

Soil and water bioengineering: Practice and research needs for reconciling natural hazard control and ecological restoration

Freddy Rey, C. Bifulco, Gian Baptista Bischetti, Franck Bourrier, Giovanni de Cesare, F. Florineth, F. Graf, M. Marden, S.B. Mickovski, C. Phillips, et al.

► To cite this version:

Freddy Rey, C. Bifulco, Gian Baptista Bischetti, Franck Bourrier, Giovanni de Cesare, et al.. Soil and water bioengineering: Practice and research needs for reconciling natural hazard control and ecological restoration. Science of the Total Environment, 2019, 648, pp.1210-1218. 10.1016/j.scitotenv.2018.08.217. hal-02089533

HAL Id: hal-02089533 https://hal.science/hal-02089533v1

Submitted on 19 Feb2024

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



University for the Common Good

Soil and water bioengineering: practice and research needs for reconciling natural hazard control and ecological restoration

Rey, Freddy; Bifulco, Carlo; Bischetti, Gian Batista; Bourrier, Frank; De Cesare, Giovanni; Florineth, Florin; Graf, Frank; Marden, Mike; Mickovski, Slobodan B.; Pillips, Chris; Peklo, Klaus; Poesen, Jean; Polster, David; Preti, Federico; Rauch, Hans Peter; Raymond, Pierre; Sangalli, Paola; Tardio, Guillermo; Stokes, Alexia

Published in: Science of the Total Environment

DOI: 10.1016/j.scitotenv.2018.08.217

Publication date: 2019

Document Version Author accepted manuscript

Link to publication in ResearchOnline

Citation for published version (Harvard):

Rey, F, Bifulco, C, Bischetti, GB, Bourrier, F, De Cesare, G, Florineth, F, Graf, F, Marden, M, Mickovski, SB, Pillips, C, Peklo, K, Poesen, J, Polster, D, Preti, F, Rauch, HP, Raymond, P, Sangalli, P, Tardio, G & Stokes, A 2019, 'Soil and water bioengineering: practice and research needs for reconciling natural hazard control and ecological restoration', *Science of the Total Environment*, vol. 648, pp. 1210-1218. https://doi.org/10.1016/j.scitotenv.2018.08.217

General rights

Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

Take down policy

If you believe that this document breaches copyright please view our takedown policy at https://edshare.gcu.ac.uk/id/eprint/5179 for details of how to contact us.

1	Soil and water bioengineering: practice and research needs for
2	reconciling natural hazard control and ecological restoration
3	F. Rey ^a , C. Bifulco ^b , G.B. Bischetti ^c , F. Bourrier ^d , G. De Cesare ^e , F. Florineth ^f , F. Graf ^g , M. Marden ^h , S.B.
4	Mickovski ⁱ , C. Phillips ^j , K. Peklo ^k , J. Poesen ^l , D. Polster ^m , F. Preti ⁿ , H.P. Rauch ^o , P. Raymond ^o , P. Sangalli ^q ,
5	G. Tardio ^r , A. Stokes ^s
6	^a Univ. Grenoble Alpes, Irstea, UR LESSEM, 2 rue de la Papeterie, BP 76, 38402 Saint-Martin-d'Hères,
7	France; <u>freddy.rey@irstea.fr</u>
8	^b Universidade de Lisboa, Instituto Superior de Agronomia, Centro de Ecologia Aplicada Prof. Baeta
9	Neves, Lisboa, Portugal ; <u>carlo.bifulco@gmail.com</u>
10	^c Department of Agricultural and Environmental Science,Università degli Studi di Milano, Milan, Italy ;
11	<u>bischetti@unimi.it</u>
12	^d Univ. Grenoble Alpes, Irstea, UR LESSEM, 2 rue de la Papeterie, BP 76, 38402 Saint-Martin-d'Hères,
13	France; <u>franck.bourrier@irstea.fr</u>
14	^e Laboratory of Hydraulic Constructions LCH, École Polytechnique Fédérale de Lausanne EPFL, Station
15	18, CH-1015 Lausanne, Switzerland ; giovanni.decesare@epfl.ch
16	^f Institute of Soil Bioengineering and Landscape Construction, Department of Civil Engineering and
17	Natural Hazards, University of Natural Resources and Life Sciences, Vienna, Austria;
18	florin.florineth@boku.ac.at
19	^g WSL Institute for Snow and Avalanche Research SLF, Flüelastrasse 11, CH-7260 Davos Dorf, Switzerland;
20	graf@slf.ch
21	^h Landcare Research, PO Box 445, Gisborne 4040, New Zealand; <u>mardenm@landcareresearch.co.nz</u>

- ¹School of Engineering and Built Environment, Glasgow Caledonian University, 70 Cowcaddens Rd,
- 23 Glasgow G4 0BA, Scotland, UK; slobodan.mickovski@gcu.ac.uk
- ^jLandcare Research, PO Box 69040, Lincoln, 7640, New Zealand; <u>phillipsC@landcareresearch.co.nz</u>
- ²⁵ ^kI.C.E Klaus PEKLO, Soil and Fluvial Bioengineering Consultancy SARL, Lasmarios, 82160 Parisot, France;
- 26 <u>sbf.peklo@orange.fr</u>
- ¹Department of Earth and Environmental Sciences, KU Leuven, Celestijnenlaan 200E, B-3001 Heverlee,
- 28 Belgium; jean.poesen@kuleuven.be
- ^m Polster Environmental Services, 6015 Mary Street, Duncan, BC. V9L 2G5, Canada; <u>d.polster@telus.net</u>
- ⁿ University Firenze GESAAF, Engineering for Agro-Forestry and Biosystems division, WaVe research
- 31 unit, via san Bonaventura 13, 50145 Firenze, Italy; federico.preti@unifi.it
- ^oInstitute of Soil Bioengineering and Landscape Construction, Department of Civil Engineering and
- 33 Natural Hazards, University of Natural Resources and Life Sciences, Vienna, Austria;
- 34 <u>hp.rauch@boku.ac.at</u>
- 35 ^P Terra Erosion Control Ltd, 308 Hart Street, Nelson, British Columbia, V1L5N5, Canada;
- 36 pierre@terraerosion.com
- ^q Sangalli Coronel y AsociadosSL, Bioingeniería y Paisaje Montesol 24-20016 San Sebastian, Spain;
- 38 <u>sangalli@sangallipaisaje.com</u>
- ^r Technical University of Madrid, Avenida Niceto Alcalá Zamora 6 4D, Getafe, Madrid 28905, Spain;
- 40 gtarcer@gmail.com
- 41 ^sInra, Amap, Cnrs, Ird, University Montpellier, Cirad, Montpellier, France ; <u>alexia.stokes@cirad.fr</u>
- 42 Corresponding author: Freddy Rey, e-mail: <u>freddy.rey@irstea.fr</u>

44 Abstract

45 Soil and water bioengineering is a technology that encourages scientists and practitioners to combine 46 their knowledge and skills in the management of ecosystems with a common goal to maximize benefits 47 to both man and the natural environment. It involves techniques that use plants as living building 48 materials, for: (i) natural hazard control (e.g., soil erosion, torrential floods and landslides) and 49 (ii) ecological restoration or nature-based re-introduction of species on degraded lands, river 50 embankments, and disturbed environments. For a bioengineering project to be successful, engineers 51 are required to highlight all the potential benefits and ecosystem services by documenting the 52 technical, ecological, economic and social values. The novel approaches used by bioengineers raise 53 questions for researchers and necessitate innovation from practitioners to design bioengineering 54 concepts and techniques. Our objective in this paper, therefore, is to highlight the practice and research 55 needs in soil and water bioengineering for reconciling natural hazard control and ecological restoration. 56 Firstly, we review the definition and development of bioengineering technology, while stressing issues 57 concerning the design, implementation, and monitoring of bioengineering actions. Secondly, we 58 highlight the need to reconcile natural hazard control and ecological restoration by posing novel 59 practice and research questions.

60

61 Keywords

62 Be

Benefits; Biodiversity; Ecological engineering; Ecosystem services; Erosion; Vegetation

63 **1.** Development and issues in soil and water bioengineering

64 **1.1. Definition, applications and benefits of soil and water bioengineering**

65 Soil and water bioengineering combines the implementation of techniques using plants as living 66 building materials, through knowledge of their mechanical and/or biological properties (Figure 1) 67 (Barker et al., 2004; Stokes et al. 2004). Bioengineering is a well-recognized component of ecological 68 engineering, itself defined as "the design of sustainable systems, consistent with ecological principles, 69 which integrate human society with its natural environment for the benefit of both" (Mitsch and 70 Jørgensen, 2003; Mitsch, 2012). Bioengineering is used to: (i) control natural hazards (e.g., Norris et 71 al., 2008; Dhital et al., 2013), (ii) restore or reintroduce plant and animal species onto degraded lands 72 and disturbed environments (e.g., Li et al., 2006; Rauch et al., 2014), and (iii) increase soil, air and 73 water quality (e.g., Pretty et al., 2003; Woolsey et al., 2007).

74 Natural hazards such as soil erosion, torrential floods, and landslides, are phenomena that have 75 severe consequences globally (Poesen et al., 2003; Smith and Katz, 2013; Poesen, 2017). The use of 76 vegetation for protecting against natural hazards and attaining economic and/or social goals is typical 77 of traditional forest and hydraulic engineering programs in Europe, such as the 'Restauration des 78 terrains en montagne' (RTM) in France (Vallauri et al., 2002), 'Wildbach und Lawinen Verbauung' 79 (WLV) in Germany and Austria, or 'Sistemazioni Idraulico-Forestali' (SIF) in Italy (Bresci and Preti, 80 2010; Bischetti et al., 2014). Today, the control of these types of hazards using herbaceous and woody 81 vegetation through bioengineering remains a major challenge in areas where technical, 82 socioeconomic, and ecological issues are confounding factors that can hinder success (Phillips et al., 83 2013; Dhital and Tang, 2015). Bioengineering in areas that are difficult to access, e.g. torrential 84 catchments, riverbanks and lakes, as well as on disturbed lands, such as agricultural zones, road and 85 rail embankments, ski slopes, mines, quarries and in urban areas (Lin et al., 2006), requires 86 understanding of the interdependency of hydrological, ecological, and biophysical processes 87 underway at the site. Bioengineering solutions should provide a combination of the benefits of 88 immediate hazard control, comprising techniques such as (Figure 1): (i)brush layers (that provide 89 deep-seated protection), (ii) drain fascines or live pole drains (which drain excess water to allow 90 vegetation establishment), (iii) vegetated crib walls (that immediately protect stream banks), (iv) 91 brush mattresses (providing roughness from establishment against flow), and the long-term 92 stabilization due to plant reinforcement effects. As with any stabilization technique, there is a stress 93 (or load) transfer between the soil and the structure, but, in contrast to other solutions, this initial 94 response is modified by the evolving role of the living material used in the bioengineering structure 95 (Preti and Giadrossich, 2009; Graf and Frei, 2013; Yildiz et al., 2015, 2018; Tardio and Mickovski, 96 2016). This latter feature must be reflected in the bioengineering work design methodologies.

97 Ecological restoration encompasses all actions for repairing degraded lands, with the aim of 98 reestablishing both form and function to attain autonomous and stable ecosystems (Clewell and 99 Aronson, 2013). Ecological engineering in general and bioengineering in particular, can be employed 100 for the restoration of a degraded environment (Mitsch and Jørgensen, 2004). Actions can include: (i) 101 rehabilitating degraded land, which requires techniques aimed at recovering the natural succession of 102 the ecosystem, especially by installing pioneer vegetation and enhancing its development; and (ii) 103 monitoring and maintaining the rehabilitated land, thereby guiding the natural dynamics of the 104 degraded systems so that they recover with a structural and functional autonomy (Aronson et al., 105 1993). A significant advantage of bioengineering actions is the incorporation of vegetation 106 establishment and succession processes into the design stage. Therefore, the need for further 107 intervention or maintenance is reduced and a long-term solution is provided. Bioengineering could 108 also make restoration faster in the sense that, if the path of plant succession is known, it is possible to 109 establish vegetation at the most advanced stage which will be compatible with the soil and 110 microclimatic conditions of the site. Moreover, considering the energy balance of any civil 111 engineering construction, a benefit from using soil bioengineering techniques is that although we 112 need energy for its construction, we "save" energy with the development of the plant biomass. Finally

113 these actions not only include ecological restoration, but also water and soil quality restoration or 114 depollution (e.g., Wang et al., 2008).

115 Soil and water bioengineering is an emerging discipline globally, with regulatory frameworks 116 (including the European Water Framework Directive or more recently the European Green 117 Infrastructure Strategy) introducing the need to implement "soft" techniques for natural hazard 118 control instead of "hard" techniques (engineered concrete and steel structures such as check dams), 119 in the pursuit of restoring degraded environments or preventing further degradation during new 120 construction. Prioritizing soil and water bioengineering techniques is now highly encouraged in the 121 European Community and in many countries worldwide, often promoted through various incentives 122 (e.g., European Commission, 2013). The International Union for the Conservation of Nature (IUCN) is 123 proactively endorsing the use of Nature-based solutions for disaster risk reduction (EcoDRR), and 124 includes soil bioengineering as a technique for protecting against natural hazards (Furuta et al., 2016; 125 Renaud et al., 2016). Questioning practitioners and scientists about their experience, successes, and 126 failures will allow a better understanding of the multiple benefits and services that natural habitats and 127 human populations derive from bioengineering actions (Table 1).

128 **1.2.** An interactive process between researchers and practitioners

Soil and water bioengineering implies an interface between the researcher and the practitioner, i.e. between the improved knowledge base and its application (Stokes et al., 2013, 2014; Mitsch, 2014). The questions raised are increasingly complex and many practitioners are now involved in research projects, improving dialogue and mediation between different stakeholders. Bioengineering projects could also benefit from more multi- and inter-disciplinary approaches, as well as from a better understanding of practical issues experienced by practitioners (e.g., choice of materials, costs, insurance, health and safety, and management of human resources).

136 For those working in the field of bioengineering, a specific framework is required at three levels:

137 (i) <u>questions</u>: identification of the technical, socio-economic, and ecological problems, evaluated by
 138 the practitioners. Researchers and practitioners should then work together to solve specific
 139 objectives.

140 applied research: translation of the technical, socioeconomic, and ecological concerns into (ii) 141 scientific questions, and increase of knowledge by observations and experiments. Soil and water 142 bioengineering can be implemented based on relevant knowledge from several scientific 143 domains, especially geosciences (e.g., geomorphology and soil science), ecology (e.g., 144 restoration ecology, landscape ecology, and plant sciences), engineering (e.g., fluvial hydraulics, 145 civil and geotechnical engineering), sociology (e.g. community engagement and social 146 acceptability of the methods and tools proposed; legitimate design approach), and economics 147 (e.g. project financial management, carbon accounting) (Petrone and Preti, 2010; Stock and 148 Burton, 2011).

149 (iii) management in bioengineering: as soil and water bioengineering is consistent with policies 150 aimed at encouraging "soft" solutions, in particular by including environmental concerns into 151 standard technical practices such as civil engineering, it is essential that current and improved 152 knowledge is included at the work design stage. For example, features such as natural wood 153 deterioration rates (Barré et al., 2017; Tardio and Mickovski, 2016), plant development and 154 successional trajectories (Walker et al., 2009; Gonzalez-Ollauri and Mickovski, 2017a), must be 155 incorporated into the routines and protocols of bioengineering projects. Research results should 156 then be used to develop methods and tools to assist management, conceptualization and action. 157 Adaptation of these tools needs to be performed in collaboration with practitioners, while the 158 knowledge transfer and learning should occur in training courses at every educational level 159 (Mitsch, 2014; Mickovski et al., 2018). The application of knowledge to real cases using newly 160 gained expertise should be verified. Long term monitoring programs, with accurate benchmark 161 data, are also required to compare similar case studies and establish databases on the successes

and failures of various techniques and plant material used (Tardío-Cerrillo and García-Rodriguez,
2016; Perez et al., 2017; Tardio et al., 2017). Such considerations should improve practitioners'
understanding of recent knowledge in ecology and geosciences, as well as increasing scientists'
understanding of practical needs in soil and water bioengineering.

For soil and water bioengineering, all the above features are essential and, for a project to be successful, close interactions between stakeholders and bioengineers are necessary. Improved communication and interaction between the stakeholders will allow bioengineering interventions to both become more effective over time and take advantage of the accumulated experience within the sector.

171

172

2 **2.** Natural hazard control and/or ecological restoration?

173 Depending on the precise objective of a project, the choice of the bioengineering intervention and 174 the desired long-term strategy can vary considerably. For example, soil or streambed erosion can 175 cause different types of damage: (i) loss of topsoil, organic matter and nutrients, which lowers soil 176 quality and hence crop yields and, in turn, threatens agricultural activities (e.g., Jin et al., 2008); it can 177 also cause imminent risk of structural failure of roads, bridges, and railway lines (e.g. Mickovski, 178 2014); (ii) topographic changes (terrain deformation) in the case of gully channel development, 179 landslide triggering and suffusion risk phenomena (e.g. Poesen et al., 2003); (iii) biodiversity loss, 180 which affects vegetation and animal habitats (Mkanda, 2002); (iv) silting of reservoirs, as a 181 consequence of soil erosion and sediment transport, compromising the functioning of these 182 structures (e.g. Schleiss et al., 2016) and (v) increased floods, caused by sediment deposition in river 183 channels (e.g. Steiger et al., 2001). Strategies to control soil erosion rates will vary depending on the 184 type of problem requiring action. For example, if reducing the sediment yield in rivers and reservoirs 185 is the final objective, then the only intervention required is sediment control. Hence, it may be possible to allow hillslope erosion to take place, but to aim at trapping and retaining sediment before it reaches the river channel (Rey, 2009). If riverbed erosion causes lateral displacement and bank failures, and impacts nearby infrastructure, the main objective is to protect facilities with highly specialized and adapted bioengineering solutions, e.g. based on geotechnical preliminary investigations during design and monitoring stage (Peklo, 2015). However, if soil and biodiversity conservation are the final objectives, then both erosion control and ecological restoration are required to prevent soil particles from being detached and removed (Petrone and Preti, 2010).

193 The current challenge when using bioengineering techniques is to define rules that satisfy a set of 194 diverse functions and benefits, particularly those that reconcile natural hazard control and ecological 195 restoration (Figure 2). This approach requires innovation from the practitioners, and also raises new 196 questions for scientists, as part of an interactive process that necessitates the designing and testing of 197 bioengineering actions that reconcile the competing demands of both natural hazard control and 198 ecological restoration. As soil and fluvial bioengineering operates on complex systems (ecosystems) 199 intrinsic adaptive management strategies and feedback loops are necessary to ensure that the 200 project and the intervention is well informed.

201

202

3. From practice to research needs

203 One of the currently pressing challenges is to define bioengineering actions for a range of different 204 situations. Although techniques are well described (e.g., Schiechtl and Stern, 1996, 1997; Gray and 205 Sotir, 1996; Zeh, 2007; Florineth, 2007; Hacker and Johanssen, 2012; EFIB, 2015), quantitative 206 recommendations on how and which materials to use in specific situations are lacking, especially 207 when the objective is to reconcile natural hazard mitigation and ecological restoration. To overcome 208 this knowledge gap, scientists should heed practitioners' needs through discussions during projects,

209 conferences or training, and conduct research at different spatial scales, with specific objectives in210 mind (Figure 3).

3.1. Selection of plant species

212 Numerous studies have already dealt with the performance of specific species in protecting against 213 different hazards (e.g. plant species potential to control gully erosion rates in a Mediterranean 214 ecosystem; De Baets et al. 2009), in a specific climate (Gonzalez-Ollauri and Mickovski, 2016), 215 environment or topographic location (Bochet et al. 2009). Furthermore, an open access plant species 216 database has recently been developed by Perez et al. (2017), allowing users to access the database or 217 add information about species suitable for controlling erosion and shallow landslides in different 218 climates. Traditionally, a limited number of plant species has been used for this purpose, although 219 there are countless species available that could perform equally well, many of which have not yet 220 been tested for suitability (Preti and Petrone, 2013; Perez et al., 2017). In general, although using 221 pioneering species in soil bioengineering projects is sometimes necessary to initiate the successional 222 processes that will maintain vegetation on the site, native species should be preferred over exotic 223 species and ecological succession trajectories should be included in the bioengineering intervention 224 design (Clemente et al., 2016; Xiao et al., 2017).

225 When choosing which species to use on a site, and considering ecological restoration principles, the 226 local and regional environmental conditions need to be considered carefully, so that an optimal and 227 sustainable system is created (Mickovski and van Beek, 2006). For the successful creation of a 228 bioengineering system, the initial phase is of major importance. For the last 30 years, practitioners 229 and scientists have been studying the installation phase, by e.g. examining the relationship between 230 the richness of pioneer species and soil aggregate stability (Pohl et al., 2012), or the efficacy of using 231 mycorrhizal inoculations to improve plant growth and soil structure on eroded soil (Powell, 1980; 232 Yildiz et al., 2015; Bast et al., 2016; Demenois et al., 2017). These studies concluded that the key to

fully understanding ecological processes at eroded sites requires similar long-term experiments in thefield.

235 Species adapted for hazard protection are not necessarily those used for ecological restoration. One 236 of the most difficult questions a manager can ask is: is it better to use only one or a few species that 237 can efficiently mitigate a specific hazard, or should a diverse range of species, sometimes less 238 efficient, be used? Promoting species diversity is generally recommended in ecological restoration 239 projects, but could an increased diversity result in less effective hazard control? Investigating this 240 problem, Erktan et al. (2013) showed that a morphological diversity of plant species used in 241 vegetation barriers did not increase sediment retention in eroded marly gully floors in the French 242 Southern Alps, compared to monospecific barriers (Figure 4). Such a consideration of plant 243 biodiversity is a critical issue, as it generally corresponds to a more ecologically stable system (Preti 244 and Petrone, 2013). A stable and healthy planting system would be less vulnerable to abiotic (e.g., 245 flooding, storms, snow loading and landslides) and biotic stress (e.g., pathogens and grazing). A 246 diversity of plant species would also enable a manager to "cover all options"; for example in the case 247 of a particular species becoming susceptible to abiotic/biotic factors, the loss of one species would be 248 less likely to compromise the aims of the project. Referring to the study of Erktan et al. (2013), Rey 249 and Labonne (2015) suggested using only one species to build brush layers and mats in eroded gully 250 floors, but to use different species between structures along the gully floor, thus reconciling natural 251 hazard mitigation with improved biodiversity.

252

3.2. Selection of bioengineering structures

The choice of the appropriate structure to use in a bioengineering project largely depends on the objective. When considering natural hazard control, the first principle to follow is to use structures and plants that have sufficient mechanical resistance to withstand gravitational or hydrological forces linked to the hazard process. Firstly, although technical drawings describing structures and their mechanical resistance exist in guidelines, we do not necessarily know their origins and performance 258 or reliability for the myriad of field situations likely to be encountered (Schaff et al., 2013). With the 259 exception of gravitational structures (e.g. crib walls and fascines), whose design procedures are well 260 established, and apart from a few cases (e.g. brushlayers and riverbank protection, Bischetti et al., 261 2010), most bioengineering techniques have not yet been sufficiently studied. At the individual plant 262 scale, it is not well known to which topographic (e.g. Bochet et al. 2009, Nadal Romero et al., 2014) 263 and hydrological forces plants resist before failing, and which plant traits are important for 264 mechanical resistance (Burylo et al., 2014), therefore this topic needs significantly more attention 265 from the scientific community. Questions also remain concerning the types of hazard and their 266 different return periods depending on geographic situation and climate, especially under extreme 267 climates such as in tropical countries. This knowledge gap calls for large-scale experiments taking into 268 consideration all the variables and elements to which the structures are subjected (Schwarz et al., 269 2012). Secondly, it should be kept in mind that certain plant species or conditions may destabilize a 270 structure. For example, along river embankments, vegetated crib walls (Figure 1D) can act as a slope 271 buttress or slope break when placed on an eroding embankment to mitigate gully erosion processes 272 (Florineth, 2007). Vegetated crib walls help protect the shoreline and promote revegetation because 273 plants are incorporated within the structure and root growth stabilizes soil (Stangl, 2007). Although 274 some long-term observations have demonstrated that root development and tree stem growth did 275 not adversely affect the structure of vegetated crib walls (K. Peklo, unpublished data), practitioners 276 may hesitate to choose this type of structure because of the way that vegetation interacts with the 277 structure over time.

There is a need to assess more precisely the interrelationships between inert and living materials in bioengineering structures. Questions remain, in particular with regard to wooden structures, where wood decay has to be assessed over time, as vegetation grows and develops around the structure (Barré et al., 2017). Although the role of plants in stabilizing slopes over the long-term is crucial, the growth dynamics of plants used in bioengineering structures are basically unknown and more research is needed to address this gap. Questions particularly arise with regard to the desirable 284 biodegradation of wooden structures. When inert structures are used to enhance vegetation 285 development, for example certain small-scale wooden structures, their initial rigidity has to allow the 286 triggering of new natural processes such as an improved resilience, an improved ecological 287 functioning, and vegetation succession processes, before these inert structures disappear (Stokes et 288 al., 2014). The initial rigidity achieved by means of the inert elements used in the bioengineering 289 intervention must exist long enough so that plants are capable of developing their reinforcing effect. 290 This evolution must be reflected well at the design stage related to a predefined construction aim. 291 Deterioration models of the material used in the work (such as wood) must also be included in the 292 design (Tardio and Mickovski, 2016). All these considerations call for research to evaluate the level 293 and speed of decay of wooden structures, as well as the dynamics of vegetation within 294 bioengineering structures, in relation to the development and diversification of vegetation and to the 295 desired stage of natural hazard mitigation (Barré et al., 2017).

Finally, a maintenance schedule of the living material used in bioengineering structures may be required. These maintenance tasks are usually needed to avoid vegetation becoming too heavy, resulting in the overturning of the bioengineering structure. On riverbanks, this maintenance should also aim at keeping vegetation flexible enough to avoid excessive hydraulic resistance, which can cause an increase of water levels, and to reduce stem and branch breakage, which produce debris obstructing bridges and narrow sections of the river (EFIB, 2015).

302

2 **3.3.** Design of bioengineering structures

To improve the adoption of bioengineering methods by a wider community, new tools (e.g. soilvegetation interaction models, technological frameworks, enhanced methodological approaches and guidelines) must be developed for use in the design of bioengineering structures. In particular, it is necessary to know how to use living plants to attain the expected objectives, and to predict the spatial-temporal development of the installed bioengineering structure, while considering the climate and ecological conditions of the site. During the pre-design phase, the designer must be able to

decide if the bioengineering techniques are feasible or must be used in combination with other conventional techniques (so called 'mixed' techniques), in order to improve the resistance of the structure and resilience of the system. Finally, a global, long-term vision of the project at all spatial scales, from local to catchment, is needed.

313 Improved knowledge is required to design bioengineering structures and optimize their performance 314 in terms of hazard mitigation while enhancing plant diversity. For example, where is it most effective 315 to install a structure at a site? How many plants or cuttings should be used within a structure? 316 Particular attention should be paid to improving our understanding of the efficiency of different plant 317 species and their traits, depending on the final goal of the intervention (Burylo et al., 2014). To design 318 technical solutions, the bioengineer can sometimes use physical scale models. This approach is often 319 not feasible, as plant effects on e.g. slope stability or erosion control and the impact of vegetation on 320 discharge capacity cannot be downscaled appropriately at reasonable laboratory scales such as 1:30 321 to 1:40 (Wilson et al., 2006). Prototype scale (1:1) tests remain a viable solution for scientists and 322 practitioners, but are often not feasible because of time, space, and cost (Schwarz et al., 2012). 323 Therefore, most engineers use readily available numerical geotechnical models that include the 324 effects of vegetation. Different types of models have been implemented over the years to predict 325 landslide risk (see Stokes et al., 2014; Gonzalez-Ollauri and Mickovski, 2017b), many of which 326 calculate a global Factor of Safety (FoS) at the slope level, but are not suitable for calculating the 327 efficacy of individual bioengineering structures. Several uncertainties exist in model parameters, 328 which can be overcome by using a probabilistic approach to e.g. synchronize the mechanical behavior 329 of roots and soil throughout the development of the shear surface (Tardio and Mickovski, 2015). 330 Further information on the hydrological effects of root water uptake is also required, particularly for 331 herbs, shrubs and trees (Chirico et al., 2013; Tron et al., 2014; Arnone et al., 2016; Kim et al., 2017). 332 Upscaling to the catchment level is still a significant challenge, partially because of a lack of suitable 333 data to either parameterize or validate models, but also because of a lack of understanding of 334 biophysical processes at different scales. However, Rossi et al. (2017) demonstrate that the physical

landslide model LAPSUS_LS is suitable for calculating the effects of vegetation on slope stability at the
 catchment level. Nevertheless, parallel studies should investigate how soil hydrological and physical
 processes at the slope level are altered over a larger scale and vice versa (Bogaard and Greco, 2015).

338

3.4. Reconciling qualitative experience and quantitative engineering

339 Statutory constraints related to building structures used for natural hazard control require quantitative 340 engineering methodology, whereas bioengineering often, but not always, comprises more qualitative 341 experience. Thus there is a need to develop research that will reconcile these qualitative and 342 quantitative issues for natural hazard control, but also for ecosystem restoration. Methods for 343 identifying precise performance thresholds for bioengineering installations at local scales could be very 344 helpful. Questions to ask include: (i) is a stepwise strategy necessary and therefore an initial plant 345 protection plan needed? (ii) what is the most efficient spatial distribution of bioengineering structures 346 and plants for hazard control with regard to the physical forces to which they are subjected? (iii) what is 347 the necessary, but adequate, rate of vegetation cover to control a given natural hazard, while 348 considering also ecological processes at this spatial level? Answers to these questions are strongly 349 related to the objectives, which will be different in the case of natural hazard mitigation or ecological 350 restoration. A need also expressed by practitioners is to define appropriate indicators that allow 351 managers to determine the thresholds of efficiency when reconciling natural hazard control and 352 ecological restoration. Finally, the need for case study analysis in terms of bioengineering work 353 performance has been suggested as a useful tool for proposing improvements at the design, 354 construction, and monitoring stages.

355 One further crucial issue determining the success of a bioengineering project is to know if bioengineers 356 can have the financial freedom to create the "best" solution to each problem (EFIB, 2014). For example, 357 for a project in which ecological restoration is required and when significant financial means can be 358 used to implement optimal actions, such actions are able to achieve effective restoration, and enable 359 the damaged ecosystem to recover its original condition. But when constraints are imposed in 360 budgeting and financing bioengineering projects, such as in underdeveloped countries, it is necessary to 361 minimize interventions. In this later case, complete 'restoration' is not economically possible. Therefore, 362 recovering the ecosystem to a state before degradation is not the final objective, but reaching a specific 363 ecosystem objective in line with the current technical, socio-economic and ecological problems, such as 364 controlling a specific natural hazard, becomes the main aim (Rey, 2009). This situation also calls for 365 reconciling possible qualitative issues linked to the definition of precise objectives of a bioengineering 366 project, and quantitative engineering, corresponding to the design of the bioengineering structures and 367 the related financial means.

368

58 **3.5.** Defining actions at the catchment and landscape scales

369 Soil and water bioengineering techniques are usually targeted to discrete locations, whereas their 370 design often needs to be considered at the catchment and landscape scales (Bifulco et al., 2015). The 371 variables with the greatest influence on a bioengineering structure's final design are usually 372 structural, hydraulic or related to plant characteristics, especially when natural hazard mitigation is 373 the main objective. Using bioengineering for natural hazard mitigation at the catchment scale, 374 especially with regard to water and sediment transport, implies taking into account the connectivity 375 between slopes and the river or gully channel, and between upstream and downstream parts of the 376 catchment. For example with respect to flood reduction, a key objective can be to reduce runoff and 377 particularly to interrupt the fine sediment connectivity between various parts of the catchment (e.g. 378 Verstraeten et al., 2006; Borselli et al. 2008; Rey and Burylo, 2014; Mekonnen et al. 2015). The spatial 379 distribution of control measures such as planted areas also needs to be considered at the wider 380 catchment scale, as does the climatic regime. In some cases it will be possible to provide a level of 381 control for more frequent climatic events that generate sediment. Nevertheless, designing structures 382 or bioengineering solutions to deal with extreme, infrequent or high magnitude events remains 383 problematic.

384 The need to define strategies for bioengineering at the catchment and landscape scales is well 385 illustrated in the framework of Green infrastructure (GI), interconnected networks of natural or semi-386 natural sites able to provide environmental, social, and economic benefits to human populations 387 (known as ecosystem services). Soil bioengineering can play a fundamental role in creating networks 388 because their capability is to bridge natural and man-made environments. It allows the restoration of 389 ecosystems and is an effective tool in the implementation of GI (EFIB, 2014). Although GI is included 390 in several EU and International Agencies programs and policies (e.g. EU 2020 Biodiversity Strategy, 391 Natura 2000, UNEP-DHI 2014 Green Infrastructure Guide for Water Management), the use of soil 392 bioengineering techniques is generally implicit. The success in implementing a GI strategy strongly 393 depends on the achievements of bioengineers. Soil bioengineering techniques, by including 394 vegetation as an intrinsic component of installations, are able to provide several functions, such as 395 slope and riverbank stabilization and protection from soil erosion, as well as habitat for animals, 396 microclimate regulation and recreational use (Table 1, Stokes et al., 2014). Therefore, design of 397 bioengineering structures can be thought for allowing them to provide both slope stabilization and 398 ecological restoration (Figure 5).

399 The Blue infrastructures (BI) framework is also a good illustration of the complex questions facing 400 bioengineers. Worldwide, two contrary eco-geomorphological management practices co-exist for 401 rivers. In some catchments, aggradation of the river channel occurs, a phenomenon caused by 402 excessive fine sediment in the river. Habitats for fish reproduction may be damaged, flood risks 403 increase, and hydroelectric reservoirs can fill with sediment (Rey, 2009). In contrast, other catchments 404 suffer from a lack of bed load in the river. As a consequence, groundwater levels can decrease and 405 river beds incise, causing damage such as bridge destabilization (Liébault et al., 2005). Vegetation 406 cover in the surrounding landscape is an important factor controlling the erosion responsible for 407 sediment yield in rivers. In case of bedload excess, eroding slopes and riverbanks are controlled 408 through bioengineering measures and revegetation efforts, considered as restoration actions on 409 degraded land (e.g., Vallauri et al., 2002). Conversely, where a deficit of bedload exists, slope erosion

410 can be reactivated by destroying the vegetation on highly erodible soils. Sometimes, both situations 411 co-exist within the river's catchment, but generally in different parts and at different times, as stated 412 by Liébault et al. (2005) who recorded both aggradation followed by degradation in the same 413 catchment pre-reforestation and post-reforestation. Thus the influence of vegetation on a river's 414 sediment production, especially in mountainous areas, is often difficult to understand.

415 Another situation showing the difficulty in reconciling natural hazard mitigation and ecological 416 restoration is the management of dams on rivers. These structures often have a role in managing 417 floods and many have been constructed over a century ago in different countries (e.g, Vallauri et al., 418 2002 on French experience). However, today these structures are blamed for representing obstacles 419 to aquatic fauna, and programs for removing them are developed, calling into question the impact of 420 this kind of action on the river's stability. One solution consists in replacing dams with rough rock 421 ramps with integrated fishway (Figure 6). All these examples call for more discussion between 422 scientists and practitioners, as well as a better assessment of current knowledge. There is an urgent 423 need to consider new research strategies and to determine whether ecological restoration actions 424 should be carried out on areas where natural hazards occur, with different ecological and socio-425 economic issues calling for different solutions in the management of these hazards.

426 **4.** Conclusion

In the sections above, we highlighted the practice and research needs in soil and water bioengineering through a critical review of the definition and development of bioengineering technology, while stressing the issues about the design, implementation, and monitoring of bioengineering actions. Based on the critical analysis presented above, we conclude that there is a need to reconcile natural hazard control and ecological restoration by posing new applied research questions aimed at meeting this purpose. More importantly, there is a need to define sound techniques that reconcile natural hazard control and ecological restoration. The key considerations helping succeeding bioengineering actions in the future can be summarized as: (i) considering a multidisciplinary approach for soil and water bioengineering projects, (ii) establishing practical guidelines and tools for designing bioengineering structures, (iii) implementing monitoring stages in bioengineering projects, (iv) transmitting knowledge and know-how on soil and water bioengineering, (v) analyzing existing bioengineering works in terms of their performance, successes and failures, and (vi) continuing to identify the needs of the bioengineering professional sector.

440

441 Acknowledgements

442 We thank the French "Ministère de la transition écologique et solidaire" and the H2020 Erasmus 443 project ECOMED (575976-EPP-1-2016-1-ES-EPPKA2-K1) for financial support. We also thank Mélanie 444 Burylo for ideas and graphical assistance in Figure 2. This paper has been initiated from a discussion 445 at the 4th International Conference "Soil bio- and eco-engineering: the use of vegetation to improve 446 slope stability" which was held from 11–14 July 2016 at the University of Sydney, Australia. We are 447 grateful to members of the European Federation for Soil Bioengineering (EFIB) for useful discussion. 448 Funding was provided (A.S.) by theTALVEG2project (funded by Région Languedoc-Roussillon-Midi-449 Pyrénées and Projet Collaboratif Régional fond FEDER) and the BMU (Germany) International Climate 450 Initiative funded project "Ecosystems Protecting Infrastructure and Communities" (EPIC). Finally the 451 authors greatly thank the four anonymous reviewers for their helpful comments.

452

453

454 **References**

- 455 Arnone, E., Caracciolo, D., Noto, L.V., Preti, F., Bras, R.L., 2016. Modeling the hydrological and 456 mechanical effect of roots on shallow landslides. Water Resources Research 52,8590–8612.
- Aronson, J., Floret, C., Le Floc'h, E., Ovalle, C., Pontanier, R., 1993. Restoration and rehabilitation of
 degraded ecosystems in arid and semi-arid lands. I. A view from the South. Restoration
 Ecology 1, 8-17.
- Barker, D.H., Watson, A.J., Sombatpanit, S., Northcutt, B., Magliano, A.R., (Eds). 2004. Ground and
 water bioengineering for erosion control and slope stabilization. Science Publishers Inc,
 Enfield, NH.
- Barré, J.B.B., Bourrier, F., Bertrand, D., Thévenon, M.F., Rey, F., 2017. Assessment of the level of decay
 of small-diameter silver fir logs degraded in natural conditions in the French Northern Alps
 using two different methods. Ecological engineering 109, 240-248.
- Bast, A., Wilcke, W., Graf, F., Lüscher, P., Gärtner, H., 2016. Does mycorrhizal inoculation improve plant
 survival, aggregate stability, and fine root development on a coarse-grained soil in an alpine
 eco-engineering field experiment? Journal of Geophysical Research Biogeosciences 121,
 2158–2171.
- Bifulco, C., Pereira, A., Ferreira, V., Mota, A., Martins, L., Rego, F., 2015. Renewing dangerous highway
 slopes in Portugal. Ingenieurbiologie 3, 43-50.
- Bischetti, G.B., Chiaradia, E.A, D'Agostino V., Simonato, T., 2010.Quantifying the effect of brush
 layering on slope stability. Ecological Engineering 36, 258-264.
- 474 Bischetti, G.B., Di Fi Dio, M., Florineth, F. 2014. On the origin of soil bioengineering. Landscape
 475 Research,1-13.

- 476 Bochet, E., Garcia-Fayos, P., Poesen, J., 2009. Topographic thresholds for plant colonization on semi-477 arid eroded slopes. Earth Surface Processes and Landforms 34, 1758-1771.
- 478 Bogaard, T.A., Greco, R., 2015. Landslide hydrology: from hydrology to pore pressure. Wiley 479 Interdisciplinary Reviews: Water 3, 439-459.
- Borselli, L., Cassi, P., Torri, D., 2008. Prolegomena to sediment and flow connectivity in the landscape:
 A GIS and field numerical assessment. Catena 75, 268–277.
- Bresci, F., Preti, F., 2010. A historical survey on the evolution of some forest watershed management
 techniques (part II: stream channel works). Journal of Agricultural Engineering1, 13-22.
- Burylo, M., Dutoit, T., Rey, F., 2014. Species traits as practical tools for ecological restoration of marly
 eroded lands. Restoration Ecology 22, 633-640.
- Chirico, G.B., Borga, M., Tarolli, P., Rigon, R., Preti, F., 2013. Role of vegetation on slope stability under
 transient unsaturated conditions. Procedia Environmental Sciences 19, 932–941.
- 488 Clemente, A.S., Moedas, A.R., Oliveira, G., Martins-Loução, M.A., Correia, O., 2016. Effect of
 489 hydroseeding components on the germination of Mediterranean native plant species.
 490 Journal of Arid Environments 125, 68-72.
- Clewell, A.F., Aronson J., 2013. Ecological restoration: principles, values, and structure of an emerging
 profession. Society for Ecological Restoration, 336 p.
- 493 De Baets, S., Poesen, J., Reubens, B., Muys, B., De Baerdemaeker, J. & Meersmans, J., 2009.
 494 Methodological framework to select plant species for controlling rill and gully erosion:
 495 application to a Mediterranean ecosystem. Earth Surface Processes and Landforms 34, 1374496 1392.

- Demenois J., Rey F., Stokes A. Carriconde F., 2017. Does arbuscular and ectomycorrhizal fungal
 inoculation improve soil aggregate stability? A case study on three tropical species growing in
 ultramafic Ferralsols. Pedobiologia 64, 8-14.
- 500 Dhital, Y.P., Kayastha, R.B., Shi, J., 2013. Soil bioengineering application and practices in Nepal. 501 Environmental Management 51, 354-364.
- 502 Dhital, Y.P., Tang, Q., 2015. Soil bioengineering application for flood hazard minimization in the 503 foothills of Siwaliks, Nepal. Ecological Engineering 74, 458-462.
- 504 European Commission (2013). Communication from the Commission to the European Parliament, the
- 505 Council, the European Economic and Social Committee and the Committee of the Regions

506 'Green Infrastructure (GI) — Enhancing Europe's Natural Capital' (COM/2013/0249 final).

- 507 EFIB. 2015. European guidelines for soil and water bioengineering. European Federation of Soil 508 Bioengineering.
- 509 EFIB. 2014. Soil and Water Bioengineering and Green infrastructure. VIIIe Congress EFIB-AEIP-APENA
 510 Vitoria Gasteiz, Spain.
- Erktan, A., Cécillon, L., Roose, E., Frascaria-Lacoste, N., Rey, F., 2013. Morphological diversity of plant
 barriers does not increase sediment retention in eroded marly gullies under ecological
 rehabilitation (Southern Alps, France). Plant and soil 370, 653-669.
- 514 Florineth, F., 2007. Bioengineering methods of strengthening slopes in the Alps: protection of deep 515 soil and subsoil horizons. Environmental Planning and Management 2, 35-45.
- Furuta, N., Monty, F., Murti, R., 2016. Helping nature help us: transforming disaster risk reduction
 through ecosystem management. IUCN Publication, Switzerland.

- 518 Gonzalez-Ollauri, A., Mickovski, S.B., 2016. Using the root spread information of pioneer plants to 519 quantify their mitigation potential against shallow landslides and erosion intemperate humid 520 climates. Ecological Engineering 95, 302-315.
- 521 Gonzalez-Ollauri, A., Mickovski, S.B., 2017a. Plant-soil reinforcement response under different soil 522 hydrological regimes. Geoderma 285, 141-150.
- 523 Gonzalez-Ollauri, A., Mickovski, S.B., 2017b. Shallow landslides as drivers for slope ecosystems 524 evolution and biophysical diversity. Landslides 14, 1699-1714.
- 525 Graf, F., Frei, M.,2013. Soil aggregate stability related to soil density, root length, and mycorrhiza using
 526 site-specific Alnus incana and Melanogaster variegatus s.l. Ecological Engineering 57, 314527 323.
- 528 Gray, D.H., Sotir, R.B., 1996. Biotechnical and soil bioengineering slope stabilization: a practical guide 529 for erosion control. John Wiley & Sons, New York, NY.

530 Hacker, E., Johanssen, R., 2012. Ingenieurbiologie. Eugen Ulmer KG, Stuttgart.

- Jin, K., Cornelis, W.M., Gabriels, D., Schiettecatte, W., De Neve, S., Lu, J., Buysse, T., Harmann, R.,
 2008. Soil management effects on runoff and soil loss from field rainfall simulation. Catena75,
 191-199.
- Kim, J.H., Fourcaud, T., Jourdan, C., Maeght, JL., Mao, Z., Metayer, J., Meylan, L., Pierret, A., Rapidel,
 B., Roupsard, O., de Rouw, A., Villatoro Sanchez, M., Wang, Y., Stokes, A., 2017. Vegetation as
 a driver of temporal variations in slope stability: The impact of hydrological processes.
 Geophysical Research Letters 44, 4897-4907.
- Li, X., Zhang, L., Zhang, Z., 2006. Soil bioengineering and the ecological restoration of riverbanks at
 the Airport Town, Shanghai, China. Ecological Engineering 26, 304-314.

- Liébault, F., Gomez, B., Page, M., Marden, M., Peacock, D., Richard, D. and Trotter, C. M., 2005. Landuse change, sediment production and channel response in upland regions. River Research and Applications 21, 739–756.
- Lin, W.T., Lin, C.Y., Chou, W.C., 2006. Assessment of vegetation recovery and soil erosion at landslides
 caused by a catastrophic earthquake: A case study in Central Taiwan. Ecological Engineering
 28, 79-89.
- 546 Mekonnen, M., Keesstra, S., Stroosnijder, L., Baartman, J., Maroulis, J., 2015. Soil conservation 547 through sediment trapping: a review. Land Degradation and Development 26, 544-556.
- 548 Mickovski, S.B. 2014. Resilient design of landslip prevention measures: a case study. Proceedings of
 549 the ICE Forensic Engineering 168, 96-106.
- Mickovski, S.B., Van Beek, L.P.H., 2006. A decision support system for the evaluation of eco engineering strategies for slope protection. Geotechnical and Geological Engineering 24, 483 498.
- Mickovski, S.B., Tardio Cerrillo, G., Sangalli, P., Perez, J., Thomson, C.S., Gallagher, C., 2018.
 Development of Eco-Engineering Sector Specific Routines and Curricula for the
 Mediterranean Region. In: Kallel A., Ksibi M., Ben Dhia H., Khélifi N. (eds) Recent Advances in
 Environmental Science from the Euro-Mediterranean and Surrounding Regions. EMCEI 2017.
 Advances in Science, Technology & Innovation (IEREK Interdisciplinary Series for Sustainable
 Development). Springer, Cham. pp 937-939.
- 559 Mitsch, W.J., 2012. What is ecological engineering? Ecological Engineering 45, 5-12.
- 560 Mitsch, W.J., 2014. When will ecologists learn engineering and engineers learn ecology? Ecological
 561 Engineering 65, 9-14.

- 562 Mitsch, W.J., Jørgensen, S.E., 2003. Ecological engineering: a field whose time has come. Ecological 563 Engineering 20, 363-377.
- 564 Mitsch, W.J., Jørgensen S.E., 2004. Ecological engineering and ecosystem restoration, John Wiley &
 565 Sons.
- 566 Mkanda, F.X., 2002. Contribution by farmers' survival strategies to soil erosion in the Linthipe River
 567 Catchment: Implications for biodiversity conservation in Lake Malawi/Nyasa. Biodiversity and
 568 Conservation 11, 1327-1359.
- Nadal-Romero, E., Petrlic, K., Verachtert, E., Bochet, E., Poesen, J., 2014. Effects of slope angle and
 aspect on plant cover and species richness in a humid Mediterranean badland. Earth Surface
 Processes and Landforms 39, 1705–1716.
- Norris, J.E., Stokes, A., Mickovski, S.B., Cammeraat, E., van Beek, R., Nicoll, B.C., Aachim, A., (Eds).
 2008. Slope stability and erosion control: ecotechnological solutions. Springer, Dordrecht, The
 Netherlands.
- 575 Peklo, K., 2015. L'apport d'un programme de modélisation hydrodynamique bidimensionnel au travail
 576 de l'ingénieur. Ingenieurbiologie 3, 13-17.
- 577 Perez J., Condes Salazaar R., Stokes A., 2017. An open access database of plant species useful for 578 controlling soil erosion and substrate mass movement. Ecological Engineering 99, 530-534.
- Petrone, A., Preti, F., 2010. Soil bioengineering for risk mitigation and environmental restoration in a
 humid tropical area. Hydrology and Earth System Sciences 14, 239-250.
- Phillips, C.J., Rey, F., Marden, M., Liébault, F., 2013. Revegetation of degraded steeplands in France
 and New Zealand: a comparison of geomorphic and policy responses. New Zealand Journal of
 Forestry Science 43, 14.

- Poesen, J. 2017. Soil erosion in the Anthropocene: research needs. Earth Surface Processes and
 Landforms 43, 64-84.
- 586 Poesen, J., Nachtergaele, J., Verstraeten G., Valentin, C., 2003. Gully erosion and environmental 587 change: Importance and research needs. Catena 50, 91-133.
- Pohl, M., Graf F., Buttler A., Rixen C., 2012. The relationship between plant species richness and soil
 aggregate stability can depend on disturbance. Plant and Soil 355, 87-102.
- 590 Powell, C.L.L., 1980. Mycorrhizal infectivity of eroded soils. Soil Biology and Biochemistry 12, 247-250.
- 591 Preti, F., Giadrossich, F., 2009. Root reinforcement and slope bioengineering stabilization by Spanish

592 Broom (Spartium junceum L.). Hydrology and Earth System Sciences 13, 1713-1726.

- 593 Preti, F., Petrone, A., 2013. Soil bioengineering for watershed management and disaster mitigation in
 594 Ecuador: a short-term species suitability test. iForest 6, 95-99.
- 595 Pretty, J.L. Harrison, S.S.C., Shepherd, D.J., Smith, C., Hildrew, A.G., Hey, R.D., 2003. River
 596 rehabilitation and fish populations: assessing the benefit of instream structures. Journal of
 597 Applied Ecology 40, 251-265.
- Rauch, H.P., Sutili, F.J., Hörbinger, S., 2014. Installation of a riparian forest by means of soil bio
 engineering techniques—Monitoring results from a river restoration work in Southern Brazil.
 Open Journal of Forestry 4, 161-169.
- Renaud, F.G., Sudmeier-Rieux, K., Estrella, M., Nehren, U. (Eds.) 2016. Ecosystem-Based Disaster Risk
 Reduction and Adaptation in Practice. Advances in Natural and Technological Hazards
 Research. Springer.

- Rey, F., 2009. A strategy for fine sediment retention with bioengineering works in eroded marly
 catchments in a mountainous Mediterranean climate (Southern Alps, France). Land
 Degradation and Development 20, 210-216.
- Rey, F., Burylo, M., 2014. Can bioengineering structures made of willow cuttings trap sediment in
 eroded marly gullies in a mountainous and Mediterranean climate (Southern Alps, France)?
 Geomorphology 204, 564-572.
- 610 Rey, F., Labonne, S., 2015. Resprout and survival of willow (Salix) cuttings in marly gullies in a 611 mountainous Mediterranean climate: a real size experiment in the Francon catchment 612 (Southern Alps, France). Environmental Management 56, 971-983.
- Rossi, L.M.W., Rapidel, B., Roupsard, O., Villatoro, M., Mao, Z., Nespoulous, J., Perez, J., Metselaar, K.,
 Schoorl, JM., Claessens, L., Stokes, A., 2017. Sensitivity of the landslide model LAPSUS_LS to
 vegetation and soil parameters. Ecological Engineering 109, 249-255.
- Schaff, S.D., Pezeshki, S.R., Shields F.D., 2013. Effects of soil conditions on survival and growth of black
 willow cuttings. Environmental Management 31, 748-763.
- Schiechtl, H.M., Stern, R., 1996. Ground bioengineering techniques for slope protection and erosion
 control. Blackwell Science, Oxford, UK, 146 p.
- Schiechtl, H.M., Stern R., 1997. Water bioengineering techniques for watercourse bank and shoreline
 protection. Blackwell Science, Oxford, UK, 186 p.
- Schleiss, A. J., Franca, M. J., Juez, C. and De Cesare, G., 2016. Reservoir sedimentation, vision paper.
 Journal of Hydraulic Research 54, 595-614.
- Schwarz, M., Cohen, D., Or, D. 2012. Spatial characterization of root reinforcement at stand scale:
 theory and case study. Geomorphology 171–172, 190-200.

- Smith, A.B., Katz, R.W., 2013. US billion-dollar weather and climate disasters: data sources, trends,
 accuracy and biases. Natural Hazards 67, 387-410.
- Staab, K., Yannelli, F.A., Lang, M., Kollmann, J., 2015. Bioengineering effectiveness of seed mixtures
 for road verges: Functional composition as a predictor of grassland diversity and invasion
 resistance. Ecological Engineering 84, 104-112.
- 631 Steiger, J., Gurnell, A.M., Petts, G.E., 2001. Sediment deposition along the channel margins of a reach
 632 of the middle river Severn, UK. River Research and Applications 17, 443-460.
- Stock, P., Burton, R.J.F., 2011. Defining Terms for Integrated (Multi-Inter-Trans-Disciplinary)
 Sustainability Research. Sustainability 3, 1090-1113.
- 635 Stokes, A., Douglas, G., Fourcaud, T., Giadrossich, F., Gillies, C., Hubble, T., Kim, J.H., Loades, K., Mao,
- 636 Z., McIvor, I., Mickovski, S.B., Mitchell, S., Osman, N., Phillips, C., Poesen, J., Polster, D., Preti,
 637 F., Raymond, P., Rey, F., Schwarz, M., Walker, L.R., 2014. Ecological mitigation of hillslope
 638 instability: ten key issues facing practitioners and researchers. Plant and Soil 377, 1-23.
- Stokes, A., Raymond, P., Polster, D., Mitchell, S.J., 2013. Engineering the ecological mitigation of
 hillslope stability research into the scientific literature. Ecological Engineering 61, 615-620.
- Stokes, A., Spanos, I., Norris, J.E., Cammeraat E., (Eds). 2004. Eco- and ground bioengineering: the use
 of vegetation to improve slope stability. Developments in plant and soil sciences 103, CDROM.
- Tardío-Cerrillo, G. and García-Rodriguez, J.L., 2016. Monitoring of erosion preventive structures based
 on eco-engineering approaches: the case of the mixed check dams of masonry and forest
 residues. Journal of Engineering Science and Technology Review 9, 103–107.

- Tardio, G., Mickovski, S.B., 2015. Method for synchronisation of soil and root behaviour for
 assessment of stability of vegetated slopes. Journal of Ecological Engineering 82,222–230.
- Tardio, G., Mickovski, S.B., 2016. Implementation of eco-engineering design into existing slope
 stability design practices. Ecological Engineering 92, 138-147.
- Tardio, G., Mickovski, S.B., Stokes, A., Devkota, S., 2017. Bamboo structures as a resilient erosion
 control measure. Proceedings of the Institution of Civil Engineers 170, 72-83.
- Tron, S., Dani, A., Laio, F., Preti, F., Ridolfi, L., 2014.Mean root depth estimation at landslide slopes.
 Ecological Engineering 69, 118-125.
- Vallauri, D., Aronson, J., Barbero, M., 2002. An analysis of forest restoration 120 years after
 reforestation on badlands in the Southwestern Alps. Restoration Ecology 10, 16-26.
- Vandaele, K., 2010. From mudflow prevention to ecosystem services development, The Melsterbeek
 catchment, Sint-Truiden, Belgium. The Economics of Ecosystems & Biodiversity (TEEB).
 TEEBcase: Changed agro-management to prevent muddy floods, Belgium. Available at:
 www.TEEBweb.org.
- Verstraeten, G., Poesen, J., Gillijns, K. and Govers, G., 2006. The use of riparian vegetated filter strips
 to reduce river sediment loads: an overestimated control measure? Hydrological Processes
 20, 4259-4267.
- Walker, L.R., Velázquez, E., Shiels, A.B., 2009. Applying lessons from ecological succession to the
 restoration of landslides. Plant and Soil 324, 157-168.
- Wang, J.H., Lu, X.G., Tian, J.H. 2008. Theory and application technology of riparian wetland. Wetland
 Science 6, 97-104.

- Wilson, C., Yagci, O., Rauch, H. & Stoesser, T., 2006. Application of the drag force approach to model
 the flow-interaction of natural vegetation. Journal of River Basin Management, Band Issue 2,
 1-10.
- Woolsey, S., Capelli, F., Gonser, T., Hoehn, E., Hostmann, M., Junker, B., Paetzold, A., Roulier, C.,
 Schweizer, S., Tiegs, S. D., Tockner, K., Weber, C., Peter, A., 2007. A strategy to assess river
 restoration success. Freshwater Biology 52, 752-769.
- Xiao, H., Huang, J., Ma, Q., Wan, J.,Li, L., Peng, Q., Rezaeimalek, S., 2017. Experimental study on the
 soil mixture to promote vegetation for slope protection and landslide prevention. Landslides
 14, 287-297.
- Yildiz, A., Askarinejad, A., Graf, F., Rickli, C., Springman, S.M., 2015. Effects of roots and mycorrhizal
 fungi on the stability of slopes, in: Winter, M.G., Smith, D.M., Eldred, P.J.L., Toll, D.G. (Eds.),
 XVI ECSMGE Geotechnical Engineering for Infrastructure and Development. ICE Publishing,
 pp. 1693–1698.

Yildiz, A., Rickli, C., Graf, F., Springman, S.M., 2018. An assessment of plant-induced suction and its
effects on the shear strength of root-permeated soils. Geoderma. submitted in revised form
08 Feb 2018 (GEODER_2017_1623).

Zeh, H. 2007. Soil Bioengineering Construction Type Manual.Verein für Ingenieurbiologie Vdf-EFIB,
441 p.

686

Table 1: Illustration of some ecosystem services provided by grass buffer strips, grassed waterways, and small flood retention ponds (bioengineering techniques) installed to reduce soil erosion rates by water and muddy floods in the loess belt of Belgium (based on Vandaele (2010) and various unpublished data).

691

Maintenance and improvement of soil ecosystem services: e.g. food, fiber, fuel and other biomass production; environmental interactions such as water filtering, carbon storage (e.g., due to the change of cropland to grassland) and nutrient cycling (e.g. N and P), transformation of substances, biological habitat for soil micro-organisms, fauna and gene pool; archive of our past (artefacts and indicators of environmental change).

Maintenance and improvement of hydrological systems: e.g. on site water infiltration, retention
 and storage, flow energy dissipation, off site flood control through reduced peak flow discharge and
 reduced sediment overloads.

* Increase of biodiversity: e.g. vegetation (such as properly managed species-rich grasses, herbs and multiple cover crops), providing food and habitats for spiders, insects (e.g. bees, ground beetles, lchneumonidae, ladybirds that are important for pollination and pest control), birds (such as skylarks, partridges and birds of prey), mammals as well as amphibians (in ecologically designed flood retention pools).

* Increase of ecological connectivity facilitating circulation of fauna in landscapes dominated by crops. Creation of ecological corridors for various kinds of animals, including potentially slow moving earth or water-bound species. Increase of genetic exchanges between distant populations of the same species.

- 709 * Adsorption of pollutants transported by runoff and wind (dust), hence cleaner surface water,
 710 groundwater and air
- 711 * **Reduction of negative off-site effects** such as sediment deposits on cropland, infrastructure and
- private property, psychological stress to inhabitants that were frequently affected by muddy floods.
- 713 * Enhanced quality of landscapes predominantly consisting of cropland through the installation of
- 714 green corridors (grass buffer strips, grassed waterways and cover crops) and blue measures (such as
- 715 flood retention ponds), which lead to an improved recreational attractiveness.

717 Figure 1. Examples of bioengineering structures and installations worldwide. A. Palissades (France) 718 (photo: F. Rey); B. Grass buffer strips (Belgium) (photo: J. Poesen); C. Green steel structure and log 719 branch/dormant cuttings after completion (France) (photo: K. Peklo); D. Prefabricated wooden 720 structure (Italy) (photo: F. Preti); E. Modified brush layers (Canada) (photo: D. Polster); F. Sowings 721 with straw mats and vegetated bench (Portugal) (photo C. Bifulco); G. Mixed check dam (Canary 722 Islands) (photo: G. Tardio); H. Planting with willow cuttings and coconut tissue (Switzerland) (photo: 723 G. De Cesare); I. Brushlayer, straw and wattle (Canada) (photo: P. Raymond); J. Vegetated crib wall 724 (Austria) (photo: H.P. Rauch); K. Hydroseeding (Scotland) (photo: S. Mickovski); L. River modeling 725 (Austria) (photo: F. Florineth).



Figure 2: Which bioengineering structures to use to restore this degraded stream while protecting the



739 railway against floods? (photo: F. Rey)

Figure 3: Questions arising from practitioners showing how research objectives should be defined throughout the life of a bioengineering intervention, and the consequences for the ecological trajectory of a bioengineering structure. Such an approach would lead to improved natural hazard control and ecological restoration of a degraded site (Burylo and Rey, unpublished)



747

- **Figure 4**: Biodiversity can have a negative effect on sediment trapping performance of vegetation
- barriers, as monospecific barriers involving one very performant species are more efficient than
- 750 plurispecific ones including more or less efficient species (photo: F. Rey)





- **Figure 5**: Design of bioengineering structures along riverbanks can be thought for allowing them to
- 755 provide both slope stabilization and ecological restoration (photo: K. Peklo)



- **Figure 6**: Rough rock ramp with integrated fishway, providing positive action for both riverbed
- 759 stabilization and fish movement (photo: K. Peklo)

