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Akinetes and ancient DNA reveal toxic cyanobacterial recurrences and their potential for resurrection in a 6700-year-old core from a eutrophic lake

Running title: Cyanobacterial recurrence and survival over 6700yr

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1 Abstract

In order to evaluate the recurrence of toxic cyanobacterial blooms and to determine the 2 3 survival capabilities of the resistance cells through time, a sedimentary core spanning 6700 4 years was drilled in the eutrophic Lake Avdat. A multiproxy approach (density, magnetic 5 susceptibility, XRF, pollen and non-pollen palynomorph analyses), was used initially to 6 determine the sedimentation model and the land uses around the lake. Comparison with the 7 akinete count revealed that Nostocales cyanobacteria have been present in Lake Aydat over a six thousand year period. This long-term cyanobacterial recurrence also highlights the past 8 9 presence of both the anaC and mcyB genes, involved in anatoxin-a and microcystin 10 biosynthesis, respectively, throughout the core. The first appearance of cyanobacteria seems to 11 be linked to the natural damming of the river, while the large increase in akinete density 12 around 1800 cal.yr BP can be correlated with the intensification of human activities 13 (woodland clearance, crop planting, grazing, etc.) in the catchment area of the lake, and marks 14 the beginning of a long period of eutrophication. This first investigation into the viability and 15 germination potential of cyanobacteria over thousands of years reveals the ability of intact 16 akinetes to undergo cell divisions even after 1800 years of sedimentation, which is 10 times 17 longer than previously observed. This exceptional cellular resistance, coupled with the longterm eutrophic conditions of this lake, could partly explain the past and current recurrences of 18 19 cyanobacterial proliferations.

20

21 Keywords

22 Sediment, past bloom, akinete, past toxic potential, ancient akinete germination

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1. Introduction

25 Environmental problems caused by cyanobacteria proliferations have increased over 26 the last few decades due to anthropization of watersheds and aquatic systems (Merel et al., 27 2013; Huisman et al., 2018). However, cyanobacteria have existed on Earth for at least 2.5 28 billion years and past proliferations have also occurred naturally far from anthropic pressure 29 (Canfield, 2005). Numerous environmental factors, both physicochemical and biological ones, 30 can drive cyanobacterial proliferations, so the functioning of long-term cyanobacteria 31 dynamics remains difficult to predict. Lacustrine sediments represent a means of 32 reconstituting a lake's history, and benthic sediments are known to store evidence of past 33 aquatic ecosystems, including microbial diversity (Domaizon et al., 2013; Martinez De la 34 Escalera et al., 2014; Savichtcheva et al., 2015; Legrand et al., 2017a). Retro-observations 35 obtained from paleolimnology make it possible to reconstruct the long-term dynamics of 36 populations and to identify the background state and the tipping points for a lake. The two 37 main ways used to trace the long-term dynamics of lacustrine organisms are by means of fossils, when they can be identified, and the analysis of DNA preserved in sediment. Both 38 39 approaches have been applied to cyanobacteria, as certain species, including Nostocales and 40 Stigonematales, can produce resistant cells, termed akinetes. These cells are able to resist unfavorable environmental conditions, such as anoxia, desiccation and lack of light (Kaplan-41 42 levy et al., 2010). Thus, akinetes present in sediments can be used as a record of past 43 nostocalean recurrences (Capo et al., 2015). Some studies have already investigated past fossil akinete distributions in lacustrine sediments (Van Geel et al., 1994; Eilers et al., 2004; Miras 44 45 et al., 2015) and have described nostocalean dynamics over a period of 1950 years calibrated 46 Before Present (cal.yr BP). Investigating these resting cells in ancient sediments also makes it 47 possible to look into their viability and of the idea of resurrection ecology. Wood et al. (2009) 48 have shown the surprising viability of ancient akinetes whose cell germination goes back to

49 120 BP. Legrand et al. (2017a) have also investigated the long-term persistence of akinete 50 integrity and viability in sediment using a 220 year-old core. However, while these findings 51 are promising, very few other studies exist, especially for sediments dating back more than a 52 century, that could help determine the driving factors involved in long-term persistence.

53 On the other hand, ancient cyanobacterial DNA can give relevant and complementary 54 information about the diversity and toxicity of past cyanobacterial proliferations. Based on the 55 detection of genes involved in cyanotoxin biosynthesis, several studies have revealed the 56 recurrence of potentially toxic cyanobacterial taxa throughout the last century in Lake Zurich, 57 involving the mcyA gene (microcystins biosynthesis) (Monchamp et al., 2016), and during the 58 last 220 years in Lake Aydat (France), involving the mcyA and anaC genes (anatoxin-a 59 biosynthesis) (Legrand et al., 2017a). The presence of genes involved in cyanotoxin 60 biosynthesis is probably closely related to the recurrence of toxic cyanobacterial blooms (such 61 as Dolichospermum (i.e Anabaena) in Lake Aydat) over the same period, and the recent 62 increase in anthropic pressure around lakes (Legrand et al., 2017a). While this relationship 63 seems clear in the current context, it is more challenging to establish over a time period of 64 thousands of years, given all the uncertainties with respect to past land use and human 65 activities during that time. Nevertheless, tracing the long-term cyanobacterial dynamics would 66 help to improve our understanding of their driving forces (climatic versus anthropic) and to 67 create predictive models to better understand patterns in species diversity and toxicity.

In this context, it is essential to simultaneously integrate the quantitative characterization of past cyanobacterial proliferations, their past toxic potential and their capacity to survive long term in sediment. For this purpose, we used both microscopic and molecular approaches in order to reconstruct long term nostocalean recurrences in terms of abundance, viability and toxic potential in Lake Aydat (France). We identified the long term akinete recurrence in a sediment core covering 6700 years and went on to characterize the 74 physiological state of akinetes in the sediment, including the identification of a pool of intact 75 akinetes. In these samples, we detected the presence of biosynthetic genes of two cyanotoxins: anatoxin-a and microcystin. The last challenging step was to link these cyanobacterial patterns 76 77 with a number of anthropization indicators already described in Miras et al., (2015) (pollen, 78 non-pollen palynomorphs) in the watershed area, to determine the potential factors driving the 79 long-term recurrence of Nostocales in Lake Aydat. In parallel, in order to test the long-term 80 viability of akinetes and the potential of cyanobacterial resurrection through time, we 81 conducted germination experiments on selected samples.

82 *2.* **Methods:**

83 2.1 Sampling and study site

84 Lake Aydat (maximum depth: 15m; area: 6.105 m²; N45°39.809'/E 2°59.106'/837masl) 85 is a eutrophic naturally dammed lake located in the French Massif Central where recurrent 86 blooms of cyanobacteria have occurred in the late summer and autumn for the last few 87 decades (Gerphagnon et al., 2013b). The lake was created by the damming of the Veyre river 88 by a basaltic lava flow 8551 ± 400 cal.yr BP ago (Boivin et al., 2004). A detailed description 89 of the lake's features (i.e geological and geomorphological setting) is given in Lavrieux et al., 90 (2013a). Coring was performed in the central, deepest part of the lake (Figure 1A). Based on 91 sedimentological features, the record is divided into 2 main units separated by an erosive mass 92 wasting deposit: i) the lower one (1076-1974cm depth which corresponds to 777-1600cm in 93 the master core), and ii) the upper unit (0-829cm depth which corresponds to 0-775cm in the 94 master core) (Lavrieux et al., 2013a). Akinete and molecular analysis were carried out on both 95 the lower and the upper units (Figure 1B) to complement existing pollen, non-pollen 96 palynomorph and diatom analyses (Miras et al., 2015).

97 *Chronology and accumulation rate*

98 Details of the chronology and accumulation rate can be found in Lavrieux et al., (2013a). Briefly, the age-depth model was firstly constructed from ¹⁷AMS radiocarbon dates, based on 99 100 analyses of leaf and wood fragments, distributed through the core. The radiocarbon dating was performed at the Laboratoire de Mesure du ¹⁴Carbone (Gif-sur-Yvette, France) and 101 calibrated (2 σ) using CalPal Online. ¹³⁷Cs; ²⁴¹Am, ²¹⁰Pb and ⁷Be measurements and 102 103 correlation with major historical floods were also used for the chronology of the upper part of 104 the core. The two sedimentary units were respectively dated at between 6700 ± 200 and 3180105 \pm 90 cal.yr BP (lower unit) and 1770 \pm 60 cal.yr BP to present (upper unit), and are separated 106 by an erosive mass wasting deposit. In the lower unit and for the basal part of the upper unit, the mean sedimentation rate was 0.20cm.year⁻¹. There was a sudden increase to 0.52cm.year⁻¹ 107 between 660cm depth and the top, which is similar to the rate of 0.46cm.year⁻¹ estimated by 108 109 Sarazin et al., (1992) for the 50 last years. Flood events caused the sedimentation rate to reach up to 0.6cm.year⁻¹ (Lavrieux et al., 2013a). 110

111 2.2 Akinete extraction and counting

112 Akinetes were extracted from 72 sediment samples (Figure 1C) along the core using 113 protocols from Legrand et al., (2016a). To summarize, 0.5g of fresh sediment was diluted with 4mL of ludox TM-50 (Sigma-Aldrich) and 9.5mL of distilled water. Solutions were sonicated 114 115 and centrifuged. Then, the first 4mL of ludox supernatant were pipeted and homogenized and divided into 2 replicates. Two extractions were performed for each sample to obtain a total of 116 4 replicates. 1µL of Sytox Green® (50mM) was then added to 1mL replicate solution 117 (Gerphagnon et al., 2013a). Each replicate was filtered through 8µm (TEPT filters Merck, 118 119 Milipore, Tullagreen, Ireland). Akinetes were counted in 40 fields per replicate using an 120 epifluorescence microscope (Zeiss Axiovert 200M; HBO 103 mercury vapor short-arc lamp, 121 200x magnification, Oberkochen, Germany). Two counts were performed for each field: i) at 546nm to highlight chlorophyll autofluorescence in akinetes; and ii) at 488 nm to distinguish 122

damaged akinetes using Sytox Green® binding. The SYTOX green, which stained the nucleic 123 124 acids of permeabilized cells after only 30 min of incubation, allowed the damaged akinetes to 125 be clearly identified. In contrast, the intact akinetes (with intact cell walls) were not rendered 126 visible by the SYTOX green staining, but they nevertheless contain photosynthetic pigments 127 that are visible under green light (Legrand et al., 2016b). Akinetes were counted and discriminated based on morphological criteria (Cirés et al., 2013; Legrand et al 2017, a,b; 128 129 Ramm et al., 2017) using taxonomic keys found in reference books (Geitler, 1932). 130 Hormogonia and young filaments observed during germination experiments (see section 2.4) 131 enable the specific identification of 2 akinete morprhotypes: Dolichospermum macrosporum 132 and Dolichospermum flos-aquae (Figure 1B). The observed filaments are the same those observed in studies which focused specifically on these 2 species (Legrand et al., 2016a, 133 134 Legrand et al., 2017, a,b).

135 2.3 *Reviviscence of preserved cells*

136 Akinete germination experiments were performed on 7 samples which were selected for 137 their high intact akinete concentrations and their diversity of akinete morphotypes. Samples 138 were respectively dated to 318 ± 40 , 566 ± 60 , 897 ± 60 , 1633 ± 60 , 1769 ± 60 , 4274 ± 119 139 and 5510 ± 156 cal.yr BP. For the akinete germination experiment, cells were extracted 140 following the same ludox protocol described above, but using 4.5 g of sediment. Supernatants 141 were filtered through a 10µm nylon filtration tissue. This tissue containing the akinetes was then washed with a minimal volume of distilled water. The solution was cleaned through a 142 143 100µm and then a 50µm nylon filtration tissue to remove a maximum of residual sedimentary 144 particles. Finally, the cells were put in 4mL of BG11 medium with cycloheximid (50µg.mL⁻¹). 145 Five replicates of 800µL of solution consisting of akinetes in BG11 medium were placed in a 146 plate with 48 wells using the following procedure. At Days 0, 1 and then every 2 days until 147 day 18, 50μ L of homogenized solution were introduced through a 8μ L TETP filter (Milipore, Tullagreen, Ireland). Intact, germinating akinetes and hormogonia were counted using an
inverse epifluorescence microscope (Zeiss Axiovert 200M; 200x magnification, Oberkochen,
Germany). Counts were performed with an excitation light wavelength of 546nm. Samples
were incubated at 20°C, 6µmol photons.m⁻².s⁻¹ with white light and a photoperiod of 16 hours
of light and 8 hours of dark.

153 2.4 DNA extraction

The same 72 sediment samples described above were submitted to DNA analysis. Genomic DNA from these samples was extracted with the FastDNA[®] Spin kit for soil (MP biomedicals[®]) following the manufacturer's instructions. Precautions laid out in Domaizon et al. (2017) concerning ancient DNA were applied. A control extraction was performed and checked with all samples in order to verify that there was no contamination from modern DNA.

The DNA concentration in each sample was estimated using a UV spectrophotometer
 (Nanodrop[®] ND 2000), and samples were kept at -20°C until further analysis.

162 **2.5** *Target gene amplification*

Three genes were targeted by classic or nested PCR during this study. The 163 164 cyanobacterial 16S rRNA gene was amplified in sediment samples using the cya359F and 165 cva781R set of primers (Nubel et al., 1997) (Table 1). Then one gene involved in the biosynthesis of anatoxin-a, anaC, and one of microcystin, mcyB gene, were also investigated. 166 167 The first was detected with the sets of primers anxgen-F2/R and anaC 80F/706R for nested 1 and 2 PCR respectively. For mcyB nested PCR, MCY F1/R1 then MCY F1/R2 were used. 168 PCR mixtures were the same as in Legrand et al., (2016a) for the two anaC nested PCRs. For 169 use with mcyB, the mix for both PCRs was composed of 5µL of colorless Go Taq[®] felxi 5X 170 171 Buffer, 2.5mM of MgCl₂ solution, 0.2mM of dNTPs, 0.5µM of each primer (Table 1), 1

mg.mL⁻¹ of Bovine Serum Albumin (BSA), and 1.5 U of GoTag[®] G3 Hot Start Polymerase in 172 173 a final volume of 25µL. The PCR programs used were referenced in Legrand et al., (2016 a,b) 174 for *anaC*. For *mcyB*, the program for PCR 1 was composed of an initial hybridization step of 94°C for 5 min, 35 cycles of 94°C for 30 sec, 50°C for 1 min, 72°C for 30 sec, and a final 175 176 extension of 72°C for 5 min, and for PCR 2 of an initial hybridization step of 94°C for 5 min, 35 cycles of 94°C for 30 sec, 54°C for 1 min, 72°C for 20 sec, and a final extension of 72°C 177 for 5 min. All PCR products were revealed with 0.5% agarose gel with 0.3 mg.L⁻¹ ethidium 178 179 bromide and migrated in a TAE buffer 0.5x at 100V for 30 min. Gels were put under UV light 180 and investigated using Gel Imager (ScienceTech) with the software VisionWorksLS vers.6.5 (Upland, USA). Band localization and intensity in electrophoresis gels were defined with 181 182 GelAnalyser vers. 2010 (http://www.gelanalyzer.com/) (Legrand et al., 2016a). For anaC, 183 only those PCR products which had been digested with the restriction enzyme EcoRV 184 (Biolabs, Ipswich, USA), following manufacturer's instructions, were considered as being 185 positive. This gene possesses a restriction site for this enzyme which reinforces the detection 186 of the expected gene (Figure S1). Furthermore, several samples have been sequenced to confirm the amplification of the expected genes (Legrand et al., in prep). Some of these 187 188 sequences have been deposited in the GenBank-EMBL database under the accession numbers: MK646071-MK646074 and MK636679-MK636688. 189

190 *3.* **Results**

191 3.1 Total akinete abundance through the core

The total akinete distribution was heterogeneous through the core and can be divided into two parts based on the two major units (Figure 2): i) the lower unit, between 6700 ± 200 and 3180 ± 90 cal.yr BP) is characterized by low akinete abundance, with between 197 and 97 900 akinetes per gram of dried sediment (gDW⁻¹) with a mean value of 12 300

akinetes.gDW⁻¹ of sediment; and ii) the upper unit (from 1770 ± 60 cal.yr BP to present) is 196 characterized by a significantly higher mean value of akinetes corresponding to 52 600 197 akinetes.gDW⁻¹ of sediment. However, the upper part also shows a greater fluctuation, from 198 22 000 akinetes.gDW⁻¹ (between 1391 \pm 50 and 670 \pm 50 cal.yr BP) to around 200 000 199 akinetes.gDW⁻¹ of sediments (in the sections comprised between 1769 ± 60 and 1507 ± 40 200 201 cal.yr BP, and 107 ± 40 and 10 ± 40 cal.yr BP). Four akinete morphotypes were counted along 202 the core corresponding respectively to Dolichospermum macrosporum, Dolichospermum flos-203 aquae, Dolichospermum sp. and Aphanizomenon sp. (Figure 1B). The vertical distribution of these four morphotypes was different, for example D. macrosporum was particularly 204 dominant in the upper unit where it composed up to 100% of the akinete morphotypes (Figure 205 206 S2).

207 It can be noted that the total akinete abundance and the D. macrosporum akinete 208 abundance are very similar (SC= 0.92, p < 0.001). In the lower unit, D. macrosporum akinetes were often absent with values no higher than 15 500 akinetes.gDW⁻¹ of sediment. D. flos-209 aquae was almost always present through the core, but its akinete abundance was generally 210 low, with between several hundred and a few thousand cells.gDW⁻¹ of sediment. 211 212 Nevertheless, D. flos-aquae akinete abundance reached occasionally high values at 1339, 213 1054, 840 and 767cm depth, levels respectively dated to 5510 ± 156 , 4274 ± 119 , 3421 ± 97 214 and 1728 ± 60 cal.yr BP, with respective concentrations of 76 400, 42 300, 23 600 and 97 400 cells.gDW-1 of sediment. Finally, Dolichospermum sp. and Aphanizomenon sp. remained at 215 216 low concentrations and presented a similar pattern: akinete presence of these 2 species was higher in the lower part, where they were found in 10 and 18 samples respectively. Akinete 217 218 abundance of these 2 species was also generally low, between several hundred and several thousand, with a slight peak of 23 300 akinetes.gDW⁻¹ for Aphanizomenon sp. at 4648 ± 129 219 220 cal.yr BP.

222 Overall, the number of intact akinetes in the lower unit, where D. macrosporum was not 223 dominant, followed the same pattern as that of total akinetes (SC= 0.70 p < 0.0001) (Figure 2). The number of intact akinetes was higher in the lower unit than in the upper unit, with a mean 224 number of 8 620 and 3 300 cells.gDW⁻¹ of sediment respectively. However, there is no 225 statistical difference due to large variations (from 0 to 86 200 akinetes.gDW⁻¹ of sediment) 226 especially in the lower unit (Figure 2). These large variations are the result of numerous intact 227 akinete peaks, especially for D. flos-aquae. For example, the number of intact cells for this 228 species reached 41 500 and 76 400 akinetes.gDW⁻¹ at 4274 \pm 119 and 5510 \pm 156 cal.yr BP 229 respectively. These variations seem to reflect a significant difference in integrity related to the 230 231 akinete morphotype (Kruskal-Wallis test p <0.0001) (Figure 3). The mean percentages of 232 intact akinetes through the core for D. macrosporum, D. flos-aquae, Dolichospermum sp. and Aphanizomenon sp. are 5.1, 54.1, 90.1, and 83.7% respectively. In addition, high temporal 233 234 variations in the percentage of intact akinetes were observed, especially for D. macrosporum 235 and D. flos-aquae (Figure 3). The variations for D. flos-aquae are marked between the two units (with a mean of 35 and 72% intact akinetes in the upper and lower unit respectively 236 237 (Mann-Whitney test, p=0.002). The percentage of *D. macrosporum* is also lower in the upper 238 unit, where the percentage of intact akinetes is low, between 0 and 13%. In the lower unit, 239 there are alternations between an absence of this morphotype and a percentage of intact 240 akinetes mostly comprised between 10 and 63%.

241

1 3.3 Potential of reviviscence of past cyanobacteria

At the beginning of the experiment, the number of akinetes.mL⁻¹ in BG11 was very different from one sample to another (Figure 4). *D. macrosporum* was largely dominant in 318, 566 and 1663 cal.yr BP whereas *D. flos-aquae* was dominant in 897 and 1769 cal.yr BP

samples. Akinete distribution was heterogenous in 4274 and 5510 cal.yr BP samples. Four of 245 246 the 7 studied samples showed no evidence of akinete germination after 18 days incubation (Figure 4B). Some germinating akinetes were observed in the samples dated to 318 ± 40 247 cal.vr BP and 1769 \pm 60 cal.vr BP after respectively 3-5 and 7-13 days after the beginning of 248 249 the experiment. All the germinating akinetes corresponded to *Dolichospermum macrosporum* 250 for 318 \pm 40 cal.yr BP and to Dolichospermum flos-aquae for 1769 \pm 60 cal.yr BP. These germinating akinetes were observed with 2 or 4 cells present inside the envelope (Figure 4) 251 252 but no young filaments were observed. The ratio of germinating akinetes to total intact 253 akinetes was very low, and varied between 6 and 8% for the 318 ± 40 cal.yr BP sample, and 254 between 0.31 and 1.23% for the 1769 \pm 60 cal.yr BP sample (Figure 4B). Some D. flos-aquae short filaments of a few cells in length (hormogonia) were observed in the sample dated to 255 897 ± 50 cal.yr BP after between 3 to 5 days of incubation. This percentage of young 256 257 filaments remained relatively low, between 6 and 8%.

258 3.4 Target gene detection in sediment

The 16 rRNA gene was detected in all the studied samples through the core. The *mcyB* gene was detected with nested PCR in 19 samples through the entire core (Figure 5): it was detected in several samples of the lower unit (from 6702 ± 195 to 3269 ± 93 cal.yr BP) and also in the upper unit (at 1316 ± 40 cal BP). This gene recurred at the top of the core with the 12 more recent samples being positive (from 422 ± 40 to 10 ± 40 cal.yr BP). The *anaC* gene was found in 11 samples, mostly located in the lower unit but also in recent layers from $227 \pm$ 40 cal.yr BP.

266 *4*. **Discussion**

267 4.1 *Reconstruction of long-term nostocalean recurrences in Lake Aydat*

268 The overall objective of this study was to characterize past recurrences of toxigenic 269 nostocalean cyanobacteria in a 6700-year-old sediment core from Lake Aydat (France). 270 Previously, the oldest evidence of akinete persistence had been found in Lake Gosciaz 271 (Poland), with akinetes approximately 4000 years old (Van Geel, 1994) So, to our knowledge, 272 our study is the first report of such a long-term recurrence of four morphotypes of akinetes 273 belonging to two genus, *Dolichospermum* and *Aphanizomenon*. These 4 morphotypes have 274 already been observed in the water column of Lake Aydat (Sabart et al., 2015; Lafforgue et 275 al., 1995). Moreover, akinetes from Dolichospermum have already been observed in surface 276 sediment (Legrand et al., 2016a), in a 220-year-old core (Legrand et al., 2017a) and in the 277 same 6700-year-old core, but using a palynological approach (Miras et al., 2015). Total 278 akinete profiles from Miras et al. (2015) are very similar to our results in the upper unit of the 279 core and confirm the recent nostocalean bloom recurrences. However, Miras et al. (2015) did 280 not observe akinetes in the lower unit, which suggests that it is possible to achieve a better 281 resolution of akinete extraction and identification with ludox and Sytox-Green protocols than 282 with the classical palynological method traditionally used in paleoecological research (Van 283 Geel et al., 1994, 1996; Revelles et al., 2016). Standard palynological treatment of samples, 284 which often includes acetolysis (Faagri and Iversen, 1989), seems thus to induce a loss of 285 akinete material and must consequently be avoided for the characterization of the long-term 286 evolution of akinetes. Our method using Sytox green also makes it possible to evaluate the 287 number of intact akinetes present in the sediment, which provides fresh insight into the 288 understanding of the long-term resistance of nostocalean blooms.

289 4.2 Establishment of baseline conditions at Lake Aydat

One surprising result was to find sizeable concentrations of akinetes in the oldest part of the core, between ca 6500 and 6000 cal.yr BP. This suggests that only around 2 millennia after its formation, the baseline water conditions at Lake Aydat, defined as conditions prevailing 293 prior to human disturbances, were characterized by high nutrient levels (Miras et al., 2015), 294 which is common for lakes formed by the damming of a river (Smol, 2014). This is also 295 supported by other paleo-ecological bio-indicators such as different non-pollen palynomorphs 296 (mainly Turbellaria (Haas, 1996), HdV-179 (Van Geel et al., 1983), and diverse rotifer resting 297 eggs such as Anuraeopsis fissa-type, Filinia longiseta-type and Keratella-type (Barbiero et 298 al., 2011; Haberman and Haldna, 2014) and diatom data (Stephanodiscus medius, S. minutulus 299 and S. parvus) which are recorded as early as 6700 cal.yr BP (Miras et al., 2015). 300 Stephanodiscus species are encountered in lakes with elevated nutriment concentrations and 301 are typically found in eutrophic or hypertrophic conditions (Kirilova et al., 2009). This high 302 trophic status is mainly driven by an influx of nutrients carried by the water flow into the 303 drainage basin, which has been very vegetation-rich in Aydat since 6700 cal.yr BP. Pollen and 304 non-pollen palynomorph data clearly indicate that woods (mainly constituted of beech and 305 oak) were already present around the lake at this time (Miras et al., 2015).

306 4.3 Chronological comparison between akinete concentrations and human impact in the 307 catchment basin

308 The oldest significant increase in akinete concentrations, dating from 5500 cal.yr BP, is coincident with the first palynological evidence of middle Neolithic human activity in the 309 310 area (Miras et al., 2015). Grazed pastures and small-scale cultivated fields were created within 311 the progressively more open wooded landscape of the Aydat catchment area (Figure S3). The 312 combined effect of this human impact with a climatic trend toward cooler and more humid 313 conditions, especially around 5600-5300 cal.yr BP (Magny et Hass, 2004), may have 314 increased soil erosion on slopes within the catchment area. Higher detrital input may have 315 consequently induced the increase in the trophic state of Lake Aydat, which is demonstrated 316 by non-pollen palynomorphs (NPP) and the presence of eutrophic diatoms (Miras et al., 2015). This direct relationship between a high number of akinetes and phases of human 317

318 activity is also found in subsequent periods. High concentrations of akinetes observed in the 319 lower unit at 4648 \pm 129, 4274 \pm 119 and 3421 \pm 97 cal.yr BP are close to, or concomitant with, different episodes of human activity (both temporary farming and agriculture 320 321 accompanied by woodland clearances) in the Avdat catchment area, as revealed by pollen and 322 non-pollen palynomorph data (Miras et al., 2015). The same trend is observed for the upper 323 unit, with an increase in anthropogenic pressure and diversification of land use. High amounts 324 of akinetes between 1769 \pm 60 and 1507 \pm 40 cal.yr BP and between 670 \pm 50 cal.yr BP and 325 the sub-recent period (639 \pm 50 cal.yr BP) are respectively encountered during phases of 326 significant deforestation and diversification of human practices, such as the introduction of 327 hemp retting (Lavrieux et al., 2013b). Hemp needs a large amount of nitrogen, phosphorus and potassium (Barron et al., 2012) and its cultivation in the catchment area, and in particular 328 329 immediately surrounding Lake Aydat, could have caused an increase in nutrient loading into 330 the lake through soil leaching. Moreover, hemp cultivation needs several periods of retting in 331 the lake water, generally during May, August and October. These periods may induce nutrient 332 enrichment in the water column that can be directly used by cyanobacterial proliferations. 333 Thus, it is already well known that hemp retting provokes eutrophication of lakes and pools 334 (Van Geel et al., 1994). This process can also be reinforced through the gradual development 335 of an agricultural and grass-rich landscape. The recurrence of eutrophic conditions is also 336 confirmed by the increasing number of diatoms through the core, related to higher phosphorus 337 concentrations estimated in the upper unit relative to the lower unit (Miras et al., 2015). Eutrophic species such as Aulacoseira subarctica and Fragilaria recta, that require a high 338 339 phosphorus content, become more dominant towards the top of the core (Kauppila, 2006). 340 The dominance of *D. macrosporum* in this upper unit seems to attest to monospecific blooms 341 and confirms stable eutrophic conditions. However, in this upper unit, there are strong 342 temporal variations in akinete concentration, despite the consistently high number of diatoms,

343 which implies high total phosphorus concentrations. Other environmental parameters may 344 thus have affected cyanobacterial growth, such as water column instability, light intensity and water temperature. For example, on the basis of sedimentological parameters, the period from 345 346 1391 ± 50 to 670 ± 50 cal.yr BP is mainly characterized by an increase in the amount of 347 detrital input and thus in the frequency of flooding events (Lavrieux et al., 2013a,b). It is 348 noteworthy that a doubling of the sedimentation rate since 1200 cal.yr BP took place at the 349 end of a wetter period, in a context of high anthropogenic pressure. Recurrent flooding may 350 have had an effect on the water column stability, creating unfavorable conditions for the 351 development of cyanobacterial biomass (Mitrovic et al., 2003; Lewis et al., 2011; Paerl et al., 352 2013). Nevertheless, the persistence of long-term high trophic conditions allowed 353 cyanobacteria to regrow as soon as conditions become favorable again, as can be observed at 354 the end of the Medieval period and during Modern Times.

355 4.4 Long term akinete integrity: persistence and germination potential

356 Very few studies have considered the potential viability of ancient cyanobacterial cells. 357 Rücker et al., (2009) and Ramm et al. (2012, 2017), focused on akinete integrity but were not 358 able to discriminate intact akinetes from lysed ones. On the other hand, Legrand et al. (2017b) 359 characterized akinete integrity in sediments, but only using the previous year's sediment 360 surface. So, to our knowledge, this study is the first one to investigate the viability and 361 germination potential of cyanobacteria over thousands of years. Interestingly, in Lake Aydat, 362 the percentage of intact akinetes of D. macrosporum are in the same range at the sediment 363 surface (around 7%) (Legrand et al., 2017b) and the upper unit of the core (a mean of 5%). 364 This suggests that akinete degradation does not increase with time spent in sediment. Instead, 365 the percentage of intact benthic akinetes seems to be linked to processes occurring in the 366 water column and is thus representative of their state at the end of their sedimentation out of 367 the water column. This hypothesis may also be confirmed by the higher percentage of 368 integrity of akinetes in the lower units, especially given the presence of *Dolichospermum flos*-369 aquae, Dolichospermum sp. and Aphanizomenon sp. These species composed between 60 and 370 100 % of the intact akinetes, whereas D. macrosporum (dominant in the upper part) had a 371 mean of 6%. This suggests that the different species had different resistance capacities as also 372 suggested in Legrand et al., (2016b). The latter study showed that akinetes from D. 373 macrosporum were less resistant than those of D. flos-aquae. In addition, the different species 374 are known to have varying abilities to produce akinetes. Thus, some species produce a large 375 number of akinetes, but which are less resistant (such as D. macrosporum), while others 376 produce less akinetes, but ones that are more resistant to unfavorable environmental 377 conditions (such as *D. flos-aquae*). These differences between species can be integrated into a 378 broader ecological strategy allowing each species to adapt to its preferred environment.

379 Despite the variable percentages of integrity, the number of intact akinetes able to 380 germinate remained extremely low in the oldest sediment (1% maximum) compared to 381 surface sediments (up to 70%) (Tjsujimura et al., 2003; Legrand et al., 2017b). Even though 382 old akinetes might have retained an intact envelope, they probably lost their potential for 383 reviviscence with time. Current germination has been recorded after a storage time of 64 years 384 (Livingstone and Jaworski, 1980) and 120 cal.yr BP (i.e. around 180 years) (Wood et al., 385 2009), but not longer. Nevertheless, in our study, some intact akinetes from D. flos-aquae 386 showed the ability to undergo cell divisions after 1830 years of sedimentation (1769 cal.yr 387 BP), which is 10 times more than that previously observed in the literature. Therefore, while 388 the culture conditions and/or cellular capacity were able to produce hormogonia only in 1 389 sample (897 cal.yr BP) (i.e. a young filament (Hence and Beckmann, 2006)), our results show 390 that akinetes can remain alive 1800 years after their formation. Interestingly, these results may be due to the differential loss of viability depending on the species. For example, D. 391 392 macrosporum only presents germinating akinetes in the more recent samples, whereas the 393 number of intact akinetes of this species was higher in older samples. On the contrary, D. flos-394 aquae filaments were observed in samples where intact akinete concentration was very low. 395 The only presence of germinating akinetes without visible filament formation may be due to a 396 loss of viability and/or to inappropriate culture conditions. Thus, samples which contain only 397 germinating akinetes may represent the limit of viability for a species. This indicates that D. 398 flos-aquae akinetes may be able to stay viable longer than D. macrosporum akinetes. This 399 hypothesis may confirm observations in Legrand al. (2016a, 2017a,b) that akinete integrity 400 could reflect different ecological strategies between theses 2 species .

Finally, this work highlights the unexpected viability of akinetes through hundreds of years and may be the starting point for other research focusing on the isolation and characterization of ancient cyanobacterial strains. Knowing the host of conditions favorizing akinetes germination (Kaplan-levy et al., 2010; Sukenik et al., 2019), other investigations will be necessary to test the long-term viability of akinetes of other cyanobacterial species.

406 4.5 Cyanotoxin genes detection as a marker of past toxic genotypes.

407 This study also highlighted the past recurrences of the anaC and mcyB genes which are 408 involved in anatoxin-a and microcystin biosynthesis, respectively. For the first time, results revealed the presence of these genes back to 6700 cal yr. BP. Previously, the presence of the 409 410 mcyA gene had only been demonstrated from the last century (Monchamp et al., 2016), and 411 from 220 years ago for the anaC gene and mcyA (Legrand et al., 2017a). So, the presence of 412 these genes in Lake Aydat provides the first long-term dynamics of both microcystin and 413 anatoxin genotypes. Interestingly, anaC and mcyB genes are both detected in one of the oldest 414 samples, 6702 ± 195 cal.yr BP, revealing an unexpected co-occurrence, or at least a close succession, of genotypes carrying these two genes. Numerous studies have already 415

416 highlighted the co-occurrence of these two cyanotoxins in the water column (Park et al., 1993,

417 1997; Pawlik-Skowronska, 2004) including at Lake Aydat (Sabart et al., 2015)

418 Nevertheless, except for short periods of co-occurrence, anaC and the mcy genes are 419 mainly found independent of each other. The anaC gene was mostly observed in deep 420 sediment (from 5674 \pm 161 to 4802 \pm 134 cal.yr BP) rather than in surface sediments, 421 whereas mcyB has been recurrent over at least the last 500 years. Monchamp et al. (2016) 422 have linked the presence of mcyA gene from 1972 to the presence of Planktothrix rubescens. 423 The same applies to *Microcystis* and the detection of *mcyA* in a sediment core spanning 30 424 years from Lake Erie, USA (Rinto-Kanto et al., 2009). Aphanizomenon and Dolichospermum, 425 genera found throughout the core from Lake Aydat, are both able to produce these 426 cyanotoxins (Rapala et al., 1993, 1997; Ballot et al., 2010; Bernard, 2014). Furthermore, mcyA gene and anaC have already been associated with the recurrence of D. macrosporum 427 428 and D. flos-aquae in Lake Aydat (Legrand et al., 2017a). So, the Nostocalean species may be 429 at least partly responsible for these toxic patterns found through the studied core. 430 Nevertheless, we were not looking for, and have no evidence of, the presence of non-431 Nostocalean groups in our sediment core. Microcystis, Woronichinia and Pseudanabaena have already been observed in the Lake Aydat water column (Sabart et al., 2015; Lafforgue et 432 433 al., 1995) and are potential producers of the studied cyanotoxins (Bernard, 2014). So, it is also 434 possible that these non-nostocalean cyanobacteria are also partly responsible for the toxic potential detected in the core. First results from sequencing of *anaC* seem to be in accordance 435 436 with a combination of several species possessing this gene. This analysis from 4000-year-old 437 sediment highlighted that the Oscillatoria and Aphanizomenon genera possessed this gene (Figure S4). More investigations are required to identify a formal link between past 438 439 nostocalean biomass and the presence of genes which could create cyanotoxins.

440 5. Conclusion

Cyanobacteria are known to have been on Earth for more than 2.5 billion years, but 441 442 much evidence suggests that harmful cyanobacterial blooms have been more recurrent and 443 frequent on a global scale over the last few decades (Huisman et al., 2018). Our results 444 confirm this trend and show a long recurrence of cyanobacteria through time, with the oldest 445 noticeable concentration of akinetes at 6700 cal.yr BP. So the presence of cyanobacteria in 446 Lake Aydat is not restricted to the modern period, but originates very early on in the history of 447 the lake. The natural damming of the river probably provided the first favorable conditions for 448 their development, which was subsequently accentuated by human activity (woodland 449 clearance, crop planting, grazing around the lake, etc.) and diverse land use through time 450 (hemp culture and retting). The next step is to characterize the ancient cyanobacterial strains 451 through this long time period, using 16S phylogenetic analysis, to find out more about the total diversity of cyanobacteria, not just Nostocaleans. Demonstrating the akinetes' potential 452 453 to germinate and the exceptional longevity of these cells in sediments (over 1800 years) 454 illustrates the huge in situ stock of viable akinetes able to germinate again and to sustain 455 future cyanobacterial blooms. The long-term recurrence of cyanobacteria, with the dominance 456 of the same species over the last 1800 years, reveals the that eutrophic conditions have been maintained over a long time period, illustrating the current difficulty to mitigate 457 eutrophication. Thus, sedimentary archives constitute a good way to improve our 458 459 understanding of the dynamic functioning of individual lakes and also to indicate means of 460 management adapted to their life history and cyanobacterial sedimentary stock.

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468

469 **Conflict of Interest**

470 The authors declare no conflict of interest.

471 **References**

472	Ballot, A., Fastner	r, J., Lentz, M.	and Wiedner, C.	2010. First Rep	port of Anatoxin-a-
473	Producing	Cyanobacterium	Aphanizomenon	Issatschenkoi	in Northeastern
474	Germany. To	oxicon. 56 (6): 964	-71. https://doi.org	g/10.1016/j.toxico	on.2010.06.021.

475 Barbiero, R. P., Lesht, B. M., and Warren, G.J. 2011. Evidence for bottom–up control of
476 recent shifts in the pelagic food web of Lake Huron. Journal of Great Lakes Research.
477 37(1), 78-85.

Barron, J.A., Bukry, D., Field, D.B., and Finney, B. 2013. Response of diatoms and
silicoflagellates to climate change and warming in the California Current during the
past 250 years and the recent rise of the toxic diatom *Pseudo-nitzschia australis*.
Quaternary international. 310, 140-154.

- 482 Bernard C., 2014. Les Cyanobactéries et Leurs Toxines. RFL Revue Francophone Des
 483 Laboratoires 2014. (460): 53–68. https://doi.org/10.1016/S1773-035X(14)72405-0.
- 484 Canfield, D.E., 2005. The early history of atmospheric oxygen: Homage to Robert M.
 485 Garrels. Annual Review of Earth and Planetary Sciences. 33 (1): 1–36.
 486 https://doi.org/10.1146/annurev.earth.33.092203.122711.
- 487 Capo, E., Debroas, D., Arnaud, F., and Domaizon, I., 2015. Is Planktonic Diversity Well
 488 Recorded in Sedimentary DNA? Toward the Reconstruction of Past Protistan
 489 Diversity. Microbial ecology. 865–75. https://doi.org/10.1007/s00248-015-0627-2.
- 490 Cirés, S., Lars, W., Agha, W., and Quesada, A., 2013. Overwintering Populations of
 491 *Anabaena, Aphanizomenon* and *Microcystis* as Potential Inocula for Summer Blooms.
- 492 Journal of Plankton Research. 35: 1254–66. https://doi.org/10.1093/plankt/fbt081.

493	Domaizon, I., Savichtcheva, O., Debroas, D., Arnaud, F., Villar, C., Pignol, C., Alric, B., and
494	Perga, M., 2013. DNA from Lake Sediments Reveals the Long-Term Dynamics and
495	Diversity of Synechococcus Assemblages. Biogeosciences Discussions. 10 (2): 2515-
496	64. https://doi.org/10.5194/bgd-10-2515-2013.

- 497 Domaizon, I., Winegardner, A., Capo, E., Gauthier, J., and Gregory-Eaves, I., 2017. DNA498 Based Methods in Paleolimnology: New Opportunities for Investigating Long-Term
 499 Dynamics of Lacustrine Biodiversity. Journal of Paleolimnology. 58 (1).
 500 https://doi.org/10.1007/s10933-017-9958-y.
- 501 Eilers, J., Kann, J., Cornett, J., Moser, K., and Amand, A., 2004. Paleolimnological Evidence
- of Change in a Shallow, Hypereutrophic Lake : Upper Klamath Lake , Oregon , USA.
 Hydrobiologia. 520 (1-3):7–18.
- Faagri, K., and Iversen, J., 1989. Textbook of pollen analysis, 4th edn, by Faegri K, Kaland
 PE, Krzywinski K.
- 506 Geitler, L. 1932. Cyanophyceae. New York;
- Gerphagnon, M., Latour, D., Colombet, J., and Sime-Ngando, T., 2013a. A Double Staining
 Method Using SYTOX Green and Calcofluor White for Studying Fungal Parasites of
 Phytoplankton. Applied and Environmental Microbiology 79 (13): 3943–51.
 https://doi.org/10.1128/AEM.00696-13.
- 511 Gerphagnon M, Latour D, Colombet J, Sime-Ngando T., 2013b. Fungal Parasitism: Life
 512 Cycle, Dynamics and Impact on Cyanobacterial Blooms. PloS One 8 (4): e60894.
 513 https://doi.org/10.1371/journal.pone.0060894.

Haas, J. N., 1996. Neorhabdocoela oocytes—palaeoecological indicators found in pollen
preparations from Holocene freshwater lake sediments. Review of Palaeobotany and
Palynology. 91(1-4), 371-382.

- Haberman, J., and Haldna, M., 2014. Indices of zooplankton community as valuable tools in
 assessing the trophic state and water quality of eutrophic lakes: long term study of
 Lake Võrtsjärv. Journal of Limnology, 73(2).
- Hense, I., and Beckmann, A., 2006. Towards a Model of Cyanobacteria Life Cycle Effects
 of Growing and Resting Stages on Bloom Formation of N 2 -Fixing Species.
 Ecological Modeling. 195: 205–18. https://doi.org/10.1016/j.ecolmodel.2005.11.018.
- Huisman, J., Codd, G. Paerl, H.W., Ibelings, B.W., Verspagen, Visser, P., 2018.
 Cyanobacterial Blooms. Nature Reviews Microbiology. 16 (8): 471–83.
 https://doi.org/10.1038/s41579-018-0040-1.
- Kaplan-levy, R., Hadas, O., Summers, M.L, Rücker, J. and Sukenik, A., 2010. Akinetes:
 Dormant Cells of Cyanobacteria. Dormancy and Resistance in Harsh Environments.
 Springer B (November 2014): 5–27. https://doi.org/10.1007/978-3-642-12422-8.
- Kauppila, T., 2006. Sediment-Based Study of the Effects of Decreasing Mine Water
 Pollution on a Heavily Modified, Nutrient Enriched Lake. Journal of Paleolimnology,
 35(1), 25-37.
- Kirilova, E., Heiri, O., Enters, D., Cremer, H., Lotter, A. F., Zolitschka, B., and Hübener, T.,
 2009. Climate-induced changes in the trophic status of a Central European lake.
 Journal of Limnology, 68(1), 71-82.

535	Lafforgue, M., Szeligiewicz, W. Devaux, J., and Poulin. M., 1995. Selective Mechanisms
536	Controlling Algal Succession in Aydat Lake. Water Science and Technology 32(4):
537	117-27. https://doi.org/10.1016/0273-1223(95)00688-5.

- 538 Lavrieux, M., Disnar, J. Chapron, E., Bréheret, J.G., Jacob, J., Miras, Y., Reyss, J.L, Andrieu-
- Ponel, V., and Fabien, A., 2013a. 6,700-Year Sedimentary Record of Climatic and
 Anthropic Signals in Lake Aydat (French Massif Central) Signals in Lake Aydat
 (French Massif Central). The Holocene, 23(9): 1317-1328.
- 542 Lavrieux, M., Jacob, J., Disnar, J., Bréheret, J., Milbeau, C., Miras Y., et al., 2013b.
 543 Sedimentary cannabinol tracks the history of hemp retting. Geology.41(7):751–5.
- Legrand, B., Lamarque, A., Sabart, M., and Latour, D., 2016a. Characterization of Akinetes
 from Cyanobacterial Strains and Lake Sediment: A Study of Their Resistance and
 Toxic Potential. Harmful Algae. 59: 42–50. https://doi.org/10.1016/j.hal.2016.09.003.
- Legrand, B., Lesobre, J., Colombet, J., Latour, D., and Sabart, M., 2016b. Molecular Tools to
 Detect Anatoxin-a Genes in Aquatic Ecosystems: Toward a New Nested PCR-Based
 Method. Harmful Algae. 58: 16–22. https://doi.org/10.1016/j.hal.2016.07.002.
- Legrand, B., Lamarque, A., Sabart, M., Latour, D., 2017a. Benthic Archives Reveal
 Recurrence and Dominance of Toxigenic Cyanobacteria in a Eutrophic Lake over the
 Last 220 Years. Toxins 9(9): 271. https://doi.org/10.3390/toxins9090271.
- Legrand, B., Le Jeune, A.H., Colombet, J., Thouvenot, A., Latour, D., 2017b. Akinetes May
 Be Representative of Past Nostocalean Blooms: A Case Study of Their Benthic
 Spatiotemporal Distribution and Potential for Germination in a Eutrophic Lake.
 Applied and Environmental Microbiology 83(23): 1–14.
 https://doi.org/10.1128/AEM.01571-1

558	Lewis, W. M., Wurtsbaugh, W.A, and Paerl, H.W., 2011. Rationale for Control of
559	Anthropogenic Nitrogen and Phosphorus to Reduce Eutrophication of Inland Waters.
560	Environmental Science and Technology 45: 10300–305.

- 561 Livingstone, D., and Jaworski, G.H.M., 1980. The Viability of Akinetes of Blue-Green Algae
 562 Recovered from the Sediments of Rostherne Mere. British Phycological Journal. 15
- 563 (4): 357–64. https://doi.org/10.1080/00071618000650361.
- Magny, M., 2006. Holocene fluctuations of lake levels in west-central Europe: methods of
 reconstruction, regional pattern, palaeoclimatic significance and forcing factors. 13891399
- Magny, M., and Haas, J.N., 2004. A major widespread climatic change around 5300 cal. yr BP
 at the time of the Alpine Iceman. Journal of Quaternary Science, 19(5), 423-430.
- Martinez De La Escalera, G., Antoniades, D., Bonilla, S., and Piccini, C., 2014. Application
 of Ancient DNA to the Reconstruction of Past Microbial Assemblages and for the
 Detection of Toxic Cyanobacteria in Subtropical Freshwater Ecosystems. Molecular
 Ecology. 23(23): 5791–5802. https://doi.org/10.1111/mec.12979.
- Merel, S., Walker, D., Chicana, R., Snyder, S., Baurès, E., and Thomas, O., 2013. State of
 Knowledge and Concerns on Cyanobacterial Blooms and Cyanotoxins. Environment
 International 59: 303–27. https://doi.org/10.1016/j.envint.2013.06.013.
- 576 Miallier, D., Pilleyre, T., Sanzelle, S., Boivin, P., and Lanos, P., 2012. Revised chronology of
 577 the youngest volcanoes of the Chaîne des Puys (French Massif Central). Quaternaire.
 578 Revue de l'Association française pour l'étude du Quaternaire, 23(4), 283-290.
- 579 Miras, Y., Beauger, A., Berthon, V., and Ledger, P.M., 2015. Tracking Long-Term Human
 580 Impacts on Landscape, Vegetal Biodiversity and Water Quality in the Lake Aydat

- 581 Catchment (Auvergne, France) Using Pollen, Non-Pollen Palynomorphs and
 582 Diatoms Palaeogeography Palaeoclimatology Palaeoecology 424:76–90.
 583 https://doi.org/10.1016/j.palaeo.2015.02.016.
- 584 Mitrovic, S. M., Oliver, R.L., Rees, C., Bowling, L.C, and Buckney, R., 2003. Critical Flow
- Velocities for the Growth and Dominance of *Anabaena Circinalis* in Some Turbid
 Freshwater Rivers. Freshwater Biology 48: 164–74.
- Monchamp, M.E., Walser, J.C, and Pomati, F., 2016. Sedimentary DNA Reveals
 Cyanobacterial Community Diversity over 200 Years in Two Perialpine Lakes. 82(21):
 6472–82. https://doi.org/10.1128/AEM.02174-16.Editor.
- Nonneman, D., and Zimba, P.V., 2002. A {PCR}-Based Test to Assess the Potentional for
 Microcystin Occurence in Channel Catfish Production Ponds. Journal of Phycology 38
 (1): 230–33. http://onlinelibrary.wiley.com/doi/10.1046/j.15298817.2002.01138.x/full.
- Nubel, U, Muyzer, G, Garcia-pichel, F., and Muyzer, G., 1997. PCR Primers to Amplify 16S
 RRNA Genes from Cyanobacteria PCR Primers To Amplify 16S RRNA Genes from
 Cyanobacteria. Microbiology 63(8): 3327–32.
- T.G., 597 2013. Harmful Paerl, H.W. and Otten, Cyanobacterial Blooms: Causes, 598 Consequences, Controls: Microbial ecology. 995-1010. and 599 https://doi.org/10.1007/s00248-012-0159-y.
- Park, H., Kim, B., Kim, E., and Okino, T., 1997. Hepatotoxic Microcystins and Neurotoxic
 Anatoxin-a in Cyanobacterial Blooms from Korean Lakes. Environmental Toxicology
 and Water Quality: An International Journal, 13(3), 225-234.

603	Park,	H., V	Watanabe,	M.F	Harada,	Nagai,	K.,	and	Suzuki,	М.,	1993.	Hepato	toxin
604		(Micr	rocystin) a	nd Ne	urotoxin	(Anatoxi	n-a)	Conta	ined in N	Jatural	Bloom	is and S	trains
605		of Cy	yanobacteri	ia Fro	m Japane	ese Fresh	wate	rs (A)). Park, I	ł. D.	Natural	toxins,	1(6),
606		353-	360.										

- Pawlik-Skowrońska, B., 2004. Relationship between cyanobacterial bloom composition and
 anatoxin-a and microcystin occurrence in the eutrophic dam reservoir (SE Poland).
 Polish Journal of Ecology, 52(4), 479-490.
- Ramm, J., Lupu, A., Hadas, O., Ballot, A., Rücker, J., Wiedner, C., and Sukenik, A., 2012. A
 CARD-FISH Protocol for the Identification and Enumeration of Cyanobacterial
 Akinetes in Lake Sediments. FEMS Microbiology Ecology 82(1): 23–36.
 https://doi.org/10.1111/j.1574-6941.2012.01401.x.
- Ramm, J., Rücker, J., Knie, M., and Nixdorf, B., 2017. Lost in the Dark: Estimation of the
 Akinete Pool for the Recruitment of Nostocales Populations (Cyanobacteria) in a
 Temperate Deep Lake. Journal of Plankton Research, 39(3), 392-403.
- Rantala-Ylinen, A., Kana, S., Wang, H., Rouhiainen, L., Wahlsten, E. Rizzi, et al., 2011.
 Anatoxin-a Synthetase Gene Cluster of the Cyanobacterium *Anabaena Sp.* Strain 37
 and Molecular Methods To Detect Potential Producers. Applied and Environmental
 Microbiology 77(20): 7271–78. https://doi.org/10.1128/AEM.06022-11.
- Rapala, J, Sivonen, K., Lyra, C., and Niemelä, S., 1997. Variation of Microcystins in *Anabaena Spp* as a Function of Growth Stimuli. Applied and Environmental
 Microbiology 63 6): 2206–12.
- Rapala, J., Sivonen, K., Luukkainen, R., and Niemela, S., 1993. Anatoxin-a Concentration in
 Anabaena and *Aphanizomenon* under Different Environmental Conditions and

- 626 Comparison of Growth by Toxic and Non-Toxic *Anabaena*-Strains a Laboratory
 627 Study. Applied Phycology 5: 581–91. https://doi.org/10.1007/BF02184637.
- Revelles, J., and Van Geel, B. 2016. Human impact and ecological changes in lakeshore
 environments. The contribution of non-pollen palynomorphs in Lake Banyoles (NE
 Iberia). Review of Palaeobotany and Palynology, 232, 81-97.
- Rinta-Kanto, J., Saxton, M., DeBruyn, J., Smith, J., Marvin, C., Krieger, K., Sayler, G.,
 Boyer, G., and Wilhelm. S., 2009. The Diversity and Distribution of Toxigenic *Microcystis Spp.* in Present Day and Archived Pelagic and Sediment Samples from
 Lake Erie. Harmful Algae 8(3): 385–94. https://doi.org/10.1016/j.hal.2008.08.026.
- Rücker, J., Tingwey, E., Wiedner, C., Anu, C., and Nixdorf, B., 2009. Impact of the Inoculum
 Size on the Population of Nostocales Cyanobacteria in a Temperate Lake. Journal of
 Plankton Research 31 (10): 1151–59. https://doi.org/10.1093/plankt/fbp067.
- Rühland, K., Paterson, A.M., and Smol, J.P.. 2015. Lake Diatom Responses to Warming:
 Reviewing the Evidence. Journal of Paleolimnology 54(1).
 https://doi.org/10.1007/s10933-015-9837-3.
- Sabart, M., Crenn, K., Perrière, F., Abila, A., Leremboure, M., Colombet, J., Jousse, C., and
 Latour, D., 2015. Co-Occurrence of Microcystin and Anatoxin-a in the Freshwater
 Lake Aydat (France): Analytical and Molecular Approaches during a Three-Year
 Survey. Harmful Algae 48: 12–20. https://doi.org/10.1016/j.hal.2015.06.007.
- Sarazin, G., Michard, G., AlGharib, I., and Bernat, M., 1992. Sedimentation Rate and Early
 Diagenesis of Particulate Organic Nitrogen and Carbon in Aydat Lake (Puy de Dome,
 France). Chemical Geology 98: 307–16.

Savichtcheva, O., Debroas, D., Perga, M.E, Arnaud, F., Villar, C., Lyautey, E., et al., 2015.
Effects of Nutrients and Warming on *Planktothrix* Dynamics and Diversity: A
Palaeolimnological View Based on Sedimentary DNA and RNA. Freshwater Biology
60(1): 31–49. https://doi.org/10.1111/fwb.12465.

- Sukenik, A., Rücker, J., and Maldener I., 2019. "Dormant Cells (Akinetes) of Filamentous
 Cyanobacteria Demonstrate a Great Variability in Morphology, Physiology, and
 Ecological Function." Cyanobacteria. Academic Press. 65-77.
- Tsujimura, S., and Okubo, T., 2003. Development of *Anabaena* Blooms in a Small Reservoir
 with Dense Sediment Akinete Population, with Special Reference to Temperature and
 Irradiance. Journal of Plankton Research 25(9): 1059–67.
 https://doi.org/10.1093/plankt/25.9.1059.
- Van Geel B, Mur LR, Ralska-jasiewiczowa M, Goslar T., 1994. Fossil akinetes of *Aphanizomenon* and *Anabaena* as indicators for medieval phosphate-eutrophication of
 Lake Gosciaz (Central Poland). Rev Paleobotany Palynol. 83:97–105.
- Van Geel, B., Buurman, J., and Waterbolk, H. T., 1996. Archaeological and palaeoecological
 indications of an abrupt climate change in The Netherlands, and evidence for
 climatological teleconnections around 2650 BP. Journal of Quaternary Science, 11(6),
 451-460.
- Van Geel, B., Hallewas, D.P., and Pals, J.P., 1983. A late Holocene deposit under the
 Westfriese Zeedijk near Enkhuizen (Prov. of Noord-Holland, The Netherlands):
 palaeoecological and archaeological aspects. Review of Palaeobotany and
 Palynology, 38(3-4), 269-335.

670	Wood, S.A., Jentzsch, K., Rueckert, A., Hamilton, D.P., and Cary, S.C., 2009. Hindcasting
671	Cyanobacterial Communities in Lake Okaro with Germination Experiments and
672	Genetic Analyses. FEMS Microbiology Ecology 67(1980): 252–60.
673	https://doi.org/10.1111/j.1574-6941.2008.00630.x.

674

675 Figure legends

676 Graphical abstract: Evolution of akinetes from a cyanobacterial bloom to several thousand years after677 landfill (DNA was preserved as free DNA in sediment or in intact akinete).

678 Table 1: Primers used to amplify targeted genes.

679 Figure 1: (A) Bathymetric map of Lake Aydat (the star represents the location of the core sample); (B)

680 age model of the core from Lavrieux et al., (2013a) (blue lines represent analyzed samples, and blue

681 circles samples used in the germination experiment); (C) different akinete morphotypes encountered:

682 (a) Dolichospermum macrosporum, b) D. flos-aquae c) Dolichospermum sp. d) Aphanizomenon sp.).

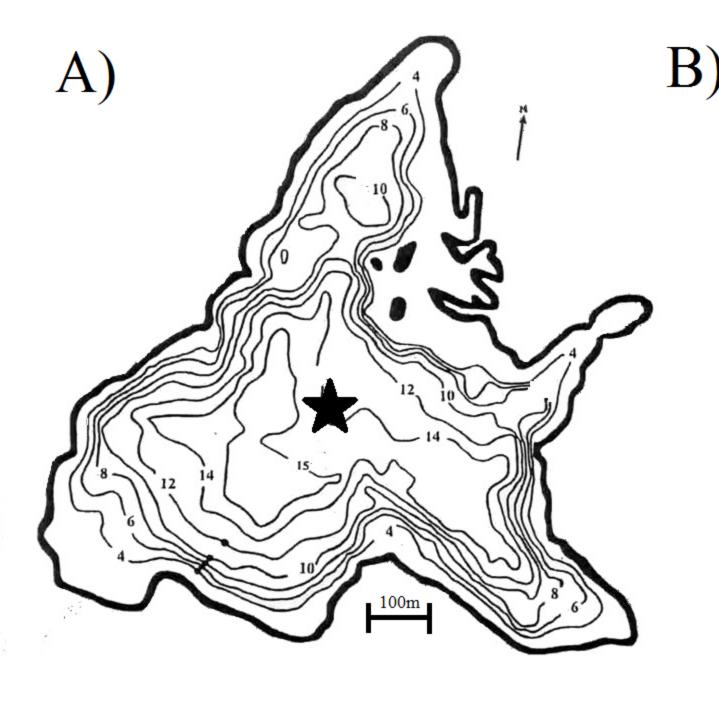
Figure 2: Distribution of *D. macrosporum* and *D. flos-aquae, Dolichospermum sp.* and *Aphanizomenon sp.* akinetes through the sediment core. Light color represents total akinete abundance for one condition whereas deep color represents intact akinete distribution. Intact akinetes are defined as akinetes which have an intact cellular wall (Legrand et al., 2016b).

687 Figure 3: Evolution of the percentage of intact akinetes for each studied species through the core.

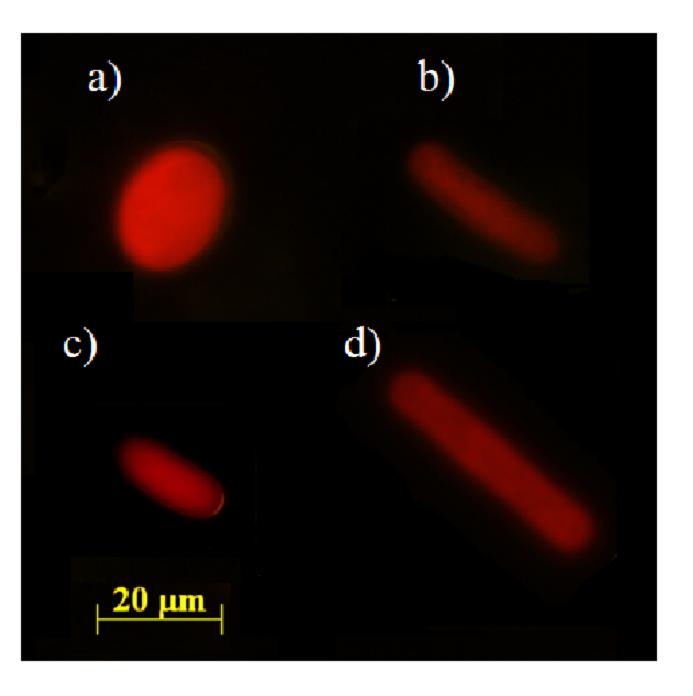
Figure 4: Proportion of akinete morphotypes observed in the sediment layers (A) and during the germination experiment (B). Red color represents samples for which intact akinetes have not been enumerated.

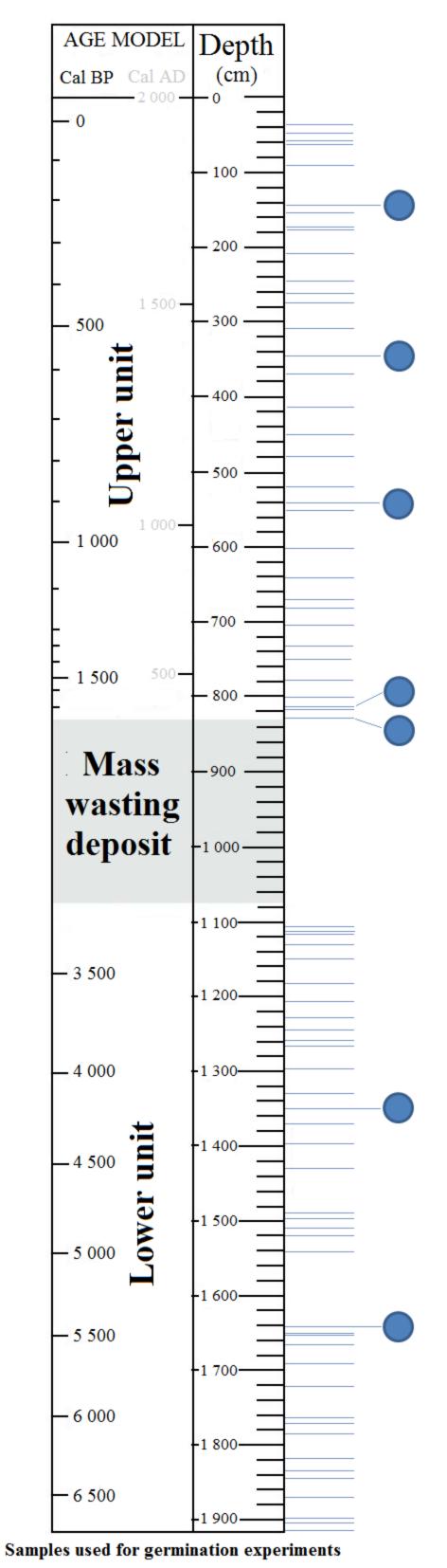
691 Figure 5: Detection of the targeted genes in total sediment through the core.

692

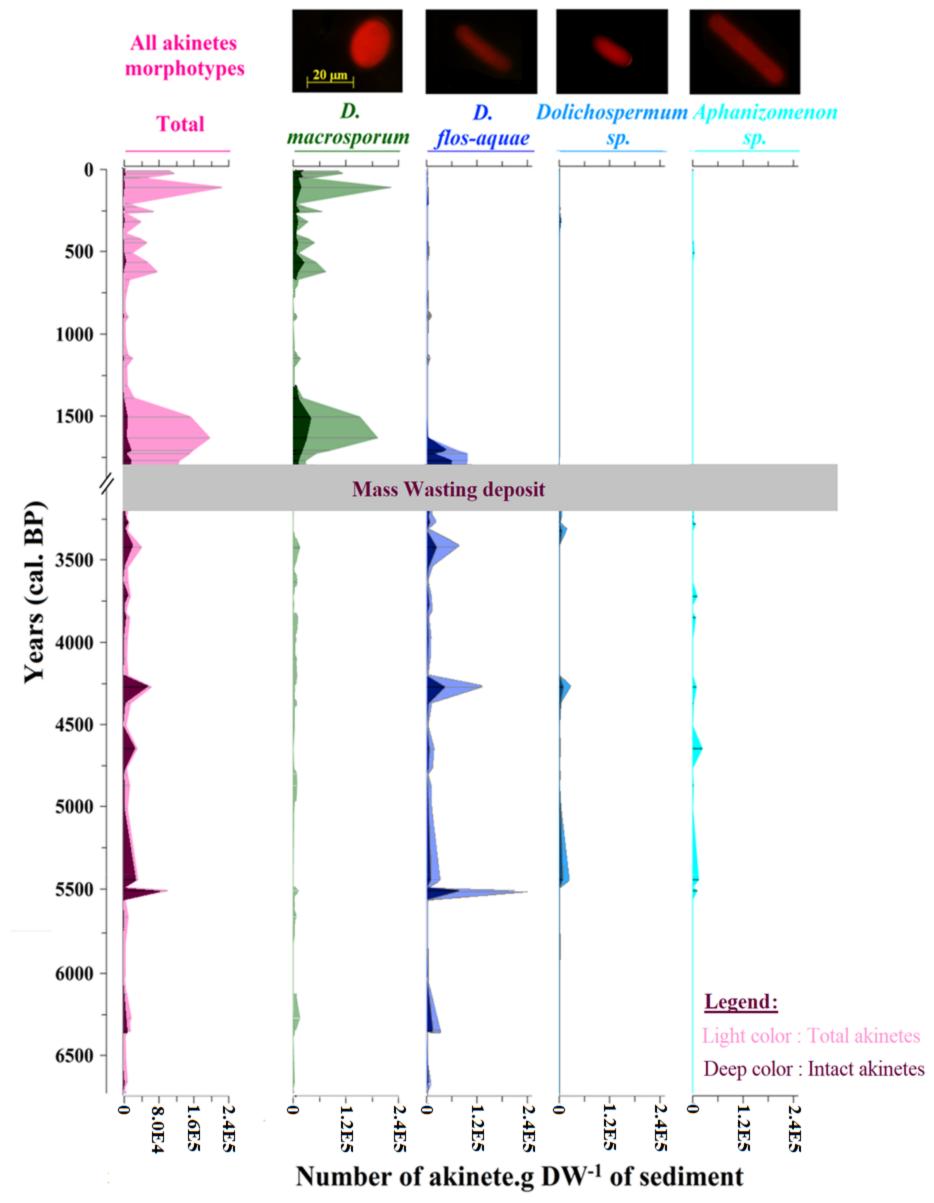


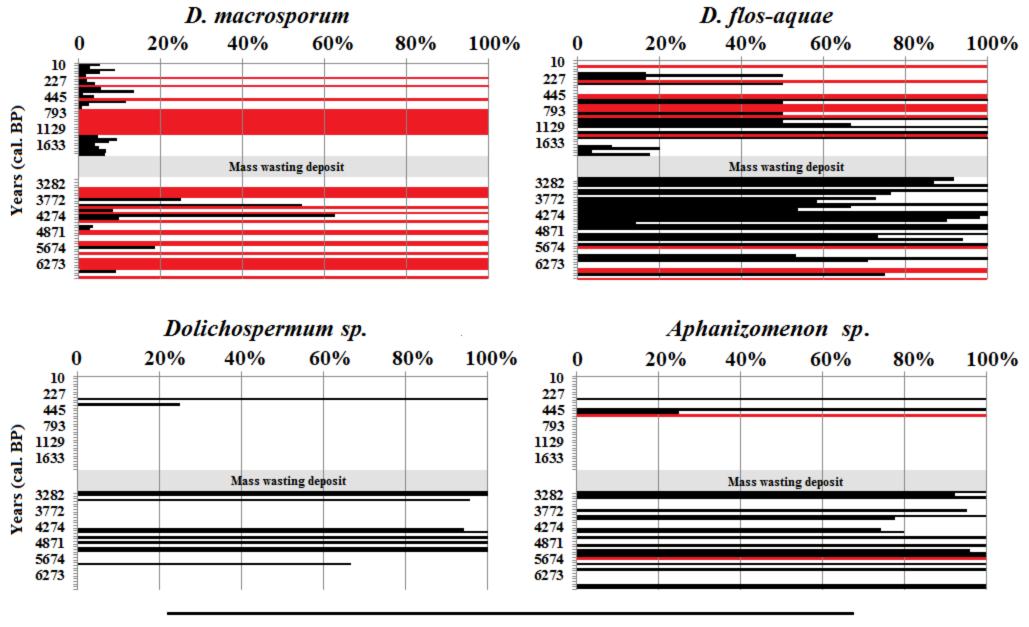
C)



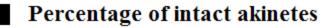


Legend: — Analyzed samples

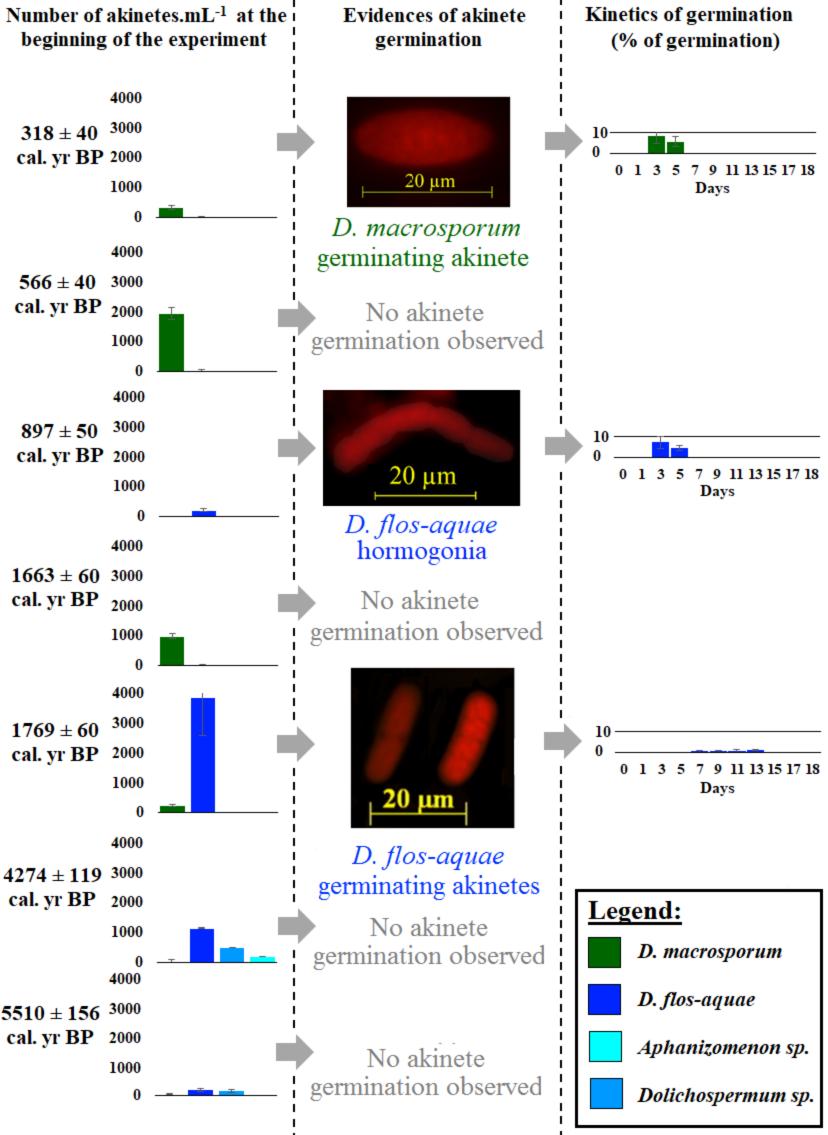




Percentage of intact akinetes



Samples characterized with 0% of intact akinetes



Upper	' Unit		Lower	Unit	
Age (Years		~	Age (Years		
cal. BP)	тсуВ	anaC	cal. BP)	тсуВ	anaC
10			3238		
25			3269		
48			3282		
52			3322		
107			3421		
208			3544		
227			3638		
258			3721		
268			3772		
318			3820		
318			3848		
			3976		
422			4102		
445			4208		
566			4274		
620			4374 4512		
670			4648		
675			4768		
739			4802		
793			4871		
864			5011		
897			5446		
924			5492		
1011			5510		
1080			5565		
1129			5674		
1151			5822		
1194			6013		
1316			6032		
1391			6126		
1507			6273		
1633			6354		
1708			6388		
			6547		
1728			6659		
1769	4. 1		6702		
Mass wa	isting de	posit	6727		

Positive with PCR detection

	Primer name	Sequence (5'-3')	Target gene	Gene Amplicon size (bp)	References
Classic PCR	cya359F cya781R	GGGGAATYTTCCGCAATGGG GACTACTGGGGTATCTAATCCCATT	Cyanobacterial 16S rRNA gene	403	Nubel et al., (1997)
Classic PCR	anaC-gen-F2 anaC-gen-R2	TCTGGTATTCAGTMCCCTCYAT CCCAATARCCTGTCATCAA	anaC	366	Sabart et al., (2015) modified from Rantala-Ylinen et al., (2011)
Nested PCR 1	anxgen-F2 anxgen-R	ATGGTCAGAGGTTTTACAAG CGACTCTTAATCATGCGATC	anaC	861	Rantala-Ylinen et al., (2011)
Nested PCR 2	anaC 80F anaC 706R	CCAATAGCCTGTCATCAACC TATACATCTGGCTCAACAGG	anaC	626	This study
Nested PCR 1	MCY F1 MCY R1	TGGGAAGATGTTCTTCAGGTATCCAA AGAGTGGAAACAATATGATAAGCTAC	mcyB	319	Nonneman et al., (2002)
Nested PCR 2	MCY F1 MCY R2	TGGGAAGATGTTCTTCAGGTATCCAA GAGATCCATCTGTTGCAAGACATAG	mcyB	287	Nonneman et al., (2002)

