mL of saline buffer and agitated overnight at 4°C. After 10 min of decantation, 7.5 mL of supernatant were added without mixing to 17.5 mL of Nycodenz solution (Axis-Shield, Dundee, Scotland), and then centrifuged for 30 min at $10,000 \times g$. The cellular fraction (nycodenz/sample interface) was removed and washed by adding two volumes of NaCl 0.9% and centrifuged for 15 min at $10,000 \times g$ at 4°C. Proteins were extracted from this cellular fraction as previously described [58], further purified using the 2-D Clean-up kit (GE Healthcare), and resuspended in rehydration buffer (364 g L⁻¹ thiourea, 1,000 g L⁻¹ urea; 25 g L⁻¹ CHAPS, 0.6% v/v IPG buffer Pharmalyte, 10 g L⁻¹ of DTT, and 0.01% bromophenol blue). Protein concentrations were quantified using the 2-D Quant kit (GE Healthcare). These proteins were separated by 2-D gel electrophoresis as previously described [58] and finally stained with silver nitrate. Gels were analyzed using an Image Scanner, LabScan v 3.0 (GE Healthcare), and the ImageMaster 2D platinum software program (v. 6.01, GE Healthcare). The spots selected were cut out of the 2-D gels and stored at −20°C. Eighty-one spots were analyzed by performing nanoLC-Chip-MS/MS.

In-Gel Digestion, Mass Spectrometry Analysis, and Protein Identification

Unless otherwise specified, all chemicals were obtained from Sigma (St. Louis, MO, USA). In-gel digestion of gel spots was performed as previously described [58]. The resulting peptides were analyzed by performing nanoLC-MS/MS on an Agilent 1100 Series HPLC-Chip/MS system (Agilent Technologies, Palo Alto, USA) coupled to an HCT Ultra ion trap (Bruker Daltonics, Bremen, Germany). The MS/MS data were analyzed using the MASCOT 2.2.0 algorithm giving a maximum of one missed cleavage, with a mass tolerance of 0.5 Da for MS and MS/MS data and carbamidomethylation of cysteines and oxidation of methionines were specified as the variable modifications. MS/MS data searches were performed against two in-house generated databases. The first database was composed of the protein sequences of all the organisms related to the groups identified by performing 16S rRNA encoding gene analysis on the Reigous creek sediments and water (α -, β -, δ -, and γ-proteobacteria, Bacilli, Clostridia, Actinobacteria,

Table 1 Diversity indices calculated from the two clone libraries in sediment and water at the Carnoulès mine drainage creek

Clone library	Taxa	Total no. of clones	$Singletons$ Dominance (D) Coverage (C)		Shannon (H)	Equitability	Simpson $(1-D)$
Sediments		229	0.5503	97.8	1.146	0.4231	0.4497
Water	18	221	0.2551	98.6	.949	0.6744	0.7449

Nitrospira), as well as unclassified bacteria from http://beta. uniprot.org/, Thiomonas sp. from http://www.genoscope.cns. fr/ (FP475956–FP475957), Euglenozoa, and Viridiplantae. The second database included all bacterial and archaeal GroEL chaperonins (12501 and 291 sequences, respectively) extracted from the Uniprot database (http://www.uniprot.org/ uniprot).

To assess the false positive rate in the protein identification, a target-decoy database search was performed [25]. With this approach, peptides are matched against a database consisting of the native protein sequences detected in the database (target) and the sequence-reversed entries (decoy). Protein identification was confirmed when at least two peptides with a minimum Mascot ion score of 30 were detected. In the case of one-peptide hits, the score of the unique peptide had to be greater than the 95% significance Mascot threshold level. All the proteins identified were added to the "InPact" proteomic database developed at our laboratory (http://inpact.u-strasbg.fr/~db/) [8].

Phylogenetic Analyses

A search for GroEL homologs and 16S rRNA encoding sequences was carried out in the Uniprot and NCBI databases, respectively. A total number of 530 reviewed GroEL bacterial sequences were retrieved from the RefSeq database. These sequences were 500–550 amino acids in length. Only one sequence representative of each genus (259 sequences in all) was kept. GroEL and 16S rRNA encoding sequences were aligned using ClustalW [55]. Alignments were checked by hand and positions with more than 1% of gaps were automatically removed. Neighborjoining trees were constructed with 185 amino acids in the case of the GroEL sequences and with 310 nt in that of the 16S rRNA encoding sequences. Trees were drawn up using the iTOL website (http://itol.embl.de/) [44].

Results and Discussion

Physical and Chemical Characteristics of Samples

The physicochemistry of the Reigous Creek water at the time of sampling was typical of that revealed during the long-term monitoring study [23]. The water sample was acid (pH=3.28) and moderately oxygenated (dissolved oxygen concentration=3.5±0.5 mg L^{-1}); it contained extremely high concentrations of SO_4^2 ⁻ (2700 ± 300 mg L⁻¹), Fe (620±30 mg L⁻¹), and As (140± 4 mg L⁻¹), with a large predominance of Fe(II) (90±10% of total Fe concentration) and equal proportion of As(III) and As(V). The removal of As during the course of the Reigous Creek from its source to the sampling station COWG reached 38%, corroborating the typical removal rates measured during the long-term monitoring study [23]. The pale-yellow loosely packed sediments previously observed during the dry season at the sampling point chosen in this study (COWG $~30$ m downstream of the spring), consisted of an amorphous $Fe(III)$ –As(V) hydroxysulfate mineral with an As/Fe molar ratio of 0.5 to 0.6 [48]. However, various other biominerals may be formed from dissolved Fe(II) and As(III) [47, 48]. The exact nature and structure of the sediment sample studied was, therefore, further investigated. X-ray powder diffraction (data not shown) and X-ray absorption spectroscopy data obtained at the As–K edge (Fig. 1) showed that arsenic was present in these samples in an amorphous Fe(III)–As (V) hydroxysulfate phase as previously observed [48]. These analyses showed that despite the presence of a minor As(III) impurity, the oxidized arsenic form As(V) predominates in this sediment (Fig. 1a). In a previous study, it has been shown that the catalytic oxidation of As(III) by Thiomonas sp. strains accelerates such As–Fe precipitation process [48]. Therefore, using extended X-ray absorption fine structure analysis at the As–K edge (Fig. 1b), the structure observed in our samples was compared with that of the minerals obtained after As(III) oxidation by the Thiomonas sp. strain B2 in bioassays in which sterilized Carnoulès Creek water was used [48]. These comparisons (Fig. 1b, c) showed that the molecular structure of the amorphous Fe(III)–As(V) hydroxysulfate phase observed in these sediments was similar to that of the $Fe(III)$ –As(V) hydroxysulfate obtained in the presence of Thiomonas, further supporting its role in situ.

Composition of Bacterial Communities in Reigous Creek Water and Sediments

Two 16S rRNA encoding gene libraries were constructed (Table 2), containing 229 clones in the sediment library and 221 in that of the water library. The Shannon index (see "Methods") and equitability values were greater in the water library than in the sediment library (Table 1), which suggests that the bacterial diversity was lower in the sediment than in the water samples. Eleven different species were identified in the sediments and 13 in the water (Table 2, Fig. 2).

Several of the bacteria identified in the present study in both the waters and the sediments have been previously detected in the Reigous waters [12, 14, 15, 18, 21, 22]. Classified first by abundance order, several sequences were affiliated to Thiobacillus sp. ML2-16. This bacterium has been frequently reported to occur in AMD [5]. The presence of several strains affiliated to this proteobacteria is in agreement with the results of previous studies and shows that these bacteria persist in this

Figure 1 X-ray absorption spectra at the As–K edge of the sample, (COWG April 11, 2006,) showing similarities with the X-ray amorphous Fe(III)–As(V) hydroxysulfate phases obtained after incubating sterilized Carnoulès water with the Thiomonas sp. strain B2 isolated at the site [48]. These phases reached a molar As/Fe ratio of 0.8, as described in [48]. a Linear least-squares fitting of XANES data showed that the largest arsenic fraction $(90\pm2\%)$ was in the As(V) oxidation state. A small arsenic fraction $(10\pm2\%)$ was in the As(III) oxidation state, which resulted in a slight decrease in the amplitude in the EXAFS spectrum. b Shell by shell fit of the EXAFS spectra in k-

ecosystem [12]. Secondly, bacteria affiliated to Gallionella capsiferriformans were detected in both the water and sediments. G. *capsiferriformans* is an oxygen-dependent ferrous iron-oxidizing bacterium that grows at circumneutral pH [59]. Relatives of Gallionella, such as G. ferruginea in particular, have often been detected, sometimes as the dominant group in microbial mine water communities, including Carnoulès [12, 33–35]. Bacteria related to At. ferrooxidans as well as Thiomonas strains have been previously isolated from this site [14, 18, 21, 22]. At. ferrooxidans, which was the first microorganism to be isolated from an acidic leaching environment, occurs ubiquitously in AMD, as does Thiomonas [32]. In addition to these groups, other species that have not previously been described at this site were detected in this study. Some of these species were found to occur in both sediments and water. For example, bacteria affiliated to the Firmicutes Alicyclobacillus sp. BRG 73 were identified. This genus, found in AMD [5], is characterized by moderately thermophilic, acidophilic, strictly aerobic, and endosporeforming bacilli [30]. Likewise, bacteria related to "Ferrovum myxofaciens" PSTR were detected in both the waters and sediments. "Ferrovum myxofaciens" is an autotrophic ironoxidizer which predominates in some AMD and is able to grow litho-autotrophically, using ferrous iron as an electron

space for the COWG sediment sample and the Thiomonas sp. strain B2 precipitate sample. c Corresponding Fourier transforms of the experimental and fit curves. Dotted lines experimental; solid lines fitting curves. The local structure of both the COWG and the laboratory Thiomonas sp. samples includes bidentate arsenate– oxygen–iron complexes characterized by ~1.5–2.0 Fe atoms at an As–Fe distance of 3.31 ± 0.02 Å. A small arsenic fraction $(10 \pm 2\%)$ was in the As(III) oxidation state, which resulted in a slight decrease in the amplitude in b the EXAFS spectrum and in c the corresponding Fourier transform

donor [32]. Lastly, bacteria affiliated to Leptospirillum ferrooxidans, an iron-oxidizing member of the Nitrospirae [36], were also present. This bacterium has been found to occur in several acidic environments and in biofilms originating from AMD [5, 9, 32].

Other newly characterized groups were identified only in sediments (Table 2). The presence in acidic mine waters of Acidocella sp., a non-iron-oxidizing heterotrophic acidophile is quite common in AMD [33, 35]. Ferrimicrobium is an iron-oxidizing heterotroph that can also use iron as an electron acceptor [20]. Other prokaryotes detected in sediments were affiliated to Acidiphilium sp. CCP3, a non-iron-oxidizing heterotrophic acidophile that is also quite common in AMD [33] and Dokdonella koreensis, a γ-proteobacteria.

Six newly characterized groups were identified only in water. Some of these bacteria were related to Sideroxydans lithotrophicus LD-1, an oxygen-dependent ferrous ironoxidizing bacterium that grows at circumneutral pH [59]; Rhodoferax ferrireducens, a psychrotolerant, facultative anaerobic bacterium which is able to oxidize acetate with the reduction of Fe(III) [28]; and an Acidobacteriaceae bacterium, CH1. Members of Acidobacteria have previously been reported in AMD [5, 32]. Finally, three sulfate-reducing bacteria were related to strain JHA1, Desulfomonile limimaris,

Sequences closely related to 16S rRNA genes from *Euglena* spp. chloroplast were also detected (data not shown). The 16S rRNA encoding gene of chloroplasts is closely related to the bacterial
16S rRNA encoding gene and can Sequences closely related to 16S rRNA genes from *Euglena* spp. chloroplast were also detected (data not shown). The 16S rRNA encoding gene of chloroplasts is closely related to the bacterial 16S rRNA encoding gene and can therefore be amplified by primers 8F and 1489R

oroun versus the total number of Table 2 Bacterial clones detected at the Carnoulès mine drainage with their phylogenetic group, the closest isolated relative and the relative abundance of each group versus the total number of Table 2. Bacterial clones detected at the Camoules mine drainage with their phylogenetic group the closest isolated relative and the relative ahundance of each

an anaerobic dehalogenating bacterium from marine sediments [54] and *Desulfuromonas svalbardensis* 60, a psychrophilic, Fe(III)-reducing bacterium isolated from Arctic sediments [56] (Table 2).

All in all, 17 species of bacteria were identified in the water and sediments sampled at the Reigous creek. Only seven genera were found to be present in both phases, six were found only in water, and four only in the sediments (Table 2). Most of these species are common residents of AMD [5, 32]. This quite low bacterial diversity was probably due to the high concentration of toxic compounds in this AMD and was consistent with previous observations showing that the biodiversity of acidic, metal-rich mine waters is mainly restricted to specialized prokaryotes and some eukaryotes such as Euglena [15, 52], which has been detected in this study (data not shown). In the Reigous system, in both sediments and water, the populations observed were mainly involved in the Fe, As, and S cycles. The populations involved in Fe(II) oxidation were related to Gallionella, At. ferrooxidans, Ferrimicrobium, Leptospirillum, Sideroxydans lithotrophicus, or "Ferrovum myxofaciens" [32, 59], whereas ferric iron reduction has been described for populations like Acidiphilium spp., Acidocella, Desulfuromonas svalbardensis, Rhodoferax ferrireducens, or even At. ferrooxidans and Ferrimicrobium acidiphilum [20, 28, 56]. Since the Carnoulès creek spring contains mainly Fe(II) and As(III) in the form of dissolved species [18], the Fe(III) and As(V) may be formed as the result of microbial oxidation processes via the activity of acidophilic iron- and arsenite-oxidizing bacteria [24, 48]. In other words, the large amounts of soft pale-yellow As(V)–Fe (III) hydroxysulfate sediments analyzed here (Fig. 1) were probably formed by the joint activities of iron-oxidizing (e.g., At. ferrooxidans or Gallionella) and arsenic-oxidizing (e.g., Thiomonas sp.) microorganisms. Concerning S cycling, we found populations able to oxidize the reduced inorganic sulfur compounds, like Thiobacillus, Thiomonas, or At. ferrooxidans [32, 42, 46]. Sulfate-reducing bacteria such as Desulfomonile limimaris or Desulfuromonas svalbardensis were also found to be present in sediments and may be involved in sulfate consumption [54, 56].

Among these prokaryotes, some bacteria may be present but not functionally active, and it was, therefore, crucial to differentiate between dead or inactive cells and functional cells. To determine which organisms play a significant role in the natural remediation processes such as the Fe(II) oxidation processes observed at the study site, a metaproteomic approach was used to list the bacterial population expressing proteins, i.e., those which were active. The metaproteomic approach was possible with sediments but failed with the water samples because larger numbers of bacterial cells were recovered from sediments than from water (data not shown).

Figure 2 Phylogenetic tree based on 16S rRNA-encoding sequences. Sequences were aligned using ClustalW. Alignments were checked by hand and positions with more than 1% of gaps were automatically removed. Neighbor-joining trees were drawn up with 310 nt using ITOL (http://itol.embl.de/) [44]. Accession numbers: see supplementary data. In red: bacteria identified based on the metaproteomic (GroEL identifications) approach; in blue: bacteria identified using the 16S rRNA encoding gene library; in black: sequences detected in NCBI databases which are closely related to the bacteria present in the Reigous sediment

Characterization of the Main Proteins Expressed by the Sediment Community

Proteins expressed by this community were liable to correspond to orthologs originating from diverse prokaryotes and to have similar amino acid sequences. For those reasons and to improve their characterization, proteins were separated by performing 2-D gel electrophoresis (supplementary Fig. 1). A total number of 89 proteins were identified, 44% of which (39 proteins, Table 3) originated from bacteria and 39% (35 proteins) from protists, while 15 proteins originated from higher plants, probably from decomposed plant debris present at the Reigous creek. One third of the proteins originated from protists, mainly consisting of E. mutabilis detected in our samples (data not shown) and could not be completely removed using the Nycodenz Gradient. E. mutabilis is a common inhabitant of AMD [2, 11, 15, 39] and was also present at the surface of sediments. These Euglena proteins are involved in various metabolic processes, suggesting that this eukaryote may play a relevant role in this ecosystem, in agreement with recent results (Gouhlen-Chollet and Bertin, unpublished data). Only proteins originating from bacteria were further analyzed in this study. The 2-D gel electrophoresis approach used here allowed identifying from a single sample only the most abundant cytoplasmic proteins. Therefore, some relevant proteins that might be expressed in this environment, such as rusticyanin from At. ferrooxidans, which is known to be involved in Fe(II) oxidation and was previously thought to possibly play a functional role at this site [21, 24, 48], were not detected in this study. This membrane protein may not be resolved in the 2-D gel, or it may not be abundant at the sampling point. In addition, the majority of the bacteria forming the Carnoulès community have never been grown and studied in vitro so far. Their genome sequences, and hence their protein sequences, which are required for MS identification purposes, may not be available in public databases, except for those of At. ferrooxidans, Thiomonas, and Gallionella strains. All these hypotheses may explain why so few bacterial proteins were identified in this study.

The bacterial proteins identified originated from Aquificae were Actinobacteria, Deinococcus, Synergistetes, Firmicutes, Bacteroidetes, and α -, β -, and γ -proteobacteria. Among the

Table 3 Bacterial proteins identified in the Reigous sediment's microbial community Table 3 Bacterial proteins identified in the Reigous sediment's microbial community

Table 3 (continued)

^aMS/MS data were searched against an in-house database including all the bacterial and archaeal GroEL sequences obtained from Uniprot MS/MS data were searched against an in-house database including all the bacterial and archaeal GroEL sequences obtained from Uniprot

b Spots from 1 to 65 originated from a pH 4 to 7 gradient gel (Supplementary Fig. 1), spots from NC1 to NC14 originated from a pH 3 to 10 gradient gel (data not shown) b Spots from 1 to 65 originated from a pH 4 to 7 gradient gel (Supplementary Fig. 1), spots from NC1 to NC14 originated from a pH 3 to 10 gradient gel (data not shown)

AGBTC9, AGTGF6, ÁGVTK5, ATEKR4, ATMQX5, A7ZIA8, Á7ZXY8, Á8AJ84, A8GB83, PÓA836, PÓA837, PÓA838, PÓA839, P58593, P66869, P66870, QOZK73, QOT6W6, QOTJW6,
QICAGI, QICFMO, QIITL3, QIREJ8, Q2IIW6, Q2NUM2, Q2SD35, Q32414, Q32IK3 ^c Several proteins may correspond to this identification: Q3KFU6, A1A8Y0, A1FGM7, A1JRB6, A1KTM6, A2UEB0, A3HHN5, A3M887, A4TNT8, A4W879, A4XV90, A5W114, A5WC33, Several proteins may correspond to this identification: Q3KFU6, A1A8Y0, A1FGM7, A1JRB6, A1KTM6, A2UEB0, A3HHN5, A3M887, A4TNT8, A4W879, A4XV90, A5W114, A5WC33, A6BTC9, A6T6F6, A6V7K5, A7FKR4, A7MQX5, A7ZJA8, A7ZXY8, A8AJ84, A8GB83, P0A836, P0A837, P0A838, P0A839, P53593, P66869, P66870, Q02K73, Q0T6W6, Q0TJW6, Q1CAG1, Q1CFM0, Q1I7L3, Q1REJ8, Q21IW6, Q2NUM2, Q2SD35, Q324I4, Q32IK3, Q3Z476, Q48K68, Q4FVH9, Q4KFY6, Q4ZUW7, Q57RL3, Q5F878, Q5PCM7, Q66DA0, Q6D7G2, Q6F8L4, Q7N6V5, Q7NZ47, Q883Z4, Q88FB2, Q8ZH00, Q9JUT0, Q9JZP4

39 bacterial proteins detected, there were two distinct ATP synthases, two distinct 50 S ribosomal proteins, one phosphoglycerate kinase, and one succinyl-CoA synthetase, which are involved in energy metabolism, translation, glycolysis, and TCA cycle, respectively (Table 3). These proteins may not play a specific role in this environment since they are known to be housekeeping proteins in bacteria. Among the other proteins, one outer protein OrpI, one uncharacterized protein with an unknown function, one glutathione-dependent, and one glutathione-independent formaldehyde dehydrogenases were identified. The latter two proteins belong to the formaldehyde detoxification pathway. Although no correlation with the environmental conditions might explain the functional specificities of these proteins, it has been reported that the arsenite-oxidizing bacteria H. arsenicoxydans synthesizes alcohol dehydrogenase and glutathione-dependent formaldehyde dehydrogenase when grown in the presence of arsenic [58].

It is worth noting that more than half of the identified bacterial proteins were similar to the 60-kDa GroEL chaperonin. These data suggest that multiple chaperonins of various genetic origins are expressed by the Reigous creek community. GroEL is known to be ubiquitously present in Bacteria and Archaea. These proteins are generally abundantly expressed in bacterial cells, especially under stress conditions such as those occurring in this particularly toxic environment [3]. The *groEL* gene is conserved in prokaryotes, and has been found to be present in one copy in the majority of sequenced genomes, except in the case of some pathogens [45, 60]. Because of its conservation properties (supplementary Fig. 2), this gene is often used as a phylogenetic marker [31]. In addition, it has been previously established that some of the amino acids stretch occurring in GroEL are specific to one genus or family of bacteria. These peptide sequences can therefore be used as the signature of a specific phylogenetic group. These GroEL identifications (Table 3, Fig. 3) were, therefore, considered for use as a possible taxonomic tool in addition to the 16S rRNA-based taxonomic approach. To determine which bacteria in the whole community were active, i.e., able to express proteins, the organisms identified using the 16S rRNA encoding gene library- and GroELbased approaches (Fig. 2) were compared.

Most of the bacteria identified based on GroEL belonged to five phyla divisions. Bacteria belonging to Deinococcus and Synergistetes did not feature among those identified based on the 16S rRNA encoding gene. One possible explanation for this discrepancy may be that a PCR or cloning bias may have prevented those bacteria from being detected with this method. These findings suggest that metaproteomic methods used as taxonomic tools can provide a useful complementary tool in addition to the 16S rRNA encoding gene approach. Interestingly, Thiomonas, At. ferrooxidans, Acidiphilium, and Gallionella, expressed proteins and were, therefore, active.

16S rRNA-encoding gene analysis showed that these bacteria abound in this ecosystem. Bacteria affiliated to At. ferrooxidans and Gallionella are able to oxidize iron. In addition, many strains of the Thiomonas genus are able to oxidize As(III) into As(V) under laboratory conditions [6, 14, 22]. The fact that their proteins were detected shows that these bacteria were viable and metabolically active. This finding supports the hypothesis that the oxidation of Fe(II) to Fe(III) catalyzed by iron-oxidizing microorganism such as At. ferrooxidans and Gallionella and oxidation of As(III) into As(V) by As(III) oxidizers such as Thiomonas, probably leads to the precipitation of the more or less ordered iron oxy-hydroxides (Fe(III)–As(V) hydroxysulfate) detected in this study (Fig. 1) [48]. This finding is in agreement with previous data showing that Gallionella ferruginea efficiently remove Fe, As(III), and As(V) in water [41]. The present data show for the first time that this bacterium is active and probably plays a functional role in the sediments of the Reigous creek. Some heterotrophic bacteria such as Acidiphilium were also found to be active in this AMD, suggesting that they could cope with the low amount of organic carbon (dissolved organic carbon concentration 1.7 ± 0.4 mg/L [16]) of Reigous creek water. It has previously been suggested that these acidophilic heterotrophic bacteria may be involved in organic carbon turnover processes [32]. Interestingly, these four bacteria (At. ferrooxidans, Thiomonas, Gallionella, and Acidiphilium) found to be active members of this AMD community have been previously identified in AMD, but some of them were thought to have different optimum pH levels. Indeed, G. ferruginea is a neutrophilic bacterium which oxidizes Fe, but relatives of Gallionella, have often been detected in AMD [12, 33-35]. The strain occurring at Carnoulès showed less than 97% homology with G. capsiferriformans and its physiological characteristics are probably different. It seems probable that an acid-tolerant relative of this bacterium is able to oxidize iron under acid pH conditions.

In addition to the bacteria belonging to these four genera, other bacteria were also found to be active, but a discrepancy was again observed between the bacteria identified based on the results of 16S rRNA encoding gene analysis and the metaproteomic approach. The GroEL protein sequences of some bacteria identified using the 16S rRNA encoding gene library were not available in the Uniprot database, which might explain this discrepancy. However, phylogenetic comparisons between the 16S rRNA and metaproteomic data obtained (Fig. 2) suggested which of the bacteria present in this ecosystem may express a GroEL identified in the metaproteomic study. Based on these comparisons, it seems likely that in addition to Thiomonas, At. ferrooxidans, Acidiphilium, and Gallionella, clones related to β -, γ -, and δ -proteobacteria, such as Limnobacter or Methylococcus (At. ferrooxidans DSM

onins. Thirty-six sequences matched the MS results: 29 Proteins were identified in Table 3: 22 of them unambiguously matched one protein (the bacterial name of which is given in orange), whereas fifteen identifications matched at least two proteins (22 bacterial names given in black). Thiomonas sp. 3As data were added for the sake of comparison (in *blue*). Thirty-seven sequences in all were aligned. These sequences were compared with ClustalW2 using the default parameters (http://www.ebi.ac.uk/Tools/clustalw2/index.html). Alignment was performed using the Jalview software program [19]. The blue highlighted letters correspond to identical residues among the 37 orthologs (BLOSUM62 score option). Highlighted letters corresponding to the peptide sequences identified by MS allowed distinguishing several orthologs. First, 23 proteins were identified because at least one identified peptide matched only this GroEL: these peptides (labeled in red) corresponding to signature sequences specific to one species, genus, or family of bacteria are located in the same homology regions but differed from other peptides by at least one amino acid substitution. Secondly, 13 of the proteins identified were confirmed, since several identified peptides (labeled in green) matched this GroEL. Each of these individual peptides matched several proteins; however, only one protein detected in Uniprot contained all these amino-acid sequences, which suggests that a relative of this protein was probably present in this extract. Since the full amino-acid sequences of almost all the

chaperonins expressed by the bacteria present at the study site are unknown, the possibility cannot be ruled out that two identified peptides were erroneously assigned to two distinct chaperonins, whereas these peptides may in fact have originated from one protein, the amino acid sequence of which has not yet been included in the databases. Nevertheless, because of the high level of conservation observed in GroEL proteins (Supplementary Fig. 2), the proteins identified may be expressed by a close relative of the organism identified using the Uniprot database. GroEL originated from Acidiphilium cryptum; Acidithiobacillus ferrooxidans ATCC23270T; Acidithiobacillus ferrooxidans ATCC53993; Acidovorax avenae subsp. avenae; Acidovorax avenae subsp. citrulli; Alkaliphilus metalliredigens; Alkaliphilus oremlandii; Bacillus halodurans; Bacillus pseudofirmus; Bacillus selenitireducens; Bordetella petrii; Clostridium botulinum; Clostridium hiranonis; Clostridium papyrosolvens; Deinococcus desertii; Deinococcus sp. A62; Gallionella ferruginea; Haliangium ochraceum; Lactobacillus delbrueckii subsp. delbrueckii; Lactobacillus delbrueckii subsp. indicus; Leptothrix cholodnii; Limnobacter sp. MED105; Methylibium petroleiphilum; Methylococcus capsulatus; Methylotenera mobilis; Propionibacterium freudenreichii; Ralstonia picketii; Rhodobacterales bacterium HTCC2083; Sinorhizobium medicae; Stackebrandtia nassauensis; Streptosporangium roseum; Thermanaerovibrio acidaminovorans; Thiomonas 3As; Thiomonas intermedia; Verminephrobacter eiseniae; Virgibacillus pantothenticus; and Gammaproteobacterium HTCC2207

Figure 3 (continued)

2392 affiliated bacteria), Methylotenera (Thiobacillusrelated clones), D. koreensis (which is affiliated to the γ-proteobacteria HTCC2207), Ferrimicrobium-like bacteria (which are related to the Actinobacteria S. nassauensis and S. roseum), and Haliangium (clones affiliated to sulfatereducing bacteria JHA1, D. limimaris and D. svalbardensis) may play a role in this ecosystem. All in all, the data obtained here show that Firmicutes, which could be affiliated to Alicyclobacillus ferrooxidans, are also active. These bacteria are able to oxidize ferrous iron [37] and may, therefore, participate in the transformation of the iron present in high concentrations in these waters. The active population as a whole was not only composed of several ironoxidizers in addition to At. ferrooxidans but also contained iron reducers, one known arsenite-oxidizer, sulfate-reducing, and sulfur compound oxidizers, and both autotrophic and heterotrophic bacteria. All these bacteria

may contribute importantly to the remediation process observed in situ.

In Conclusion

The active bacterial species inhabiting Carnoulès AMD ecosystem were identified in this study using highsensitivity nanoLC-chip-MS/MS methods combined with a 16S rRNA based phylogenetic approach. The metaproteomic data obtained here show for the first time that Gallionella, Thiomonas, At. ferrooxidans, and Acidiphilium actively express proteins in situ. Previous hypotheses based on experiments performed under laboratory conditions [14, 18, 21, 22, 24, 48] suggest that microbial activity may contribute to the arsenite oxidation and As immobilization occurring in the heavily contaminated AMD at the Carnoulès mine via iron oxidation processes.

Since these bacteria were found to be active and to express proteins which are among the most abundant proteins encountered at this site, it seems likely that the large amounts of pale-yellow As(V)–Fe(III) hydroxysulfate sediments forming at Carnoulès, which were characterized here, may result from the conjugate activities of ironoxidizing microorganisms (such as At. ferrooxidans, Alicyclobacillus ferrooxydans, Ferrimicrobium, or Gallionella) and arsenic-oxidizing microorganisms (such as Thiomonas sp.). Several bacteria may be responsible in situ for changing the ratio between the oxidized and reduced forms of iron, arsenic, and sulfur compounds, promoting the formation of the Fe (III)–As(V) hydroxysulfate precipitates detected in this study. These bacteria are therefore of prime importance in the partial but efficient natural process of remediation undergone by the contaminated Carnoulès ecosystem. In addition, autotrophic iron, arsenic, and sulfur oxidizers may provide the organic carbon sources required by the functional heterotrophs such as Acidiphilium present in this ecosystem. All in all, the present findings provide evidence that the functional genomics approach provides a useful means of describing bacterial communities such as those inhabiting the Reigous creek and determining their contribution to natural attenuation processes. It is now proposed to use approaches of this kind in future studies to complete this picture of the functional processes at work in this ecosystem, as well as to investigate the role played by the less abundant active bacteria identified in this study.

Acknowledgements The study was financed by the EC2CO program ("Institut National des Sciences de l'Univers," CNRS), the "Observatoire de Recherche Méditerranéen en Environnement" (OSU-OREME), and by the ANR 07-BLANC-0118 project ("Agence Nationale de la Recherche"). Sébastien Gallien and Aurélie Volant were supported by a grant from the French Ministry of Education and Research. This work was performed in the framework of the "Groupement de recherche: Métabolisme de l'Arsenic chez les Microorganismes" (GDR2909-CNRS).

References

- 1. Altschul SF, Gish W, Miller W, Myers EW, Lipman DJ (1990) Basic local alignment search tool. J Mol Biol 215:403–410
- 2. Amaral Zettler LA, Gomez F, Zettler E, Keenan BG, Amils R, Sogin ML (2002) Microbiology: eukaryotic diversity in Spain's River of Fire. Nature 417:137
- 3. Arsène F, Tomoyasu T, Bukau B (2000) The heat shock response of Escherichia coli. Int J Food Microbiol 55:3–9
- 4. Ashelford KE, Chuzhanova NA, Fry JC, Jones AJ, Weightman AJ (2005) At least 1 in 20 16S rRNA sequence records currently held in public repositories is estimated to contain substantial anomalies. Appl Environ Microbiol 71:7724–7736
- 5. Baker BJ, Banfield JF (2003) Microbial communities in acid mine drainage. FEMS Microbiol Ecol 44:139–152
- 6. Battaglia-Brunet F, Dictor MC, Garrido F, Crouzet C, Morin D, Dekeyser K, Clarens M, Baranger P (2002) An arsenic(III) oxidizing bacterial population: selection, characterization, and performance in reactors. J Appl Microbiol 93:656–667
- 7. Bednar AJ, Garbarino JR, Ranville JF, Wildeman TR (2002) Preserving the distribution of inorganic arsenic species in groundwater and acid mine drainage samples. Environ Sci Technol 36:2213–2218
- 8. Bertin PN, Médigue C, Normand P (2008) Advances in environmental genomics: towards an integrated view of microorganisms and ecosystems. Microbiology 154:347–359
- 9. Bond PL, Smriga SP, Banfield JF (2000) Phylogeny of microorganisms populating a thick, subaerial, predominantly lithotrophic biofilm at an extreme acid mine drainage site. Appl Environ Microbiol 66:3842–3849
- 10. Bowell RJ (1994) Sorption of arsenic by iron oxides and oxyhydroxides in soils. Appl Geochem 9:279–286
- 11. Brake SS, Hasiotis ST (2010) Eukaryote-dominated biofilms and their significance in acidic environments. Geomicrobiol J 27:534–558
- 12. Bruneel O, Duran R, Casiot C, Elbaz-Poulichet F, Personné JC (2006) Diversity of microorganisms in Fe-As-rich acid mine drainage waters of Carnoules, France. Appl Environ Microbiol 72:551–556
- 13. Bruneel O, Pascault N, Egal M, Bancon-Montigny C, Goni-Urriza MS, Elbaz-Poulichet F, Personné JC, Duran R (2008) Archaeal diversity in a Fe-As rich acid mine drainage at Carnoules (France). Extremophiles 12:563–571
- 14. Bruneel O, Personné JC, Casiot C, Leblanc M, Elbaz-Poulichet F, Mahler BJ, Le Flèche A, Grimont PA (2003) Mediation of arsenic oxidation by Thiomonas sp. in acid-mine drainage (Carnoulès, France). J Appl Microbiol 95:492–499
- 15. Casiot C, Bruneel O, Personné JC, Leblanc M, Elbaz-Poulichet F (2004) Arsenic oxidation and bioaccumulation by the acidophilic protozoan, Euglena mutabilis, in acid mine drainage (Carnoules, France). Sci Total Environ 320:259–267
- 16. Casiot C, Egal M, Bruneel O, Bancon-Montigny C, Cordier MA, Gomez E, Aliaume C, Elbaz-Poulichet F (2009) Hydrological and geochemical controls on metals and arsenic in a Mediterranean river contaminated by acid mine drainage (the Amous river, France); preliminary assessment of impacts on fish (Leuciscus cephalus). Appl Geochem 24:787–799
- 17. Casiot C, Lebrun S, Morin G, Bruneel O, Personné JC, Elbaz-Poulichet F (2005) Sorption and redox processes controlling arsenic fate and transport in a stream impacted by acid mine drainage. Sci Total Environ 347:122–130
- 18. Casiot C, Morin G, Juillot F, Bruneel O, Personné JC, Leblanc M, Duquesne K, Bonnefoy V, Elbaz-Poulichet F (2003) Bacterial immobilization and oxidation of arsenic in acid mine drainage (Carnoulès creek, France). Water Res 37:2929–2936
- 19. Clamp M, Cuff J, Searle SM, Barton GJ (2004) The Jalview Java alignment editor. Bioinformatics 20:426–427
- 20. Coupland K, Johnson DB (2008) Evidence that the potential for dissimilatory ferric iron reduction is widespread among acidophilic heterotrophic bacteria. FEMS Microbiol Lett 279:30–35
- 21. Duquesne K, Lebrun S, Casiot C, Bruneel O, Personné JC, Leblanc M, Elbaz-Poulichet F, Morin G, Bonnefoy V (2003) Immobilization of arsenite and ferric iron by Acidithiobacillus ferrooxidans and its relevance to acid mine drainage. Appl Environ Microbiol 69:6165–6173
- 22. Duquesne K, Lieutaud A, Ratouchniak J, Muller D, Lett MC, Bonnefoy V (2008) Arsenite oxidation by a chemoautotrophic moderately acidophilic Thiomonas sp.: from the strain isolation to the gene study. Environ Microbiol 10:228–237
- 23. Egal M, Casiot C, Morin G, Elbaz-Poulichet F, Cordier MA, Bruneel O (2010) An updated insight into the natural attenuation of As concentrations 3 in Reigous Creek (southern France). Appl Geochem 25:1949–1957
- 24. Egal M, Casiot C, Morin G, Parmentier M, Bruneel O, Lebrun S, Elbaz-Poulichet F (2009) Kinetic control on the formation of

tooeleite, schwertmannite and jarosite by Acidithiobacillus ferrooxidans strains in an As(III)-rich acid mine water. Chem Geol 265:432–441

- 25. Elias JE, Gygi SP (2007) Target-decoy search strategy for increased confidence in large-scale protein identifications by mass spectrometry. Nat Methods 4:207–214
- 26. Felsenstein J (1992) Estimating effective population size from samples of sequences: a bootstrap Monte Carlo integration method. Genet Res 60:209–220
- 27. Ferris FG, Hallberg RO, Lyvén B, Pedersen K (2000) Retention of strontium, cesium, lead and uranium by bacterial iron oxides from a subterranean environment. Appl Geochem 15:1035–1042
- 28. Finneran KT, Johnsen CV, Lovley DR (2003) Rhodoferax ferrireducens sp. nov., a psychrotolerant, facultatively anaerobic bacterium that oxidizes acetate with the reduction of Fe(III). Int J Syst Evol Microbiol 53:669–673
- 29. Good IJ (1953) The population frequencies of species and the estimation of population parameters. Biometrika 40:237–264
- 30. Goto K, Mochida K, Kato Y, Asahara M, Fujita R, An SY, Kasai H, Yokota A (2007) Proposal of six species of moderately thermophilic, acidophilic, endospore-forming bacteria: Alicyclobacillus contaminans sp. nov., Alicyclobacillus fastidiosus sp. nov., Alicyclobacillus kakegawensis sp. nov., Alicyclobacillus macrosporangiidus sp. nov., Alicyclobacillus sacchari sp. nov. and Alicyclobacillus shizuokensis sp. nov. Int J Syst Evol Microbiol 57:1276–1285
- 31. Gupta RS (1998) Protein phylogenies and signature sequences: a reappraisal of evolutionary relationships among archaebacteria, eubacteria, and eukaryotes. Microbiol Mol Biol Rev 62:1435– 1491
- 32. Hallberg KB (2010) New perspectives in acid mine drainage microbiology. Hydrometallurgy 104:448–453
- 33. Hallberg KB, Coupland K, Kimura S, Johnson DB (2006) Macroscopic streamer growths in acidic, metal-rich mine waters in north wales consist of novel and remarkably simple bacterial communities. Appl Environ Microbiol 72:2022–2030
- 34. He Z, Xiao S, Xie X, Zhong H, Hu Y, Li Q, Gao F, Li G, Liu J, Qiu G (2007) Molecular diversity of microbial community in acid mine drainages of Yunfu sulfide mine. Extremophiles 11:305–314
- 35. Heinzel E, Hedrich S, Janneck E, Glombitza F, Seifert J, Schlomann M (2009) Bacterial diversity in a mine water treatment plant. Appl Environ Microbiol 75:858–861
- 36. Hippe H (2000) Leptospirillum gen. nov. (ex Markosyan 1972), nom. rev., including Leptospirillum ferrooxidans sp. nov. (ex Markosyan 1972), nom. rev. and Leptospirillum thermoferrooxidans sp. nov. (Golovacheva et al. 1992). Int J Syst Evol Microbiol 50(Pt 2):501–503
- 37. Jiang CY, Liu Y, Liu YY, You XY, Guo X, Liu SJ (2008) Alicyclobacillus ferrooxydans sp. nov., a ferrous-oxidizing bacterium from solfataric soil. Int J Syst Evol Microbiol 58:2898–2903
- 38. Johnson DB (1995) Acidophilic microbial communities: candidates for bioremediation of acidic mine effluents. Int Biodeterior Biodegrad 35:41–58
- 39. Johnson DB, Hallberg KB (2003) The microbiology of acidic mine waters. Res Microbiol 154:466–473
- 40. Johnson DB, Hallberg KB (2005) Acid mine drainage remediation options: a review. Sci Total Environ 338:3–14
- 41. Katsoyiannis IA, Zouboulis AI (2004) Application of biological processes for the removal of arsenic from groundwaters. Water Res 38:17–26
- 42. Kelly DP, Wood AP (2000) Confirmation of Thiobacillus denitrificans as a species of the genus Thiobacillus, in the betasubclass of the Proteobacteria, with strain NCIMB 9548 as the type strain. Int J Syst Evol Microbiol 50(Pt 2):547–550
- 43. Lane DJ, Pace B, Olsen GJ, Stahl DA, Sogin ML, Pace NR (1985) Rapid determination of 16S ribosomal RNA sequences for phylogenetic analyses. Proc Natl Acad Sci USA 82:6955–6959
- 44. Letunic I, Bork P (2007) Interactive tree of life (iTOL): an online tool for phylogenetic tree display and annotation. Bioinformatics 23:127–128
- 45. Lund PA (2009) Multiple chaperonins in bacteria–why so many? FEMS Microbiol Rev 33:785–800
- 46. Moreira D, Amils R (1997) Phylogeny of Thiobacillus cuprinus and other mixotrophic thiobacilli: proposal for Thiomonas gen. nov. Int J Syst Bacteriol 47:522–528
- 47. Morin G, Calas G (2006) Arsenic in soils, mine tailings, and former industrial sites. Elements 2:97–101
- 48. Morin G, Juillot F, Casiot C, Bruneel O, Personné JC, Elbaz-Poulichet F, Leblanc M, Ildefonse P, Calas G (2003) Bacterial formation of tooeleite and mixed arsenic(III) or arsenic(V)-iron (III) gels in the Carnoules acid mine drainage, France. A XANES, XRD, and SEM study. Environ Sci Technol 37:1705–1712
- 49. Ona-Nguema G, Morin G, Juillot F, Calas G, Brown GE (2005) EXAFS analysis of arsenite adsorption onto two-line ferrihydrite, hematite, goethite, and lepidocrocite. Environ Sci Technol 39:9147–9155
- 50. Rodier J, Broutin JP, Chambon P, Champsaur H, Rodi L (1996) L'Analyse des Eaux. Dunod, Paris, p 1383
- 51. Rowe OF, Johnson DB (2008) Comparison of ferric iron generation by different species of acidophilic bacteria immobilized in packed-bed reactors. Syst Appl Microbiol 31:68–77
- 52. Rowe OF, Sanchez-Espana J, Hallberg KB, Johnson DB (2007) Microbial communities and geochemical dynamics in an extremely acidic, metal-rich stream at an abandoned sulfide mine (Huelva, Spain) underpinned by two functional primary production systems. Environ Microbiol 9:1761–1771
- 53. Sánchez España JS, LópezPamo E, Santofimia Pastor E, Reyes Andrés J, Rubi J (2005) The natural attenuation of two acidic effluents in Tharsis and La Zarza-Perrunal mines (Iberian Pyrite Belt, Huelva, Spain). Environ Geol 49:253–266
- 54. Sun B, Cole JR, Tiedje JM (2001) Desulfomonile limimaris sp. nov., an anaerobic dehalogenating bacterium from marine sediments. Int J Syst Evol Microbiol 51:365–371
- 55. Thompson JD, Plewniak F, Thierry J, Poch O (2000) DbClustal: rapid and reliable global multiple alignments of protein sequences detected by database searches. Nucleic Acids Res 28:2919–2926
- 56. Vandieken V, Mussmann M, Niemann H, Jorgensen BB (2006) Desulfuromonas svalbardensis sp. nov. and Desulfuromusa ferrireducens sp. nov., psychrophilic, Fe(III)-reducing bacteria isolated from Arctic sediments, Svalbard. Int J Syst Evol Microbiol 56:1133–1139
- 57. Weisburg WG, Barns SM, Pelletier DA, Lane DJ (1991) 16S ribosomal DNA amplification for phylogenetic study. J Bacteriol 173:697–703
- 58. Weiss S, Carapito C, Cleiss J, Koechler S, Turlin E, Coppee J-Y, Heymann M, Kugler V, Stauffert M, Cruveiller S, Médigue C, Van Dorsselaer A, Bertin PN, Arsène-Ploetze F (2009) Enhanced structural and functional genome elucidation of the arseniteoxidizing strain Herminiimonas arsenicoxydans by proteomics data. Biochimie 91:192–203
- 59. Weiss JV, Rentz JA, Plaia T, Neubauer SC, Merrill-Floyd M, Lilburn T, Bradburne C, Megonigal JP, Emerson D (2007) Characterization of neutrophilic Fe(II)-oxidizing bacteria isolated from the rhizosphere of wetland plants and description of Ferritrophicum radicicola gen. nov. sp. nov., and Sideroxydans paludicola sp. nov. Geomicrobiol J 24:559-570
- 60. Williams TA, Codoner FM, Toft C, Fares MA (2010) Two chaperonin systems in bacterial genomes with distinct ecological roles. Trends Genet 26:47–51