



**HAL**  
open science

# How do multi-criteria assessments address landscape-level problems? A review of studies and practices

Sandrine Allain, Gaël Plumecocq, Delphine Burger-Leenhardt

► **To cite this version:**

Sandrine Allain, Gaël Plumecocq, Delphine Burger-Leenhardt. How do multi-criteria assessments address landscape-level problems? A review of studies and practices. *Ecological Economics*, 2017, 136, pp.282-295. 10.1016/j.ecolecon.2017.02.011 . hal-01605875

**HAL Id: hal-01605875**

**<https://hal.science/hal-01605875>**

Submitted on 26 May 2020

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - ShareAlike 4.0 International License

1 **How do multi-criteria assessments address landscape-level problems? A review of**  
2 **studies and practices**

3

4 Sandrine ALLAIN

5 Gaël PLUMECOCQ

6 Delphine LEENHARDT

7

8 Professional affiliation:

9 AGIR, Toulouse University, INRA, INPT, INP-EI PURPAN

10

11 Address:

12 BP 52627, F-31326 Castanet-Tolosan cedex, France

13

14 Corresponding author: Sandrine ALLAIN

15 Contact: sandrine.allaingranet@gmail.com

16

Comment citer ce document :

Allain, S. (Auteur de correspondance), Plumecocq, G., Burger-Leenhardt, D. (2017). How do multi-criteria assessments address landscape-level problems? A review of studies and practices. Ecological Economics, 136, 282-295. DOI : 10.1016/j.ecolecon.2017.02.011

17 **ABSTRACT:**

18 Viewing the landscape as a spatialized social-ecological system allows identification of  
19 specific management challenges: integration of multiple views, multiple levels of  
20 organization, complex spatial-temporal patterns and uncertainties. Multi-criteria assessments  
21 (MCAs), which allow the comparison of alternative actions when multiple interests collide,  
22 are considered adequate to support landscape management. However, there is no consensus  
23 about how they should be applied and can integrate both multiple views and spatial  
24 dimension. We conducted an extensive quantitative and qualitative literature review targeting  
25 MCAs with a participatory and spatial approach. Our results suggest that (1) for sustainability  
26 assessments, participatory and spatial approaches endorse different rationales and hybrid  
27 methods are not so common; (2) within those methods, only scenario-selection methods (as  
28 opposed to design methods) can integrate spatially-explicit, spatially-implicit, place-specific,  
29 and overall values; and (3) current applications, which aggregate values ignoring their spatial  
30 and social distribution, do not coincide with the nature of landscape-management challenges.  
31 In addition, they give little importance to the structuration of information and to collective  
32 deliberation. We conclude that, in the absence of a good match between spatiality and  
33 participation, MCAs should, for now, be handled as insightful but distorted tools to explore  
34 and structure landscape-level management problems.

35  
36

37 **KEYWORDS:**

38 landscape management, integrated assessment, spatialization, participation, social-ecological  
39 system

40

41        1. **INTRODUCTION**

42 Landscapes are complex social-ecological systems (SES) because many human and natural processes  
43 mutually interact (Bastian, 2001; Cumming et al., 2012; Naveh, 2000; Wu, 2006) and because they are  
44 shaped by a social history (Antrop, 2000; Pedroli et al., 2006), which promotes perceptions, values or  
45 expectations that differ spatially and among individuals. This makes it challenging to collectively  
46 define a desirable future for a given landscape. Many possible ways exist to address this issue; one of  
47 them is multi-criteria assessment (MCA). Nonetheless, there is no consensus on how to apply this  
48 method given the nature of landscape-level challenges.

49 A first challenge comes from the complexity of processes and multiple interactions. Because  
50 landscapes are characterized by interdependencies between human societies and their environment that  
51 originate from a coevolutionary history (Berkes et al., 2000; Costanza et al., 1998; Daily, 1997; Kallis  
52 and Norgaard, 2010), they exhibit non-linear and cascading effects that make their trajectories of  
53 change impossible to predict (Kinzig et al., 2006; Walker et al., 2004). It is generally acknowledged  
54 that command-and-control approaches (Folke et al., 2002; Holling and Meffe, 1996) and risk  
55 assessment (Linkov et al., 2014) are ill-suited in this context. Instead, scholars advocate exploring  
56 possibilities for change (Berkes et al., 2002; Olsson et al., 2006) and adopting an adaptive  
57 management approach (Linkov et al., 2006; Plummer, 2009).

58 Another challenge with landscape-level problems arises because the different groups concerned hold  
59 different and sometimes irreconcilable values (Gómez-Sal et al., 2003; Hunziker et al., 2008;  
60 Swedeen, 2006). This problem of multiple views, which is common to many decision-making  
61 situations, has led to recommendations to include non-experts in evaluations of the quality of decisions  
62 (Funtowicz and Ravetz, 1990) and explicitly consider incommensurabilities<sup>1</sup> (Martinez-Alier et al.,  
63 1998; Munda, 2004). Deliberative approaches and MCAs involving multiple stakeholders are  
64 considered particularly well-suited to operationalize these principles (Frame and Brown, 2008; Munda,  
65 2004).

---

1    1 Specific concepts are defined in the appendix.

66 The landscape is a complex system that has another challenging characteristic: material resources and  
67 populations are distributed in space. Human and natural systems can interact “through” the spatial  
68 dimension: social and ecological processes increasingly overlap as perspective widens to a global scale  
69 (Alessa et al., 2008), and spatial mismatches can have far-reaching consequences (Cumming et al.,  
70 2012; Wilson et al., 1999). Likewise, human and natural systems can interact “within” the spatial  
71 dimension: processes such as species migration, farming dynamics or social exclusion are closely  
72 related to spatial patterns, such as habitat heterogeneity and the spatial distribution of crops or  
73 infrastructure networks (Benoît et al., 2012; Cumming, 2011). Accounting for complexity at the  
74 landscape level therefore requires considering these different spatial interactions. This is one reason  
75 why geographic information systems (GIS), given the wide possibilities they offer to investigate  
76 spatial relationships, have become key tools to analyze and resolve landscape-level management  
77 problems (Malczewski, 2006; Malczewski and Rinner, 2015).

78 Because social-ecological interactions have a spatial dimension, diverse and potentially conflicting  
79 representations of space, i.e. new types of incommensurabilities, coexist within a landscape. Because  
80 people relate to places in many different ways – not only through their actions, but also through their  
81 perceptions and history (Antrop, 2005) – they do not have the same definition of boundaries,  
82 meaningful zoning, significant places, features of identity, etc. The same occurs with expert  
83 descriptions: relevant extents, resolutions and locations differ when describing water dynamics or  
84 pollination. The ecological economics community does not formally address these types of  
85 incommensurabilities specific to spatial problems, though it is aware of “scale biases” when  
86 stakeholders express value judgments (Hein et al., 2006; McFadden, 1994; Zia et al., 2011).  
87 Incommensurabilities are not well integrated into spatial decision support systems either, because the  
88 latter are designed as “expert systems” that rely on a uniform understanding of space (Ramsey, 2009).  
89 Applying MCAs to landscape management problems raises the fundamental challenge of integrating  
90 spatiality with multiple views. In an initial step to meet this challenge, we investigate current practices  
91 of MCA reported in the scientific literature that combine a spatial approach with multi-stakeholder or  
92 participatory approaches. More specifically, we address the following issues: how, and how well,  
93 MCA practices reflect landscape-specific challenges.

- 94 These issues are addressed following three nested analyses that enable us to:
- 95 (i) Position spatial and participatory approaches within the broad scope of multi-criteria methods
- 96 (ii) Distinguish types of MCA methods that combine spatial and participatory approaches
- 97 (iii) Clarify how MCAs are applied to assess landscape-management scenarios
- 98 (iv) Generate suggestions for using MCAs at the landscape level.
- 99

Comment citer ce document :

Allain, S. (Auteur de correspondance), Plumecocq, G., Burger-Leenhardt, D. (2017). How do multi-criteria assessments address landscape-level problems? A review of studies and practices. *Ecological Economics*, 136, 282-295. DOI : 10.1016/j.ecolecon.2017.02.011

100

## 101 1. MATERIALS AND METHODS

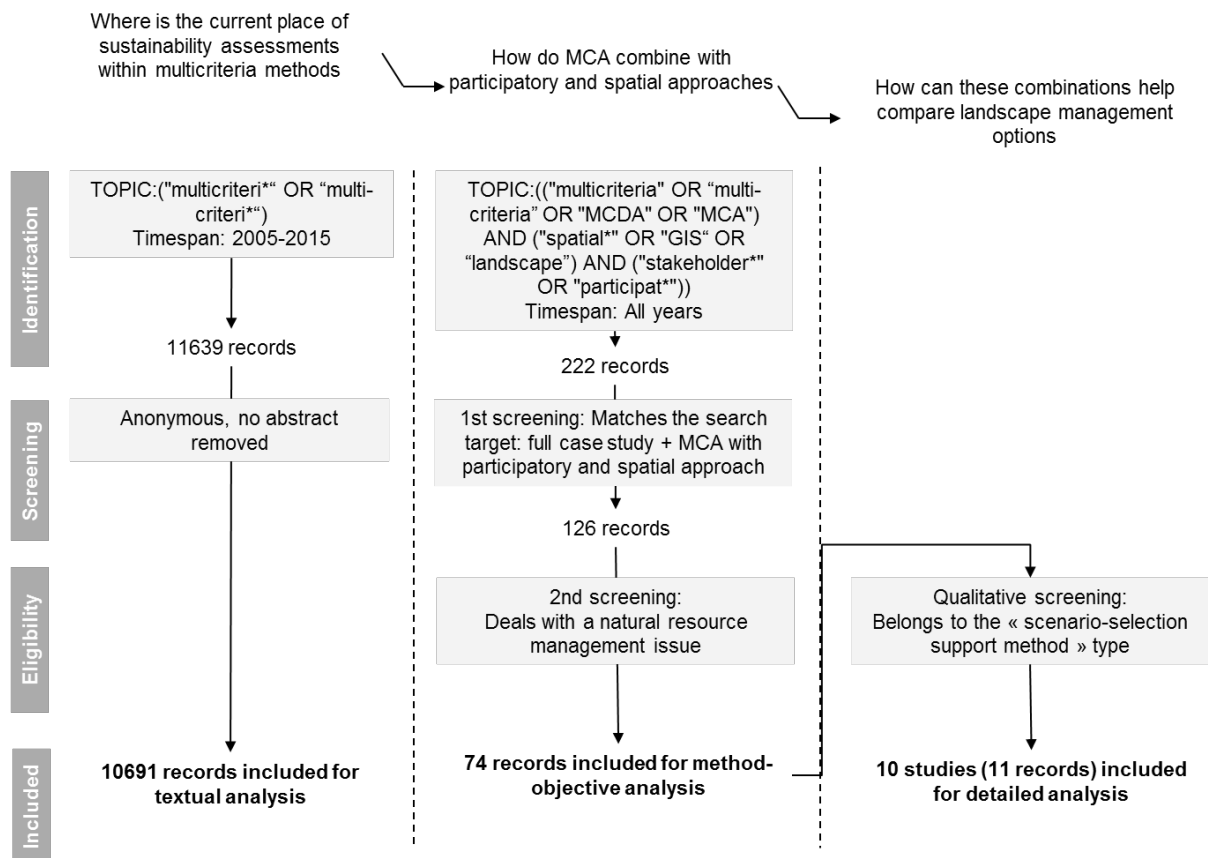
102 This literature review follows three steps (the overall method is described in Fig. 1). First, we  
103 performed a lexicometric analysis of a large sample of studies to characterize the position of  
104 sustainability assessments, participatory approaches and spatial approaches within the wide  
105 spectrum of multi-criteria methods (Section 2.2). Second, we qualitatively classified  
106 applications of multi-criteria approaches mixed with participatory and spatialization methods  
107 in the field of natural resource management (Section 2.3). Our aim was to provide a typology  
108 of existing methods with their general steps. Third, we focused on a specific type of methods  
109 arising from the typology, “scenario-selection support method”, and undertook a detailed  
110 qualitative analysis of the corresponding case-studies (Section 2.4).

### 111 1.1. *Bibliographical data.*

112 We generated three datasets of studies of decreasing size using the Web of Science database,  
113 corresponding to the three steps of our review. For better traceability, we summarized this  
114 selection process in a PRISMA diagram<sup>2</sup> (Moher et al., 2009) (Fig. 1). A large dataset of  
115 abstracts (10,691) was selected to analyze recent trends (2005-2015) in sustainability  
116 assessments within multi-criteria methods. For the two subsequent steps, the time span was  
117 extended to all available years (1975-2005) to embrace a wider diversity of research;  
118 nonetheless, few records were published before 2005.

---

2 2 The “PRISMA (Preferred Reporting Items for Systematic Reviews and Meta-Analyses) statement” includes a  
3 flow chart that maps out the number of records identified, selected for analysis and excluded. The general aim is to  
4 improve the reporting of systematic reviews and to help the reader identify bias in the selected materials.



119 Fig. 1 PRISMA flow diagram for each objective addressed in the review (database searched in June 2015).

120 The three steps reflected in the PRISMA diagram are detailed in the three following sub-  
 121 sections.

122 1.2. **Step 1. Quantitative analysis of multi-criteria methods.**

123 Lexicometric analysis compresses complex information contained in large numbers of texts.  
 124 Reorganizing and classifying segments of texts and providing textual statistics promotes  
 125 understanding of the meaning and context of use of utterances (Lebart et al., 1998). We used  
 126 this method to identify trends in the current use of multi-criteria methods throughout the  
 127 observation of lexical similarities among abstracts.

128 We collected information that characterizes the heterogeneity of research referring to multi-  
 129 criteria methods. We compiled the 10,691 abstracts we selected into a single textual corpus  
 130 and used IRaMuTeQ software (Ratinaud and Marchand, 2012) to implement Reinert's method  
 131 (1993) of textual clustering. This clustering technique proceeds from a contingency table



132 showing the presence or absence of words (except pronouns, conjunctions, and some  
133 adjectives) in a given abstract. IRaMuTeQ clusters abstracts by performing descending  
134 hierarchical classification iteratively: abstracts are divided into two significantly different  
135 groups ( $p$ -value  $< 0.05$ ) according to the presence or absence of words, and then the cluster  
136 with the most abstracts is divided into two new significantly different groups, etc. Clustering  
137 stops when the number of clusters predefined by the analyst is reached or when the largest  
138 cluster contains no significantly different vocabulary. A chi-square test is then calculated for  
139 the classification of each word to assess the significance of the association between a word  
140 and a semantic class. This enabled us to obtain a broad view of multi-criteria methods and  
141 then identify subgroups relevant for analysis.

142 We also assessed the classification via correspondence analysis, in which variability in the  
143 vocabulary and its distribution among clusters (assessed as a chi-square value) is statistically  
144 explained by inertia. For each inertia factor, correspondence analysis provides a statistical  
145 basis for interpreting why certain clusters are similar to or different from others.

146

147 **1.3. Step 2. Qualitative classification of multi-criteria assessments that use**  
148 ***participatory and spatial approaches.***

149 In the second step of our analysis, we emphasize the participatory and spatial approaches of  
150 MCA.

151 We collected research that links these three aspects (i.e., multi-criteria, spatialization,  
152 participation) through a database search and manual screening. “Spatialization” meant that  
153 alternatives and/or the assessment results had to be spatialized; “participation” meant that  
154 people other than the authors had to be involved at some stage of the research. We also  
155 removed all reviews, theoretical developments and studies that introduced a multi-criteria tool  
156 without applying it (even though they could provide interesting insights for interpretation).

157 Among the 222 studies collected, 126 met those eligibility conditions, and among them 74  
158 belonged to the field of natural resource management. The others referred to waste- or  
159 pollution-management issues, urban or infrastructure development and health.

160 We systematically read the abstracts of the 74 studies dealing with natural resource  
161 management, with a special focus on their objectives and methods; when necessary, we also  
162 read the studies' methods sections. We focused more on the stages in which the inclusion of a  
163 spatial or participatory component to MCA adds to or modifies the way the method is  
164 developed. We summarized the most frequent method patterns according to their objective.

165

#### 166 **2.4. Step 3. Systematic qualitative analysis of case studies for scenario** 167 **exploration.**

168 To address our third objective, we analyzed a subset of the previous studies. We only  
169 investigated articles that developed a method for exploring landscape management  
170 alternatives and possibly selecting one of them. We ended up with 10 studies, some of them  
171 covered by more than one article. When necessary, we used additional articles referring to the  
172 same research project and contacted the corresponding author to clarify certain points or to  
173 obtain additional information.

174 Two types of characteristics were examined for each study. First, general characteristics of  
175 MCA were analyzed: the problem structure (i.e. alternatives, criteria, stakeholders and their  
176 judgments) and the decision analysis (the weighting system and the aggregation procedure  
177 used to produce the final evaluation of alternatives) (Kiker et al., 2005; Malczewski, 1999).  
178 Second, we investigated characteristics related to the landscape viewed as a spatialized SES,  
179 focusing on participation of multiple stakeholders (i), representation of multiple levels in the  
180 assessment(ii), consideration of spatial and temporal patterns and relationships (iii), and  
181 management of uncertainties (iv).

182 We created an analytical table (provided in the appendix) with columns referring to the 10  
183 studies reviewed and rows featuring the characteristics of each case-study in terms of:

- 184 - general characteristics (objective, type of issue addressed, area)
- 185 - problem structure (definition of alternatives, criteria (and indicators), stakeholders,  
186 value judgments)
- 187 - decision rules (weighting, aggregation method)
- 188 - involvement of stakeholders (participants, stages with participation, differentiation  
189 between participant input, participatory settings)
- 190 - multi-level system (levels of assessment, upscaling methods)
- 191 - spatiotemporal patterns (knowledge sources for spatial data, accounting for spatial  
192 patterns (heterogeneity, distribution of values), use of visualization tools, accounting  
193 for temporal behavior)
- 194 - integration of uncertainties (sensitivity analysis, uncertainties in outcomes,  
195 uncertainties/inconsistencies in judgments, flexibility of tools)

196 This way of synthesizing information allowed us to identify trends and gaps in the ways  
197 landscape-management options are assessed with a multi-criteria approach.

198 3. **RESULTS AND DISCUSSION**

199 **3.1. Results of lexicometric analysis.**

200 3.1.1 *Main clusters of multi-criteria methods: the vocabulary of sustainability*  
 201 *assessments*

202 In the abstract corpus (10,691 records), we performed a three-cluster classification, which  
 203 allowed 9,450 abstracts (88%) to be classified (Fig. 2). The most significant forms of words,  
 204 tool-words excluded, define three distinct “lexical worlds”.

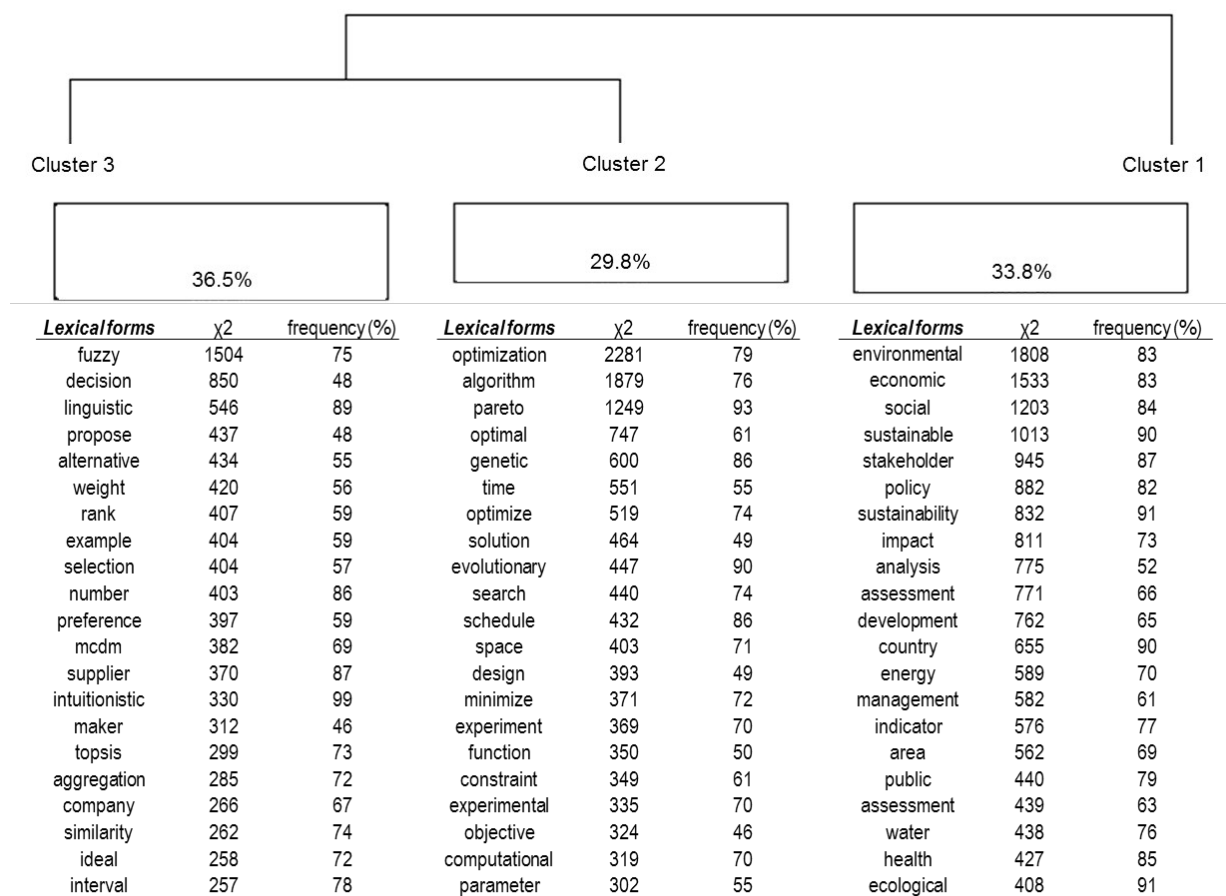


Fig. 2 Dendrogram resulting from descending classification of the abstract corpus, showing the 20 most significant lexical forms of each cluster. Frequency is the number of abstracts classed in a given cluster that contains at least one time a given word divided by the total number of abstracts. Chi-squared values assess the significance of the association between a given word form and a cluster (a chi-squared value of 3.84 corresponds to a probability of 5% that the association of a word to a certain category occurred randomly)

205  
 206 We identified a first cluster of abstracts (3,189 records, 34% of the total) with the most  
 207 significant forms relating to:

- 208 - integrating economic, environmental, social and institutional issues (“environmental”,  
 209 “economic”, “social”, “policy”, “public”)  
 210 - considering future consequences and uncertainties (“impact”, “assessment”, “options”,  
 211 “management”, “potential”)  
 212 - engaging the public (“stakeholder”, “public”)

213 Gasparatos et al. (2008) depicted these different aspects as shared properties of sustainability  
 214 assessments and added equity considerations (intra- and inter-generational) to this list. Equity  
 215 considerations, however, are not directly reflected by the most significant words in the  
 216 “sustainability assessment” cluster. The first word in this semantic field, “equity”, appears in  
 217 312<sup>th</sup> position (still significant for this cluster) because it occurs rarely. We conclude that  
 218 social justice issues are, in most cases, absent from sustainability assessments based on a  
 219 multi-criteria method. Sustainability remains mainly understood as the fulfilment of  
 220 competing goals rather than a trajectory of change that reconfigures social interactions.  
 221 A second cluster of abstracts focuses on optimization problems (“optimization”, “Pareto”,  
 222 “solution”, “design”, etc.) aiming at designing Pareto-optimal<sup>3</sup> solutions for a limited set of  
 223 objectives and constraints (Linkov et al., 2006). Decision-analysis procedures (“decision”,  
 224 “rank”, “selection”, “aggregation”) are grouped together in the last cluster of abstracts (cluster  
 225 3). Considering that no optimal solution can be found, a variety of procedures exist to  
 226 rank/sort alternatives (Guitouni and Martel, 1998), each of them corresponding to a specific  
 227 definition of what makes the best compromise (the closest from an ideal point, the one that is  
 228 not outranked by others, etc.).

229

### 230 3.1.2. *Rationales underlying MCA methods: the specificity of spatial approaches*

---

5 <sup>3</sup> Resource allocation is considered optimal (*sensu* Pareto) when an increase in one objective cannot be achieved  
 6 without worsening another objective. The total number of situations that satisfy the Pareto criterion forms a  
 7 boundary limit from which no improvement can be made given the available resources. The existence of such a  
 8 boundary assumes that objectives are perfectly substitutable.

231 We generated an additional classification to see how the sustainability assessment cluster  
 232 would split up, adding two more successive phases of clustering. The abstract corpus was  
 233 consequently divided into 5 classes (8,964 abstracts included in the new classification, ~84%  
 234 of the corpus). We were then able to distinguish two new subgroups of abstracts: strategic  
 235 business decision-making (within the “decision analysis” cluster, former cluster 3) and spatial  
 236 approaches (within the “sustainability assessment” class), the latter being of special interest  
 237 for our purpose. The “optimization problem” cluster, (former cluster 2) remained unaffected  
 238 by this new classification. Projection on the two first axes of the correspondence analysis  
 239 (Fig. 3), which summarize 69% of the total inertia, reveals that spatial assessments constitute  
 240 both a marginal (approx. 1/4 of the abstracts on sustainability assessments) and specific  
 241 branch of sustainability-assessment approaches.

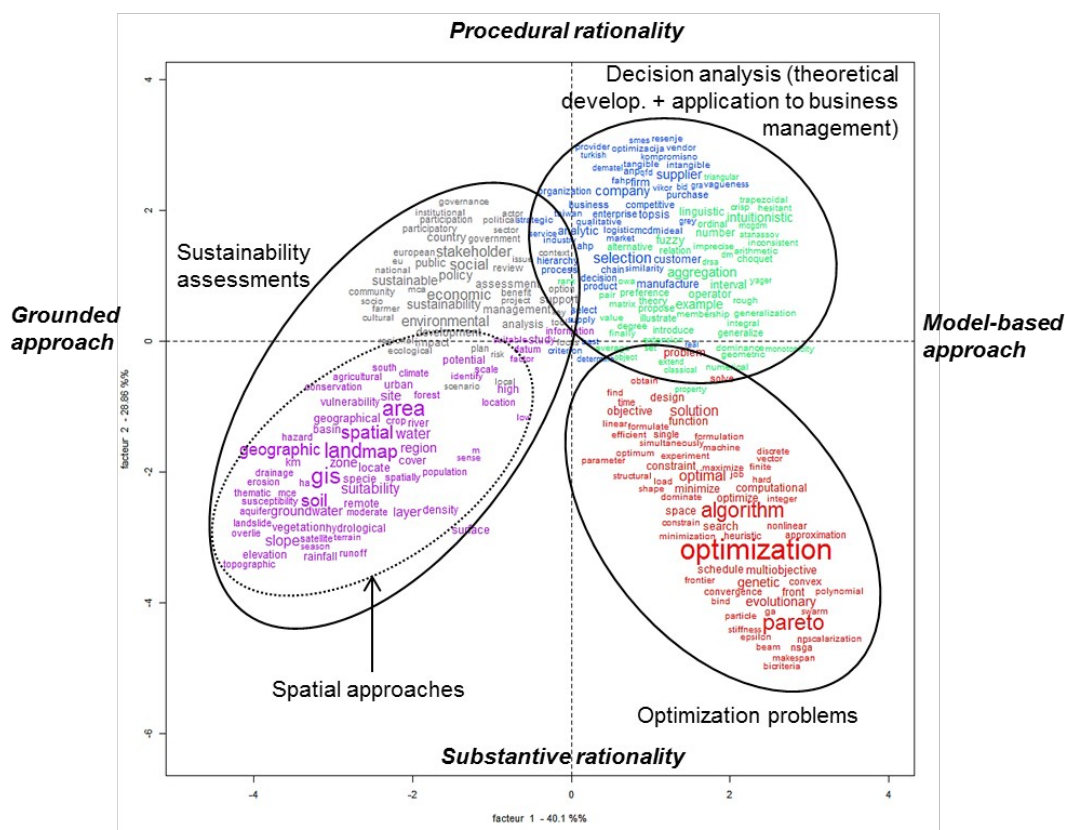


Fig. 3: Distribution of active forms along the two first axes of the correspondence analysis (5-class clustering). These axes do not separate the classes of decision analysis theories and their application to business management.

242

243 We interpret the two first factors as reflecting what is understood as a rational decision (axis  
244 2, 29% of total variance) and what makes a scientifically-sound decision (axis 1, 40% of total  
245 variance).

246 The first axis differentiates multi-criteria methods that rely on grounded, contextualized  
247 information from those that mainly rely on mathematical equations to find a solution. By  
248 “grounded”, we mean that theory proceeds from the accumulation and analysis of place-based  
249 data. Since they seek to capture multiple indicators and multiple views, sustainability  
250 assessments imply such bottom-up view of science, providing they not only try to formulate  
251 ad-hoc solutions but also create new knowledge. The opposite pole of axis 1 represents  
252 another view of how science interacts with decision-making, in which there are “rules of  
253 nature” that can be formalized (by scientists) and then applied to any situation (e.g. to support  
254 decision-making). According to this view, generic models and algorithms are the elements  
255 that validate the results obtained. Decision analysis, which focuses on finding the best  
256 compromise when no optimum solution exists, and optimization methods, which focus on  
257 finding solutions that maximize utility under a set of constraints, are two different methods  
258 that share the assumption that mathematical formalization is the best way to reach a valid  
259 solution.

260 Axis 2 reflects the opposition between substantive rationality and procedural rationality  
261 behind multi-criteria methods (Simon, 1976). Under a substantive rationality hypothesis,  
262 rational behaviors allow the fulfillment of a set of goals under a set of constraints (from the  
263 external environment, its perception, or internal characteristics of the decision maker). This  
264 realm assumes a “heroic” (Béjean et al., 1999) decision-maker who, according to Simon  
265 (1990: 195): “contemplates, in one comprehensive view, everything that lies before him. [...]”  
266 *He has reconciled or balanced all his conflicting partial values and synthesized them into a*  
267 *single utility function that orders, by his preference for them, all these future states of the*

268 *world*". This idea is clearly embedded in optimization procedures. However, the substantive  
269 rationality paradigm is also dominant in spatial assessments. The use of spatial data is thought  
270 to "empower the decision-maker" (Densham, 1991) by providing both more precise and more  
271 usable information. The environment (objective and subjective) is considered fixed and the  
272 coexistence of different value schemes and conflicting preferences not allowed.

273 In the realm of procedural rationality, behaviors are rational because they come from an  
274 appropriate individual deliberative process in the face of incomplete information, an unknown  
275 future, or shifting beliefs and values (Simon, 1976). Studies related to decision analysis, since  
276 they model uncertainty and the interaction of multiple views towards a compromise, typically  
277 assume a procedural rationality (extended from the individual to the group) (Garmendia and  
278 Gamboa, 2012; Munda, 2004). Sustainability assessments following a participatory approach  
279 (Fig. 3, upper-left corner) also fall in this category because deliberation is the pivotal process  
280 for integrating heterogeneous information and values (Vatn, 2005). Nonetheless, these  
281 approaches often fail to accommodate the evolutionary nature of preferences and the  
282 existence of non-utilitarian values (values that cannot be reflected in competing individual  
283 interests). Rather, they assume that one's motivation can be captured in a unique order of  
284 preferences (Béjean et al., 1999). They therefore only partially fall under procedural  
285 rationality paradigm.

286 We have just demonstrated how two branches of sustainability assessment distinguish  
287 themselves. Though common in their place-based anchorage, sustainability assessments  
288 following a spatial approach and those following a participatory approach endorse  
289 fundamentally different philosophies about what constitutes a good decision (the process or  
290 the availability of information) and the context under which a decision is to be taken  
291 (irreducible uncertainties and conflicting values or a well-defined decision space).



292 Nonetheless, bridges between these two branches exist and are crucial if one wants to address  
293 multi-stakeholder landscape-level issues. We examine these bridges in the following section.

294

295 **3.2. How MCA combines with participatory and spatial approaches. Synthesis of**  
296 **methods and objectives.**

297 In this section, we present results of our qualitative analysis that classified studies using  
298 participatory and spatial approaches (Fig. 4). We distinguish three types of objectives, hence  
299 three types of methods: scenario-design support, scenario-selection support, and spatial  
300 assessment of real situations. The categories of both scenario-design support and scenario-  
301 selection support methods are future-oriented, leading to management recommendations.  
302 Method steps can be shared between real-world and future-oriented methods even when their  
303 objectives differ (prescriptive as opposed to descriptive). We found two studies (3 records)  
304 combining scenario-design support and scenario-selection support objectives (detailed in  
305 section 3.3.2).

306 Regarding the two families of future-oriented methods, their differences in objectives are  
307 reflected in different method steps (Fig. 5). Scenario-design support methods deal with a  
308 location or spatial allocation problem (van Herwijnen and Janssen, 2007): Which areas should  
309 be managed or preserved first (prioritization)? How best to allocate land-uses in the landscape  
310 (zoning/ mapping proposals)? Those questions refer respectively to multi-attribute or multi-  
311 objective problems. In the first case, every location is scored according to a set of attributes  
312 (van Herwijnen and Janssen, 2007). Then, value maps are overlaid and aggregated, generally  
313 with a weighted linear combination procedure (Malczewski, 2000), to define the best area(s)  
314 or rank sites for a given purpose. This method is also used for suitability assessments. In the  
315 second case, a set of goals are defined and multi-objective optimization algorithms produce  
316 land-use plans or spatially-explicit management proposals.

Conversely, scenario-selection support methods differentiate and often rank management alternatives based on their effects (impact assessment) and sometimes their implementation (generally cost, but sometimes applicability or flexibility; e.g. Sahin and Mohamed, 2013). A scenario can be evaluated and possibly selected according to both spatially-explicit and non-spatial criteria. An aggregation procedure can be performed first on criteria (producing a landscape-level score for each criteria) or on alternatives (producing total performance maps) (see van Herwijnen, 1999, cited by van Herwijnen and Jansen, 2007), for different possible pathways).

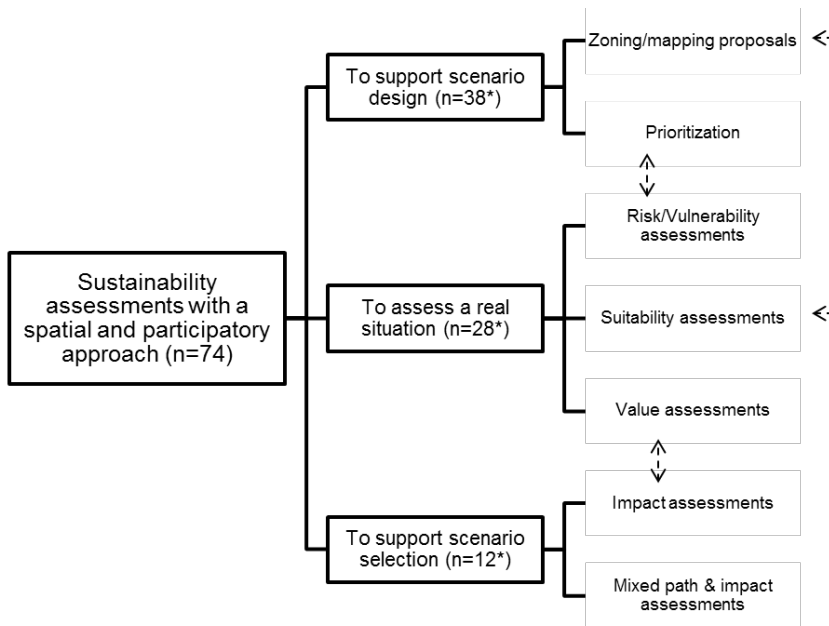


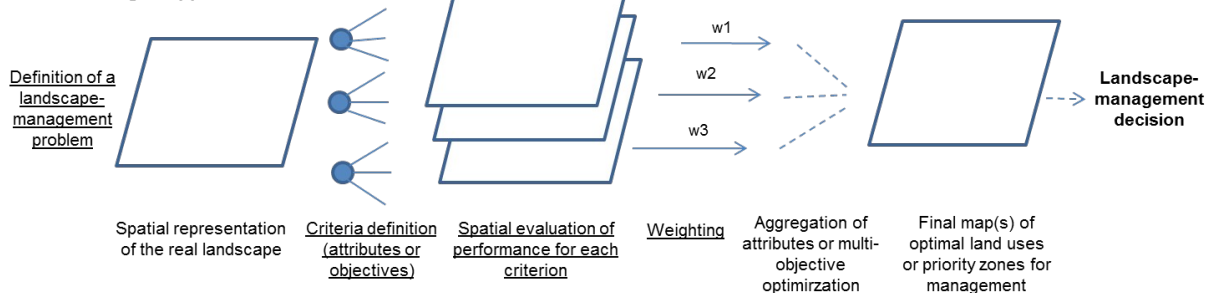
Fig. 4. Classification of sustainability assessment methods combining a participatory and spatial approach according to their general objective for decision-making (supporting scenario design, assessing a real situation and supporting scenario selection) and the type of results produced.

Double-headed arrows link categories that share similar method steps but differ in their objective.

\* 4 records were assigned to 2 different categories.

317

#### Scenario-design support



#### Scenario-selection support

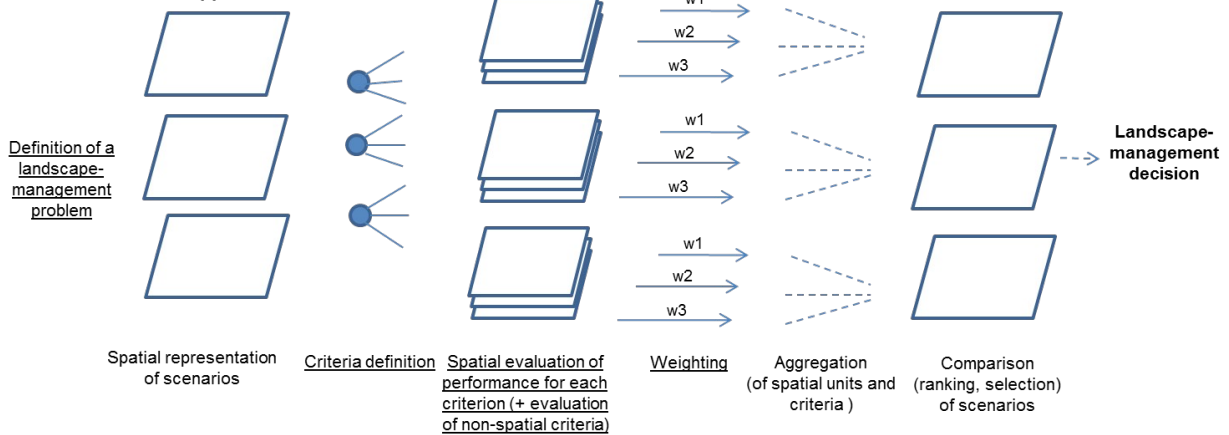


Fig. 5. Classic steps in scenario design support (top) and scenario selection support (bottom). Participatory settings were used at different stages (underlined in the figure) varying between studies.

Spatial diagnosis methods follow the same steps though objectives differ. When the aim is to establish risk, vulnerability or suitability maps, method steps are similar to the ones of the "scenario-design support" method. For value assessment, the steps follow the "scenario-selection support" method except that the input landscape(s) is (are) real.

318

319 The involvement of non-academic participants, including local experts, can occur at different  
320 stages. In such cases, identified criteria and/or weighting are designed to reflect stakeholder  
321 preferences. The issues of how to select participants, how to elicit their preferences and how  
322 to combine them are the same for non-spatial evaluation (see e.g. Garmendia and Gamboa,  
323 2012; Hajkowicz and Collins, 2006; Reichert et al., 2015). Also, some criteria can be  
324 evaluated on a participatory basis when they integrate a sensible dimension (e.g. visual aspect  
325 (Sheppard and Meitner, 2005)); risk perception (Raaijmakers et al., 2008)); place attachment  
326 (Newton et al., 2012; Nordström et al., 2011) or when the evaluation of performances relies  
327 on the statements of local experts (e.g. Arciniegas and Janssen, 2012).

328 Through this clarification exercise, we highlighted consistent patterns in scenario-design and  
329 scenario-selection methods. The way stakeholders are involved does not appear to  
330 differentiate methods. Stages with participation – and the extent of that participation – reflect  
331 the diversity of scientific postures in action-research, which is independent from the  
332 operational objective pursued. In contrast, the position of the MCA within the decision-  
333 making process and the nature of criteria and values do distinguish scenario-design support  
334 from scenario-selection support. In the former, alternatives are generally not defined *a priori*,  
335 and the decision-making process is in its exploratory stage; also, criteria and values are  
336 necessarily spatially-explicit. In the latter, alternatives are established *a priori*. They can be  
337 realistic plans identified by managers (when the decision process is in an advanced stage) or  
338 contrasting images of the future created to support learning. With scenario-selection support  
339 methods, alternatives can be spatially-explicit or not (e.g. a normative change) and the  
340 assessment of alternatives can accommodate both spatially-explicit and non-spatial values.  
341 Assessment is therefore more comprehensive, which suits a holistic and integrative view of  
342 the landscape. Because of these possibilities, we focus on the potential of scenario-selection  
343 methods to compare landscape-management alternatives.

344 3.3. *How MCA combining participatory and spatial approaches can help compare*  
 345 *landscape-management alternatives. Detailed analysis of current practices.*

346

347 3.3.1 *Overview of selected studies*

348 In our corpus, we found 11 studies (12 records) that implement a multi-criteria scenario-  
 349 selection method at the landscape level, i.e. studies that use, at some stage, the input of non-  
 350 academics to compare a set of landscape-management scenarios to highlight those best-suited  
 351 (Table 1). The study of Raaijmakers et al. (2008) was excluded as there was no integration  
 352 between the stakeholder assessment of alternatives (not spatial) and the simulation of their  
 353 impacts using a GIS-based model. The 10 remaining studies deal with land-use planning (4),  
 354 ecosystem conservation (3), water management (2) and forest management (2).

355 Table 1. General description of selected studies (LU = land-use). Italics indicate studies with a “light”  
 356 participatory approach. Additional articles used for the analysis are also mentioned.

Selected studies	Goal	Issue addressed	Geographic area
Ahrens and Kantelhardt, 2009	Assessing the impact of farming production options on landscape functions	Agricultural landscape planning	Bayerisches Donauried (Germany)
Arciniegas et al., 2011	Designing LU plans of increased quality	LU planning	Bodegraven Polder (Netherlands)
Arciniegas and Janssen, 2012	Assessing impacts of water-allocation alternatives on watershed sustainability	Water management	Ichkeul Basin (Tunisia)
Chakroun et al., 2015	Assessing effects of forest LU management on ecosystem services	Regional LU planning	Saxony (Germany)
Fürst et al., 2013	Assessing impacts of alternative water-management regimes on ecological functions	Wetland ecological functioning	Wormer and Jisperveld fen meadow (Netherlands)
Fürst et al., 2010	Assessing impacts of alternative water-management regimes on ecological functions	Wetland ecological functioning	Wormer and Jisperveld fen meadow (Netherlands)
Janssen et al., 2005	Assessing impacts of alternative water-management regimes on ecological functions	Wetland ecological functioning	Wormer and Jisperveld fen meadow (Netherlands)
Goosen et al., 2007	Assessing impacts of alternative water-management regimes on ecological functions	Wetland ecological functioning	Wormer and Jisperveld fen meadow (Netherlands)
Linhoss et al., 2013	Selecting a management strategy for the Snowy Plover conservation	Species conservation and coastal conservation	Florida Gulf Coast (USA)
Manoli et al., 2005	Assessing possible water-management interventions	Water management	Paros Island (Greece)
Newton et al., 2012 (+ supporting information)	Assessing costs and benefits of alternative ecological network plans	Ecological restoration	River Frome watershed (United Kingdom)
Nordström et al., 2010	Designing and assessing alternative forest-management plans	Strategic forest planning	Lycksele forest (Sweden)
Nordström et al., 2011	Designing and assessing alternative forest-management plans	Strategic forest planning	Lycksele forest (Sweden)
Sheppard and Meitner, 2005	Reducing conflicts and assessing alternative forest-management plans	Sustainable forest management	Lemon Landscape Unit (Canada)
Sheppard et al., 2006	Reducing conflicts and assessing alternative forest-management plans	Sustainable forest management	Lemon Landscape Unit (Canada)

357

358 Our sample covers the gradient of research postures that characterizes integrated assessments  
 359 (Barreteau et al., 2010; van Asselt and Rijkens-Klomp, 2002). Some of the studies we  
 360 analyzed hence implemented a “lighter” participatory approach, corresponding to a situation

361 of consultation or information (Barreteau et al., 2010). In these cases, stakeholders are only  
 362 present at the very beginning of the process (associated with generating alternatives) or  
 363 merely exist as assumed recipients/users of assessment results. They can also be absent but  
 364 different values nonetheless considered through archetypical weighting schemes (in the case  
 365 of Linhoss et al., 2013).

366 We identified two studies that combine scenario-selection and scenario-design approaches,  
 367 and hence promote iteration for adaptive management. One (Nordström et al., 2011, 2010)  
 368 used a scenario-design approach to generate and select three realistic but contrasting scenarios  
 369 of forest management, which were consequently assessed. The other (Arciniegas et al., 2011;  
 370 Arciniegas and Janssen, 2012) followed the opposite approach: scenario assessment was used  
 371 as an exploratory phase prior to generating land-use plans. Although they are rarely used  
 372 jointly, scenario-design and scenario-selection methods are not mutually exclusive.

### 373 3.3.2. Results of qualitative analysis

374 Table 2 summarizes the options adopted in the case studies reviewed.

375 Table 2. Qualitative analysis of the multi-criteria structure and incorporation of landscape-management  
 376 characteristics of the 10 selected case studies. Options in bold refer to the most frequently observed  
 377 practice. def.: definition; AHP: analytic hierarchy process; MAUT: multi-attribute utility theory; CP:  
 378 compromise programming; MCDA: multi-criteria decision analysis; SDSS: spatial decision support  
 379 system.

MULTI-CRITERIA STRUCTURE	Alternative def.	<ul style="list-style-type: none"> <li>▪ Directly developed by researchers according to local issues (consultation of stakeholders, literature review, etc.)</li> <li>▪ Developed by selected stakeholders or experts</li> <li>▪ Exploration-Selection process (computer-aided, matrix of scenarios)</li> </ul>
	Criteria def.	<ul style="list-style-type: none"> <li>▪ Defined by researchers (according to local issues)</li> <li>▪ Based on a scientific reference framework (more or less adapted to a specific context)</li> <li>▪ Defined by stakeholders in participatory settings (workshops, interviews)</li> </ul>
	Indicator def.	<ul style="list-style-type: none"> <li>▪ <b>Defined or specified (following consultation) by experts or researchers</b></li> <li>▪ Defined by reference framework</li> </ul>
	Stakeholder def.	<ul style="list-style-type: none"> <li>▪ <b>Absence of a formalized method</b></li> <li>▪ Snowball sampling, Stakeholder analysis followed by grouping and selection of participants</li> </ul>
	Value judgment def.	<ul style="list-style-type: none"> <li>▪ <b>Importance</b> or perceived performance (scoring/allocation of 30 points)</li> <li>▪ Preference (pairwise comparison)</li> </ul>
DECISION RULES	Weighting systems	<ul style="list-style-type: none"> <li>▪ <b>Average of scores</b> (for all stakeholders or at the group level)</li> <li>▪ No weights / Exploratory weights / Weighting schemes reflecting different focuses</li> <li>▪ Group average score x Group importance score</li> <li>▪ Market value (0 weight to non-marketable ecosystem services)</li> </ul>

INV OLV EME NT OF STA KEH OLD ERS	<b>Aggregation method</b>	<ul style="list-style-type: none"> <li>When value maps are produced: <b>weighted summation</b> or absence of aggregated results</li> <li>When assessment is not spatially-explicit : varied methods hypothesizing strong to weak compensation (AHP, MAUT, CP, Electre-II, Stochastic MCDA)</li> </ul>
	<b>Participating stakeholders</b>	<b>Managers</b> , Users of different types, Experts of different types, Representatives of interest groups, Farmers, Landowners
	<b>Differentiation of stakeholder input</b>	<ul style="list-style-type: none"> <li><b>No</b></li> <li>Distinct role of different stakeholder groups</li> <li>Evaluation at the group level (through averaging)</li> </ul>
	<b>Stages with participation</b>	<ul style="list-style-type: none"> <li><b>Criteria def. and/or Alternative def.</b></li> <li>Assessment, Feedback, Selection of participants</li> </ul>
	<b>Participatory methods</b>	<ul style="list-style-type: none"> <li><b>Workshops</b>, Interviews, Surveys/Inquiry</li> </ul>
HIE RAR CHI CAL ORG ANI ZATI ON	<b>Levels of assessment</b>	<ul style="list-style-type: none"> <li>When assessment of criteria is spatially-explicit : <b>basic units/compartments + Landscape/Region</b></li> <li>When assessment is not spatially-explicit : entire landscape/Region</li> </ul>
	<b>Upscaling/downscaling methods</b>	<ul style="list-style-type: none"> <li><b>Area-weighted aggregation of basic units' normalized performance</b></li> <li>Temporal aggregation according to statistical metrics</li> <li>Corrected spatial aggregation according to landscape metrics</li> </ul>
	<b>Knowledge sources of spatial data</b>	<ul style="list-style-type: none"> <li><b>Generic (from database)</b></li> <li>Local (expert opinion, field survey, online stakeholder survey)</li> <li>Simulation outputs</li> </ul>
SPA TIAL & TEM POR AL PAT TER NS	<b>Spatial distribution</b>	<ul style="list-style-type: none"> <li><b>No</b></li> <li>Qualitative analysis of value maps, Use of landscape metrics, Use of spatially-explicit metapopulation models</li> </ul>
	<b>Visualization tools</b>	<ul style="list-style-type: none"> <li><b>Value maps</b></li> <li><b>No</b></li> <li>Landscape visualization</li> </ul>
	<b>Temporal dynamics</b>	<ul style="list-style-type: none"> <li><b>No</b></li> <li>Statistical metrics describing temporal behavior</li> </ul>
	<b>Sensitivity analysis in MCA</b>	<ul style="list-style-type: none"> <li><b>No</b></li> <li>Varying weights, Exploration of the whole weight space, Varying aggregation methods,</li> </ul>
UNC ERT AIN TIES	<b>Uncertainties in outcome</b>	<ul style="list-style-type: none"> <li><b>No</b></li> <li>Assessment under varying futures</li> </ul>
	<b>Uncertainties in judgment</b>	<ul style="list-style-type: none"> <li><b>No</b></li> <li>Consistency ratio to eliminate inconsistent stakeholders</li> <li>Use of different methods to reveal preferences</li> <li>Stochastic MCDA allowing undefined weights</li> </ul>
	<b>Flexibility</b>	<ul style="list-style-type: none"> <li><b>No</b></li> <li>Possible modification in the SDSS, Use of interactive devices</li> </ul>

380 (i) Multi-criteria structure

381

382 The case studies highlight typical limitations of MCAs in general (i.e. not specifically applied  
383 to landscape problems) about problem structuring. It is however acknowledged that the way  
384 in which problems are structured, including how to define alternatives or criteria, how to  
385 determine who is relevant and what legitimacy they have, how to elicit judgments and which  
386 decision rules to choose, is central to any MCA (Choo et al., 1999; Garmendia and Gamboa,  
387 2012; Guitouni and Martel, 1998; Munda, 2004; Vatn, 2009).

388 In the 10 studies we reviewed, alternatives and criteria are either developed by researchers or  
389 by a set of stakeholders. Specific reference frameworks (for ecosystem services, ecological  
390 functions or management standards) are sometimes used. In some cases, definition of  
391 alternatives is a central part of the assessment method and results from a formalized  
392 exploration-selection-specification process (in Fürst et al., 2013; Nordström et al., 2011).  
393 Often, but not always, criteria are translated into indicators (in other cases, this hierarchical  
394 structure is absent); definition of these indicators is managed by researchers and experts,  
395 without feedback from other participants.

396 The process for selecting the stakeholders included in the assessment is often not formally  
397 addressed. Two studies (Nordström et al., 2010; Sheppard and Meitner, 2005), which defined  
398 their study area as conflict-ridden (both around the issue of forest management), used  
399 stakeholder analysis to identify stakeholders and group them according to their assumed  
400 interests.

401 (ii) Weighting and aggregation

402 In 6 of the studies analyzed, criteria weights reflect stakeholder preferences. Individual  
403 elicitation of weights is preferred over collective deliberation, so that group-level values are  
404 actually the average of individual scores. The only case of using collective decision-making to



405 derive weights was found in an expert workshop in Arciniegas and Janssen (2012). Generally,  
406 stakeholders are asked to score criteria according to their importance. In one study  
407 (Nordström et al., 2010), stakeholders were asked to state preferences using a pairwise  
408 comparison of criteria.

409 More originally, Newton et al. (2012) and Linhoss et al. (2013) used archetypical weighting  
410 systems to reflect different ethical positions but not any particular stakeholder preferences.

411 Linhoss et al. (2013) produced weighting schemes for human-focused, bird-focused and  
412 mixed values, while Newton et al. (2012) included a weighting scheme based on purely  
413 monetary valuation.

414 In all the other cases reviewed, authors chose to use equal weights or exploratory weights  
415 (weights are modified to determine how they influence the final ranking).

416 When value maps are produced as intermediate or final results of the assessment, weighted  
417 summing is the common aggregation method. Studies producing spatially-aggregated,  
418 spatially-implicit or non-spatial values to assess each criterion use different decision analysis  
419 possibilities. The majority of those procedures assume that values associated with criteria can  
420 compensate each other (compensatory procedures). Nonetheless, Ahrenz and Kantelhardt  
421 (2009) specifically address the question of compensability between criteria by comparing the  
422 outcomes of an additive model (multi-attribute utility theory) with those of an outranking  
423 model (Electre-II).

424  
425 (iii) Interaction with stakeholders

426

427 In most of the studies we collected, stakeholders of different status and with different interests  
428 in the problem are considered and consulted during workshops. Interviews and surveys are  
429 other observed practices. Most commonly, participating stakeholders are managers (who are  
430 involved in the decision-making process) and representatives of interest groups (representing

431 users – farmers, timber-harvesters, hunters, etc. – or a political view – environmentalists,  
432 bankers, community leaders, etc.); they are involved in the definition of criteria and/or  
433 alternatives. Though researchers generally assign them to a category of stakeholders, Ahrenz  
434 and Kantelhardt (2009) instead let each participant choose the group to which he/she  
435 belonged. A few studies introduce hierarchy among stakeholders, either by differentiating the  
436 stages when they intervene (e.g., in Arciniegas et al. (2011), “public” stakeholders explored  
437 and designed alternatives, researchers introduced criteria, and experts defined indicators and  
438 weights) or by determining importance coefficients (in Nordström et al. (2010)), a steering  
439 group was established to define relevant stakeholder groups and their relative importance).

440 (iv) Multiple levels and spatio-temporal patterns

441 Consideration of the multi-level nature of the landscape system is generally simplified into a  
442 two-scale problem: elementary units and landscape/region. The study of Ahrenz and  
443 Kantelhardt (2009) is an exception, as authors consider an intermediate level – the farm – to  
444 derive production responses and calculate socio-economic indicators. Nonetheless, results are  
445 produced at the landscape-level only.

446 In the spatially-explicit assessments we analyzed, basic units or compartments are defined  
447 according to their homogeneous functioning (Chakroun et al., 2015; Janssen et al., 2005;  
448 Manoli et al., 2005) or their use (Arciniegas and Janssen, 2012). Another option is to use a  
449 cell grid (Fürst et al., 2009; Newton et al., 2012). Those units constitute the resolution of the  
450 spatial assessment. Value maps are then generated, showing the distribution of performances  
451 in the entire landscape. Subsequent spatial aggregation consists of adding the value of all  
452 units (with their size as a coefficient when it is not uniform). Some authors acknowledge the  
453 limits of such additive aggregation. For instance, Janssen et al. (2005) insist on the need to  
454 give disaggregated results to decision-makers, while Fürst et al. (2013) “corrected” the results

455 of spatial aggregation for two ecosystem services using landscape metrics<sup>4</sup>. Limits of spatial  
456 aggregation can also be overcome through an overall evaluation: Sheppard and Meitner  
457 (2005) attempted to do so by asking stakeholders to directly assess semi-realistic landscape  
458 visualizations of competing alternatives.

459 Temporal patterns are conspicuous by their absence, with the notable exception of Manoli et  
460 al. (2005), who first spatially aggregated each criterion considered (e.g., cost, water  
461 availability) and then considered second-level criteria describing temporal behavior.

462 (v) Uncertainties

463 Finally, uncertainties are generally not considered. When they are, the most common practice  
464 is to perform a sensitivity analysis on the weighting systems, which reflect uncertainties in the  
465 preferences for each criteria (Chakroun et al., 2015; Janssen et al., 2005; Newton et al., 2012).

466 Linhoss et al. (2013) chose to use of a stochastic multi-criteria decision analysis in order to  
467 test a case of uninformed weights (the whole weighting space is explored and a rank  
468 acceptability index is calculated, describing the percentage of times one alternative ranks as  
469 “most-preferred”). Uncertainties about future changes are addressed in two studies (Linhoss et  
470 al., 2013; Manoli et al., 2005), in which climate is considered the main driver of change.

471 Uncertainties in judgment are addressed in the two studies that attempted to structure  
472 stakeholder participation in a conflict-ridden context (Nordström et al., 2010; Sheppard and  
473 Meitner, 2005) and in the study of Linhoss et al. (2013) that uses a stochastic procedure.

474 Another way to handle uncertainties is to design a flexible decision-support tool that allows  
475 modifications. Some tools are also user-friendly (e.g. the touchtable of Arciniegas and Janssen  
476 (2009); the "Pimp your landscape" software of Fürst et al. (2009)). It is not clear, however,  
477 whether these tools can be modified by end-users.

---

9 4 Landscape metrics refer to measures or indices that consider the distribution of patches (or groups of patches)  
10 within a landscape mosaic. They can characterize either a landscape's composition (e.g., abundance of forested  
11 patches, evenness of different patch types) or spatial configuration (e.g., measures of patch density, shape, or  
12 connectivity).

478 3.4. ***Research gaps to fill to effectively address landscape-level management problems.***

479 ***Critical discussion of observed practices.***

480 This systematic qualitative analysis enables us to identify and discuss several research gaps in the  
481 application of MCA methods to assess landscape management alternatives.

482 (i) Lack of attention to problem structuring

483 Problems observed in MCA in general are also found in our analysis. Among them, the lack of  
484 attention to problem structuring, which is crucial for establishing a shared platform for deliberation  
485 and effectively linking assessment to decision making, is striking (Borouhaki and Malczewski, 2010;  
486 Giampietro, 2003; Hajkowicz and Collins, 2006; Ramsey, 2009). This loophole is particularly  
487 prevalent in the generation of alternatives, selection of stakeholders, and definition of indicators.

488 The selection of stakeholders, alternatives, criteria and indicators constitute a “compression” of the  
489 information space (Giampietro, 2003) that conditions the quality of the MCA outcome. As shown by  
490 this literature review, alternatives and criteria are often defined with relatively little involvement of  
491 local stakeholders, although the conditions of their involvement are often not fully clarified or  
492 formalized. Institutional analysis – at worst to identify the diversity of interests and “legitimate”  
493 people to participate in the evaluation and at best to gain knowledge about their power relations – is a  
494 noteworthy method to justify the choices made about participants, scenarios and criteria (De Marchi et  
495 al., 2000; Munda, 2004; Xenarios and Tziritis, 2007). Also, if alternatives are used to expand the  
496 decision-space – i.e. alternatives represent contrasting images that stimulate stakeholders’ creativity –  
497 then formalized procedures (computer-assisted or not) that enable moving out of the classic set of  
498 “sketched” alternatives can be useful (Groot and Rossing, 2011).

499 Defining indicators related to criteria is commonly considered the task of researchers and experts,  
500 which is consistent with a positivist view of science. However, using MCA, especially in conjunction  
501 with participatory methods, should force researchers to endorse a post-normal paradigm for evaluating  
502 sustainability (Funtowicz and Ravetz, 1994). Following Le Bellec et al. (2012), we suggest that MCA  
503 would gain in quality – under a post-normal assumption – if stakeholders were involved in the design  
504 and validation of the indicator set. Also, if we acknowledge that a set of indicators reflects a specific

505 way of organizing information (i.e. a specific view of the problem), then it modifies the status of  
506 indicators in MCA: indicators are not only “proxies” that provide a measurement of a criterion’s  
507 performance but are also objects for collective deliberation (Frame and O’Connor, 2011).

508 (ii) Mismatches in the use of aggregation procedures

509 Another recurrent problem with current practices is that common aggregation procedures (i.e.  
510 compensatory methods) are not consistent with the way value judgments are made or with  
511 sustainability issues characterized by incommensurabilities. The first problem that was repeatedly  
512 observed lies in the meaning of weights, which relates to the type of question stakeholders are asked  
513 (Choo et al., 1999). In most studies that integrate stakeholder preferences, stakeholders are asked to  
514 allocate points between criteria and then to score or rank them. Compensability is generally not  
515 addressed as such because scientists do not ask “By how many times is this more important?” or “How  
516 many units of one criteria would compensate for losing one unit of another?”, etc. (Choo et al., 1999;  
517 Garmendia and Gamboa, 2012). Nonetheless, compensatory methods are usually used to aggregate  
518 preferences. Participants could learn about the conceptual bases that distinguish, for instance,  
519 compensatory aggregation from outranking methods through the questions they answer. Therefore,  
520 defining a “good” question to elicit preferences could improve the general challenge of making the  
521 assessment process more transparent for and controllable by those involved in decision-making (Bell  
522 et al., 2001).

523 A second inconsistency lies in the idea that compensatory aggregation methods suit sustainability  
524 assessments. If we consider the question of sustainability under a “non-substitutability” hypothesis  
525 (strong sustainability, *sensu* Neumayer, 2003), then values of one criteria cannot compensate values of  
526 another criteria, which excludes compensatory aggregation. Garmendia and Gamboa (2012) and  
527 Munda (2004) suggest using outranking methods to avoid this bias. Ahrenz and Kantelhardt (2009)  
528 tested one outranking method (Electre-II) and concluded that the transparency of the method is  
529 limited, so that the choice between compensatory and non-compensatory methods for landscape  
530 planning decisions should account for the context of the assessment. Likewise, the ordered weighted  
531 averaging aggregation method allows partial compensability among criteria. In a GIS environment, it  
532 has been used mainly for scenario-design support (e.g. prioritizing areas for forest conservation,

533 Averna Valente and Vettorazzi (2008) adjusting forest-management plans, Greene et al. (2010).  
534 Nonetheless, its mathematical and computational sophistication (Aliyu and Ludin, 2015) seriously  
535 limits its use in participatory settings. This trade-off between transparency and integration of  
536 complexity should therefore be clearly addressed and discussed by scientists to justify their  
537 aggregation choices.

538 (iii) Lack of consideration of the potential of GIS to enhance deliberation

539 In sustainability assessments, MCA methods often aim to produce group or societal solutions  
540 (Borouhaki and Malczewski, 2010; Garmendia and Gamboa, 2012). In the studies selected, the  
541 “group” solutions achieved are intermediate positions between the preferences of the various  
542 stakeholders (e.g. a “compromise”, *sensu* Vodoz, 1994). Preferences are considered immutable  
543 throughout the evaluation process, which can, at best, enhance the mutual understanding of  
544 participants. Only one of our 10 studies effectively links spatial evaluation to deliberation (Arciniegas  
545 and Janssen, 2012). In this study, maps not only provide a spatialized representation of performances,  
546 but are also intermediate objects for exchanging views (for analysis and negotiation). In the other  
547 studies, social learning remains understood as a “by-product” and not as a core process fundamentally  
548 attached to the evaluation activity.

549 We support the rehabilitation of deliberation in spatially-explicit evaluations, understood as an  
550 opportunity to dialogue, understand the position of others and eventually produce novel solutions  
551 underlying a change of view. Such a deliberative perspective reconnects spatial assessments with the  
552 philosophy of collaborative GIS (Borouhaki and Malczewski, 2010). Conditions to foster collective  
553 deliberation encompass: establishing inclusive platforms for exchanging views (considering the  
554 dominated actors and the silent mass of future generations), tracing the history of changes in the  
555 evaluation matrix, engaging reiterations, and analyzing the reasons that led to a group solution or  
556 impeded its emergence. This methodological posture suggests that the coupling of GIS tools with  
557 MCA should turn toward the former’s potential to stimulate exploration, understanding and  
558 redefinition of the decision problem by those who are involved in the evaluation (Malczewski, 2006;  
559 Ramsey, 2009).

560 (iv) Forming stakeholder groups based on an *a priori* affiliation

561 Delineation of the stakeholder groups in the evaluation is crucial to allow for comparison of  
562 contrasting views. This task differs from the identification of knowledge-holders (to define indicators)  
563 and stakeholders (to choose diverse criteria and elicit judgments about them). The delineation of  
564 groups for comparing alternatives comes later, as a way to organize the diverse opinions previously  
565 explored and to identify possible alliances and conflicts. The common practice for forming groups that  
566 we observed (when groups are considered) is reliance on social and/or professional affiliation (e.g.,  
567 farmers, environmentalists, water managers). However, it supposes that social and/or professional  
568 affiliation serves as a good proxy for value systems, i.e. value systems are considered homogenous  
569 within each social or professional group.

570 Leaning on other research works, we suggest alternative guidelines to form groups of participants and  
571 reflect the diversity of competing value systems. The first option is strategic: it consists in analyzing  
572 stakeholders' interests and relationships to identify "key" participants, e.g. participants who have a  
573 legitimate influence on the problem (e.g. De Marchi et al., 2000). In a landscape context, this option  
574 also supposes to make clear what scale of interest drives one's value judgments (Zia et al., 2011). A  
575 second option is to construct preferences archetypes as done in Ravier et al. (2015) from the statistical  
576 clustering of individual preference trees. The various preferences archetypes reflect the diversity of  
577 value systems without *a priori* assuming that these archetypes would coincide with socially instituted  
578 categorization. We believe this method is useful in conflict-prone situations, which are often  
579 encountered at the landscape level. A third option is to let participants define the group to which they  
580 belong, as done by Ahrenz and Kantelhardt (2009). This way, part of the structuring bias is transferred  
581 from researchers to the participants, which is more consistent with bottom-up approaches to multi-  
582 actor problems.

583 (v) Difficulty integrating multiple levels in assessments

584 Given the complex nature of landscapes, one must account for multiple levels in the assessment. At the  
585 landscape scale, two types of aggregation occur: vertical (among criteria, already well addressed by  
586 the MCA literature) and horizontal (across space, a core component of spatial analysis). Landscape-

587 level MCAs mainly address the latter issue via spatially aggregative methods, such as weighted linear  
588 combination, that are fully compensatory (i.e. increased performance at one location compensates for  
589 decreased performance at another location). However, the overall performance of a landscape  
590 management alternative cannot be reduced to the sum of performances of its spatial units at a given  
591 moment. Such additive aggregation to combine attribute maps ignores the existence of spatial  
592 relationships and patterns (Malczewski, 2000).

593 Two directions appear for reducing this gap. The first “corrects” the aggregation result by adding other  
594 evaluation techniques. This option is illustrated by Fürst et al. (2013), who use landscape metrics to  
595 improve assessment of esthetic and biological integrity services. The second acknowledges that “gaps”  
596 between the multiple levels of assessment are irreducible; as a consequence, levels of assessment  
597 should be explicitly distinguished. Though formal methods exist to identify gaps between  
598 organizational levels (e.g. MuSIASEM, Giampietro et al., 2009), relying on management levels or  
599 stakeholders’ levels of interests can be more operational (see point iv). Using different criteria to  
600 assess different levels of the system could however undermine the consistency and intelligibility of the  
601 evaluation system. López-Ridaura et al. (2005) define disciplinary- and scale-independent criteria for  
602 sustainability assessments: productivity, stability, reliability, resilience and adaptability. This idea gave  
603 birth to the MESMIS framework (Astier et al., 2012), in which farm, community and regional  
604 sustainability are evaluated with the same set of criteria but with different indicators. Similarly, Manoli  
605 et al. (2005) use response properties (reliability, vulnerability, and resilience) to assess the temporal  
606 variability of management performances. These options offer a generic multi-level framework but are  
607 at odds with a bottom-up definition of criteria.

608 (vi) Uncertainties are under-considered

609 Finally, the question of uncertainties remains under-considered, or at best is reduced to a sensitivity  
610 analysis of the weighting systems and flexible support tools. We assert that sensitivity to scenarios of  
611 future change (as in Linhoss et al., 2013; Manoli et al., 2005), to resolution and spatial aggregation  
612 units (Malczewski, 2000), and to group delineation, among others, can enrich the deliberative process.  
613 Moreover, since judgments can change during the deliberative process, solutions to capture  
614 uncertainties may also lie in the participatory settings (Garmendia and Stagl, 2010) and the research



615 posture (e.g. “co-design”, Barreteau et al., 2010). Specific mathematical procedures (e.g., fuzzy sets,  
616 stochastic methods) can help address the vague or fuzzy nature of human judgments and information  
617 (Ascough et al., 2008; Geneletti et al., 2003).

618

619 **CONCLUSIONS**

620 We critically examined the place of sustainability assessments within multi-criteria methods, trends in  
621 methods that combine a participatory and spatial approach, and current practices for assessing  
622 landscape-management alternatives through a systematic review of the literature, combining  
623 quantitative textual statistics from large datasets and qualitative analysis of specific research.

624 Drawing upon the results of textual statistics, we suggest that conceptual assumptions constitute the  
625 primary discriminating element of multi-criteria methods and that methodological elements should be  
626 both adapted to those conceptual bases (i.e. the nature of the problem tackled), as well as operational  
627 questions (i.e. the decision context). We offer insight into multi-criteria methods that differ from the  
628 classic multi-attribute/multi-objective typology or aggregation procedures. We consider the  
629 dichotomies of substantive/procedural rationality and grounded/positivist scientific legitimacy to  
630 explain the diversity of methods and approaches. In this respect, though both are grounded in real-life  
631 problems, spatial sustainability assessments and participatory sustainability assessments differ in what  
632 they consider a “rational” decision-process. Spatial assessments focus on accumulating as much  
633 information as possible to provide manager(s) with “all the elements” to make a substantively rational  
634 decision. In contrast, participatory approaches recognize deliberation as the principal justification of  
635 results. Consequently, cross-approaches are not so common.

636 When attempting to classify participatory spatial MCA methods, we noted that to assess a real  
637 situation, support design of a management alternative or accompany selection of an alternative, a wide  
638 variety of research postures towards society can be adopted. Unlike treatment of the participatory  
639 aspect, treatment of the spatial aspect depends to some extent on the MCA goal. For landscape-  
640 management issues, the method chosen should capture the diverse ways people relate to spaces and  
641 locations. Scenario-selection methods have the greatest potential to accommodate spatially-explicit,  
642 spatially-implicit, place-specific, overall visual and non-spatial values. However, putting this into  
643 practice in a consistent and feasible way remains in an embryonic stage in the literature. The choice of  
644 participants, criteria, indicators and competing alternatives can also support the integration of diverse  
645 spatial values and multiple interest levels. Several initiatives for problem structuring could be

646 formalized to better address the spatiality of problems. As a result, current cases of landscape-level  
647 MCAs resemble “patchworks” of different methods with distinct underlying hypotheses.  
648 The main unaddressed issue of landscape-level MCAs is strikingly the question of distribution and  
649 heterogeneity. How are benefits distributed in space, time, and between social actors? How do spatial  
650 patterns and temporal behavior influence the overall “performance” of a management alternative?  
651 These questions remain under-considered. It is implicitly assumed that increased overall value is  
652 desirable, irrespective of social or spatial justice issues. Similarly, most studies assume that the  
653 performance and acceptability of a management alternative only depends on the total extent of each  
654 land cover or land-use type, irrespective of its shape, location, or configuration. Assessments are  
655 grounded in a weak conception of sustainability and a reductionist view of the landscape. While  
656 MCAs provide formal methods to overcome problems of vertical aggregation (how to aggregate  
657 results for different criteria), horizontal aggregation (how to aggregate results from different locations)  
658 is rarely considered. In addition, although (or maybe because) more complex aggregation algorithms  
659 exist, they are not accessible or transparent to the wider public. Other important aspects of MCA for  
660 landscape-management issues remain under-investigated: how stakeholders are involved, the group  
661 values produced, the maps and other spatial representations valued, and the uncertainties addressed.  
662 To conclude, we emphasize that MCA at the landscape-level do not succeed in addressing the joint  
663 issues of spatiality and multiple views, and thus has important scientific loopholes to close. MCA is  
664 nonetheless a promising method to structure landscape problems, to explore management options and  
665 to foster social learning. Even with these objectives, it is necessary to reveal biases and ethical  
666 positions implicitly assumed in the assessment process.

667

## 668 **ACKNOWLEDGEMENTS**

669 This review is part of a PhD research funded by the French Ministry of Higher Education and  
670 Research. We warmly thank two anonymous reviewers whose comments helped improve and clarify  
671 this manuscript.

672 References

- 673 Ahrenz, H., Kantelhardt, J., 2009. Accounting for farmers' production responses to  
 674 environmental restrictions within landscape planning. *Land Use Policy* 26, 925–934.  
 675 doi:10.1016/j.landusepol.2008.11.003
- 676 Alessa, L., Kliskey, A., Brown, G., 2008. Social–ecological hotspots mapping: A spatial  
 677 approach for identifying coupled social–ecological space. *Landsc. Urban Plan.* 85, 27–39.  
 678 doi:10.1016/j.landurbplan.2007.09.007
- 679 Aliyu, M., Ludin, A.N.B.M., 2015. A Review of Spatial Multi Criteria Analysis (SMCA)  
 680 Methods for Sustainable Land Use Planning (SLUP). *planning* 2.
- 681 Allan, C., Curtis, A., 2003. Learning to implement adaptive management. *Nat. Resour.*  
 682 *Manag.* 6, 25–30.
- 683 Allen, C.R., Fontaine, J.J., Pope, K.L., Garmestani, A.S., 2011. Adaptive management for a  
 684 turbulent future. *J. Environ. Manage.*, 92, 1339–1345. doi:10.1016/j.jenvman.2010.11.019
- 685 Anderies, J.M., Janssen, M.A., Ostrom, E., 2004. A framework to analyze the robustness of  
 686 social-ecological systems from an institutional perspective. *Ecol. Soc.* 9, 18.
- 687 Antrop, M., 2005. Why landscapes of the past are important for the future. *Landsc. Urban*  
 688 *Plan.*, *Rural Landscapes: past processes and future strategies* 70, 21–34.  
 689 doi:10.1016/j.landurbplan.2003.10.002
- 690 Antrop, M., 2000. Background concepts for integrated landscape analysis. *Agric. Ecosyst.*  
 691 *Environ.* 77, 17–28. doi:10.1016/S0167-8809(99)00089-4
- 692 Arciniegas, G., Janssen, R., 2012. Spatial decision support for collaborative land use planning  
 693 workshops. *Landsc. Urban Plan.* 107, 332–342. doi:10.1016/j.landurbplan.2012.06.004
- 694 Arciniegas, G., Janssen, R., Omtzigt, N., 2011. Map-based multicriteria analysis to support  
 695 interactive land use allocation. *Int. J. Geogr. Inf. Sci.* 25, 1931–1947.  
 696 doi:10.1080/13658816.2011.556118
- 697 Arciniegas, G.A., Janssen, R., 2009. Using a touch table to support participatory land use  
 698 planning. 18th World Imacs Congr. Modsim09 Int. Congr. Model. Simul. Interfacing Model.  
 699 *Simul. Math. Comput. Sci.* 2206–2212.
- 700 Ascough, J.C., Maier, H.R., Ravalico, J.K., Strudley, M.W., 2008. Future research challenges  
 701 for incorporation of uncertainty in environmental and ecological decision-making. *Ecol.*  
 702 *Model.* 219, 383–399. doi:10.1016/j.ecolmodel.2008.07.015
- 703 Astier, M., Garcia-Barrios, L., Galvan-Miyoshi, Y., Gonzalez-Esquivel, C.E., Masera, O.R.,  
 704 2012. Assessing the Sustainability of Small Farmer Natural Resource Management Systems.  
 705 A Critical Analysis of the MESMIS Program (1995-2010). *Ecol. Soc.* 17, 25. doi:10.5751/ES-  
 706 04910-170325
- 707 Avena Valente, R. de O., Vettorazzi, C.A., 2008. Definition of priority areas for forest  
 708 conservation through the ordered weighted averaging method. *For. Ecol. Manag.* 256, 1408–  
 709 1417. doi:10.1016/j.foreco.2008.07.006
- 710 Barreteau, O., Bots, P., Daniell, K., 2010. A Framework for Clarifying Participation in  
 711 Participatory Research to Prevent its Rejection for the Wrong Reasons. *Ecol. Soc.* 15, 22 p.
- 712 Bastian, O., 2001. Landscape Ecology – towards a unified discipline? *Landsc. Ecol.* 16, 757–  
 713 766. doi:10.1023/A:1014412915534
- 714 Bejean, S., Midy, F., Peyron, C., 1999. La rationalité simonienne: Interprétations et enjeux  
 715 épistémologiques (LATEC - Document de travail - Economie (1991-2003) No. 1999–14).

- 716 Bell, M.L., Hobbs, B.F., Elliott, E.M., Ellis, H., Robinson, Z., 2001. An evaluation of multi-  
717 criteria methods in integrated assessment of climate policy. *J. Multi-Criteria Decis. Anal.* 10,  
718 229–256. doi:10.1002/mcda.305
- 719 Benoît, M., Rizzo, D., Marraccini, E., Moonen, A.C., Galli, M., Lardon, S., Rapey, H.,  
720 Thenail, C., Bonari, E., 2012. Landscape agronomy: a new field for addressing agricultural  
721 landscape dynamics. *Landsc. Ecol.* 27, 1385–1394. doi:10.1007/s10980-012-9802-8
- 722 Berkes, F., Colding, J., Folke, C., 2002. *Navigating Social-Ecological Systems: Building  
723 Resilience for Complexity and Change.* Cambridge University Press.
- 724 Berkes, F., Folke, C., Colding, J., 2000. *Linking Social and Ecological Systems: Management  
725 Practices and Social Mechanisms for Building Resilience.* Cambridge University Press.
- 726 Bhave, A.G., Mishra, A., Raghuvanshi, N.S., 2014. A combined bottom-up and top-down  
727 approach for assessment of climate change adaptation options. *J. Hydrol.* 518, 150–161.  
728 doi:10.1016/j.jhydrol.2013.08.039
- 729 Boroushaki, S., Malczewski, J., 2010. Measuring consensus for collaborative decision-  
730 making: A GIS-based approach. *Comput. Environ. Urban Syst.* 34, 322–332.  
731 doi:10.1016/j.compenvurbsys.2010.02.006
- 732 Chakroun, H., Chabaane, Z.L., Benabdallah, S., 2015. Concept and prototype of a spatial  
733 decision support system for integrated water management applied to Ichkeul Basin, Tunisia.  
734 *Water Environ. J.* 29, 169–179. doi:10.1111/wej.12095
- 735 Choo, E.U., Schoner, B., Wedley, W.C., 1999. Interpretation of criteria weights in multicriteria  
736 decision making. *Comput. Ind. Eng.* 37, 527–541.
- 737 Costanza, R., d’Arge, R., De Groot, R., Farber, S., Grasso, M., Hannon, B., Limburg, K.,  
738 Naeem, S., O’Neill, R.V., Paruelo, J., others, 1998. The value of ecosystem services: putting  
739 the issues in perspective. *Ecol. Econ.* 25, 67–72.
- 740 Cumming, G.S., 2011. Spatial resilience: integrating landscape ecology, resilience, and  
741 sustainability. *Landsc. Ecol.* 26, 899–909. doi:10.1007/s10980-011-9623-1
- 742 Cumming, G.S., Olsson, P., Chapin, F.S., Holling, C.S., 2012. Resilience, experimentation,  
743 and scale mismatches in social-ecological landscapes. *Landsc. Ecol.* 28, 1139–1150.  
744 doi:10.1007/s10980-012-9725-4
- 745 Daily, G., 1997. *Nature’s services: societal dependence on natural ecosystems.* Island Press.
- 746 De Marchi, B., Funtowicz, S.O., Lo Cascio, S., Munda, G., 2000. Combining participative  
747 and institutional approaches with multicriteria evaluation. An empirical study for water issues  
748 in Troina, Sicily. *Ecol. Econ.* 34, 267–282. doi:10.1016/S0921-8009(00)00162-2
- 749 Densham, P.J., 1991. Spatial decision support systems. *Geogr. Inf. Syst. Princ. Appl.* 1, 403–  
750 412.
- 751 Folke, C., Carpenter, S., Elmqvist, T., Gunderson, L., Holling, C.S., Walker, B., 2002.  
752 *Resilience and Sustainable Development: Building Adaptive Capacity in a World of  
753 Transformations.* *AMBIO J. Hum. Environ.* 31, 437–440. doi:10.1579/0044-7447-31.5.437
- 754 Frame, B., Brown, J., 2008. Developing post-normal technologies for sustainability. *Ecol.  
755 Econ.* 65, 225–241. doi:10.1016/j.ecolecon.2007.11.010
- 756 Frame, B., O’Connor, M., 2011. Integrating valuation and deliberation: the purposes of  
757 sustainability assessment. *Environ. Sci. Policy* 14, 1–10. doi:10.1016/j.envsci.2010.10.009
- 758 Funtowicz, S.O., Ravetz, J.R., 1994. The worth of a songbird: ecological economics as a post-

- 759 normal science. *Ecol. Econ.* 10, 197–207. doi:10.1016/0921-8009(94)90108-2
- 760 Funtowicz, S.O., Ravetz, J.R., 1990. *Uncertainty and Quality in Science for Policy*. Springer  
761 Science & Business Media.
- 762 Fürst, C., Frank, S., Witt, A., Koschke, L., Makeschin, F., 2013. Assessment of the effects of  
763 forest land use strategies on the provision of ecosystem services at regional scale. *J. Environ.*  
764 *Manage., Integrated land-use and regional resource management – A cross-disciplinary*  
765 *dialogue on future perspectives for a sustainable development of regional resources* 127,  
766 Supplement, S96–S116. doi:10.1016/j.jenvman.2012.09.020
- 767 Fürst, C., Nepveu, G., Pietzsch, K., Makeschin, F., 2009. Pimp your Landscape: a software  
768 application for interactive landscape management - potential and limitations. *Rev. For.*  
769 *Francaise* 61, 21–36.
- 770 Garmendia, E., Gamboa, G., 2012. Weighting social preferences in participatory multi-criteria  
771 evaluations: A case study on sustainable natural resource management. *Ecol. Econ., The*  
772 *Economics of Degrowth* 84, 110–120. doi:10.1016/j.ecolecon.2012.09.004
- 773 Garmendia, E., Stagl, S., 2010. Public participation for sustainability and social learning:  
774 Concepts and lessons from three case studies in Europe. *Ecol. Econ.* 69, 1712–1722.  
775 doi:10.1016/j.ecolecon.2010.03.027
- 776 Gasparatos, A., El-Haram, M., Homer, M., 2008. Critical review of reductionist approaches  
777 for assessing the progress towards sustainability. *Environ. Impact Assess. Rev.* 28, 286–311.  
778 doi:10.1016/j.eiar.2007.09.002
- 779 Geneletti, D., Beinat, E., Chung, C.F., Fabbri, A.G., Scholten, H.J., 2003. Accounting for  
780 uncertainty factors in biodiversity impact assessment: lessons from a case study. *Environ.*  
781 *Impact Assess. Rev.* 23, 471–487.
- 782 Giampietro, M., 2003. *Multi-Scale Integrated Analysis of Agroecosystems*. CRC Press.
- 783 Giampietro, M., Mayumi, K., Ramos-Martin, J., 2009. Multi-scale integrated analysis of  
784 societal and ecosystem metabolism (MuSIASEM): Theoretical concepts and basic rationale.  
785 Energy, WESC 2006 Advances in Energy Studies 6th World Energy System Conference 5th  
786 workshop on Advances, Innovation and Visions in Energy and Energy-related Environmental  
787 and Socio-Economic Issues 34, 313–322. doi:10.1016/j.energy.2008.07.020
- 788 Gibson, C.C., Ostrom, E., Ahn, T.K., 2000. The concept of scale and the human dimensions of  
789 global change: a survey. *Ecol. Econ.* 32, 217–239. doi:10.1016/S0921-8009(99)00092-0
- 790 Gómez-Sal, A., Belmontes, J.-A., Nicolau, J.-M., 2003. Assessing landscape values: a  
791 proposal for a multidimensional conceptual model. *Ecol. Model., Landscape Theory and*  
792 *Landscape Modelling* 168, 319–341. doi:10.1016/S0304-3800(03)00144-3
- 793 Greene, R., Luther, J.E., Devillers, R., Eddy, B., 2010. An approach to GIS-based multiple  
794 criteria decision analysis that integrates exploration and evaluation phases: Case study in a  
795 forest-dominated landscape. *For. Ecol. Manag.* 260, 2102–2114.  
796 doi:10.1016/j.foreco.2010.08.052
- 797 Groot, J.C.J., Rossing, W.A.H., 2011. Model-aided learning for adaptive management of  
798 natural resources: an evolutionary design perspective. *Methods Ecol. Evol.* 2, 643–650.  
799 doi:10.1111/j.2041-210X.2011.00114.x
- 800 Guitouni, A., Martel, J.-M., 1998. Tentative guidelines to help choosing an appropriate  
801 MCDA method. *Eur. J. Oper. Res.* 109, 501–521. doi:10.1016/S0377-2217(98)00073-3
- 802 Hajkowicz, S., Collins, K., 2006. A Review of Multiple Criteria Analysis for Water Resource

- 803 Planning and Management. *Water Resour. Manag.* 21, 1553–1566. doi:10.1007/s11269-006-  
804 9112-5
- 805 Hein, L., van Koppen, K., de Groot, R.S., van Ierland, E.C., 2006. Spatial scales, stakeholders  
806 and the valuation of ecosystem services. *Ecol. Econ.* 57, 209–228.  
807 doi:10.1016/j.ecolecon.2005.04.005
- 808 Holling, C. s., Meffe, G.K., 1996. Command and Control and the Pathology of Natural  
809 Resource Management. *Conserv. Biol.* 10, 328–337. doi:10.1046/j.1523-  
810 1739.1996.10020328.x
- 811 Holling, C.S., 2001. Understanding the Complexity of Economic, Ecological, and Social  
812 Systems. *Ecosystems* 4, 390–405. doi:10.1007/s10021-001-0101-5
- 813 Hunziker, M., Felber, P., Gehring, K., Buchecker, M., Bauer, N., Kienast, F., 2008. Evaluation  
814 of Landscape Change by Different Social Groups. *Mt. Res. Dev.* 28, 140–147.  
815 doi:10.1659/mrd.0952
- 816 Janssen, R., Goosen, H., Verhoeven, M.L., Verhoeven, J.T.A., Omtzigt, A.Q.A., Maltby, E.,  
817 2005. Decision support for integrated wetland management. *Environ. Model. Softw.* 20, 215–  
818 229. doi:10.1016/j.envsoft.2003.12.020
- 819 Kallis, G., Norgaard, R.B., 2010. Coevolutionary ecological economics. *Ecol. Econ.*, Special  
820 Section: Coevolutionary Ecological Economics: Theory and Applications 69, 690–699.  
821 doi:10.1016/j.ecolecon.2009.09.017
- 822 Kiker, G.A., Bridges, T.S., Varghese, A., Seager, P.T.P., Linkov, I., 2005. Application of  
823 multicriteria decision analysis in environmental decision making. *Integr. Environ. Assess.*  
824 *Manag.* 1, 95–108. doi:10.1897/IEAM\_2004a-015.1
- 825 Kinzig, A.P., Ryan, P.A., Etienne, M., Allison, H.E., Elmqvist, T., Walker, B.H., 2006.  
826 Resilience and regime shifts: assessing cascading effects. *Ecol. Soc.* 11.
- 827 Le Bellec, F., Rajaud, A., Ozier-Lafontaine, H., Bockstaller, C., Malezieux, E., 2012.  
828 Evidence for farmers' active involvement in co-designing citrus cropping systems using an  
829 improved participatory method. *Agron. Sustain. Dev.* 32, 703–714. doi:10.1007/s13593-011-  
830 0070-9
- 831 Lebart, L., Salem, A., Berry, L., 1998. Exploring Textual Data, Text, Speech and Language  
832 Technology. Springer Netherlands, Dordrecht.
- 833 Linhoss, A.C., Kiker, G.A., Aiello-Lammens, M.E., Chu-Agor, M.L., Convertino, M., Muñoz-  
834 Carpena, R., Fischer, R., Linkov, I., 2013. Decision analysis for species preservation under  
835 sea-level rise. *Ecol. Model.* 263, 264–272. doi:10.1016/j.ecolmodel.2013.05.014
- 836 Linkov, I., Bridges, T., Creutzig, F., Decker, J., Fox-Lent, C., Kröger, W., Lambert, J.H.,  
837 Levermann, A., Montreuil, B., Nathwani, J., others, 2014. Changing the resilience paradigm.  
838 *Nat. Clim. Change* 4, 407–409.
- 839 Linkov, I., Satterstrom, F.K., Kiker, G.A., Bridges, T.S., Benjamin, S.L., Belluck, D.A., 2006.  
840 From optimization to adaptation: Shifting paradigms in environmental management and their  
841 application to remedial decisions. *Integr. Environ. Assess. Manag.* 2, 92–98.  
842 doi:10.1002/ieam.5630020116
- 843 Liu, J., Dietz, T., Carpenter, S.R., Alberti, M., Folke, C., Moran, E., Pell, A.N., Deadman, P.,  
844 Kratz, T., Lubchenco, J., Ostrom, E., Ouyang, Z., Provencher, W., Redman, C.L., Schneider,  
845 S.H., Taylor, W.W., 2007. Complexity of Coupled Human and Natural Systems. *Science* 317,  
846 1513–1516. doi:10.1126/science.1144004

- 847 López-Ridaura, S., Keulen, H.V., Ittersum, M.K. van, Leffelaar, P.A., 2005. Multiscale  
 848 Methodological Framework to Derive Criteria and Indicators for Sustainability Evaluation of  
 849 Peasant Natural Resource Management Systems. *Environ. Dev. Sustain.* 7, 51–69.  
 850 doi:10.1007/s10668-003-6976-x
- 851 Malczewski, J., 2006. GIS-based multicriteria decision analysis: a survey of the literature. *Int.*  
 852 *J. Geogr. Inf. Sci.* 20, 703–726. doi:10.1080/13658810600661508
- 853 Malczewski, J., 2000. On the use of weighted linear combination method in GIS: common  
 854 and best practice approaches. *Trans. GIS* 4, 5–22.
- 855 Malczewski, J., 1999. *GIS and Multicriteria Decision Analysis*. John Wiley & Sons.
- 856 Malczewski, J., Rinner, C., 2015. *Multicriteria decision analysis in geographic information*  
 857 *science*. Springer.
- 858 Manoli, E., Katsiardi, P., Arampatzis, G., Assimacopoulos, D., 2005. Comprehensive water  
 859 management scenarios for strategic planning, in: Lekkas, T.D. (Ed.), *Proceedings of the 9th*  
 860 *International Conference on Environmental Science and Technology, Vol A - Oral*  
 861 *Presentations, Pts A and B*. pp. A913–A920.
- 862 Martinez-Alier, J., Munda, G., O’Neill, J., 1998. Weak comparability of values as a  
 863 foundation for ecological economics. *Ecol. Econ.* 26, 277–286. doi:10.1016/S0921-  
 864 8009(97)00120-1
- 865 McFadden, D., 1994. Contingent valuation and social choice. *Am. J. Agric. Econ.* 76, 689–  
 866 708.
- 867 Munda, G., 2004. Social multi-criteria evaluation: Methodological foundations and  
 868 operational consequences. *Eur. J. Oper. Res.* 158, 662–677. doi:10.1016/S0377-  
 869 2217(03)00369-2
- 870 Munda, G., Nijkamp, P., Rietveld, P., 1994. Qualitative multicriteria evaluation for  
 871 environmental management. *Ecol. Econ.* 10, 97–112.
- 872 Naveh, Z., 2000. What is holistic landscape ecology? A conceptual introduction. *Landsc.*  
 873 *Urban Plan.* 50, 7–26. doi:10.1016/S0169-2046(00)00077-3
- 874 Neumayer, E., 2003. *Weak Versus Strong Sustainability: Exploring the Limits of Two*  
 875 *Opposing Paradigms*. Edward Elgar Publishing.
- 876 Newton, A.C., Hodder, K., Cantarello, E., Perrella, L., Birch, J.C., Robins, J., Douglas, S.,  
 877 Moody, C., Cordingley, J., 2012. Cost-benefit analysis of ecological networks assessed  
 878 through spatial analysis of ecosystem services. *J. Appl. Ecol.* 49, 571–580.  
 879 doi:10.1111/j.1365-2664.2012.02140.x
- 880 Nitschelm, L., Aubin, J., Corson, M.S., Viaud, V., Walter, C., 2016. Spatial differentiation in  
 881 Life Cycle Assessment LCA applied to an agricultural territory: current practices and method  
 882 development. *J. Clean. Prod.* 112, Part 4, 2472–2484. doi:10.1016/j.jclepro.2015.09.138
- 883 Nordstrom, E.-M., Eriksson, L.O., Ohman, K., 2011. Multiple Criteria Decision Analysis with  
 884 Consideration to Place-specific Values in Participatory Forest Planning. *Silva Fenn.* 45, 253–  
 885 265. doi:10.14214/sf.116
- 886 Nordström, E.-M., Eriksson, L.O., Öhman, K., 2010. Integrating multiple criteria decision  
 887 analysis in participatory forest planning: Experience from a case study in northern Sweden.  
 888 *For. Policy Econ.* 12, 562–574. doi:10.1016/j.forpol.2010.07.006
- 889 Olsson, P., Gunderson, L.H., Carpenter, S.R., Ryan, P., Lebel, L., Folke, C., Holling, C.S.,



- 890 2006. Shooting the rapids: navigating transitions to adaptive governance of social-ecological  
891 systems. *Ecol. Soc.* 11, 18.
- 892 Ostrom, E., 2009. A General Framework for Analyzing Sustainability of Social-Ecological  
893 Systems. *Science* 325, 419–422. doi:10.1126/science.1172133
- 894 Pedroli, B., Pinto-Correia, T., Cornish, P., 2006. Landscape – What’s in it? Trends in  
895 European Landscape Science and Priority Themes for Concerted Research. *Landsc. Ecol.* 21,  
896 421–430. doi:10.1007/s10980-005-5204-5
- 897 Plummer, R., 2009. The adaptive co-management process: an initial synthesis of  
898 representative models and influential variables. *Ecol. Soc.* 14, 24.
- 899 Raaijmakers, R., Krywkow, J., van der Veen, A., 2008. Flood risk perceptions and spatial  
900 multi-criteria analysis: an exploratory research for hazard mitigation. *Nat. Hazards* 46, 307–  
901 322. doi:10.1007/s11069-007-9189-z
- 902 Ramsey, K., 2009. GIS, modeling, and politics: On the tensions of collaborative decision  
903 support. *J. Environ. Manage., Collaborative GIS for spatial decision support and visualization*  
904 90, 1972–1980. doi:10.1016/j.jenvman.2007.08.029
- 905 Ratinaud, P., Marchand, P., 2012. Application de la méthode ALCESTE à de «gros» corpus et  
906 stabilité des «mondes lexicaux»: analyse du «CableGate» avec IRaMuTeQ. *Actes 11e Journ.*  
907 *Int. Anal. Stat. Données Textuelles JADT 2012.*
- 908 Ravier, C., Prost, L., Jeuffroy, M.-H., Wezel, A., Paravano, L., Reau, R., 2015. Multi-criteria  
909 and multi-stakeholder assessment of cropping systems for a result-oriented water quality  
910 preservation action programme. *Land Use Policy* 42, 131–140.  
911 doi:10.1016/j.landusepol.2014.07.006
- 912 Reichert, P., Langhans, S.D., Lienert, J., Schuwirth, N., 2015. The conceptual foundation of  
913 environmental decision support. *J. Environ. Manage.* 154, 316–332.  
914 doi:10.1016/j.jenvman.2015.01.053
- 915 Reinert, M., 1993. Les “mondes lexicaux” et leur “logique” à travers l’analyse statistique d’un  
916 corpus de récits de cauchemars.” *Lang. Société* 66, 5–39. doi:10.3406/lsoc.1993.2632
- 917 Sahin, O., Mohamed, S., 2013. A spatial temporal decision framework for adaptation to sea  
918 level rise. *Environ. Model. Softw.* 46, 129–141. doi:10.1016/j.envsoft.2013.03.004
- 919 Sheppard, S.R.J., Meitner, M., 2005. Using multi-criteria analysis and visualisation for  
920 sustainable forest management planning with stakeholder groups. *For. Ecol. Manag., Decision*  
921 *Support in Multi Purpose Forestry Decision Support in Multi Purpose Forestry Selected papers*  
922 *from the symposium on “Development and Application of Decision Support Tools in Multiple*  
923 *Purpose Forest Management”* 207, 171–187. doi:10.1016/j.foreco.2004.10.032
- 924 Simon, H.A., 1990. Alternative visions of rationality, in: Moser, P.K. (Ed.), *Rationality in*  
925 *Action: Contemporary Approaches.* Cambridge University Press, pp. 189–204.
- 926 Simon, H.A., 1976. From substantive to procedural rationality, in: Kastelein, T.J., Kuipers,  
927 S.K., Nijenhuis, W.A., Wagenaar, G.R. (Eds.), *25 Years of Economic Theory.* Springer US,  
928 pp. 65–86.
- 929 Swedeen, P., 2006. Post-normal science in practice: A Q study of the potential for sustainable  
930 forestry in Washington State, USA. *Ecol. Econ.* 57, 190–208.  
931 doi:10.1016/j.ecolecon.2005.04.003
- 932 van Asselt, M., Rijkens-Klomp, N., 2002. A look in the mirror: reflection on participation in  
933 Integrated Assessment from a methodological perspective. *Glob. Environ. Change* 12, 167–

- 934 184. doi:10.1016/S0959-3780(02)00012-2
- 935 van Herwijnen, M., Janssen, R., 2007. The use of multi-criteria evaluation in spatial policy,  
936 in: Asian Conference on Remote Sensing. Kuala Lumpur.
- 937 Vatn, A., 2009. An institutional analysis of methods for environmental appraisal. *Ecol. Econ.*  
938 68, 2207–2215. doi:10.1016/j.ecolecon.2009.04.005
- 939 Vatn, A., 2005. *Institutions and the Environment*. Edward Elgar Publishing.
- 940 Vodoz, L., 1994. La prise de décision par consensus: pourquoi, comment, à quelles  
941 conditions. *Environ. Société* 13, 57–66.
- 942 Walker, B., Holling, C.S., Carpenter, S.R., Kinzig, A., 2004. Resilience, adaptability and  
943 transformability in social–ecological systems. *Ecol. Soc.* 9, 5.
- 944 Wilson, J., Low, B., Costanza, R., Ostrom, E., 1999. Scale misperceptions and the spatial  
945 dynamics of a social–ecological system. *Ecol. Econ.* 31, 243–257. doi:10.1016/S0921-  
946 8009(99)00082-8
- 947 Wu, J. (Jingle), 2006. *Landscape Ecology, Cross-disciplinarity, and Sustainability Science*.  
948 *Landsc. Ecol.* 21, 1–4. doi:10.1007/s10980-006-7195-2
- 949 Xenarios, S., Tziritis, I., 2007. Improving pluralism in Multi Criteria Decision Aid approach  
950 through Focus Group technique and Content Analysis. *Ecol. Econ.* 62, 692–703.  
951 doi:10.1016/j.ecolecon.2006.08.017
- 952 Zia, A., Hirsch, P., Songorwa, A., Mutekanga, D.R., O’Connor, S., McShane, T., Brosius, P.,  
953 Norton, B., 2011. Cross-Scale Value Trade-Offs in Managing Social-Ecological Systems: The  
954 Politics of Scale in Ruaha National Park, Tanzania. *Ecol. Soc.* 16. doi:10.5751/ES-04375-  
955 160407
- 956

Comment citer ce document :

Allain, S. (Auteur de correspondance), Plumecocq, G., Burger-Leenhardt, D. (2017). How do multi-criteria assessments address landscape-level problems? A review of studies and practices. *Ecological Economics*, 136, 282-295. DOI : 10.1016/j.ecolecon.2017.02.011