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1	Effects of two common fungicides on the reproduction of Aporrectodea caliginosa in
2	natural soil
3	Sylvain Bart ^{a*} , Alexandre Barraud ^a , Joël Amossé ^a , Alexandre R.R. Péry ^a , Christian Mougin ^a ,
4	Céline Pelosi ^a
5	^a UMR ECOSYS, INRA, AgroParisTech, Université Paris-Saclay, 78026, Versailles, France
6	
7	* Corresponding author: UMR1402 ECOSYS INRA AgroParistech, Bâtiment 6, RD 10, 78026
8	Versailles cedex, France. Tel: (+33)1.30.83.32.72. ; Fax: (+33)1.30.83.32.59. E-mail address:
9	sylvain1.bart@outlook.fr
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12	Highlights
13	• The effects of two commercial fungicides were tested on earthworm reproduction
14	• Adult individuals and "naïve" cocoons were exposed
15	• Negative effects were found on cocoon production, hatching time, hatching success
16	• "Naïve" cocoons were less affected than cocoons produced by exposed adults
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28 Abstract

29 The use of pesticides in agroecosystems can have negative effects on earthworms, which play key roles in soil functioning such as organic matter decomposition. The aim of this study was 30 to assess the effects of two fungicides (Cuprafor micro[®], composed of copper oxychloride, and 31 Swing Gold[®], composed of epoxiconazole (EPX) and dimoxystrobin (DMX)) on earthworm 32 reproduction by exposing adults and cocoons. First, adult Aporrectodea caliginosa individuals 33 were exposed for 28 days to 3.33, 10 and 30 times the recommended dose (RD) of Cuprafor 34 micro[®] corresponding to 25.8, 77.5 and 232.5 mg kg⁻¹ dry soil of copper, respectively, and 0.33, 35 1 and 3 times the RD of Swing Gold[®] (corresponding to 5.2×10^{-2} mg DMX kg⁻¹ + 1.94×10^{-2} 36 mg EPX kg⁻¹, 1.55 x 10⁻¹ mg DMX kg⁻¹ + 5.81 x 10⁻² mg EPX kg⁻¹ and 4.62 x 10⁻¹ mg DMX 37 $kg^{-1} + 1.74 \times 10^{-1} mg EPX kg^{-1}$ respectively), in addition to a control soil with no fungicide 38 treatment. Cocoon variables (production, weight, hatching success, hatching time) were 39 monitored. Second, "naïve" cocoons produced by uncontaminated earthworms were exposed 40 to soils contaminated by the same concentrations of the two fungicides, and we assessed 41 hatching success and hatching time. In the first experiment, cocoon production was halved at 42 the highest copper concentration (232.5 mg Cu kg⁻¹ of dry soil) as compared to the control. 43 Cocoons took 5 more days to hatch, and the hatching success decreased by 35% as compared 44 to the control. In the Swing Gold[®] treatments, cocoon production was reduced by 63% at 3 45 times the RD, and the hatching success significantly decreased by 16% at the RD. In the second 46 experiment, only the hatching success of cocoons was impacted by Swing Gold[®] at 3 times the 47 48 RD (30% less hatching). It is concluded that the cocoon stock in the soil is crucial for the renewal of populations in the field. The most sensitive endpoint was the hatching success of the 49

cocoons produced by exposed adults. This endpoint and the effects observed on the "naïve"
cocoons could be taken into account in pesticide risk assessment.

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53 Keywords: Ecotoxicology; Ecological risk assessment; Soil invertebrates; Lumbricidae;
54 Pesticide

55 **1. Introduction**

56 During the last decades, the increasing demand for food products has led to changes in land use and management. Along with an increase in cultivated areas, the enhancement of yields 57 58 has been promoted by the use of chemical inputs such as pesticides (Tilman et al., 2002). In temperate soils, earthworms represent a large proportion of the living biomass (Gobat et al., 59 2004). These primary consumers of organic matter are involved in soil structuring, and thus 60 maintain soil health (Bertrand et al., 2015; Blouin et al., 2013). While earthworm densities 61 generally range from 200 to 400 individuals m⁻² in permanent grasslands (Amossé et al., 2018; 62 Beylich and Graefe, 2012; Edwards and Bohlen, 1996), they can be lower than 100 individuals 63 m⁻² in conventional agricultural fields (Scullion et al., 2002). Mechanical tillage such as 64 ploughing has a direct negative impact on earthworm populations (Chan, 2001; Mele and 65 Carter, 1999). However, the impact of pesticides on earthworm populations under field 66 conditions is much harder to assess and rarely documented. Using a concentration-response-67 based field experiment, Schnug et al. (2015) showed that a mixture of three biocides (i.e., the 68 insecticide esfenvalerate, the fungicide picoxystrobin and the bactericide triclosan) affected 69 earthworm abundance and juvenile proportions. These authors also highlighted that pesticide 70 concentrations corresponded to the EC₅₀ for *E. fetida* reproduction (Schnug et al., 2013; Schnug 71 et al., 2014) and had effects on earthworms in the field exceeding 50%, indicating that the 72 effects observed in the field could be greater than those observed under laboratory conditions. 73

This result showed that some endpoints such as reproduction were not sufficiently and correctlycovered in laboratory experiments.

76

77 The reproduction test proposed in the ISO guideline with the *Eisenia fetida* species (ISO 11268-2, 2012; OECD 222, 2016) currently allows assessing the number of juveniles produced 78 by adults exposed to chemicals. But the procedure suffers from at least two weaknesses. First, 79 this test procedure does not allow for specifically assessing the impacts of chemicals on cocoon 80 production and then on their viability, which is a knowledge gap if we seek to understand the 81 82 impact of pesticides on the earthworm reproduction process. Secondly, the cocoons produced before the chemical applications, which end up in contact with them, could be affected. No 83 assessment on these "naïve" cocoons is currently performed. Nevertheless, the soil harbours a 84 85 cocoon stock that is necessary for the populations to be renewed (Bouché, 1972; Boström and Lofs Holmin, 1996) in case of disturbances (e.g., drought, freezing temperatures, chemical 86 stress) that directly affect adult and juvenile individuals. Consequently, not only is it relevant 87 88 to assess the impacts of pesticides on cocoon production and hatching success by exposing adult earthworms, but it is also equally relevant to study the viability of cocoons by exposing them 89 directly to chemicals. Finally, the tests proposed by the ISO guidelines are performed with 90 *Eisenia fetida*, which is not representative of agricultural fields and is less sensitive than other 91 92 soil-dwelling species to pesticides (Pelosi et al., 2013). The relevance of using species other 93 than E. fetida in pesticide risk assessment is also increasingly supported in the scientific literature (Lowe and Butt, 2007; Spurgeon et al., 2003). Aporrectodea caliginosa is one of the 94 increasingly advocated species in ecotoxicity tests. It is highly representative of temperate 95 agricultural fields (Bart et al., 2018; Decaëns et al., 2011; Perez-Losada et al., 2009) and 96 according to Pelosi et al. (2013), it is three times more sensitive to pesticides than E. fetida. 97

In this context, our study aimed to assess the impacts of two commercial formulations 98 of fungicides (Cuprafor micro[®], composed of copper oxychloride, and Swing Gold[®], composed 99 of epoxiconazole and dimoxystrobin) on A. caliginosa reproduction. We selected copper 100 101 because it can negatively impact earthworm biomass and growth (Bart et al., 2017; Eijsackers et al., 2005; Khalil et al., 1996a) and accumulates in the soil, potentially leading to high 102 concentrations (Brun et al., 1998). We selected Swing Gold[®] because its active substances are 103 recognized as dangerous for the environment, and dimoxystrobin is considered as reprotoxic. 104 Both fungicides are widely used in cultivated fields in Europe. Finally, the choice to use 105 106 commercial formulations was given preference over pure molecules so as to be as close as possible to field conditions, and this generates relevant data for risk assessment. We assessed 107 the impact of the two fungicides on A. caliginosa reproduction by exposing adult earthworms 108 109 to contaminated soils, and also by exposing "naïve" cocoons. Thereby, we tested the assumption that the hatching success and time to hatch of cocoons produced by exposed adults could be 110 more affected by pesticides than those of "naïve" cocoons because "naïve" cocoons were 111 produced by unexposed adults and could better tolerate unfavorable environmental conditions 112 (e.g., drought, freezing temperatures, chemical stress). 113

114

115 **2. Materials and methods**

116 2.1. Soil, animals, and pesticides

117 A loamy soil (FAO soil classification) was used in all experiments. It was collected from 118 a permanent meadow in Versailles (48°48′ N, 2°5′ E) where any synthetics and metals 119 compounds had been applied for more than 20 years. The soil was collected from the top 0-20 120 cm, air-dried and crushed at 2 mm. Its main physico-chemical characteristics were: pH 7.5, organic matter 32.6 g kg⁻¹, C/N 12.7, 29% sand, 48% silt, 23% clay, and total Cu 25.2 mg kg⁻¹ (see Bart et al., 2017 for more details).

A. caliginosa individuals came from an agricultural field in Estrées-Mons, France
 (49°52' N 3°01'E). They were mature, with a fully developed clitellum, and their weight ranged
 from 600 to 1 000 mg. They were bred by incorporating horse dung to the above-mentioned
 soil at 15°C for 14 days; the full breeding procedure is presented in Bart et al. (2018).

We studied two fungicide formulations. The first was Cuprafor micro[®] (Quimicas del Valles, 50% copper oxychloride), used to prevent fungal and bacterial germination in conventional and organic farming. The French Recommended Dose (RD) for this product is 10 kg ha⁻¹ on potato and in vineyards (ANSES, 2017b). The laboratory RD was calculated with the following formula:

$$[mg kg^{-1}] = \frac{[mg ha^{-1}]}{area (dm^2) x depth (dm) x soil density (kg dm^{-3})}$$

with mg ha⁻¹ = 10 000 000 mg Cu, area = 1 000 000 dm², a soil density of 1.29 kg dm⁻³, and 133 considering that copper was mainly found in the top 0.5 dm of soil (Couto et al., 2015). The 134 RD was thus calculated as 15.5 mg kg⁻¹ of dry soil (corresponding to 7.75 mg kg⁻¹ of copper). 135 We used the following concentrations for all experiments: 3.33, 10 and 30 times the RD 136 abbreviated as Cu3.33, Cu10 and Cu30, which correspond to 25.8, 77.5 and 232.5 mg kg⁻¹ of 137 copper, respectively. These concentrations were expected to be sub-lethal, considering that the 138 LC50 for *Eisenia fetida* has been estimated to be $> 486.6 \text{ mg kg}^{-1}$ of copper (which corresponds 139 to 63 times the RD) (PPDB, 2017). The second fungicide was Swing Gold[®] (BASF Agro SAS, 140 dimoxystrobin 133 g L^{-1} , epoxiconazole 50 g L^{-1}), used to protect cereal crops in conventional 141 farming. The French Recommended Dose (RD) for this product is 1.5 L ha⁻¹ on wheat (ANSES, 142 2017a). Using the same formula as for copper, the laboratory RD was calculated as $1.16 \ 10^{-3}$ 143 mL kg⁻¹ of dry soil, with the same soil density and considering that the active compounds of 144

this fungicide were mainly found in the top 1 dm of soil (Chabauty et al., 2016; McDonald et
al., 2013). We used the following concentrations for all experiments: 0.33, 1 and 3 times the
RD, abbreviated as SG0.33, SG1 and SG3, respectively. These concentrations were expected
to be sub-lethal, considering that the 14-day LC50 was estimated to be 6.3 times the RD for *A*. *caliginosa* (Bart et al., 2017).

The dry soil was spiked with aqueous solutions of fungicides, and the soil moisture was adjusted concomitantly at 70% of the water-holding capacity (WHC) (corresponding to 28% of water content) in all experiments. Only tap water was used for the control treatments.

153

154 2.2. Cocoon production, hatching success and time to hatch through adult exposure

Five replicates were performed for each concentration and the control. For each 155 replicate, 500 g of soil (dry mass) and 24 g of horse dung were mixed and placed in a 1-L plastic 156 vessel (15 x 10 x 7 cm) with a removable perforated cover for gas exchanges. Horse dung used 157 as a food source was frozen - defrosted twice and then milled (< 1 mm) before use (Lowe and 158 Butt, 2005). The food amount corresponded to 6 g ind⁻¹ month⁻¹, which is enough to maintain 159 A. caliginosa weight. This food amount was previously determined and close to that proposed 160 by Lowe and Butt (2005). Four adult individuals with a visible clitellum were weighed and 161 introduced in each vessel and stored in a climate room (15±1 °C, 24 h darkness, Bart et al., 162 2018) for 28 days. The soil moisture (i.e., 70% of WHC) was checked weekly by monitoring 163 the mass of the vessels throughout the experiment and adjusted with tap water if necessary. At 164 the end of the exposure period, individuals were removed, counted, and weighed. The 165 166 difference between the individual weights at the beginning and at the end of the experiment was used to calculate the percentage of adult weight change. The wet sieving method was used to 167 retrieve produced cocoons, and results were expressed as numbers of cocoons per vessel. The 168

169 cocoons were then weighed and incubated at $20^{\circ} \pm 0.5$ C in petri dishes in aqueous tap water 170 solution (i.e., semi-immersed). The cocoons were monitored every two days over a 60-day 171 period. After that period, we considered that the unhatched cocoons would never hatch. Mean 172 hatching time and hatching success were calculated for each replicate.

- 173
- 174

2.3. Cocoon hatching success and time to hatch through "naïve" cocoon exposure

More than 120 cocoons were produced over a period of 7 days in the A. caliginosa 175 breeding culture. In order to get a homogenous cocoon cohort, we discarded the 5% of cocoons 176 177 that were too big or too small. The sinks of 24-well microtiter plates were filled with 2 g of soil (dry mass). All the plates had a removable perforated cover for gas exchanges. Five replicates 178 were set up for each fungicide and concentration tested, and for the control. Three cocoons 179 randomly selected among the cocoon cohort were introduced in each sink. The cocoons were 180 gently buried in the soil. The plates were incubated at 20 ± 0.5 °C, and the soil moisture (i.e., 181 182 70% WHC) was checked weekly by monitoring the mass of the plates, and adjusted with tap water if necessary. The hatching cocoons were checked every 2 days, and the experiment was 183 stopped after 60 days of incubation, considering that the remaining cocoons would never hatch 184 beyond that delay. The mean hatching time and hatching success per replicate were calculated 185 for each modality. 186

187

188 2.4. Statistical analyses

All data were analysed with the statistical software program R Core Team (2015). Differences were considered significant if the p-value < 0.05. When the homogeneity of variance (Bartlett-test, Snedecor and Cochran, 1989) and the normality of residuals (Shapiro test) were respected, ANOVA and Fisher's least significant difference (LSD) post-hoc tests 193 were performed to analyse the differences among treatments. When normality and 194 homoscedasticity conditions were not satisfied, the non-parametric Kruskal-Wallis test was 195 used.

196

197 **3. Results**

198 *3.1. Cocoon production, hatching success and time to hatch through adult exposure*

The effects of Cuprafor micro[®] on earthworm biomass, cocoon production, and cocoon weight are presented in Fig. 1. The hatching time and success of the produced cocoons are presented in Table 1, and the hatching dynamics (all replicates included) in Fig. S1.

At the end of the experiment, a mortality rate of 5% was recorded in the Cu3.33 and the 202 control treatments. No mortality occurred in the Cu10 and Cu30 treatments. No significant 203 difference in adult weight change was recorded between the control and Cuprafor micro® 204 treatments. However, adult weight change was significantly lower in the Cu30 treatment (-205 7.3%) than in the Cu10 treatment (+9.6%, Fig. 1A). Earthworms produced 15.9 cocoons (mean 206 number of cocoons per vessel) on average in the control treatment. Cocoon production was 207 significantly two times lower in the Cu30 treatment (corresponding to 30 times the RD or 232.5 208 mg kg⁻¹ of copper) than in the control (Fig. 1 B). We did not find any significant effect of the 209 copper fungicide on the mean weight of the produced cocoons (Fig. 1 C). 210

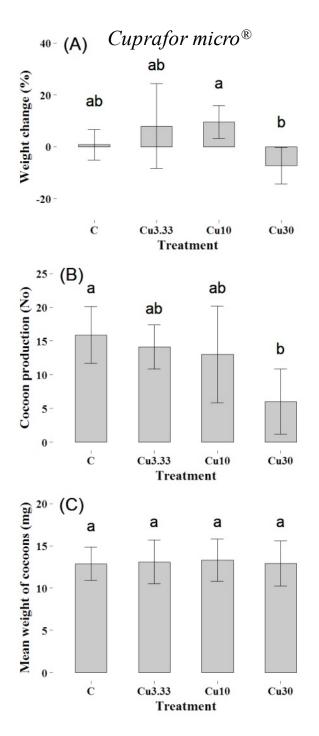


Fig. 1. (A) Weight change (% of difference between the beginning and the end of the experiment, 28 days), (B) cocoon production (number of cocoons per vessel produced by 4 individuals), and (C) mean weight of the cocoons produced by earthworms exposed to control soil or soils spiked with different concentrations of Cuprafor micro[®] (3.33, 10 and 30 times the RD, corresponding to 25.8, 77.5 and 232.5 mg kg⁻¹ of copper, abbreviated Cu3.3, Cu10 and Cu30, respectively). n=5 ± SD; different letters indicate significant differences among the concentrations (p<0.05).

219	The hatching dynamics of the cocoons (all replicates included) in the Cu10 and Cu30
220	treatments differed from the dynamics of the control treatment (Fig. S1). More specifically, the
221	mean hatching time of the cocoons produced in the Cu30 treatment was significantly longer (5
222	more days) than in the control treatment (Table 1). Moreover, the hatching success was
223	significantly two times lower in the Cu10 treatment than in the control (Table 1). Although the
224	mean hatching success was 61% and 94% in the Cu30 and control treatments, respectively, the
225	difference was not significant due to high variability in the Cu30 treatment: only one cocoon
226	was produced in one replicate (N=5) and did not hatch (hatching success of 0%), while 2
227	cocoons were produced in another replicate and hatched (hatching success of 100%).

Table 1. Mean hatching time (days) and mean hatching success (%) of cocoons produced by *A*.
 caliginosa adults exposed to Cuprafor micro^{®a}

		Treatment ^b				
		С	Cu3.33	Cu10	Cu30	
	Hatching time (days)	$23.5 \pm 1.7 \text{ A}$	25.6 ± 1.8 AB	$26.8 \pm 3.5 \text{ AB}$	$28.5 \pm 2.4 \text{ B}$	
	Hatching success (%)	94.4 ± 5.3 A	$88.5 \pm 13.7 \text{ A}$	$45.9\pm18.9~B$	$61.0 \pm 38.5 \text{ AB}$	
231	^a Cocoons were proc	luced over a 28	-day period in a	control soil C and	l in soils spiked with	
232	different Cuprafor m	icro [®] concentrati	ions (3.33, 10 and	30 times the RD, o	corresponding to 25.8,	
233	77.5 and 232.5 mg kg	⁻¹ of copper abbr	eviated Cu3.3, Cu	10 and Cu30, resp	ectively). C = Control	
234	treatment.					
235	^b Different letters ind	icate significant	differences amon	g concentrations (p<0.05).	
236 237						
238	The effects o	f Swing Gold [®]	on earthworm bio	omass, cocoon pro	oduction, and cocoon	
239	weight are presented	in Fig. 2. The	hatching time and	d success of the p	produced cocoons are	
240	presented in Table 2,	and the hatching	g dynamics (all re	plicates included)	in Fig. S2.	
241	At the end o	f the experimer	nt, no mortality w	vas recorded in th	ne SG0.33 treatment.	
242	Mortality rates of 5%	% were found in	the SG1 and cont	trol treatments, an	d of 20% in the SG3	

- treatment. No significant difference in adult weight change was recorded between the control
- and the Swing Gold[®] treatments. However, adult weight was significantly lower in the SG3
- treatment (-13.4%) than in the SG0.33 treatment (+7.9%, Fig. 2 A). Earthworms produced
- significantly three times less cocoons at 3 times the recommended dose (SG3) as compared to
- the control soil (Fig. 2 B). As for the copper fungicide, the different Swing Gold[®] concentrations
- 248 did not influence cocoon weight (Fig. 2 C).

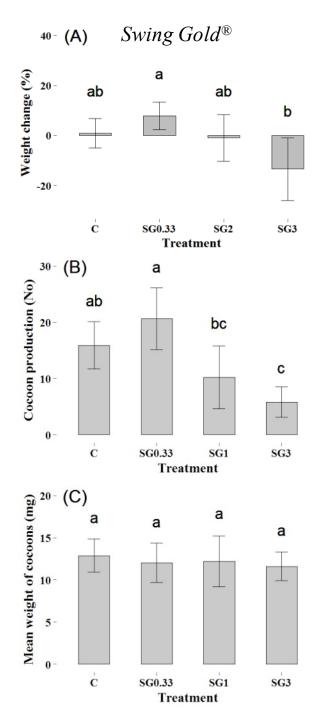


Fig. 2. (A) Weight change (% of the difference between the beginning and the end of the experiment, 28 days), (B) cocoon production (number of cocoons per vessel produced by 4 individuals), and (C) mean weight of the cocoons produced by earthworms exposed to control soil or soils spiked with different Swing Gold[®] concentrations (0.33, 1 and 3 times the RD, abbreviated SG0.33, SG1 and SG3, respectively). n=5 ± SD; different letters indicate significant differences among the concentrations (p<0.05).

The hatching dynamics of the cocoons (all replicates included) in the SG1 and SG3 treatments differed from the dynamics of the control treatment (Fig. S2). More specifically, the Swing Gold[®] fungicide negatively affected the hatching time of cocoons: they took 5 more days to hatch in the SG3 treatment than in the other treatments (Table 2). The hatching success was 16% and 36% lower for the cocoons produced by individuals exposed to the SG1 and SG3 treatments, respectively, as compared to the control treatment (Table 2).

Table 2. Mean hatching time (days) and mean hatching success (%) of the cocoons produced
 by *A. caliginosa* adults exposed to Swing Gold^{®a}.

	Treatment ^b					
		С	SG0.33 22.2 ± 1.4 A	SG1	SG3	
	Hatching time	$23.5 \pm 1.7 \text{ A}$	$22.2 \pm 1.4 \text{ A}$	22.9 ± 1.7 A	$28.8\pm3.4~\mathrm{B}$	
	(days) Hatching success _(%)		91.1 ± 15.6 AB	$78.9\pm6.8~B$	60.3 ± 14.4 C	
266	^a Cocoons were pro	duced over a 28	8-day period in a c	ontrol soil C an	d in soils spiked with	
267	different Swing Gol	d [®] concentration	ns (0.33, 1 and 3 tin	nes the RD, abb	reviated SG0.33, SG1	
268	and SG3, respectivel	5 /				
269	^b Different letters inc	licate significan	t differences among	the concentration	ons (p<0.05).	
270 271						
272	3.2. Cocoon hatching	g success and tin	me to hatch through	n "naive" cocoon	1 exposure	
273	The effects of	f Cuprafor micro	$\infty^{\mathbb{R}}$ on the hatching the	me and success of	of "naïve" cocoons are	
274	presented in Table 3, and the hatching dynamics (all replicates included) are presented in Fig					
275	S3 A.					
276	The hatching	dynamics (all re	eplicates included)	of "naïve" cococ	ons produced in a non-	
277	contaminated soil and then exposed to Cu3.33 treatment, Cu10 treatment, or to a control soil					
278	were similar (Fig. S3	A). The dynami	ics in the Cu30 treat	ment was slower	and stopped at around	
279	80% of the hatching success, versus 100% in the control treatment. However, there was no					
280	significant effect on	the hatching tim	e and success, what	ever the concent	ration tested (Table 3).	

282	The effects of Swing Gold® on the hatching time and success of "naïve" cocoons are
283	presented in Table 4, and the hatching dynamics (all replicates included) are presented in Fig
284	S3 B.
285	The hatching dynamics (all replicates included) of "naive" cocoons produced in a non-
286	contaminated soil and then exposed to SG0.33 or SG1 treatment were close to the dynamics of
287	the control (Fig. S3 B). The hatching dynamics of the cocoons exposed to the SG3 treatment
288	was slower than the dynamics of the control treatment (Fig. S3 B), and the hatching success
289	was significantly affected, with only 67% of hatching success (Table 4).

Table 3. Mean hatching time (days) and mean hatching success (%) of "naïve" *A. caliginosa* cocoons^a exposed to Cuprafor micro^{®b}.

		Treatment ^c					
		С	Cu3.33	Cu10	Cu30		
	Hatching time (days)	31.1 ± 1.1 A	32.4 ± 1.0 A	32.7 ± 3.6 A	33.1 ± 1.5 A		
	Hatching success (%)	$100 \pm 0.0 \text{ A}$	93.3 ± 14.9 A	86.7 ± 18.3 A	80 ± 18.25 A		
293	^a Cocoons were proc	luced over a 7-da	ay period in the br	eeding culture			
294	^b Cocoons were incu	bated at 20°C in	the control soil C o	or in soils spiked w	with different Cuprafor		
295	micro [®] concentration	ns (3.33, 10 and	30 times the RD,	corresponding to	25.8, 77.5 and 232.5		
296	mg kg ⁻¹ of copper, abbreviated Cu3.3, Cu10 and Cu30, respectively).						
297	^c Different letters indicate significant differences among the concentrations ($p < 0.05$).						
298							
299							
300							
301							
302	Table 4. Mean hatcl	hing time (days)	and mean hatchin	ng success (%) of	"naïve" A. caliginosa		
303	cocoons ^a exposed to	Swing Gold ^{®b} .					
	Treatment ^c						

Treatment ^c					
С	SG0.33	SG1	SG3		

Hatching time 31.1 ± 1.1 A 31.3 ± 1.4 A 32.6 ± 3.0 A 34.4 ± 2.6 A(days)Hatching success 100 ± 0.0 A 100 ± 0.0 A 66.7 ± 0.0 B(%)

^a Cocoons were produced over a 7-day period in the breeding culture

^b Cocoons were incubated at 20°C in the control soil C or in soils spiked with different Swing
Gold[®] concentrations (0.33, 1 and 3 times the RD, abbreviated SG0.33, SG1 and SG3,
respectively).

^c Different letters indicate significant differences among the concentrations (p < 0.05).

309

310 4. Discussion

The impacts of pesticides on earthworm reproduction are difficult to assess under 311 natural conditions because several factors can influence on it at the same time (e.g., agro-312 pedoclimatic factors). In this study, we used an ecologically relevant earthworm species, a 313 natural soil, and commercial formulations of pesticides so as to be closer to field conditions. 314 The results revealed negative effects of two fungicides on A. caliginosa reproduction at 315 concentrations not far from those applied in the field. For the Swing Gold[®] fungicide, effects 316 were observed starting from the RD. Concerning the active substances, Silva et al. (2019) found 317 a median concentration of 0.02 mg kg⁻¹ of epoxiconazole and a maximum of 0.16 mg kg⁻¹ in 318 agricultural soil, which is in the range of the RD (0.058 mg kg⁻¹ of epoxiconazole, Table S1). 319 Furthermore, according to the EFSA Scientific Report 46, 1-82 (2005), the no-observed-effect 320 concentration (NOEC) of dimoxystrobin for E. fetida reproduction in a formulation with the 321 same properties as Swing Gold[®] (i.e., 133 g L⁻¹ of dimoxystrobin and 50 g L⁻¹ of epoxiconazole) 322 was estimated to be $<0.0887 \text{ mg kg}^{-1}$ of dimoxystrobin, which is equivalent to half the RD in 323 our experimental conditions (i.e., 10 cm depth and 1.29 kg dm⁻³ soil density). Finally, as regards 324 epoxiconazole, the NOEC for *E. fetida* reproduction was estimated to be 0.084 mg kg⁻¹ (PPDB, 325 326 2017). We carried out our experiments with the endogeic species A. caliginosa, which inhabits agroecosystems and will consequently come in contact with this pesticide. Our results strongly 327

support the conclusion that the Swing Gold[®] formulation is potentially harmful for earthworm reproduction in the field. Concerning the coper fungicide, the effects were observed from 10 times the RD, corresponding to 77.5 mg kg⁻¹ of Cu, which added up to the soil background Cu concentration, leading to around 102.5 mg kg⁻¹. This value is a bit far from the average maximum Cu concentration in vineyard soils in Europe, i.e., 49.26 mg kg⁻¹ of Cu (Ballabio et al., 2018), but could be reached in the field because copper accumulates treatment after treatment.

335

In our experimental conditions (i.e., soil characteristics, 28 % water content, and 336 temperature), the individuals of the control group did not lose weight, indicating that the amount 337 of food was sufficient for individuals. The pesticide concentrations were sublethal for A. 338 caliginosa. Even if the Swing Gold® treatment was applied at three times the recommended 339 dose, it killed 20% of earthworms. This result is in accordance with Bart et al. (2017), who 340 estimated an LC50 at 6.3 times the RD for A. caliginosa. Furthermore, a recent field experiment 341 showed that this fungicide applied at 10 times the recommended dose in the field dramatically 342 affected earthworm density, especially A. caliginosa density (Amossé et al., 2018). 343

A. caliginosa adults produced one cocoon per individual per week in the control 344 345 treatment. This is in accordance with other studies (Lofs Holmin, 1982; Spurgeon et al., 2000), which found between 0.6 to 2.6 cocoons per individual per week at 15 °C in natural soils for A. 346 caliginosa. Cocoon production sharply dropped at the two highest concentrations of both 347 fungicides, although partially explained by the mortality rate in the SG3 treatment. Other 348 studies with the same species reported EC₅₀ values for cocoon production between 68 and 186 349 mg Cu kg⁻¹ (Khalil et al., 1996b; Ma 1988). Similarly, Martin (1986) estimated a NOEC of 50 350 mg Cu kg⁻¹ for Aporrectodea tuberculata, which belongs to the A. caliginosa complex of 351

species (Sims and Gerard, 1999). Mirmonsef et al. (2017) found an EC₅₀ value of 220 mg kg⁻¹ for *Aporrectodea tuberculata*. In our experiment, copper toxicity was a little lower than previously reported because the addition of horse manure may have decreased copper bioavailability (Ma and Rao, 1997) for earthworms and thus alleviated its ecotoxicological effects. Copper toxicity highly depends on the copper form and the soil (EU 2008).

The weight of the cocoons produced by adults exposed to fungicides or chemicals as an 357 ecotoxicological endpoint is not documented in the scientific literature. However, it can reveal 358 key information on the reproductive strategy of A. caliginosa. Neither fungicide impacted the 359 weight of the cocoons produced by the exposed adults, even at the highest concentrations. When 360 361 a cocoon is produced, the vitellus allows for the development of the embryo (Bouché, 1972). A hypothesis could be that when A. caliginosa are exposed to fungicides, their strategy could be 362 to produce fewer cocoons but with the required amount of vitellus to ensure the embryo 363 364 development, but further experiments would be required to investigate that point. The cocoons produced by exposed adults took longer to hatch and a had a lower hatching success, even at 365 the recommended dose of Swing Gold[®]. Previous studies showed that the sperm quality and 366 quantity of earthworms, including A. caliginosa, can be affected by pesticides (Espinoza-367 Navarro and Bustos-Obregon, 2004; Espinoza-Navarro and Bustos-Obregon, 2005; Mosleh et 368 al., 2003; Venter and Reinecke, 1985) and might explain our results. 369

In the second experiment, in which "naive" cocoons were exposed to the two fungicides, we found a mean incubation period of 31 days in the control treatment. This number is underestimated because cocoons were produced over the 7-day period at 15°C before the beginning of the incubation, so that some cocoons had already started incubating. The observed incubation period is in accordance with Holmstrup et al. (1991), who found a cocoon incubation period in petri dishes on a moist filter paper of 36 days at 20 °C for cocoons of *A. caliginosa*.

Several authors recorded a hatching success of more than 90% in a natural soil (Boström and 376 Lofs Holmin, 1996) or in petri dishes (Holmstrup, 2001; Jensen and Holmstrup, 1997) at 15°C. 377 This is in accordance with our results in the control treatment, where 100% of the cocoons 378 hatched. To our knowledge, the monitoring of "naive" cocoons incubated in a contaminated 379 soil has never been documented in the scientific literature. Nevertheless, cocoons from "naive" 380 earthworms can be exposed to pesticides after being produced, although this is less probable 381 than the previous situation (cocoons produced by exposed earthworms). These "naive" cocoons 382 appeared to be less affected by the fungicides than the cocoons produced by the exposed adults. 383 with only significant effects for the Swing Gold[®] fungicide at 3 times the RD. These results 384 mean that the ecotoxicological effects are mainly related to the quality of the cocoons rather 385 than the environment in which they are deposited. Nevertheless, the effects observed with the 386 Swing Gold[®] fungicide indicate that this endpoint should be monitored in pesticide risk 387 assessment since the cocoon stock is crucial for the renewal of populations in the field. This 388 result was significant for the hatching success in the SG3 treatment. In addition, the cocoons 389 390 produced by adults exposed to pesticides under field conditions are generally laid in the same contaminated soil. Thus, the impacts on reproduction could be greater than our observations 391 appear to indicate. 392

393

394 5. Conclusions

In risk assessment procedures, the reproduction test with *Eisenia fetida* directly assesses the impact of chemicals on the number of produced juveniles. This does not provide information about the mechanisms whereby pesticides affect earthworm reproduction. In this study, we took one step further by taking the cocoon stage (i.e., weight, hatching success and time) into account after cocoon production, and by separately exposing "naïve" cocoons to contaminated soils so as to allow for a better understanding of the impact of pesticides on earthworm reproduction.
Our results showed that earthworm reproduction could be affected through adult fertility
(cocoon production and viability of the produced cocoons), but also through the cocoon stock
in the soil. These results highlight that there is a need to assess the impact of pesticides on
earthworm reproduction through both adult exposure and cocoons.

405

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supplementary material

Swing Gold [®]						
Abbreviation	Times RD		nmercial form	Dimo (mg l	2	Epoxiconazole (mg kg ⁻¹)
SG0.33	0.33	0.39	9 x 10 ⁻³	5.2 x	10-2	1.94 x 10 ⁻²
SG1	1	1.16	6 x 10 ⁻³	1.55	x 10 ⁻¹	5.81 x 10 ⁻²
SG3	3	3.48	8 x 10 ⁻³	4.62	x 10 ⁻¹	1.74 x 10 ⁻¹
Cuprafor micro [®]						
Abbreviation	Times I	RD	Commercial (mg kg ⁻¹)	form	Copper (mg kg ⁻¹)	
Cu3.33	3.33		52		25.8	
Cu10	10		155		77.5	
Cu30	30		465		232.5	

Table S1. Concentrations of fungicides tested as commercial forms and active compounds.

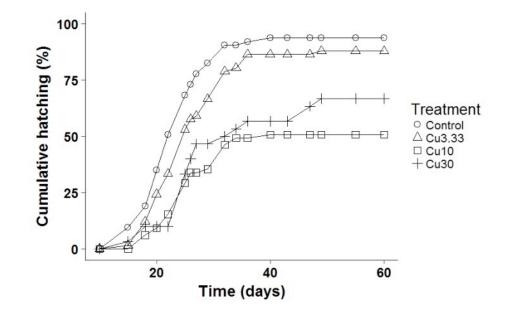
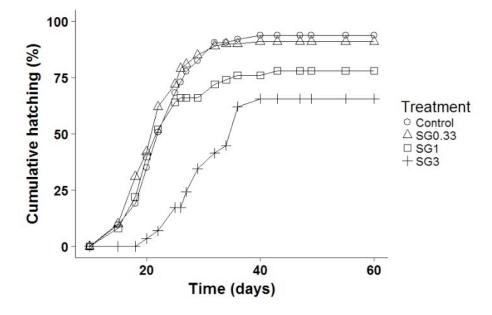
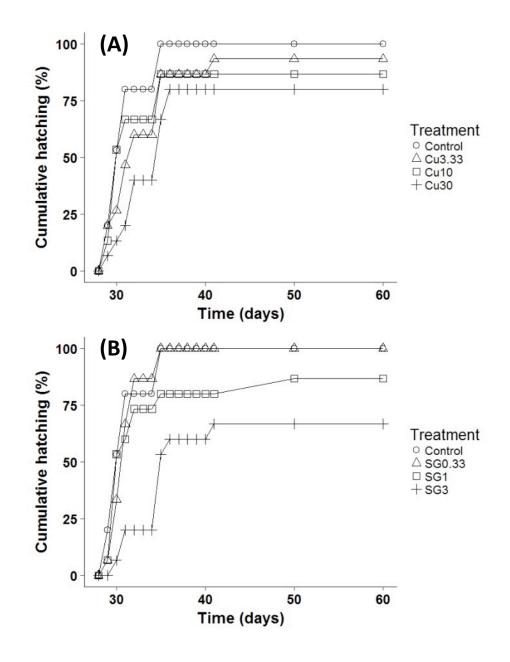


Fig. S1. Hatching dynamics of cocoons produced for a 28 days period in the control soil and in soils spiked with different Cuprafor micro[®] concentrations (3.33, 10 and 30 times the RD corresponding to 25.8, 77.5 and 232.5 mg kg⁻¹ of copper abbreviated Cu3.3, Cu10 and Cu30 respectively).



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Fig. S2. Hatching dynamics of cocoons produced for a 28 days period in the control soil and in
soils spiked with different Swing Gold[®] concentration (0.33, 1 and 3 times the RD abbreviated
SG0.33, SG1 and SG3 respectively).



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Fig. S3. Hatching dynamics of cocoons produced for a 7 days period in the breeding culture and incubated at 20°C in the control soil or in soils spiked with **(A)** different Cuprafor micro[®] concentrations (3.33, 10 and 30 times the RD corresponding to 25.8, 77.5 and 232.5 mg kg⁻¹ of copper abbreviated Cu3.3, Cu10 and Cu30 respectively) or **(B)** different Swing Gold[®] concentrations (0.33, 1 and 3 times the RD abbreviated SG0.33, SG1 and SG3 respectively).

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