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# **Using blood and feathers to investigate large-scale Hg contamination in Arctic seabirds: A review**

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**Abstract:** Mercury (Hg), because of its deleterious effects on wildlife and its high concentrations in polar regions, has been widely studied in the Arctic. This provided important information regarding food web contamination, spatial and temporal trends of Hg in ecosystems or risk assessments for wildlife and Humans. Among the Arctic biota, seabirds have been among the most studied species due to their sensitivity to this toxicant, their role as bioindicators of the contamination status of their environment, and their consumption by Arctic communities. However, most studies that investigated Hg in Arctic seabirds focused on measurements in internal organs or in eggs, while few investigations have been performed on blood and feathers, despite the relevant and complementary information they provide. Here, we first provide a detailed overview of the specific information blood and feathers can bring when investigating Hg contamination of Arctic seabirds, including new knowledge on the poorly studied non-breeding period. Second, we perform a comprehensive review of the use of blood and feathers as non-lethal tissues to study Hg in Arctic seabirds. This review demonstrates important interspecific variations in Hg blood concentrations according to seabird trophic status, with seaducks generally presenting the lowest Hg concentrations while auks have the highest ones. However, all the observed Hg concentrations are below the admitted toxicity thresholds. Hg concentrations in feathers follow similar trends and gulls appear to be the most contaminated species, likely as a consequence of contrasting migratory and overwintering strategies. This review also confirms strong spatial variations with higher concentrations found in the Canadian Arctic and Pacific waters than in Greenland and the European Arctic. It also identifies some major understudied areas such as West Greenland, Aleutian Islands and Russia. Finally, we provide a thorough review of the current knowledge regarding molting patterns in Arctic seabirds, which is an essential information to interpret Hg concentrations measured in feathers. Overall, our results point out the importance of blood and feathers in seabird ecotoxicological assessments and highlight the need for large scale international collaborations and research programs.

**Keywords:** Seabirds; Spatial ecotoxicology; Molting pattern; Mercury

## **1. Introduction**

Mercury (Hg) is a metallic trace element naturally found in the environment, released by volcanic eruptions or weathering. However, human activities, such as coal burning or gold mining, have been responsible for a large increase in Hg emissions, especially since the 19<sup>th</sup> century (UNEP, 2013). Because of its solubility, its ability to enter and biomagnify along the food chains (as methylmercury – MeHg (Morel et al., 1998)), and its toxicity for human and wildlife (e.g., Tan et al., 2009; Wolfe et al., 1998), this pollutant requires close monitoring. More specifically, a major emphasis has been put towards the understanding of Hg levels in marine organisms, underlying modulators and resulting impacts on nervous systems, reproduction or development (Eagles-Smith et al., 2018). Anthropogenic Hg is mainly emitted by midlatitude northern countries, with still increasing emissions in some regions such as Asia (UNEP, 2013). It is then transported over large distances to polar regions by atmospheric and oceanic currents and, in the case of the Arctic, by rivers (Sonke et al., 2018; UNEP, 2013). Therefore, polar top predators are among the most exposed organisms to Hg (Cherel et al., 2018) and have thus been largely studied both in toxicity and conservation contexts, but also as efficient indicators to monitor Hg contamination in polar marine ecosystems (e.g., Carravieri et al., 2013; Dietz et al., 2013; Mallory et al., 2017; Sun et al., 2006; Verreault et al., 2010). Among those polar top predators, Arctic seabirds, being long-lived species with a circumpolar distribution, have been widely used to study long-term and relatively large-scale Hg exposure (see Becker, 2003). However, these studies mainly focused on the period during which seabirds are at their breeding site (hereafter the breeding season), even though this period represents a part of their annual cycle only. Conversely, investigations for the non-breeding period are still scarce, mainly because most of seabirds are offshore during this period, making their sampling logistically difficult. Previous studies suggested that Arctic seabirds could be exposed to higher Hg levels during the non-breeding period with

subsequent impacts on their reproduction (e.g. Fort et al., 2014). Hence, improving our knowledge of bird contamination during the non-breeding period – which mostly occurs outside the Arctic – is crucial, first to determine its role on seabird exposure to Hg, but also to better understand the origin of this contamination, in space and time.

Several tissues were targeted within the various studies which investigated Hg in Arctic seabirds, providing information about seabird Hg contamination at different temporal scales. Indeed, once ingested through the diet (the main source of Hg intake), Hg enters the bloodstream where it stays a few weeks/months (e.g. half-life of 30 days in great skua (*Stercorarius skua*) (Bearhop et al., 2000) and of 40–65 days in Cory's shearwaters (*Calonectric borealis*) (Monteiro and Furness, 2001)). Hg in blood thus informs about recent intakes (e.g. from late migration to chick-rearing when blood is sampled during the breeding season). Once Hg is assimilated, it is transported and distributed to the different organs, where it can be demethylated (i.e. protective mechanism against Hg toxicity) and/or stored (Dietz et al., 2013). While internal organs such as liver, brain, kidney and muscle give access to longer term Hg contamination compared to blood (Mallory et al., 2018), the liver also informs, depending on the species, about Hg demethylation abilities (Dehn et al., 2005; Kim et al., 1996b; Thompson and Furness, 1989). Finally, egg production and feather synthesis (molt) are two major excretion pathways for Hg. The study of eggs therefore informs about seabird Hg contamination during the laying time (and/or during the pre-breeding time in case of “capital” breeders (i.e. allocation of stored nutrients to egg formation) (Becker, 2003; Bond and Diamond, 2010; Mallory et al., 2017)). Cover feathers provide information about Hg contamination during a relatively longer period (over several weeks/months according to species and molting patterns (Furness et al., 1986, see below, Annex 1)). Because of toxicological risks for seabirds and consumers (seabirds and their eggs are consumed in several Arctic regions (Bard, 1999; Chan, 1998; Mallory and Braune, 2012)), internal tissues

and eggs have received most of the attention (Figs. 1 and 2). Conversely, non-consumed tissues such as blood and feathers received much less attention, despite the relevant and complementary spatio-temporal information they could provide, and the fact that they can be non-lethally sampled. In the present study, we thus review and discuss the current knowledge of Hg contamination in Arctic seabirds, with a specific focus on the information provided by blood and feather tissues. We will also highlight how these tissues and their combination with new approaches could help us improving knowledge of seabird yearly (breeding and non-breeding) Hg contamination.

## **2. Methods**

### **2.1. Reviewing process**

A review of the literature was done on Scopus searching for three keywords: “seabirds”, “Arctic” and “mercury” on December 2018. The Arctic boundaries used were as defined by the Conservation of Arctic Flora and Fauna (CAFF) working group (CAFF, 1996), and only the species breeding within these boundaries and spending most of the year feeding in oceanic food webs were included in the analysis (Annex 1). Blood and feather Hg concentrations were extracted and, when necessary, blood values were converted to dry weight using a moisture of 79.13% (Eagles-Smith et al., 2008) to allow comparison between published values. We considered feather concentrations provided in wet weight directly comparable to those in dry weight as moisture content is negligible in feathers (e.g. Stettenheim, 2000). A review of molting patterns for the retained species was also performed and is summarized in Annex 1.

### **2.2. Statistical analyses**

The dataset does not provide sufficient information to test any temporal or spatial trend in Hg concentrations (underrepresentation of some families, species, countries or years). However,

for both blood and feathers, we used ANOVA to compare Hg concentrations between families (see Figs. S1 and S2). More specifically, for blood, we used mean values provided by each study for each species and then compared Hg concentrations between families being adults anatids (n=29) and larids (n=19) and removed the alcids (n=4) and the procellariids (n=1) because of limited data (see Fig. S1). For feathers, we used mean values provided by each study for each species and then compared Hg concentrations between families being adults alcids (n=31), anatids (n=8) and larids (n=19) and removed Hg concentrations for hydrobatids (n=1) and stercorariids (n=2) because of limited data (see Fig. S2). Statistical significance was assumed at  $p < 0.05$  and analyses were run with R studio (version 1.2.1335).

### **3. Results and discussion**

#### **3.1. Hg assessment in blood and feathers of Arctic seabirds**

The present review shows that 101 published articles (see Annex 2) reported Hg concentrations in Arctic seabirds (i.e. species/populations breeding in the Arctic) since 1984. In total, 42 species (list provided in Annex 1) from all around the Arctic were assessed, yet with an emphasis on the Canadian Arctic (n=38 publications), Svalbard (n=23), Aleutian Islands (n=16), Alaska (n=14) and West Greenland (n=7). In addition, internal tissues and eggs were more investigated than blood and feathers (81, 30, 22 and 19 studies, respectively), also integrating a wider spatial scale (Figs. 1 and 2). Among studied species, three families received most of the attention: gulls (hereafter larids) (n=56 publications), auks (hereafter alcids) (n=49) and seaducks (hereafter anatids) (n=44), covering all foraging strategies (benthic, pelagic, coastal, oceanic, divers, surface feeders, scavengers) and diets (insectivorous, molluscivorous, planktivorous, piscivorous, omnivorous).

Blood Hg concentrations were investigated in a total of 12 seabird species but anatids and larids were the most studied (see Table 1). For the Anatid family (n=12 studies), the common

eider (*Somateria mollissima*), the king eider (*Somateria spectabilis*), the Steller's eider (*Polysticta stelleri*), the spectacled eider (*Somateria fischeri*) and the long-tailed duck (*Clangula hyemalis*) were studied, covering Alaska, Svalbard and the Canadian Arctic. For the Larid family, investigations focused on the black-legged kittiwake (*Rissa tridactyla*) (n=8 studies) covering both the Canadian Arctic and Svalbard, and the ivory gull (*Pagophila eburnea*) (n=1 study) in Svalbard. Finally, four studies investigated Hg contamination of alcids breeding in the Canadian Arctic (Brünnich's guillemots (*Uria lomvia*) and black guillemots (*Cepphus grylle*)), East Greenland (little auks (*Alle alle*)) and Svalbard (Mandt's black guillemot (*Cepphus grylle mandtii*)), and one study investigated Hg contamination in procellariids (northern fulmars (*Fulmarus glacialis*)) from the Canadian Arctic.

Most studies which measured Hg concentrations in feathers focused on body (n=18 publications) and flight (n=4) feathers. They covered more areas than blood investigations and were mainly performed in the Aleutian Islands (n=7 publications) and in Alaska (n=3), followed by East Greenland (n=3), Svalbard (n=2), Russia (n=2), the Canadian Arctic (n=2) and Norway (n=1). They encompassed alcids (n=10), larids (n=9), anatids (n=8), procellariids (n=2) and stercorariids (n=1) (see Tables 2 and 3). There was a large heterogeneity between the different regions. In the Pacific Arctic, the Aleutian Islands are the only area where Hg contamination has been measured in feathers, targeting alcids (Kittlitz's murrelet (*Brachyramphus brevirostris*), pigeon guillemot (*Cepphus columba*), tufted puffin (*Fratercula cirrhata*)), anatids (common eider (*Somateria mollissima*)), hydrobatids (fork-tailed storm-petrel (*Oceanodroma furcata*)) and larids (blacklegged kittiwakes (*Rissa tridactyla*), glaucous-winged gull (*Larus glaucescens*)). In the Atlantic Arctic, each country was studied, providing information for alcids (Brünnich's guillemot, common guillemot (*Uria aalge*), little auk), anatids (king eider, common eider, long-tailed duck), larids (black-legged kittiwake, Arctic tern (*Sterna paradisae*), Sabine's gull (*Xema sabini*), ivory gull, glaucous



gull (*Larus hyperboreus*), herring gull (*Larus argentatus*)), procellariids (northern fulmar) and stercorariids (parasitic jaeger (*Stercorarius parasiticus*), long-tailed jaeger (*Stercorarius longicaudus*)). The common eider and the black-legged kittiwake are the only two species studied in both the Atlantic and Pacific Arctic. In Russia, only one site (Chaun, Siberia) was studied with feather Hg measurements performed in anatids (long-tailed duck and king eider), larids (herring gull, Arctic tern, glaucous gull, Sabine's gull) and stercorariids (long-tailed and parasitic jaeger). Flight feathers were processed in three studies that only focused on alcids (Brünnich's guillemot, Kittlitz's murrelet), anatids (long-tailed-duck, common eider), larids (herring gull, Arctic tern, black-legged kittiwakes, ivory gull and glaucous gull) and procellariids (northern fulmar) (Table 3) (Kenney et al., 2018; Kim et al., 1996b; Mallory et al., 2015). Finally, three studies focused on the non-breeding period only, and all were performed on little auks breeding in East Greenland (Fort et al., 2014, 2016; Amélineau et al., 2019).

### **3.2. Blood as a tool to monitor Hg contamination during the breeding season**

During the breeding season, seabirds aggregate in colonies where blood samples can be more easily collected. Sampling periods extend from the pre-laying to the late chick-rearing stages, allowing investigations on adult seabird short-term Hg contamination (i.e. from late migration to chick-rearing (Bearhop et al., 2000; Evers et al., 2008; Monteiro and Furness, 1995)). Hg analyses are done on whole blood or on blood cells (concentrations in these two matrices being highly correlated (Sire et al. unpublished)). Blood is also a good proxy for Hg concentrations in other internal tissues. Indeed, Hg concentrations in blood have been found to be strongly correlated with Hg concentrations in brain and muscle (e.g., Fort et al., 2015; Mallory et al., 2018). In addition, as a non-lethally sampled tissue, blood has several advantages. First, individuals can be sampled every year for long-term Hg monitoring at the

individual and population scales or several times along the breeding season. This provides specific information about Hg dynamics along the pre-laying, incubation and chick-rearing periods (see Lavoie et al., 2014). Second, a large number of individuals and species can be sampled at a broad scale, allowing spatial comparisons of Hg concentrations and thus large-scale monitoring. For instance, Ackerman et al. (2016) reviewed bird Hg contamination across the Northwestern America to map the distribution of Hg concentrations and highlighted hotspots of Hg contamination.

Our results show that anatids had lower Hg concentrations in blood than larids ( $F_{1,46}=14.36$ ,  $p < 0.001$ ). For anatids, Hg concentrations ranged from  $0.57 \pm 0.04 \mu\text{g/g dw}$  (long-tailed duck, Franson et al., 2004) to  $2.25 \pm 0.29 \mu\text{g/g dw}$  (common eider, Provencher et al., 2017) (Table 1, Figs. 3 and 4) with an average value of  $1.04 \pm 0.33 \mu\text{g/g dw}$ . For larids, concentrations ranged from  $0.81 \pm 0.05 \mu\text{g/g dw}$  (ivory gull, Lucia et al., 2016) to  $3.04 \pm 1.41 \mu\text{g/g dw}$  (black-legged kittiwake, Mallory et al., 2018) (Table 1, Figs. 3 and 4) with an average value of  $1.57 \pm 0.48 \mu\text{g/g dw}$ .

Finally, Hg concentrations in alcid blood samples ranged from  $0.32 \pm 0.80 \mu\text{g/g dw}$  to  $4.58 \mu\text{g/g dw}$ , both measured in black guillemots (Eckbo et al., 2019 and Mallory et al., 2018, respectively), with an average value for this family of  $2.16 \pm 1.58 \mu\text{g/g dw}$ . Overall, these data show a high inter-species variability (see Table 1, Figs. 3 and 4). Even if alcids have the highest mean blood Hg concentration, too few values ( $n=4$ ) were reported to confirm that they are significantly more contaminated than the two other families. In addition, such variations in blood Hg concentrations among taxa are in accordance with the feeding ecology of these species. Indeed, anatids feed at low trophic levels (i.e. mostly on molluscs and crustaceans) during the breeding season. Guillemots feed at intermediate levels (mostly on planktivorous fish species) (Gaston and Jones, 1998) while larids occupy the highest trophic levels (see Braune et al., 2016; Burgess et al., 2013; Hobson et al., 2002, 1994 for a review) (Fig. 4). It

should also be noted that all families reviewed in this study showed blood Hg concentrations below toxicity thresholds associated with substantial impairments to health and reproduction ( $2 \mu\text{g/g ww}$ , equivalent to approximately  $9.6 \mu\text{g/g dw}$ ; Ackerman et al., 2016). These results also highlight spatial variations through the Arctic, with higher levels found in the Canadian Arctic and Alaska than in East Greenland or Svalbard, and thus confirms previous investigations using eggs which found higher values in the Canadian Arctic (Arctic Monitoring and Assessment Programme (AMAP), 2018; Provencher et al., 2014). They are also in agreement with previous investigations performed on internal tissues where higher Hg concentrations were found in the liver of Canadian Arctic seabirds (Northern Baffin Bay) compared to those from the European Arctic (Central Barents Sea) (Borgå et al., 2006). Although more analyses are needed to confirm this spatial trend by including various missing Arctic regions (e.g. Russia, Northwest Greenland, Aleutian Islands, Alaska), these observations provide additional elements towards the understanding of Hg spatial distribution in the Arctic.

Finally, Braune et al. (2007) adjusted Hg concentrations by trophic status in seabird eggs ( $\delta^{15}\text{N}$ ) to take into account any change in their trophic position and, hence, to study temporal changes in Hg contamination. This method was subsequently used on different species and sites (Braune et al., 2016; Burgess et al., 2013). Blood, as eggs, represents relatively local (in the vicinity of the colony) and short-term dietary exposure. Blood Hg concentrations strongly depend on seabird diet and trophic position (i.e. biomagnification process), which can vary between individuals, populations and species. Hence, and similarly to Braune and colleagues' method, we propose that the adjusted Hg concentrations in blood could be used to monitor Hg temporal and spatial trends not only in seabirds but also in the Arctic marine food chains.

### **3.3. Feathers as a tool to monitor Hg contamination at different seasons**

#### **3.3.1. Hg excretion to feathers, molting patterns and feathers to monitor Hg contamination through the Arctic**

Approximately 70–90% of the Hg available in the bloodstream and internal tissues is excreted into the plumage during molt (Agusa et al., 2005; Braune, 1987; Honda et al., 1986). Hg binds to disulphide bridges of feather keratins (Crewther et al., 1965) and becomes stable once the feather is grown (Appelquist et al., 1984). Hence, Hg concentrations in feathers represent the Hg accumulated between two molts and inform about the Hg body burden at the molting time (Agusa et al., 2005; Furness et al., 1986). As a consequence, Hg concentrations in feathers are not affected by the age of the individual (Honda et al., 1986; Bustamante et al., 2016). As they can be non-lethally sampled, feathers have also been proposed as a good proxy to study seabird exposure to Hg (Monteiro and Furness, 1995). The Hg excreted into feathers is mainly under its organic and toxic MeHg form (> 80%) (Bond and Diamond, 2009; Renedo et al., 2017; Thompson and Furness, 1989). This means that measuring total Hg concentrations in seabird feather samples provides a good information about their contamination and exposure to MeHg. Similar to blood, feathers can be collected every year on the same individuals (Bearhop et al., 2000), hence allowing long-term assessment of individual Hg exposure. In addition, depending on the species, molting patterns and the type of feather considered, feathers can inform about Hg exposure during different periods of the year (Lavoie et al., 2014; Fort et al., 2014). Some Arctic seabirds molt once a year only, i.e. post or pre-breeding molt, as is the case for female anatids or ivory gulls, respectively. Some other species (most of the alcids, larids, and male anatids) molt twice a year (see the complete review of molting patterns in Annex 1). As an example, alcids have a first partial molt occurring before the breeding period and leading to a nuptial plumage, when only head and neck feathers are replaced. A second complete molt occurs after the breeding period during

which all feathers, including body, head and flight feathers, are renewed. Conversely, the northern gannet (*Morus bassanus*) is the only species breeding in the Arctic that presents a continuous molt of body feathers (see Annex 1), which leads to a more complex interpretation of the feather Hg concentrations (Cherel et al., 2018).

Three different types of feather are commonly used in Hg assessments: flight (primary and secondary) feathers, body feathers and head feathers. Head feathers of most alcid and larid species (Annex 1) provide information about Hg contamination of individual seabirds specifically during the non-breeding period (Fort et al., 2014, Fig. 5). However, their use has been largely neglected to date. Primary and body feathers inform about seabird Hg contamination either over the breeding season for species molting twice a year or over a full year for those molting once a year (Fig. 5). Nevertheless, primary feathers show a much higher inter-feather variability of Hg concentrations compared to body feathers. Feathers are sequentially molted from primary 1 (P1) to primary 10 (P10), with higher Hg quantities excreted into P1 and then decreasing from P2 to P10 (Furness et al., 1986) (Fig. 5). Conversely, body feathers are molted at the same period and show lower variability (Brasso et al., 2013; Carravieri et al., 2014; Furness et al., 1986). Therefore, if primary feathers are chosen to investigate seabird exposure to Hg, they must be carefully selected in order to make Hg measurements comparable between the different individuals. In addition, some species also have a sequential molt of their body and flight feathers over several months or an unknown molting pattern. The use of these species for feather Hg assessment is thus more complicated. This is the case of the Atlantic puffin (*Fratercula arctica*), the whiskered auklet (*Aethia pygmaea*), the Leach's storm-petrel (*Oceanodroma leucorhoa*) and the Fork-tailed storm-petrel (*Oceanodroma furcata*), the pelagic cormorant (*Stercorarius longicaudus*) and the red-faced cormorant (*Stercorarius pomarinus*) (Gaston and Jones, 1998; Harris, 1974; Pyle, 2016; Toochin and Fenneman, 2014; Van Tets, 1959). For instance, the Atlantic puffin,

as most of the alcids, undergoes two molts per year but this species presents unsynchronized post-breeding molts between individuals (Gaston and Jones, 1998; Harris and Yule, 1977).

### **3.3.2. Hg contamination of Arctic seabirds during the breeding season or through the entire year**

At the Arctic scale, Hg concentrations measured in body feathers significantly differed between bird families (ANOVA,  $F_{2,55}=10.19$ ,  $p < 0.001$ ), with the lowest mean Hg concentrations measured in anatids ( $0.92 \pm 0.19 \mu\text{g/g dw}$ ), followed by alcids ( $2.20 \pm 1.23 \mu\text{g/g dw}$ ) and larids ( $3.53 \pm 1.74 \mu\text{g/g dw}$ ).

More specifically, while Hg concentrations measured in female anatids (for which body feathers represent yearly local exposure) did not exceed  $2 \mu\text{g/g dw}$  and showed little intra- and interspecific variations, alcids (for which body feathers represent contamination during breeding period only), reached concentrations of  $7 \mu\text{g/g dw}$  and showed higher interspecific variations (See Table 2, Figs. 6 and 7). As observed for blood, these results are consistent with the knowledge of species' diet and trophic position. For instance, little auks ( $1.00 \pm 0.22$  to  $2.11 \pm 1.49 \mu\text{g/g dw}$ ; Fort et al., 2016; Amélineau et al., 2019) feed mostly on zooplankton, while tufted puffins ( $2.48 \pm 0.25$  to  $2.59 \pm 0.29 \mu\text{g/g dw}$ ; Burger and Gochfeld, 2009) and Kittlitz's murrelets (from  $1.66 \pm 1.28$  (Kaler et al., 2014) to  $5.15 \pm 2.51 \mu\text{g/g dw}$  (Kenney et al., 2018)) feed on a mixed diet of fish and invertebrates (Gaston and Jones, 1998; Hobson et al., 1994). However, these differences might also be explained by the spatial variations of Hg contamination of their prey. For instance, common ( $0.88 \pm 0.19 \mu\text{g/g dw}$ ; Wenzel and Gabrielsen, 1995), Brünnich's ( $0.78 \pm 0.18 \mu\text{g/g dw}$ ; Wenzel and Gabrielsen, 1995), and pigeon guillemots ( $7.72 \pm 0.84 \mu\text{g/g dw}$ ; Burger and Gochfeld, 2009) are all mostly piscivorous (Gaston, 2000; Hobson et al., 1994) species during the breeding season. However, Hg concentrations measured in pigeon guillemots and Kittlitz's murrelets were about six and

ten times higher than in common and Brünnich's guillemots, reflecting a hotspot of Hg contamination in the Aleutian Islands (Ackerman et al., 2016), where tufted puffins were also sampled. Those Hg concentrations in feathers are in accordance with previous works showing that pigeon guillemots breeding in the Pacific Ocean (Western North America) are facing potential toxicological risk linked to higher Hg concentrations in this region (Ackerman et al., 2016; Arctic Monitoring and Assessment Programme (AMAP), 2018).

The larid family showed the highest Hg concentrations in body feathers as well as the highest intra- and interspecific variations, with values ranging from  $0.89 \pm 0.15 \mu\text{g/g dw}$  (Arctic tern, Kim et al., 1996a, 1996b) to  $11.66 \pm 6.52 \mu\text{g/g dw}$  (ivory gull, Mallory et al., 2015) (See Table 2, Figs. 6 and 7). Interestingly, the lowest values were measured in Arctic terns breeding in Russia, while this species is reaching toxicological concerns in the Canadian Arctic, where it faces both high Hg concentrations (in both eggs and liver) and decreasing population trends (Provencher et al., 2014, Akearok et al., 2010). In addition, the highest Hg concentrations were measured in ivory gulls breeding in the Canadian Arctic, while the lowest values for this species were measured in birds from Svalbard. Again, these results are consistent with Hg spatial trends highlighted from the study of other seabird tissues (Provencher et al., 2014) or other species (Arctic Monitoring and Assessment Programme (AMAP), 2018), showing that the Canadian Arctic marine biota presents higher Hg contamination compared to other Arctic regions. High Hg concentrations measured in black-legged kittiwakes, ivory gulls, glaucous, glaucous-winged and herring gulls are in accordance with their mostly piscivorous diet (opportunistic diet, from young to older fish, see Braune et al., 2016; Burgess et al., 2013; Hobson et al., 2002, 1994 for a review) and with previous studies focused on bird eggs and liver (Provencher et al., 2014).

For the other seabird families, data are much more limited. Hg concentrations in body feathers were investigated only once in hydrobatids, including the study of fork-tailed storm-petrels

breeding in the Aleutian Islands, for which the molting pattern is unclear (see Annex 1, but see Harris, 1974) and exhibited the highest Hg concentrations ( $6.70 \pm 1.63 \mu\text{g/g dw}$ ; Kaler et al., 2014), close to concentrations observed in some larids. This species mostly feeds on relatively low trophic level prey (amphipods, small fish or cephalopods - (Hobson et al., 1994)), and like pigeon guillemots and Kittlitz's murrelet, might be exposed to higher Hg concentrations in the Aleutian Islands. Similarly, stercorariids were barely investigated, with feather Hg measurements in long-tailed and parasitic jaegers breeding in Russia only. As these species have two molts per year, we suggest that body feathers represent Hg exposure during the breeding period, as for most of alcid and larid species. Long-tailed and parasitic jaegers showed Hg concentrations of  $1.95 \pm 0.54$  and  $1.84 \mu\text{g/g dw}$ , respectively (Kim et al., 1996a, 1996b). The former species preferentially feeds on lemmings during the breeding season (see Gilg et al., 2013 for a review) while the latter mostly consumes robbed fishes caught by other seabirds (Davis et al., 2005). Their Hg concentrations, close to the ones of Sabine's gulls (also having two molts per year – see Annex 1) from the same region and also feeding on small fish, suggests that these jaeger populations also feed on intermediate trophic levels (see Figs. 6 and 7).

Eisler (1987) (see also Burger and Gochfeld, 1997) proposed a toxicity threshold in feathers of  $5 \mu\text{g/g dw}$ . Here, all of the anatids were below this limit. In the case of alcids and larids however, bird vulnerability depended on their breeding regions. Brünnich's and common guillemots from Norway, little auks from East Greenland and tufted puffin from Aleutian Islands were below the toxicity threshold. In Kittlitz's murrelets, most of the measured Hg concentrations were below the threshold, except for the population breeding in the Adak Island. For the pigeon guillemot, the Alaska population was below the threshold, while all the colonies from the Aleutian Islands were above. For larids, Hg concentrations of black-legged kittiwakes, Arctic tern, glaucous-winged and Sabine's gull were not of concern, while Hg



concentrations of herring and glaucous gulls from Russia, and ivory gull from the Canadian Arctic, were exceeding the threshold. Finally, the fork-tailed storm-petrel from the Aleutian Islands showed Hg concentrations above this toxicity threshold. Overall, this suggests that some species and/or areas need to be carefully monitored to follow any trend in Hg contaminations and therefore permanent toxicity risks for these populations.

Flight feathers provide complementary information to other feathers, but have only been used in three studies within the Arctic (Kenney et al., 2018; Kim et al., 1996b; Mallory et al., 2015). While one of these studies sampled primary 1 or primary 3 only (Mallory et al., 2015), another focused on primary 5 (Kim et al., 1996b), and secondary feathers (Kenney et al., 2018). This makes comparisons difficult as those feathers do not molt at the same time. Nevertheless, a similar trophic/spatial pattern than in body feathers can be observed in primary feathers. The lowest Hg concentrations were found in anatids ( $0.59 \pm 0.21 \mu\text{g/g dw}$  and  $0.70 \pm 0.20 \mu\text{g/g dw}$  respectively for common eiders of the Canadian Arctic; Mallory et al., 2015; and for long-tailed ducks in Russia; Kim et al., 1996a, 1996b) and the highest concentrations were observed in larids, with ivory gulls being the most contaminated species ( $15.79 \pm 14.13 \mu\text{g/g dw}$ ; Mallory et al., 2015). Brünnich's guillemots and northern fulmars had intermediate Hg concentrations ( $1.94 \pm 0.63$  and  $2.71 \pm 0.72 \mu\text{g/g dw}$ , respectively; Mallory et al., 2015). Secondary feathers have only been used for the Kittlitz's murrelet breeding on the Aleutian Islands, with the highest Hg concentration being  $37.18 \mu\text{g/g dw}$  (Kenney et al., 2018). This unique value is in agreement with the Hg concentrations measured in Kittlitz's murrelet body feathers from the Aleutian Islands. However, as this type of feather has not been used in any other study, the comparison of its Hg concentration with other feathers remains uncertain.

#### **4. Hg contamination of Arctic seabirds during the non-breeding period**

As illustrated above through blood and body feather measurements, the breeding time is the most studied period in Arctic seabird Hg monitoring. However, investigations covering their entire annual cycle, and thus both the breeding and non-breeding periods, are important in order to have a complete understanding of seabird exposure to Hg and of the associated risks, both in the Arctic and outside of the Arctic. Indeed, most seabirds breeding in the Arctic migrate southwards to spend the non-breeding time outside the Arctic (Egevang et al., 2010; Gilg et al., 2013). Previous studies used bird carcasses legally shot or beached along the coasts (Bond et al., 2015; Fort et al., 2015) to get information for this specific period. However, for obvious ethical reasons or because stranded dead birds could show a bias in their Hg concentrations (e.g. starvation might affect Hg concentrations of internal tissues, Fort et al., 2015), an alternative solution is required. Seabird bycatch could provide opportunities to collect unbiased bird carcasses but would restrict samples to diving and piscivorous species and has never been used or validated until now. In that context, analysis of feathers appears as a useful and non-lethal approach to obtain information about bird non-breeding contamination, by the use of 1) head feathers for species molting twice a year, or 2) body feather for species molting once a year only. In this latter case, body feathers integrate information about Hg accumulated over the year and extraction of the non-breeding information should be performed. As proposed for the clapper rails (*Rallus longirostris obsoletus*) by Ackerman et al. (2012), it could be relevant to use published equations to predict Hg concentrations in different seabird tissues (i.e. head and body feathers, blood) and thus obtain information for specific seasons.

Head feathers have been used in only three studies focused on little auks from East Greenland to assess Hg contamination during the nonbreeding period (Table 4), with values ranging from  $2.27 \pm 0.41$  to  $3.99 \pm 1.96$   $\mu\text{g/g dw}$  (Amélineau et al., 2019; Fort et al., 2016, 2014).

These studies showed that non-breeding (head feathers) concentrations were significantly higher than breeding (body feathers) ones and suggested a higher exposure to Hg during the non-breeding period (Fort et al., 2014). These values are below the 5 µg/g toxicity threshold proposed by Eisler (1987). A long-term study of the same population did not find a link between Hg concentrations and adult survival (Amélineau et al., 2019). Even though this method has been scarcely used in the Arctic, other studies combining carbon and nitrogen stable isotope analyses together with Hg measurements (both in feathers and/or in blood), demonstrated important seasonal variations in Hg contamination in Double-crested Cormorants (*Phalacrocorax auratus*) (Lavoie et al., 2014; Ofukany et al., 2012), Caspian terns (*Hydroprogne caspia*) (Lavoie et al., 2014) and common terns (*Sterna hirundo*) (Nisbet et al., 2002). This highlights the importance to further study such seasonal variations in Hg exposure, which could be performed by the analysis of feathers.

## **5. Chicks as relevant stage for assessing Hg local contamination**

Chicks, through measurements in blood, down and fully-grown feathers, can provide valuable information about pre-breeding and local Hg contamination (see Bond and Diamond, 2010 for the eggs). In the case of capital breeders (i.e. allocation of stored nutrients to egg formation) (Akearok et al., 2010), Hg accumulated by females during the late migration and egg formation is transferred to the eggs (Burger et al., 2008b). In the case of income breeders (i.e. allocation of local nutrient to egg formation), eggs represent female dietary Hg intakes during the egg formation only (Bond and Diamond, 2010). Several species (e.g. Atlantic puffin or razorbill) have an intermediate strategy (Bond and Diamond, 2010). As the down is grown into the egg before its hatch, Hg concentrations in eggs and chick down are correlated (Furness, 1997; Stewart et al., 1997). Hence, down was proposed as an alternative to egg collection (Monteiro and Furness, 1995), as it can be less destructive to assess Hg

contamination in declining species (such as the ivory gull). For conservation purpose, eggshells could also be a useful non-lethal tissue to monitor Hg concentrations (Peterson et al., 2017).

During chick growth, the down is replaced by feathers, which thus integrate Hg from the chick diet. Therefore, chick feathers are useful indicators of the local Hg contamination during the chick rearing period (Blévin et al., 2013; Furness, 1997). Indeed, most adults feed their chicks with relatively local food (except in a few species such as for instance northern fulmars which can travel several hundreds of kilometers to find their prey (Weimerskirch et al., 2001)). Unlike eggs, chick feathers and blood have been scarcely studied in the Arctic (Kaler et al., 2014; Tartu et al., 2016; Wenzel and Gabrielsen, 1995). Chick feathers constitute a good alternative to adult feathers to assess Hg local contamination (Blévin et al., 2013; Monteiro and Furness, 1995), with the advantage to integrate, in most cases, the local Hg contamination only (see above) while adult body feathers can integrate Hg exposure over much larger geographical scales, especially in species molting once a year. Finally, chick blood informs about recent intakes and local contamination only and presents the same advantages than fully-grown feathers. However, collecting blood in growing chicks could be more challenging and stressful for the bird than collecting feathers or down.

## **6. Future direction: spatial ecotoxicology to link seabird**

Blood and feathers thus appear as important tissues for the monitoring of seabird Hg contamination and its temporal variations along their whole biological cycle. However, to further comprehend the vulnerability of these species to Hg marine contamination, it is essential to understand where individuals, populations or species get contaminated. Indeed, seabirds can use contrasting feeding areas, both at small spatial scale over a season and large scale when migrating thousands of kilometers between their breeding, molting and wintering

grounds. Therefore, they potentially experience various levels of Hg environmental contamination (e.g. Leat et al., 2013). Different tools exist to track movements in Arctic seabirds. For instance, miniaturized electronic devices (e.g. GLS, GPS, Argos) are used to track large scale movements of birds over days, months or years and can be combined to Hg measurements in blood and feathers to highlight the role played by specific regions on their contamination (Fort et al., 2014).

Furthermore, Hg isotopes have been widely used in the last decades as a new and powerful perspective to explore the cycle of this element (Blum et al., 2014) and could thus be used to track the trophic and spatial origins of Hg in polar seabirds (Renedo et al., 2018). Hg has seven naturally occurring stable isotopes that undergo both mass-dependent fractionation (MDF) and mass-independent fractionation (MIF), both related to distinct Hg isotopic ratios. Hg MDF (normally reported as  $\delta^{202}\text{Hg}$ ) is induced during physicochemical reactions including environmental processes such as volatilization, redox transformations, methylation and demethylation, photochemical reactions and biological processes (Kritee et al., 2009, 2007; Perrot et al., 2015; Rodríguez-González et al., 2009; Zheng et al., 2007). Conversely, Hg MIF is essentially attributed to the photochemical reduction of inorganic Hg and photodemethylation of MeHg in aquatic systems before its incorporation into the food web (Bergquist and Blum, 2007; Zheng and Hintelmann, 2010). These two processes cause a significant “anomaly” in the fractionation of odd Hg isotopes (reported as  $\Delta^{199}\text{Hg}$  and  $\Delta^{201}\text{Hg}$ ) which represents the deviation from the theoretical MDF. Further, as Hg MIF is not produced during biological processes (Kritee et al., 2009, 2007), its signature is conserved along the food webs. Therefore, the measurement of Hg MIF magnitudes and ratios in the tissues of top predators permits to trace Hg photochemical conditions in their foraging habitats (Masbou et al., 2018; Perrot et al., 2016; Renedo et al., 2018). Since seabirds forage in diverse locations at sea, both in the horizontal and vertical scales, their use for Hg isotopic investigations

present important advantages for the investigation of Hg sources and its associated geochemical processes in the ocean (Renedo et al., 2018). In the Arctic, Point et al. (2011) were pioneers on the use of avian samples (eggs) as bioindicators for Hg isotopic analyses. These authors revealed the influence of ice cover on Hg marine photochemistry by the differences of Hg odd-MIF signatures between ice-covered and non-ice-covered oceanic areas of the Alaskan Arctic Ocean, using Brünnich's and common guillemot as bioindicators. Likewise, another study on seabird eggs from these species identified variations of Hg isotopic signatures between coastal and open ocean environments attributed to both different sources and processes (photochemistry) between these two reservoirs in the Alaskan region (Day et al., 2012). In the Southern Ocean, Hg isotopic signatures in blood samples of subantarctic adult penguins also permitted to illustrate different MeHg biogeochemical characteristics across inshore-offshore marine compartments depending on specific foraging behaviour variability within a same ecosystem (Renedo et al., 2018). Significant latitudinal variations of Hg isotopic signatures have also been observed in seabird tissues across distant colonies of the Southern Ocean (Renedo et al., submitted). These observations indicate that both specific foraging habitats and latitudinal movements of seabirds highly determine their exposure to distinct environmental MeHg sources in variable marine ecosystems.

Hg isotopic composition in seabirds has demonstrated to be a potent tool for the definition of MeHg sources and to highlight the variability of processes among the different marine compartments. Therefore, the use of Hg isotopic approaches in migratory seabirds, feather Hg isotopic signatures will reflect the integration of Hg from diverse sources during an entire year for species molting once a year. This integrative effect of adult feathers could be interesting for tracking whole annual Hg contamination, i.e. including the wintering areas. Furthermore, analyses of different type of feather of species moulting twice a year such as the case of alcids, appears as a suitable approach to obtain information about both the breeding

and non-breeding Hg contamination. Measurements of Hg isotopic composition in feathers of flying migratory seabirds could then open a new horizon for exploring Hg exposure pathways at their breeding sites in summer but also in large zones of the marine environment during their non-breeding period outside the Arctic. Finally, stable isotopes of carbon and nitrogen (respectively indicators of foraging habitat and trophic level), together with the use of Hg isotopes (proxies of Hg biogeochemical characteristics in each marine compartment), would enable to help elucidating the Hg exposure pathways of migratory seabirds during their whole annual cycle, both during their breeding and non-breeding periods.

## **7. Limitations**

Although highly valuable to inform about seabird Hg contamination through the Arctic, the use of blood and feathers also shows some limitations. Even if seabirds have been used in ecotoxicology since decades, detailed information about their trophic and behavioral ecology is essential to properly interpret their exposure to the environmental contamination. Such information is nonetheless often missing or ignored. For instance, information about seabird diet and molting patterns, mostly during the non-breeding period, are still scarce in many species. The present paper reviewed this existing knowledge of Arctic seabird molting patterns (pre- and/or post-breeding) (Annex 1) and should help future investigations using feathers to track seabird Hg contamination. Furthermore, Hg measurements are not always performed under the same protocols (e.g. different sample preparations leading to contrasting units (dry vs wet weights)). Although correction factors exist, they add uncertainties in the data and complicate comparisons. Such potential bias could be easily avoided by the use of homogenized methods, ensuring comparability of Hg measurements between areas and species. Finally, spatial ecotoxicology approaches combining biologging with Hg analyses should be developed to track the spatial origin of Hg contamination. However, large scale

movements such as migration could also be associated to changes in diet, possibly affecting Hg intakes. Although nitrogen stable isotopes can be used to investigate seabird trophic status, isotopic values cannot be directly compared to Hg measurements in feathers as they do not provide information for the same periods (Bond, 2010). Therefore, feathers cannot be corrected to take into account potential diet effects as we can do for blood samples.

## **8. Conclusion**

Although largely neglected in comparison to internal tissues or eggs, feathers and blood are two complementary and important tissues allowing large spatio-temporal investigations of Hg contamination in Arctic seabirds during both the breeding and non-breeding periods. They also provide information about risks encountered by seabirds depending on their Arctic breeding site, their migratory movements and their wintering grounds, within or outside the Arctic. Further studies monitoring long-term and large scale Hg contamination and potential subsequent risks for Arctic seabirds are now necessary, and require large international collaborations and monitoring programmes such as those implemented in Canada (Braune et al., 2001, Northern Contaminants Program (Chételat and Braune, 2012)), Norway (SEAPOP (Anker-Nilssen et al., 2017)) or at the pan-Arctic scale (newly established ARCTOX program). Such initiatives are indeed essential in order to determine temporal and spatial trends, to define baseline and hotspots of Hg contamination, and to allow international actions (see the Minamata Convention, the Arctic Monitoring Assessment Program or the Global Mercury Assessment).



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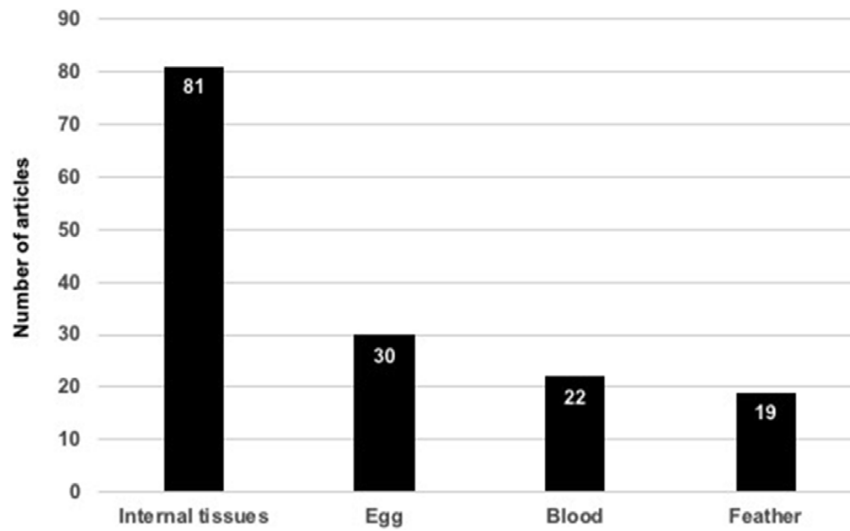
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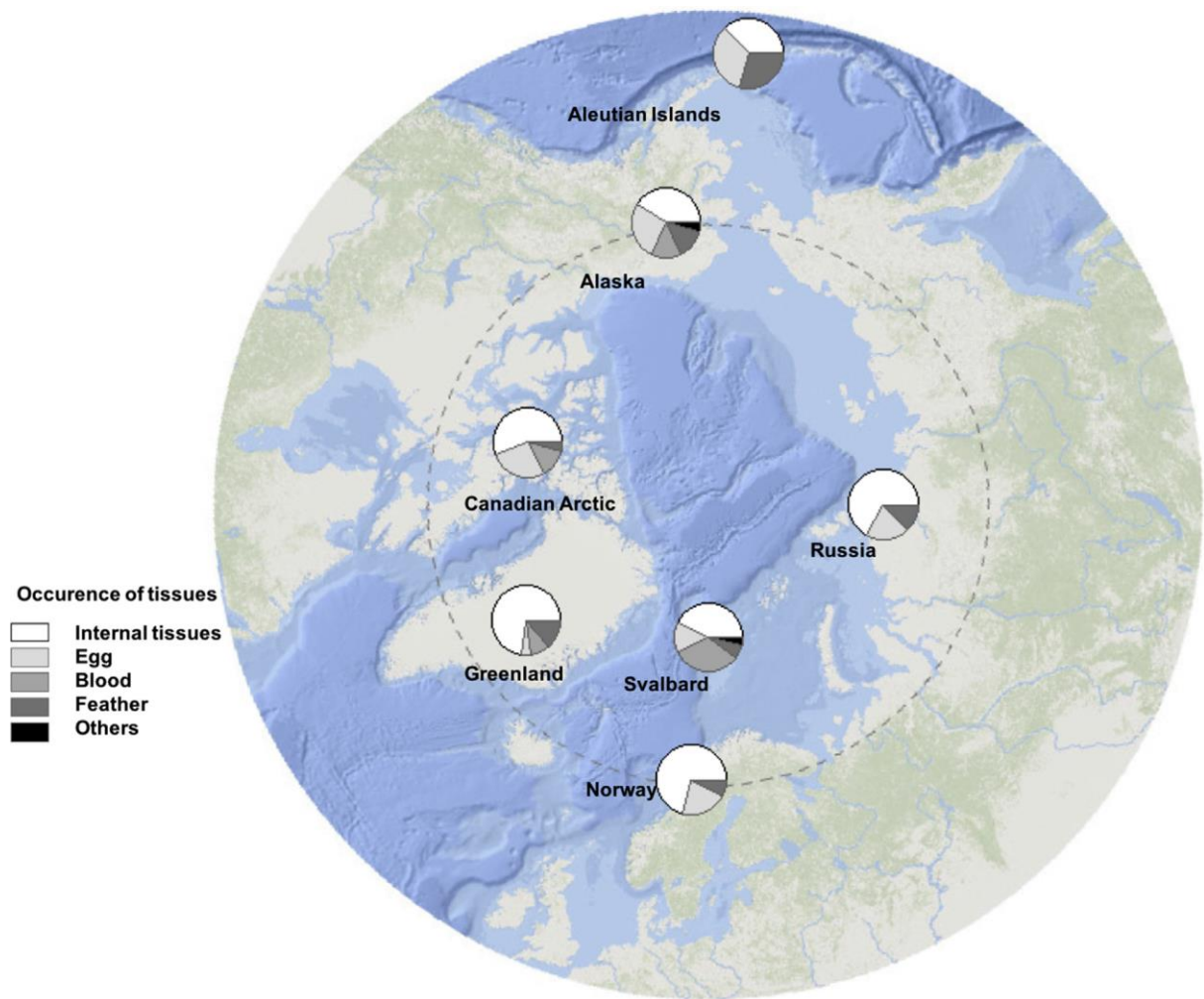
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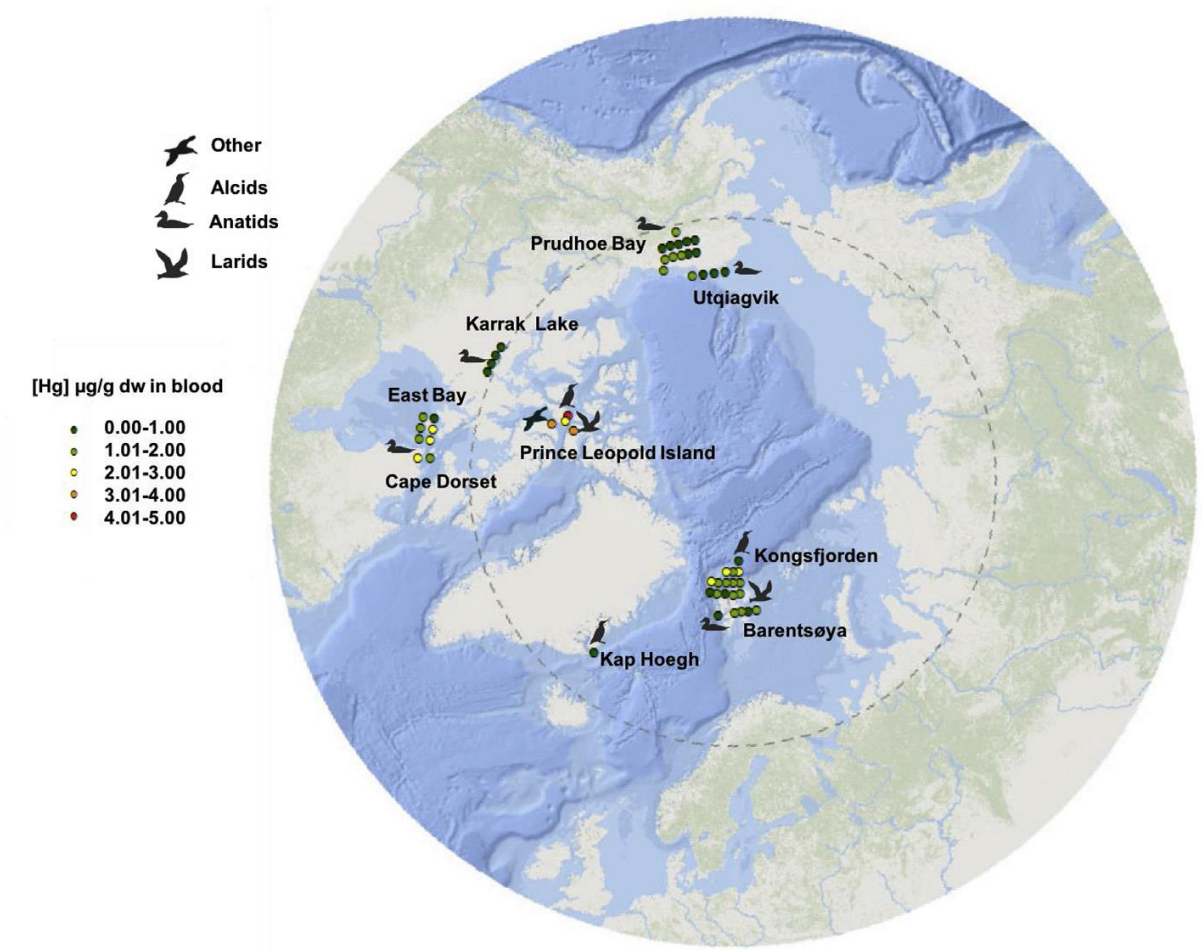
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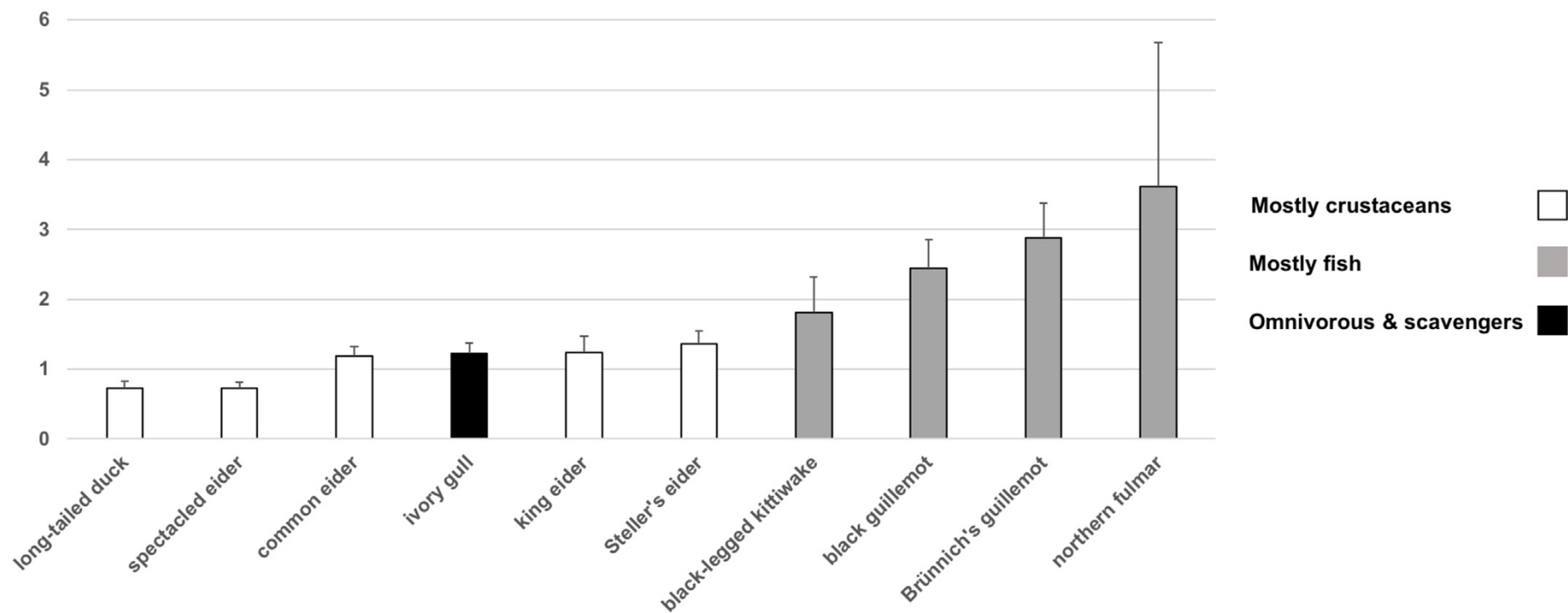
**Fig. 1.** Occurrence of tissues analysed for Hg in Arctic seabirds in the 101 reviewed articles. Some publications studied several tissues. Internal tissues comprise liver (n=38), kidneys (n=10), muscle (n=25), heart (n=1), bones (n=1), brain (n=4), gonads (n=1) and lungs (n=1). Feathers comprise body feathers (n=14), head feathers (n=3) and flight feathers (n=4).



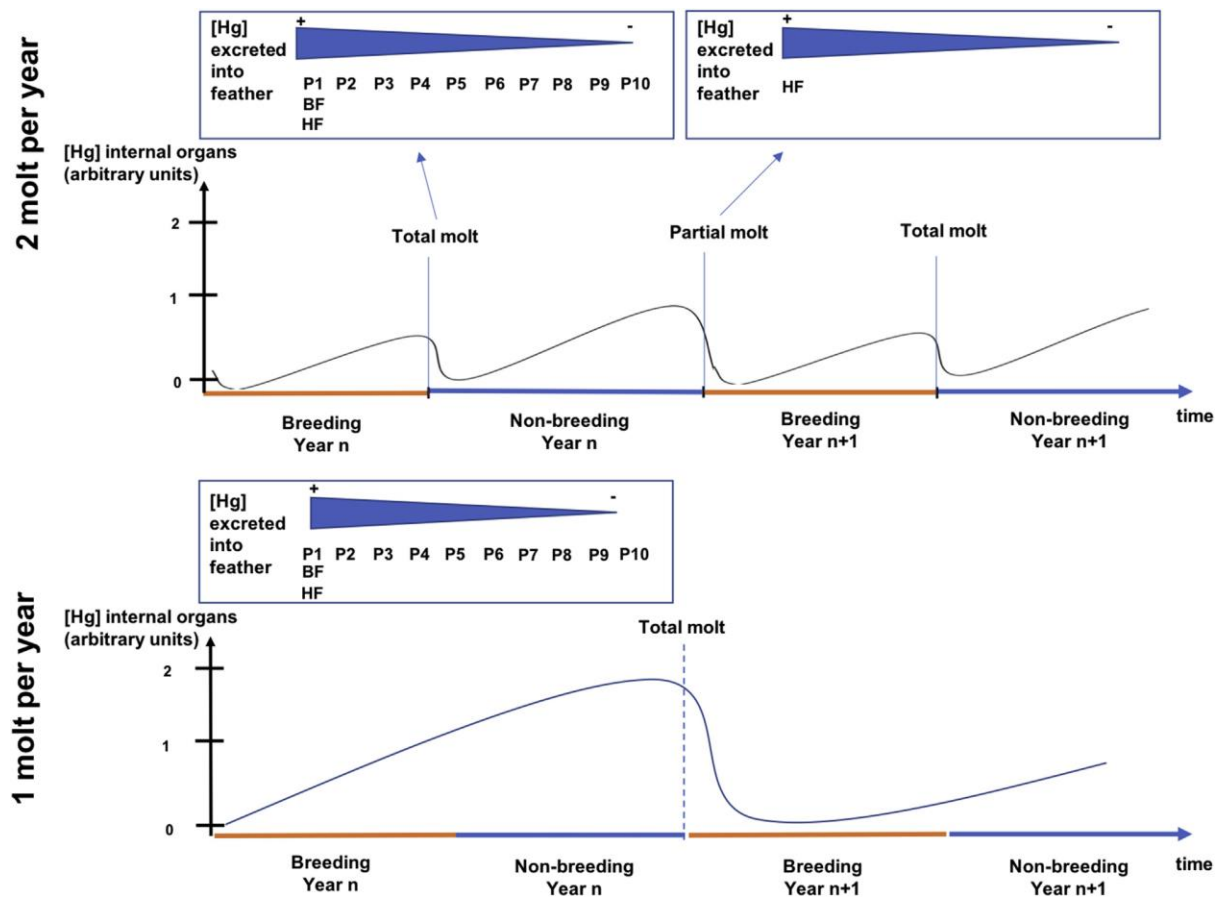
**Fig. 2.** Occurrence of tissues analysed for Hg in Arctic seabirds in the 101 reviewed articles, by Arctic region. Some publications studied several tissues. Internal tissues comprise liver (n=38), kidneys (n=10), muscle (n=25), heart (n=1), bones (n=1), brain (n=4), gonads (n=1) and lungs (n=1). Others (excrements, n=2). Feathers comprise body feathers (n=14), head feathers (n=3) and flight feathers (n=4).



**Fig. 3.** Spatial distribution of Hg concentrations measured in blood.

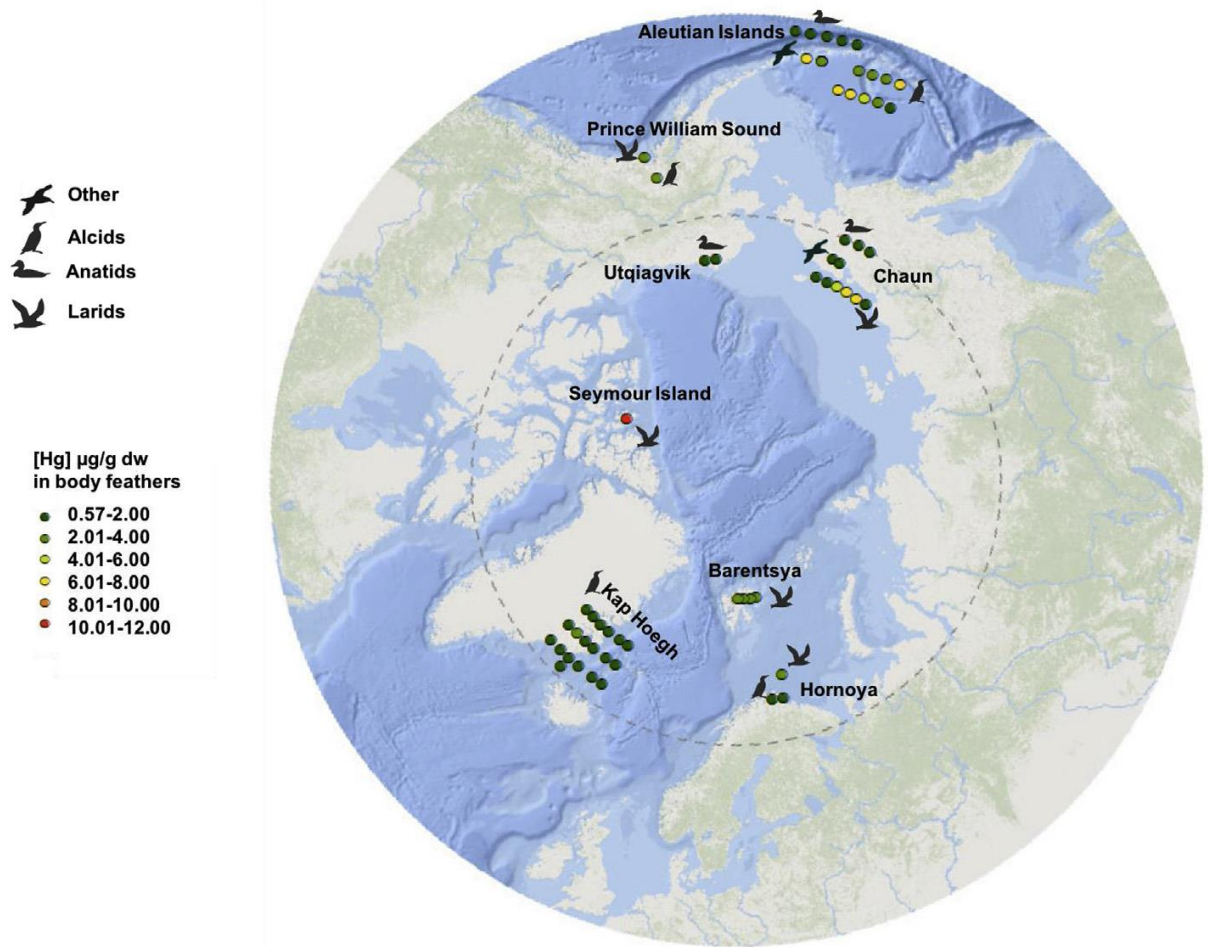


**Fig. 4.** Mean Hg concentrations (mean  $\pm$  SD  $\mu\text{g/g dw}$ ) measured in blood by species, in relation to bird diet during the breeding season.

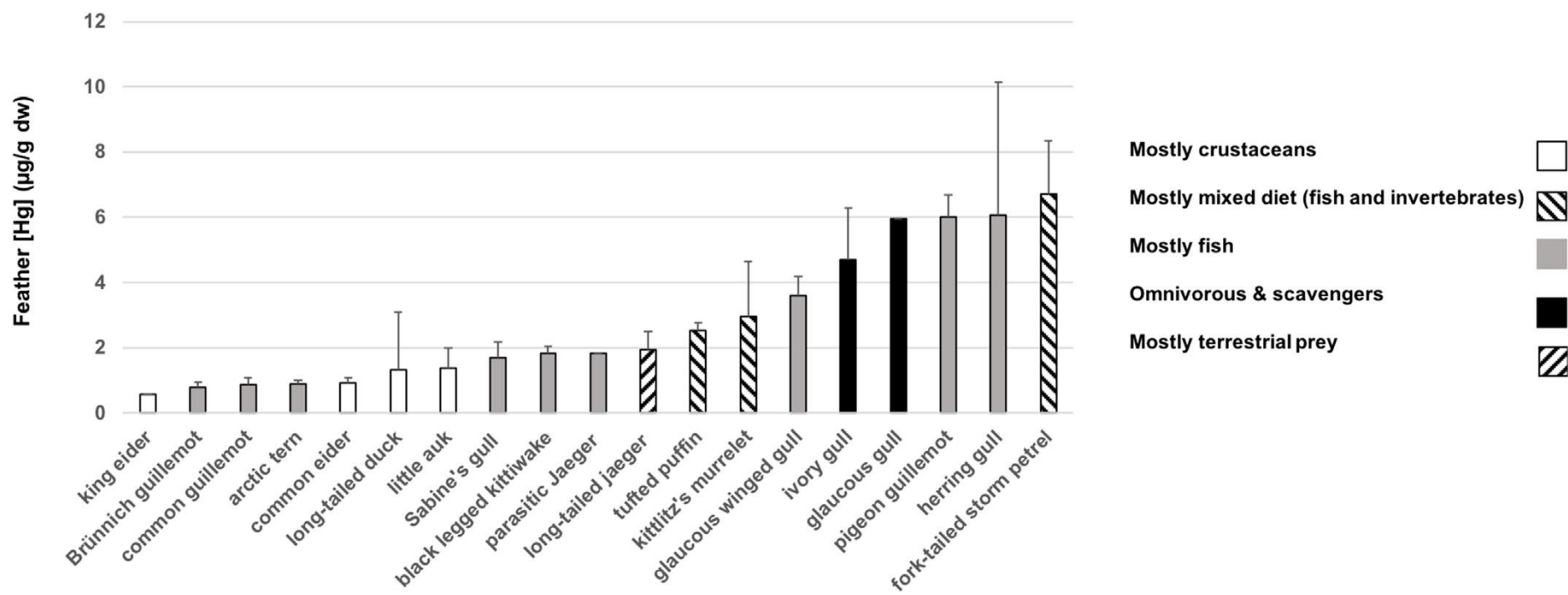


**Fig. 5.** Schematic representation of Hg excretion from internal tissues and organs to body, head and primary feathers for seabirds with one and two molts per year.





**Fig. 6.** Spatial distribution of Hg concentrations in body feathers.



**Fig. 7.** Mean Hg concentrations (mean  $\pm$  SD  $\mu\text{g/g dw}$ ) measured in body feathers, by species, in relation to bird diet during the breeding season.

**Table 1.** Total Hg concentrations measured in blood samples (whole blood  $\mu\text{g/g}$  dw mean  $\pm$  SD or SE; red blood cell  $\mu\text{g/g}$  dw mean  $\pm$  SD; sample size is given into bracket) for seabirds breeding in the Arctic. If sexes were differentiated, F for female or M for male is added. \* wet weight converted to dry weight following a moisture content of 79.13% (Eagles-Smith et al., 2008), a nmol/g to  $\mu\text{g/g}$ , b overall sample size=105.

Family	Species	Sampling year	Colony location	Country	Whole blood	Red blood cell	References	
Alcid	Mandt's black guillemot ( <i>Cepphus grylle mandtii</i> )	2015	Kongsfjorden	Svalbard		0.32 $\pm$ 0.80 (10)	Eckbo et al. (2019)	
	Black guillemot ( <i>Cepphus grylle</i> )	NA	Prince Leopold Island	Canada	4.58 (1)		Mallory et al. (2018)	
	Brünnich guillemot ( <i>Uria lomvia</i> )	NA	Prince Leopold Island	Canada	2.88 $\pm$ 0.50 (SD) (6)		Mallory et al. (2018)	
Anatid	Little auk ( <i>Alle alle</i> )	2010	Kap Hoegh	Greenland	0.84 $\pm$ 0.20 (SD) (35)		Fort et al. (2014)	
	Common eider ( <i>Somateria mollissima</i> )	2014	Kaktovik	Alaska	1.13 $\pm$ 0.04 (SD) F (20)		Miller et al. (2019)	
	Common eider ( <i>Somateria mollissima</i> )	NA	Cape Dorset	Canada	1.22 $\pm$ 0.33 (SD) (9)		Mallory et al. (2018)	
	Common eider ( <i>Somateria mollissima</i> )	2011	Storholmen, Kongsfjorden	Svalbard	0.57 $\pm$ 0.14 (SD) (26) F a;		Fenstad et al. (2017), 20:	
					0.30-0.94 (range) F (29) *			
	Common eider ( <i>Somateria mollissima</i> )	2013	East bay	Canada	2.06 $\pm$ 0.29 (SE) (11) F *		Provencher et al. (2017)	
	Common eider ( <i>Somateria mollissima</i> )	2014	East bay	Canada	2.25 $\pm$ 0.29 (SE) (10) F *		Provencher et al. (2017)	
	Common eider ( <i>Somateria mollissima</i> )	2013	East bay	Canada	0.91 $\pm$ 0.24 (SD) (98) F *		Provencher et al. (2016)	
	Common eider ( <i>Somateria mollissima</i> )	2014	East bay	Canada	1.05 $\pm$ 0.29 (SD) (92) F *		Provencher et al. (2016)	
	Common eider ( <i>Somateria mollissima</i> )	2000	Western Prudhoe Bay Oil Field	Alaska	0.87 $\pm$ 0.04 (SE) (20) F		Franson et al. (2004)	
	Common eider ( <i>Somateria mollissima</i> )	2000	Eastern Prudhoe Bay Oil Field	Alaska	0.73 $\pm$ 0.03 (SE) (20) F		Franson et al. (2004)	
	Common eider ( <i>Somateria mollissima</i> )	1997	East bay	Canada	1.10 $\pm$ 0.05 (SE) (11) F *		Wayland et al. (2001)	
	Common eider ( <i>Somateria mollissima</i> )	1998	East bay	Canada	1.10 $\pm$ 0.10 (SE) (15) F *		Wayland et al. (2001)	
	King eider ( <i>Somateria spectabilis</i> )	2010–2014	Utqiagvik	Alaska	0.64 $\pm$ 0.03 (SD) F (28)		Miller et al. (2019)	
	King eider ( <i>Somateria spectabilis</i> )	NA	Cape Dorset	Canada	2.04 $\pm$ 0.76 (SD) (6)		Mallory et al. (2018)	
	King eider ( <i>Somateria spectabilis</i> )	2001–2003	Karrak lake	Canada	0.81 $\pm$ 0.24 (SD) (147) F *		Wayland et al. (2008a, 2)	
	King eider ( <i>Somateria spectabilis</i> )	2001	Karrak lake & Adventure lake	Canada	0.65 (0.60–0.70) (SE) (63) F *		Wayland et al. (2008a, 2)	
	King eider ( <i>Somateria spectabilis</i> )	2002	Karrak lake & Adventure lake	Canada	0.90 (0.85–0.95) (SE) (69) F *		Wayland et al. (2008a, 2)	
	King eider ( <i>Somateria spectabilis</i> )	2003	Karrak lake & Adventure lake	Canada	0.85 (0.80–0.90) (SE) (74) F *		Wayland et al. (2008a, 2)	
	King eider ( <i>Somateria spectabilis</i> )	1996 (prebreeding)	Prudhoe Bay Oil Field	Alaska	1.44 $\pm$ 0.19 (SE) (7) M *		Wilson et al. (2004)	
	King eider ( <i>Somateria spectabilis</i> )	1996 (prebreeding)	Prudhoe Bay Oil Field	Alaska	1.05 $\pm$ 0.09 (SE) (4) F *		Wilson et al. (2004)	
	King eider ( <i>Somateria spectabilis</i> )	1996 (nesting)	Prudhoe Bay Oil Field	Alaska	1.48 $\pm$ 0.09(SE) (4) F *		Wilson et al. (2004)	
	Long-tailed duck ( <i>Clangula hyemalis</i> )	2010–2012; 2014	Utqiagvik	Alaska	0.95 $\pm$ 0.18 (SD) F (15)		Miller et al. (2019)	
	Long-tailed duck ( <i>Clangula hyemalis</i> )	2000	Western Prudhoe Bay Oil Field	Alaska	0.67 $\pm$ 0.06 (SE) (20) M		Franson et al. (2004)	
	Long-tailed duck ( <i>Clangula hyemalis</i> )	2000	Eastern Prudhoe Bay Oil Field	Alaska	0.57 $\pm$ 0.04 (SE) (20) M		Franson et al. (2004)	
	Spectacled eiders ( <i>Somateria fischeri</i> )	2010–2014	Utqiagvik	Alaska	0.49 $\pm$ 0.08 (SD) (42) F		Miller et al. (2019)	
	Spectacled eiders ( <i>Somateria fischeri</i> )	1995 (prebreeding)	Prudhoe Bay Oil Field	Alaska	0.96 $\pm$ 0.05 (SE) (14) M *		Wilson et al. (2004)	
	Spectacled eiders ( <i>Somateria fischeri</i> )	1995 (prebreeding)	Prudhoe Bay Oil Field	Alaska	0.57 $\pm$ 0.09 (SE) (6) F *		Wilson et al. (2004)	
	Spectacled eiders ( <i>Somateria fischeri</i> )	1995 (nesting)	Prudhoe Bay Oil Field	Alaska	0.72 $\pm$ 0.09 (SE) (9) F *		Wilson et al. (2004)	
	Spectacled eiders ( <i>Somateria fischeri</i> )	1995 (brood)	Prudhoe Bay Oil Field	Alaska	1.05 $\pm$ 0.14 (SE) (10) F *		Wilson et al. (2004)	
	Steller's eider ( <i>Polysticta stelleri</i> )	2008; 2010–2012; 2014	Utqiagvik	Alaska	1.36 $\pm$ 0.19 (SD) (36) F		Miller et al. (2019)	
	Larid	Black-legged kittiwakes ( <i>Rissa tridactyla</i> )	NA	Prince Leopold Island	Canada	3.04 $\pm$ 1.41 (SD) (9)		Mallory et al. (2018)
		Black-legged kittiwakes ( <i>Rissa tridactyla</i> )	2015	Kongsfjorden	Svalbard	2.00 $\pm$ 0.59 (SD) (20) M		Blévin et al. (2018)
Black-legged kittiwakes ( <i>Rissa tridactyla</i> )		2015	Kongsfjorden	Svalbard	1.43 $\pm$ 0.38 (SD) (20) F		Blévin et al. (2018)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> )		2012	Kongsfjorden	Svalbard		0.89 $\pm$ 0.05 (SE) (22) F	Blévin et al. (2017)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> )		2012	Kongsfjorden	Svalbard		1.14 $\pm$ 0.07 (SE) (22) M	Blévin et al. (2017)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> ) ADULT		2012 (chick rearing)	Kongsfjorden	Svalbard		0.90 $\pm$ 0.25 (SD) (48) F	Tartu et al. (2016)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> ) ADULT		2012 (incubation)	Kongsfjorden	Svalbard		1.25 $\pm$ 0.33 (SD) (48) F	Tartu et al. (2016)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> ) ADULT		2012 (chick rearing)	Kongsfjorden	Svalbard		1.13 $\pm$ 0.32 (SD) (44) M	Tartu et al. (2016)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> ) ADULT		2012 (incubation)	Kongsfjorden	Svalbard		1.82 $\pm$ 0.46 (SD) (44) M	Tartu et al. (2016)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> ) CHICK		2012	Kongsfjorden	Svalbard		NA	Tartu et al. (2016)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> ) ADULT		2013	Kongsfjorden	Svalbard		NA	Tartu et al. (2016)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> )		2012	Kongsfjorden	Svalbard		1.80 $\pm$ 0.45 (SD) (34) M	Tartu et al. (2015)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> )		2008	Kongsfjorden	Svalbard	2.06 $\pm$ 0.44 (SD) M b		Goutte et al. (2015)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> )		2009	Kongsfjorden	Svalbard	2.33 $\pm$ 0.55 (SD) M b		Goutte et al. (2015)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> )		2008	Kongsfjorden	Svalbard	1.97 $\pm$ 0.44 (SD) F b		Goutte et al. (2015)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> )		2009	Kongsfjorden	Svalbard	2.01 $\pm$ 0.41 (SD) F b		Goutte et al. (2015)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> )		2008	Kongsfjorden	Svalbard		NA	Tartu et al. (2013)	
Black-legged kittiwakes ( <i>Rissa tridactyla</i> )		2011	Kongsfjorden	Svalbard		0.60–3.30	Tartu et al. (2013)	
Ivory gull ( <i>Pagophila eburnea</i> )		April 2011	Barentsøya	Svalbard	1.68 $\pm$ 0.27 (SD) (6)		Lucia et al. (2016)	
Ivory gull ( <i>Pagophila eburnea</i> )		April 2012	Barentsøya	Svalbard	0.81 $\pm$ 0.05 (SD) (4)		Lucia et al. (2016)	
Ivory gull ( <i>Pagophila eburnea</i> )		April 2013	Barentsøya	Svalbard	1.06 $\pm$ 0.16 (SD) (5)		Lucia et al. (2016)	
Ivory gull ( <i>Pagophila eburnea</i> )		April 2014	Barentsøya	Svalbard		1.20 $\pm$ 0.07 (SD) (42)	Lucia et al. (2016)	
Procellariid		Northern fulmar ( <i>Fulmarus glacialis</i> )	NA	Prince Leopold Island	Canada	3.62 $\pm$ 2.06 (SD) (7)		Mallory et al. (2018)

**Table 2.** Total Hg concentrations measured in body feathers (BF) (mean  $\pm$  SD or SE  $\mu\text{g/g}$  dw; sample size is given into bracket) for seabirds breeding in the Arctic. If sexes were differentiated, F for female or M for male is added. a: ppb to  $\mu\text{g/g}$ ; b: fw.

Family	Species	Sampling year	Colony location	Country	Body feathers	References	
Alcid	Brünnich guillemot ( <i>Uria lomvia</i> )	1992 & 1993	Homøya	Norway	0.78 $\pm$ 0.18 (SD) (14)	Wenzel and Gabrielsen (1995)	
	Common guillemot ( <i>Uria aalge</i> )	1992 & 1993	Homøya	Norway	0.88 $\pm$ 0.19 (SD) (10)	Wenzel and Gabrielsen (1995)	
	Kittlitz's murrelet ( <i>Brachyramphus brevirostris</i> ) ADULT	August 2008–2011	Agattu	Aleutian Islands	1.66 $\pm$ 1.28 (SE) (29) a, b	Kaler et al. (2014)	
	Kittlitz's murrelet ( <i>Brachyramphus brevirostris</i> ) CHICK	August 2008–2011	Agattu	Aleutian Islands	1.47 $\pm$ 0.67 (SE) (8) a, b	Kaler et al. (2014)	
	Kittlitz's murrelet ( <i>Brachyramphus brevirostris</i> )	2008–2011	Agattu	Aleutian Islands	2.06 $\pm$ 1.28 (SD) (37)	Kenney et al. (2018)	
	Kittlitz's murrelet ( <i>Brachyramphus brevirostris</i> )	2011	Adak Island	Aleutian Islands	5.15 $\pm$ 2.51 (SD) (7)	Kenney et al. (2018)	
	Little auk ( <i>Alle alle</i> )	2007	Kap Höegh	Greenland	1.00 $\pm$ 0.22 (SD) (20)	Amélineau et al. (2019)	
	Little auk ( <i>Alle alle</i> )	2008	Kap Höegh	Greenland	1.13 $\pm$ 0.24 (SD) (19)	Amélineau et al. (2019)	
	Little auk ( <i>Alle alle</i> )	2009	Kap Höegh	Greenland	1.04 $\pm$ 0.26 (SD) (19)	Amélineau et al. (2019)	
	Little auk ( <i>Alle alle</i> )	2010	Kap Höegh	Greenland	1.39 $\pm$ 0.73 (SD) (49)	Amélineau et al. (2019)	
	Little auk ( <i>Alle alle</i> )	2011	Kap Höegh	Greenland	1.84 $\pm$ 0.85 (SD) (18)	Amélineau et al. (2019)	
	Little auk ( <i>Alle alle</i> )	2012	Kap Höegh	Greenland	1.25 $\pm$ 0.51 (SD) (20)	Amélineau et al. (2019)	
	Little auk ( <i>Alle alle</i> )	2013	Kap Höegh	Greenland	1.69 $\pm$ 0.80 (SD) (20)	Amélineau et al. (2019)	
	Little auk ( <i>Alle alle</i> )	2014	Kap Höegh	Greenland	1.15 $\pm$ 0.33 (SD) (19)	Amélineau et al. (2019)	
	Little auk ( <i>Alle alle</i> )	2015	Kap Höegh	Greenland	1.25 $\pm$ 0.53 (SD) (20)	Amélineau et al. (2019)	
	Little auk ( <i>Alle alle</i> )	2016	Kap Höegh	Greenland	1.60 $\pm$ 0.62 (SD) (20)	Amélineau et al. (2019)	
	Little auk ( <i>Alle alle</i> )	2007	Kap Höegh	Greenland	1.00 $\pm$ 0.22 (SD) (20)	Fort et al. (2016)	
	Little auk ( <i>Alle alle</i> )	2008	Kap Höegh	Greenland	1.13 $\pm$ 0.24 (SD) (19)	Fort et al. (2016)	
	Little auk ( <i>Alle alle</i> )	2009	Kap Höegh	Greenland	1.04 $\pm$ 0.26 (SD) (19)	Fort et al. (2016)	
	Little auk ( <i>Alle alle</i> )	2010	Kap Höegh	Greenland	1.37 $\pm$ 0.70 (SD) (45)	Fort et al. (2016)	
	Little auk ( <i>Alle alle</i> )	2011	Kap Höegh	Greenland	1.70 $\pm$ 0.66 (SD) (17)	Fort et al. (2016)	
	Little auk ( <i>Alle alle</i> )	2012	Kap Höegh	Greenland	1.25 $\pm$ 0.51 (SD) (20)	Fort et al. (2016)	
	Little auk ( <i>Alle alle</i> )	2013	Kap Höegh	Greenland	1.69 $\pm$ 0.80 (SD) (20)	Fort et al. (2016)	
	Little auk ( <i>Alle alle</i> )	2014	Kap Höegh	Greenland	2.11 $\pm$ 1.49 (SD) (20)	Fort et al. (2016)	
	Little auk ( <i>Alle alle</i> )	2010–2011	Kap Höegh	Greenland	1.53 $\pm$ 0.84 (SD) (78)	Fort et al. (2014)	
	Pigeon guillemot ( <i>Cepphus columba</i> )	July 2004	Aleutian Islands	Aleutian Islands	7.11 $\pm$ 0.66 (SE) (38) a	Burger et al. (2009)	
	Pigeon guillemot ( <i>Cepphus columba</i> )	June 2004	Amchitka Island	Aleutian Islands	7.72 $\pm$ 0.84 (SE) (21) a	Burger et al. (2007)	
	Pigeon guillemot ( <i>Cepphus columba</i> )	June 2004	Kiska Island	Aleutian Islands	6.36 $\pm$ 1.03 (SE) (17) a	Burger et al. (2007)	
	Pigeon guillemot ( <i>Cepphus columba</i> )	2004	Prince William Sound	Alaska	2.81 $\pm$ 0.25 (SE) (40) a	Burger et al. (2007)	
	Tufted puffin ( <i>Pratercula cirrhata</i> )	July 2004	Aleutian Islands	Aleutian Islands	2.54 $\pm$ 0.19 (SE) (39) a	Burger et al. (2009)	
	Tufted puffin ( <i>Pratercula cirrhata</i> )	July 2004	Amchitka Island	Aleutian Islands	2.59 $\pm$ 0.29 (SE) (22)	Burger et al. (2009)	
	Tufted puffin ( <i>Pratercula cirrhata</i> )	July 2004	Kiska Island	Aleutian Islands	2.48 $\pm$ 0.25 (SE) (17)	Burger et al. (2009)	
	Anatid	Common eider ( <i>Somateria mollissima</i> )	July 2007	Amchitka Island	Aleutian Islands	0.89 $\pm$ 0.16 (SE) (22) F a	Burger et al. (2008b)
		Common eider ( <i>Somateria mollissima</i> )	July 2007	Kiska Island	Aleutian Islands	1.24 $\pm$ 0.32 (SE) (8) F a	Burger et al. (2008a)
		Common eider ( <i>Somateria mollissima</i> )	July 2004	Aleutian Islands	Aleutian Islands	0.84 $\pm$ 0.08 (SE) (26) a	Burger et al. (2009)
		Common eider ( <i>Somateria mollissima</i> )	July 2004	Amchitka Island	Aleutian Islands	0.85 $\pm$ 0.14 (SE) (13)	Burger et al. (2009)
		Common eider ( <i>Somateria mollissima</i> )	July 2004	Kiska Island	Aleutian Islands	0.83 $\pm$ 0.08 (SE) (13)	Burger et al. (2009)
		Common eider ( <i>Somateria mollissima</i> )	2011	Storholmen, Kongsfjorden	Svalbard	0.50–1.90 $\mu\text{g/g}$ ww (29) F	Fenstad et al. (2017)
		King eider ( <i>Somateria spectabilis</i> )	2014	Utqiagvik	Alaska	0.80 $\pm$ 0.37 (SD) (28) F	Miller et al. (2019)
		King eider ( <i>Somateria spectabilis</i> )	Jun, July 1993	Chaun	Russia	0.57 (1)	Kim et al. (1996a, 1996b)
		Long-tailed duck ( <i>Clangula hyemalis</i> )	Jun, July 1993	Chaun	Russia	1.96 $\pm$ 3.31 (SD) (8)	Kim et al. (1996a, 1996b)
Spectacled eider ( <i>Somateria fischeri</i> )		2010–2014	Utqiagvik	Alaska	1.35 $\pm$ 1.00 (SD) (42) F	Miller et al. (2019)	
Hydrobatid		Fork-tailed storm-petrel ( <i>Oceanodroma furcata</i> )	August 2008–2011	Agattu	Aleutian Islands	6.70 $\pm$ 1.63 (SE) (12) a, b	Kaler et al. (2014)
		Arctic tern ( <i>Sterna paradisae</i> )	Jun, July 1993	Chaun	Russia	0.89 $\pm$ 0.15 (SD) (10)	Kim et al. (1996a, 1996b)
Larid		Black-legged kittiwakes ( <i>Rissa tridactyla</i> )	2004	Prince William Sound	Alaska	2.91 $\pm$ 0.19 (SE) (61) a	Burger et al. (2008b)
		Black-legged kittiwakes ( <i>Rissa tridactyla</i> ) ADULT	1992 & 1993	Homøya	Norway	2.03 $\pm$ 0.40 (SD) $\mu\text{g g}^{-1}$ ww (27)	Wenzel and Gabrielsen (1995)
		Black-legged kittiwakes ( <i>Rissa tridactyla</i> ) FLEDGLING	1992 & 1993	Homøya	Norway	0.55 $\pm$ 0.10 (SD) $\mu\text{g g}^{-1}$ ww (10)	Wenzel and Gabrielsen (1995)
		Glaucous gull ( <i>Larus hyperboreus</i> )	Jun, July 1993	Chaun	Russia	5.96 (1)	Kim et al. (1996a, 1996b)
		Glaucous-winged gulls ( <i>Larus glaucescens</i> )	July 2004	Aleutian Islands	Aleutian Islands	3.68 $\pm$ 0.37 (SE) (63) a	Burger et al. (2009)
	Glaucous-winged gulls ( <i>Larus glaucescens</i> ) ADULT	NA	Adak Island	Aleutian Islands	2.97 $\pm$ 0.57 (SE) (16) a	Burger et al. (2009)	
	Glaucous-winged gulls ( <i>Larus glaucescens</i> ) ADULT	NA	Amchitka Island	Aleutian Islands	4.01 $\pm$ 0.54 (SE) (30) a	Burger et al. (2009)	
	Glaucous-winged gulls ( <i>Larus glaucescens</i> ) ADULT	NA	Kiska Island	Aleutian Islands	3.76 $\pm$ 0.86 (SE) (17) a	Burger et al. (2009)	
	Glaucous-winged gulls ( <i>Larus glaucescens</i> ) CHICKS	NA	Amchitka Island	Aleutian Islands	2.19 $\pm$ 0.26 (SE) (21) a	Burger et al. (2009)	
	Glaucous-winged gulls ( <i>Larus glaucescens</i> ) CHICKS	NA	Kiska Island	Aleutian Islands	1.68 $\pm$ 0.23 (SE) (15) a	Burger et al. (2009)	
	Herring gull ( <i>Larus argentatus</i> )	June 1993	Chaun	Russia	6.10 $\pm$ 4.60 (SD) (5)	Kim et al. (1996a, 1996b)	
	Herring gull ( <i>Larus argentatus</i> )	Jun, July 1993	Chaun	Russia	6.06 $\pm$ 4.60 (SD) (6)	Kim et al. (1996a, 1996b)	
	Ivory gull ( <i>Pagophila eburnea</i> )	April 2011	Barentsøya	Svalbard	2.65 $\pm$ 0.42 (SD) (19)	Lucia et al. (2016)	
	Ivory gull ( <i>Pagophila eburnea</i> )	April 2012	Barentsøya	Svalbard	3.03 $\pm$ 0.41 (SD) (40)	Lucia et al. (2016)	
	Ivory gull ( <i>Pagophila eburnea</i> )	April 2013	Barentsøya	Svalbard	2.79 $\pm$ 0.24 (SD) (47)	Lucia et al. (2016)	
	Ivory gull ( <i>Pagophila eburnea</i> )	April 2014	Barentsøya	Svalbard	3.39 $\pm$ 0.30 (SD) (39)	Lucia et al. (2016)	
	Ivory gull ( <i>Pagophila eburnea</i> )	2010	Seymour Island	Canada	11.66 $\pm$ 6.52 (SD) (8)	Mallory et al. (2015)	
	Sabine's gull ( <i>Xema sabini</i> )	Jun, July 1993	Chaun	Russia	1.70 $\pm$ 0.47 (SD) (2)	Kim et al. (1996a, 1996b)	
	Stercorariid	Long-tailed jaeger ( <i>Stercorarius longicaudus</i> )	Jun, July 1993	Chaun	Russia	1.95 $\pm$ 0.54 (SD) (5)	Kim et al. (1996a, 1996b)
Parasitic jaeger ( <i>Stercorarius parasiticus</i> )		Jun, July 1993	Chaun	Russia	1.84 (1)	Kim et al. (1996a, 1996b)	

**Table 3.** Total Hg concentrations (mean  $\pm$  SD  $\mu\text{g/g}$  dw; sample size is given into bracket) measured in flight feathers (a: primary or b: secondary) for seabirds breeding in the Arctic.

Family	Species	Sampling year	Colony location	Country	Flight feathers	References
Alcid	Brünnich's guillemot ( <i>Uria lomvia</i> )	NA	Prince Leopold Island	Canada	1.94 $\pm$ 0.63 (10) a	Mallory et al. (2015)
	Kittlitz's murrelet ( <i>Brachyramphus brevirostris</i> )	2011	Adak Island	Aleutian Islands	37.18 (1) b	Kenney et al. (2018)
Anatid	Long-tailed duck ( <i>Clangula hyemalis</i> )	June 1993	Chaun	Russia	0.70 $\pm$ 0.20 (5) a	Kim et al., 1996a, 1996b
	Common eider ( <i>Somateria mollissima</i> )	NA	Cape Dorset, Nunavut	Canada	0.59 $\pm$ 0.21 (10) a	Mallory et al. (2015)
Larid	Herring gull ( <i>Larus argentatus</i> )	June 1993	Chaun	Russia	6.10 $\pm$ 4.60 (5) a	Kim et al., 1996a, 1996b
	Arctic tern ( <i>Sterna paradisae</i> )	June 1993	Chaun	Russia	0.90 $\pm$ 0.10 (5) a	Kim et al., 1996a, 1996b
	Black-legged kittiwakes ( <i>Rissa tridactyla</i> )	NA	Prince Leopold Island	Canada	3.58 $\pm$ 0.92 (2) a	Mallory et al. (2015)
	Ivory gull ( <i>Pagophila eburnea</i> )	NA	Seymour Island	Canada	15.79 $\pm$ 14.13 (8) a	Mallory et al. (2015)
	Glaucous gull ( <i>Larus hyperboreus</i> )	NA	Nasaruvaalik Island	Canada	2.31 $\pm$ 1.68 (4) a	Mallory et al. (2015)
Procellariid	Northern fulmar ( <i>Fulmarus glacialis</i> )	NA	Prince Leopold Island	Canada	2.71 $\pm$ 0.72 (10) a	Mallory et al. (2015)

**Table 4.** Total Hg concentrations measured in head feathers (HF) (mean  $\pm$  SD  $\mu\text{g/g}$  dw, sample size is given in brackets) for seabirds breeding in the Arctic.

Family	Species	Sampling year	Location	Country	Head feather	References
Alcid	Little auk ( <i>Alle alle</i> )	2007	Kap Hoegh (East Greenland)	Greenland	3.73 $\pm$ 1.33 (20)	<a href="#">Amélineau et al. (2019)</a>
	Little auk ( <i>Alle alle</i> )	2008	Kap Hoegh (East Greenland)	Greenland	2.86 $\pm$ 0.71 (20)	<a href="#">Amélineau et al. (2019)</a>
	Little auk ( <i>Alle alle</i> )	2009	Kap Hoegh (East Greenland)	Greenland	3.06 $\pm$ 0.99 (20)	<a href="#">Amélineau et al. (2019)</a>
	Little auk ( <i>Alle alle</i> )	2010	Kap Hoegh (East Greenland)	Greenland	3.17 $\pm$ 0.82 (40)	<a href="#">Amélineau et al. (2019)</a>
	Little auk ( <i>Alle alle</i> )	2011	Kap Hoegh (East Greenland)	Greenland	3.21 $\pm$ 0.99 (25)	<a href="#">Amélineau et al. (2019)</a>
	Little auk ( <i>Alle alle</i> )	2012	Kap Hoegh (East Greenland)	Greenland	2.27 $\pm$ 0.41 (20)	<a href="#">Amélineau et al. (2019)</a>
	Little auk ( <i>Alle alle</i> )	2013	Kap Hoegh (East Greenland)	Greenland	2.75 $\pm$ 0.87 (20)	<a href="#">Amélineau et al. (2019)</a>
	Little auk ( <i>Alle alle</i> )	2014	Kap Hoegh (East Greenland)	Greenland	3.99 $\pm$ 1.96 (20)	<a href="#">Amélineau et al. (2019)</a>
	Little auk ( <i>Alle alle</i> )	2015	Kap Hoegh (East Greenland)	Greenland	3.06 $\pm$ 0.86 (20)	<a href="#">Amélineau et al. (2019)</a>
	Little auk ( <i>Alle alle</i> )	2007	Kap Hoegh (East Greenland)	Greenland	3.73 $\pm$ 1.33 (20)	<a href="#">Fort et al. (2016)</a>
	Little auk ( <i>Alle alle</i> )	2008	Kap Hoegh (East Greenland)	Greenland	2.86 $\pm$ 0.71 (20)	<a href="#">Fort et al. (2016)</a>
	Little auk ( <i>Alle alle</i> )	2009	Kap Hoegh (East Greenland)	Greenland	3.06 $\pm$ 0.99 (20)	<a href="#">Fort et al. (2016)</a>
	Little auk ( <i>Alle alle</i> )	2010	Kap Hoegh (East Greenland)	Greenland	3.17 $\pm$ 0.82 (40)	<a href="#">Fort et al. (2016)</a>
	Little auk ( <i>Alle alle</i> )	2011	Kap Hoegh (East Greenland)	Greenland	3.21 $\pm$ 0.99 (25)	<a href="#">Fort et al. (2016)</a>
	Little auk ( <i>Alle alle</i> )	2012	Kap Hoegh (East Greenland)	Greenland	2.27 $\pm$ 0.41 (20)	<a href="#">Fort et al. (2016)</a>
	Little auk ( <i>Alle alle</i> )	2013	Kap Hoegh (East Greenland)	Greenland	2.60 $\pm$ 0.56 (19)	<a href="#">Fort et al. (2016)</a>
	Little auk ( <i>Alle alle</i> )	2014	Kap Hoegh (East Greenland)	Greenland	3.02 $\pm$ 1.14 (16)	<a href="#">Fort et al. (2016)</a>
	Little auk ( <i>Alle alle</i> )	2010–2011	Kap Hoegh (East Greenland)	Greenland	3.17 $\pm$ 0.83 (81)	<a href="#">Fort et al. (2014)</a>

**Annex 1:** List of the 42 species referenced in the 101 articles who studied Hg contamination in seabirds breeding in the Arctic. For each species, pre or post breeding have been assessed.

Family	Species	Prebreeding molt	Postbreeding molt	Molt pattern unclear or unknow	References
Alcid	Atlantic puffin ( <i>Fratercula arctica</i> )			x	<a href="#">Gaston and Jones, 1998</a> ; <a href="#">Harris and Yule, 1977</a>
	Black guillemot ( <i>Cepphus grylle</i> )	x	x		<a href="#">Gaston and Jones, 1998</a>
	Mandt's black guillemot ( <i>Cepphus grylle mandtii</i> )	x	x		<a href="#">Gaston and Jones, 1998</a>
	Brünnich guillemot ( <i>Uria lomvia</i> )	x	x		<a href="#">Gaston and Jones, 1998</a>
	Common guillemot ( <i>Uria aalge</i> )	x	x		<a href="#">Gaston and Jones, 1998</a>
	Crested auklet ( <i>Aethia cristatella</i> )	x	x		<a href="#">Gaston and Jones, 1998</a>
	Horned puffin ( <i>Fratercula corniculata</i> )	x	x		<a href="#">Piatt and Kitaysky, 2002</a>
	Kittlitz's murrelet ( <i>Brachyramphus brevirostris</i> )	x	x		<a href="#">Gaston and Jones, 1998</a>
	Little auk ( <i>Alle alle</i> )	x	x		<a href="#">Gaston and Jones, 1998</a>
	Parakeet auklet ( <i>Aethia psittacula</i> )	x	x		<a href="#">Gaston and Jones, 1998</a>
	Pigeon guillemot ( <i>Cepphus columba</i> )	x	x		<a href="#">Gaston and Jones, 1998</a>
	Razorbill ( <i>Alca torda</i> )	x	x		<a href="#">Gaston and Jones, 1998</a>
	Tufted puffin ( <i>Fratercula cirrhata</i> )	x	x	Post-breeding apparently complete	<a href="#">Gaston and Jones, 1998</a>
Whiskered auklet ( <i>Aethia pygmaea</i> )				x <a href="#">Gaston and Jones, 1998</a>	
Anatid	Black Scoter ( <i>Melanitta nigra</i> )		x		<a href="#">Sea Duck Joint Venture, 2016</a>
	Common eider ( <i>Somateria mollissima</i> ) female		x		<a href="#">Waltho and Coulson, 2015</a>
	Common eider ( <i>Somateria mollissima</i> ) male	x	x	Experience an eclipse plumage before going back to breeding plumage	<a href="#">Waltho and Coulson, 2015</a>
	King eider ( <i>Somateria spectabilis</i> )		x		<a href="#">Fenneman and Tootchin, 2013</a>
	Long-tailed duck/oldsquaw ( <i>Clangula hyemalis</i> )	x	x		<a href="#">Howell et al., 2003</a>
	Spectacled eiders ( <i>Somateria fischeri</i> )		x		<a href="#">US Fish &amp; Wildlife Service, 2012</a>
	Steller's eiders ( <i>Polysticta stelleri</i> )		x		<a href="#">US Fish &amp; Wildlife Service, 2012</a>
	White-winged Scoter ( <i>Melanitta fusca</i> )		x		<a href="#">Brown and Fredrickson, 1997</a> ; <a href="#">Dickson et al., 2012</a>

Hydrobatid	Fork-tailed storm-petrel ( <i>Oceanodroma furcata</i> )			x	But see Harris, 1974
	Leach's Storm-petrel ( <i>Oceanodroma leucorhoa</i> )			x	<u>Harris, 1974</u>
Larid	Arctic tern ( <i>Sterna paradisae</i> )	x	x		<u>Voelker, 1997</u>
	Black-legged kittiwakes ( <i>Rissa tridactyla</i> )	x	x		<u>Cramp and Simmons, 1983</u>
	Glaucous gull/Arctic gull ( <i>Larus hyperboreus</i> )	x	x		<u>Hume, 1975</u>
	Glaucous-winged gulls ( <i>Larus glaucescens</i> )	x	x		<u>Gabrielson and Lincoln, 1959</u>
	Herring gull ( <i>Larus argentatus</i> )	x	x		Cramp and Simmons, 1983
	Iceland gull ( <i>Larus glaucooides</i> )	x	x		Cramp and Simmons, 1983
	Ivory gull ( <i>Pagophila eburnea</i> )	x			Cramp and Simmons, 1983
	Ross's gull ( <i>Rhodostethia rosea</i> )	x	x		Cramp and Simmons, 1983
	Sabine's gull ( <i>Xema sabini</i> )	x	x		Cramp and Simmons, 1983
	Thayer's gull ( <i>Larus thayeri</i> )	x	x		<u>Howell and Gunn, 2007</u>
Phalacrocoracid	European shag ( <i>Phalacrocorax aristotelis aristotelis</i> )			x	<u>Potts, 1970</u>
	Pelagic cormorant ( <i>Phalacrocorax pelagicus</i> )	x	x		<u>Rohwer, 1999; Van Tets, 1959</u>
	Red-faced cormorants ( <i>Phalacrocorax urile</i> )	x	x		<u>Toochin and Fenneman, 2014</u>
Procellariid	Northern fulmar ( <i>Fulmarus glacialis</i> )			x	Initiated around one week before hatching <u>Allard et al., 2008</u>
Stercorariid	Long-tailed Jaeger ( <i>Stercorarius longicaudus</i> )	x	x		<u>Pyle, 2016</u>
	Parasitic Jaeger ( <i>Stercorarius parasiticus</i> )	x	x		Pyle, 2016
	Pomarine skua ( <i>Stercorarius pomarinus</i> )	x	x		Pyle, 2016
Sulid	Northern gannet ( <i>Morus bassanus</i> )				Continuous molt from post breeding to pre-breeding period <u>Nelson, 2010</u>



**Annex 2:** List of the 101 articles studying Hg in seabirds breeding in the Arctic.

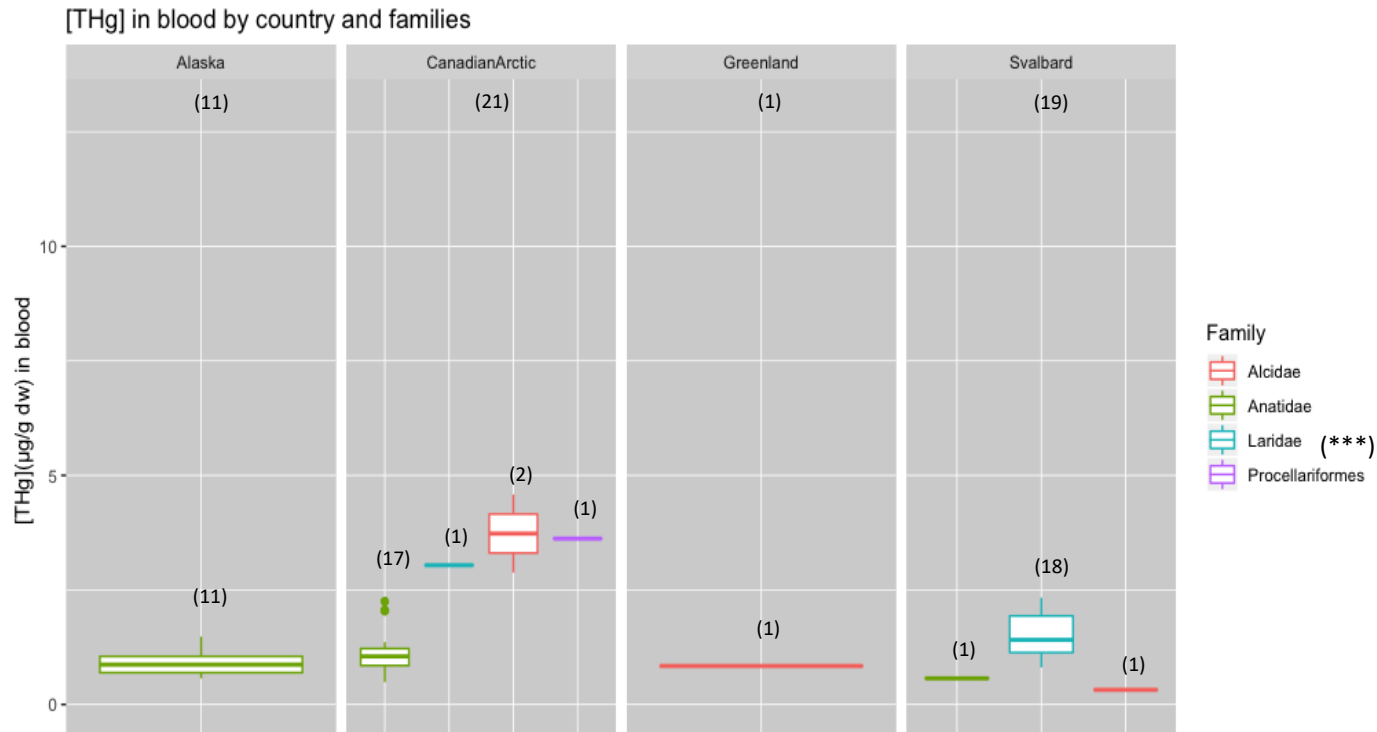
Article Nb	Authors	Title
1	Amélineau et al., 2019	Arctic climate change and pollution impact little auk foraging and fitness across a decade
2	Eckbo et al., 2019	Individual variability in contaminants and physiological status in a resident Arctic seabirds species
3	Miller et al., 2019	Interspecific patterns of trace elements in sea ducks: can surrogate species be used in contaminants monitoring?
4	Blévin et al., 2018	Organochlorines, perfluoroalkyl substances, mercury and egg incubation temperatures in an Arctic seabirds: insight from data loggers.
5	Braune B.M. et al., 2018	Variation in organochlorine and mercury levels in first and replacement eggs of a single-egg clutch breeder, the thick-billed murre, at a breeding colony in the Canadian Arctic
6	Keney et al., 2018	Mercury concentrations in multiple tissues of Kittlitz's murrelets ( <i>Brachyramphus brevirostris</i> )
7	Mallory et al., 2018	Mercury concentrations in blood, brain and muscle tissues of coastal and peagic birds from northeastern Canada
8	Blévin et al., 2017	Contaminants and energy expenditure in an Arctic seabird: organochlorine pesticides and perfluoroalkyl substances are associated with metabolic rate in a contrasted manner.
9	Fenstad A.A. et al., 2017	Blood and feather concentrations of toxic elements in a Baltic and an Arctic seabird population
10	Ishii C., et al. 2017	Contamination status and accumulation characteristics of heavy metals and arsenic in five seabird species from the central Bering Sea
11	Mallory M.L. and Braune B.M. 2017	Do concentrations in eggs and liver tissue tell the same story of temporal trends of mercury in high Arctic seabirds?
12	Mallory C.D. et al., 2017	Hepatic trace element concentrations of breeding female common eiders across a latitudinal gradient in the eastern Canadian Arctic
13	Provencher J.F. et al 2017	Anti-parasite treatment, but not mercury burdens, influence nesting propensity dependent on arrival time or body condition in a marine bird
14	Braune B.M. et al., 2016	Temporal trends of mercury in eggs of five sympatrically breeding seabird species in the Canadian Arctic
15	Fenstad A.A. et al., 2016	DNA double-strand breaks in incubating female common eiders ( <i>Somateria mollissima</i> ): Comparison between a low and a high polluted area
16	Fort J. et al., 2016	Does temporal variation of mercury levels in Arctic seabirds reflect changes in global environmental contamination, or a modification of Arctic marine food web functioning?
17	Lucia M. et al., 2016	Trace Element Concentrations in Relation to the Trophic Behaviour of Endangered Ivory Gulls ( <i>Pagophila eburnea</i> ) During Their Stay at a Breeding Site in Svalbard
18	Miller et al., 2016	Trace elements in Sea ducks of the Alaskan Arctic Coast: patterns of variation among species, sexes, and ages
19	Provencher J.F. et al., 2016	Implications of mercury and lead concentrations on breeding physiology and phenology in an Arctic bird
20	Peck L.E. et al., 2016	Persistent organic pollutant and mercury concentrations in eggs of ground-nesting marine birds in the Canadian high Arctic
21	Provencher J.F. et al., 2016	Direct and indirect causes of sex differences in mercury concentrations and parasitic infections in a marine bird
22	Tartu S. et al., 2016	Mercury exposure, stress and prolactin secretion in an Arctic seabird: An experimental study

23	Bond A.L. et al., 2015	Rapidly increasing methyl mercury in endangered ivory gull ( <i>Pagophila eburnea</i> ) feathers over a 130 year record
24	Clayden M.G. et al., 2015	Mercury bioaccumulation and biomagnification in a small Arctic polynya ecosystem
25	Goutte A. et al., 2015	Survival rate and breeding outputs in a high Arctic seabird exposed to legacy persistent organic pollutants and mercury
26	Huber S. et al., 2015	A broad cocktail of environmental pollutants found in eggs of three seabird species from remote colonies in Norway
27	Lucia M. et al., 2015	Circumpolar contamination in eggs of the high-arctic ivory gull <i>Pagophila eburnea</i>
28	Mallory M.L. et al., 2015	Mercury concentrations in feathers of marine birds in Arctic Canada
29	Øverjordet I.B. et al., 2015	Effect of diet, location and sampling year on bioaccumulation of mercury, selenium and cadmium in pelagic feeding seabirds in Svalbard
30	Overjordet I.B. et al., 2015	Toxic and essential elements changed in black-legged kittiwakes ( <i>Rissa tridactyla</i> ) during their stay in an Arctic breeding area
31	Ruus A. et al., 2015	Methylmercury biomagnification in an Arctic pelagic food web
32	Tartu S. et al., 2015	Increased adrenal responsiveness and delayed hatching date in relation to polychlorinated biphenyl exposure in Arctic-breeding black-legged kittiwakes ( <i>Rissa tridactyla</i> )
33	Braune B.M. et al., 2014	A geographical comparison of mercury in seabirds in the eastern Canadian Arctic
34	Braune B.M. et al., 2014	Changes in food web structure alter trends of mercury uptake at two seabird colonies in the Canadian Arctic
35	Fort J. et al., 2014	Spatial ecotoxicology: Migratory arctic seabirds are exposed to mercury contamination while overwintering in the northwest Atlantic
36	Kaler R.S.A. et al., 2014	Mercury concentrations in breast feathers of three upper trophic level marine predators from the western Aleutian Islands, Alaska
37	Provencher J.F. et al., 2014	Trace element concentrations and gastrointestinal parasites of Arctic terns breeding in the Canadian High Arctic
38	Brasso R.L. and Polito M.J., 2013	Trophic calculations reveal the mechanism of population-level variation in mercury concentrations between marine ecosystems: Case studies of two polar seabirds
39	Tartu S. et al., 2013	To breed or not to breed: Endocrine response to mercury contamination by an Arctic seabird
40	Braune B.M. et al., 2012	Toxicity of methylmercury injected into eggs of thick-billed murrelets and arctic terns
41	Day R.D. et al., 2012	Regional, temporal, and species patterns of mercury in Alaskan seabird eggs: Mercury sources and cycling or food web effects?
42	Day R.D. et al., 2012	Mercury stable isotopes in seabird eggs reflect a gradient from terrestrial geogenic to oceanic mercury reservoirs
43	Miljeteig C. et al., 2012	Eggshell thinning and decreased concentrations of vitamin E are associated with contaminants in eggs of ivory gulls
44	Point D. et al., 2011	Methylmercury photodegradation influenced by sea-ice cover in Arctic marine ecosystems
45	Akearok J.A. et al., 2010	Inter- and intraclutch variation in egg mercury levels in marine bird species from the Canadian Arctic
46	Braune B.M. et al., 2010	Persistent halogenated organic contaminants and mercury in northern fulmars ( <i>Fulmarus glacialis</i> ) from the Canadian Arctic
47	Wayland M. et al., 2010	Evidence of weak contaminant-related oxidative stress in glaucous gulls ( <i>Larus hyperboreus</i> ) from the Canadian arctic
48	Miljeteig C. et al., 2009	High levels of contaminants in ivory gull <i>Pagophila eburnea</i> eggs from the Russian and Norwegian arctic

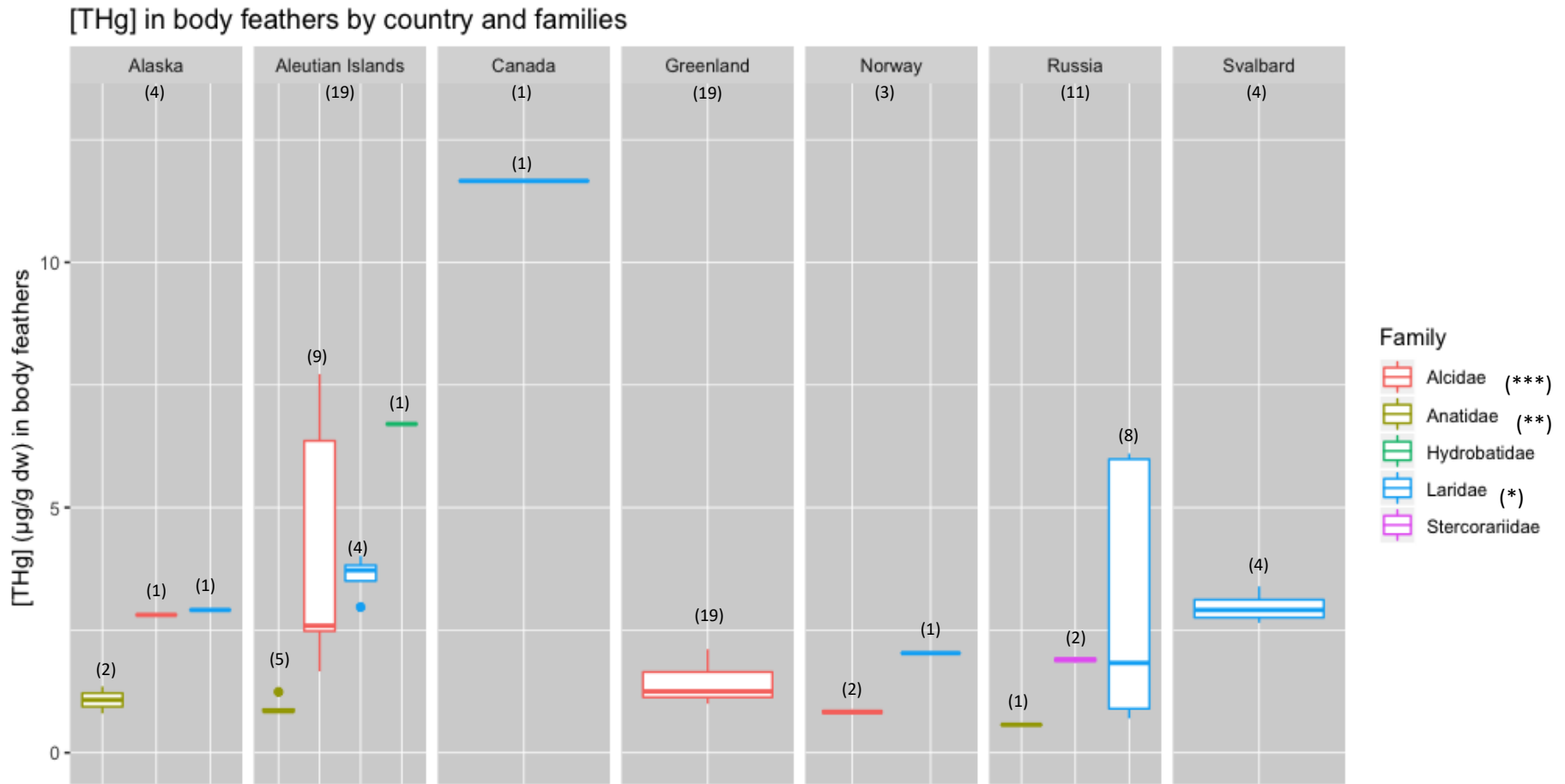
49	Burger J. and Gochfeld M., 2009	Mercury and other metals in feathers of common eider ( <i>Somateria mollissima</i> ) and tufted puffin ( <i>Fratercula cirrhata</i> ) from the Aleutian chain of Alaska
50	Burger J. et al., 2009	Mercury and other metals in eggs and feathers of glaucous-winged gulls ( <i>Larus glaucescens</i> ) in the Aleutians
51	Burger J. and Gochfeld M., 2009	Comparison of arsenic, cadmium, chromium, lead, manganese, mercury and selenium in feathers in bald eagle ( <i>Haliaeetus leucocephalus</i> ), and comparison with common eider ( <i>Somateria mollissima</i> ), glaucous-winged gull ( <i>Larus glaucescens</i> ), pigeon guillemot ( <i>Cepphus columba</i> ), and tufted puffin ( <i>Fratercula cirrhata</i> ) from the Aleutian chain of Alaska
52	Jæger I. et al., 2009	Biomagnification of mercury in selected species from an Arctic marine food web in Svalbard
53	Sagerup K. et al., 2009	Persistent organic pollutants and mercury in dead and dying glaucous gulls ( <i>Larus hyperboreus</i> ) at Bjørnøya (Svalbard)
54	Sagerup K. et al., 2009	Persistent organic pollutants, heavy metals and parasites in the glaucous gull ( <i>Larus hyperboreus</i> ) on Spitsbergen
55	Braune B.M. and Scheuhammer A.M., 2008	Trace element and metalloprotein concentrations in seabirds from the Canadian Arctic
56	Burger J. et al., 2008	Assessment of metals in down feathers of female common eiders and their eggs from the Aleutians: Arsenic, cadmium, chromium, lead, manganese, mercury, and selenium
57	Burger J. et al., 2008	Arsenic, cadmium, chromium, lead, manganese, mercury, and selenium in feathers of Black-legged Kittiwake ( <i>Rissa tridactyla</i> ) and Black Oystercatcher ( <i>Haematopus bachmani</i> ) from Prince William Sound, Alaska
58	Helgason L.B. et al., 2008	Levels and temporal trends (1983-2003) of persistent organic pollutants (POPs) and mercury (Hg) in seabird eggs from Northern Norway
59	Ricca M.A. et al., 2008	Sources of organochlorine contaminants and mercury in seabirds from the Aleutian archipelago of Alaska: Inferences from spatial and trophic variation
60	Wayland M. et al., 2008	Survival rates and blood metal concentrations in two species of free-ranging North American sea ducks
61	Wayland M. et al., 2008	Trace element concentrations in Blood of nesting king eiders in the Canadian Arctic
62	Yin X. et al., 2008	Animal excrement: A potential biomonitor of heavy metal contamination in the marine environment
63	Braune B.M., 2007	Temporal trends of organochlorines and mercury in seabird eggs from the Canadian Arctic, 1975-2003
64	Burger J. et al., 2007	Mercury levels and potential risk from subsistence foods from the Aleutians
65	Burger J. et al., 2007	Mercury, arsenic, cadmium, chromium lead, and selenium in feathers of pigeon guillemots ( <i>Cepphus columba</i> ) from Prince William Sound and the Aleutian Islands of Alaska
66	Burger J. and Gochfeld M., 2007	Metals and radionuclides in birds and eggs from Amchitka and Kiska Islands in the Bering Sea/Pacific Ocean ecosystem
67	Rigét F. et al., 2007	Transfer of mercury in the marine food web of West Greenland
68	Wayland M. et al., 2007	Year-to-year correlations in blood metal levels among individuals of two species of North American sea ducks
69	Borgå K. et al., 2006	Regional and species specific bioaccumulation of major and trace elements in arctic seabirds
70	Braune B.M. et al., 2006	Elevated mercury levels in a declining population of ivory gulls in the Canadian Arctic
71	Braune B.M. and Malone B.J., 2006	Mercury and selenium in livers of waterfowl harvested in northern Canada
72	Day R.D. et al., 2006	Murre eggs ( <i>Uria aalge</i> and <i>Uria lomvia</i> ) as indicators of mercury contamination in the Alaskan marine environment

73	Braune B.M. et al., 2005	Regional differences in collagen stable isotope and tissue trace element profiles in populations of long-tailed duck breeding in the Canadian Arctic
74	Campbell L.M. et al., 2005	Mercury and other trace elements in a pelagic Arctic marine food web (Northwater Polynya, Baffin Bay)
75	Rothschild R.F.N. and Duffy L.K., 2005	Mercury concentrations in muscle, brain and bone of Western Alaskan waterfowl
76	Wayland M. et al., 2005	Concentrations of cadmium, mercury and selenium in common eider ducks in the eastern Canadian arctic: Influence of reproductive stage
77	Braune B.M. et Simon M., 2004	Trace elements and halogenated organic compounds in Canadian Arctic seabirds
78	Franson J.C. et al., 2004	Contaminants in molting long-tailed ducks and nesting common eiders in the Beaufort Sea
79	Johansen P. et al., 2004	Human exposure to contaminants in the traditional Greenland diet
80	Mallory M.L. et al., 2004	Trace elements in marine birds, arctic hare and ringed seals breeding near Qikiqtarjuaq, Nunavut, Canada
81	Rocque D.A. et Winker K., 2004	Biomonitoring of contaminants in birds from two trophic levels in the North Pacific
82	Wilson H.M. et al., 2004	Concentrations of metals and trace elements in blood of spectacled and king eiders in northern Alaska, USA
83	Savinov V.M. et al., 2003	Cadmium, zinc, copper, arsenic, selenium and mercury in seabirds from the Barents Sea: Levels, inter-specific and geographical differences
84	Wayland M. et al., 2003	Biomarker responses in nesting, common eiders in the Canadian arctic in relation to tissue cadmium, mercury and selenium concentrations
85	Braune B.M et al., 2002	Contaminant residues in seabird eggs from the Canadian Arctic. II. Spatial trends and evidence from stable isotopes for intercolony differences
86	Christopher S.J. et al., 2002	Determination of mercury in the eggs of common murre (Uria aalge) for the seabird tissue archival and monitoring project
87	Stout J.H. et al., 2002	Environmental contaminants in four eider species from Alaska and arctic Russia
88	Wayland M. et al., 2002	Immune function, stress response, and body condition in arctic-breeding common eiders in relation to cadmium, mercury, and selenium concentrations
89	Braune B.M. et al., 2001	Contaminant residues in seabird eggs from the Canadian Arctic. Part I. Temporal trends 1975-1998
90	Wayland M. et al., 2001	Concentrations of cadmium, mercury and selenium in blood, liver and kidney of common eider ducks from the Canadian arctic
91	Wayland M. et al., 2001	Trace elements in king eiders and common eiders in the Canadian arctic
92	Riget F. and Dietz R., 2000	Temporal trends of cadmium and mercury in Greenland marine biota
93	Trust K.A. et al., 2000	Contaminant exposure and biomarker responses in spectacled eiders (Somateria fischeri) from St. Lawrence Island, Alaska
94	Donaldson G.M. et al., 1997	Organochlorine and heavy metal residues in breast muscle of known-age thick-billed murre (Uria lomvia) from the Canadian Arctic
95	Barrett R.T. et al., 1996	Recent changes in levels of persistent organochlorines and mercury in eggs of seabirds from the Barents Sea
96	Dietz R. et al., 1996	Lead, cadmium, mercury and selenium in Greenland marine animals
97	Kim E.Y. et al., 1996	Mercury levels and its chemical form in tissues and organs of seabirds

98	Kim E.-Y. et al., 1996	Metal accumulation in tissues of seabirds from Chaun, northeast Siberia, Russia
99	Wenzel C. and Gabrielsen G.W., 1995	Trace element accumulation in three seabird species from Hornøya, Norway
100	Dietz R. et al., 1990	Organic mercury in Greenland birds and mammals
101	Norheim G. and Kjos-Hanssen B., 1984	Persistent chlorinated hydrocarbons and mercury in birds caught off the west coast of Spitsbergen



**Figure S1.** Hg concentration ( $\mu\text{g/g dw}$ ) in blood, in addition to sample size (into bracket), per country and family. Hg concentrations comparisons have been done between the adults Anatids ( $n=29$ ) and the Larids ( $n=19$ ). Alcids ( $n=4$ ) and the Procellariiformes ( $n=1$ ) have been removed of the statistics because of limited data. Significance in the legend into brackets.



**Figure S2:** Hg concentration ( $\mu\text{g/g dw}$ ) in body feathers, in addition to sample size (into bracket), per country and family. Hg concentrations comparisons have been done between the adults of Alcids ( $n=31$ ), anatids ( $n=8$ ) and Larids ( $n=19$ ). The Hydrobatids ( $n=1$ ) and the Stercorariids ( $n=2$ ) have been removed from the statistical analyses because of limited data. Significance in the legend into brackets.

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