



HAL
open science

Sentinel-2 image capacities to predict common topsoil properties of temperate and Mediterranean agroecosystems

Emmanuelle Vaudour, Cécile Gomez, Youssef Fouad, Philippe Lagacherie

► **To cite this version:**

Emmanuelle Vaudour, Cécile Gomez, Youssef Fouad, Philippe Lagacherie. Sentinel-2 image capacities to predict common topsoil properties of temperate and Mediterranean agroecosystems. *Remote Sensing of Environment*, 2019, 223, pp.21-33. <10.1016/j.rse.2019.01.006>. <hal-01990153>

HAL Id: hal-01990153

<https://hal.science/hal-01990153v1>

Submitted on 21 Oct 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons CC BY-NC 4.0 - Attribution - Non-commercial use - International License

1 **SENTINEL-2 image capacities to predict common topsoil properties of temperate and**
2 **Mediterranean agroecosystems**

3

4 E. Vaudour^{1*}, C. Gomez², Y. Fouad³, P. Lagacherie⁴

5

6 ¹UMR ECOSYS, AgroParisTech, INRA, Université Paris-Saclay, 78850, Thiverval-Grignon,
7 France *Email: Emmanuelle.Vaudour@agroparistech.fr

8 ²IRD, UMR LISAH (INRA-IRD-SupAgro), F-34060 Montpellier, France

9 ³UMR SAS, Agrocampus Ouest, INRA, 35042, Rennes, France

10 ⁴INRA, UMR LISAH (INRA-IRD-SupAgro), F-34060 Montpellier, France

11

12

13 **Abstract**

14 To be fully operational for facilitating decisions made at any spatial level, models and
15 indicators of soil ecosystem functions require the use of precise spatially referenced soil
16 information as inputs. This study aimed at exploring the capacity for SENTINEL-2A (S2A)
17 multispectral satellite images to predict several topsoil properties in two contrasted
18 pedoclimatic environments: a temperate region marked by intensive annual crop cultivation
19 patterns and soils derived from loess or colluvium and/or marine limestone or chalk
20 (Versailles Plain, 221 km²); and a Mediterranean region marked by vineyard cultivation and
21 soils derived from lacustrine limestone, calcareous sandstones, colluvium, or alluvial deposits
22 (Peyne catchment, 48 km²). Prediction models of soil properties based on partial least squares
23 regressions (PLSR) were built from S2A spectra of 72 and 143 sampling locations across the
24 Versailles Plain and Payne catchment, respectively. Eight soil surface properties were
25 investigated in both regions: pH, cation exchange capacity (CEC), texture fractions (Clay,

26 Silt, Sand), Iron, Calcium Carbonate (CaCO_3) and Soil Organic Carbon (SOC) content.
27 Predictive abilities were studied according to the root mean square error of cross-validation
28 (RMSECV) tests, cross-validated coefficient of determination (R^2_{cv}) and ratio of performance
29 to deviation (RPD). Intermediate prediction performance outcomes (R^2_{cv} and RPD greater
30 than or equal to 0.5 and 1.4, respectively) were obtained for 4 topsoil properties found across
31 the Versailles Plain (SOC, pH, CaCO_3 and CEC), and near-intermediate performance
32 outcomes ($0.5 > R^2_{cv} > 0.39$, $1.4 > \text{RPD} > 1.3$) were yielded for 3 topsoil properties (Clay,
33 Iron, and CEC) found across the Peyne catchment and for 1 property (Clay) found across the
34 Versailles Plain. The study results show what can be expected from SENTINEL2 images in
35 terms of predictive capacities at the regional scale. The spatial structure of the estimated soil
36 properties for bare soils pixels is highlighted, promising further improvements made to spatial
37 prediction models for these properties based on the use of Digital Soil Mapping (DSM)
38 techniques.

39

40 Key words: soil properties, SENTINEL-2A, partial least squares regression, temperate soils,
41 Mediterranean soils, agroecosystems.

42

43 **1. Introduction**

44 Soils carry out a number of key environmental functions that are essential for human
45 subsistence (e.g., food, fiber and timber production, water storage and redistribution, pollutant
46 filtering and carbon storage). The understanding and modeling of physical, chemical and
47 biological processes underlying these different functions have been the subject of many
48 scientific works now that a number of models or indicators that represent these functions are
49 available (Sanchez et al., 2009; Adhikari and Hartemink, 2016; Baveye et al. 2016). To be
50 fully operational in facilitating decisions made at local, national and global levels, these

51 models and indicators require the use of precise spatially referenced soil information as
52 inputs.

53 Given the lack of precision and extension of existing soil databases, Digital Soil
54 Mapping (Mc Bratney et al., 2003) methods have been recognized as the most adequate and
55 operational response to these requirements (Arrouays et al., 2014, Minasny & McBratney,
56 2016). However, even though digital soil maps of soil properties are now being produced
57 across the globe (Arrouays et al., 2017) and for large areas—regions, countries and even
58 continents—at high resolutions (100 m), they still provide very uncertain estimations of soil
59 properties due to the limited availability of sites offering soil property measurements used for
60 calibrating DSM models (Vaysse et al., 2015).

61 Visible, Near-infrared and Short-wave Infrared (VNIR/SWIR) hyperspectral remote
62 sensing methods have been demonstrated to serve as promising means to significantly
63 increase the number of sites offering quantitative estimations of topsoil properties (e.g., Soil
64 Organic Carbon (SOC), pH, Cation Exchange Capacity (CEC), texture fractions (clay, silt,
65 sand), Iron and Calcium Carbonate (CaCO_3)) (e.g., Selige et al., 2008; Stevens et al., 2010;
66 Gomez et al., 2012; Vaudour et al., 2016). For bare soil surfaces, accurate local estimates
67 ($R^2_{cv} > 0.7$) have been obtained for soil properties i) related to a chemical species that impacts
68 soil surface reflectance values through absorption bands (e.g., OH⁻ ion for clay) (Ben-Dor et
69 al., 2002) or ii) highly correlated with the latter (e.g., CEC with clay content) (Ben-Dor et al.,
70 2002), iii) showing a sufficient level of variability across study regions (Gomez et al., 2012a
71 & b). For instance, Selige et al. (2006) mapped SOC content, a soil property having specific
72 spectral behaviour and high spatial variability over their study area. In addition, they also
73 mapped Total Nitrogen, a soil property deprived of specific spectral features but, due to an
74 associated chemical or physical structure, correlated to a soil property having such features
75 (being SOC). DSM approaches are designed to employ this new source of soil data to allow

76 for the exhaustive mapping of surface (Walker et al., 2017) and subsurface (Lagacherie et al.,
77 2013) properties.

78 Although hyperspectral VNIR/SWIR imagery has been proven to serve as a valuable
79 tool for mapping key soil surface properties, it cannot be applied for large surface mapping or
80 for temporal monitoring because it is expensive to apply and because hyperspectral
81 VNIR/SWIR imaging data are not widely available. Indeed, only one hyperspectral
82 VNIR/SWIR satellite sensor is currently operational. The HYPERION sensor offers a spatial
83 resolution of 30 m, a spectral resolution of 10 nm and a swath of 7.5 km (Folkman et al.,
84 2001). Other hyperspectral VNIR/SWIR imaging sensors are airborne sensors (e.g., the
85 Hymap, AISA-DUAL and HySpex sensors) with spectral resolutions of 5 to 10 nm, spatial
86 resolutions of approximately 4-5 m (depending to the flight altitude) and flight prints of
87 generally less than 400 km² (depending to the studied case). Furthermore, few studies have
88 obtained successful results in terms of mapping soil properties from VNIR-SWIR
89 multispectral satellite imagery that allow one to capture a larger area from a single nearly
90 instantaneous glance (e.g., Vaudour et al., 2013; Shabou et al., 2015). This can be attributed
91 to i) the low revisit frequency of sensors, which compromises the acquisition of images under
92 clear conditions, while time ranges of bare soil availability, and particularly for annual crop
93 systems, are short and ii) the coarse spectral and spatial resolutions of most historical
94 multispectral sensors, which limit prediction performance outcomes (Gomez et al, 2015;
95 Adeline et al., 2017).

96 The recent release of multispectral satellite SENTINEL-2A data by the European
97 Space Agency last November 2015 is revolutionary in developing images of large areas with
98 a 290 km-swath width at high revisit frequencies (planned to be near-daily when the full
99 satellite constellation is in orbit). The multispectral SENTINEL-2A satellite provides 13

100 spectral bands at spectral resolutions of between 20 and 180 nm and at spatial resolutions of
101 10 to 60 m (§2.3).

102 The objective of this study is to assess the potential for SENTINEL-2A data to
103 produce estimations of topsoil properties for large sets of sites. The study focuses on two
104 French regions—the Versailles Plain and the Peyne Valley—that represent very different
105 agroecosystems and that have both been previously used as test areas for topsoil mapping
106 from VNIR/SWIR hyperspectral data (Vaudour et al, 2016 and Gomez et al, 2012,
107 respectively).

108

109

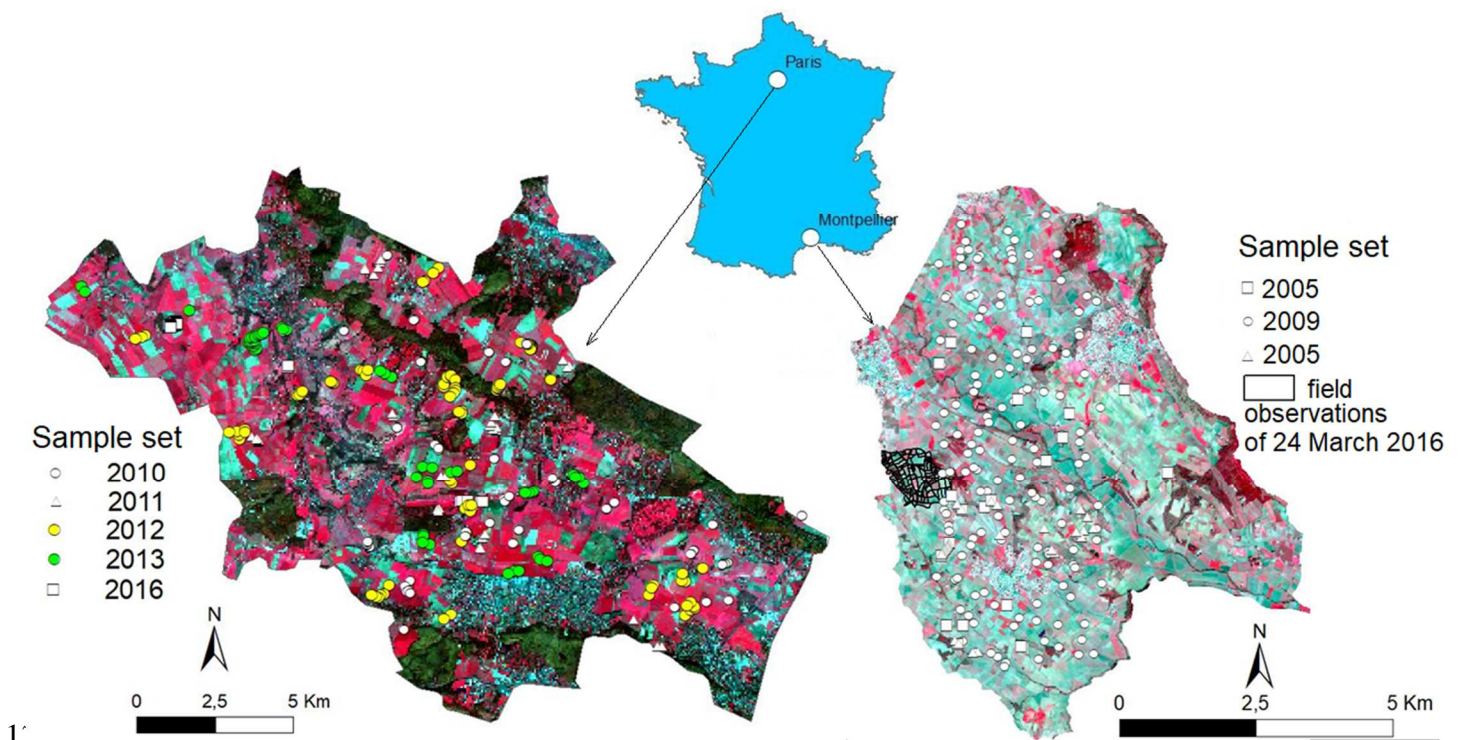
110 **2. Materials**

111 *2.1 Study areas*

112 2.1.1 Versailles Plain

113 The Versailles Plain is an agricultural peri-urban region of 221 km² located west of Paris
114 (North of France) that is characterized by intensive annual crop cultivation across an area of
115 ~99 km² (Fig. 1, left). The main crop rotations in the area involve winter wheat, winter
116 rapeseed, winter and spring barley and maize on occasion (Vaudour et al., 2015). Cultivation
117 practices are mainly conventional, with early winter plowing applied at least 1 year out of
118 every 3. The Versailles Plain is composed of two plateaus levels: one at a 120-m elevation in
119 the center (lower limestone plateau) and another at a 170 m-elevation along the northern and
120 southern edges of the area (upper millstone clay plateau). Both are oriented in the NW-SE
121 direction toward the Palace of Versailles. Quaternary loessic deposits and loessic colluvium
122 leave a mark on all landforms in the area and particularly on local plateaus, which have
123 evolved into haplic or glossic luvisols according to the FAO classification (World Reference
124 Base (WRB) (Crahet, 1992; Vaudour et al., 2013, 2014). The largely forested upper plateau

125 flanks are characterized by Fontainebleau acid sands that give rise to arenic cambisols, while
 126 the lower plateau flanks, which are largely cultivated, are composed of calcaric cambisols
 127 derived from limestone, colluvial materials and, along the lower slopes, chalk. Throughout the
 128 lower slopes and valleys, stagnic colluvic cambisols are derived from marls or alluvio-
 129 colluvial materials.
 130



131
 132 Figure 1. Locations of soil samples from the Versailles Plain (left, infrared colored S2A
 133 image from 12 March 2016) and the lower Peyne Valley (right, infrared colored S2A image
 134 from 19 March 2016), ©ESA 2016

135
 136 2.1.2 La Peyne Valley

137 The lower valley of the Peyne River is a viticultural region of 48 km² located west of
 138 Montpellier (South of France) (Fig. 1, right) and characterized by Plio-Quaternary alluvial
 139 terraces and fan terraces, the deposits of which have undergone fersiallitization and which

140 have evolved into red Mediterranean soils or chromic luvisols (Coulouma, 2008). The valley
141 base has been composed of marine Miocene sediments (marl, sandy loam, and calcareous
142 sandstone) and marine Pliocene marl refillings since the Messinian salinity crisis. Gently
143 undulating forms of hills and colluvial fans characterize landscape units in the area and give
144 rise to several types of soils rich in calcium carbonate with a loamy sand texture and
145 increasing clay levels toward the bases of the hillslopes: calcaric leptosols, calcaric regosols,
146 and calcisols. A cuesta that characterizes the landscape unit has formed on lacustrine
147 limestone of the Upper Miocene, where calcisols and calcaric regosols have developed with
148 irregular bedrock depths creating variations in stoniness and calcium carbonate content levels.
149 Alluvial terraces and alluvio-colluvial fans originating from Plio-Quaternary clays and
150 alluvial deposits correspond to deep stony decarbonated chromic and/or colluvial luvisols,
151 fluvisols, and some endogleyic calcisols. While early Quaternary terraces of the Payne River
152 have produced stony and clayey chromic luvisols, newer terraces have balanced textured
153 fluvisols with carbonate content.

154

155 *2.2 Soil samples and ground observations*

156 *2.2.1 Soil sampling*

157 This study gathered two soil data sets that were collected for the purpose of earlier studies (for
158 Versailles, Vaudour et al. 2013, 2014a, 2016; for La Payne, Gomez et al, 2012). Despite some
159 differences in chemical analysis, it was found useful to gather them in view of enlarging the
160 range of the tested pedological situations. A total of 72 and 143 topsoil samples were
161 collected from the Versailles Plain and Payne Valley, respectively (Fig. 1). For the Versailles
162 Plain, 69 topsoil samples were collected from 2010 to 2013, and 13 topsoil samples were
163 collected in 2016. All of the samples were composed of roughly 10 sub-samples collected to a
164 depth of 8 cm from random locations within a 2.7×2.7 m square area centered at the

165 sampling plot as recorded at its center using a Trimble Pathfinder®Power DGPS of 50 cm
 166 precision (Vaudour et al., 2014). From the Peyne Valley, 39 topsoil samples were collected in
 167 2005, and 104 topsoil samples were collected in 2009. All of the samples were composed of
 168 five sub-samples collected to a depth of 5 cm from random locations within a 10 × 10 m
 169 square centered on the geographical position of the sampling plot as recorded by a Garmin
 170 GPS instrument.

171

172

173 2.2.2 Physico-chemical analysis

174 Topsoil samples collected from both study areas were air-dried and then crushed and sieved to
 175 2 mm. Soil property determinations were performed through a classical physico-chemical
 176 analysis (Baize and Jabiol, 1995). In total, 8 common soil properties were considered for
 177 spectral modeling (Table 1): content of 3 particle sizes or granulometric fractions of 2 mm-
 178 sieved fine earth (total silt, 2 - 50 µm; total sand, 50 µm - 2 mm and clay, < 2 µm); calcium
 179 carbonate (CaCO₃) content; free iron content (Mehra-Jackson); soil organic carbon (SOC)
 180 content; pH; and cation exchange capacity (CEC).

181 Granulometric fractions (clay, silt, and sand) and CaCO₃ content levels were measured
 182 from 143 soils samples collected from the Peyne Valley (Table 1). The three other soil
 183 properties (iron, CEC and pH) were measured from 104 soil samples collected from the Peyne
 184 Valley. For the Versailles Plain, all soil properties except for CEC (39 samples) were
 185 measured (Table 1).

186

187 Table 1. Soil datasets used for models and statistics on soil properties

Soil property	Description	Unit	Versailles Plain					Peyne Valley						
			Sample size	Min	Q1	Median	Q3	Max	Sample size	Min	Q1	Median	Q3	Max
SOC	soil organic C	g.kg ⁻¹	72	7.0	12.5	15.9	20.1	31.9	104	4.0	7.0	8.9	11.4	21.8

CaCO ₃	content total CaCO ₃	g.kg ⁻¹	72	0	0	27.1	67.2	530	143	0	22.2	136.0	251.0	473
Clay	content granulometric fraction <2 µm	g.kg ⁻¹	72	132	192.5	249.5	281.0	391	143	67	185.0	231.0	290.5	452
Silt	granulometric fraction 2-50 µm	g.kg ⁻¹	72	228	474.8	542.0	603.0	774	143	141	252.0	322.0	384.5	655
Sand	granulometric fraction 50 µm – 2 mm	g.kg ⁻¹	72	17	87.8	157.5	209.0	512	143	106	364.5	433.0	516.0	737
Iron	« free iron » content	g/100 g	72	0.51	0.75	0.87	0.97	1.29	104	0.03	0.75	1.07	1.58	2.92
pH	water-pH	-	72	5.61	6.91	7.99	8.18	8.34	104	5.37	8.3	8.46	8.54	8.70
CEC	cation exchange capacity	cmol.kg ⁻¹	39	6.1	11.0	17.1	22.1	28.3	104	6.1	8.3	10.7	12.9	19.1

188

189

190 2.2.3 Ground observations

191 For the Versailles Plain, soil surface conditions were observed on 15 and 17 March 2016
192 across 6 agricultural fields and for other fields planned for spectral measurements. For the
193 Peyne Valley, vineyard soil surface conditions were observed on 24 March 2016 across the
194 0.9 km² sub-catchment of Roujan, which includes 75 viticultural fields. These observations
195 enabled us to define the threshold of the vegetation index for retrieving bare soil pixels from
196 S2A images. At the time that the images were acquired, the budburst stage had just begun for
197 vines in the Peyne Valley, and hence vine vegetation had not yet developed. For the
198 Versailles Plain, spring barley sowing had just started and maize and pea sowing had not
199 started, hence allowing for the observation of bare soils used for spring crops. For both study
200 areas, winter crops and grasslands were photosynthetically active and covering, and thus
201 easily identifiable and removable by means of NDVI.

202

203 2.3 Remotely sensed images and geographic data

204 Two SENTINEL-2A (S2A) images were acquired under nearly clear conditions on 12 and 19
205 March 2016 from the Versailles Plain and Peyne Valley, respectively (Table 2). The

206 multispectral instrument aboard the S2A satellite has 13 spectral bands, including 4 bands of a
 207 10 m resolution and 6 bands of a 20 m resolution (Table 3) retained for spectral modeling.
 208 Both images were atmospherically corrected using the ATCOR2/3 radiative transfer model
 209 (Richter and Schläpfer, 2016).

210 Both study sites are characterized by urban activity and by the presence of bare soil,
 211 water and vegetation. To mask no-soil pixels from the S2 data, the following approach was
 212 applied. First, urban and water areas were masked using Land Parcel Registry maps. In turn,
 213 only cropland and vineyards were maintained across the study areas. Second, pixels with
 214 normalized difference vegetation index (NDVI) values exceeding an expert-calibrated
 215 threshold were masked. A value of 0.35 was determined for both sites after considering field
 216 observations (section 2.2.3). The NDVI was retrieved using bands of 842 nm and 665 nm.
 217 Finally, bare soils of the Versailles Plain represent 8.05% of our study area and 44,979 S2A
 218 pixels, while bare soils of the Payne catchment represent 38.6% of our study area and 46,971
 219 S2A pixels.

220

221 Table 2. Main characteristics of the studied scenes

Imaging date	Sensor	Output resolution (m)	Time of acquisition (U.T GMT)	Viewing incidence angle (°)	Sun azimuth (°)	Sun elevation (°)
12 March 2016	S2A	20	10:50:37	<5.1	160.5	36.1
19 March 2016	S2A	20	10:40:32	<3.3	157.6	43.7

222

223 Table 3. Characteristics of the Multi-Spectral Instrument aboard the SENTINEL2 satellite.

224 Spectral S2A bands used are shown in bold.

Spectral band	Spatial resolution (m)	Central wavelength (nm)	Band width (nm)
b1	60	443	20
b2	10	490	65
b3	10	560	35
b4	10	665	30
b5	20	705	15
b6	20	740	15
b7	20	783	20
b8	10	842	115
b8A	20	865	20
b9	60	945	20

b10	60	1380	30
b11	20	1610	90
b12	20	2190	180

225

226

227

228

229

230

231

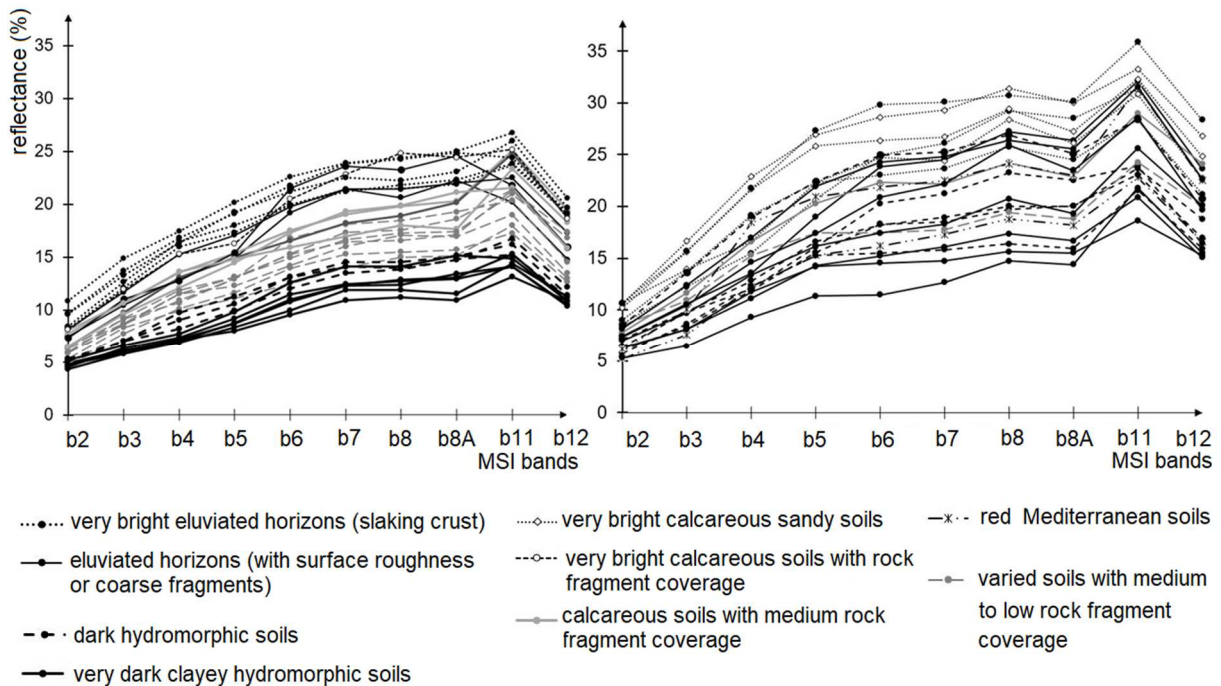
232

233

234

235

A selection of various reflectance spectra comprising the darkest and the brightest soils is shown in Fig. 2. The brightest soils comprise: (i) calcic or calcareous sandy soils originating from calcareous sandstone (La Peyne); (ii) silt-loam luvisols (Versailles) or eluviated horizons of the chromic luvisols (La Peyne), developing a slaking crust ; (iii) rendzic or calcareous cambisols with very high topsoil calcium carbonate content or outcropping underlying chalk (Versailles). The darkest soils are clayey and alluvial and, for Versailles, can have high organic carbon content (>20 g kg⁻¹) and/or high roughness dating from late winter ploughing. Intermediate soil spectra comprise soils with either moderate rock fragment content (10–25%), or varied textures and intermediate to low roughness (both regions), and also red Mediterranean soils (La Peyne).



236

237

238

239

240

Figure 2. Selection of reflectance image spectra covering the whole dataset reflectance range and the main soil types

3. Methods

241 *3.1 Models for soil property prediction*

242 The partial least squares regression (PLSR) method was used to construct prediction models
243 based on bare soil samples drawn from each study region. The PLSR method relates two data
244 matrices, the matrix of explanatory variables X (here the soil spectra) and the matrix of
245 dependent variables Y (here the soil properties), through a linear multivariate model, but it
246 extends beyond traditional regression methods in that it also models the structures of X and Y
247 (Geladi and Kowalski, 1986; Wold et al., 2001). For Y in particular, PLSR extracts
248 orthogonal or latent predictor variables accounting for variation in the dependent variable.
249 The PLSR approach enables one to analyze data based on several noisy collinear features
250 (Wold et al., 2001), and it has been used widely for hyperspectral remote sensing (e.g., Selige
251 et al., 2008; Stevens et al., 2010; Gomez et al., 2012; Vaudour et al., 2016).

252 Prior to PLSR modeling, in order to maximize the fitting of target values against spectra, a
253 series of transforms were tested for either spectra or target values. Reflectance values of the
254 whole spectra were centered. For the Versailles Plain, Clay, SOC and CaCO₃ content levels
255 and CEC were log-transformed, while for both areas, the square root of iron content was used.
256 For La Peyne, log-transform of target values was not retained, as it did not bring any
257 improvement of the fitting. Outliers were detected for sets of image spectra or for sets of
258 measured soil properties. This was based on the 5% threshold of the standardized
259 Mahalanobis distance (Mark and Tunnel, 1985) and on the multivariate outlier detection
260 approach elaborated by Filzmoser et al. (2005). Outliers selected through both approaches
261 were retained as outliers.

262 For each target soil property Y, a PLSR model was constructed from preprocessed image
263 spectra with 10 selected spectral bands (Table 3). The optimal number of latent variables was
264 determined from the prediction residual error sum of squares (PRESS). Due to the limited
265 number of bare soil samples available, and because this study was more focused on

266 comparing performance levels between regions than on carrying out external validation
267 assessments, a leave-one-out cross-validation procedure was applied (Wold, 1978; Gomez et
268 al., 2012). The quality of model fit was evaluated from the root mean squared error of cross-
269 validation (RMSECV), from the coefficient of determination of cross-validation (R_{cv}^2) and
270 from the residual prediction deviation (RPD), i.e., the ratio between the standard deviation of
271 the calibration dataset to the RMSECV. According to Chang et al. (2001) and Viscarra-Rossel
272 et al. (2006), values of 1.4 to 1.8 denote models exhibiting moderate levels of predictive
273 capacity; values of between 1 and 1.4 denote models exhibiting poor levels of predictive
274 capacity, and values of <1 denote very poor models that should not be used.

275 PLSR models were used through R version 3.2.1 (R Development Core Team, 2015)
276 employing the “pls” package (Mevik and Wehrens, 2007).

277

278 *3.2 Variograms of predicted soil properties*

279 For the best predicted soil properties, PLSR models were applied to pixels of bare agricultural
280 soil. The spatial structure of the predicted pixels was examined and compared with that of
281 measured properties for the same locations. Spatial structures can be described from a
282 variogram that relates the spatial dependence or semi-variance between points at a distance of
283 h , i.e., the average variance between any pair of sampling points (Goovaerts, 1997):

$$284 \gamma(h) = \frac{1}{2n} \sum_{i=1}^n \{p(x_i) - p(x_i + h)\}^2 \quad (1)$$

285 where $\gamma(h)$ is the average semi-variance of the soil property, n is the number of pairs of
286 sampling points, p is the value of the property P , x is the coordinate of the point, and h is the
287 distance between pairs or the lag value.

288 For each best predicted soil property, two empirical variograms were built from
289 equation (1): the first was built from 143 or 104 values of the soil samples measured by
290 physico-chemical analysis; the second was built from predicted values for the same locations.

291 Simple omnidirectional variograms were calculated, and following Gomez et al. (2012), the
292 spherical or exponential model used was fitted through a weighted least squares procedure
293 (Pebesma, 2004) using the R statistical software program and the Gstats package. The
294 universality condition (no drift) was assumed.

295 For the Versailles Plain, it was not possible to obtain experimental variograms suitable
296 for fitting a model due to the limited number of sites involved and the clustered sampling
297 scheme employed. Variograms were therefore only examined for the Peyne Valley. To
298 examine the robustness of spatial structures at the scale of the entire study region, variograms
299 were calculated from all bare soil pixels.

300

301 **4. Results**

302 *4.1 Description of soils samples*

303 Basic statistics on all of the soil properties measured are shown in Table 2. For the Peyne
304 Valley, the range of clay contents present is higher than that of the Versailles Plain, with the
305 highest value recorded as 452 g.kg⁻¹, though the 1st quartile and median values (185.0 and
306 231.0 g.kg⁻¹) are slightly lower than those for the Versailles Plain (192.5 and 249.5 g.kg⁻¹),
307 while the 3rd quartile is slightly higher (290.5 g.kg⁻¹) than that for the Versailles Plain (281
308 g.kg⁻¹). SOC, silt and CEC levels are measured within a narrower and lower range for the
309 Peyne Valley and with a much lower median (SOC, 8.92 g.kg⁻¹; silt content, 322 g.kg⁻¹; CEC,
310 10.65 cmol+.kg⁻¹) than that for the Versailles Plain (SOC, 15.9 g.kg⁻¹; silt content, 542 g.kg⁻¹;
311 CEC, 17.10 cmol+.kg⁻¹). Conversely, sand and calcium carbonate content levels are higher in
312 the Peyne Valley, with a higher median (sand content, 433 g.kg⁻¹; calcium carbonate content,
313 136 g.kg⁻¹) than that for the Versailles Plain (sand content, 157.5 g.kg⁻¹; calcium carbonate
314 content, 27 g.kg⁻¹). Consequently, pH values are higher and less variable for the Peyne
315 Valley, while soils of the Versailles Plain vary more in terms of pH levels. Red Mediterranean

316 soils resulting from geochemical weathering contain more free iron, and hence their presence
 317 has induced a higher range of Fe content levels, with the 3rd quartile reaching a value of 1.58
 318 g/100 g compared to a value of 0.97 g/100 g found for the Versailles Plain.

319 Moreover, coarse fragment content levels in topsoil can reach percentages of as high
 320 as 20% and 30% for the 3rd quartile of the Versailles and Payne sets, respectively.

321

322

323

324

325 Table 4. Pearson correlation table of common soil properties for the Versailles Plain

Variables	Clay	Silt	Sand	CaCO ₃	SOC	Fe	pH	CEC
Clay	1.00							
Silt	-0.24	1.00						
Sand	-0.34	-0.27	1.00					
CaCO ₃	-0.02	-0.70	-0.33	1.00				
SOC	0.53	-0.38	-0.44	0.46	1.00			
Fe	0.48	0.24	0.03	-0.52	0.00	1.00		
pH	0.46	-0.43	-0.50	0.60	0.66	-0.25	1.00	
CEC	0.77	-0.32	-0.52	0.33	0.82	0.12	0.77	1.00

326 *Values in bold are different from 0 at a significance level of alpha=0.05*

327

328 Table 5. Pearson correlation table of common soil properties of the Payne Valley

Variables	Clay	Silt	Sand	CaCO ₃	SOC	Fe	pH	CEC
Clay	1.00							
Silt	-0.14	1.00						
Sand	-0.54	-0.76	1.00					
CaCO ₃	-0.08	0.53	-0.40	1.00				
SOC	-0.19	0.08	0.06	0.11	1.00			
Fe	0.42	-0.26	-0.06	-0.50	-0.11	1.00		
pH	0.02	0.42	-0.37	0.52	0.03	-0.42	1.00	
CEC	0.89	-0.16	-0.45	0.00	-0.02	0.16	0.14	1.00

329 *Values in bold are different from 0 at a significance level of alpha=0.05*

330

331 For the Versailles Plain, the highest correlations ($|R| > 0.7$) were observed between CEC and
332 SOC content levels, between CEC and clay levels, between CEC and pH levels and between
333 CaCO_3 and silt levels (Table 4). Modest correlations ($0.7 > |R| > 0.5$) were observed between
334 pH and SOC levels, between pH and CaCO_3 levels, between pH and sand levels, between Iron
335 and CaCO_3 levels and between SOC and clay levels (Table 4). For the Payne catchment, the
336 highest correlations ($|R| > 0.7$) were observed between CEC and clay levels and between sand
337 and silt levels (Table 5). Modest correlations ($0.7 > |R| > 0.5$) were observed between pH and
338 CaCO_3 levels, between iron and CaCO_3 levels, between sand and clay levels and between silt
339 and CaCO_3 levels (Table 5).

340

341 *4.2 Performance of the prediction models*

342 A PLSR model was built from S2A spectra and from each observed soil property associated
343 with the available soil samples. No spectral outliers were removed from the calibration
344 database, regardless of the study area concerned, and the number of latent variables was
345 determined following the rule of the first local minimum of the RMSECV, which varied
346 between 2 and 9 (Table 6). The predictive capacities of the PLSR models ranged from low to
347 medium to high, with RPD values of between 1.0 and 2.0 and with R^2_{cv} of between 0.02 and
348 0.75 (Table 6). Three groups of variables can be identified from the cross-validation
349 performance of the PLSR models (Table 6, Fig. 3): i) soil property models yielding $R^2_{cv} \geq 0.5$
350 and $\text{RPD} \geq 1.4$ (rounded values) or models with intermediate to high levels of predictability;
351 ii) soil property models with R^2_{cv} and RPD values of 0.4 and 1.3, respectively (rounded
352 values) or models with poor predictive capacities just below thresholds commonly accepted in
353 the literature or 'near intermediate' models; and iii) soil property models with $R^2_{cv} < 0.4$ and
354 RPD values of roughly 1 or models exhibiting poor or very poor levels of predictive capacity.

355 In overall, the Versailles Plain includes all 4 properties of group i (SOC, pH, CEC and
 356 CaCO₃) and one property of group ii (Clay), whereas the Peyne region includes 3 properties
 357 of group ii (Clay, Iron and CEC), with the remaining properties (silt and sand content for both
 358 regions and SOC, pH, and CaCO₃ content for the Peyne site) are categorized as group iii.

359

360

361

362

363

364

365 Table 6. Cross-validation performance statistics derived from the PLSR algorithm for soil
 366 property prediction (NL_v, number of latent variables; RMSE_{cv} expressed in original units).

367 R²_{cv} and RPD rounded values of greater than or equal to 0.5 and 1.4, respectively, are shown
 368 in bold (group *i*). RPD rounded values of greater than or equal to 1.3 are shown in italics
 369 (group *ii*).

Soil property	Versailles Plain				Peyne Valley			
	R ² _{cv}	RMSE _{cv}	RPD	NL _v	R ² _{cv}	RMSE _{cv}	RPD	NL _v
SOC	0.56	1.23	1.51	4	0.02	3.71	1.00	4
CaCO ₃	0.48	20.3	1.39	4	0.15	122.2	1.08	8
Clay	<i>0.39</i>	1.23	<i>1.30</i>	6	<i>0.42</i>	56.4	<i>1.31</i>	5
Silt	0.14	103.0	1.09	3	0.11	91.9	1.06	9
Sand	0.22	81.7	1.14	4	0.03	113.0	1.02	7
Iron content	0.05	0.09	1.02	2	<i>0.45</i>	0.04	<i>1.34</i>	6
pH	0.51	0.51	1.43	4	0.08	0.64	1.03	4
CEC	0.75	1.23	2.00	6	<i>0.41</i>	2.41	<i>1.31</i>	6

370

371

372

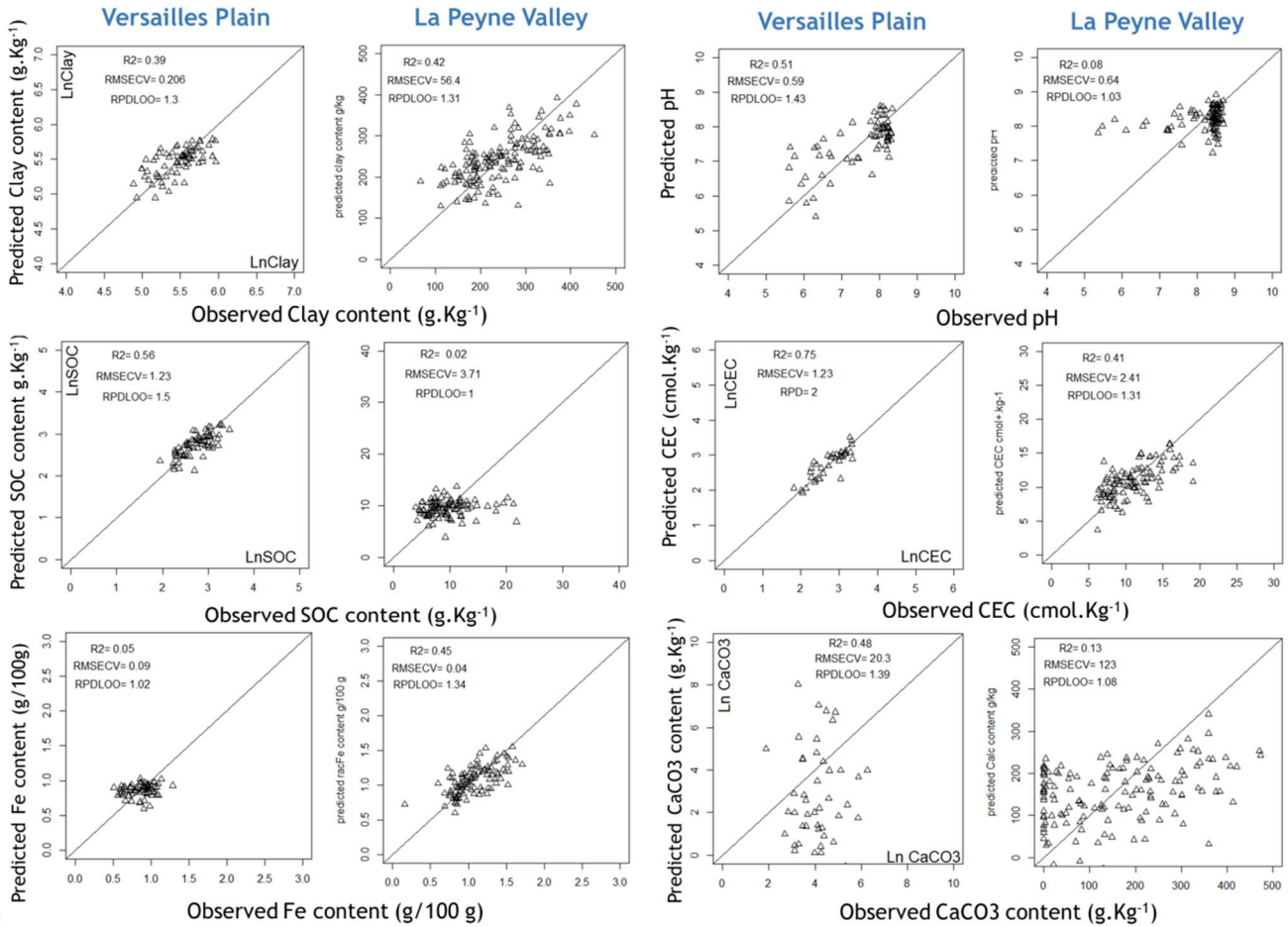
373

374

375

376

377



378

379 Figure 3. Predicted vs. observed soil properties for the Versailles Plain (left) and Peyne
 380 Valley (right) for 6 soil properties best predicted for at least one region

381

382 A model property that exhibits intermediate or near intermediate predictive capacity for one
 383 region can exhibit poor capacities for another. This is the case for pH, SOC, iron and CaCO₃
 384 content levels. A model property that exhibits high predictive capacity for one region can
 385 exhibit near intermediate capacities for another: this is the case for CEC levels.

386

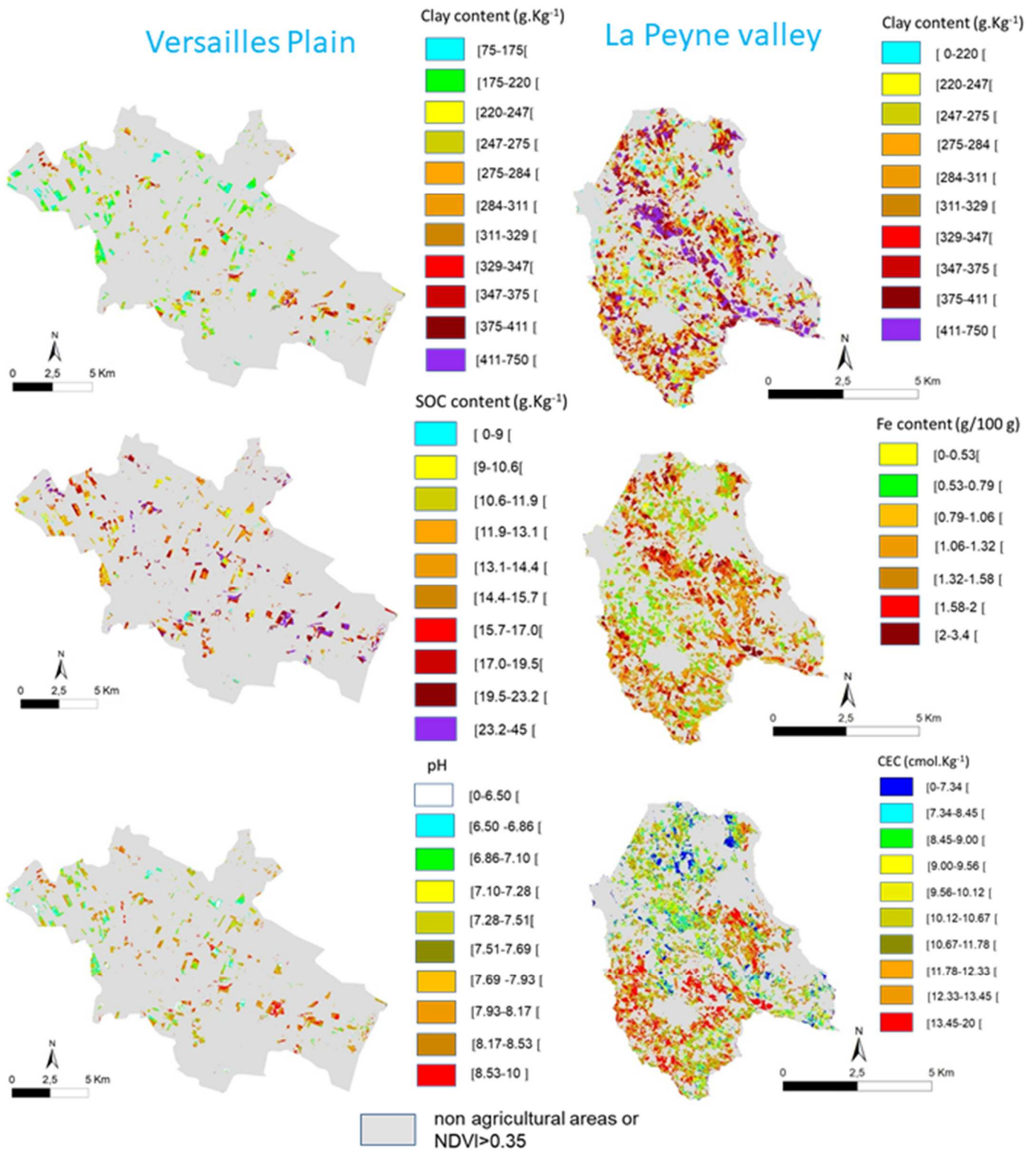
387 4.3 Soil property mapping from SENTINEL2 data

388 While retaining PLSR models for groups i and ii, equations were applied to bare cropland and
 389 vineyard pixels (Fig. 4).

390 To relate the output maps to soil data and available soil maps and to facilitate the
391 visual analysis of results for the other maps, and particularly for the Versailles Plain, for
392 which the percentage of mapped pixels is very low, values were simplified to decile classes of
393 predicted bare soil pixels for each study zone (Fig. 4).

394 The spatial patterns of clay correspond with known soil patterns for both regions. In
395 the Peyne Valley, spatial patterns clearly show chromic luvisols developed along Plio-
396 Quaternary alluvial terraces and characterized by high clay, iron and CEC levels. This result
397 contrasts well with areas with Miocene marine deposits of calcisols and calcareous leptosols
398 with less clay, iron and CEC. However, very high clay and iron content levels predicted for
399 the minor riverbed of the Peyne River (the violet band in the clay map shown in Fig. 4) are
400 not in accordance with our soil knowledge, likely revealing a perturbation related to soil
401 moisture levels.

402 In the Versailles Plain, the highest predicted clay content levels match our
403 observations of stagnic colluvic cambisols in lower slopes and valleys, whereas the lowest
404 predicted clay content levels as expected were found along the plateaus with luvisols of loess
405 origin. The highest SOC content levels were predicted both for stagnic colluvic cambisols in
406 the valleys and for calcareous cambisols across the slopes, while low SOC content levels were
407 predicted for luvisols originating from loess sources, in accordance with previous studies of
408 the same region (Vaudour et al., 2013, 2016; Zaouche et al, 2017).



41

410 Figure 4. Maps of soil properties of the Versailles Plain (left) and Peyne Valley (right)

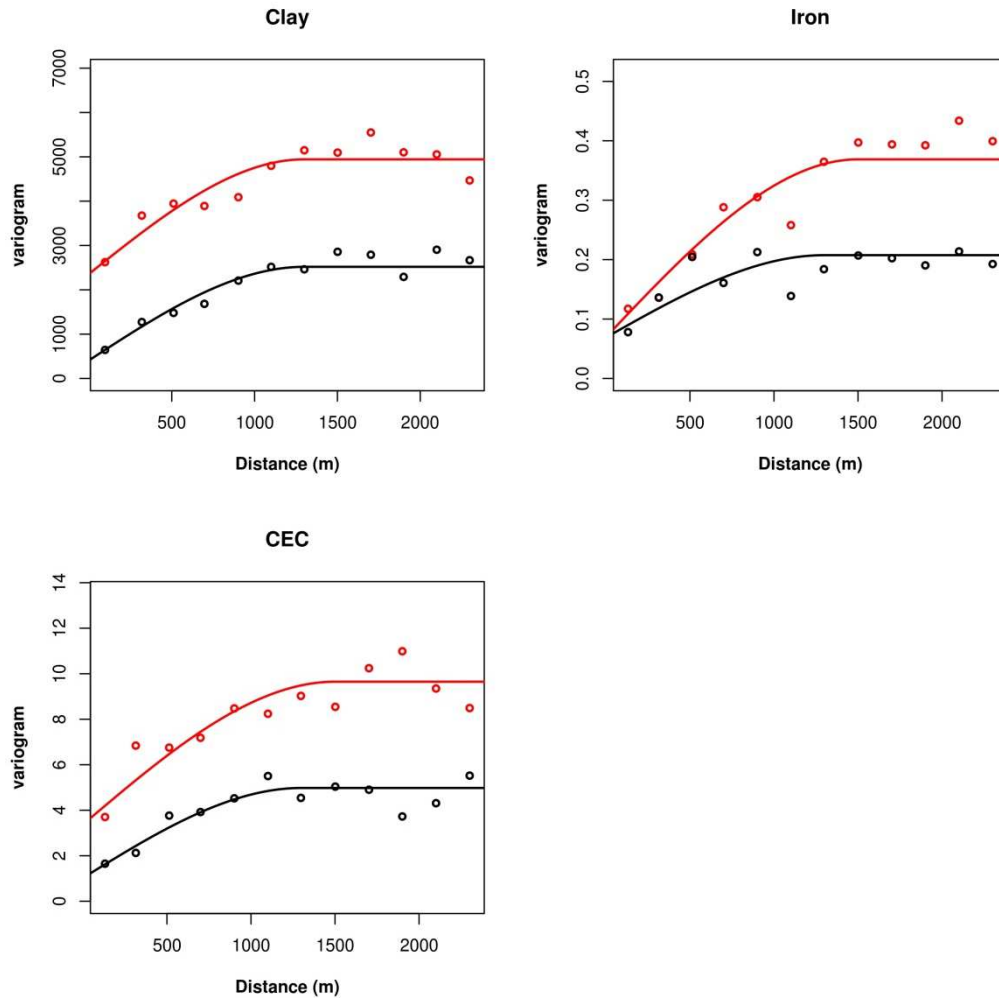
411

412

413 *4.4 S2A capacities to retrieve spatial structures*

414 *4.4.1 Comparisons between empirical variograms derived from measured and predicted*
415 *values*

416 Following from the map examinations, comparisons drawn between variograms of predicted
417 values of the three best predicted soil properties for the Payne Valley (clay, Fe, and CEC
418 content, Fig. 5, black curves) and those of measured values for the same locations (Fig. 5 red
419 curves) confirm that the soil patterns are captured by the predictions fairly well. Variograms
420 for clay and CEC are very similar in shape and only show a decrease in the ‘nugget’ value (at
421 the origin) of the predicted variogram. These differences in variograms have been observed
422 through tests of soil predictions from hyperspectral data (Gomez et al, 2012). The predicted
423 iron variogram has the same nugget and range as the measured one, with only a difference in
424 ‘sill’ values (the slope of the variogram). No clear explanation can be given for this difference
425 in behavior.



426

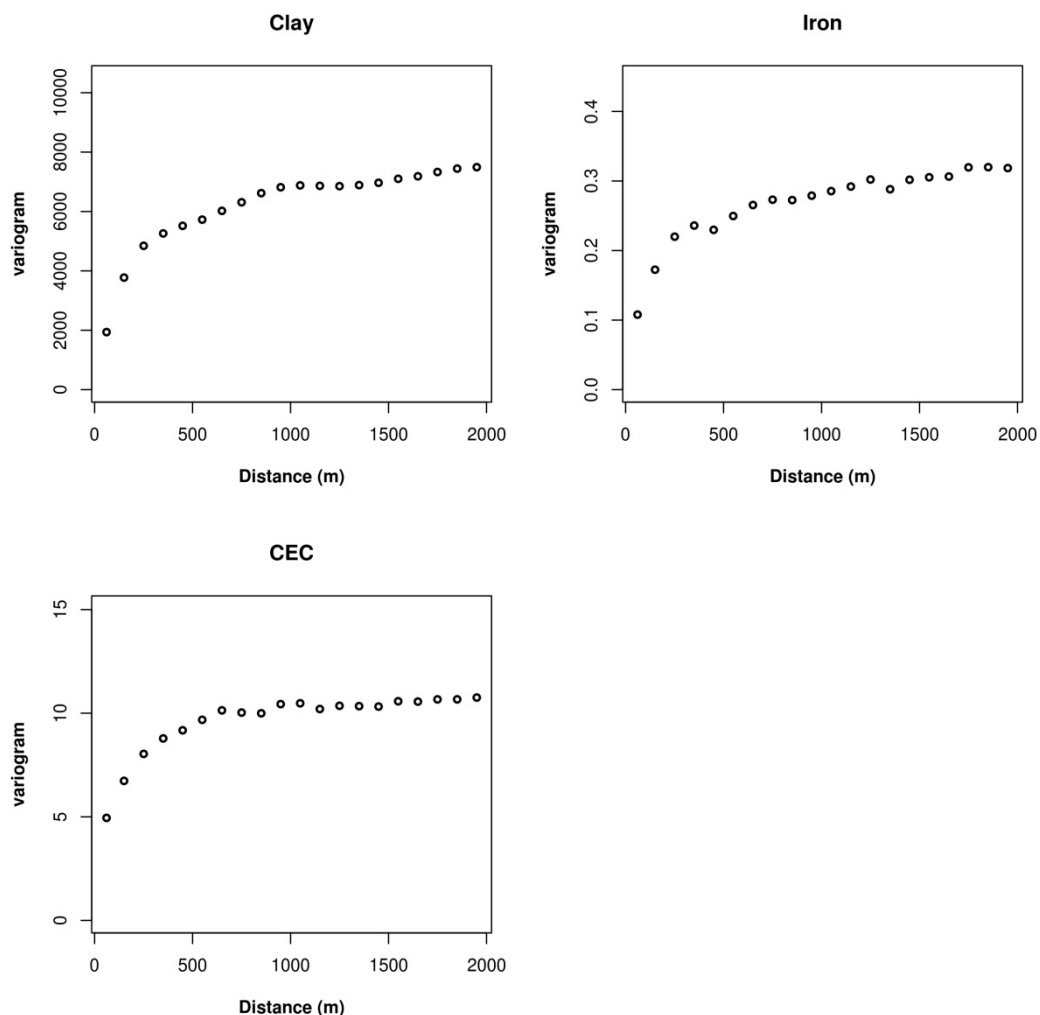
427 Figure 5. Empirical variograms calculated for clay, iron and CEC for the Payne Valley from
 428 physico-chemical values of the collected soil samples (red points) and from predicted values
 429 based on S2A spectra (black points) and fitted theoretical variograms (lines). Lag = 200 m.
 430 Distance max: 2000 m.

431

432 4.4.2 Spatial structures retrieved from the set of predicted pixels

433 Empirical variograms were computed from the 46,971 pixels of bare soil in the Payne
 434 catchment (Fig. 6). They exhibit different spatial structures from those revealed by
 435 variograms built from the measured sites only (Fig. 5). From an increase in the number of
 436 sites tested, short-scale variations are found across the clay and iron variograms as a result of
 437 a first slope break at 250 m that is not shown in the previous variograms (Fig. 6). Another

438 slope break at approximately 1,000 m in the variograms shown in Fig. 6 is still visible but is
439 less pronounced. This more complex spatial structure reveals a combination of soil variations
440 acting at different scales in accordance with our understanding. The broad-scale variability
441 found could be related to the spatial distribution of parent materials of soils with dissimilar
442 clay and iron content levels. Short-range variations may result from erosion-redeposition
443 processes occurring along slopes that involve neighboring parent materials with dissimilar
444 iron and clay content levels.
445



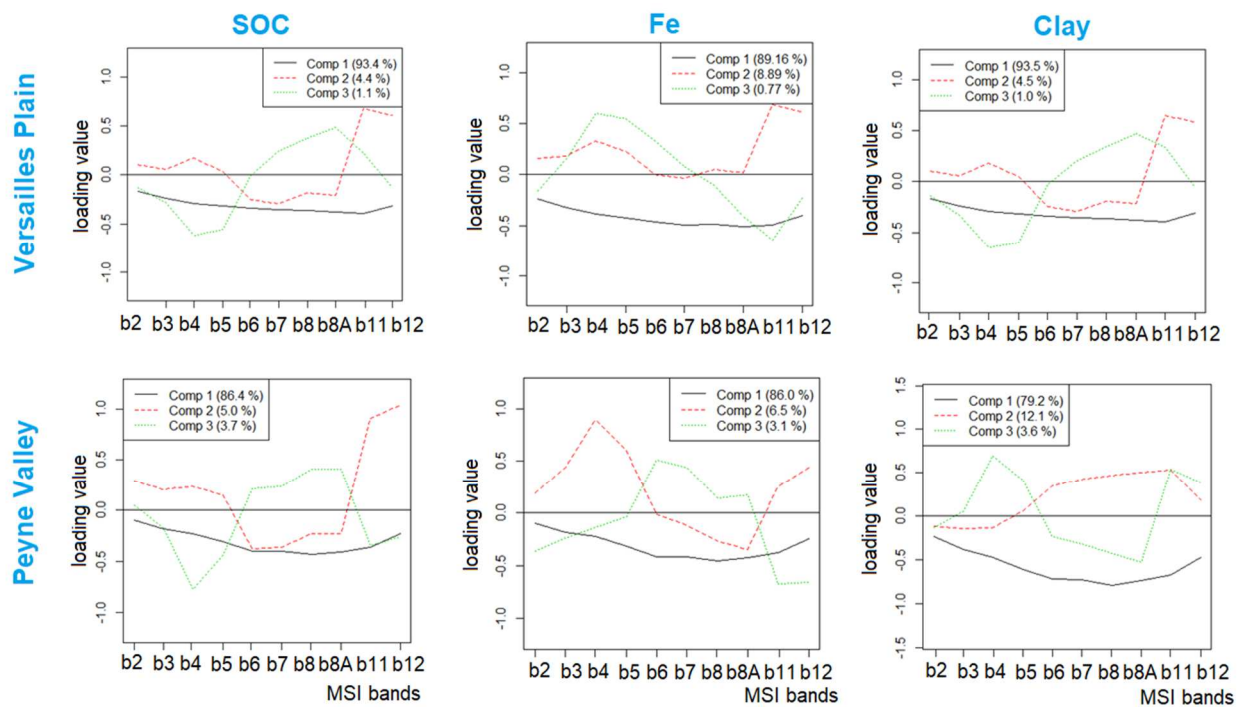
446
447 Figure 6. Empirical variograms computed for clay, iron and CEC from predicted soil
448 properties of bare soils of the Peyne. Lag = 100 m. Distance max: 2,000 m.

449

450 **5. Discussion**

451 *5.1 Performance of the S2A-derived PLSR models*

452 Comparisons of the performance of the S2A-derived models for two contrasting
453 agroecosystems suggest that the predictability of a given property does not solely rely on the
454 property itself or particularly on its spectral behavior as a “chromophore” according to Ben
455 Dor et al. (2009). Rather, if such conditions could explain predictive capacities, both iron and
456 SOC content levels would be predicted for any study area, as iron oxides and organic matter
457 are spectrally influent due to their chemical compositions and arrangements, while their
458 specific influent absorption bands are covered by MSI bands. In particular, SOC is known to
459 be spectrally sensitive across the entire VNIR-SWIR spectral range (Ben Dor et al., 2009;
460 Demattê et al., 2015), while the specific absorption wavelengths of iron oxides (550 nm and
461 roughly 860-900 nm) are overlain by green band b3 and near-infrared band b8A. As was
462 expected for SOC, loading values for SOC prediction models were found to be higher overall
463 for the visible b4, b5, near infrared b8A and two SWIR bands (b11, b12) for the Versailles
464 Plain than for the Peyne valley. For iron, loading values were found to be marked for the b3
465 and b8A bands for the Peyne Valley while they were weakly expressed for the Versailles
466 Plain (Fig. 7). This does not apply for both areas, suggesting that more conditions must be
467 applied to confer predictive capacity. Moreover, clay predictions may be linked to the
468 physical structure of soils in conjunction with a chromophore of approximately 2,200 nm that
469 is fully covered by the b12 band (Table 1): this was revealed by higher loading values (Fig. 7)
470 . However, performance of clay prediction was only found to be near-intermediate ($0.5 > R^2_{cv}$
471 > 0.4) for both study areas, suggesting that other conditions may hamper predictive capacity
472 levels.



473

474 Figure 7. Loading plots for SOC, Fe and Clay of both study areas

475

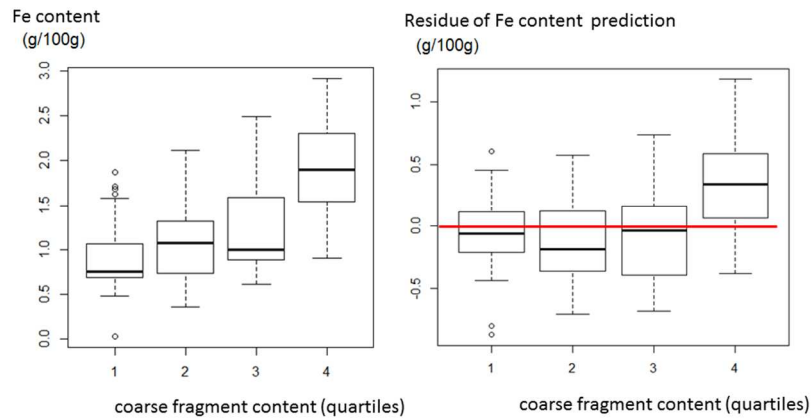
476 First, as previously found by Gomez et al. (2012) for the Payne Valley, intra-regional
 477 variances between the considered properties are large unless considered variables (even when
 478 spectrally sensitive) are not predictable. This is the case for both SOC and iron content levels,
 479 which are neither abundant nor variable across the Payne Valley and Versailles Plain,
 480 respectively.

481 Second, if the property considered is correlated with another well-predicted property,
 482 this might improve its predictive ability. Performance levels obtained for pH (Versailles
 483 Plain) and CEC (Versailles Plain, Payne Valley) are based on levels of variability across the
 484 study regions and correlations to either SOC and/or clay. As a matter of fact, pH and CEC are
 485 not spectrally sensitive. Moreover, while CaCO_3 has a specific absorption band (2,330 nm),
 486 its wavelength is not covered by any MSI band; its predictive ability might therefore be
 487 attributed to its correlation to pH (0.6) for the Versailles Plain and other disturbing factors
 488 might play a role for the Payne Valley. Although the prediction model for CaCO_3 yielded

489 intermediate performance figures-of-merit (when rounded) for the Versailles Plain, a careful
490 examination of the PLSR residues suggests that only soils with no or very low CaCO₃ content
491 levels present low predictions errors (< 35 g.kg⁻¹), and they represent 56% of the dataset,
492 resulting in apparently average intermediate performance (Table 6). Specifically, calcareous
493 soils of the Versailles Plain are as poorly predicted as those of the Peyne Valley. In addition,
494 CaCO₃ degrades model accuracy levels for SOC content, as the lab uncertainty of SOC
495 measurements increases with CaCO₃ content levels (Vaudour et al., 2016).

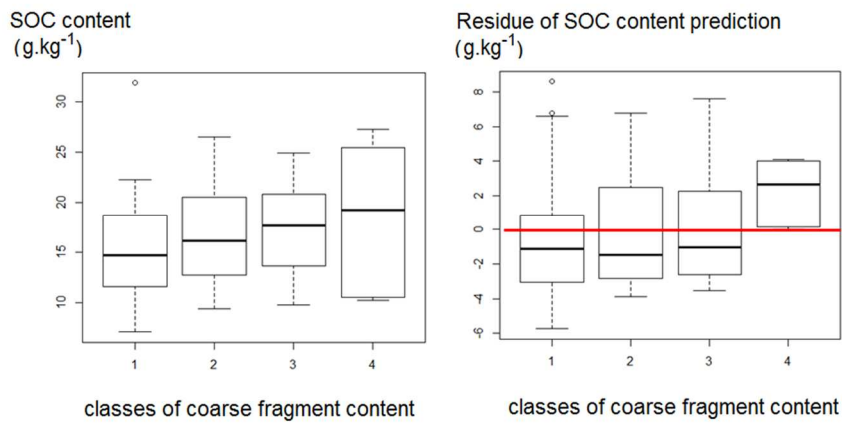
496 Third, the number and composition of sample sets is likely to influence prediction
497 performance. For the Versailles Plain, strong performance recorded for CEC must be
498 considered with caution, as it is based on a small sample size (39 samples) and the model may
499 be overfit.

500 Fourth, soil surface conditions can be disturbed by soil roughness, soil moisture,
501 emerging and non-covering vegetation, dry vegetation, vine stock and coarse fragment areal
502 coverage. In addition to commonly known difficulties associated with PLSR models in terms
503 of predicting high soil property values (e.g., for SOC (Stevens et al., 2012)), which are
504 inherent to linear models, some additional disturbances can arise from these specific soil
505 surface conditions. As a matter of fact, in the Peyne Valley, higher iron content levels are
506 found among very stony soils with coarse fragment content levels of greater than 40%: the
507 higher the coarse fragment content level, the more significant iron content prediction errors
508 become (Fig. 8). In the Versailles Plain, because calcareous soils are stony and as their rock
509 fragments are made of limestone, such relationships merge with those of calcium carbonate
510 content: the higher the coarse fragment content level, the higher the CaCO₃ content level and
511 the higher the SOC content level. The highest levels of CaCO₃ content (> 136 g. kg⁻¹) show
512 the highest levels of SOC residue prediction (Fig. 9).



513

514 Figure 8. Relationships between quartiles of coarse fragment content and Fe content (left) and
 515 Fe content prediction errors (right) for the Payne Valley. Quartile thresholds of coarse
 516 fragment content: 1, < 7.5%; 2, 7.5-18%; 3, 18-40%; 4, > 40%.



517

518 Figure 9. Relationships between quartiles of coarse fragment content and SOC content (left)
 519 and effects of coarse fragment content on SOC content prediction errors (right) for the
 520 Versailles Plain. Classes of coarse fragment content: 1, < 7.5%; 2, 7.5-18%; 3, 18-40%; 4, >
 521 40%.

522

523 Of course, it cannot be inferred from Figs. 8 and 9 that coarse fragment content alone causes
524 prediction errors, but presumably it contributes to such errors.

525 In terms of vegetation features, the composition and adequacy of the dataset is conditioned by
526 the NDVI threshold. Regarding dry vegetation, while the b12 band covers the 2,100 nm
527 specific absorption wavelength for cellulose (Daughtry 2001), the MSI bands do not enable
528 one to derive an index of dry vegetation such as the Cellulose Absorption Index ($CAI = 0.5$
529 $(R_{2000\text{ nm}} + R_{2200\text{ nm}}) - R_{2100\text{ nm}}$) elaborated by Nagler et al. (2000). The CAI relies on two
530 wavelengths covered by the b12 band (2,100 and 2,200 nm), but it requires the use of the
531 2,000 nm wavelength not covered by any MSI band. However, other indexes, such as the
532 Normalized Difference Tillage Index defined for Landsat Thematic Mapper (TM) bands
533 ($NDTI = (TM5 - TM7) / (TM5 + TM7)$) (Van Deventer et al., 1997), could be adapted and tested
534 for discriminating crop residue cover. The S2A data thus show limits in accounting for crop
535 residue cover and vine wood vegetation more than the MSI spatial resolution does not enable
536 one to discriminate between vine rows and vine interrows to remove the spectral influence of
537 vine woods.

538

539 In conclusion, our use of two contrasting agroecosystems datasets denotes what is achievable
540 from S2A data in terms of predictive capacities for a single considered date. The soils
541 considered in this study are located in French regions but widely develop abroad, across the
542 temperate and Mediterranean areas. Assessment of Sentinel-2 performance is developing for
543 other French regions with other soil types and further for tropical soils.

544

545

546

547

548 *5.2 Comparisons with other sensors*

549 Our model performs in a similar manner as those employed in other studies. It
550 performs better than SPOT multispectral satellite images in measuring SOC, and it performs
551 similarly or less optimally than hyperspectral airborne images in measuring all properties.

552 For the Versailles Plain in particular, previous models of SOC content were derived
553 from multispectral or hyperspectral fields or from image spectra. In terms of RMSECV and
554 RPD values in particular, the regional PLSR model obtained SOC content levels from S2A
555 spectra (1.23 g.kg⁻¹ and 1.51, respectively) more accurately than real SPOT or simulated
556 SPOT spectra (Vaudour et al., 2013: error range 4.5-6.0 g.kg⁻¹, RPD 0.9-1.3) for the same
557 region. This is presumably attributable to the beneficial role of its SWIR bands and to
558 differences in soil surface conditions, which are similar to those obtained from a VNIR
559 hyperspectral AISA airborne image (no SWIR) for the same region (Vaudour et al., 2016:
560 error range 2.82-3.79 g.kg⁻¹, RPD 1.3-1.6).

561 For the Peyne Valley, performance outcomes obtained from Sentinel-2 are lower
562 than those obtained from a VNIR-SWIR Hymap airborne image (Gomez et al., 2012),
563 regardless of soil properties involved with dramatic changes in performance found for CaCO₃
564 content effectively predicted from the hyperspectral spectra but not predicted from S2A
565 spectra (see §5.1). However, due to intra-regional variances (SOC) and an absence of
566 correlations with more effectively predicted properties (silt), neither SOC nor silt content
567 levels were predicted from hyperspectral and S2A images. Performance degradation from
568 airborne to spaceborne models has also been found by Steinberg et al. (2016) for SOC, clay
569 and iron content levels, though these authors used simulated EnMAP satellite spectra.

570 It must be noted that our models were constructed from actual spaceborne images that
571 account for the effects of atmospheric, signal/noise ratio, and soil surface conditions and

572 notably coarse fragment content and vegetation. This distinguishes our models from those
573 constructed from simulated MSI spectra derived from lab spectral libraries, as anticipated by
574 Castaldi et al. (2016) for a number of forthcoming or newly launched satellites, including
575 Sentinel-2. However, with the addition of noise and atmospheric effects and from the use of
576 untransformed lab reflectance spectra sampled from central and southern Italy, these authors
577 obtained validation results similar to our performance results for clay (R^2 0.36-0.42; RPD
578 1.26-1.35) that are inferior to those found for the Versailles Plain in terms of SOC content
579 levels (R^2 0.36; RPD 1.26) in the best case or that are not as predictive as found for the Payne
580 (R^2 0.13; RPD 1.09) in the worst case. Gomez et al. (2018) used VNIR/SWIR hyperspectral
581 airborne data (initial spectral resolution of approximately 5 nm) to simulate, among others,
582 Sentinel-2, Landsat 8 and Landsat 7 spectral resolution data. They showed that these
583 multispectral sensors provide very modest performances of clay content estimations, as these
584 spectral resolutions do not provide clay absorption feature useful in regression models.

585

586 *5.3 Maximizing the mapped area*

587 As the presence of vegetation severely hampers the prediction of soil properties from
588 VNIR/SWIR remote sensing data, it is important to retrieve as many bare soil signals as
589 possible. When only one image is considered, as was done in this study, this involves a
590 careful selection of acquisition periods. Images of the Versailles Plain and Payne Valley were
591 acquired for the spring period to maximize the size of the area with bare soils. In temperate
592 agroecosystems like that of the Versailles Plain, this corresponds to the sowing of spring
593 cereals, with the winter period being less suitable due to impacts of the Nitrate Directive that
594 requires intermediate crop cultivation (Vaudour et al., 2013). In a Mediterranean
595 agroecosystem like that of the Payne Valley, the spring period corresponds to maximum tilled
596 vineyard coverage for removing weeds (Paré, 2007) while green vegetation is not yet present.

597 Despite this careful selection of periods, coverage levels found for bare soil areas remain
598 modest, especially for the Versailles Plain (Fig. 4).

599 Bare soil areas could be maximized (at least for temperate agroecosystems) by
600 aggregating multiple acquisition dates. This would allow one to accumulate the spectra of soil
601 areas that change every year and along the crop cycle. Sentinel-2 is particularly well suited to
602 this task, as it provides an image every 5 days, thus increasing the likelihood of acquiring
603 images under clear conditions and across the crop rotation cycle.

604 Such an aggregation of images for maximizing bare soil areas was already attempted
605 by Diek et al. (2016). These authors doubled the bare soil area by combining hyperspectral
606 APEX acquisitions over 3 successive years and performed intercalibration from linear
607 regressions of atmospherically corrected spectra for different years. This raises issues relating
608 to accurate atmospheric corrections across dates and to the number and anteriority values of
609 dates. This issue is being studied in reference to the Versailles Plain to consider the
610 directional effects of soil roughness in line with tillage operations (Vaudour et al., 2014a).
611 Indeed, soil roughness levels can be spatially and temporally retrieved from optical/radar
612 pairs such as S2/S1 (Vaudour et al., 2014b), which could be accounted for during soil
613 properties prediction.

614 Bare soil coverage levels can also be increased by applying unmixing techniques that
615 isolate bare soil signals from semi-vegetated pixels (Bartholomeus et al 2011, Ouerghemmi et
616 al., 2016). For the Payne Valley, such techniques are used to predict clay content levels from
617 hyperspectral data for semi-vegetated areas ($NDVI < 0.55$), expanding the studied surface
618 from 4% (bare soil) to 63% of the total area (Ouerghemmi et al., 2016). A similar approach
619 could be applied for S2A soil predictions.

620

621

622

623 *5.4. Incorporating S2A data into the framework of digital soil mapping*

624 The soil predictions obtained from S2A are neither precise nor extended enough to be
625 exploited “as is” by end users. They must rather be considered as a new source of soil data
626 that should be used for Digital Soil Mapping (Mc Bratney et al, 2003, Lagacherie et al, 2007),
627 a larger methodological framework that can produce more exhaustive and precise soil maps.

628 If we examine the soil property predictions obtained in this study with such a perspective in
629 mind, S2A soil predictions obtained from this study could serve as valuable inputs for digital
630 soil mapping for two reasons.

631 First, S2A data offer insight on the spatial patterns of certain soil properties that could be
632 quantified from variograms that closely resemble those obtained from real measurements.
633 These variograms could be used to optimize further sampling efforts, as they allow one to
634 compute error variances (kriging variance) for sampling schemes of different densities and
635 spatial distributions and thus determine the sampling efforts required to meet target levels of
636 precision in spatial predictions (McBratney et al., 1981).

637 Second, S2A products could serve as surrogate data to improve the precision of Digital Soil
638 Mapping models (Lagacherie & Gomez; 2013). For example, Walker et al. (2017) found the
639 use of hyperspectral data as soft data for co-kriging or cokriging with external drift models to
640 improve prediction performance. Although less pronounced, improvements were still
641 observable when using hyperspectral-based estimations of moderate quality ($R^2 < 0.5$) like
642 many of those obtained in this study. It is therefore expected that some S2A soil outputs could
643 make improvements while being more available than hyperspectral data.

644 Finally, the potential of S2 data as a co-variable in DSM model could further be evaluated.
645 While previous researches used MODIS or LANDSAT data in DSM models, either as
646 spectral indexes, such as the NDVI (e.g., Mishra et al., 2010), or as spectral bands (e.g., Were

647 et al., 2015). Zaouche et al (2018) found improvement incorporating the green band of SPOT
648 into a joint Bayesian model, and development of their approach is in progress for Sentinel2.

649

650 **6. Conclusion**

651 S2A image spectra acquired from two French study areas representative of temperate and
652 Mediterranean agroecosystems—the Versailles Plain and the Payne Valley—were tested as
653 input data of a chemometric model (PLSR) to predict 8 topsoil properties of bare soil areas:
654 8% and 39% of the Versailles Plain and the Payne Valley, respectively. Six of these 8 soil
655 properties (clay, SOC, iron, CaCO₃, pH, and CEC levels) can be predicted with varying
656 success depending on their intrinsic spectral properties, intra-regional variances, correlations
657 and soil surface conditions. Some of the best predictions were also found to serve as good
658 approximations of the spatial patterns of soil properties.

659 Although a significant decline in the prediction performance of those obtained from
660 hyperspectral data for the same study areas was observed, S2A soil predictions can serve as
661 valuable inputs for Digital Soil Mapping. With the current absence of efficient hyperspectral
662 satellites that can deliver hyperspectral data, it is advantageous to allow for the mosaicking of
663 multirate acquisition and for the selection of the best acquisitions over a time series. We
664 therefore recommend their use to improve the performance of Digital Soil mapping
665 predictions.

666

667 **Acknowledgment**

668 This study was carried out in the framework of the TOSCA “Cartographie Numérique des
669 Sols (CNS)” program of the French Space Agency (CNES). It also benefited support from the
670 TOSCA-PLÉIADES-CO program of the CNES for the Versailles Plain. Special thanks to
671 Dalila Hadjar, Arthur Scriban, Paulina Michura, for technical assistance in the field. We are

672 also indebted to Yves Blanca, IRD-UMR LISAH Montpellier, for the soil samples sampling
673 in 2009 over the vineyard Plain of Languedoc.

674

675 **References**

676 AFNOR, 1999. Qualité des sols. AFNOR, Paris, vol. 1, 565 p. (in French)

677 Adeline K., Gomez C., Gorretta N., Roger J.M., 2017. Predictive ability of soil properties to
678 spectral degradation from laboratory Vis-NIR spectroscopy data. *Geoderma*. 288, pp.
679 143–153.

680 Adhikari, K., Hartemink, A.E., 2016. Linking soils to ecosystem services - a global review.
681 *Geoderma* 262, 101–111.

682 Arrouays, D., Grundy, M.G., Hartemink, A.E., Hempel, J.W., Heuvelink, G.B.M., Young
683 Hong, S., Lagacherie, P., Lelyk, G., McBratney, A.B., McKenzie, N.J., Mendonça-
684 Santos, M.L., Minasny, B., Montanarella, L., Odeh, I.O.A., Sanchez, P.A., Thompson,
685 J.A., Zhang, G.L., 2014. GlobalSoilMap: toward a fine-resolution global grid of soil
686 properties. *Advances in Agronomy*, 125, 93-134.

687 Arrouays, D., Lagacherie, P., Hartemink, A.E., 2017. Digital soil mapping across the globe.
688 *Geoderma Reg.* 9, 1–4.

689 Bartholomeus, H., Kooistra, L., Stevens, A., Leeuwen, M., Wesemael, B., Ben-Dor, E.,
690 Tychon, B., 2011. Soil organic carbon mapping of partially vegetated agricultural fields
691 with imaging spectroscopy. *International Journal of Applied Earth Observation and*
692 *Geoinformation* 13 (1), 81–88.

693 Baveye, P., Baveye, J., Gowdy, J., 2016. Soil “ecosystem” services and natural capital:
694 critical appraisal of research on uncertain ground. *Frontiers in Environmental Science*, 4
695 (41).

696 Ben-Dor, E., Chabrillat, S., Demattê, J.A.M., Taylor, G.R., Hill, J., Whiting, M.L., Sommer,
697 S., 2009. Using imaging spectroscopy to study soil properties. *Remote Sens. Environ.*
698 113, S38-S55.

699 Castaldi, F., Palombo, A., Santini, S., Pignatti, S., Casa, R., 2016. Evaluation of the potential
700 of the current and forthcoming multispectral and hyperspectral imagers to estimate soil
701 texture and organic carbon. *Remote Sensing of Environment*, 179, 54-65.

702 Chang, C.W., Laird, D.A., Mausbach, M.J., Hurburgh, C.R., 2001. Near infrared reflectance
703 spectroscopy: principal components regression analysis of soil properties. *Soil Sci. Soc.*
704 *Am. J.* 65, 480-490.

705 Crahet, M., 1992. Soil Map of the Versailles Plain. Scale 1:50 000. Institut National
706 Agronomique Paris-Grignon, Grignon, Internal report.

707 Daughtry, C.S.T., 2001. Discriminating crop residues from soil by shortwave infrared
708 reflectance. *Agron. J.* 93, 125-131.

709 Demattê, J., Morgan, C.L.S. Chabrillat, S., Rizzo, R., Franceschini, M.H.D., Terra, F.daS.,
710 Vasques, G.M., Wetterlind J., 2015. Spectral sensing from ground to space in soil
711 science: State of the art, applications, potential and perspectives. In book: *Remote*
712 *Sensing Handbook - Land Resources Monitoring, Modeling, and Mapping with Remote*
713 *Sensing*, Publisher: CRC Press, Editors: P.S. Thenkabail, pp.661-732.

714 Diek, S., Schaepman, M.E., de Jong, R., 2016. Creating Multi-Temporal Composites of
715 Airborne Imaging Spectroscopy Data in Support of Digital Soil Mapping. *Remote Sens.*
716 8, 906.

717 Filzmoser, P., Garrett, R.G., Reimann, C., 2005. Multivariate outlier detection in exploration
718 geochemistry. *Computers&Geosciences* 31, 579-587.

719 Folkman, M., Pearlman, J., Liao, L., Jarecke, P., 2001. EO1/Hyperion hyperspectral imager
720 design, development, characterization and prediction. In: Smith, W.L., Yasuoka, Y.
721 (Eds.), *Hyperspectral Remote Sensing of the Land and Atmosphere*. SPIE Proceeding,
722 vol. 4151, pp. 40–51.

723 Geladi, P., Kowalski, B. R., 1986. Partial least squares regression: a tutorial. *Analytica*
724 *Chimica Acta* 185, 1-17.

725 Gomez, C., Adeline, K., Bacha, S., Driessen, B., Gorretta, N, Lagacherie, P, Roger, JM,
726 Briottet, X., 2018. Sensitivity of clay content prediction to spectral configuration of
727 VNIR/SWIR imaging data, from multispectral to hyperspectral scenarios. *Remote*
728 *Sensing of Environment* 204, 18-30.

729 Gomez, C., Viscarra Rossel, R.A., McBratney, A.B., 2008. Soil organic carbon prediction by
730 hyperspectral remote sensing and field vis-NIR spectroscopy: an Australian case study.
731 *Geoderma* 146, 403–411.

732 Gomez, C., Coulouma, G., Lagacherie, P., 2012a. Regional predictions of eight common soil
733 properties and their spatial structures from hyperspectral Vis–NIR data, *Geoderma*,
734 189–190, 176-185.

735 Gomez, C., Lagacherie P., Bacha, S., 2012b. Using an VNIR/SWIR hyperspectral image to
736 map topsoil properties over bare soil surfaces in the Cap Bon region (Tunisia). In
737 “Digital Soil Assessments and Beyond” Minasny B., Malone B.P., McBratney A.B.
738 (Ed.). Springer, 387-392.

739 Gomez, C., Oltra-Carrió, R., Bacha, S., Lagacherie, P., Briottet, X., 2015. Evaluating the
740 sensitivity of clay content prediction to atmospheric effects and degradation of image
741 spatial resolution using hyperspectral VNIR/SWIR imagery. *Remote Sens. Environ.*
742 164, 1-15.

743 Goovaerts, P., 1997. Geostatistics for natural resources evaluation. Oxford University Press,
744 New York.

745 Hagolle, O., 2013. Séries temporelles, SPOT4 Take Five blog, URL: <http://www.cesbio.upstlse.fr/multitemp/>

746

747 Hagolle, O., Huc, M., Villa Pascual, D., Dedieu, G., 2010. A multi-temporal method for cloud
748 detection, applied to FORMOSAT-2, VEN μ S, LANDSAT and SENTINEL-2 images.
749 Remote Sens. Environ. 114, 1747-1755.

750 Kennard R. W., Stone L. A., 1969. Computer aided design of experiments. Technometrics 11,
751 1, 137–148.

752 Lagacherie, P., Sneep, A.-R., Gomez, C., Bacha, S., Coulouma, G., Hamrouni, M.H., Mekki,
753 I., 2013. Combining Vis–NIR hyperspectral imagery and legacy measured soil profiles
754 to map subsurface soil properties in a Mediterranean area (Cap-Bon, Tunisia).
755 Geoderma 209–210, 168–176.

756 Mark, H.L., Tunnell, D., 1985. Qualitative near-infrared reflectance analysis using
757 Mahalanobis distances. Anal. Chem. 57, 1449-1456.

758 McBratney, A. B., Webster, R., Burgess, T. M.: The design of optimal sampling schemes for
759 local estimation and mapping of regionalized variables. I. Theory and Method,
760 Computers and Geosciences, 7, 331–334, 1981.

761 McBratney, A.B., Mendonca Santos, M.L., Minasny, B., 2003. On digital soil mapping.
762 Geoderma 117, 3–52.

763 Mehra O.P., Jackson M.L., 1960. Iron oxide removal from soils and clays by a dithionite-
764 citrate system buffered with sodium bicarbonate. In Clays and Clay Minerals;
765 Proceedings of the 7th Conference, Natl Acad. Sci-Natl Res. Council Publ.,
766 Washington, DC, pp. 317-327.

767 Mevik B.H., Wehrens R., 2007. The pls package: principal component and partial least
768 squares regression in R. *Journal of Statistical Software* 18, 2, 1-24.

769 Minasny, B., McBratney, A.B., 2016. Digital soil mapping: A brief history and some lessons.
770 *Geoderma* 264, 301–311.

771 Nagler, P.L., Daughtry, C.S.T., Goward, S.N., 2000. Plant litter and soil reflectance. *Remote*
772 *Sens. Environ.* 71, 207-215.

773 Mishra, U., Lal, R., Liu, D., Van Meirvenne, M., 2010. Predicting the Spatial Variation of the
774 Soil Organic Carbon Pool at a Regional Scale. *Soil Science Society of America Journal*
775 74, 906. <https://doi.org/10.2136/sssaj2009.0158>

776 Ouerghemmi, W., Gomez, C., Naceur, S., Lagacherie, P., 2016. Semi-blind source separation
777 for the estimation of the clay content over semi-vegetated areas using VNIR/SWIR
778 hyperspectral airborne data. *Remote Sens. Environ.* 181, 251–263.

779 Paré, N., 2007. Etude de la variabilité et de la dynamique des états de surface des sols
780 viticoles méditerranéens. Master Thesis AgroParisTech Paris.

781 Pebesma, E.J., 2004. Multivariable geostatistics in S: the gstat package. *Comput. Geosci.* 30,
782 683–691.

783 R Development Core Team, 2015. The Comprehensive R Archive Network. The R
784 Foundation for Statistical Computing, Wirtschaft Universitat, Vienna, Austria. URL:
785 <http://www.r-project.org/>

786 Ramirez-Lopez, L., Schmidt, K., Behrens, T., Van Wesemael, B., Demattê, J.A.M., Scholten,
787 T., 2014. Sampling optimal calibration sets in soil infrared spectroscopy. *Geoderma*
788 226-227, 140-150.

789 Richter, R., Schläpfer, D., 2016. ATCOR2/3 User guide, v.9.0.2, March 2016, 263 p.

790 Sanchez, P.A., Ahamed, S., Carré, F., Hartemink, A.E., Hempel, J., Huising, J., Lagacherie,
791 P., McBratney, A.B., McKenzie, N.J., Mendonca Santos, M.L., Minasny, B.,
792 Montanarella, L., Okoth, P., Palm, C.A., Sachs, J.D., Shepherd, K.D., Vågen, T.-G.,
793 Vanlauwe, B., Walsh, M.G., Winowiecki, L.A., Zhang, G.-L., 2009. Environmental
794 science. Digital soil map of the world. *Science* 325, 680–1.

795 Selige, T., Böhner, J., Schmidhalter, U., 2006. High resolution topsoil mapping using
796 hyperspectral image and field data in multivariate regression modeling procedures.
797 *Geoderma* 136, 235–244. <https://doi.org/10.1016/j.geoderma.2006.03.050>

798 Shabou, M., Mougenot, B., Chabaane, Z.L., Walter, C., Boulet, G., Aissa, N.B. Zribi, M.
799 2015. Soil Clay Content Mapping Using a Time Series of Landsat TM Data in Semi-
800 Arid Lands. *Remote Sensing*. 7, 6059-6078.

801 Steinberg, A., Chabrillat, S., Stevens, A., Segl, K., Foerster, S., 2016. Prediction of common
802 surface soil properties based on Vis-NIR airborne and simulated EnMAP imaging
803 spectroscopy data: prediction accuracy and influence of spatial resolution. *Remote*
804 *Sensing*, 8, 613. Doi:10.3390/rs8070613

805 Stevens, A., Miralles, I., Van Wesemael, B., 2012. Soil organic carbon predictions by
806 airborne imaging spectroscopy: comparing cross-validation and validation. *Soil Sci.*
807 *Soc. Am. J.* 76, 2174-2183.

808 Stevens, A., Udelhoeven, T., Denis, A., Tychon, B., Liroy, R., Hoffman, L., Van Wesemael,
809 B., 2010. Measuring soil organic carbon in croplands at regional scale using airborne
810 imaging spectroscopy. *Geoderma* 158, 32-45.

811 Van Deventer, A.P., Ward, A.D., Gowda, P.H., Lyon, J.G., 1997. Using Thematic Mapper
812 data to identify contrasting soil plains and tillage practices. *Photogrammetric*
813 *Engineering & Remote Sensing* 63, 1, 87-93.

814 Vaudour, E., Bel, L., Gilliot, J.M., Coquet, Y., Hadjar, D., Cambier, P., Michelin, J., Houot,
815 S., 2013. Potential of SPOT multispectral satellite images for mapping topsoil organic
816 carbon content over peri-urban croplands. *Soil Sci. Soc. Am. J.* 77, 2122-2139.

817 Vaudour, E., Gilliot, J.M., Bel, L., Bréchet, L., Hadjar, D., Hamiache, J., Lemonnier, Y.,
818 2014a. Uncertainty of soil reflectance retrieval from SPOT and RapidEye multispectral
819 satellite images using a per-pixel bootstrapped empirical line atmospheric correction
820 over an agricultural region. *International Journal of Applied Earth Observation and*
821 *Geoinformation* 26, 217-234.

822 Vaudour, E., Baghdadi, N., Gilliot, J.M., 2014b. Mapping tillage operations over a peri-urban
823 region using combined SPOT4 and ASAR/ENVISAT images. *Int. J. Appl. Earth*
824 *Observ.Geoinf.* 28, 43-59.

825 Vaudour, E., Gilliot, J.M., Bel, L., Lefevre, J., Chehdi, K., 2016. Regional prediction of soil
826 organic carbon content over temperate croplands using visible near-infrared airborne
827 hyperspectral imagery and synchronous field spectra. *Int. J. Appl. Earth Observ.Geoinf.*
828 49, 24-38.

829 Vaudour, E., Noirot-Cosson, P.E., Membrive, O., 2015. Early-season mapping of crops and
830 cultural operations using very high spatial resolution Pléiades images. *Int. J. Appl. Earth*
831 *Observ.Geoinf.* 42, 128-141.

832 Wold, S., Sjöström, M., Eriksson, L., 2001. PLS-regression: a basic tool of chemometrics.
833 *Chemometrics and Intelligent Laboratory Systems* 58, 109-130.

834 World Reference Base (WRB), 2014. World reference base for soil resources. A framework
835 for international classification, correlation and communication. Food and Agriculture
836 Organization of the United Nations, Rome, 128 p.

837 Zaouche, M., Bel, L., Vaudour, E., 2017. Geostatistical mapping of topsoil organic carbon
838 and uncertainty assessment in Western Paris croplands (France). *Geoderma Regional*
839 10, 126-137.

840 Were, K., Bui, D.T., Dick, Ø.B., Singh, B.R., 2015. A comparative assessment of support
841 vector regression, artificial neural networks, and random forests for predicting and
842 mapping soil organic carbon stocks across an Afromontane landscape. *Ecological*
843 *Indicators* 52, 394–403. <https://doi.org/10.1016/j.ecolind.2014.12.028>

844 Zaouche, M., Bel, L., Tressou, J., Vaudour, E., 2018. Soil organic carbon modelling using
845 jointly different sources. METMA2018, Montpellier (France), poster.
846 <http://metma2018.sfds.asso.fr/assets/Download/proceedingsMETMA9.pdf>

847 **List of figure captions**

848 Figure 1. Locations of soil samples from the Versailles Plain (left, infrared colored S2A
849 image from 12 March 2016) and the lower Peyne Valley (right, infrared colored S2A image
850 from 19 March 2016), ©ESA 2016

851 Figure 2. Selection of reflectance image spectra covering the whole dataset reflectance range
852 and the main soil types

853 Figure 3. Predicted vs. observed soil properties for the Versailles Plain (left) and Peyne
854 Valley (right) for 6 soil properties best predicted for at least one region

855 Figure 4. Maps of soil properties of the Versailles Plain (left) and Peyne Valley (right)

856 Figure 5. Empirical variograms calculated for clay, iron and CEC for the Peyne Valley from
857 physico-chemical values of the collected soil samples (red points) and from predicted values
858 based on S2A spectra (black points) and fitted theoretical variograms (lines). Lag = 200 m.
859 Distance max: 2000 m.

860 Figure 6. Empirical variograms computed for clay, iron and CEC from predicted soil
861 properties of bare soils of the Peyne. Lag = 100 m. Distance max: 2,000 m.

862 Figure 7. Loading plots for SOC, Fe and Clay of both study areas

863 Figure 8. Relationships between quartiles of coarse fragment content and Fe content (left) and

864 Fe content prediction errors (right) for the Peyne Valley. Quartile thresholds of coarse

865 fragment content: 1, < 7.5%; 2, 7.5-18%; 3, 18-40%; 4, > 40%.

866 Figure 9. Relationships between quartiles of coarse fragment content and SOC content (left)

867 and effects of coarse fragment content on SOC content prediction errors (right) for the

868 Versailles Plain. Classes of coarse fragment content: 1, < 7.5%; 2, 7.5-18%; 3, 18-40%; 4, >

869 40%.

870