

# **How good is your marine protected area at curbing threats?**

Mirta Zupan, Fabio Bulleri, Julian Evans, Simonetta Fraschetti, Paolo Guidetti, Antoni Garcia-Rubies, Marta Sostres, Valentina Asnaghi, Anthony Caro, Salud Deudero, et al.

# **To cite this version:**

Mirta Zupan, Fabio Bulleri, Julian Evans, Simonetta Fraschetti, Paolo Guidetti, et al.. How good is your marine protected area at curbing threats?. Biological Conservation, 2018, 221, pp.237–245. 10.1016/j.biocon.2018.03.013. hal-01826691

# **HAL Id: hal-01826691 <https://hal.science/hal-01826691v1>**

Submitted on 15 Apr 2021

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

# **How good is your marine protected area at curbing threats?**

Mirta Zupan<sup>a\*</sup>, Fabio Bulleri<sup>b</sup>, Julian Evans<sup>c</sup>, Simonetta Fraschetti<sup>d</sup>, Paolo Guidetti<sup>e</sup>, Antoni Garcia-Rubies<sup>f</sup>, Marta Sostres<sup>a</sup>, Valentina Asnaghi<sup>g</sup>, Anthony Caro<sup>a</sup>, Salud Deudero<sup>h</sup>, Raquel Goñi<sup>h</sup>, Giuseppe Guarnieri<sup>d</sup>, Francois Guilhaumon<sup>i</sup>, Diego Kersting <sup>j,k</sup>, Athina Kokkali<sup>l</sup>, Claudia Kruschel<sup>m</sup>, Vesna Macic<sup>n</sup>, Luisa Mangialajo<sup>o,p</sup>, Sandra Mallol<sup>h</sup>, Enrique Macpherson<sup>f</sup>, Antonella Panucci<sup>l</sup>, Mirko Radolovic<sup>q</sup>, Mohamed Ramdani<sup>s</sup>, Patrick J. Schembri<sup>c</sup>, Antonio Terlizzi<sup>tu</sup>, Elisa Villa<sup>v</sup>, Joachim Claudet<sup>a</sup>

# **Affiliations**

<sup>a</sup> National Center for Scientific Research, PSL Research University, CRIOBE, USR 3278 CNRS-EPHE-UPVD, 66860 Perpignan, France

<sup>b</sup> Dipartimento di Biologia, Università di Pisa, Via Derna 1, 56126, Pisa

c Department of Biology, University of Malta, Msida MSD2080, Malta

<sup>d</sup> Department of Biological and Environmental Sciences and Technologies, University of Salento, CoNISMa, 73100, Lecce, Italy

<sup>e</sup> CoNISMa (Interuniversity National Consortium of Marine Sciences, 00196 Rome, Italy)

f Centre d'Estudis Avançats de Blanes, CEAB-CSIC, C. acces Cala S. Francesc 14, 17300 Blanes, Spain

g Department for the Earth, Environment and Life Sciences, University of Genoa, CoNISMa, 16132, Genoa, Italy

h Instituto Español de Oceanografía - Centro Oceanográfico de Baleares. Muelle de Poniente s/n 07015 Palma de Mallorca, Spain.

<sup>i</sup> UMR 9190 MARBEC, IRD-CNRS-IFREMER-UM, Université de Montpellier, 34095 Montpellier, France

<sup>j</sup> Section Paleontology, Institute of Geological Sciences, Freie Universität Berlin, 12249 Berlin, Germany

<sup>k</sup> Departament de Biologia Evolutiva, Ecologia i Ciències Ambientals. Universitat de 08028 Barcelona, Spain

l Hellenic Centre for Marine Research, 46.7 km. Athens Sounio, PO Box 712 19013 Anavyssos, Attiki, Greece

m Department of Ecology, Agronomy and Aquaculture, University of Zadar, Trg kneza Višeslava 9, 23000 Zadar, Croatia

<sup>n</sup> Institute of marine biology, University of Montenegro, Dobrota b.b.; 85330 Kotor; Montenegro

o Université Côte d'Azur, CNRS, ECOMERS, Parc Valrose 28, 06108 Nice Cedex, France

<sup>p</sup> Sorbonne Universités, UPMC Univ Paris 06, CNRS, LOV, 06230 Villefranche-surmer, France

<sup>q</sup> Administrative department for sustainable development, Region of Istria, Flanatička 29, 52100 Pula, Croatia

<sup>s</sup> Institut scientifique de Rabat, Avenue Ibn Batouta, Rabat, Morocco

t Department of Life Sciences, University of Trieste, CoNISMa, 34127 Trieste, Italy u Stazione Zoologica Anton Dohrn, 80121, Napoli, Italy

v Microbia Environnement, Observatoire Océanologique, F-66651 Banyuls/mer, France

\*is corresponding author

# **ABSTRACT**

Marine protected areas (MPAs) are key tools to mitigate human impacts in coastal environments, promoting sustainable activities to conserve biodiversity. The designation of MPAs alone may not result in the lessening of some human threats, which is highly dependent on management goals and the related specific regulations that are adopted. Here, we develop and operationalize a local threat assessment framework. We develop indices to quantify the effectiveness of MPAs (or individual zones within MPAs in the case of multipleuse MPAs) in reducing anthropogenic extractive and non-extractive threats operating at local scale, focusing specifically on threats that can be managed through MPAs. We apply this framework in 15 Mediterranean MPAs to assess their threat reduction capacity. We show that fully protected areas effectively eliminate extractive activities, whereas the intensity of artisanal and recreational fishing within partially protected areas, paradoxically, is higher than that found outside MPAs, questioning their ability at reaching conservation targets. In addition, both fully and partially protected areas attract non-extractive activities that are potential threats. Overall, only three of the 15 MPAs had lower intensities for the entire set of eight threats considered, in respect to adjacent control unprotected areas. Understanding the intensity and occurrence of human threats operating at the local scale inside and around MPAs is important for assessing MPAs effectiveness in achieving the goals they have been designed for, informing management strategies, and prioritizing specific actions.

**Keywords**: fully protected area; partially protected area; management; extractive activities; non-extractive activities; marine use;

# **Highlights**

- We present and operationalize a cost-effective framework to quantify local threats inside and outside MPAs.
- We assess how good MPAs are at curbing extractive and non-extractive threats.
- Fully protected areas effectively eliminate extractive threats.
- Fully protected areas attract non-extractive threats, when allowed.
- Small scale fishing intensity is larger in partially protected areas than outside.

# **1. Introduction**

The effects of human coastal activities often combine into cumulative impacts on many marine ecosystems (Halpern et al., 2015; 2008). Marine protected areas (MPAs) represent the most common tool used in marine spatial planning to mitigate human impacts on marine ecosystems (Lubchenco and Grorud-Colvert, 2015; Lubchenco et al., 2003) and are being increasingly used worldwide both for conservation and fisheries management (Boonzaier and Pauly, 2015). Understanding how MPAs target threats they've been designed to address is essential to inform decision-making and optimize conservation outcomes (Guarderas et al., 2008; Hockings et al., 2004). Some MPAs are located in regions of high cumulative human impact, as shown both in the Mediterranean Sea (Coll et al., 2012; Rodríguez-Rodríguez et al., 2015) and in the Atlantic (Batista et al., 2014). This led to debate about their appropriateness at effectively reducing threats (Agardy et al., 2011; Jameson et al., 2002; Rodríguez-Rodríguez et al., 2015).

Most studies that aimed at evaluating MPAs effectiveness at reducing threat intensities were based on large scale assessments, with resolutions of 500 m (Batista et al., 2014) or 1 km2 grid cells (Micheli et al., 2013; Portman and Nathan, 2015). Little emphasis has been given on mapping MPAs' specific threats acting at smaller spatial scales -- possibly with some heterogeneity across MPA zones in the case of multiple-use MPAs --, those at which MPAs and their respective management actions are primarily designed to operate in (MPAs being a local spatial management tool; Olsen et al., 2013). This can result in difficulties to translate research findings into management actions, compromising potential benefits of MPAs (Agardy et al., 2011; Freed and Granek, 2014; Kearney et al., 2012; Mills et al., 2010).

The effectiveness of MPAs in reducing threats should be assessed at a local scale, where protection schemes are implemented. To achieve this goal, it is crucial to understand the differences in occurrence and intensity of human activities between the protected and unprotected areas (Claudet and Guidetti, 2010; Portman and Nathan, 2015). This helps to determine whether an MPA is actually successful in mitigating threats or whether the trends observed are merely an indicator of what is occurring outside of the protected area, at larger scales (Hargreaves-Allen et al., 2011).

Two broad types of MPAs exist. First, fully protected areas (FPAs); where all extractive activities (e.g., fishing) are prohibited and where some non-extractive actives (e.g., diving) can be allowed. They are also known as no-take areas or marine reserves. Second, partially protected areas (PPAs); where some activities are prohibited (e.g., spearfishing), others regulated (e.g., fishing with trammel nets) and others allowed (e.g., boating). Those PPAs can be further classified down according to the impact allowed and regulated uses have on species and habitats (Horta e Costa et al., 2016). Different levels of partial protection, together with full protection, can be combined spatially within multiple-use MPAs. Accordingly, the capacity of MPAs to reduce threats will differ depending on their type, design, regulations and level of enforcement (Di Franco et al., 2016; Guidetti et al., 2008; Scianna et al., 2015). Therefore, information on the intensity of threats within each MPA zone and in the surrounding external areas is necessary to assess MPA effectiveness in reducing threats.

Obtaining detailed information on threats is resource demanding, both in terms of time and costs, as the sources of information are largely heterogeneous (Levin et al., 2014). There is a need for a reliable, cost-effective method to assess the threat-reduction capacity of MPAs, robust to the heterogeneity of data sources and associated levels of confidence based on data quality. In addition, methods need to be standardized across MPAs as to allow both threat comparisons, among different zones within individual MPAs, and across MPAs.

Here, we developed a cost-effective framework to quantify threats at local scale and assess how MPAs are good (or not) at mitigating extractive and non-extractive local area-based threats (Fig. 1). We trialed the framework on 15 Mediterranean MPAs. MPAs in the Mediterranean Sea are a good model since this basin combines a high intensity of human uses (Claudet and Fraschetti, 2010; Portman and Nathan, 2015) together with a high conservation priority (Coll et al., 2010). This information is essential to inform local management as well as forthcoming regional marine spatial planning (European Commission, 2017).

# **2. Materials and methods**

# *2.1. Data collection*

We first identified human threats that affect marine ecosystems at a local scale and that can be managed by the MPA staff through regulations. This allowed the identification of 8 threats (Table 1). Threats were either assigned to extractive (i.e., professional and recreational fishing) or non-extractive uses (i.e., activities related to touristic frequentation).

We then developed indicators for each threat accounting for both availability of the data and the quality of the information given. The set of indicators selected was tailored to the data context of the study case, here the Mediterranean, where data availability can be poor in some regions. Data relevant for quantifying threat indicators were collected by means of questionnaires distributed to local managers and scientists. Local expert had to preferably obtain threat indicator values from scientific studies, technical reports or other official documents. When such sources were not available, expert opinion was considered. Three levels of confidence were applied for the estimated threat values (qualitative: high, medium, low). Threat indicator values were considered as high confidence when they were obtained directly from recent quantitative data (e.g., from monitoring data), as medium confidence when estimated from less recent quantitative data and as low confidence when no quantitative data were available and local experts of the particular MPA provided the estimation of the threat indicator value.

Threat indicators were quantified both within and outside 15 coastal MPAs in the Mediterranean Sea (MPAs listed in Appendix A). In case of multiple-use MPAs (n=13), data was obtained for each full (no-take/no-entry or no-take zones) or partial protection level. We used two approaches to delineate the outside area, depending on the characteristics of the threat: i) for commercial fishing with trawlers and purse seiners a 10 km radius surrounding the MPAs was applied, while ii) for all other threats originating closer to the shore the coastal section at a maximum distance from the shore equal to the MPAs most offshore limit was considered (Fig C.1.). This specific outside areas were chosen following consultations with at least one expert of each MPA. The two approaches aimed to reflect the nature and occurrence of that threat in order to avoid over- or underestimation of threat intensities, respectively, in the outside areas.

The total size of each zone within each MPA was obtained directly from managers and/or

from management plans, whilst the outside surface areas were calculated using QGIS 2.8/1 (QGIS Development Team, 2015).

All the raw data collected in this study can be found in Appendix B.

#### *2.2. Threat indices*

#### 2.2.1. Threat intensity

Using the raw threat values (Table B.1), we calculated the intensity  $T_{ijk}$  of each threat *i* within each protection level *j* (full protection, partial protection and no protection-outside) for each MPA *k*, as follows:

$$
\mathrm{TI}_{ijk} = \frac{\mathrm{T}_{ijk}}{\mathrm{A}_{jk}}
$$

where  $T_{ijk}$  is the value of threat *i* in zone *j* of MPA *k*, and  $A_{ik}$  is the area (km<sup>2</sup>) of zone *j* in MPA *k*. We have then normalized threat intensity values *(TI)* by diving each value with the maximum threat intensity value *(TI)* of each threat, which resulted in a standardized scale of [0;1]. These normalized values were then used to calculate the mean threat intensity of each threat in each protection level to explore the variability of threat intensity amongst protection levels. Detailed descriptive statistics of the normalized threat intensity index can be found in Appendix C.

#### 2.2.2. Threat reduction capacity

The threat reduction capacity *TRijk* of each protection level *j* (excluding the outside area) of each MPA *k* was calculated as:

$$
TR_{ijk} = -\left(1 - \frac{TI_{ijk}}{TI_{o,ik}}\right)
$$

where  $TI_{ijk}$  is the intensity of threat *i* inside zone *j* of MPA  $k$ , and  $TI_{o,ik}$  is the threat intensity of threat *i* outside of the MPA *k*. A negative value of TR indicates that the threat intensity is reduced inside the protected area relative to the outside; a positive value of  $TR$  indicates that the threat intensity is higher inside the MPA relative to the outside. The mean threat reduction capacity values (*TRijk*) were used to compare the reduction capacity of the fully and partially protected zones for every threat. Detailed descriptive statistics of the threat reduction capacity index can be found in Appendix C.

#### 2.2.3. Zone threat reduction score

The zone threat reduction score (*zT*) was calculated as:

$$
zT_{jk} = \frac{\sum_i TR_{ijk} * w_i}{\sum_i w_i}
$$

,

where *TR<sub>ijk</sub>* is the threat reduction capacity defined above and  $w_i$  is the weight associated to threat *i* (Table 1). The weights for each threat were derived from Horta e Costa et al. (2016) and were rescaled to values between 0 and 1, with 1 being the weight assigned to the threat with the greatest ecological impact (i.e. trawling). The weights were used to discriminate between different potential ecological impacts of each threat and were obtained with expert

knowledge, but based on previous studies (see detailed explanation in Appendix A of Horta e Costa et al., 2016).

#### 2.2.4. MPA threat reduction score

The MPA threat reduction score *mpaT* was calculated as:

$$
\text{mpaT}_k = \frac{\sum_j \ zT_{jk} * A_{jk}}{\sum_j A_{jk}},
$$

where  $zT_{ik}$  is the threat reduction score of zone *j* of MPA *k*, and  $A_{jk}$  is the surface area of zone *j* of MPA *k*. When the zone or MPA threat reduction score is negative, threats are being effectively reduced inside the zone and/or MPA compared to outside. When the threat reduction score is positive the MPA is actually enhancing threats rather than reducing them.

#### 2.2.5. Local threat index

To account for the context in which MPAs are sited we calculated a local threat index *lTI* as follows:

$$
lTI_{ik} = \frac{TI_{o,ik}}{\max TI_{o,i}}
$$

,

where the  $TI_{o,ik}$  is the intensity of threat *i* outside MPA *k* and max  $TI_{o,i}$  is the maximum intensity value of threat *i* recorded from all the outside areas of all considered MPAs. The local threat index ranges from 0 to 1, indicating a local low and high intensity, respectively, of a particular threat in the area outside an MPA compared to the broader context.

#### 2.2.6. MPA local threat index

Local threat indices were aggregated at each MPA scale to calculate the MPA local threat index *mpa.lTI* as follows:

mpa. 
$$
lTI_k = \frac{\sum_i lTI_{ik} * w_i}{\sum_i w_i}
$$

where  $lTI_{ik}$  is the local threat index of threat *i* of MPA *k*, as calculated above, and  $w_i$  is the weight of threat *i*. The values of the MPA local threat indices were normalized by dividing by the maximum local threat index value, resulting in a standard scale of 0-1, with low scores indicating a low local threat intensity, that is an overall low intensity of threats in the area outside the MPA, when compared to threat intensities in area outside the other MPAs in the region.

#### *2.3. Data analyses*

The relationship between the MPA threat reduction score (*mpaT*) and the MPA local threat index *(mpa.lTI)* were explored using a linear model. Additionally, the relationship between the threat reduction score, *mpaT*, and the age and size of the MPA were also tested with a regression to investigate whether these design characteristics affect the threat reduction capacity of MPAs.

The data on the confidence level associated with each threat type and zone protection level was analysed using chi-square tests to assess whether there is an association between data quality and zone type or threat type, respectively.

All analyses were conducted using R (R Development Core Team, 2016). The R script is available in Appendix D.

# **3. Results**

Mean threat intensities (*TI*) differed amongst protection levels threat type (Fig. 2). Fully protected zones had no extractive threats (by definition), however they had the highest intensity of all non-extractive threats. On average, the intensity of non-extractive threats was 2.6 and 16 times greater in fully protected zones compared to partially protected zones and the area outside of the MPAs, respectively.

All threats were present in partially protected zones, with the highest levels of both recreational and artisanal fishing (Fig. 2). In the area outside MPAs, the intensity of extractive threats was up to 18 times greater than the intensity of non-extractive threats.

The threat reduction capacity (*TR*) differed across protection level and threat type (Fig. 3). Fully protected areas removed all extractive threats, while on average, partially protected zones did not reduce extractive threats. The intensities of large commercial fisheries (trawling and purse-seine) were reduced in partially protected zones compared to the outside, yet the intensities of artisanal and recreational fishing were approximately 4.9 times greater than in the surrounding areas. All non-extractive threats were increased within fully and partially protected zones, compared to the outside surrounding areas.

The mean zone threat reduction score  $(zT)$  for non-extractive threats was, on average, 4.8 greater for fully protected zones compared to partially protected zones, yet this value was highly variable (Fig. 3).

The MPA threat reduction score (*mpaT*) ranged between -0.8 and 116 (Fig. 4). Overall, only three MPAs reduced all threats relative to the outside. Five MPAs reduced all extractive threats and two MPAs reduced all non-extractive threats relative to their outside area.

The intensity of threats in the areas outside of MPAs varied greatly among MPAs (Fig. C.2 in Appendix C). There were no significant relationship between the MPAs' threat reduction score (*mpaT*) and the MPA local threat index (*mpa.lTI*), indicating that the capacity of MPAs to reduce threats was independent on the intensity of threats in their surrounding areas  $(p>0.05$ , see Fig. C.3 in Appendix C). Likewise, there were no significant relationships between the MPA threat reduction score and the age or size of the MPA (p>0.05, see Fig. C.4 and C.5 in Appendix C).

The confidence levels associated with estimates of threat intensities differed among zones of MPAs and the outside areas (chi-squared test,  $\chi^2$ =153.54, p<0.001). More than half of the values from the areas outside were given low confidence (Fig. 5). In contrast within the MPAs, the majority of threat values had high confidence, 65% and 50% in fully and partially protected zones, respectively.

Confidence levels also differed amongst threats (chi-squared test,  $\chi^2$ =233.63, p<0.001). For all extractive threats, the greatest proportion of values had high confidence (Fig. 6). Scubadiving and commercial boating were among the non-extractive threats with the highest proportion of high confidence, while the values attributed to private boating and bathing had low confidence.

### **4. Discussion**

This study was designed to provide a framework of the effectiveness of MPAs at curbing threats at the scale they are designed, developing indices of the intensity of threats at the scale of single MPAs (Fig. 1). The information provided was scaled to support management actions designed to enhance the success of each MPA in achieving its goals (Hockings et al., 2004).

Our most compelling result is that threats can indeed increase in MPAs compared to outside areas. Even if absent for fully protected zones, many extractive threats were larger in partially protected zones compared to outside. Contrary to fully protected zones, partially protected zones do not necessarily have the objective to eliminate threats (e.g., via a fishing ban), but to regulate their intensity towards sustainable levels. Implementing management strategies that allow and maintain ecologically sustainable uses ensures long-term benefits from ecosystem services benefiting local economies and, hence, local communities (Roberts et al., 2005). In the Mediterranean Sea, the regulatory regimes of partially protected areas vary from MPA to another, however there are some common modalities (Portman et al., 2012).

Artisanal and recreational fishing are permitted in partially protected zones of many MPAs, although most of the time subjected to restrictions in the type of gear allowed. The intensity of artisanal and recreational fishing was up to 4.8 times higher inside partially protected zones compared to outside control areas. While the regulatory regime in partially protected zones may be stricter than in the surrounding outside areas, the former still attract more fishers and hence increase fishing pressure. Most of the partially protected zones considered in this study surround fully protected areas (Appendix A) and may attract fishermen. Full protection enhance fish abundance and size (Guidetti et al., 2014; Sala et al., 2012), often resulting in the spillover of larvae and adults to adjacent unprotected areas (Di Franco et al., 2015; 2012; Di Lorenzo et al., 2016; Garcia-Rubies et al., 2013; Goñi et al., 2010; Stobart et al., 2009), hence concentrating fishing effort (Kellner et al., 2007; Stelzenmüller et al., 2007).

The intensity of non-extractive threats was also higher within MPAs than in the areas outside, and at least 4 times higher in fully protected zones compared to partially protected zones*.* By definition, fully protected areas are places aiming to protect the full spectrum of biodiversity by limiting all kinds of extractive and destructive activity within them (Allison et al., 1998), yet this does not exclude all human activity. Indeed, most fully protected zones allow and, as we show here, *attract* visitation with the intent of non-extractive activities (Thurstan et al., 2012). Increasing non-extractive touristic activities are viewed as a positive socioeconomic output of MPAs in the Mediterranean Sea and worldwide (Leisher et al., 2007; Pascual et al., 2016), as they promote education, support employment and generate revenue (Hargreaves-Allen et al., 2011; Pascual et al., 2016; Spalding et al., 2017). Increasing tourism is, therefore, often listed as a looked-for objective of MPAs (Hargreaves-Allen et al., 2011). Thurstan et al.

(2012) suggest that unless all activities are adequately managed, regulation of extractive activities alone may not guarantee the levels of biodiversity protection expected from fully protected zones.

Only three MPAs were effective in reducing the intensity of threats in respect to adjacent unprotected areas. Examining the context in which MPAs are established is, nevertheless, central to understanding the capacity of MPAs in addressing threats and their conservation potential (Portman and Nathan, 2015). The majority of MPAs in this study are multiple-use MPAs that accommodate a variety of uses. This type of MPAs is common for densely populated systems (Agardy et al., 2003), such as the Mediterranean Sea, where reducing all threats is unlikely. Portman et al. (2015) showed that the variety of activities occurring within the Mediterranean MPAs is often greater than in the coastal unprotected areas of many Mediterranean countries. Yet, multiple use MPAs and particularly partially protected areas alone are now the most common type of MPAs being implemented worldwide (Claudet, 2017). They are being established to meet the international targets of protection (Agardy et al. 2016), with the stated objectives of biodiversity conservation. Our results point that the achievability of their goals should be questioned, as here, we show that not only the variety, but also intensity of threats is generally greater within MPAs than in the areas outside (see also Mora et al., 2006) that leads to the low threat reduction capacity of MPAs.

The capacity to reduce (or not) threats was not dependent on the threat local context, the age or the size of the MPA. This suggests that the relationship between threat-intensity and reduction-capacity is mainly a function of management objectives and capacity (Gill et al., 2017), corresponding regulations (Horta e Costa et al. 2016) and enforcement (Guidetti et al., 2008) of each MPA rather than the magnitude of the outside threat intensities or design features. Hargreaves et al. (2017) examined the differences in threats inside and outside coral reef MPAs and they showed that MPAs with reduced threats had more staff and invested more funds into active management. Indeed, management has been identified as one of the most important factors affecting the effectiveness of MPAs (Gill et al., 2017). Understanding how different management regimes affect the threat reduction capacity of zones in multipleuse MPAs should be further investigated as it could shed light onto the highly variable responses we have observed among protection levels.

Our framework uses threats, and their respective indicators, that are most common and were easily obtained in the Mediterranean MPAs. While we were able to trial our framework using the information available, the assessment of the MPAs threat reduction capacity could be refined with more detailed and robust data. Besides, we do acknowledge that other types of threats may be present in other systems yet this issue can be easily overcome, as our framework can be adapted to suit a variety of systems, by incorporating new threats and their corresponding indicators.

We did not include poaching as one of the threats in our framework. Poaching levels can often be higher than assumed in MPAs (Bergseth et al., 2017) and represent an important threat to the effectiveness of MPAs. While we initially planned to collect data on poaching, we faced strong difficulties to quantitatively standardize this threat across MPAs. If standardized data can be obtained across a range of MPAs, this can easily be added in the framework as an additional threat.

Our results point out that we have greater confidence about what is happening inside MPAs than in their surrounding areas. The low confidence associated with threat values in the

unprotected, outside, areas might have affected the precision of our results. The areas outside MPAs are usually less monitored or completely unmonitored and estimating threat values for these areas was challenging. Good management should require regular monitoring, both for the ecological and environmental status of the MPA habitats (Fraschetti et al., 2013), but also should incorporate monitoring of the main activities that can lead to potential threats both inside and beyond the borders of the MPA (Claudet and Guidetti, 2010; Hargreaves-Allen et al., 2017; Parravicini et al., 2013).

Evaluating the success of MPAs is essential to maximize their conservation potential (Agardy et al., 2016; Pomeroy et al., 2005), especially within an adaptive management framework (Scianna et al., 2015). Contrary to previous assessments (Coll et al., 2012; Micheli et al., 2013) our approach uses data specific to each zone of the MPA, where strong heterogeneity among threats intensity can occur, and accounts for manageable threats at manageable scales relevant for MPA managers. We therefore believe this work has general and wide applications, especially for managers and planners who need to assess the success of MPAs at achieving marine spatial planning goals (Partelow et al., 2015).

# **Management implications**

By providing the ability to easily compare threat intensities and the capacity of a given MPA to reduce threats, our proposed framework can help prioritize management actions. Management actions should first be directed towards reducing threats that would prevent achieving the MPA objectives. Besides, the threat assessment framework can help identify unexpected indirect effects of MPA creation, such as increased non-extractive threat due to increase attendance.

By providing the ability to easily compare threat reduction score among zones, in the case of multiple-use MPAs, our proposed framework can help guide local spatial planning. Assigning different uses in different zones can have trade-offs in terms of biodiversity conservation and ecosystem services delivery and threat assessment is a first step towards the identification of acceptable thresholds of uses. When scaled-up regionally, the threat assessment framework can help guide regional spatial planning.

The threat assessment framework could be incorporated into monitoring programs. First, monitoring threat evolution over time helps identify the management actions able to reduce threats. Second, this would necessarily imply to monitor and collect data, which is a condition of proper MPA effectiveness assessments.

#### **Acknowledgments**

This work was financially supported by the ERA-Net BiodivERsA (BUFFER project) and the EU-FP7 (COCONET project).

#### **References**

- Agardy, T., Bridgewater, P., Crosby, M.P., Day, J., Dayton, P.K., Kenchington, R., Laffoley, D., McConney, P., Murray, P.A., Parks, J.E., Peau, L., 2003. Dangerous targets? Unresolved issues and ideological clashes around marine protected areas. Aquatic Conserv: Mar. Freshw. Ecosyst. 13, 353–367. doi:10.1002/aqc.583
- Agardy, T., Claudet, J., Day, J.C., 2016. "Dangerous Targets" revisited: Old dangers in new contexts plague marine protected areas. Aquatic Conserv: Mar. Freshw. Ecosyst. 26, 7– 23. doi:10.1002/aqc.2675
- Agardy, T., di Sciara, G.N., Christie, P., 2011. Mind the gap: Addressing the shortcomings of marine protected areas through large scale marine spatial planning. Mar. Policy 35, 226– 232. doi:10.1016/j.marpol.2010.10.006
- Allison, G.W., Lubchenco, J., Carr, M.H., 1998. Marine Reserves are Necessary but not Sufficient for Marine Conservation. Ecol. Appl. 8, S79-92
- Batista, M.I., Henriques, S., Pais, M.P., Cabral, H.N., 2014. Assessment of cumulative human pressures on a coastal area: Integrating information for MPA planning and management. Ocean Coast. Manag. 102, 248–257. doi:10.1016/j.ocecoaman.2014.09.020
- Boonzaier, L., Pauly, D., 2015. Marine protection targets: an updated assessment of global progress. Oryx 50, 27–35. doi:10.1017/S0030605315000848
- Bergseth, B.J., 2017 Effective marine protected areas require sea change in compliance management. ICES J. Mar. Sci. doi:10.1093/icesjms/fsx105
- Claudet, J., Fraschetti, S., 2010. Human-driven impacts on marine habitats: A regional metaanalysis in the Mediterranean Sea. Biol. Cons. 143, 2195–2206. doi:10.1016/j.biocon.2010.06.004
- Claudet, J., Guidetti, P., 2010. Improving assessments of marine protected areas. Aquatic Conserv: Mar. Freshw. Ecosyst. 20, 239–242. doi:10.1002/aqc.1087
- Claudet, J., 2017. Six conditions under which MPAs might not appear effective (when they are). ICES J. Mar. Sci. doi: 10.1093/icesjms/fsx074.
- Coll, M., Piroddi, C., Albouy, C., Ben Rais Lasram, F., Cheung, W.W.L., Christensen, V., Karpouzi, V.S., Guilhaumon, F., Mouillot, D., Paleczny, M., Palomares, M.L., Steenbeek, J., Trujillo, P., Watson, R., Pauly, D., 2012. The Mediterranean Sea under siege: spatial overlap between marine biodiversity, cumulative threats and marine reserves. Glob. Ecol. and Biogeogr. 21, 465–480. doi:10.1111/j.1466-8238.2011.00697.x
- Coll, M., Piroddi, C., Steenbeek, J., Kaschner, K., Ben Rais Lasram, F., Aguzzi, J., Ballesteros, E., Bianchi, C.N., Corbera, J., Dailianis, T., Danovaro, R., Estrada, M., Froglia, C., Galil, B.S., Gasol, J.M., Gertwagen, R., Gil, J., Guilhaumon, F., Kesner-Reyes, K., Kitsos, M.-S., Koukouras, A., Lampadariou, N., Laxamana, E., López-Fé de la Cuadra, C.M., Lotze, H.K., Martin, D., Mouillot, D., Oro, D., Raicevich, S., Rius-Barile, J., Saiz-Salinas, J.I., San Vicente, C., Somot, S., Templado, J., Turon, X., Vafidis, D., Villanueva, R., Voultsiadou, E., 2010. The Biodiversity of the Mediterranean Sea: Estimates, Patterns, and Threats. PLOS ONE 5, e11842–36. doi:10.1371/journal.pone.0011842
- Di Franco, A., Calò, A., Pennetta, A., De Benedetto, G., Planes, S., Guidetti, P., 2015. Dispersal of larval and juvenile seabream: Implications for Mediterranean marine protected areas. Biol. Conserv. 192, 361–368. doi:10.1016/j.biocon.2015.10.015
- Di Franco, A., Coppini, G., Pujolar, J.M., De Leo, G.A., Gatto, M., Lyubartsev, V., Melià, P., Zane, L., Guidetti, P., 2012. Assessing Dispersal Patterns of Fish Propagules from an Effective Mediterranean Marine Protected Area. PLOS ONE 7, e52108–14. doi:10.1371/journal.pone.0052108
- Di Franco, A., Thiriet, P., Di Carlo, G., Dimitriadis, C., Francour, P., Gutiérrez, N.L., de Grissac, A.J., Koutsoubas, D., Milazzo, M., del Mar Otero, M., Piante, C., Plass-Johnson, J., Sainz-Trapaga, S., Santarossa, L., Tudela, S., Guidetti, P., 2016. Five key attributes can increase marine protected areas performance for small-scale fisheries management. Sci. Rep. 6, 1–9. doi:10.1038/srep38135
- Di Lorenzo, M., Claudet, J., Guidetti, P., 2016. Spillover from marine protected areas to adjacent fisheries has an ecological and a fishery component. J. Nat. Conserv. 32, 62–66. doi:10.1016/j.jnc.2016.04.004
- European Commission, 2017. Maritime Affairs, Maritime spatial Planning. https://ec.europa.eu/maritimeaffairs/policy/maritime\_spatial\_planning\_en (accessed 10.07.2017)
- Fraschetti, S., Guarnieri, G., Bevilacqua, S., Terlizzi, A., Boero, F., 2013. Protection enhances community and habitat stability: evidence from a mediterranean marine protected area. PLOS ONE 8, e81838. doi:10.1371/journal.pone.0081838
- Freed, S., Granek, E.F., 2014. Effects of Human Activities on the World's Most Vulnerable Coral Reefs: Comoros Case Study. Ocean Coast. Manage. 42, 280–296. doi:10.1080/08920753.2014.904261
- Garcia-Rubies, A., Hereu, B., Zabala, M., 2013. Long-term recovery patterns and limited spillover of large predatory fish in a Mediterranean MPA. PLOS ONE 8, e73922. doi:10.1371/journal.pone.0073922
- Gill, D.A., Mascia, M.B., Ahmadia, G.N., Glew, L., Lester, S.E., Barnes, M., Craigie, I., Darling, E.S., Free, C.M., Geldmann, J., Holst, S., Jensen, O.P., White, A.T., Basurto, X., Coad, L., Gates, R.D., Guannel, G., Mumby, P.J., Thomas, H., Whitmee, S., Woodley, S., Fox, H.E., 2017. Capacity shortfalls hinder the performance of marine protected areas globally. Nature. 000, 1–16. doi:10.1038/nature21708
- Gonson, C., Pelletier, D., Gamp, E., Preuss, B., Jollit, I., Ferraris, J., 2016. Decadal increase in the number of recreational users is concentrated in no-take marine reserves. Mar. Poll. Bull. 107, 1–11. doi:10.1016/j.marpolbul.2016.04.007
- Goñi, R., Hilborn, R., D i az, D., Mallol, S., Adlerstein, S., 2010. Net contribution of spillover from a marine reserve to fishery catches. Mar. Ecol. Prog. Ser. 400, 233–243. doi:10.3354/meps08419
- Guarderas, A.P., Hacker, S.D., Lubchenco, J., 2008. Current Status of Marine Protected Areas in Latin America and the Caribbean. Conserv. Biol. 22, 1630–1640. doi:10.1111/j.1523-1739.2008.01023.x
- Guerrero, A.M., McAllister, R.R.J., Corcoran, J., Wilson, K.A., 2013. Scale Mismatches, Conservation Planning, and the Value of Social-Network Analyses. Conserv. Biol. 27, 35–44. doi:10.1111/j.1523-1739.2012.01964.x
- Guidetti, P., Baiata, P., Ballesteros, E., Di Franco, A., Hereu, B., Macpherson, E., Micheli, F., Pais, A., Panzalis, P., Rosenberg, A.A., Zabala, M., Sala, E., 2014. Large-Scale Assessment of Mediterranean Marine Protected Areas Effects on Fish Assemblages. PLOS ONE 9, e91841–14. doi:10.1371/journal.pone.0091841
- Guidetti, P., Milazzo, M., Bussotti, S., Molinari, A., Murenu, M., Pais, A., Spanò, N., Balzano, R., Agardy, T., Boero, F., Carrada, G., Cattaneo-Vietti, R., Cau, A., Chemello, R., Greco, S., Manganaro, A., Notarbartolo di Sciara, G., Russo, G.F., Tunesi, L., 2008. Italian marine reserve effectiveness: Does enforcement matter? Biol. Conserv. 141, 699– 709. doi:10.1016/j.biocon.2007.12.013
- Halpern, B.S., Frazier, M., Potapenko, J., Casey, K.S., Koenig, K., Longo, C., Lowndes, J.S., Rockwood, R.C., Selig, E.R., Selkoe, K.A., Walbridge, S., 2015. Spatial and temporal changes in cumulative human impacts on the world's ocean. Nat. Commun. 6, 7615. doi:10.1038/ncomms8615
- Halpern, B.S., Walbridge, S., Selkoe, K.A., Kappel, C.V., Micheli, F., D'Agrosa, C., Bruno, J.F., Casey, K.S., Ebert, C., Fox, H.E., Fujita, R., Heinemann, D., Lenihan, H.S., Madin, E.M.P., Perry, M.T., Selig, E.R., Spalding, M., Steneck, R., Watson, R., 2008. A Global Map of Human Impact on Marine Ecosystems. Science 319, 948–952. doi:10.1126/science.1149345
- Hargreaves-Allen, V., Mourato, S., Milner-Gulland, E.J., 2011. A Global Evaluation of Coral Reef Management Performance: Are MPAs Producing Conservation and Socio-Economic Improvements? Environ. Manage. 47, 684–700. doi:10.1007/s00267-011- 9616-5
- Hargreaves-Allen, V.A., Mourato, S., Milner-Gulland, E.J., 2017. Drivers of coral reef marine protected area performance. PLOS ONE 12, e0179394–21. doi:10.1371/journal.pone.0179394
- Hockings, M., Stolton, S., Dudley, N., 2004. Management Effectiveness: Assessing Management of Protected Areas? J. Environ. Policy Plan. 6, 157–174. doi:10.1080/1523908042000320731
- Horta e Costa, B., Claudet, J., Franco, G., Erzini, K., Caro, A., Gonçalves, E.J., 2016. A regulation-based classification system for Marine Protected Areas (MPAs). Mar. Policy 72, 192–198. doi:10.1016/j.marpol.2016.06.021
- Jameson, S.C., Tupper, M.H., Ridley, J.M., 2002. The three screen doors: can marine "protected" areas be effective? Mar. Poll. Bull. 44, 1177–1183. doi:10.1016/S0025- 326X(02)00258-8
- Kearney, R., Buxton, C.D., Farebrother, G., 2012. Australia's no-take marine protected areas Appropriate conservation or inappropriate management of fishing? Mar. Policy 36, 1064–1071. doi:10.1016/j.marpol.2012.02.024
- Keller, B.D., Gleason, D.F., McLeod, E., Woodley, C.M., Airame, S., Causey, B.D., Friedlander, A.M., Grober-Dunsmore, R., Johnson, J.E., Miller, S.L., Steneck, R.S., 2009. Climate change, coral reef ecosystems, and management options for marine protected areas. Environ. Manage. 44, 1069–1088. doi:10.1007/s00267-009-9346-0
- Kellner, J.B., Tetreault, I., Gaines, S.D., Nisbet, R.M., 2007. Fishing the line near marine reserves in single and multispecies fisheries, Ecol. Appl.. doi:10.1890/05-1845
- Leisher, C., Van Beukering, P., Scherl, L.M., 2007. Nature's Investment Bank: How Marine Protected Areas Contribute to Poverty Reduction. Arlington, USA: The Nature **Conservancy**
- Levin, N., Coll, M., Fraschetti, S., Gal, G., Giakoumi, S., Göke, C., Heymans, J.J., Katsanevakis, S., Mazor, T., Öztürk, B., Rilov, G., Gajewski, J., Steenbeek, J., Kark, S., 2014. Biodiversity data requirements for systematic conservation planning in the Mediterranean Sea. Mar. Ecol. Prog. Ser. 508, 261–281. doi:10.3354/meps10857
- Lubchenco, J., Grorud-Colvert, K., 2015. Making waves: The science and politics of ocean protection. Science 350, 382–383. doi:10.1126/science.aad5443
- Lubchenco, J., Palumbi, S.R., Gaines, S.D., Andelman, S., 2003. Plugging a hole in the ocean: the emerging science of marine reserves. Ecol. Appl. 13, 3–7. doi:10.1890/1051- 0761(2003)013[0003:PAHITO]2.0.CO;2
- Micheli, F., Halpern, B.S., Walbridge, S., Ciriaco, S., Ferretti, F., Fraschetti, S., Lewison, R., Nykjaer, L., Rosenberg, A.A., 2013. Cumulative Human Impacts on Mediterranean and Black Sea Marine Ecosystems: Assessing Current Pressures and Opportunities. PLOS ONE 8, e79889. doi:10.1371/journal.pone.0079889
- Mills, M., Pressey, R.L., Weeks, R., Foale, S., Ban, N.C., 2010. A mismatch of scales: challenges in planning for implementation of marine protected areas in the Coral Triangle. Conserv. Lett. 3, 291–303. doi:10.1111/j.1755-263X.2010.00134.x
- Mora, C., Andrèfouët, S., Costello, M.J., Kranenburg, C., Rollo, A., Veron, J., Gaston, K.J., Myers, R.A., 2006. Coral reefs and the global network of Marine Protected Areas. Science 312, 1750–1751. doi:10.1126/science.1125295
- Olsen, E.M., Fonseca, L., Goni, R., Johnson, D., Katsanevakis, S., Macpherson, E., Pelletier, D., Rabaut, M., Chantal-Ribeiro, M., Weaver, P., Zaharia, T., 2013. Achieving Ecologically Coherent MPA Networks in Europe: Science Needs and Priorities. European Marine Board Position Paper 18. Larkin, K.E, and McDonough, N. (Eds.). European Marine Board, Ostend, Belgium
- Parravicini, V., Micheli, F., Montefalcone, M., Morri, C., Villa, E., Castellano, M., Povero, P., Bianchi, C.N., 2013. Conserving biodiversity in a human-dominated world: degradation of marine sessile communities within a protected area with conflicting human uses. PLOS ONE 8, e75767. doi:10.1371/journal.pone.0075767
- Partelow, S., Wehrden, von, H., Horn, O., 2015. Pollution exposure on marine protected areas: A global assessment. Mar. Poll. Bull. 100, 352–358. doi:10.1016/j.marpolbul.2015.08.026
- Pascual, M., Rossetto, M., Ojea, E., Milchakova, N., Giakoumi, S., Kark, S., Korolesova, D., Melià, P., 2016. Socioeconomic impacts of marine protected areas in the Mediterranean and Black Seas. Ocean Coast. Manage. 133, 1–10. doi:10.1016/j.ocecoaman.2016.09.001
- Pomeroy, R.S., Watson, L.M., Parks, J.E., Cid, G.A., 2005. How is your MPA doing? A methodology for evaluating the management effectiveness of marine protected areas. Ocean Coast. Manage. 48, 485–502. doi:10.1016/j.ocecoaman.2005.05.004
- Portman, M.E., Nathan, D., 2015. Conservation "identity" and marine protected areas management: A Mediterranean case study. J. Nat. Conserv. 24, 109–116. doi:10.1016/j.jnc.2014.10.001
- Portman, M.E., Nathan, D., Levin, N., 2012. From the Levant to Gibraltar: A Regional Perspective for Marine Conservation in the Mediterranean Sea. Ambio 41, 670–681. doi:10.1007/s13280-012-0298-x
- QGIS Development Team, 2015. QGIS Geographic Information System. Open Source Geospatial Foundation. https://qgis.osgeo.orh
- R Core Team, 2016. R: A language and environment for statistical computing. Foundation for Statistical Computing, Vienna, Austria
- Roberts, C.M., Hawkins, J.P., Gell, F.R., 2005. The role of marine reserves in achieving sustainable fisheries. Philos. Trans. R. Soc. Lond., B, Biol. Sci. 360, 123–132. doi:10.1098/rstb.2004.1578
- Rodríguez-Rodríguez, D., Sánchez-Espinosa, A., Schröder, C., Abdul Malak, D., Rodríguez, J., 2015. Cumulative pressures and low protection: a concerning blend for Mediterranean MPAs. Mar. Poll. Bull. 101, 288–295. doi:10.1016/j.marpolbul.2015.09.039
- Sala, E., Ballesteros, E., Dendrinos, P., Di Franco, A., Ferretti, F., Foley, D., Fraschetti, S., Friedlander, A., Garrabou, J., Güçlüsoy, H., Guidetti, P., Halpern, B.S., Hereu, B., Karamanlidis, A.A., Kizilkaya, Z., Macpherson, E., Mangialajo, L., Mariani, S., Micheli, F., Pais, A., Riser, K., Rosenberg, A.A., Sales, M., Selkoe, K.A., Starr, R., Tomas, F., Zabala, M., 2012. The Structure of Mediterranean Rocky Reef Ecosystems across Environmental and Human Gradients, and Conservation Implications. PLOS ONE 7, e32742–13. doi:10.1371/journal.pone.0032742
- Scianna, C., Niccolini, F., Gaines, S.D., Guidetti, P., 2015. Organization Science': A new prospective to assess marine protected areas effectiveness. Ocean Coast. Manage. 116, 443–448. doi:10.1016/j.ocecoaman.2015.09.005

Spalding, M., Burke, L., Wood, S.A., Ashpole, J., Hutchison, J., zu Ermgassen, P., 2017. Mapping the global value and distribution of coral reef tourism. Mar. Policy 82, 104– 113. doi:10.1016/j.marpol.2017.05.014

Stelzenmüller, V., Maynou, F., Mart i n, P., 2007. Spatial assessment of benefits of a coastal Mediterranean Marine Protected Area. Biol. Conserv. 136, 571–583. doi:10.1016/j.biocon.2007.01.002

- Stobart, B., Warwick, R., González, C., Mallol, S., Díaz, D., Reñones, O., Goni, R., 2009. Long-term and spillover effects of a marine protected area on an exploited fish community. Mar. Ecol. Prog. Ser. 384, 47–60. doi:10.3354/meps08007
- Thurstan, R.H., Hawkins, J.P., Neves, L., Roberts, C.M., 2012. Are marine reserves and nonconsumptive activities compatible? A global analysis of marine reserve regulations. Mar. Policy 36, 1096–1104. doi:10.1016/j.marpol.2012.03.006

**Tables**

**Table 1.** Local threats that can affect marine protected area (MPA) effectiveness and that can be managed by the MPA regulations. Indicators were chosen considering trade-offs between reliability and data availability. Weights reflect the potential impact of a given activity on both species and habitats and were rescaled after Horta e Costa et al. (2016)



### **Figures**

# **How good is your MPA at curbing threats?**

**Step 1.** Identify threats



#### **Step 2.** Quantify threats inside and outside the MPA



Fig. 1. The framework to assess how good are marine protected areas (MPAs) at curbing threats.



**Fig. 2.** Mean threat intensities (± SE) across fully protected, partially protected and outside zones of 15 marine protected areas.



**Threats** 

**Fig. 3.** Mean threat reduction capacity (± SE) for each threat across zones of different protection levels. The last bar of each panel represents the zone threat reduction score (*zT)*. Negative values of the threat reduction capacity indicate that the threat is reduced in the marine protected area (MPA) relative to the outside, while positive values indicate that threats are higher inside the MPA relative to the outside.



**Fig. 4.** Threat reduction score of marine protected areas (MPAs) for all threats combined (top panel), extractive (bottom left panel) and non-extractive threats (bottom right panel). Negative values of the threat reduction score indicate that threats are being effectively reduced within an MPA relative to the outside, while positive threat reduction score values indicate that the MPA is not effective at reducing threats compared to the outside.



**Fig. 5.** Level of confidence (low, medium and high) associated to threat values corresponding to the fully protected zones, partially protected zones and the outside areas of the marine protected areas.



**Fig. 6.** Confidence levels (low, medium and high) associated to threat values for each threat. The red and blue outlines separate extractive and non-extractive threats.

# **Appendix A** Database of marine protected areas used in the present study

Table A.1: The characteristics of each of the marine protected areas (MPAs) included in this study. Zone name represents the official name on the basis of the establishment decree. Zones were classified according to the level of protection: fully protected zones (no-entry and no-take areas) and partially protected zones. The size of each zone within the MPA was obtained from managers and/or management plans, whilst the outside surface areas were calculated using QGIS 2.8/1 (QGIS Development Team, 2015). Commercial fisheries using trawlers and purse seiners were assessed in a 10km radius outside area whereas recreational and artisanal fishing and tourism were assessed in the coastal outside area.









# **Appendix B**

# **Raw threat data collected of marine protected areas used in the present study**

Table B.1. Raw data collected for all threat indicators for each zone of all marine protected areas (MPAs) used in this study. The source of data and confidence level associated to each threat value are also presented. Zone name represents the official name on the basis of the establishment decree. Zones were classified according to the level of protection: fully protected zones (no-entry and no-take areas), partially protected zones and the unprotected-outside areas. The size of each zone within the MPA was obtained from managers and/or management plans, whilst the outside surface areas were calculated using QGIS 2.8/1 (QGIS Development Team, 2015). Commercial fisheries using trawlers and purse seiners were assessed in a 10km radius outside area whereas recreational and artisanal fishing and tourism were assessed in the coastal outside area.





























































































### **References**

Association AGIR. 2013. Appui et accompagnement à la création/post création d'activités génératrices de revenus et de coopérative les pêcheurs néo-alphabétisés au sein du littoral du Rif centrale et orientale Marocain. Rapport global final. 120p. http://panorama.solutions/sites/default/files/rapport-final-projet-agir\_-app.pdf

Cappanera, V., Venturini, S., & Campodonico, P., Blini, V., Ortenzi, 2010. Valutazione dell'impatto antropico legato alle attività svolte nell'Area Marina protetta Portofino (AMP). Le attività di fruizione 2010-2011. Rapporto tecnico: 194 pp.

Cappanera, V., Venturini, S., & Campodonico, P. 2014. Valutazione dell'impatto antropico e socio-economico nell'Area Marina Protetta del Promontorio di Portofino. Portofino MPA, Annual Report, 1–51 pp.

Cattaneo-Vietti, R. (ed.). 2013. La pesca ricreativa nel Mar Ligure. In Modelli di governante e monitoraggio e valorizzazione delle risorse ittiche (pp. 1–84). Regione Liguria: Progetto strategico MARTE+.

Franzosini, C., Limam, A. 2006. Plan de gestion de la composante marine du Parc National d'Al Hoceima. Rapport du Projet Régional pour le Développement d'Aires Protégées Marines et Côtières dans la Région Méditerranéenne (Projet MedMPA). 92 p.

Font, T., Lloret, J., Piante, C. 2012. Recreational fishing within Marine Protected Areas in the Mediterranean. MedPAN North Project. WWFFrance: 168 pp

Nibani, H. 2010. Rapport des ateliers de communication sur la gestion du Parc Nationale d'Al Hoceima. Rapport final, 31p. http://www.mapama.gob.es/es/parques-nacionales-oapn.

Sacanell, M. 2010. Avaluació de les captures de la pesca artesanal en la zona d'influència de l'Àrea Protegida de les Illes Medes.

## **Appendix C**

## **Supplementary Figure and Descriptive Statistics**

# **Supplementary Figures**



**Fig. C.1.** Example of outside areas of a marine protected area, as used for the study. We used two approaches to delineate the outside area, depending on the characteristics of the threat: 1) for commercial fishing with trawlers and purse seiners a 10 km radius surrounding the MPAs was applied (dashed area), while 2) for all other threats originating closer to the shore we considered the coastal section up to a maximum distance from the shore equal to the MPAs most offshore limit (brown area).



**Fig. C.2.** Local threat index of marine protected areas (MPAs) for all threats combined (top panel), extractive (bottom left panel) and non-extractive threats (bottom right panel). A low/high index mean that the MPA have a low/high threat intensities in its outside area, when compared to the threat intensities found in the outside areas of the other MPAs included in this study.



**Fig. C.3.** Relationships between the marine protected area (MPA) threat reduction score and the MPA local threat index for all threats combined (top panel), extractive (bottom left panel) and non-extractive threats (bottom right panel). No significant linear relationships were found  $(p>0.05)$ .



**Fig. C.4.** Relationships between the threat reduction score and the age of the marine protected areas (MPAs) for all threats combined (top panel), extractive (bottom left panel) and non-extractive threats (bottom right panel). No significant linear relationships were found  $(p>0.05)$ .



**Fig. C.5.** Relationships between the threat reduction score and the size of the marine protected areas (MPAs) for all threats combined (top panel), extractive (bottom left panel) and non-extractive threats (bottom right panel). No significant linear relationships were found  $(p>0.05)$ .

## **Descriptive statistics for the threat intensity index and threat reduction capacity index**

Table C.1. Descriptive statistics for the standardized threat intensity index. The standardized threat intensity values were obtained by diving with the maximum threat intensity of each threat. This resulted in a standardized scale of 0-1. These normalized values were then used to calculate the mean threat intensity of each threat in each protection level to explore the variability of threat intensity amongst protection levels (Fig. 1).





Table C.2. Descriptive statistics for the threat reduction capacity index. A negative value of index indicates that the threat intensity is reduced inside the protected area relative to the outside; a positive value of index indicates that the threat intensity is higher inside the MPA relative to the outside. The mean threat reduction capacity values were used to compare the reduction capacity of the fully and partially protected zones for every threat (Fig. 2).





#### APPENDIX D

Title: How good is your marine protected area at curbing threats?

Authors: Zupan, M., Bulleri, F., Evans, J., Fraschetti, S., Guidetti, P., Garcia-Rubies, #A., Sostres, M., Asnaghi, V., Caro, A., Deudero,S., Goñi,R., Guarnieri, G., Guilhaumon, F., Kersting, D., Kokkali, A., Kruschel, C., Macic, V., Mangialajo, L., Mallol, S., Macpherson, E., Panucci, A., Radolovic, M., Ramdani, M., Schembri, P.J., Terlizzi, A., Villa, E., Claudet, J.

Journal: Biological Conservation

# This R script can be used to calculate the following indices

# 1. Threat intensity index # 2. Threat reduction capacity index # 3. Zone threat reduction score # 4. MPA threat reduction score # 5. Local threat index # 6. MPA Local threat index

# Open and load table with raw data on MPA threats assessment

MPAthreats

#Load library(plyr)

```
############################# ############################# 
############################# #############################
```
# Prep the table

############################# #############################

############################# #############################

# Add threat weights; Import threat weight value of each threat (Table 1; https://www.dropbox.com/s/1301uzj0s907gbp/MPAthreats.xlsx?dl=0)

Threats<-MPAthreats[,"Threats"];Threats<-as.character(Threats); sort(unique(Threats))

```
Threats[Threats=="Trawlers/purse-seiners"]= 1 
Threats[Threats=="Artisanal fishing"]= 0.82 
Threats[Threats=="Recreational fishing: hook and line"]= 0.56
Threats[Threats=="Recreational fishing: spearfishing"]= 0.33 
Threats[Threats=="Tourism: commercial boating"]= 0.22 
Threats[Threats=="Tourism: private boating"]= 0.22
Threats[Threats=="Tourism: scuba-diving"]= 0.11
Threats[Threats=="Tourism: bathing/trampling"]= 0.11 
T
hreats<-as.numeric(Threats)
MPAthreats$Weight<-Threats
# Assign threat type (extractive or non-extractive)
```

```
MPAthreats$Type[MPAthreats$Threats == "Trawlers/purse-seiners"] <- "extractive"
MPAthreats$Type[MPAthreats$Threats == "Recreational fishing: spearfishing"] <- "extractive"
MPAthreats$Type[MPAthreats$Threats == "Recreational fishing: hook and line"] <- "extractive"
MPAthreats$Type[MPAthreats$Threats == "Artisanal fishing"] <- "extractive"
MPAthreats$Type[MPAthreats$Threats == "Tourism: scuba-diving" ] <- "non-extractive"
MPAthreats$Type[MPAthreats$Threats == "Tourism: private boating" ] <- "non-extractive"
MPAthreats$Type[MPAthreats$Threats == "Tourism: commercial boating" ] <- "non-extractive"
```
MPAthreats\$Type[MPAthreats\$Threats == "Tourism: bathing/trampling" ] <- "non-extractive"

## ##

# 1. Threat Intensity

## ##

# Threat intensity (TI) was calculated as TI ijk= T ijk/A jk; # where T ijk is the value of threat i in zone j of MPA k, and A jk is the area (km2) of zone j in MPA k.

MPAthreats\$TI= MPAthreats\$Value/MPAthreats\$Zone.size

# To compare threat intensity amongst protection levels, we have then normalized threat # intensity values (TI) by diving each value with the maximum threat intensity value (TI) # of each threat, which resulted in a standardized scale of  $[0,1]$ .

# First, we extract the maximum threat intensity of each threat Threat\_Max<-aggregate(MPAthreats\$TI,by=list(Threats=MPAthreats\$Threats),FUN=function(x) c(max(x,na.rm=TRUE)))

# We merge this information with our database MPAthreats<-merge(MPAthreats,Threat\_Max,c("Threats"),all.x=TRUE) names(MPAthreats)[14]<-"TI.Max" # name the new column

#divide each threat intensity with the maximum intensity of that threat MPAthreats\$norm.TI = MPAthreats\$TI/MPAthreats\$TI.Max

# what we have now is normalized threat intensity values ranging from 0-1. # we have used this to generate descriptive statistics in Appendix C and Figure 1

## ##

# 2. Threat reduction capacity

## ##

# Threat reduction capacity (TR) was calculated as: # TR ijk = - (1- (TI jk/ TI o,ik)) # where TI ijk is the intensity of threat i inside zone j of MPA k, # and TI o,ik is the threat intensity of threat i outside of the MPA k.

threat.red = MPAthreats # create another table with same data

# We are comparing threats from inside and outside the MPA, so we need to split our data set between inside and outside of the MPA

Threat\_Out<-threat.red[threat.red\$Protection.level=="Outside",] # separate the Outside Values Threat\_Out<-Threat\_Out[,c("MPA.name","Threats","Value", "TI")] # retain the important columns

threat.red<- threat.red[threat.red\$Protection.level!="Outside",] # keep only Inside values

reduction<-merge(threat.red,Threat\_Out,c("MPA.name","Threats"),all.x=TRUE) # combine the Outside and Inside values

# Renaming columns names(reduction)[13]<-"TI.In" # threat intensity inside names(reduction)[17]<-"TI.Out" # threat intensity outside

## Threat reduction capacity

```
reduction$TR = -(1-(reduction$TI.In)/(reduction$TI.Out))
# this data was used to generate Figure 2.
# The descriptive statistics can be found in Appendix C
########################################## ########################################## 
# 3. Zone threat reduction score
########################################## ########################################## 
# The zone threat reduction score (zT) was calculated as:
# zT jk= sum (TR ijk* wi) / sum (wi)
# where TR ijk is the threat reduction capacity defined above and wi is the weight associated 
to threat i (Table 1).
# First, we calculate the zone threat reduction score for each threat type separately 
(extractive, non-extractive)
zone = reduction # create new table to work with
# Calculate numerator 
zone$wTR = (zone$TR) * (zone$Weight) 
numerator = ddply (zone,c("MPA.name", "Protection.level",
                                 'Zone.size', "MPA.size", 'Type', 'Zone.name'), summarise,
                   num = sum(wTR, na.rm = T))# Calculate denominator 
# Now we need to divide the wTR with the sum of the weights in each zone
# in some cases, the threats are not present, so we should not account for them
# to do this we assign a 0 value to the weight when no threat is present
zone$Weight.zone= ifelse (is.nan(zone$TR), 0, zone$Weight)
## now we sum the Weight.zone variable in each zone
denominator = ddply (zone,c("MPA.name", "Protection.level",
                                  'Zone.size', "MPA.size", 'Type', 'Zone.name'), summarise,
                      den = sum(Weight.zone, na.rm = T))
# Now we merge the tables with the numerator and denominator
Zone.score=merge(numerator, denominator)
Zone.score$zT=Zone.score$num/Zone.score$den 
# Now we calculate the zone threat reduction score for the all threats combined 
# same as above
# sum wTr in each zone 
numerator.all = ddply (zone, c("MPA.name", "Protection.level",
                           'Zone.size', "MPA.size", 'Zone.name'), summarise,
                   num = sum(wTR, na.rm = T))## now we sum the Wight.all variable in each zone
denominator.all = ddply (zone,c("MPA.name", "Protection.level",
                             'Zone.size', "MPA.size", 'Zone.name'), summarise,
                     den = sum(Weight.zone, na.rm = T))
```
# Now we merge the tables with the numerator and denominator

Zone.score.all=merge(numerator.all, denominator.all)

Zone.score.all\$zT=Zone.score.all\$num/Zone.score.all\$den

## ## # 4. MPA threat reduction score ## ## # The MPA threat reduction score is calculated (mpaT) # mpaT k= sum (zT\*A jk) / sum (A jk) # where is the threat reduction score of zone j of MPA k, and A jk is the surface area of zone j of MPA k. # First, we calculate the MPA threat reduction score for extractive and non-extractive threats, separately MPA.type = Zone.score # create a new table to work on # calculate numerator MPA.type\$zT.Area = MPA.type\$zT \* MPA.type\$Zone.size MPA.type.score = ddply ( MPA.type, c("MPA.name", 'Type'), summarise,  $num = sum(zT.Area)$ , den = unique (MPA.size)) MPA.type.score\$mpa.T.type = MPA.type.score\$num / MPA.type.score\$den # Now we calculate the MPA threat reduction score for the threats combined  $MPA. all = Zone.score. all # create a new table to work on$ # calculate numerator MPA.all\$zT.Area = MPA.all\$zT \* MPA.all\$Zone.size MPA.all.score = ddply ( MPA.all, c("MPA.name"), summarise,  $num = sum(zT.Area)$ , den = unique (MPA.size)) # Calculate the MPA threat reduction score for extractive and non-extractive threats separately MPA.all.score\$mpa.T = MPA.all.score\$num / MPA.all.score\$den ## ## 5. Local threat index ## ## # The local threat index was calculated as: # lTI ik = TI o,ik / max TI o,i # where the T Io,ik is the intensity of threat i outside MPA k and is the maximum intensity value of threat i recorded from all the outside areas of all considered MPAs # We have already calculated the intensity of each threat, so we will now just extract it local = MPAthreats[MPAthreats\$Protection.level=='Outside',] # The numerator is the variable TI # The denominator: we need to extract the maximum threat intensity of every threat from the outside area

Threat\_OutMax<-aggregate(local\$TI,by=list(Threats=local\$Threats),FUN=function(x) c(max(x,na.rm=TRUE))) #merge tables local<-merge(local,Threat\_OutMax,c("Threats"),all.x=TRUE) names(local)[16]<-"Outside\_Max" # Now we can look at the regional context of each threat local\$LocMax = local\$TI/local\$Outside\_Max # this would be used to plot with the threat figure ## ## # 6. MPA local threat index ## ## # Calculated as: # mpa.lTI k = sum (local threat index \* weight) / sum (weight) # where lTIik is the local threat index of threat i of MPA k, as calculated above, and wi is the weight of threat i. # MPA local threat index for extractive and non-extractive threats, separately local\$w.LocMax= local\$LocMax \* local\$Weight Type.local = ddply (local,c("MPA.name", 'Type'), summarise, sumLocMax = sum(w.LocMax, na.rm = T)) # denominator for the two threat types Type.local.weight = ddply (local,c("MPA.name", 'Type'), summarise, W = sum(Weight, na.rm = T)) Type.local=merge(Type.local, Type.local.weight, all=TRUE) # MPA local threat index for extractive and non-extractive threats separately Type.local\$mpa.local.type = Type.local\$sumLocMax / Type.local\$W # MPA local threat index for all threats combined All.local = ddply (local,c("MPA.name"), summarise,  $sumLocMax = sum(w.LocMax, na.rm = T))$ # denominator All.local.weight = ddply (local,c("MPA.name"), summarise,  $W = sum(Weight, na.m = T))$ All.local=merge(All.local, All.local.weight, all=TRUE) # MPA local threat index for all threats combined All.local\$mpa.local = All.local\$sumLocMax / All.local\$W