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Short-term climate-induced change in French plant communities

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#### Abstract

Latitudinal and altitudinal range shifts in response to climate change have been reported for numerous animal species, especially those with high dispersal capacities. In plants, the impact of climate change on species distribution or community composition has been documented mainly over long periods (decades) and in specific habitats, often forests. Here, we broaden the results of such long-term, focused studies by examining climate-driven changes in plant community composition over a large area (France) encompassing multiple habitat types and over a short period (2009-2017). To this end, we measured mean community thermal preference, calculated as the community-weighted mean of the Ellenberg temperature indicator value, using data from a standardized participatory monitoring scheme. We report a rapid increase in the mean thermal preference of plant communities at national and regional scales, which we relate to climate change. This reshuffling of plant community composition corresponds to a relative increase in the abundance of warm- vs. cold-adapted species. However, support for this trend was weaker when considering only the common species, including common annuals. Our results thus suggest for the first time that the response of plant communities to climate change involves subtle changes affecting all species rare and common, which can nonetheless be detected over short time periods. Whether such changes are sufficient to cope with the current climate warming remains to be ascertained.


## KEYWORDS

Plants, monitoring, mean community thermal preference, climate change, lifespan, citizen science.

## INTRODUCTION

Climate change is considered a significant global threat to biodiversity [1-3], with impacts from individuals to ecosystems. In ecological communities, climate change drives compositional change, in particular via latitudinal and altitudinal shifts, especially in species with high dispersal capacities, including birds, insects or marine invertebrates [3-5]. In plants, which are less mobile, elevation has shifted upward during the $20^{\text {th }}$ century, e.g. in highland French temperate forests [6], although these shifts lag temperature change. In contrast, lowland forest plant communities have not shifted northward, highlighting their greater inertia [7] or higher shading mitigating climate warming [8]. Whether and how plant communities are able to track climate change in all habitats is still unclear.

Owing to the scarcity of available data, most studies demonstrating an effect of climate change on plant communities used heterogeneous data collected over decades and focused on a particular habitat. To understand better how climate change and other drivers influence biodiversity, real-time monitoring data using standardized schemes are required across large spatio-temporal scales [9]. Citizen science, the involvement of volunteers in research [10], represents a powerful tool that can contribute to investigating the impacts of climate change on biodiversity [4,11]. Such monitoring makes it possible to report subtle responses to climate change.

In this study, we examine nine years of change in plant community composition across France, using data collected through a participatory monitoring scheme, to assess whether such data can be used to detect short-term effects of climate warming beyond alpine or forest areas. The systematic sampling of monitoring sites particularly encompasses numerous lowland open habitats, which have received limited attention in studies of the response of plant communities to climate change. We address the following questions: (1) has the mean plant community thermal preference changed over the past nine years? (2) Can this change be
related to climatic variables? (3) Are the temporal changes in mean community thermal preference and in abundances of individual species related to species lifespan?

## MATERIALS AND METHODS

Monitoring data
We used data from Vigie-flore, a French citizen science program monitoring wild flora. The data were collected yearly by 321 skilled amateur botanists between 2009 and 2017 in 586 1 kmx 1 km squares sampled from a systematic grid (one square every 10 km ), which ensures representative sampling of habitats. Each square contains eight systematically distributed 10 $\mathrm{m}^{2}$ plots (Figure S1) divided into ten $1 \mathrm{~m}^{2}$-quadrats. In each quadrat, the presence of all vascular plants was recorded, as well as habitat type following the CORINE biotope nomenclature, a European hierarchical classification of vegetation types [12]. Most squares were in open habitats ( $27 \%$ artificial land cover, $29 \%$ farmland, $16 \%$ meadows vs. $22 \%$ in forests, Electronic Supplementary Material I). For each plot, the number of quadrats in which a species was observed provides a proxy for species abundance. Individual plots were recorded at different intensities over time, depending on the behaviour of recorders, such that on average there were 2.9 years of observation per plot, and 5.3 plots sampled per $1 \mathrm{~km}^{2}$ square, for a total of 3,118 plots.

## Species characteristics

We collected information for two species-specific attributes most likely to be involved in plant species response to climate change: (1) the Ellenberg temperature indicator value (hereafter ETIV), characterizing the optimum temperature class for growth and survival of a species (nine classes), and (2) lifespan (Electronic Supplementary Material II and Table S1). ETIV and lifespan were available for 1,709 and 1,780 out of the 2,428 species sampled,
representing $85.6 \%$ and $89.3 \%$ of total observations, respectively. We calculated the community-weighted Mean Thermal Preference (hereafter MTP), i.e. the abundance-weighted sum of ETIV of all species [13]. We also calculated MTP with presence/absence data instead of abundance, to test whether temporal trends could be detected with lower resolution data [14]. Finally, we calculated MTP within two extreme lifespan classes: annuals vs. perennials. The survey sites are not distributed homogeneously in space and time, which may cause spurious temporal trends in MTP due to a spatial displacement of sampled sites. To control for this, we performed the same analysis at a regional level, in Île-de-France, the region with the highest site density (188 squares sampled throughout the nine years, Figure S1).

## Climate data

For each plot, we collected daily precipitation sum, mean, minimum and maximum temperature between 2009 and 2017, from which we calculated 10 climatic variables per plot. Two variables used annual data: mean annual temperature and annual temperature anomaly; while the remaining eight used information from January to May (the growing period of most plant species): mean temperature, temperature anomaly, mean maximum temperature, number of heatwave days $\left(\mathrm{T}^{\circ} \mathrm{C}>27^{\circ} \mathrm{C}\right)$, freezing days $\left(\mathrm{T}^{\circ} \mathrm{C}<0^{\circ} \mathrm{C}\right)$, rainy days, drought days, and pluviometry. We used a random forest analysis to extract the most important climatic variable explaining MTP, which was mean annual temperature (Electronic Supplementary Material III and Table S2).

## Statistical analyses

We first checked for a potential bias attributable to uneven sampling effort by verifying that there was no temporal trend in the average latitude of sampled sites. We then used Bayesian hierarchical models to estimate the temporal trend in MTP (1) at a national scale and within

Île-de-France, (2) using either abundance data or presence/absence data and (3) in communities where only annuals vs. only perennials species were retained (Electronic Supplementary Material IV). Note that this approach captures changes in the relative abundance of species with fixed thermal preference rather than changes in species-specific adaptation to new climatic conditions. We then examined the relationship between (1) the temporal change in MTP in a plot between 2009 and 2017, and (2) the temporal change in mean annual temperature over the same period, using Spearman's correlation coefficient. To further investigate the role of species lifespan in the community-level trends, we used Bayesian models to estimate the temporal trend in abundance for the 550 most common species ( 75 annuals and 349 perennials) observed at least four out of nine years and in at least ten squares (Electronic Supplementary Material IV). We tested whether species trends were related to lifespan (annuals vs. perennials), to ETIV, or to the interaction of both variables by a linear regression in which each species was weighted by the inverse of the standard error associated with the estimated temporal trend. All analyses were performed using R [15] (Electronic Supplementary Material V).

## RESULTS

At the community level, MTP increased through time at the national scale, both with abundance and presence/absence data (Figure 1a,c), particularly in northern France (Electronic Supplementary Material IV). This increase was also observed within Île-de-France (Figures 1c and S2). The increase in MTP of plant communities occurred simultaneously with an increase in mean annual temperature (Electronic Supplementary Material III, Table S3, Figure S3), such that there was a significant positive but weak correlation between MTP change and the mean annual temperature change over time (Figure 1d): plots in which the temperature has increased tended to be also plots for which MTP has increased. Moreover, we
found a significant increase in MTP of annual, but not of perennial species, at national and regional scales (Figure 1b,c).

Changes at the community level were weakly explained by differences in temporal trends for common species considered independently. We found a significant relationship between species temporal trends and lifespan (Estimate $=0.194, n=424, \mathrm{df}=1.402, F=3.978, P=0.047$, Figure 2). The interaction between ETIV and lifespan was marginally significant (Estimate=$0.028, n=424, \mathrm{df}=1.402, F=3.013, P=0.083$ ), but no relation was found between temporal trends and ETIV.

## DISCUSSION

Our work complements longer-term, single habitat studies by documenting a small but steady increase in MTP of plant communities over all habitat types at national and regional scales, detectable over nine years only, and most likely driven by temperature change. This increase is measurable even with presence/absence data, suggesting a true reshuffling of plant communities, as opposed to mere abundance fluctuations of common species. In contrast to Bertrand and collaborators [7], who observed no response of lowland forest plant communities to climate change over a longer period, we show with a dataset encompassing mostly low-altitude sites that lowland plant communities can be modified by climate change, despite their limited dispersal capabilities. The difference may arise because our dataset contained a high proportion of open habitats, in which plant communities may experience higher turnover than in forests [8]. MTP changes resulted from local increases in the relative abundance of warm- vs. cold-adapted species, which, over nine years, are detectable in annuals but not perennials. Yet, examination of the temporal trends of the 424 most common annual or perennial species did not reveal this pattern. Non-congruence of the community vs.
species-level analysis, the latter covering only $55 \%$ of the observations, suggests that all species, including rarer species, are affected and calls for a more thorough analysis of species traits likely to drive the observed changes.

The general trend of increased relative abundance of warm- vs. cold-adapted species is consistent with longer-term studies documenting responses to climate change via species latitudinal and altitudinal range shifts, or local changes in abundance (including local decline and extinction of cold-adapted species) [6-8,16,17]. Yet, the causal role of climate change in the observed increase in the MTP of plant communities still has to be confirmed, for example by checking that this increase is sustained in the near future, or by examining possible confounding drivers, such as eutrophication or urbanization. Urbanization is unlikely to be responsible for increased MTP, because only a quarter of plots are located in urban areas. Although the Ellenberg nitrogen indicator value is weakly but positively correlated with ETIV, the mean community nitrogen preference or the abundance of nitrophilous species did not increase over the same period (Electronic Supplementary Material VI and Figure S4). Therefore, eutrophication should have a limited role in the observed increase in MTP.

The detection of this subtle but rapid change in plant communities was made possible by citizen science monitoring and a large sampling effort in space and time. Such data can be of great value [18] but come with several caveats deserving further attention. For example, possible biases could be related to uneven sampling effort, identification errors or uncomplete trait data. The effect of spatial heterogeneity in the sampling effort was tested by examining trends within a small region. Identification errors and missing trait data are unlikely to be systematically biased in favour of high or low species thermal preference, such that these problems should not influence our results. Finally, the observed trend in MTP could be caused
partly by temporal changes in species phenology. Because the Vigie-flore protocol recommends a single annual survey, there may be a phenology-influenced bias in species detection and identifiability against undeveloped or non-flowered species. Climate change has led to phenological shifts in flowering plants [19], which could increase the detectability of warm-adapted species as the climate warms.

Whether such rapid community changes are sufficient a response to current climate warming, or instead reflect mostly a maladaptive loss of cold-adapted species, remains to be ascertained. Future research efforts could also be focused on analysing trends in MTP across habitats or protection regimes and on identifying possible consequences of changes for biotic homogenization or plant-pollinator interactions.

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(a)

(c)

(b)

(d)


Figure 1. National temporal trend in MTP of plant communities from 2009 to 2017 calculated with abundance data, (a) for all species or (b) annual species only. The blue line is a generalized additive model estimate of the temporal trend, with its associated standard error (blue band). The black dots and error bars correspond to the inter-annual variations of the variable, with its associated standard error. (c) Temporal trends in MTP from 2009 to 2017
estimated using Bayesian models. N/R: National/Regional scale, Ab/PA:
Abundance/Presence-Absence data. Dots indicate the mean of the posterior distribution, and bars correspond to the 95 \% credible interval for the temporal trends. (d) Correlation between the plant MTP change and the annual mean temperature change over time. Each point represents one plot.


Figure 2. Relationship between the temporal trends of the most common species and their ETIV for annuals (orange) and perennials (blue). Each dot represents the temporal trend of a species, sized proportionally to the inverse of the standard error. Lines correspond to the regressions (with slopes not significantly different from zero in both cases), with their associated standard error (grey band).

ELECTRONIC SUPPLEMENTARY MATERIALS

Short-term climate-induced change in French plant communities

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## I. Habitat of sampled plots

A total of 3,118 plots were sampled in 586 1kmx1km squares between 2009 and 2017 as part of the Vigie-flore program. Habitat type following the CORINE biotope nomenclature [1] was reported by observers for 2,985 plots (ca. $96 \%$ of plots). These plots were distributed among the first levels of the CORINE biotope classification as follows: 21 plots in coastal habitats, 13 plots in wetland, 488 plots in grassland and shrub, 682 plots in woodland, 24 plots in marsh and bog, 36 plots in rocky, habitats, 895 plots in agricultural habitats and 826 plots in artificial land use.

## II. Species attributes

We collected information for two species-specific attributes: (1) the Ellenberg temperature indicator values (hereafter ETIV), characterizing the optimum temperature class for growth and survival of a species (nine classes), from the database Baseflor [2], and (2) lifespan, described using seven broad classes obtained by combining classes from the Biolflor [3] and LEDA [4] databases (Table S1).

Table S1. Procedure used to combine lifespan information from Biolflor and LEDA traitbases.

| Lifespan <br> (Biolflor database) | Lifespan (LEDA database) | Single or combined <br> lifespan categories | Number of <br> species |
| :--- | :--- | :--- | :---: |
| a (annual) | Annuals | annual | 17 |
| a (annual) - b (biennial) | Annuals | annual/biennial | 1 |
| b (biennial) | Annuals | annual/biennial | 1 |
| p (perennial-pollakanthic) | Annuals | annual/perennial | 1 |
| a (annual) | annuals + perennials | annual/perennial | 1 |
| p (perennial-pollakanthic) | annuals + perennials | 1 |  |
| p (perennial-pollakanthic) | annuals + perennials + poly-annuals < <br> years (short-lived perennials) | annual/perennial | 2 |
| a (annual) - b (biennial) <br> (perennial-pollakanthic) | annuals + perennials + strict monocarpic <br> bi-annuals and poly-annuals | annual/biennial/perennial | 1 |
| a (annual) - p (perennial- <br> pollakanthic) | annuals + perennials + strict monocarpic <br> bi-annuals and poly-annuals | annual/biennial/perennial | 2 |
| p (perennial-pollakanthic) | annuals + perennials + strict monocarpic <br> bi-annuals and poly-annuals | annual/biennial/perennial | 5 |


| p (perennial-pollakanthic) | annuals + perennials + strict monocarpic bi-annuals and poly-annuals + polyannuals $<5$ years (short-lived perennials) | annual/biennial/perennial | 1 |
| :---: | :---: | :---: | :---: |
| p (perennial-pollakanthic) | annuals + perennials + strict monocarpic bi-annuals and poly-annuals + polyannuals < 5 years (short-lived perennials) + poly-annuals 5-50 years (medium-lived perennials) | annual/biennial/perennial | 1 |
| p (perennial-pollakanthic) | annuals + perennials + strict monocarpic bi-annuals and poly-annuals + polyannuals 5-50 years (medium-lived perennials) | annual/biennial/perennial | 2 |
| a (annual) | annuals + perennials + summer annuals | annual/perennial | 2 |
| a (annual) | annuals + perennials + summer annuals + strict monocarpic bi-annuals and polyannuals | annual/biennial/perennial | 1 |
| a (annual) - b (biennial) | annuals + perennials + summer annuals + strict monocarpic bi-annuals and polyannuals | annual/biennial/perennial | 3 |
| a (annual) - p (perennialpollakanthic) | annuals + perennials + summer annuals + strict monocarpic bi-annuals and polyannuals | annual/biennial/perennial | 1 |
| b (biennial) | annuals + perennials + summer annuals + strict monocarpic bi-annuals and polyannuals | annual/biennial/perennial | 1 |
| b (biennial) - h (perennialhapaxanthic) | annuals + perennials + summer annuals + strict monocarpic bi-annuals and polyannuals + poly-annuals $<5$ years (shortlived perennials) | annual/biennial/perennial | 1 |
| a (annual) | annuals + perennials + summer annuals + winter annuals | annual/perennial | 2 |
| a (annual) | annuals + perennials + summer annuals + winter annuals + strict monocarpic biannuals and poly-annuals | annual/biennial/perennial | 3 |
| a (annual) - b (biennial) | annuals + perennials + summer annuals + winter annuals + strict monocarpic biannuals and poly-annuals | annual/biennial/perennial | 1 |
| a (annual) | annuals + perennials + summer annuals + winter annuals + strict monocarpic biannuals and poly-annuals + poly-annuals 5 years (short-lived perennials) | annual/biennial/perennial | 1 |
| a (annual) - p (perennialpollakanthic) | annuals + perennials + summer annuals + winter annuals + strict monocarpic biannuals and poly-annuals + poly-annuals 5 years (short-lived perennials) | annual/biennial/perennial | 1 |
| $\begin{aligned} & \mathrm{a} \text { (annual) - b (biennial) - } \mathrm{h} \\ & \text { (perennial-hapaxanthic) } \end{aligned}$ | annuals + perennials + winter annuals | annual/biennial/perennial | 1 |
| a (annual) | annuals + perennials + winter annuals + strict monocarpic bi-annuals and polyannuals | annual/biennial/perennial | 1 |
| p (perennial-pollakanthic) | annuals + poly-annuals > 50 years (longlived perennials) | annual/perennial | 1 |
| a (annual) | annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 16 |
| a (annual) - b (biennial) | annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 4 |
| $\begin{aligned} & \mathrm{a} \text { (annual) - b (biennial) - p } \\ & \text { (perennial-pollakanthic) } \end{aligned}$ | pannuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 1 |
| b (biennial) | annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 3 |


| b (biennial) - h (perennialhapaxanthic) | annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 1 |
| :---: | :---: | :---: | :---: |
| a (annual) | annuals + summer annuals | annual | 34 |
| a (annual) | annuals + summer annuals + poly-annuals $>50$ years (long-lived perennials) | annual/perennial | 1 |
| a (annual) | annuals + summer annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 10 |
| a (annual) - b (biennial) | annuals + summer annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 2 |
| a (annual) | annuals + summer annuals + strict monocarpic bi-annuals and poly-annuals + poly-annuals 5-50 years (medium-lived perennials) | annual/biennial/perennial | 1 |
| a (annual) | annuals + summer annuals + winter annuals | annual | 5 |
| a (annual) | annuals + summer annuals + winter <br> annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 18 |
| a (annual) - b (biennial) | annuals + summer annuals + winter <br> annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 1 |
| a (annual) | annuals + summer annuals + winter annuals + strict monocarpic bi-annuals and poly-annuals + poly-annuals 5-50 years (medium-lived perennials) | annual/biennial/perennial | 1 |
| a (annual) | annuals + winter annuals | annual | 2 |
| a (annual) | annuals + winter annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 11 |
| a (annual) - b (biennial) | annuals + winter annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 2 |
| a (annual) | perennials | annual/perennial | 2 |
| a (annual) - p (perennial- pollakanthic) | perennials | annual/perennial | 2 |
| b (biennial) - p (perennialpollakanthic) | perennials | biennial/perennial | 1 |
| h (perennial-hapaxanthic) | perennials | perennial | 1 |
| p (perennial-pollakanthic) | perennials | perennial | 396 |
| p (perennial-pollakanthic) | perennials + poly-annuals < 5 years (shortlived perennials) | perennial | 28 |
| p (perennial-pollakanthic) | perennials + poly-annuals $<5$ years (shortlived perennials) + poly-annuals > 50 years (long-lived perennials) | perennial | 1 |
| p (perennial-pollakanthic) | perennials + poly-annuals $<5$ years (shortlived perennials) + poly-annuals $>50$ years (long-lived perennials) + polyannuals 5-50 years (medium-lived perennials) | perennial | 3 |
| p (perennial-pollakanthic) | perennials + poly-annuals < 5 years (shortlived perennials) + poly-annuals 5-50 years (medium-lived perennials) | perennial | 26 |
| p (perennial-pollakanthic) | perennials + poly-annuals $>50$ years (long-lived perennials) | perennial | 11 |
| p (perennial-pollakanthic) | perennials + poly-annuals $>50$ years (long-lived perennials) + poly-annuals 550 years (medium-lived perennials) | perennial | 2 |
| a (annual) | perennials + poly-annuals 5-50 years (medium-lived perennials) | annual/perennial | 1 |
| a (annual) - p (perennial- pollakanthic) | perennials + poly-annuals 5-50 years (medium-lived perennials) | annual/perennial | 2 |
| p (perennial-pollakanthic) | perennials + poly-annuals 5-50 years | perennial | 84 |

$\left.\begin{array}{|l|l|l|c|}\hline & \begin{array}{l}\text { (medium-lived perennials) } \\ \text { a (annual) }\end{array} & \begin{array}{l}\text { perennials + strict monocarpic bi-annuals } \\ \text { and poly-annuals }\end{array} & \text { annual/biennial/perennial }\end{array}\right] 1$

|  | perennials) |  |  |
| :---: | :---: | :---: | :---: |
| a (annual) | perennials + summer annuals + winter annuals | annual/perennial | 1 |
| b (biennial) - p (perennialpollakanthic) | perennials + summer annuals + winter annuals | annual/biennial/perennial | 1 |
| b (biennial) - h (perennialhapaxanthic) | perennials + summer annuals + winter annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 1 |
| p (perennial-pollakanthic) | perennials + winter annuals | annual/perennial | 1 |
| p (perennial-pollakanthic) | perennials + winter annuals + poly-annuals $>50$ years (long-lived perennials) | annual/perennial | 1 |
| a (annual) - b (biennial) | perennials + winter annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 1 |
| b (biennial) | perennials + winter annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 1 |
| b (biennial) - h (perennialhapaxanthic) | perennials + winter annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 1 |
| a (annual) - b (biennial) | poly-annuals $<5$ years (short-lived perennials) | annual/biennial/perennial | 1 |
| b (biennial) | poly-annuals < 5 years (short-lived perennials) | biennial/perennial | 1 |
| b (biennial) - p (perennialpollakanthic) | poly-annuals < 5 years (short-lived perennials) | biennial/perennial | 1 |
| p (perennial-pollakanthic) | poly-annuals < 5 years (short-lived perennials) | perennial | 8 |
| p (perennial-pollakanthic) | poly-annuals < 5 years (short-lived perennials) + poly-annuals 5-50 years (medium-lived perennials) | perennial | 5 |
| p (perennial-pollakanthic) | poly-annuals > 50 years (long-lived perennials) | perennial | 3 |
| p (perennial-pollakanthic) | poly-annuals $>50$ years (long-lived perennials) + poly-annuals 5-50 years (medium-lived perennials) | perennial | 2 |
| $\begin{aligned} & \text { a (annual) - p (perennial- } \\ & \text { pollakanthic) } \end{aligned}$ | poly-annuals 5-50 years (medium-lived perennials) | annual/perennial | 1 |
| b (biennial) - p (perennialpollakanthic) | poly-annuals 5-50 years (medium-lived perennials) | biennial/perennial | 1 |
| p (perennial-pollakanthic) | poly-annuals 5-50 years (medium-lived perennials) | perennial | 32 |
| a (annual) | strict monocarpic bi-annuals and polyannuals | annual/biennial/perennial | 2 |
| a (annual) - b (biennial) | strict monocarpic bi-annuals and polyannuals | annual/biennial/perennial | 1 |
| b (biennial) | strict monocarpic bi-annuals and polyannuals | biennial/perennial | 14 |
| b (biennial) - h (perennial- hapaxanthic) | strict monocarpic bi-annuals and polyannuals | biennial/perennial | 5 |
| b (biennial) - p (perennialpollakanthic) | strict monocarpic bi-annuals and polyannuals | biennial/perennial | 1 |
| p (perennial-pollakanthic) | strict monocarpic bi-annuals and polyannuals | biennial/perennial | 1 |
| a (annual) - b (biennial) -(perennial-hapaxanthic) | strict monocarpic bi-annuals and polyannuals + poly-annuals $<5$ years (shortlived perennials) | annual/biennial/perennial | 1 |
| b (biennial) - h (perennialhapaxanthic) | strict monocarpic bi-annuals and polyannuals + poly-annuals $<5$ years (shortlived perennials) | biennial/perennial | 1 |
| p (perennial-pollakanthic) | strict monocarpic bi-annuals and polyannuals + poly-annuals $<5$ years (shortlived perennials) | biennial/perennial | 1 |


| a (annual) | summer annuals | annual | 65 |
| :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { a (annual) - b (biennial) - p } \\ & \text { (perennial-pollakanthic) } \end{aligned}$ | summer annuals | annual/biennial/perennial | 1 |
| a (annual) - p (perennialpollakanthic) | summer annuals | annual/perennial | 1 |
| p (perennial-pollakanthic) | summer annuals | annual/perennial | 1 |
| a (annual) | summer annuals + poly-annuals 5-50 years (medium-lived perennials) | annual/perennial | 1 |
| a (annual) - p (perennial- pollakanthic) | summer annuals + poly-annuals 5-50 years (medium-lived perennials) | annual/perennial | 1 |
| a (annual) | summer annuals + strict monocarpic biannuals and poly-annuals | annual/biennial/perennial | 9 |
| a (annual) - b (biennial) | summer annuals + strict monocarpic biannuals and poly-annuals | annual/biennial/perennial | 1 |
| $\begin{aligned} & \text { a (annual) - b (biennial) - h } \\ & \text { (perennial-hapaxanthic) } \end{aligned}$ | hsummer annuals + strict monocarpic biannuals and poly-annuals | annual/biennial/perennial | 1 |
| b (biennial) | summer annuals + strict monocarpic biannuals and poly-annuals | annual/biennial/perennial | 1 |
| b (biennial) - p (perennialpollakanthic) | summer annuals + strict monocarpic biannuals and poly-annuals | annual/biennial/perennial | 2 |
| a (annual) - p (perennialpollakanthic) | summer annuals + strict monocarpic biannuals and poly-annuals + poly-annuals 5-50 years (medium-lived perennials) | annual/biennial/perennial | 1 |
| a (annual) | summer annuals + winter annuals | annual | 19 |
| a (annual) - b (biennial) | summer annuals + winter annuals | annual/biennial | 1 |
| a (annual) - p (perennialpollakanthic) | summer annuals + winter annuals + polyannuals $<5$ years (short-lived perennials) | annual/perennial | 1 |
| a (annual) | summer annuals + winter annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 4 |
| a (annual) - b (biennial) | summer annuals + winter annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 1 |
| a (annual) - b (biennial) | summer annuals + winter annuals + strict monocarpic bi-annuals and poly-annuals + poly-annuals $<5$ years (short-lived perennials) | annual/biennial/perennial | 1 |
| a (annual) | winter annuals | annual | 6 |
| a (annual) - b (biennial) | winter annuals + strict monocarpic biannuals and poly-annuals | annual/biennial | 1 |
| a (annual) - p (perennialpollakanthic) | winter annuals + strict monocarpic biannuals and poly-annuals | annual/biennial/perennial | 1 |
| b (biennial) | winter annuals + strict monocarpic biannuals and poly-annuals | annual/biennial/perennial | 1 |
| NA | annuals | annual | 5 |
| NA | annuals + summer annuals | annual | 3 |
| NA | perennials | perennial | 57 |
| NA | perennials + strict monocarpic bi-annuals and poly-annuals | biennial/perennial | 1 |
| NA | perennials + summer annuals + strict monocarpic bi-annuals and poly-annuals + poly-annuals 5-50 years (medium-lived perennials) | annual/biennial/perennial | 1 |
| NA | poly-annuals 5-50 years (medium-lived perennials) | perennial | 3 |
| NA | strict monocarpic bi-annuals and polyannuals | biennial/perennial | 1 |
| NA | strict monocarpic bi-annuals and polyannuals + poly-annuals $<5$ years (shortlived perennials) | biennial/perennial | 1 |
| NA | summer annuals | annual | 26 |


| NA | summer annuals + strict monocarpic biannuals and poly-annuals | annual/biennial/perennial | 6 |
| :---: | :---: | :---: | :---: |
| NA | summer annuals + winter annuals | annual | 1 |
| NA | summer annuals + winter annuals + strict monocarpic bi-annuals and poly-annuals | annual/biennial/perennial | 1 |
| NA | winter annuals | annual | 1 |
| a (annual) | NA | annual | 72 |
| a (annual) - b (biennial) | NA | annual/biennial | 7 |
| $\begin{aligned} & \text { a (annual) - b (biennial) - h } \\ & \text { (perennial-hapaxanthic) } \end{aligned}$ | NA | annual/biennial/perennial | 1 |
| a (annual) - p (perennialpollakanthic) | NA | annual/perennial | 4 |
| b (biennial) | NA | biennial | 10 |
| b (biennial) - h (perennialhapaxanthic) | NA | biennial/perennial | 2 |
| b (biennial) - p (perennialpollakanthic) | NA | biennial/perennial | 3 |
| h (perennial-hapaxanthic) - p (perennialpollakanthic) | NA | perennial | 1 |
| p (perennial-pollakanthic) | NA | perennial | 284 |
| $\begin{aligned} & \text { p (perennial-pollakanthic) } \\ & \text { - p (perennial- } \\ & \text { pollakanthic) } \end{aligned}$ | NA | perennial | 1 |
| NA | NA | NA | 917 |

## III.Climate data and Random Forest analysis

For each plot, we collected daily mean temperature, daily minimum temperature, daily maximum temperature and daily precipitation sum between 2009 and 2017 from the ENSEMBLES dataset [5], using data from the nearest E-OBS station in France on a 0.25 degree grid. Climate data were extracted using the R package climateExtract [6], and devtools [7]. We then calculated 10 climatic variables per plot, two using annual data (mean annual temperature and annual temperature anomaly) and eight using information from January to May (growing period of many plants): mean temperature, temperature anomaly, mean maximum temperature, number of heatwave days $\left(\mathrm{T}^{\circ} \mathrm{C}>27^{\circ} \mathrm{C}\right)$, freezing days $\left(\mathrm{T}^{\circ} \mathrm{C}<0^{\circ} \mathrm{C}\right)$, rainy days, drought days, and pluviometry. The temperature anomaly is the sum of the differences between the observed daily temperature and the mean temperature over a period on a site.

We then performed a Random Forest regression analysis to identify the climatic variable(s) explaining the greatest variability in MTP of plant communities. The Random Forest analysis is a nonparametric technique derived from classification and regression trees. This method, which has been applied in various ecological studies [8,9], shows high accuracy and the ability to model complex interactions between variables. To perform the model, we defined two essential parameters: mtry, i.e. the number of variables to try at each split, and ntree, i.e. the number of trees to run. To maximize the accuracy of the Random Forest analysis, i.e. minimize the measure of the prediction error (Out Of Bag (OOB) error), mtry and ntree can be varied. We determined the parameter mtry thanks to the internal RF function TuneRF [R package randomForest, 10]; this function computes the optimal number of variables starting from the default (total number of variables/3 for regression) and tests half and twice this threshold for the value with the minimum OOB error rate. It is already known that a mtry of 1 can produce good accuracy [10,11], but it is recommended to include at least two variables to avoid using also the weaker regressors as splitters [12]. Here, even if the best accuracy was obtained for a mtry of 1, we performed the Random Forest regression analysis with a mtry of 2 following Grömping [12]. The ntree parameter was set to 1000 to obtain stable results. Variable importance of the 10 climatic variables (i.e. the mean annual temperature, the annual temperature anomaly, and eight variables estimated from January to May: mean temperature, temperature anomaly, mean maximum temperature, number of heatwave days, freezing days, rainy days, drought days, and pluviometry) was estimated as the total decrease in node impurities, measured by residual sum of squares from splitting on the variable, averaged over all trees.

Finally, as a result of the Random Forest regression analysis, we selected the annual mean temperature, the most relevant climatic variable by ranking the variables according to their
importance measure (Table S2).
We performed a linear regression of the annual mean temperature against years to describe the temporal trend of climate change on the sampled plots of the study. We included geographical coordinates in the model to account for residual spatial autocorrelation. We recovered one slope per square, since climate data were only available at that scale; all plots within a scale were thus attributed the same temporal trend in climate data, describing the temporal trend of annual mean temperature.

Table S2. Importance measures of the 10 climatic variables resulting from a Random Forest regression analysis and explaining the share of variability of the MTP of plant communities.

| Climatic variables | Importance |
| :--- | :---: |
| Annual mean temperature | 725 |
| Number of freezing day from January to May | 365 |
| Mean temperature from January to May | 342 |
| Mean maximum temperature from January to May | 243 |
| Number of drought day from January to May | 204 |
| Pluviometry from January to May | 160 |
| Annual temperature anomaly | 135 |
| Temperature anomaly from January to May | 133 |
| Number of rainy day from January to May | 99 |
| Number of heatwave day from January to May | 55 |

## IV.Estimation of temporal trends in mean temperature preference and species

 occurrenceWe performed Bayesian hierarchical models to estimate the temporal trend in MTP of plant communities (1) at a national scale and within Île-de-France, (2) using either abundance data or presence/absence data and (3) of annuals vs. perennials species. Note that MTP is comparable to the Community Temperature Index [13], but is not a direct measure of mean
temperature, because ETIV consists in classes of temperature preference. The underlying model describes the MTP value as a function of year for all plant communities. We assumed that MTP of plant communities was normally distributed, with a mean that depended linearly on year, geographical coordinates and survey square, and constant variance. The linear model was the following:

MTP $_{i j k}=m+a^{*}$ year $_{i}+b^{*}$ lat $_{i j k}+c^{*}$ long $_{i j k}+d^{*}$ lat $_{i j k}{ }^{2}+e^{*}$ long $_{i j k^{2}}+$ Square $_{j}+\sigma_{i j k}$
where $M T P_{i j k}$ is the mean temperature preference of the plant community in plot $k$ of square $j$ and year $i, m$ is the intercept, lat and long are the latitude and longitude of the plot, Square is the random effect of square with variance $\sigma_{s}^{2}$ and $\sigma_{i j k}^{2}$ is the residual variance. We chose to model non-linear spatial patterns in MTP squared latitude and longitude terms over more complex models, because on visual examination of the relationship between MTP and latitude or longitude, we expected mostly a linear effect, potentially with a plateau at high or low latitudes that a second order polynomial model correctly represents.

Note that we performed a linear regression to test for an interaction between latitude and temporal trend of the MTP in order to detect spatial differences at the national scale (not detailed here). At the national scale, the temporal trend increased significantly with latitude, i.e. the trend was significantly larger in northern vs. southern France (significant latitude:year interaction, not shown).

We also performed Bayesian hierarchical models to generate estimates of the temporal trend in common species abundance, as estimated at the $10 \mathrm{~m}^{2}$ plot level by the number of $1 \mathrm{~m}^{2}$ quadrats in which these species were observed. We selected the 550 most common species that were observed at least four out of nine years and in at least ten squares. We modelled species abundance assuming a binomial distribution with $n=10$ trials $^{1}$ and a logittransformed probability $p$ depending linearly on year, survey plot, square and geographical

[^0]coordinates as follows:
$\operatorname{logit}\left(p_{i j k}\right)=m+a^{*}$ year $_{i}+b^{*}$ lat $_{i j k}+c^{*}$ long $_{i j k}+d^{*}$ latijk $^{2}+e^{*}$ long $_{i j k^{2}}+$ Square $_{j}+$ Plot $_{k}$ where $p_{i j k}$ is the probability to observe the species in each $1 \mathrm{~m}^{2}$ square of plot $k$ of square $j$ and year $i, m$ is the intercept, lat and long are the latitude and longitude of the plot, Square is the random effect of square with variance $\sigma_{s}^{2}$, Plot is the random effect of plot with variance $\sigma^{2} p$. For all Bayesian hierarchical models, we chose only non-informative priors, sampled from Gaussian distributions with a mean of 0 and a variance of $10^{10}$ for most parameters, and gamma distributions with shape parameter $k=0.001$ and rate $\beta=0.001$ for precision (inverse of variance) parameters. Parameters were estimated by fitting the model in a Bayesian mode of inference using JAGS [14], through R [15] using the package R2jags [16]. Four chains were run using priors to draw 100,000 MCMC iterations, discarding the first 50,000 iterations. Samples were thinned by 1 in 10 to reduce autocorrelation and obtain mean trends from the remaining posterior distributions. Convergence was checked using the Rubin statistics (Rhat), comparing within-chain variance to between-chain variance [17] (all Rhat < 1.2) and visual inspection of the traceplots (superposition of the 4 chains) produced through the R package mcmcplots [18]. Posterior mean community and species trends were interpreted as different from zero when $95 \%$ of the highest posterior density interval did not include zero.

## V. R packages used

All analyses were performed using R version 3.2.4 [15] and R packages Taxonstand version 2.1 [19], R packages $T R 8$ [20], R packages climateExtract version 0.2 .0 [6], R packages randomForest [10], R packages caret version 6.0-78 [21], R packages plotmo version 3.3.4 [22], R packages R2jags version 0.5-7 [16], R packages mcmcplots version 0.4 .2 [18], R packages car [23], R packages lme4 [24], R packages nlme version 3.1-131 [25], R packages
dplyr version 0.7.3 [26], R packages reshape2 [27], R packages devtools version 1.13.4 [7], R packages effects version 2.1.0 [28], R packages ggplot2 version 2.1.0 [29], R packages ggthemes version 3.4.0 [30], R packages gridExtra version 2.3 [31]. Plant functional traits were collected using the R package TR8 [20].

## VI. Relationship between Ellenberg Temperature and Ellenberg Nitrogen Indicators Values

Across the 1,473 species for which we could obtain Ellenberg indicator values (out of the 2,428 species of the Vigie-flore dataset), we observed that the Ellenberg temperature indicator value was positively correlated with the Ellenberg nitrogen indicator value (i.e. nitrogen preference, Pearson's correlation $=0.27$ ), although the correlation was not very strong. There is thus a possibility that some of the changes in plant community composition we observed are also driven by ecosystem eutrophication, in addition to climate warming. To rule out this possibility, we tested whether mean nitrogen preference changed over time in the plant communities we monitored. To this end, we calculated the Community Weighted-Mean Ellenberg nitrogen indicator value (hereafter ENIV for the species characteristic, and MNP for the community-level Mean Nitrogen Preference), in the same way as the Mean Temperature Preference. We used a Bayesian hierarchical model to estimate the temporal trend in MNP in plant communities. The linear model was similar to the one we used for Mean Temperature Preference:

MNPijk $=m+a^{*}$ year $_{i}+b^{*}$ lat $_{i j k}+c^{*}$ long $_{i j k}+d^{*}$ lat $_{i j k}{ }^{2}+e^{*}$ long $_{i j k}{ }^{2}+$ Square $_{j}+\sigma_{i j k}^{2}$ where $M N P_{i j k}$ is the mean nitrogen preference of the plant community in plot $k$ of square $j$ and year $i, m$ is the intercept, lat and long are the latitude and longitude of the plot, Square is the random effect of square with variance $\sigma_{\mathrm{s}}^{2}$ and $\sigma_{\mathrm{ijk}}^{2}$ is the residual variance.

The temporal trend in MNP was not different from zero (Figure S4, -0.0047, CI95\%=[-0.013, $0.0036]$ ), i.e. the mean Ellenberg value for nitrogen preference of plant communities has not changed significantly over the past nine years, which suggests that the increase in Mean Temperature Preference we observed is unlikely to be driven by the (weak) correlation between the thermal preference and the nitrophily of plant species. In other words, while we do observe an increase in the relative abundance of thermophilic species (increase in MTP), the relative abundance of nitrophilous species remains stable.

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Table S3. Type II Anova results for the linear mixed models testing
for a temporal change in the mean annual temperature over time.

| Fixed effects | $\boldsymbol{n}$ | Estimate | Standard error | Chisq | $\boldsymbol{P}$ value |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Year | 7854 | 0.1094 | 0.0032 | 1203.82 | $<0.001$ |
| Longitude | 7854 | 3.9921 | 0.7605 | 27.56 | $<0.001$ |
| Latitude | 7854 | -7.5612 | 1.9008 | 15.82 | $<0.001$ |
| Longitude*Longitude | 7854 | -0.0318 | 0.0096 | 10.89 | $<0.01$ |
| Latitude*Latitude | 7854 | 0.0819 | 0.0204 | 16.15 | $<0.001$ |
| Longitude*Latitude | 7854 | -0.0872 | 0.0158 | 30.30 | $<0.001$ |



Figure S1. Spatial distribution and hierarchical structure of the Vigie-flore sampling sites.
Plant species are sampled in 1 km X 1 km squares. In each square, plant species were recorded in several plots with a standardized protocol. On the map, square colours indicate the average (over all plots) temporal trend in Mean Temperature Preference (MTP) of plant communities on the study period. Six squares had average temporal trends in MTP of plant communities greater than $0.5(0.51,0.55,0.62$ and 1$)$ or below $-0.5(-0.67$ and -0.79$)$, they were capped at 0.5 and -0.5 respectively to improve readability. Plots and squares that were surveyed only once, for which no temporal trend could be calculated, are plotted as circles.


Figure S2. Regional temporal trend in MTP of plant communities from 2009 to 2017 calculated with abundance data in Île-de-France. The blue line is a generalized additive model estimate of the temporal trend, with its associated standard error (blue band). The black dots and error bars correspond to the inter-annual variations of the variable, with its associated standard error.


Figure S3. Temporal trend of annual mean temperature estimated from 2009 to 2017 in the sampling sites. The blue line is a generalized additive model estimate of the temporal trend, with its associated standard error (blue band). The black dots and error bars correspond to the inter-annual variations of the variable, with its associated standard error.


Figure S4. National temporal trend in the mean nitrogen preference of plant communities from 2009 to 2017 calculated with abundance data for all species. The blue line is a generalized additive model estimate of the temporal trend, with its associated standard error (blue band). The black dots and error bars correspond to the inter-annual variation of the variable, with its associated standard error.

List S1. The 321 Vigie-flore volunteers involved in the wild flora monitoring:
Abadie Jean-Claude, Abadie Juliet, Abraham Christelle, Aguettant Laurent, Allemand Denis, Anglaret Cédric, Arnould Loic, Arondel Bruno, Association Asterella, Association Les Amis Du Jardin Botanique Littoral Paul Jovet, Association Les Marais Du Verdier, Association Tela Botanica, Aupoix Alain, Autret Françoise, Avenas Élise, Avril Damien, Balard Yvette, Balloux Gabriel, Baradat Mathilde, Barbet-Massin Morgane, Barreda Flavie, Bastin André, Batsere Francine, Baudet Nicolas, Baurens-Franc Christophe, Bebin Clément, Bégin Sophie, Benezet Ruddy, Béraud Jean-Marc, Berger Romain, Biancardini Famille, Bircker Loïc, Bodiguel Brigitte, Boisbineuf Christelle, Boissier Marc, Bonnafoux Florence, Bonnafoux Vincent, Bonneau Pierre, Bonneil Philippe, Bos Nathalie, Bosquet Christian, Bouard Hervé, Boudies Catherine, Bouffard Frédéric, Bouligand Sandrine, Boutegege Serge, Bouveris Nicole, Bouvier Chantal, Brajon Émilie, Brasseur Jean-Emmanuel, Breuil Anne, Briand Françoise, Brin Antoine, Buckvald Nathalie, Bultel Yvan, Burchianti Mireille, Busch Geneviève, Callewaert Pierre, Calme Catherine, Calmet Claire, Camenen-Houdayer Danièle, Canel Célie, Canevet Marie-Françoise, Caron Matthieu, Casenave Isabelle, Catieau Alexis, Causeur Nicole, Cellier Pierre, Chabanois Gérard, Chamaux Jérome, Chammard Émilie, ChappertBessiere Joceline, Chapuis Gabriel, Chardinal Sophie, Chatelain Marion, Cherrier Camille, Chevalier Adélie, Chevalier Richard, Chiffoleau Bertrand, Cholet Jérémie, Clochard Alain, Cocquempot Marine, Coiffard Clément, Collectivité Territoriale (Reserve Naturelle) Poupart, Conservatoire Botanique National Du Bassin Parisien, Coupiac Louis, Couric Pauline, Crouzet Pierre, Cuille Rosalie, Danet Pierre, De Chastenet Cedissia, De Kerimel Aymeric, De La Brosse Caroline, Delafoy Arielle, Delisle Pascal, Delphin Chantal, Deschatres Aurélien, Desvois Régine, Devers Florence, Dezerald Hervé, Dole Nicolas, Domenech Boris, Douard Anne, Douarre Marie, Druez Flora, Dubosc Sarah, Duffaut Philippe, Dumas Françoise, Dumez Nicolas, Durand Alexandre, Duron Quiterie, Duytschaever Francis, Engel Geneviève,

Eriksson Marion, Esnault Sarah, Etcheverry Georgette, Farçat Charlotte, Fatz Pierre-Marie, Féral Camille, Fernandes Val, Filleur Lucas, Fournel Isabelle, Fourvel Jean-Marc, Fried Guillaume, Galerme Jean Baptiste, Galerne Jean Pierre, Garnier-Boudier Joanne, Gatto Claire, Gaudry Jérôme, Gauthier Jacques, Gelin Muriel, Gelmetti Anne-Marie, Geneix Gilles, Georganidis Michèle, Germaine Clémence, Ghesquière Jean-Christophe, Ghestem Murielle, Goguet Liselotte, Gombault Clémentine, Gorremans Jean-Luc, Gotteland Pierre, Gourdel Laurence, Gourlin Benjamin, Grange Milena, Grignon Céline, Guais Daniel, Guichard Bernard, Guiller Clémence, Guittet Jean, Haeffner Hugues, Haenggi Philippe, Hance Emmanuel, Happe David, Heuze Sylvère, Hillion Jeanne, Hubert Eric, Hugot Nicolas, Huve Rosalia, Indorf Marc-Frédéric, Jacot Pierre-Alain, Jagoda Daniel, Jardin Botanique Besançon, Jargeat Patricia, Javaux Benjamin, Jaymond Delphine, Joset Henri, Jourdan Christine, Joyaux Sandrine, Jullin Franck, Kania Daniel, Laffargue Lucie, Lafon Céline, Lagier Anne-Marie, Lallemand Johann, Lambert Cécile, Lamouroux Paule, Lanciaux Maxime, Lantuejoul Eric, Larbouillat Annick, Lavialle Jeanne, Le Fur Patrick, Le Garrec Sandra, Leblanc Marie, Leclerc Catherine, Leduc Vinciane, Lefebvre Manon, Lekieffre Alix, Lemoine Yves, Lemouzy Claire, Leneveu Justine, Lenne Patricia, Lepage Elen, Lequeuvre Charly, Lerat Christine, Lerat Elodie, Lesné Sébastien, Leveslin Gérard, Leviol Isabelle, Lévy-Leonesio Patricia, Locqueville Jonathan, Lodziak Michel, Lonchamp Jean-Pierre, Longeot Yves, Luciani Bernadette, Luigi Catherine, Mace Bruno, Machon Nathalie, Magrou Nathalie, Maguet Nicolle, Marchal Pierre-Luc, Marsteau Christine, Martin Gabrielle, Martin Hilaire, Martin Nicolas, Martinelli Marion, Masonnave François, Massouh Jacques, Master Promo, Mathieu Daniel, Mathieu Joël, Maurel Noëlie, Meyer Claire-Lise, Michel Prisca, Mistarz Margaux, Mollereau Aurore, Monferrand Christophe, Monthus Alain, Montpied Pierre, Moragues Angels, Morelon Stéphanie, Morin Delphine, Morlon Claire, Motard Éric, Muratet Audrey, Mutuel Arnaud, Naudin Marie-Noëlle, Nesty Matthieu, Nettleship-Begin Pierre, Neveu Dany,

Nonglaton Gilles, Noraz Éloise, Onf Bergerie, Ostermann Lukas, Oszczak Joëlle, Oudot Le Secq Marie-Pierre, Parachout Marie, Pascal Marie-Élise, Pelatan François, Pellecuer Hélène, Perret Jeannette, Peyrard Morgan, Picard Francis, Pichard Adeline, Pirot Manuel, Plessier Claudine, Politi Bertoncini Alzira, Ponchant Monique, Pons Virginie, Porcher Pascale, Quénaon Christine, Ravot Charlotte, Raynaud Xavier, Réal Juliette, Régentête Christine, Rémaud Dominique, Rémaud Marie-Bernadette, Renault Gérard, Réserve Naturelle Nationale Saint-Mesmin, Ressayre Patrick, Revault Pascal, Riberol Lionel, Rist Delphine, Rivière Yann, Robert Jean-Louis, Rodrigues Mélanie, Ropars Lise, Rossier Dominique, Rouger Caroline, Roussel Dany, Rouveyrol Paul, Sahl Amandine, Sanson Karine, Sapaly Élise, Sartre Isabelle, Scheidegger Anne, Serret Patrick, Signoret Sandrine, Silberfeld Thomas, Sonneville Hélène, Soulas Patrick, Spicher Fabien, Stratmains Emmanuel, Tasso Dominique, Tauran Julien, Texier-Pauton Élodie, Thébault Ludovic, Thiebaut David, Thys Évelyne, Tinchant Anne, Tinguy Hugues, Torre Fabrice, Tositti Arnaud, Toty Anne, Truong Minh-Xuan, Turcati Laure, Tzvetan Gwladys Mai, Valette Nicolas, Van Panhuys-Sigler Michèle, Vérite Vincent, Viallon Camille, Vidal Rodolphe, Videlot Vincent, Vilard Vanessa, Villeneuve Aurélie, Virton-Lavorel Anthony, Wahl Marie-Salomé, Waligorski Stanislas, Wernert Dominique, Wolff Anne, Ybres Nicolas.


[^0]:    ${ }^{1}$ The 10 quadrats within a plot are always sampled; there are thus no missing data within a plot.

