

When behavioural geographic differentiation matters: inter-populational comparison of aggressiveness and group structure in the European perch

Lola Toomey, Martin Bláha, Emilie Mauduit, Tatyana Vanina, Margot Baratçabal, Yannick Ledore, Sami Vesala, Pascal Fontaine, Alain Pasquet,

Thomas Lecocq

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20	Short title: Behavioural geographic differentiation in Perca fluviatilis
21	Lola Toomey ^{1*} , Martin Bláha ² , Emilie Mauduit ¹ , Tatyana Vanina ² , Margot Baratçabal ¹ ,
22	Yannick Ledoré ¹ , Sami Vesala ³ , Pascal Fontaine ¹ , Alain Pasquet ^{1, 4} , Thomas Lecocq ^{1*}
23	¹ Université de Lorraine, Inra, URAFPA, F-54000 Nancy, France
24	² Faculty of Fisheries and Protection of Waters, South Bohemian Research Centre of
25	Aquaculture and Biodiversity of Hydrocenoses, University of South Bohemia in Ceske
26	Budejovice, Zátiší 728/II, CZ-389 25 Vodňany, Czech Republic
27	³ Natural Resources Institute Finland, Helsinki, Finland
28	⁴ CNRS (Centre national de la Recherche Scientifique)
29	
30	* Corresponding Authors
31	lola.toomey@univ-lorraine.fr +33372745200
32	thomas.lecocq@univ-lorraine.fr +33372745696
33	University of Lorraine, Boulevard des Aiguillettes, BP 70239, 54506 VANDOEUVRE-LES-
34	NANCY CEDEX FRANCE

ABSTRACT

Domestication is still a long and difficult process and it is particularly impacted by species behavioural traits. Indeed, tolerance to high densities in intensive cultures and sociability are major features which facilitate domestication and influence the effectiveness of aquaculture production. Moreover, behavioural domestication predispositions could change at the intraspecific level. Here, we investigate three essential behavioural traits: aggressive interactions, group structure, and activity between three allopatric populations of Perca fluviatilis, a fish species at its nascent stage of production. We highlight inter-populational differences in group structure and aggressive interactions but not in activity. A more cohesive and homogeneous group structure was demonstrated for Finnish populations compared to Lake Geneva at 45-46 days post-hatching. In addition, Lake Geneva presented a higher aggressiveness. These inter-populational differences could be used in European perch aquaculture in order to improve production as well as welfare of individuals.

KEYWORDS

Aggression - aquaculture - behaviour - cannibalism - intraspecific differentiation - larvae *Perca fluviatilis* - social structure

56 **INTRODUCTION**

Increasing the agriculture sustainability, including aquaculture, relies partly on the 57 production and domestication of new species (Gepts et al. 2012). Domestication is 58 considered as the process in which populations are bred in man-controlled environment and 59 60 modified across successive generations from their wild ancestors in ways making them more useful to humans who control, increasingly during the process, their reproduction and food 61 supply (Lecocq, 2018). However, domestication remains a difficult, long, and expensive 62 process ridden by unfruitful outcomes, mostly due to zootechnical issues or taxon intrinsic 63 features (Liao and Huang 2000; Diamond 2002; Teletchea and Fontaine 2014). This is 64 particularly acute in intensive aquaculture (i.e. nowadays, intensive monoculture is the 65 primary aquaculture) in which many new species domestication trials are hampered by 66 several bottlenecks and end up being abandoned (Teletchea and Fontaine 2014). For 67 instance, some fish species display low resistance to diseases or low food conversion 68 69 efficiency, which impede or slow down their domestication (Liao and Huang 2000; Otton 70 2004). Conversely, other traits are favouring domestication such as fast growth rate and 71 acceptance of artificial feeds, and consequently make taxon production an economically viable initiative (Liao and Huang 2000; Le François et al. 2010). Among these features, some 72 behavioural traits are particularly essential since they can deeply facilitate domestication 73 (Liao and Huang 2000; Jobling 2010; Le François et al. 2010) and subsequent aquaculture 74 75 production (Huntingford et al. 2012).

Among behavioural traits, inter-individual relationships, group structure, and activity can affect directly the ability of a species to be domesticated and efficiently produced in intensive monoculture conditions. Tolerance to conspecifics in a limited area is an essential

79 parameter for production (Kristiansen et al. 2004) since it affects individual welfare (Huntingford 2004; Ashley 2007). Selecting populations which present an aggregative and 80 cohesive group structure, therefore limiting stress, would be favouring welfare. However, 81 living in group is not costless as it can trigger for instance competition for resources (Pitcher 82 and Parrish 1993; Martins et al. 2012; Ward and Webster 2016). In culture conditions, this 83 84 can lead to the emergence of aggressive behaviours (Damsgård and Huntingford 2012), such as attacks or bites, leading in some cases to cannibalism (Baras 2013). These aggressive 85 behaviours have several potentially negative consequences in fish culture such as 86 mortalities, stress, immune-suppression, or uneven competition for food (Damsgård and 87 Huntingford 2012 and references therein). Cannibalism (type I: prey is caught tail first and 88 ingested partially; type II: prey is caught by the head or tail and fully ingested; Baras et al. 89 90 2003; Baras 2013) is a major bottleneck in finfish aquaculture (Naumowicz et al. 2017) since it can lead to important losses (Baras et al. 2003; Huntingford et al. 2012). For example, 91 cannibalism can cause up to 50% losses in Perca fluviatilis (Baras et al. 2003; Kestemont et al. 92 93 2003). Finally, activity is also an important factor in aquaculture as it contributes to the total energetic budget (e.g. up to 40% of Perca flavescens budget; Boisclair and Leggett 1989). 94 95 Moreover, less active taxa could contribute to lower potential contacts and subsequent potential aggressive interactions. Therefore, in domestication processes, it is necessary to 96 97 take into account the ability for taxa to present the most suitable group structure, low aggressive interaction rate, as well as lower activity. Yet, there is an intraspecific 98 99 differentiation (differentiation between allopatric populations of conspecific individuals; 100 Mayr 1963) which could further help to improve domestication processes.

101 Behavioural intraspecific differentiation and its potential for selection of founder 102 populations have been poorly investigated to date. However, intraspecific differences in 103 aggressive behaviour (Magurran and Seghers 1991; Mandiki et al. 2004; Bell 2004), time spent foraging in an open habitat (Bell 2004; Magnhagen 2006), schooling (Magurran and 104 Seghers 1991), or boldness (Wright et al. 2003) have been already assessed for a few species 105 106 (see also Foster 1999). Abiotic factors can influence social behaviours (e.g. temperature, light, population density; Baras et al. 2003; Kestemont et al. 2003), yet a genetic basis was 107 108 also suggested since allopatric populations or geographically distinct strain differentiations were demonstrated for a few species (Amundsen et al. 1999; Damsgård and Huntingford 109 2012; Magnhagen et al. 2015). Therefore, considering such behavioural intraspecific 110 111 differentiation could allow improving aquaculture for species for which production is still 112 limited by behavioural bottlenecks.

113 The European perch, Perca fluviatilis L., is one of the fish species involved in the European aquaculture diversification (Kestemont et al 2015). Its long standing socio-economic interest 114 (high market value and recreational interest) led to the development of its aquaculture in 115 the 90's (Kestemont and Mélard 2000; Kestemont et al. 2015). However, its production is 116 still limited due to several bottlenecks including some aspects related to fish behaviour such 117 118 as aggressiveness and high cannibalism and subsequent mortalities (Kestemont et al. 2015). However, geographic differentiation has been previously observed for some of problematic 119 behavioural traits (e.g. cannibalism rate, Mandiki et al. 2004). Therefore, we aim in this study 120 121 at (i) assessing if European perch allopatric populations present differentiation for group structure and activity, as well as for aggressive interactions during first-life stages, and (ii) 122 identifying populations presenting behavioural advantages for production. 123

124 MATERIAL AND METHODS

125 Rearing conditions

Rearing parameters were chosen according to trade-offs between abiotic culture conditions used in literature (e.g. Vlavonou 1996; Kestemont et al. 2003; Kestemont et al. 2015), our practices, and fish farming practices. The rearing protocol was tested and validated with a domesticated population from the fish farm "Lucas Perches" (Hampont, France) comparing growth and survival results to literature (e.g. Vlavonou 1996; Fiogbé and Kestemont 2003).

Egg ribbons were obtained during the 2018 spawning season (May 2018) from lakes Geneva 131 (GEN; Switzerland; 46°26'N, 6°33'E), Valkea-Müstajärvi (VAL; Finland; 61°13'08"N, 132 25°07'05"E), and Iso-Valkjärvi (ISO; Finland; 60°57'21"N, 26°13'3"E). After transport, 19 egg 133 134 ribbons per lake were incubated at the Experimental Platform of Aquaculture (Unit of Animal Research and Functionality of Animal Products, University of Lorraine, Vandœuvre-135 lès-Nancy, France) in incubators (110x64x186cm; one incubator per population), containing 136 nine racks each (45x7x12cm), at 13°C. Each incubator had its own temperature control and 137 recirculated water (flow rate of 4m³.h⁻¹) system and water was UV sterilised. Oxygen rate 138 $(10.5 \pm 0.2 \text{mg.L}^{-1})$ and temperature $(13.0 \pm 0.3 \text{°C})$ were checked daily while pH was 139 measured three times a week (8.0 \pm 0.1). Ammonium (lower than 0.05mg.L⁻¹) and nitrite 140 concentrations were monitored three times a week until hatching (lower than 0.01mg.L⁻¹). 141 Photoperiod was 12L:12D and light intensity was 400 lx at the water surface. 142

143 Two independent experiments were performed in order to ensure availability of larvae 144 across the rearing period: experiment I from hatching until the end of weaning (26 days post-hatching, dph) and experiment II from 27 dph until 60 dph. All populations were rearedin independent structures.

Concerning experiment I, after hatching, larvae from the different egg ribbons were mixed 147 and transferred to three green internal-wall 71L cylindro-conical tanks (three replicates per 148 population; recirculated aquaculture system (RAS)) at a density of 50 larvae.L⁻¹. Temperature 149 150 was gradually increased during two weeks to 20°C, photoperiod was 12L:12D and light 151 intensity was 400lx. Larvae were fed with newly hatched Artemia naupli (Sep-Art, INVE) every 1h30 from 3 dph until weaning. At 16 dph, Artemia ration was decreased by 25% every 152 three days and dry feed ration (BioMar, 300µm until 21 dph, then 500µm) was increased by 153 154 the same ratio. After 25 dph, larvae were only fed with dry feed ad libitum (BioMar 500µm, then 700µm at 44 dph until the end of the experiment). At 26 dph, the larvae in cylindro-155 conical tanks were removed to start experiment II. 156

For experiment II, larvae not used for experiment I were held after hatching in 2m³ tanks (RAS) under the same temperature, feeding, light intensity, and photoperiod regimes as individuals of the experiment I. At 27 dph, these larvae were transferred at a density of 19 larvae.L⁻¹ to the three cylindro-conical tanks in order to start experiment II. Light intensity was 80 lx at water surface, all else remaining equal to experiment I (except for density).

During the two experiments, oxygen concentration $(8.7 \pm 2.3 \text{mg.L}^{-1})$ and temperature $(20.0 \pm 0.6 \text{ °C})$ were checked daily for all tanks. Ammonium $(0.14 \pm 0.1 \text{ mg.L}^{-1})$, pH $(7.2 \pm 0.9 \text{ mg.L}^{-1})$, and nitrite concentrations $(0.08 \pm 0.08 \text{ mg.L}^{-1})$ were monitored three times a week. Tanks were cleaned daily after first feeding and dead individuals were removed every morning. Survival rate for ISO, VAL and GEN were respectively 40.1 % (± 12.0), 29.4 % (± 14.5) and 6.6 % (± 3.4) for experiment I (26 dph; statistical difference between Geneva Lake and the two Finnish populations; F=7.2, df=2, *P*<0.05) and 31.5 % (\pm 4.3), 28.0 % (\pm 11.1) and 37.4 (\pm 8.1) for experiment II (60 dph; no statistical difference between the three populations; F=0.98, df=2, *P*=0.42)). These ranges of survival rates were comparable to what is found in literature (e.g. Tamazouzt et al. 2000; Baras et al. 2003; Fiogbé and Kestemont 2003).

172 Group structure and activity

For each population, three replicates for each cylindro-conical tank were performed over 173 174 two days (25 and 26 dph). At 24 and 25 dph, a total of 90 individuals (n=30 for each cylindro-175 conical tank, 10 individuals per replicate) were sampled for each population and transferred 176 to three aquaria (58 L; one aquarium per cylindro-conical tank; order of cylindro-conical replicates randomly assessed over two days; see appendix 1) with an 80lx light intensity and 177 a temperature of 20.0°C (± 0.5). Individuals were not fed from the moment they were 178 179 transferred to the beginning of the experiment the following day in order to have individuals 180 in the same energetic state. After one night of acclimatisation, individuals were tested by 181 groups of ten in circular arenas. Groups of ten individuals might not reflect faithfully what occurs in cylindro-conical tanks. However, evaluation cannot be performed directly in the 182 183 tanks and this method was previously validated (e.g. Colchen et al. 2016). Three circular arenas (30 cm diameter with 1.5cm of water depth) were used to investigate group structure 184 185 and activity (Colchen et al. 2016). Water in the arena was the same as in the aquaria, room 186 temperature was maintained at 20.0 °C (± 0.6) and arenas were lit at 10lx from underneath 187 in order to avoid shadows during recording. For each replicate, individuals were transferred from the aquarium to the arena with a beaker and a siphon. After 30min acclimatisation, 188 189 individuals were filmed for 30min using camcorders (Sony, Handycam, DCR-SR72E) located 190 50cm above the arena. The three arenas were filmed simultaneously and the order of 191 replicates tested was randomly assessed. After 1h, individuals were euthanised with an overdose of MS-222 following European rules and kept in formalin 4% for later length 192 measurements. Larvae tested from ISO, VAL and GEN were respectively 14.05 ± 0.55mm, 193 12.90 ± 0.62mm, and 13.87 ± 0.26mm. This full experiment was performed again during 194 195 experiment II with fish sampled from cylindro-conical tanks at 44 and 45 dph. For this second 196 test, individuals from ISO, VAL and GEN were respectively 26.74 ± 1.67mm, 26.28 ± 1.99mm, 197 and 22.97 ± 1.08 mm (no statistical difference between populations for the two experiments; 198 experiment I: F=0.712, df=2, P=0.528; experiment II: F=1.68, df=2, P=0.263).

199 Group structure analysis was performed using the ImageJ software. Images were extracted 200 from videos at 3-min interval (11 images per video). From each image, exact coordinates of 201 each individual were noted using the middle point between the eyes. Three parameters 202 were evaluated to assess the group structure: the nearest neighbour distance, the mean of 203 inter-individual distances, and the variance of these inter-individual distances (Buske and 204 Gerlai 2011a). Nearest neighbour distance represents the distance between a focal fish and 205 its closest neighbour and is an indicator of the group aggregation. The mean of inter-206 individual distances corresponds to the mean of distances between a focal fish and all the 207 other fish of the group and the average of values from all group members is an indicator of 208 the group cohesion. Finally, the average of variances of inter-individual distances from each fish represents the homogeneity of distribution (Buske and Gerlai 2011a). Activity was also 209 calculated in ImageJ. One image per second was extracted for six consecutive seconds every 210 211 five minutes. Coordinates of each individual were noted for each image and then distance 212 swam was calculated every second during the five seconds then averaged to obtain the mean distance swam for each individual per second. 213

All statistical analyses were performed in R 3.0.3 (R Core Team 2017). To test the normality of distributions, a Shapiro-Wilk test (R Core Team 2017) was used and homogeneity of variances was tested using the Levene test (Gastwirth et al. 2015). Then, linear mixed models were used with distances and activity as fixed factors and cylindro-conical tanks as random factor (Bates et al. 2004). There was no influence of the cylindro-conical tank on all models. Therefore, one-way analyses of variance (ANOVA F test) followed by Tukey post hoc tests were used to evaluate differences between populations (R Core Team 2017).

221 Quantification of aggressive interactions

222 Daily observations were carried out at different moments of the day previous to this experiment but did now allow identifying a cannibalism peak during the photophase. 223 Therefore, we hypothesised that the beginning of the photophase would correspond to the 224 225 cannibalism peak since individuals were not fed between 5.30pm and the next morning, that 226 they were used to be fed during the photophase and that European perch is a visual 227 predator (Graeb et al. 2005; Kestemont et al. 2008). Therefore, observations were performed after first feeding with a five-minute acclimatisation to the presence of the 228 229 observer and five minutes of focal sampling (Colchen et al., 2019). Daily observations were carried out between 8.30am and 10am from 10 dph until the end of experiment I (26 dph). 230 231 For experiment II, observations were carried out every three days (see Appendix 1). One 232 replicate per population was observed per day and the same person performed all 233 observations. Since hatching times were asynchronous between populations, the order of 234 populations or tanks observed were randomly assigned. Several aggressive behaviours were 235 noted: (i) Pursuit: an individual heads towards a conspecific, gets close and follows it when 236 the conspecific moves; this involves a change of direction of the two individuals, (ii) Attack:

237 when an individual heads towards a congener and gets rapidly close to it without necessarily 238 contact between the two individuals, (iii) Bite: when an individual catches with its mouth a part of a conspecific's body and then releases it, and (iv) Capture: when an individual ingests 239 a part or the whole conspecific (type I and II cannibalisms). Taking into account all these 240 aggressive interactions, a global daily aggressive interaction rate was calculated relatively to 241 242 the initial number of individuals in the tank. Enucleation, being a specific indicator of 243 aggressiveness in perch (Jourdan et al. 2000), was also evaluated by counting daily the 244 number of dead individuals enucleated. Enucleation rate was calculated relatively to the initial number of individuals in the tank. For phase II, type II cannibalism rate was estimated 245 by subtracting from the initial number of individuals the number of survivors and dead 246 247 individuals over phase II and calculating a rate relatively to the initial number of individuals 248 in the tank. Cannibalism rate could not be evaluated for experiment I as a precise monitoring of mortality was not possible the first week due to fast degradation of dead larvae. Finally, 249 250 mortality rates attributed to cannibalism and enucleation were also calculated. In order to meet assumptions of normality (Shapiro-Wilk test, R Core Team 2017) and homogeneity of 251 252 variances (Levene test, Gastwirth et al. 2015), data for all aggressive parameters was transformed (i.e. log(x+1)). One-way analyses of variance (ANOVA F test) followed by Tukey 253 post hoc tests were used to evaluate differences between populations (R Core Team 2017). 254 255 When assumptions were not respected (only for the aggressiveness rate during experiment I), Kruskal-Wallis H tests (R Core Team 2017) were used followed by Dunn post-hoc tests 256 257 (Pohlert 2015).

258 Compliance with ethical standards

All along experimental procedures, individuals were handled as little as possible. All procedures were in accordance with the national and international guidelines for protection of animal welfare (Directive 2010/63/EU). This study was conducted with the approval Animal Care Committee of Lorraine (CELMA n°66) and the Ministry of Higher Education, Research, and Innovation (APAFIS13368-2018020511226118).

264 **RESULTS**

265 Group structure and activity

266 <u>Experiment I (25-26 dph)</u>

Inter-individual distances (F=7.8, df=2, P<0.05), variance of inter-individual distances (F=9.9, df=2, P<0.05), and activity (F=8.2, df=2, P<0.05) are significantly lower for GEN compared to VAL and ISO (Fig. 1). There is no statistical difference between VAL and ISO. There is no statistical difference between populations for the nearest neighbour distance (F=1.4, df=2, P=0.2; Fig. 1).



Fig. 1: Boxplots representing group structure and aggressive interaction results for experiment I (n=3330). Four measures of group structure are presented: inter-individual distances, variance of inter-individual distances, nearest neighbour distances, and activity. The black line represents the median, the outsider box corresponds to lower and upper quartile values and white dots correspond to most extreme values within 1.5 times the interquartile range from the ends of the box. Different letters indicate significant differences between populations (P < 0.05) using post hoc tests.

279 *Experiment II (45-46 dph)*

Inter-individual distances (F=7.8, df=2, *P*<0.05) and variance of inter-individual distances (F=9.9, df=2, *P*<0.05) are significantly higher for GEN compared to VAL and ISO (Fig. 2). There is no statistical difference between ISO and VAL. There is no statistical difference between populations for the nearest neighbour distance (F=1.4, df=2, *P*=0.3) and activity (F=1.2, df=2, *P*=0.3; Fig. 2).

285 Quantification of aggressive interactions

286 The first cases of cannibalism (type II only) were observed at 26 dph for ISO and at 51 dph for GEN. No cannibalism case was observed for VAL. There is no difference of daily 287 aggressive interaction rate in experiment I (H=0.60, df=2, P=0.74; Fig. 1) but in experiment II 288 289 these behaviours were significantly higher in GEN than ISO and VAL (F=7.21, df=2, P<0.05; 290 Fig. 2). There is no statistical difference between ISO and VAL. Enucleation rate, which is null 291 in experiment I for all populations, is significantly higher in GEN compared to VAL and ISO (F=70.74, df=2, P<0.05; Fig. 2) in experiment II. Cannibalism rate is not statistically different 292 between populations (F=0.018, df=2, P=0.98; Fig. 2). Mortality rate attributed to cannibalism 293 294 and enucleation is not statistically different between the three populations (F=0.69, df=2, 295 P=0.59; Fig. 2).



Fig.2: Boxplots representing group structure and aggressive interaction results for 296 experiment II (n=1080). Four measures of group structure are presented: inter-individual 297 distances, variance of inter-individual distances, nearest neighbour distances and activity. 298 299 Aggressive interactions results include aggressive interaction rate, enucleation rate, cannibalism rate, and mortality rate caused by aggressive interactions. The black line 300 represents the median, the outsider box corresponds to lower and upper quartile values and 301 white dots correspond to most extreme values within 1.5 times the interquartile range from 302 the ends of the box. Different letters indicate significant differences between populations 303 304 (P < 0.05) using post hoc tests.

305 **DISCUSSION**

306 Inter-populational differentiation in behavioural traits and its potential causes

307 In this study, we highlight intraspecific differentiation between the two Finnish populations 308 and GEN for group structure (experiments I and II) and aggressive interactions (only 309 experiment II) while activity does not differ. We cannot exclude some potential biases in our experiments. For instance, (i) aggressive behaviour observations have been made after first 310 feeding but cannibalism peaks (and aggressions) might occur at different moments of the 311 312 day between populations and (ii) calculated cannibalism rate might also include dead 313 individuals eaten by conspecifics which can blurry differences in cannibalism rates. However, 314 since (i) temporal differentiation in cannibalism peaks has not been reported to date and (ii) the cannibalism estimation method, widely used across literature (e.g. Kestemont et al. 315 316 2003; Mandiki et al. 2004), allows to compare populations, we argue that bias related to 317 observations of inter-populational behavioural differentiations at two ages are limited.

318 Intraspecific differentiation in group structure and aggressive behaviours has been already 319 highlighted for several fish species (Rosenau and McPhail 1987; Magurran and Seghers 1991; 320 Amundsen et al. 1999; Lahti et al. 2001; Huizinga et al. 2009; Wark et al. 2011; Song et al. 321 2011). Here, a more cohesive and homogeneous structure is demonstrated for Finnish 322 populations compared to GEN at 45-46 dph (Fig. 2). These results, associated with a similar 323 nearest neighbour distance, indicate a structure in sub-groups in all populations but with a distance between these groups higher for GEN at 45 dph. The less homogeneous group 324 325 structure of GEN at 45 dph is quite congruent with the higher aggressiveness highlighted for 326 this population. Indeed, although daily aggressive interaction rate seems low (0.1-0.9 %; Fig. 2), the congruence between aggressive interaction patterns and enucleation rate supports 327

328 the higher aggressiveness of GEN compared to the two Finnish populations. The absence of 329 difference in cannibalism rate indicates that aggressive interactions are not necessarily followed by type II cannibalism. Therefore, our study is not congruent with Mandiki et al. 330 (2004) who showed a difference in intra-cohort cannibalism rate between different 331 European perch allopatric populations (but with different populations than the ones 332 333 investigated here). Here, we highlight differences in aggressive interactions aside from cannibalism rate. Inter-populational behavioural differences can be shaped by genetic 334 335 differentiation, by phenotypic plasticity, or by their combination.

336 On the one hand, the observed inter-populational behavioural differences could be shaped 337 by genetic differentiation. Indeed, population-specific demographic histories and potential local adaptations fostered by particular selective pressures can lead to the acquisition of 338 339 distinct behavioural phenotypic traits or development rates between allopatric conspecific 340 populations (Foster and Endler 1999; Foster 1999). For instance, it was shown a link between aggressiveness and the level of predation of the natural living site (Huntingford 1982; 341 Magurran and Seghers 1991) as well as other environmental factors such as food availability 342 and water current velocity (Lahti et al. 2001 and references therein). The occurrence of 343 inherited differences in aggressive interactions was assessed for several species (Huntingford 344 345 et al. 2012; Damsgård and Huntingford 2012). Similarly, inter-populational differences in activity were found to be connected to prey size distribution, total prey biomass, and water 346 transparency (Boisclair and Leggett 1989). Unfortunately, we do not have enough 347 348 information on the different lakes abiotic and biotic parameters to make any assessment. Another explanation of inter-populational differences could be divergences in development 349 rates potentially triggered by genetic specificities. Indeed, we compare the populations at 350

351 the same age but we do not know if the compared fishes are at the same developmental 352 stage (i.e. the lack of development table for larval and juvenile stages of *P. fluviatilis* prevent 353 us to assess if the development is synchronous between populations). The populations might have divergent development rates, which can trigger inter-populational differences in 354 parameters investigated. For instance, the higher aggressiveness in experiment II might be 355 356 related to the development of muscular and nutritional structures through the larval stage (Kestemont et al. 1996; Vlavonou 1996). It can also be related to the development of visual 357 358 structures since the visual acuity, essential for capture of prey, increases until metamorphosis (Guma'a 1982). In addition, the aggregation in sub-groups might be due to 359 several factors such as kinship (Behrmann-Godel et al. 2006), the nature of interactions (e.g. 360 aggressive interactions), spatial distribution, or differential sizes (Hinde 1976). Since these 361 362 two last factors are sensitive to developmental stage, group structure is also influenced by development rate. 363

On the other hand, phenotypic plasticity (i.e. the ability of a genotype to produce more than 364 one phenotype when exposed to different environments; Pigliucci et al. 2006; Kelly et al. 365 2012) is an alternative explanation of the observed behavioural differentiations between 366 populations (DeWitt and Scheiner, 2004) with behaviour reflecting the strategy adopted 367 368 under the influence of environmental factors. This was suggested as the driving factor for cannibalism in P. fluviatilis (Krol et al. 2015 and references therein) as well as in other 369 370 species (e.g. Svenning and Borgstrøm 2005). Since we have used an experimental transplant 371 approach (common environment), we speculate that we have minimized the effect of the environment (West-Eberhard 2003). Nevertheless, influences of past environmental 372 373 conditions (i.e. before the beginning of our experiment) cannot be ruled out. On the one

374 hand, we have collected individuals at the egg stage in the wild and phenotypic response to environmental conditions could have occurred during development (Swain and Lindsey 375 1986). On the other hand, environmental conditions experienced by the parents might have 376 377 influenced offspring phenotype (Mousseau and Fox 1998; Youngson and Whitelaw 2008). At 378 last, it was also shown an influence of maternal size on larvae performance (Olin et al. 2012). 379 Since we have no information on the parents of egg sampled, we cannot exclude the influence of maternal effects which were demonstrated in P. fluviatilis for other traits 380 381 (Babiak et al. 2004; Krol et al. 2015).

382 Overall, we cannot assess the importance of genetic differentiation, phenotypic plasticity, 383 and specific development rates in population-specific behaviour. Behavioural differentiation might be the result of the interaction of all factors (see for instance for cannibalism, Baras 384 385 and Jobling 2002; Yang et al. 2015). Moreover, effects of experience on behaviour cannot be 386 ruled out as it was previously suggested as an important factor for the behavioural variation 387 (Hellström and Magnhagen 2011; Magnhagen 2015). Further analyses over several generations with populations under identical rearing conditions as well as the establishment 388 of a development table for larval and juvenile stages of *P. fluviatilis* are needed to assess the 389 390 importance of each factor in the geographic differentiation of behaviour.

391 Differences in group structure and aggressive interactions between the two different ages

For each population, we observe a behavioural differentiation between the two studied ages. Such a differentiation has been previously observed in other species. For instance, development of shoaling (increasing protection against predators, foraging efficiency, and mate encounters) with age through a decrease of inter-individual and nearest neighbour distances has already been observed in zebrafish (Buske and Gerlai 2011b; Buske and Gerlai 397 2012). We observe similar development for ISO and VAL populations. In contrast, the opposite pattern observed for GEN population is unexpected and corresponds to the 398 399 establishment of a less homogeneous group with age. Mechanisms underlying the agedependant changes in group structure are so far unknown. Differential ontogenies of 400 sensory development might play a role into differences at the different ages (Buske and 401 402 Gerlai 2012). Several neuroanatomical, physiological or biochemical factors have been suggested to be involved (Buske and Gerlai 2012). There might also be some group 403 regulation mechanisms subsequent to weaning, which can potentially increase competition 404 for resources linked for instance to bioenergetic needs or physiological shifts. This 405 competition could explain the higher occurrence of aggressive interactions in experiment II. 406

407 Integrating inter-populational behaviour differentiation: a way to improve aquaculture408 production?

409 The variability occurring in behaviour at the intraspecific level offers the opportunity to 410 select fishes whose behaviours make them more suitable for aquaculture production (Huntingford et al. 2012). In the European perch, the lack of population-specific activity 411 412 tends to make this trait useless to select best population for aquaculture purpose. In contrast, difference in aggressiveness (i.e. and its consequences: losses due to aggressive 413 interactions ranged from about 20% to 40% depending on the population; experiment II, Fig. 414 415 2) is a potential selection criterion for farmers since such interactions are highly detrimental 416 for fish production. Similarly, the population-specific group structure should be considered as highly important information to highlight most suitable populations for intensive 417 aquaculture. Based on our result, the more cohesive group structure and less aggressive 418 interactions of Finnish P. fluviatilis make them the most suitable populations for 419

aquaculture. However, more populations need to be compared in order to identify
populations of interest across the species range. Moreover, we cannot exclude the future
potential impact of domestication since behavioural traits are modified by this process
(Kohane and Parsons 1988). Yet, taking into account behavioural intraspecific differentiation
would allow starting domestication program on populations presenting the best behavioural
pre-disposition.

Selecting best populations for aquaculture production cannot be made through only 426 427 behavioural trait comparisons. Indeed, selective breeding for low stress responsiveness has for instance been applied in several fish species but these low-stress response fish were also 428 429 the ones which were more aggressive (Huntingford et al. 2012). Intraspecific differentiation has been already assessed for several other traits of interest such as growth (e.g. Mandiki et 430 431 al. 2004; Leithner and Wanzenböck 2015), feed conversion efficiency (e.g. Imsland et al. 432 2000; Jonassen et al. 2000), or disease resistance (e.g. Imsland et al. 2002; Overturf et al. 433 2003). Therefore, the choice of the founder population must then be based on a multifunction and multi-traits approach rather than a single-trait decision framework. 434

435

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447 CONFLICT OF INTEREST AND ETHICAL STATEMENT

- 448 Conflict of Interest: The authors declare that they have no conflict of interest.
- 449 Ethical approval: "All applicable international, national, and/or institutional guidelines for
- 450 the care and use of animals were followed by the authors."

451 **ABBREVIATIONS**

- 452 GEN: Lake Geneva
- 453 VAL: Lake Valkea-Müstajärvi
- 454 ISO: Lake Iso-Valkjärvi
- 455 RAS: Recirculated Aquaculture System

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458

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APPENDIX





465 **REFERENCES**

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