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Moving beyond the human–nature dichotomy through biocultural approaches: including ecological well-being in resilience indicators

*Sophie Caillon*¹, *Georgina Cullman*^{2,3}, *Bas Verschuuren*⁴ and *Eleanor J. Sterling*²

ABSTRACT. Diverse and productive ecosystems and human well-being are too often considered opposing targets. This stems mainly from nature being perceived as separate from culture, which results in resilience indicators that focus predominantly on either ecosystems or humans, and that overlook the interplay between the two. Meanwhile, global targets for biodiversity conservation and human well-being have yet to be satisfactorily achieved. We believe that in order to develop effective, culturally appropriate, and equitable conservation strategies that ensure social-ecological resilience, conservation planners and practitioners must conceive of human and ecological well-beings as an interrelated system. By giving nature a voice, and by viewing nature and people as an undifferentiated whole, some indigenous peoples and local communities (IPLC) have philosophical bases for achieving well-being for both humans and nature. Biocultural approaches to conservation ground management in local knowledges, practices, and ontologies. These approaches encompass both the biological and cultural aspects of a system, address complex relationships and feedbacks within human and ecological well-being, and offer flexible frameworks that facilitate synthesis across different metrics, knowledge systems, and ontologies. The process of developing indicators of resilience with a biocultural approach could help (1) overcome the human–nature dichotomy that often makes global approaches incompatible with local approaches by integrating local peoples' diverse forms of relating to nature, (2) reflect two-way feedbacks between people and their environment by focusing on processes, not just final states, and (3) define, measure, and monitor ecological and human well-being as a whole. It can also facilitate dialog between IPLCs and global decision-makers who are disconnected from local realities, and between people from a diversity of disciplinary, ontological, and professional backgrounds.

Key Words: *biocultural approach; conservation; ecological well-being; human well-being; indicator; indigenous peoples and local communities; nature–culture; ontology; resilience; traditional ecological knowledge*

INTRODUCTION

Despite growing awareness and efforts to conserve biodiversity, ecosystems worldwide are still declining (Botsford et al. 1997, Vitousek et al. 1997, Butchart et al. 2010, Newbold et al. 2016) while people still struggle to lead a life in which they themselves can decide how to improve their well-being (UNDP 2016). Policy-makers and scientists too often frame the targets of diverse and productive ecosystems and healthy human societies as if they were in opposition (Raudsepp-Hearne et al. 2010, McShane 2011). This framing shapes the development of indicators for managing biodiversity and human well-being. Most internationally derived indicator sets (e.g., GDP ranking, world development indicators, IUCN Red List index) measure ecosystem and human health separately. Recent efforts to reconcile nature conservation and human development include promoting the integration of social sciences into conservation (Mascia et al. 2003, Agrawal and Ostrom 2006, Brosius 2006, Peterson et al. 2010, Barry and Born 2013, Sandbrook et al. 2013, Bennett et al. 2016, Ives et al. 2017) and the integration of local actors (e.g., community members, NGOs, local government) into research and action through social-ecological systems resilience studies, community-based management, or in situ conservation of agrobiodiversity (Altieri and Merrick 1987, Alcorn 1993, Pinedo-Vásquez and Padoch 1993, Berkes 2008, Liu and Opdam 2014, Ens et al. 2016). These efforts have led to (1) more nuanced human well-being indicators, modified from the Human Development Index, to better integrate material conditions, quality of life (e.g., spiritual dimensions, social connections, environmental quality, and subjective well-

being), and sustainability of well-being (i.e., human, social, economic, and natural capital) (Clark 2014, OECD 2015, Biedenweg et al. 2017, Gross-Camp 2017, Wali et al. 2017), and (2) sets of indicators, such as the United Nations Sustainable Development Goals, that include people-focused and ecological goals but fall short in integrating these domains through attention to the feedbacks and interactions between humans and the environment.

Defining, measuring, and monitoring biodiversity and human well-being indicators separately may result in irrelevant or disruptive actions on the local scale (Jupiter 2017, Sterling et al. 2017b). We advocate for comparable investment in indicators that integrate the specific well-being of ecosystems and the relationship between humans and nature that maintain these well-beings. Equitable conservation strategies can be achieved only if we believe in a joint future for ecological and human well-beings by (1) actively engaging with the diversity of knowledges, practices, and ontologies (i.e., different realities with complex relations between distinct categories of being; for example, between humans and nonhumans, such as landforms, spirits, rocks, trees, energy), (2) moving beyond the dichotomy between people and nature, and (3) giving nature a voice.

We are aware that ecological well-being is difficult if not impossible to define generically, but when grounded in a place and in relation to the peoples who manage that place, it is possible to articulate preferred states as well as specific notions of respect, caring, and responsibility for nature that are held by a given social

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group. We have much to learn from other knowledge traditions (Matulis and Moyer 2016), in particular, the importance of relational values (Chan et al. 2016, Pascual et al. 2017) that apply to interactions with place (i.e., the concepts of place attachment or sense of place) (cf. Massey 1991, Ingold 2000, Poe et al. 2016). Moving beyond the human–nature dichotomy in conservation practices results in a different relationship and hierarchy between humans and nature, and a contingent responsibility on the part of humans to nurture the environment and not just benefit from it (Pascual et al. 2017).

Ignoring the feedbacks and interactions between humans and ecosystems is ahistorical, a denial of the often generations-long relationships between people and their environments. It ignores omnipresent human influence on ecosystems and people's rights to their territories and resources (Gillson and Willis 2004, Heckenberger et al. 2007, van Oudenhoven et al. 2011). In addition, framing conservation objectives in terms of “reverting” to a prior, pristine nature, as is often implied in conservation and restoration strategies, both impedes us from imagining transformative futures and recreates the structural elements that have resulted in present-day environmental degradation and cultural erasure (West 2016). Here, we use the term “nature” when noting human perceptions and conceptual interactions with the environment. We use the more technical term “ecosystem” when presenting scientific realities such as “ecosystems are dynamic” or when discussing human management or use of the material component of the environment. Both terms encompass the processes and interactions that support systems, not just isolated individuals. For culture, we have adopted Bates and Plog's definition (1990:7) “the system of shared beliefs, values, customs, behaviors, and artifacts that the members of society use to cope with their world and with one another, and that are transmitted from generation to generation,” and within the same generation through learning. Thus, culture is holistic, dynamic, ubiquitous, and learned.

We explore the implications of including paired ecological and human well-being in conservation practices through biocultural approaches—ones that include the dynamic interactions between local knowledges, practices, and ontologies, and recognize the generative interrelationships between people and their environment, and the processes and feedbacks that sustain ecological and human well-being (Maffi and Woodley 2010, Gavin et al. 2015, Sterling et al. 2017b, McCarter et al. 2018). Biocultural approaches are gaining traction in resource management arenas and particularly with international conventions and platforms (e.g., Convention on Biological Diversity, Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services [IPBES], United Nations) (Gavin et al. 2015, Pascual et al. 2017, Sterling et al. 2017b). Taking into account the mutualistic relation between local peoples, particularly indigenous peoples and local communities (IPLCs), and the environment (Kuznar 2001, Cairns 2007) does not exclude scientific measurements and monitoring of ecosystem states, processes, interactions, and functions. Local knowledges and western sciences, especially the discipline of community ecology, share common concepts such as connectedness and relatedness (Pioretti and Wildcat 2000), aspects specifically recognized as a core concern to conservation (Zylstra et al. 2014). Western sciences have supported objectives defined by local peoples (Bartlett et al. 2012, Muller 2012, Verschuuren et al. 2015).

The creation and measurement of resilience indicators (capturing the capacity of systems to absorb shocks and disturbances, and to catalyze renewal, adaptation, transformation, and innovation) (cf. Béné et al. 2013) through a biocultural approach can contribute to equitable conservation strategies that achieve well-being for humans and nature. Biocultural approaches recognize and emphasize IPLCs' realities through an emic approach, which aims to illuminate IPLCs' own understandings of their relationships to nature (as opposed to a priori theoretical frameworks that underpin etic approaches). The approaches take into account interactions between humans and nature, and try to integrate connections—including both convergences and divergences—between local and scientific knowledges. The co-construction of indicators of resilience can also facilitate dialog between IPLCs dwelling with nature and other decision-makers, and between people from a diversity of backgrounds (Verschuuren et al. 2014). A focus on the interactions between people and their environment facilitates attention to processes in addition to final states. We propose that conservation planning and policy should address the diversity of knowledges (including classification systems) and practices embedded in different ontologies. Additionally, given human embeddedness within ecosystems, ecological well-being in and of itself should be considered by giving nature a voice. We believe that attention to how we measure and monitor progress toward targets can help bring these conceptual ideas into practice. The ideas in this paper reflect our experiences with peoples dwelling with nature, as well as discussions from workshops and organized sessions that convened diverse stakeholders (including scientists from different disciplines), backgrounds, and countries.

A DIVERSITY OF LOCAL KNOWLEDGES, PRACTICES, AND ONTOLOGIES

Humans in different places around the world make use of different ontologies that shape the worlds they know and see, and how they behave and interact with them. We define ontology as the conceptualization of the nature and relation of being. Ontologies are continuously in interaction with knowledges and practices, and through this interaction, distinct realities are enacted. This dynamic dialectic shapes a myriad of relationships to nature; materiality and immateriality are interconnected, indistinguishable, and produce one another. We are not describing how different groups of people have different perceptions of a single world (a multicultural etic approach), but rather are accepting the radical plurality of worlds, and thus of natures (a multinaturalist emic approach developed by Viveiros de Castro [1998]). Many IPLCs do not consider nature and humans as separate, and their views often differ from western-educated conservation agents (West 2006, Jupiter 2017). Since the 1980s, anthropologists such as Viveiros de Castro (1992), Strathern (1980), and Descola (1986, 2005) have documented such indigenous ontologies. They have described how small-scale societies conceive of themselves as part of their environment, and view their surroundings as kin. As reported by Descola (1986), the Achuar from the Ecuadorian Amazonian forest, for example, confer to nonhumans attributes that are similar to those of humans. Animals, according to the Achuar, have the same spirit, called *wakan*, as humans, and appear as animals only on the outside, as seen by others. Animals have a reflexive consciousness and intentionality; they can express emotions and communicate among themselves or with members of other species, like humans.

Early anthropologists—including Durkheim and Mauss (1901), Evans-Pritchard (1940), Conklin (1954), and Levi-Strauss (1962)—as well as later ethnobotanists (Hunn 1977, Friedberg 1990, Berlin 1992, Ellen 1998) have shown the richness of local knowledges and practices in relation to nature by using IPLCs' lenses. For example, work on ethno-classifications shows that how people know and see the world is also how they classify it (Brown et al. 1976, Friedberg 1992). As such, each ontology results in different ways of seeing, ordering, ranking, and categorizing; this has profound implications for conceiving of problems, identifying solutions, and measuring success toward envisioned futures. The original goal of these early anthropologists and ethnobiologists was to understand societies through a society's own construct of knowledge on the environment, and not to understand the environment itself and its uses (Roué 2014). These ideas were the precursors of the biocultural approach we are promoting in this paper. More recently, “multi-species ethnography” looks to better integrate the strong and changing interconnections between humans and nature. It focuses on the multiple relations people have with their environment and how these relations make people become humans (Kohn 2007, Haraway 2008, Kirksey and Helmreich 2010, Ogden et al. 2013, Tsing 2015). This movement could bring new concepts and methods, and help improve mutualism between human and ecological well-being, especially if it also includes the study of the role of nonhumans.

Conservationists have noted that some aspects of local knowledges have strong parallels with western scientific categories and have aimed to use fragments of that knowledge—particularly those termed “ecological” knowledge—in service of conservation objectives (Gadgil et al. 1993, Huntington 2000, Berkes et al. 2000, Drew 2005, Menzies 2006, Aswani and Lauer 2014). However, local knowledges are dynamic, situational, and relational, and arise from a combination of common heritage and individual experience. They change and adapt in response to social and environmental changes. Knowledge about nature is entwined and built up with other types of knowledge—such as knowledge about death and ancestors, birth and children (West 2005). The different spheres or systems of knowledge developed by IPLCs are interconnected. Moreover, when trying to understand how local knowledges could contribute to conservation, the study of the processes and relationships that are mobilized to accumulate knowledge is often more informative than the end products, such as the number of plants known at the community level (Ellen 1998). It is important to understand how people transmit knowledge vertically (from parents to offspring), horizontally (among peers of the same generation), or obliquely (between less-related generations), thereby mobilizing their kin group, relatives, or external people.

New approaches developed in the context of IPBES emphasize using validation systems from within each knowledge system rather than using one knowledge system (for instance, western science) to validate information from another system (for instance, local knowledge) (Tengö et al. 2014). In part this is because definitions of human relationships to the environment based on different ontologies may conflict because they are not addressing the same system or perhaps the same reality. For example, Blaser (2009) describes how different ontologies and knowledges created conflict around the management of hunting in a park in Paraguay. In an agreement between the managers of

the park and the Yshiro, the local indigenous people, sustainable hunting of anacondas (*Eumectes notaeus*) was explained to the Yshiro as hunting without impairing the hunt for future generations. The Yshiro thus set off to hunt a significant, but in their eyes sustainable, number of anacondas. The Yshiro believe in a culture of reciprocity, where the forest will continuously provide resources as long as relationships between all entities are in equilibrium. This involves communicating with “powers and potencies” of nonhuman entities, which is done through shamans. The conservation biologists did not agree with the resulting behaviors. Their own science-based strategy to achieve sustainability involves population viability of species calculated at a far lower uptake limit. Both parties were acting “correctly” with respect to their own understanding of sustainability. To reduce conflict, the Yshiro and the conservation biologists could have placed greater importance on the co-construction of indicators and operationalization of sustainable harvesting.

Conservation initiatives with IPLCs should work in partnership to apply knowledge from multiple knowledge systems and ontologies, and to better understand the reasons for different proposed strategies. Ignoring local ontologies by giving primacy to western framing of issues can exacerbate political, economic, religious, and educational inequalities, and ultimately frustrate conservation outcomes (Walley 2002, Atran et al. 2005, Bartlett et al. 2012, Verschuuren 2016). Creating channels toward a plurality of ontologies also means that biodiversity conservation strategies in territories of IPLCs cannot simply use conventional actions such as reducing “human pressure” (e.g., West 2005).

NATURE SHOULD BE GIVEN A VOICE

The scientific literature places a substantial emphasis on the positive impact of nature on human well-being, but only rarely the reverse (Biedenweg et al. 2014, Comberti et al. 2015). This lacuna can be explained in part by the fact that western ontologies do not give the same status to nature as to humans. In some other ontologies, rivers, trees, and rocks are animated and relate to people; nature is given a voice and acquires rights to existence, similar to humans (Emmenegger and Tschentscher 1994, Starik 1995, Cullinan 2002, Burdon 2011). Is this expansion of rights “unthinkable” beyond those cultures? No, there are parallels in western history when rights were afforded to groups previously denied similar status to other humans—e.g., slaves, women, and children (Stone 2010 [1972]).

Today, Ecuadorians have overcome the “unthinkable” with respect to nature's existence and rights. Ecuador's courts were the first to uphold the rights of nature. In the country's constitution of 2008, nature “has the right to exist, persist, maintain and regenerate its vital cycles, structure, functions and its processes in evolution” (Constitution of the Republic of Ecuador, Title II, Ch.7 “Rights of nature”: Art.71). The court succeeded in asserting a harmony between people and nature (Becker 2011), thereby establishing a legal basis for the inherent rights of nature, and recognizing the indisputable importance of those rights in the present and for future generations (Daly 2012).

New Zealanders have acknowledged a former national park, Te Urewera, as “a legal entity” with “all the rights, powers, duties, and liabilities of a legal person” (Te Urewera Act 2014: section 11(1)). Te Urewera thus gained the status of an integrated, living whole with rights and interests according to the ontology of

Tūhoe, a Māori group. Māori iwi are well-known for their close relationship with nature and their holistic ontology, called Kotahitanga, which recognizes how each individual's actions affects the collective (Marsden 1992, Mead 2003). Their relations with other humans and nonhumans is clear from the very start of the Act: Te Urewera is described as a “place of spiritual value, with its own mana and mauri” (section 3(2)), and his origin is storied as “*Te Manawa o te Ika a Māui*; it is the heart of the great fish of Maui, its name being derived from Murakareke, the son of the ancestor Tūhoe.[...] Te Urewera expresses and gives meaning to Tūhoe culture, language, customs, and identity” (section 3(4,6)). (According to the Māori dictionary [Moorfield 2011], mana means the “prestige, authority, control, power, influence, status, spiritual power, charisma—mana is a supernatural force in a person, place or object,” and mauri means “life principle, life force, vital essence, special nature, a material symbol of a life principle, source of emotions—the essential quality and vitality of a being or entity. Also used for a physical object, individual, ecosystem or social group in which this essence is located”). The law empowers Tūhoe stewardship of Te Urewera and enables the continuation of this generative relationship between people and place.

Once a natural entity gains legal standing, what are the implications in terms of assessing obligations to it? Is it appropriate to conceive of ecological well-being or ecosystem “health,” or should the scientific lexicon, such as ecosystem state/biodiversity/processes/functions/resilience, continue to predominate? The ecosystem health literature (e.g., Schaeffer et al. 1988, Costanza et al. 1992) proposes measurement in terms of vigor, organization, and resilience (Costanza 2012), and is generally associated with a utilitarian value and western understandings of ecosystems. The term ecosystem health can be used without reference to feedbacks with people, though some authors do recognize these links (Tiwari et al. 1998, Spiegel et al. 2001, Tzoulas et al. 2007). We believe that using the same term “well-being” for humans as well as nonhumans helps reinforce interactions and processes between humans and nature. In addition, the use of well-being encourages a focus not just on the absence of physical illness or decline in ecosystem state, but also on less easily translated elements such as connection to place, or mental and spiritual well-being of nonhumans (West 2005).

Again, New Zealanders are pioneering. The Whanganui River is also granted legal personhood (i.e., a “legal entity with standing in its own right” (Tūtohu Whakatupua, Whanganui Iwi and the Crown (2012): section 2.1.2), and its well-being is highlighted repeatedly in the legal agreement between Whanganui Iwi and the Crown. “*Ko au te awa, ko te awa ko au*—the health and well-being of the Whanganui River is intrinsically interconnected with the health and well-being of the people” (section 1.8.2). “Whanganui Iwi view the Whanganui River as a living being, Te Awa Tupua; an indivisible whole incorporating its tributaries and all its physical and metaphysical elements from the mountains to the sea” (section 1.2). Similarly, on 20 March 2017, a court in Uttarakhand, India ordered that the Ganga and the Yamuna, and all their tributaries, “are declared as juristic/legal persons/living entities having the status of a legal person with all corresponding rights, duties and liabilities of a living person in order to preserve and conserve river Ganga and Yamuna” (Sharma and Singh 2017:11).

But how can a land or a river talk, express its discontent—or the opposite—its well-being? These no-longer-missing actors are still mute. Stone (2010 [1972]), a legal scholar, has already raised this issue: “It is no answer to say that streams and forests cannot have standing because streams and forests cannot speak” (Stone 2010 [1972]:27). Lawyers can speak for corporations or municipalities, so they can do the same for nature in the name of guardians. Even though they are not currently recognized by law, the deities inhabiting the sacred forests and mountains in Tibet have spoken to the Tibetan people through their shamans for millennia, thus maintaining a topocosmic (i.e., the world order that places humans on the same footing as animals and plants) equilibrium (Studley and Jikmed 2016). In New Zealand, a board appointed by Tūhoe Te Uru Taumatua's trustees and by the ministers will manage the land Te Uwerera, and two river guardians (Te Pou Tupua) appointed by the legislation—one by the Crown, one by the Māori people living close to the river—will protect and promote the river's status and well-being (Tūtohu Whakatupua: section 2.21). In India, two officers, named *persons in loco parentis*, are appointed “to promote the health and well being” of the rivers (Sharma and Singh 2017:12).

Does this process once again impose anthropocentric values on nature by deciding what is good for nature and who can speak for nature? If the ventriloquist is not the IPLCs living with and for the river, should it be external specialists in ecology or global policy-makers? Regardless of who speaks for nature, we sustain nature as humans prefer it, but in the Te Urewera and Whanganui river case studies, nature and humans can access the same legal status. In the ontology of the Māori, humans and nonhumans communicate because one belongs to the other, and reciprocally: “I am the River, and the River is me” (Tūtohu Whakatupua: section 1.1). Studley and Jikmed (2016) provides more insight and examples of how juristic personhood is bestowed upon nonhuman entities and nature. In most contexts, if IPLCs self-define human and ecological well-being, and choose relevant indicators via biocultural approaches, we advance toward giving nature a voice.

BIOCULTURAL APPROACHES TO INDICATOR DEVELOPMENT

Biocultural approaches employ participatory methods for goal setting, identification of locally relevant criteria and indicators of resilience, monitoring, and evaluation, and continued adaptive management (e.g., Tipa and Nelson 2008, Verschuuren 2012, Verschuuren et al. 2014, Wali et al. 2017, Sterling et al. 2017a, McCarter et al. 2018). The process of selecting indicators for human well-being in biocultural landscapes, for example in Bolivia (Escobar 2014) as well as in Ghana (Guri and Verschuuren 2014), shows that many IPLCs recognize a direct link between human well-being and landscape quality. Culture and spirituality form key areas for the selection of community-level indicators. These authors provide examples of indicators for human well-being, such as number of sacred sites revitalized and maintained, or cultural festivals celebrated, among many others. The approaches differ significantly in the goal setting, problem identification, and potential solutions in comparison with conventional biodiversity conservation and resource management projects.

Resulting resilience indicators will also undoubtedly differ depending on the stakeholders, as each ontology relies on different ways of seeing, ordering, ranking, validating, and categorizing elements of a system. For example, 96 varieties of taro (*Colocasia esculenta*) in a village in Vanuatu hold high cultural heritage value for farmers, as expressed in varieties' names and histories, their exchange potential emphasized by their rarity, and the pride farmers develop when exhibiting their know-how in open water taro pondfields (Caillon and Lanouguère-Bruneau 2005). In contrast, agronomists note that very few clones of taro were introduced in the country, and they worry about the resulting narrow genetic base; most morphological diversity is due to mutations (Caillon et al. 2006). A resilience indicator regarding the number of named varieties in a village will not have the same value depending on the background and interests of each actor. All metrics of resilience reflect the values of the measurers and their ontologies, so it is important to attempt to accommodate diversity (Pascual et al. 2017).

Several groups have already developed culturally grounded indicators, such as the United Nations University (UNU)-United Nations Environment Programme "Learning from the Practitioners" framework, the UNU-Biodiversity Biocultural Indicator Toolkit, and the Melanesian Well-being Indicators, and via the Mauri Model decision-making framework (Morgan 2006, Subramanian and Pisupati 2009, Malvatumauri National Council of Chiefs 2012, Bergamini et al. 2013, UNU-IAS et al. 2014, Sterling et al. 2017a, b). Current methods that expand the breadth of resilience indicators available for decision-making in biocultural approaches include cultural landscape and community asset mapping, multispecies ethnographies, and the development of community well-being indicators for the conservation of biocultural landscapes (Ens 2012, Verschuuren et al. 2014, Pert et al. 2015, Wali et al. 2017, Thàch et al. 2017, McCarter et al. 2018).

Sometimes locally important, culturally grounded elements are less tangible and harder to measure than global ones, and we need to identify ways to equitably include them (Nic Eoin and King 2013, Satterfield et al. 2013). These may be locally measured and justified through local ontologies, but they are difficult to translate across scales; e.g., local to national or global (Sterling et al. 2017b, Verschuuren et al. 2014). There are ways, however, to mesh locally derived and internationally generated resilience indicators. For instance, results from cultural landscape mapping could be combined with other spatially explicit indicator compilations, such as the ambitious Biodiversity Indicators Dashboard developed by NatureServe (2012), to better visually capture both cultural and biological elements of a system. However, what remains a challenge is to determine action based on the indicators. By giving recognition to a diversity of ontologies, a new negotiation must occur to determine the indicators that significantly impact directions for collective action. This synthesis across ontologies is still a work in progress in most areas.

Similarly, development of indicators that focus on processes and not just outcomes is an emerging field and a critical part of biocultural approaches to indicator development. Most indicators focus only on the end results of such processes. By just focusing on the outcomes, one risks missing critical elements that contributed to those outcomes. For example, in the Pacific

Northwest of North America, in what is now British Columbia, Canada, humans over thousands of years enriched the terrestrial ecosystems (Trant et al. 2016). First Nations resource use, as evidenced by shell middens in nearshore habitation sites, elevated the soil nutrient composition (especially calcium and phosphorous, which are limiting in these forests otherwise), which led to better growing conditions for the forest as a whole. Through intentional burying of shells from the intertidal zone and use of fire in and near habitation sites, First Nations altered soil chemistry and nutrient availability. In particular, Trant et al. (2016) were able to show the effect of nutrient enrichment on the growth and productivity of the western redcedar (*Thuja plicata*), a culturally and economically important species. Outcomes-based indicators would focus on the growth and productivity of the cedar and perhaps fail to capture the long-term process of soil nutrient enrichment through the creation of the shell middens—the generations of care that led to their healthy state. In addition, an ecosystem services framework might emphasize the provisioning service that First Nations enjoy in terms of building materials from the forests, while overlooking the critical feedback from First Nations practices that enrich forest ecosystems. Thus, the processes encompassed by relationships between elements in a system need to be measured, in addition to the elements themselves. Indicators for capturing this process might be the number and frequency of additions to shell midden as well as the number, identity, and distribution of people who continue this practice of enriching middens. Better attention to processes and feedbacks could help us sustainably manage resources and increase the well-being of both humans and nature.

Characterization of the connections between humans and their environment, and how they evolve through time, is a product of ontological pluralism, knowledges, sciences, and the different relationships between humans and nature. Working to understand how ontologies and local viewpoints, motivations, and behaviors can improve processes and outcomes helps ensure that people can react or adapt early to a change in a system and subsequently make it more resilient.

CONCLUSION

We have argued that ecological well-being is an overlooked concept. Better accounting for how human and ecological well-being are inextricably related makes conservation approaches more socially just and equitable. We identify some barriers to considering ecological well-being, including the western dichotomy between nature and culture, and a lack of appreciation for how different constructs of nature in different ontologies permeate our values and actions. Other viewpoints on nature and conservation exist: not those that view people and nature as separate, but those in which people are part of nature, where people and nature continuously interact and produce one another. We then show how a more expansive reframing, through the adoption of biocultural approaches, reinforces a "people as part of nature" perspective, can influence the selection and use of new resilience indicators, and can inform conservation practice. In particular, we advocate for the development of indicators that (1) integrate IPLCs' diverse forms of relating to nature, (2) reflect two-way feedbacks between people and their environment (i.e., services to and from ecosystems) (Comberty 2015), (3) include foreground processes, not just outcomes, and (4) define, measure, and monitor both ecological and human well-being. We need

flexible frameworks and approaches that facilitate synthesis across different metrics, knowledge systems, and ontologies, and that contribute to the creation of a common ground, encompassing human and ecological well-being, on which a joint future for people and nature can be built.

Responses to this article can be read online at:
<http://www.ecologyandsociety.org/issues/responses.php/9746>

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