



HAL
open science

Impact of farming systems on agricultural landscapes and biodiversity: from plot to farm and landscape scales

Gilles Martel, Stéphanie Aviron, Alexandre Joannon, Etienne Lalechère,
Bénédicte Roche, Hugues Bousard

► To cite this version:

Gilles Martel, Stéphanie Aviron, Alexandre Joannon, Etienne Lalechère, Bénédicte Roche, et al.. Impact of farming systems on agricultural landscapes and biodiversity: from plot to farm and landscape scales. *European Journal of Agronomy*, 2019, 107, pp.53-62. 10.1016/j.eja.2017.07.014 . hal-01588617

HAL Id: hal-01588617

<https://hal.science/hal-01588617v1>

Submitted on 22 Oct 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

Copyright

Impact of farming systems on agricultural landscapes and biodiversity: from plot to farm and landscape scales.

Gilles Martel ^{±1}, Stéphanie Aviron ¹, Alexandre Joannon ¹, Etienne Lalechère ^{1*}, Bénédicte Roche ¹,

Hugues Boussard ¹

¹ INRA UMR0980 BAGAP, 65 rue de Saint Briec, 35042 RENNES Cedex, France

* Present address: Irstea – Lisc, 63178 Aubière – France

[±] Corresponding author: gilles.martel@inra.fr

1 **Impact of farming systems on agricultural landscapes and biodiversity: from plot to**
2 **farm and landscape scales.**

3 **Abstract**

4 Green-way policies in agricultural landscapes focus on ecological continuity between semi-natural
5 elements (hedgerows, permanent grasslands, woods) and landscape heterogeneity. These policies
6 suggest annual crops and temporary grasslands exhibit a negative or neutral impact on biodiversity.
7 However, recent studies indicated the spatial continuities between different crops (spring vs. winter)
8 showed positive impacts on biodiversity. These landscape patterns were directly related to farmers'
9 decisions regarding the crops cultivated and where the crops were distributed spatially on the farm.
10 The aim of the present study was to evaluate the impacts of different livestock farming system
11 management types on crop patterns and associated biodiversity (carabid beetles) in agricultural
12 landscapes. We combined empirical analyses of farmers' decision making and ecological data to
13 develop a modeling framework simulating crop allocation and abundance of two different carabid
14 beetle species groups (maize and woody species). Modeling included field, farm, and landscape
15 levels. We simulated different scenarios, where two livestock farming systems, swine and dairy,
16 were combined in different proportions (i.e. number of swine vs. dairy farms) in two agricultural
17 landscapes with varied hedgerow densities. Simulations showed maize carabid species abundance
18 was higher in swine production landscapes due to more frequent spatial continuities between spring
19 and winter crops. In contrast, woody carabid species were more abundant in mixed landscapes
20 (dairy and swine) under high crop diversity. For a given combination of livestock farming systems,
21 simulated landscapes were highly variable in crop acreages and spatial continuities between crops.
22 Our results emphasized the need to manage landscape at a collective level, where crop allocation
23 decisions create more interfaces without modifying livestock farming system combinations.

24 **Keywords:** Agricultural landscape modeling, Biodiversity, Carabid beetle, Livestock,
25 Heterogeneity, Mixed landscape

26 **1. Introduction**

27 Preventing further biodiversity losses in agricultural landscapes is an important social and economic
28 issue, particularly due to societal expectations regarding the provision of ecosystem services, such
29 as pollination and pest regulation from species at risk, e.g. *Apis mellifera*, *Rodolia cardinalis*
30 (Kleijn and Sutherland, 2003; Le Roux et al., 2008; Millennium Ecosystem Assessment, 2005).
31 Among the major causes responsible for the decline in biodiversity, habitat fragmentation
32 (corresponding to a reduction in suitable species habitat and increased habitat isolation; Fahrig,
33 2003) was inarguably identified as a main driver in species extinction (Fahrig, 2003; Krauss et al.,
34 2010; Tilman et al., 2001). The connectivity of remaining suitable habitat fragments — defined as
35 the degree to which the landscape facilitates or impedes species movements (Taylor et al., 1993) —
36 is therefore considered particularly critical for species survival (Fahrig and Merriam, 1985, 1994;
37 Moilanen and Hanski, 1998; Ricketts, 2001).

38 A variety of green-way policies have been developed and implemented in Europe and elsewhere to
39 promote habitat connectivity in agricultural landscapes through the design of "ecological networks"
40 (Bennett and Mulongoy, 2006). Existing policies primarily focus on the connectivity of semi-
41 natural habitats (hedgerows, permanent grasslands, woods) (e.g. the Natura 2000 network in the
42 EU), which was found crucial for many species survival (Bianchi et al., 2006; Billeter et al., 2008).
43 Under these policies, the "matrix" composed of annual crops and temporary grasslands was
44 expected to result in a negative or neutral impact on biodiversity. Some studies suggested the
45 presence of annual crops with dense cover had a positive impact on woody species (see e.g.
46 Fitzgibbon, 1997; Ouin et al., 2000). The connectivity of annual crops, which serve a role for
47 species using cropped habitats during their life cycle was also reported, such as pollinators or
48 natural enemies of crop pests (Burel et al., 2013; Maisonhaute, 2010; Varchola and Dunn, 1999;
49 Vasseur et al., 2013). Because biotic and abiotic resources are ephemeral in annual crops, the
50 survival of these species is expected to depend on their ability to colonise new suitable habitats to
51 supplement or complement resources (Dunning et al., 1992). The temporal connectivity of crops,

52 such as winter and spring crops, which offer resources at asynchronous periods, might guarantee
53 supplementation and complementation processes and therefore native species survival (Burel et al.,
54 2013; Vasseur et al., 2013).

55 Papy (2001) showed annual crop connectivity in agricultural landscapes was directly related to
56 farmers' decisions regarding the crops cultivated and where the crops were geographically located
57 on the farm. In livestock farms, crop choice was associated with animal feeding management (Papy,
58 2001). When several animal types were bred and include a dairy herd, the main objective was to
59 produce fodder (i.e. primarily grass and maize) for dairy animals (Garcia et al., 2005). In dairy
60 farms, the diet composition and field spatial organization (distance of fields to farmstead)
61 influenced crop rotation choices (Brunschwig et al., 2006). The main factor determining crop
62 rotation on swine farms was the presence of a feed production facility on the farm (Tersiguel et al.,
63 2012). Tersiguel et al. (2012) reported swine farms grow primarily wheat and maize when feed
64 production facilities were on the farm, whereas swine farms without feed production facilities grow
65 a large diversity of winter crops and less maize. In addition to feed production, crop market prices
66 (Reganold et al., 2011) and manure management (Ramonet et al., 2014) also played a role in crop
67 rotation choices for all farming systems. Crop location on the farm depended on field
68 characteristics, crop management requirements, and distance between field and facilities
69 (Brunschwig et al., 2006; Maxime et al., 1995). The types and diversity of livestock farming
70 systems associated with the field pattern and characteristics on farms might therefore be important
71 drivers of crop surface area and spatial configuration, i.e. crop connectivity. However, to our
72 knowledge, studies have not yet attempted to associate the agronomic drivers of crop patterns on
73 biodiversity, which might be due to the challenges in experimentation on these factors: the farming
74 system is a specific choice of a farmer and cannot be altered. Modeling the agricultural landscape
75 resulting from crop organization by farmers is an alternative. Several landscape models have been
76 developed in recent years, mainly based on statistical inferences (Dury, 2011). The models are
77 largely derived from data mining of land cover databases, where the farm level is lacking. This

78 implies the models cannot integrate an explicit representation of a cropping plan decision rule at the
79 farm level. Some models include the possibility to define several groups of fields, and for each,
80 specific parameter are available to simulate land cover (e.g. Castellazzi et al., 2010), but for
81 ecological analyses the models are not defined. However, a wide range of whole farm models have
82 been developed to identify the best cropping plan to fit farmer objectives (Chardon et al., 2012;
83 Pannell, 1996; Rotz et al., 1999; Rounsevell et al., 2003). Some of these models do not include a
84 spatial representation of the cropping plan (e.g. Schils et al., 2007) and cannot be used to evaluate
85 landscape configuration. However other models include a spatial representation of cropping plans,
86 but i) produce only one optimized crop allocation at the farm level, and ii) do not allow aggregation
87 of several farms within a landscape (e.g. Chardon et al., 2012). The role of agronomic factors
88 should be explored by developing a model that describes the diversity of possible crop distributions
89 within a farming system and not only an optimal model.

90 The objectives of this study were to evaluate the impact of different livestock farming management
91 systems on landscape patterns related to crops and associated biodiversity using carabid beetles
92 (Coleoptera, Carabidae) as the model species. Carabid beetles have been extensively studied in
93 agricultural landscapes (Koivula, 2011; Kromp, 1999; Rainio and Niemelä, 2003) and they are
94 known to respond to the effects of landscape composition or configuration related to semi-natural
95 habitats (Billeter et al., 2008; Duflot et al., 2015; Puech et al., 2014) or land-use diversity (Ekroos et
96 al., 2010; Maisonhaute et al., 2010). The following hypotheses were tested: i) the adoption of
97 different livestock farming systems generated changes in landscape patterns observed in relative
98 amounts of spring and winter crops and crop connectivity; ii) the observed changes were modulated
99 by the spatial characteristics of farm fields; and iii) changes influenced carabid beetle species using
100 cropped habitats, but not on species using semi-natural habitats. Our hypotheses were tested using a
101 modeling framework, which simulated the diversity of agricultural landscapes based on crop
102 allocation decision rules at the farm level and estimated biodiversity based on landscape

103 configurations. We chose two different livestock farming systems as a case study — swine and dairy
104 — in hedgerow network landscapes (“bocage” landscapes) of Brittany (France).

105

106 **2. Materials and Methods**

107 Our landscape modeling framework combined plot (crop field or woody element, here hedgerow),
108 and its landscape context (described in a circle centred on the plot, with varying radius sizes), the
109 landscape unit level (defined here as a 500m radius circle), the individual farm level, and the level
110 of several farms (eight farms in total) (Figs. 1 and 2). The model combined two ecological models
111 predicting carabid abundance at the plot level according to plot landscape context (A), a crop
112 allocation model at farm level integrating agronomic decisions and constraints (B), and digitized
113 maps of real landscapes including several farm territories (C) to predict and assess landscape
114 patterns and ecological effects produced under different scenarios of farming system combinations
115 (D and E). The combination of ecological models with farming scenario and landscape analysis
116 simulations were conducted in APILand virtual laboratory for landscape modeling (Boussard et al.,
117 2010). This modeling platform includes the following: i) a meta-model of landscape representation
118 in space, time, and theme, which facilitates a non-agricultural matrix (roads, buildings, forests, and
119 hedgerows) and farm territory combination; and ii) temporal simulation tools to combine agronomic
120 and ecological models.

121

122 **2.1. Ecological and agronomic models**

123 **2.1.1. Predictive models of carabid beetle abundances: the 'maize' & 'woody' models (A)**

124 Statistical modeling was performed to simulate carabid beetle abundance at the plot level (crop field
125 or woody elements) based on the characteristics of plot landscape contexts (amount, spatial
126 connectivity, and cultivated and uncultivated habitat heterogeneity) (see Appendix A for more
127 details). Multi-model inference (Burnham and Anderson, 2002) applying Mixed Generalized Linear
128 Models (GLMM) was used to link existing carabid data and landscape metrics (calculated in

129 different radius sizes) in landscape case studies in Brittany, Western France (Duflot et al., 2014,
130 2015, 2016). Analyses were performed for two different indicator carabid species groups, i.e.
131 species associated with woody habitats ('woody' model) and species associated with maize crops
132 ('maize' model). The final statistical models identified for the two carabid groups (best models with
133 the lowest Akaike's information criterion (AICc)) were the following:

134 • Maize species: $N_{maize} = e^{(4.98 + (6.78e^{-04} \times E_{WIC-MA500}) - (7.05e^{-06} \times C_{W500}))}$

135 where N_{maize} is the abundance of maize species, $E_{WIC-MA500}$ is the edge length between maize and
136 winter crops in a 500 m radius and C_{W500} is the connectivity of woody habitats in a 500 m radius.

137 • Woody species: $N_{woody} = e^{(5.43e^{-01} + (3.49e^{-04} \times C_{W50}) + (6.80e^{-05} \times C_{G50}) + (1.37 \times SHDI_{50}))}$

138 where N_{woody} is the abundance of woody species, and C_{W50} , C_{G50} , and $SHDI_{50}$ represent respectively
139 woody habitats and grasslands connectivity and land-use Shannon diversity in a 50 m radius.

140 The two carabid models explained a good percentage of variation in carabid abundances (pseudo-R²
141 = 0.48 for both models) (Appendix A). The predictive capacity of the two statistical models was
142 assessed by i) calculating a measure of Root Mean Square Error of Prediction (RMSEP) with k-fold
143 cross-validation (where k = 10); and ii) fitting observed and predicted values for both models. The
144 'maize' model had a RMSEP of 80.3 individuals, which corresponded to 33% of the median
145 observed abundances of maize species (min = 9, max = 494). The 'woody' model had a RMSEP of
146 20.5 individuals, which corresponded to 49% of the median observed abundances of woody species
147 (min = 0, max = 83). The comparison of observed and predicted abundances showed that both
148 models tended to underestimate predicted abundances compared to observations (Figure 1,
149 Appendix A). It also showed for both models a correct, overall fit between predicted and observed
150 values ('maize' model: pearson correlation R = 0.64; 'woody' model: R = 0.58), but weaker
151 relationships for the highest abundance values (above median values) (Figure 1, Appendix A).

152 The two models were integrated in the APILand virtual laboratory as the 'maize' and 'woody'
153 models.

154

155 **2.1.2. Model of crop allocation: the CAPFarm model (B)**

156 A crop allocation problem solver was developed to simulate crop allocations over years at the farm
157 level. The crop allocation problem was defined by Akplogan et al. (2013) as "a spatiotemporal
158 planning problem in which crops are assigned to fields over a finite time horizon" and we solved the
159 problem by assigning "crops (values) to fields (variables) over a finite time horizon" (Akplogan et
160 al., 2013).

161 We developed our crop allocation problem solver — 'CAPFarm' — applying the following steps: i)
162 the JAVA *Choco3* Constraint Satisfaction Problem (CSP) solver (Prud'homme et al., 2014) was
163 employed to build a crop allocation problem solver in the APILand platform; ii) field data from
164 farm surveys in Brittany (Western France) were applied to conduct a first parameterization of the
165 model integrating crop distribution constraints observed in livestock farming systems in the area
166 investigated (constraint list in Table 1) (Roche et al., 2013); two farms with different crop rotations
167 and location rules were considered at this stage, i.e. a swine farming system ('*swine*') and a dairy
168 farming system ('*dairy*'); and iii) crop allocation simulations were run for the two farms to generate
169 maps of possible crop allocation during five consecutive years on surveyed farms; the maps were
170 verified by farmers to modify and/or validate crop rotation and location rules.

171 The final model parameterization integrated the following rules: i) the '*swine*' system was
172 characterized by large geographic surface areas cultivating cereals (including rape seed), no
173 temporary grassland, short crop rotations, and only 50% usable agricultural area (UAA) required for
174 the minimum cropped area; and (ii) the '*dairy*' system was characterized by distance from pasture
175 fields to milking area (farmstead) constraints, large geographic surface areas cultivating temporary
176 grassland and maize for forage self-sufficiency, long crop rotations including cereal crops, and 90%
177 of the UAA required for minimum cropped area. The final set of rules for each case system is
178 shown in Table 1.

179

180 **2.2. Selection of case study sites: the 'open' & 'dense' landscape units (C)**

181 We tested the sensitivity of our simulation results to “bocage” characteristics by selecting two
182 landscape units (and their associated farms) with different densities of woody elements. The
183 landscape units were defined within 500 m radii circles to run both ecological model simulations
184 (see part 2.1). We used two existing GIS databases to produce maps with realistic landscape
185 composition and configuration regarding arable fields, roads, buildings, and woody elements: i) a
186 regional map of woody elements (hedgerows and woodlots) produced by the CNRS-LETG
187 laboratory (Hubert-Moy et al., 2012); ii) the land parcel identification system (LPIS), where farm
188 territories and arable islets are mapped. Moving window analyses were applied to these maps using
189 the *Chloe2012* software (Boussard and Baudry, 2014) to select two landscape units with varied
190 densities of woody elements, but with a similar number of farms (eight) and UAA percentage (90 %
191 of the circle areas). Finally, we digitized fields within LPIS islets and identified a farmstead for
192 each farm added from aerial photographs and satellite images. This resulted in production of two
193 divergent virtual landscape units based on real landscape composition and configuration: an 'open'
194 landscape unit with very low density of woody elements (0.5 % of the 500 m radius circle area) and
195 a 'dense' landscape unit, with higher density of woody elements (16.3 % of the 500 m radius circle
196 area). The farms composing those landscape units are noted O1 to O8 for open landscape unit and
197 F1 to F8 for dense landscape unit.

198

199 **2.3. Virtual experiment plan for landscape and ecological simulations (D)**

200 The effects of crop allocation on carabid abundance was defined by simulating the following virtual
201 experiment plan applied to the two case study landscape units (*open* and *dense*):

- 202 • At the farm level and for each of the eight farms, which composed each landscape unit, we pre-
203 simulated crop allocation for both farming systems, i.e. 'swine' and 'dairy' using the *CAPFarm*
204 solver parameterized with validated sets of agronomic constraints (B). One thousand simulation
205 iterations were run to obtain several crop allocation over 10 years for each farm and each
206 farming system. Due to variable constraint levels resulting of the combination farm / farming

207 system, we obtained from 9 to 810 successes over the 1000 iterations. These successful
208 iterations allowed the generation of various possible crop distribution patterns regarding each
209 set of agronomic constraints.

210 • At the all farms level, we simulated 256 scenarios representing all combination possibilities of
211 two farming systems across eight farms (from zero to eight farms in swine vs. dairy farming).
212 For each scenario, 250 simulation iterations were performed to ensure sufficient robustness in
213 the modeling results and allow sufficient comparisons among scenarios. Simulations were
214 conducted by randomly selecting one of the pre-simulated crop allocations for each farm over
215 10 years. This resulted in the production of 10 x 250 x 256 simulated landscape maps for each
216 landscape unit. These sets of simulated landscape maps are henceforth referred to as 'open
217 landscape' or 'dense landscape'.

218 • At the plot level, the '*maize*' and '*woody*' models were run using landscape metrics calculated on
219 simulated landscape units to predict carabid abundances at the plot level based on its landscape
220 context. For the '*maize*' model, landscape metrics were calculated within a 500 m radius around
221 the central arable field of each landscape unit. Calculations were conducted only for years
222 where these central fields were cultivated as maize according to crop allocation model
223 simulations. The '*woody*' model had a limited spatial extent (50 m) used in the metric
224 calculations. Therefore, running the simulations for a central woody element (hedge) at each
225 landscape unit could not ensure sufficient variability for plots' spatial context among scenarios.
226 Subsequently, the '*woody*' model was run on each year for several woody elements based on the
227 density of woody elements at each landscape unit (five at the '*open*' landscape unit and 29 at the
228 '*dense*' landscape unit).

229 **2.4. Data analyses (E)**

230 As agronomic rules contains some time dependencies (i.e.: crop return time), not all the rules are
231 effective at the start of the simulation. The first five years (over the ten simulated) were therefore as
232 an initialization and stabilization phase and were excluded from analyses. On the remaining five

233 years, we considered each year of a simulated landscape unit as a statistical individual. This resulted
234 in 1250 statistical individuals (i.e. simulated landscapes-year) for each scenario.

235 We analysed the variation in cultivated area amongst scenarios, by calculating the surface area of
236 crops (averaged over the 1250 simulated landscapes-year) at the all farms and landscape unit levels.

237 We distinguished three categories of crops: (i) winter crops (WIC) grouping wheat, oil seed rape
238 and barley, (ii) spring crops (SC) grouping maize-corn and maize-silage and (iii) grassland (GL)
239 grouping permanent and temporary grassland, pasture and lucerne.

240 The impacts of farming systems scenarios on carabid abundance predicted by 'maize' and 'woody'
241 models at the plot level were analyzed by testing the effects of the following two variables: i)
242 number of farms under dairy production (0 to 8 farms); and ii) percentage of geographic surface
243 area in each landscape unit managed under dairy production (see part 2.3.).

244 In a last step, we analyzed farm contributions to variation in carabid beetle abundance predicted by
245 the 'maize' model using a specific analysis. We focused on the 'maize' model, as only one
246 calculation was required per simulated landscape, in contrast to the 'woody' model (see part 2.3.).

247 Focusing on the 'maize' model allowed us to limit the potential factor number able to modify the
248 farm contributions to variation in carabid abundance predicted by ecological models. The
249 contribution of an individual farm i to the variation in predicted abundances of maize carabids was
250 assessed in several steps. First, the 256 initial scenarios were split into 128 sets of two modalities,
251 where for each set, the individual farm i was either under swine farming ('S') or dairy farming ('D') ,
252 whilst the other seven farms had fixed farming type. The 128 sets corresponded to the whole
253 possibilities regarding the type of farming system attributed to each of the other seven farms (as
254 explained in section 2.3). For instance, the following are three sets created to calculate the
255 contribution to the variation in predicted abundances of maize carabids of farm #1: {SDDDDDDD ;
256 DDDDDDDD}, {SDDDDDDS ; DDDDDDDS}, {SDDDDSD ; DDDDDSD} (where each
257 letter represents an individual farm). In a second step, we noted $N[S]_x$ the simulated abundance of
258 maize carabid beetles for the set x with the swine modality and $N[D]_x$ for the dairy modality. The

259 contribution C_{set_x} of the farm i to the variation in carabid abundances for one set x was estimated as
260 $C_{set_x} = N[S]_x - N[D]_x$. The final contribution of farm i , C_i was the average contribution of all sets x :

261
$$C_i = \frac{1}{128} \sum_{x=1}^{128} (N[S]_x - N[D]_x).$$

262 Because maize carabid abundances are related to edge length between winter and spring crops we
263 calculated several descriptors of farm characteristics within the landscape units to explain their
264 contribution to variation in carabid abundances (Table 2): i) the geographical area, ii) the total edge
265 length, with the hypothesis that farm contribution is high when the total edge length of the farm in
266 the landscape unit is high, and iii) grassland area (from minimum to maximum obtained within
267 replications) with the hypothesis that the contribution of a farm will be high if it has large grassland
268 areas when it is under dairy farming (since it will result in a large decrease of winter crop area when
269 changing from swine to dairy production).

270

271 **3. Results**

272 **3.1. Effect of farming systems on cropping plan**

273 Differences in crop surface areas (averaged over the 1250 landscapes-year) at the all farms scale
274 based on percentage of all farms geographic surface area under dairy farming in both landscape
275 units is shown in Figure 3. In each landscape unit, the transition from swine production to dairy
276 production resulted in the same trend: a decrease in winter crop areas and increase in grasslands
277 areas. Dense and open landscape units differed in grassland areas in the swine farm scenario only (0
278 % of area under dairy farming). This result was due to the smaller field size in the dense landscape
279 unit, which resulted in a higher number of fields used as permanent grassland compared with the
280 open landscape unit. Differences in grassland percentages between the two landscape units (67 % of
281 all farm areas in both landscape units) were not observed in the dairy only scenario (100 % of area
282 under dairy farming), the permanent grassland area was lower than the grassland area required by
283 these systems. Consequently, the increased grassland percentage in dense landscape unit was lower

284 than in the open landscape unit. Within a scenario, the variability in crop surfaces was low
285 (variation coefficient below 15 %).

286 Differences in each crop surface at the landscape unit level (averaged over the 1250 landscape-year)
287 based on percentage of the geographic area managed by dairy farms (a) or number of dairy farms
288 (b) is depicted in Figure 4. Results indicated crop differences were the same compared with the all
289 farms scale. Variability within scenario was higher than that observed at the all farms scale. The
290 variation coefficient for all spring crops remained high in all scenarios (~ 30 %). Spring crops are
291 necessary in both farming systems and no specific agronomic constraints are required on the
292 seedlings, therefore increased variability was observed related to a scale effect. The simulated
293 grassland percentage is correlated to the percentage of the landscape unit area managed by dairy
294 farms but locally two scenarios with almost the same percentage of the landscape unit managed by
295 dairy farms (i.e. 50%) could have very different grassland percentage (i.e. from 15 to 30%) (Fig.
296 4a). The differences were due to the spatial organization of farms fields and the farmstead position.
297 Some farmsteads were within the landscape unit and so pastures were more often seeded within
298 landscape unit. For farms with facilities outside the landscape unit, fields managed within landscape
299 unit were more often in winter or spring crops and not grasslands. The wide range of variability in
300 cropping plans for a fixed number of dairy or swine production farms is shown in Figure 4b.
301 Differences in the farm area managed at the landscape unit and the previously explained field
302 organization explain these observations.

303

304 **3.2. Predictions of carabid abundance in simulated landscapes**

305 Predicted abundance of maize and woody carabid species in both landscape units, derived from the
306 number of dairy farms (one to eight) and the geographic surface area covered by dairy production is
307 depicted in Figure 5.

308 Different responses were found between the two carabid species groups and landscape units.
309 Woody species abundance varied non-linearly based on the number of dairy farms in both

310 landscape units (polynomial regression, open landscape unit: $R^2 = 0.46$, F-statistics = 108.0, $P <$
311 0.001; dense landscape unit: $R^2 = 0.37$, F-statistics = 76.1, $P < 0.001$) (Fig. 5.1a). Responses were
312 maximized for different scenarios depending on landscape, i.e. for scenarios with six to seven dairy
313 farms (over a total of eight farms) in the open landscape unit and for scenarios with two to three
314 dairy farms (over a total of eight farms) in the dense landscape unit. In contrast, maize species
315 abundance decreased with increasing number of dairy farms in both landscape units (linear
316 regression, open landscape unit: $R^2 = 0.42$, F-statistics = 181.2, $P < 0.001$; dense landscape unit: R^2
317 = 0.38, F-statistics = 158.8, $P < 0.001$) (Fig. 5.1b). In scenarios where dairy farming dominated,
318 edge length between maize and winter crops was reduced. Hedgerow length did not vary within
319 each landscape unit, therefore decreased maize species abundance was due to decreased edge length
320 between maize and winter crops (cf. 2.1.1.). The decreased abundance of maize species was
321 reduced in dense compared with open landscape units, in relationships with the negative effect of
322 the connectivity of woody elements on maize carabid species.

323 Results regarding the surface area covered by dairy production (Figure 5.2a and 5.2b) showed that
324 abundances of woody species vary non-linearly according to the proportion of dairy farming in both
325 landscape units (polynomial regression, *open* landscape unit: $R^2 = 0.37$, F-statistics = 50.1, $p < 0.001$;
326 *dense* landscape unit: $R^2 = 0.81$, F-statistics = 557.8, $p < 0.001$). Regarding maize species, carabid
327 abundances were aggregated into two distinct clusters, but without relationships with the overall
328 variation in the proportion of surface area under dairy production in both landscape units. It
329 suggests the influence of other factors than the only proportion of surface area under dairy farming.
330 However, within each cluster, carabid abundances tended to decrease with the proportion of dairy
331 farming in the landscape.

332

333 **3.3. Farm contribution to the variation in carabid abundance predicted by the 'maize' model**

334 The 'maize' model farm contribution varied from 0 to 50.4 in the dense landscape unit and from 0.1
335 to 230.1 in the open landscape unit (Table 2). In both landscape units, the highest contributions

336 were associated with large areas farmed at a landscape unit. Five farms with more than 10 ha in a
337 landscape unit were major contributors to variation in carabid abundance. However, data showed no
338 linear relationship between farm contribution and farm area within landscape unit (see farms O4,
339 O8 in open landscape unit and F7 in dense landscape unit). Spearman correlation coefficients
340 indicated long lengths of variable edges (i.e. each field along the edge are cultivated) and increased
341 grassland frequencies were correlated with high contributions (Table 3). More precisely, grassland
342 frequency was more determinant in the open landscape and variable edge length was more
343 determinant in the dense landscape. Thus, the results indicated O8 did not contribute despite
344 farming 7.3 ha in an open landscape unit, O8 showed the absence of variable internal edges, and
345 increased grassland areas were rare (see percentile p80). F7 exhibited comparable variable edge
346 length to F5 and lower relative to F3, which explained its close contribution to F5 yet lower than
347 F3.

348

349 **4. Discussion**

350 *4.1. Effects of farming systems on landscape pattern and carabid abundance*

351 In our study, we simulated carabid abundances in fields or woody elements using statistical
352 modelling. The two models built had a good explicative value (pseudo $R^2 = 0.48$) but the RMSEP
353 analyses revealed high errors of predictions of the models. For both models, these errors were
354 higher for the highest values of predicted abundances. The higher errors of predictions for the
355 'woody' model suggests that variations in predicted abundances by this model should be interpreted
356 carefully. However, we assumed that the models remained useful for a qualitative assessment of
357 contrasted farming and landscape pattern scenarios for the two carabid groups.

358 Our overall results showed abundance of the two carabid groups differed due to landscape pattern
359 effects as a function of farming system but the two models revealed different responses. Maize
360 carabid species experienced positive abundance influence under swine farming system dominance
361 landscapes (i.e. a 500 m radius around crops). This was associated with increased winter crops

362 cultivated areas in swine systems, which resulted in increased edge lengths between winter crops
363 and spring crops. In spite of the high prediction errors (33%) of the '*maize*' model, the important
364 contrast in the predicted abundances between the most extreme farming scenarios in the open
365 landscape unit (about 750 individuals in the 100% swine scenarios *vs.* 200 individuals in the 100%
366 dairy scenarios) suggested a significant effect of farming systems on maize carabid beetles in these
367 landscapes. Vasseur et al. (2013) suggested crop spatial connectivity might be an important driver
368 of arthropod maintenance in cultivated habitats of agricultural landscapes, as it determined the
369 spatio-temporal continuity of resources for these species. At the local scale, Burel et al. (2013)
370 showed carabid beetles moved between adjacent, asynchronous crops (winter and spring), which
371 suggested spatial continuities between seasonal cultivated crops favored resource complementation
372 processes (Dunning et al., 1992) for crop carabid species. Such processes might explain the positive
373 edge length effects between asynchronous crops on crop carabid beetle abundance at the landscape
374 scale observed in our study. Duflot et al. (2016) reported crop configurations decreased carabid
375 abundance, which differed from our results. Duflot et al. (2016) examined an overall carabid beetle
376 community in cultivated wheat fields and total, gamma biodiversity, where we investigated maize
377 carabid species groups in cultivated maize and local, alpha biodiversity, variables, which likely
378 explains the different results observed between our studies. Our results emphasized the importance
379 of crop diversity and configuration in biodiversity, as reinforced by previous studies (e.g. Duflot et
380 al., 2014).

381 By contrast, abundances of woody carabid species varied non-linearly according to the number of
382 dairy farms and the proportion of area under dairy farming. Abundances of these carabid species
383 were slightly higher in landscapes with both types of farming systems (i.e. with similar relative
384 surface areas of spring crops, winter crops and grassland) than in landscapes with specialization of
385 production (either towards dairy or swine production). A positive effect of land-use diversity on
386 woody species at a fine spatial scale (i.e. in a 50 m radius around hedgerows) served to explain this
387 result. Fahrig et al. (2011) provided evidence that land-use diversity reflected compositional

388 landscape heterogeneity. Additional study indicated it a key driver of biodiversity in agricultural
389 landscapes (Benton et al., 2003; Fahrig et al., 2011) and diverse land-use showed beneficial effects
390 on carabid communities (Batáry et al., 2007; Millán de la Peña et al., 2003). However, these results
391 should be interpreted with caution since the range of variation in predicted abundances is very low
392 (a few individuals) and the errors of prediction of the 'woody' model are high (49%). This
393 underlines the difficulty to explain properly the presence and abundance of species such as woody
394 carabid species that are often very scarce in actual agricultural landscapes.

395 If the effects of farming system processes seemed similar in both landscape units, results indicated
396 the range of variation in carabid abundance across scenarios was smaller in the dense (complex)
397 landscape relative to the open landscape. This result can be compared to the mitigation role in
398 landscape complexity to the intensification effect on biodiversity (Tscharntke et al., 2005).

399 ***4.2. Farm contribution to variation in predicted carabid abundances***

400 Our modeling design was successful in elucidating the importance of individual farm level to
401 explain variations in abundances of maize carabid beetles in the agricultural mosaic. The variability
402 of contribution to deviation in predicted carabid abundance among farms was related to farm size,
403 but also to the probability to seed grassland in dairy systems and to the edge length between fields
404 where crops were sown. Brunschwig et al. (2006) and Thenail and Baudry (2004) identified a
405 relationship between grassland seeding in dairy systems and field distance to the farmstead.
406 Farmstead location to evaluate farm influence on predicted carabid abundance in a defined spatial
407 area was clearly identified in these and our study. However, databases would not provide the
408 farmstead location due to confidentiality clauses (INSEE, 2010) which limits the possible use of
409 such database to identify the main farms able to increase the carabid beetle abundance in a specific
410 area. In our simulations, we controlled farmstead location and analyzed location role to farm
411 contribution to variation in predicted carabid abundance. Farms with the farmstead close to or
412 within the processing landscape unit exhibited higher contributions to variation than farmsteads
413 with a greater distance to processing. Factors responsible for differences in farm contribution to

414 variability between landscape units can be explained by landscape structural differences. Indeed,
415 dense landscape unit revealed large surface areas of permanent grasslands and high hedgerow
416 lengths, which limited the edges or interfaces between crop fields. However, these remaining
417 interfaces between crop fields in the dense landscape unit were individually more important for the
418 abundance of crop carabid beetle than crop fields interfaces in the open landscape unit. The
419 interface length criterion was therefore more important at the farm level in the dense landscape
420 compared with the open landscape.

421

422 *4.3. Implications for biodiversity management in agricultural landscapes*

423 Predicting species abundances in agricultural landscapes using statistical modeling has limitations
424 due to the uncertainty of predictions of ecological models (Barry and Elith, 2006). Although our
425 models were built using an extensive dataset, there are still additional drivers of carabid abundances
426 that could improve the predicting ability of models. However, our results on maize carabid species
427 suggest that some specific organization of annual crops in space, maximizing the edges between
428 asynchronous crops, could increase the abundances of these species. Presently, green way policies
429 focus on the conservation of existing networks of semi-natural habitats. The present study suggests
430 that farmers might contribute to create cropped green veining for the biodiversity associated to the
431 agricultural mosaic. Our modeling approach allowed highlighting the possible contribution of
432 farming systems and individual farmers to generate landscape configurations that are important for
433 crop carabid species. These landscape configurations could result of localization of asynchronous
434 crops on juxtaposed fields by an individual farmer. If our results showed this localization is possible
435 one year, it should be tested over several years by including a new rule in our sets. As the field
436 interfaces are often longer with fields of other farms than within the farm due to fragmentation of
437 farms' land, another solution would be the coordination among farmers in association with other
438 stakeholder types. For example, a group of farmers, the mayor, and a watershed advisor reduced the
439 risk of erosion by engaging in a dialogue regarding grass strips, storage ponds, and an intercrop

440 management period (Souchère et al., 2010). For biodiversity management, this type of solution
441 should consider crop localization, under consultation with a landscape planner. Indeed, crop spatial
442 organization is easier to understand at a collective scale, rather than individually due to the spatial
443 constraints at the farm scale (e.g. farmstead distance, soil quality, accessibility, among other
444 factors). Green way policies do not include collective responses, which have to pass through
445 different mechanisms. For example, Lucas et al. (2015) described an enhancement of collective
446 farming and several farmers already engaged in collective cropping plans (Martin et al., 2015),
447 which should be viewed as an opportunity to improve biodiversity at landscape scales.

448

449 **5. Conclusion**

450 Our multi-level and agro-ecological modeling framework facilitated the evaluation of different
451 farming system (dairy and swine) impacts on landscape patterns and carabid beetle abundance. We
452 showed the over-representation of swine farming resulted in an increased edge length between
453 maize and winter cereals and consequently enhanced abundances of maize carabid beetles. Both
454 farming systems seemed necessary to favor woody carabid beetles. Moreover, our results showed,
455 given a set of farming systems in a landscape, variability in crop patterns and edge length between
456 winter crops and spring crops emerged. This indicated farmers have possibilities to modify crop
457 allocation with the objective to increase landscape patterns in relationship to biodiversity issues,
458 while maintaining their production objectives. Therefore, we recommend the following: i) crop
459 allocation modifications at the farm level, e.g. decrease large field size by dividing fields and
460 increase potential interfaces, increase field accessibility to animals from the farmstead; and ii)
461 explore modifications at a collective level, since landscape patterns result from several farmers'
462 decisions. Currently, these solutions are not integrated in policies aiming to promote biodiversity, as
463 green veining.

464

465 **Acknowledgements:** We gratefully thank Jean-Luc Roger, Gérard Savary and Eloïse Couthouis for fieldwork, Tristan Jean-Charles for testing the crop
466 decision model and Clarisse Amiotte for interviewing kindly farmers. We also thank Nicolas Parisey and Sylvain Poggi for their counsels to obtain the
467 ecological models.

468 **Funding:** This work has been funded by the French Ministry of Ecology (DIVA-Agriconnect project); INRA Payote research group; and Zone Atelier
469 Armorique (supported by CNRS and INRA).

470

471 **6. References**

- 472 Akplogan, M., de Givry, S., Metivier, J.P., Quesnel, G., Joannon, A., Garcia, F., 2013. Solving the
473 crop allocation problem using hard and soft constraints. *Rairo-Oper. Res.* 47, 151–172.
474 doi:10.1051/ro/2013032
- 475 Barry, S., Elith, J., 2006. Error and uncertainty in habitat models. *J. Appl. Ecol.* 43, 413–423.
476 doi:10.1111/j.1365-2664.2006.01136.x
- 477 Batáry, P., Báldi, A., Szél, G., Podlussány, A., Rozner, I., Erdős, S., 2007. Responses of grassland
478 specialist and generalist beetles to management and landscape complexity: Grassland
479 specialist and generalist beetles. *Divers. Distrib.* 13, 196–202. doi:10.1111/j.1472-
480 4642.2006.00309.x
- 481 Bennett, G., Mulongoy, K.J., 2006. Review of experience with ecological networks, corridors and
482 buffer zones, CBD Technical Series.
- 483 Benton, T.G., Vickery, J.A., Wilson, J.D., 2003. Farmland biodiversity: is habitat heterogeneity the
484 key? *Trends Ecol. Evol.* 18, 182–188. doi:10.1016/S0169-5347(03)00011-9
- 485 Bianchi, F.J.J., Booij, C.J., Tscharrntke, T., 2006. Sustainable pest regulation in agricultural
486 landscapes: a review on landscape composition, biodiversity and natural pest control. *Proc.*
487 *R. Soc. B Biol. Sci.* 273, 1715–1727. doi:10.1098/rspb.2006.3530
- 488 Billeter, R., Liira, J., Bailey, D., Bugter, R., Arens, P., Augenstein, I., Aviron, S., Baudry, J.,
489 Bukacek, R., Burel, F., Cerny, M., De Blust, G., De Cock, R., Diekötter, T., Dietz, H.,
490 Dirksen, J., Dormann, C., Durka, W., Frenzel, M., Hamersky, R., Hendrickx, F., Herzog, F.,
491 Klotz, S., Koolstra, B., Lausch, A., Le Coeur, D., Maelfait, J.P., Opdam, P., Roubalova, M.,
492 Schermann, A., Schermann, N., Schmidt, T., Schweiger, O., Smulders, M.J.M., Speelmans,
493 M., Simova, P., Verboom, J., Van Wingerden, W.K.R.E., Zobel, M., Edwards, P.J., 2008.
494 Indicators for biodiversity in agricultural landscapes: a pan-European study: Biodiversity in
495 European Agro-ecosystems. *J. Appl. Ecol.* 45, 141–150. doi:10.1111/j.1365-
496 2664.2007.01393.x
- 497 Boussard, H., Baudry, J., 2014. *Chloe2012* : a software for landscape pattern analysis. INRA,
498 Rennes, France.
- 499 Boussard, H., Martel, G., Vasseur, C., 2010. Spatial links specifications in the APILand simulation
500 approach: an application to the coupling of a farm model and a carabid population model.
501 *LandMod 2010 Int. Conf. Integr. Landsc. Model.* (11p).
- 502 Brunschwig, G., Josien, E., Bernhard, C., 2006. Contraintes géographiques et modes d'utilisation
503 des parcelles en élevage bovin laitier et allaitant. *Fourrages* 185, 83–95.
- 504 Burel, F., Aviron, S., Baudry, J., Le Feon, V., Vasseur, C., 2013. The Structure and Dynamics of
505 Agricultural Landscapes as Drivers of Biodiversity, in: Bojie, F., K. Bruce Jones (Eds.),
506 *Landscape Ecology for Sustainable Environment and Culture*. Springer, pp. 285–308.
- 507 Burnham, K.P., Anderson, D.R. (Eds.), 2002. *Model selection and multimodel inference: a practical*
508 *information-theoretic approach*, second edition. New-York.
- 509 Castellazzi, M.S., Matthews, J., Angevin, F., Sausse, C., Wood, G.A., Burgess, P.J., Brown, I.,
510 Conrad, K.F., Perry, J.N., 2010. Simulation scenarios of spatio-temporal arrangement of
511 crops at the landscape scale. *Environ. Model. Softw.* 25, 1881–1889.
512 doi:10.1016/j.envsoft.2010.04.006

- 513 Chardon, X., Rigolot, C., Baratte, C., Espagnol, S., Raison, C., Martin-Clouaire, R., Rellier, J.-P.,
514 Le Gall, A., Dourmad, J.Y., Piquemal, B., Leterme, P., Paillat, J.M., Delaby, L., Garcia, F.,
515 Peyraud, J.L., Poupa, J.C., Morvan, T., Faverdin, P., 2012. MELODIE: a whole-farm model
516 to study the dynamics of nutrients in dairy and pig farms with crops. *animal* 6, 1711–1721.
517 doi:10.1017/S1751731112000687
- 518 Dufлот, R., Aviron, S., Ernoult, A., Fahrig, L., Burel, F., 2015. Reconsidering the role of 'semi-
519 natural habitat' in agricultural landscape biodiversity. *Ecol. Res.* 30, 75–83.
- 520 Dufлот, R., Ernoult, A., Burel, F., Aviron, S., 2016. Landscape level processes driving carabid crop
521 communities in dynamic farmlands. *Popul. Ecol.* 58, 265–275.
- 522 Dufлот, R., Georges, R., Ernoult, A., Aviron, S., Burel, F., 2014. Landscape heterogeneity as an
523 ecological filter of species traits. *Acta Oecologica* 56, 19–26.
524 doi:http://dx.doi.org/10.1016/j.actao.2014.01.004
- 525 Dunning, J.-B., Danielson, B.-J., Pulliam, H.-R., 1992. Ecological processes that affect populations
526 in complex landscapes. *Oikos* 65, 169–175.
- 527 Ekroos, J., Hyvönen, T., Tiainen, J., Tiira, M., 2010. Responses in plant and carabid communities to
528 farming practises in boreal landscapes. *Agric. Ecosyst. Environ.* 135, 288–293.
- 529 Fahrig, L., 2003. Effects of habitat fragmentation on biodiversity. *Annu. Rev. Ecol. Evol. Syst.* 34,
530 487–515.
- 531 Fahrig, L., Baudry, J., Brotons, L., Burel, F.G., Crist, T.O., Fuller, R.J., Sirami, C., Siriwardena,
532 G.M., Martin, J.-L., 2011. Functional landscape heterogeneity and animal biodiversity in
533 agricultural landscapes: Heterogeneity and biodiversity. *Ecol. Lett.* 14, 101–112.
534 doi:10.1111/j.1461-0248.2010.01559.x
- 535 Fahrig, L., Merriam, G., 1994. Conservation of fragmented populations. *Conserv. Biol.* 8, 50–59.
- 536 Fahrig, L., Merriam, G., 1985. Habitat patch connectivity and population survival. *Ecology* 66,
537 1762–1768.
- 538 Fitzgibbon, C.D., 1997. Small mammals in farm woodlands : The effects of habitat, isolation and
539 surrounding land-use patterns. *J. Appl. Ecol.* 34, 530–539.
- 540 Garcia, F., Faverdin, P., Delaby, L., Peyraud, J.L., 2005. Tournesol: un modèle pour simuler les
541 assolements en exploitation bovine laitière. *Rencontres Autour Rech. Sur Rumin.* 195–198.
- 542 Hubert-Moy, L., Nabucet, J., Vannier, C., Lefebvre, A., 2012. Mapping ecological continuities:
543 which data for which territorial level? Application to the forest and hedge network. *Int. J.*
544 *Geomat. Spat. Anal.* 22, 619–640.
- 545 INSEE, 2010. Guide du secret statistique. INSEE, Paris (FRA).
- 546 Kleijn, D., Sutherland, W.J., 2003. How effective are European agri-environment schemes in
547 conserving and promoting biodiversity. *J. Appl. Ecol.* 40, 947–969.
- 548 Koivula, M.J., 2011. Useful model organisms, indicators, or both? Ground beetles (Coleoptera,
549 Carabidae) reflecting environmental conditions. *ZooKeys* 287–317.
- 550 Krauss, J., Bommarco, R., Guardiola, M., Heikkinen, R.K., Helm, A., Kuussaari, M., Lindborg, R.,
551 Öckinger, E., Pärtel, M., Pino, J., Pöyry, J., Raatikainen, K.M., Sang, A., Stefanescu, C.,
552 Teder, T., Zobel, M., Steffan-Dewenter, I., 2010. Habitat fragmentation causes immediate
553 and time delayed biodiversity loss at different trophic levels. *Ecol. Lett.* 13, 597–605.
- 554 Kromp, B., 1999. Carabid beetles in sustainable agriculture: a review on pest control efficacy,
555 cultivation impacts and enhancement. *Agric. Ecosyst. Environ.* 74, 187–228.
- 556 Le Roux, X., Barbault, R., Baudry, J., Burel, F., Doussan, I., Garnier, E., Herzog, F., Lavorel, S.,
557 Lifran, R., Roger-Estrade, J., Sarthou, J.P., Trommetter, M., 2008. Agriculture et
558 biodiversité. Valoriser les synergies, Rapport INRA Expertise scientifique collective.
559 France.
- 560 Maisonhaute, J.-E., 2010. Influence de la structure du paysage sur l'assemblage des prédateurs
561 terricoles dans les zones agricoles non cultivées. 100pp
- 562 Maisonhaute, J.-É., Peres-Neto, P., Lucas, É., 2010. Influence of agronomic practices, local
563 environment and landscape structure on predatory beetle assemblage. *Agric. Ecosyst.*
564 *Environ.* 139, 500–507.

565 Martin, G., Moraine, M., Ryschawy, J., Magne, M.-A., Sarthou, J.-P., Duru, M., Therond, O., 2015.
566 Coll-ICLS: a participatory tool to support crop-livestock integration among farms. Presented
567 at the 5th International Symposium for Farming Systems Design, Montpellier (FRA), pp.
568 207–208.

569 Maxime, F., Mollet, J.-M., Papy, F., 1995. Aide au raisonnement de l'assolement en grande culture.
570 Cah. Agric. 4, 351–362.

571 Millán de la Peña, N., Butet, A., Delettre, Y., Morant, P., Burel, F., 2003. Landscape context and
572 carabid beetles (Coleoptera: Carabidae) communities of hedgerows in western France.
573 Agric. Ecosyst. Environ. 94, 59–72. doi:10.1016/S0167-8809(02)00012-9

574 Millennium Ecosystem Assessment, 2005. Ecosystems and human well-being. Island Press
575 Washington, DC.

576 Moilanen, A., Hanski, I., 1998. Metapopulation dynamics: effect of habitat quality and landscape
577 structure. Ecology 79, 2503–2515.

578 Ouin, A., Paillat, G., Butet, A., Burel, F., 2000. Spatial dynamics of wood mouse (*Apodemus*
579 *sylvaticus*) in an agricultural landscape under intensive use in the Mont Saint Michel bay
580 (France). Agric. Ecosyst. Environ. 78, 159–165.

581 Pannell, D.J., 1996. Lessons from a decade of whole-farm modeling in Western Australia. Rev.
582 Agric. Econ. 18, 373–383.

583 Papy, F., 2001. Interdépendance des systèmes de culture dans l'exploitation agricole, in:
584 Modélisation des agroécosystèmes et aide à la décision, 2001, Malézieux E., Trébuil G.,
585 Jaeger M. (eds)., Editions CIRAD-INRA, collection Repères (51-74)

586 Prud'homme, C., Fages, J.-G., Lorca, X., 2014. Choco3 Documentation.

587 Puech, C., Baudry, J., Joannon, A., Poggi, S., Aviron, S., 2014. Organic vs. conventional farming
588 dichotomy: Does it make sense for natural enemies? Agric. Ecosyst. Environ. 194, 48–57.

589 Rainio, J., Niemelä, J., 2003. Ground beetles (Coleoptera: Carabidae) as bioindicators. Biodivers.
590 Conserv. 12, 487–506.

591 Ramonet, Y., Giteau, J.-L., Tersiguel, E., Martel, G., 2014. Le choix des cultures dans les
592 exploitations porcines en Bretagne. Innov. Agron. 39, 43–53.

593 Reganold, J.P., Jackson-Smith, D., Batie, S.S., Harwood, R.R., Kornegay, J.L., Bucks, D., Flora,
594 C.B., Hanson, J.C., Jury, W.A., Meyer, D., others, 2011. Transforming US agriculture.
595 Science 332, 670–671.

596 Ricketts, T.H., 2001. The matrix matters: effective isolation in fragmented landscapes. Am. Nat.
597 157, 87–99.

598 Roche, B., Amiotte, C., Boussard, H., Joannon, A., Martel, G., 2013. Crop acreage allocation
599 decisions on intensive mixed crop-livestock farms. 64th Annual Meeting of the European
600 Federation of Animal Science, Nantes, France.

601 Rotz, C., Mertens, D., Buckmaster, D., Allen, M., Harrison, J., 1999. A dairy herd model for use in
602 whole farm simulations. J. Dairy Sci. 82, 2826–2840.

603 Rounsevell, M.D.A., Annetts, J.E., Audsley, E., Mayr, T., Reginster, I., 2003. Modelling the spatial
604 distribution of agricultural land use at the regional scale. Agric. Ecosyst. Environ. 95, 465–
605 479.

606 Schils, R.L.M., Olesen, J.E., del Prado, A., Soussana, J.F., 2007. A review of farm level modelling
607 approaches for mitigating greenhouse gas emissions from ruminant livestock systems.
608 Livest. Sci. 112, 240–251. doi:10.1016/j.livsci.2007.09.005

609 Souchère, V., Millair, L., Echeverria, J., Bousquet, F., Le Page, C., Etienne, M., 2010. Co-
610 constructing with stakeholders a role-playing game to initiate collective management of
611 erosive runoff risks at the watershed scale. Environ. Model. Softw. 25, 1359–1370.
612 doi:10.1016/j.envsoft.2009.03.002

613 Taylor, P.D., Fahrig, L., Henein, K., Merriam, G., 1993. Connectivity is a vital element of
614 landscape structure. Oikos 68, 571–573.

615 Tersiguel, E., Ramonet, Y., Giteau, J.-L., Martel, G., 2012. Déterminants du choix de l'assolement
616 par les éleveurs de porcs en Bretagne, 44^{ème} Journées de la Recherche Porcine, IFIP -
617 Institut du Porc ; INRA, ACTN, pp. 103–108.

618 Thenail, C., Baudry, J., 2004. Variation of farm spatial land use pattern according to the structure of
619 the hedgerow network (bocage) landscape: a case study in northeast Brittany. *Agric.
620 Ecosyst. Environ.* 101, 53–72. doi:10.1016/S0167-8809(03)00199-3

621 Tilman, D., Fargione, J., Wolff, B., D'Antonio, C., Dobson, A., Howarth, R., Schindler, D.,
622 Schlesinger, W.H., Simberloff, D., Swackhamer, D., 2001. Forecasting agriculturally driven
623 global environmental change. *Science* 292, 281–284.

624 Tschardtke, T., Klein, A.M., Kruess, A., Steffan-Dewenter, I., Thies, C., 2005. Landscape
625 perspectives on agricultural intensification and biodiversity – ecosystem service
626 management. *Ecol. Lett.* 8, 857–874. doi:10.1111/j.1461-0248.2005.00782.x

627 Varchola, J.M., Dunn, J.P., 1999. Changes in ground beetle (Coleoptera: Carabidae) assemblages in
628 farming systems bordered by complex or simple roadside vegetation. *Agric. Ecosyst.
629 Environ.* 73, 41–49.

630 Vasseur, C., Joannon, A., Aviron, S., Burel, F., Meynard, J.M., Baudry, J., 2013. The cropping
631 systems mosaic: How does the hidden heterogeneity of agricultural landscapes drive
632 arthropod populations? *Agric. Ecosyst. Environ.* 166, 3–14. doi.org/10.1016/j.
633

634

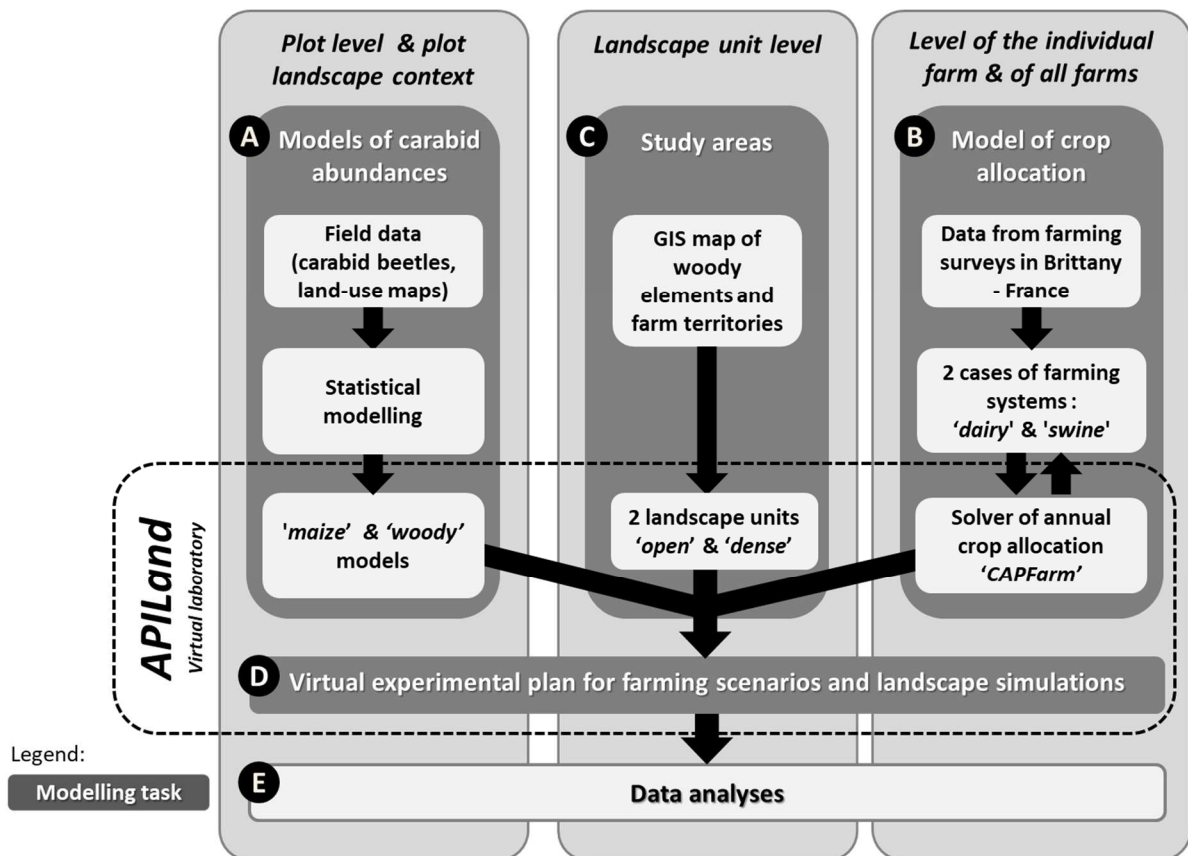


Figure 1: Overall view of the modelling approach used to simulate the diversity of agricultural landscapes based on crop allocation decision rules at farm level and to estimate biodiversity according to landscape configuration.

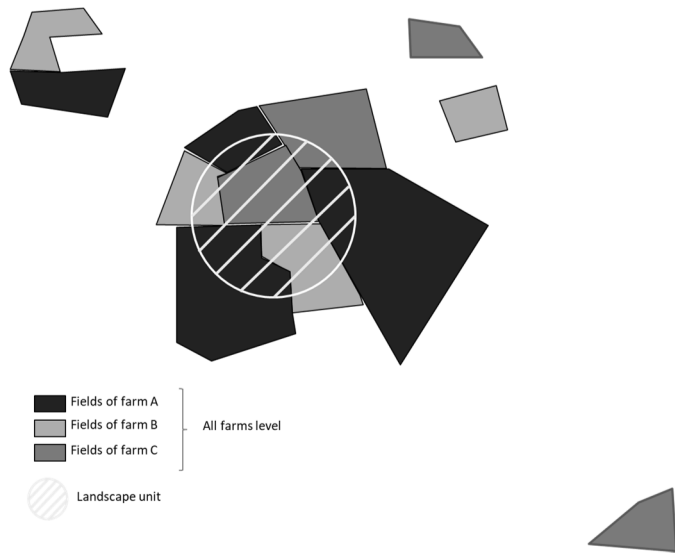


Figure 2: Landscape unit and farm / all farms levels considered in the modelling approach.

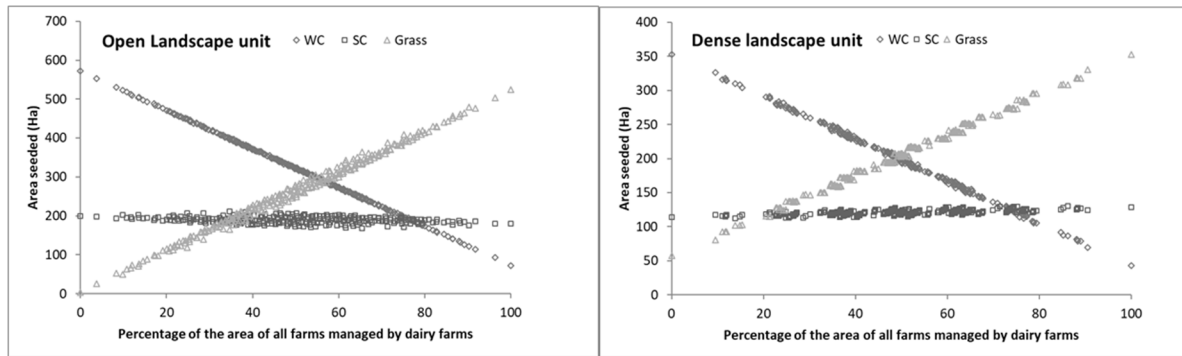
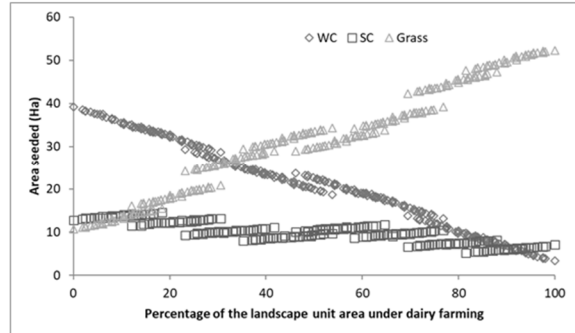
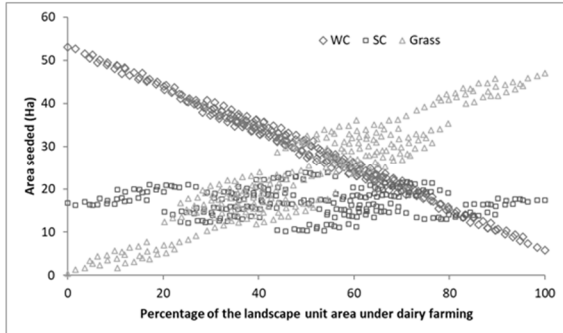


Figure 3: Averaged surface area (over the 1250 simulated landscapes x years) of winter crops (WC), spring crops (SC) and grassland (Grass) at the all farm level according to the percentage of the geographical area of all farms under dairy farming (left: *open* landscape unit, right: *dense* landscape unit).

Open Landscape unit

Dense Landscape unit

a/ Area seeded in landscape unit according to the percentage of the area managed by dairy farm



b/ Area seeded in landscape unit according to the number of dairy farm

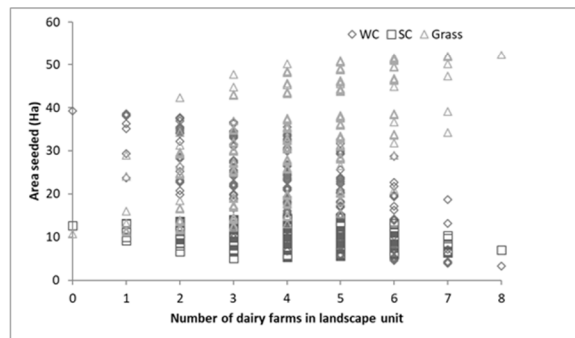
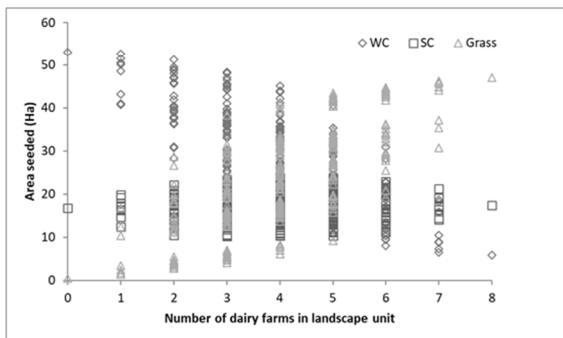
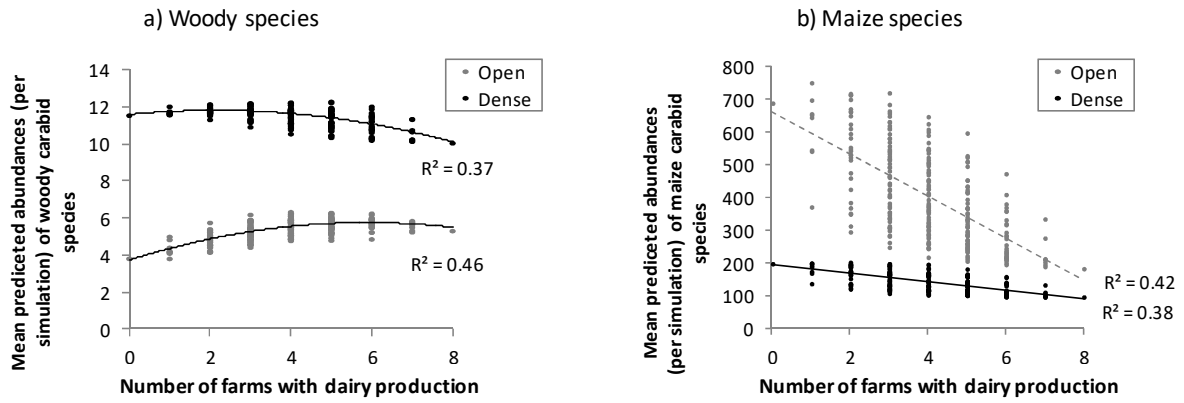


Figure 4: Average surface area (over the 1250 simulated landscapes x years) of winter crops (WC), spring crops (SC) and grassland (Grass) at the landscape unit level according to (a) the percentage of area under dairy farming and (b) the number of dairy farms in the *open* and *dense* landscape units.

1) Predicted abundances according to number of dairy farms



2) Predicted abundances according to surface area with dairy production

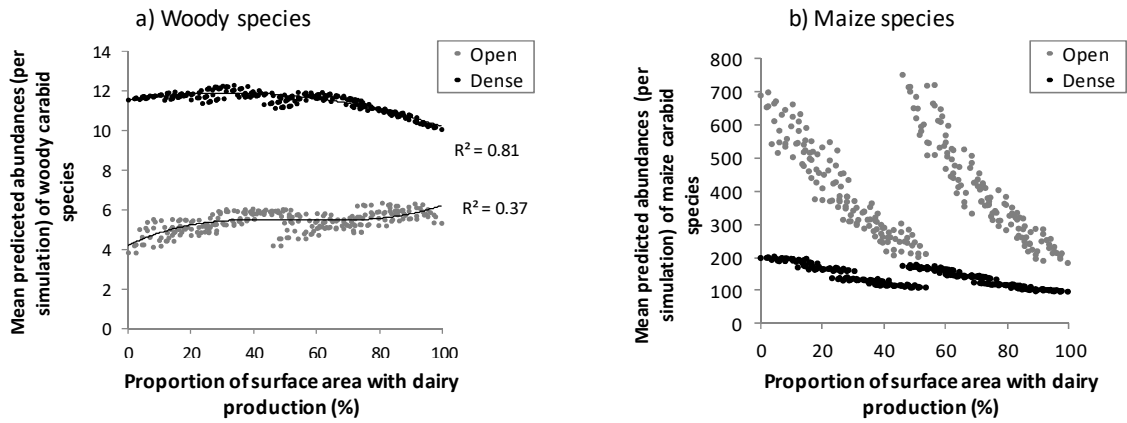


Figure 5: Predicted abundances of woody (a) and maize (b) species according to 1) the number of dairy farms and 2) the proportion of surface area under dairy farming in the dense and open landscape units.

Table 1: Set of rules in Swine and Dairy farming systems for each crop.

Group	Crop	Code	Allowed previous crop	Field size	Min total area	min return time	max rep.	min dur.	max dur.	never on	Distance from farm's head	
SFS	Maïze-Corn	MC	WH-BA	>0.5ha	max (5ha ; 5% UAA)	2	0	-	1	-	-	
	Wheat	WH	MC-RA	>0.5ha	max (5ha ; 10% UAA)	2	0	-	1	-	-	
	Barley	BA	WH	>0.5ha	max (5ha ; 5% UAA)	3	0	-	1	-	-	
	Rape oil seed	RA	WH-BA	>0.5ha	max (5ha ; 10% UAA)	3	0	-	1	Drained plot	-	
	Permanent Grassland	PG	PG	<0.5ha	-	-	-	-	-	-	-	
Maize	-	MA	-	>0.5ha	30% UAA	-	-	-	-	-	-	
Maize	Maize-corn	MC	LO-PA-MA-WH-LU	-	-	-	3	-	1	-	> 400m	
Maize	Maize-Silage	MS	LO-PA-MA-WH-LU	-	-	-	1	-	1	-	< 400m	
Weed	-	WE	-	-	33% UAA	-	-	-	-	-	-	
DFS	Weed	Lolium	LO	MA-WH	>0.5ha	-	-	0	5	7	-	< 5000m
	Weed	Pasture	PA	MA	>0.5ha	16,5% UAA	2	0	5	10	-	< 400m
	Weed	Permanent Grassland	PG	PG	<0.5ha	-	-	-	-	-	-	-
-	Wheat	WH	MA-LU	>0.5ha	max (2ha; 7% UAA)	-	-	-	1	-	-	
-	Lucerne	LU	WH	>0.5ha	max (2ha; 5% UAA)	8	-	3	4	-	> 400m	

SFS: Swine farming system; DFS: Dairy farming system; UAA: Usable Agricultural Area; Return time is the number of years to wait between two sowings of the same crop on a same field; Max rep: Maximum number of successive seeding of a same crop on the same field; min dur. and max dur.: Minimum and maximum duration of crop.

Table 2: Contribution of farms to the variation in maize carabid abundances in both landscape units and main characteristics of farms.

Landscape unit	Farm	FC	UAA-S (ha)	FI (m)	VI(m)	VI-IN (m)	Min (ha)	p20 (ha)	Mean (ha)	p80 (ha)	Max (ha)
Dense	F6	0.0	1.5	605	0	0	0.0	0.0	0.7	1.5	1.5
	F8	1.2	1.6	557	134	0	0.0	0.0	0.4	1.2	1.2
	F1	1.4	1.2	514	99	0	0.0	0.0	0.5	0.9	0.9
	F4	4.9	5.0	2232	561	0	0.0	1.5	2.5	3.4	4.7
	F2	5.5	2.3	928	275	185	0.0	0.0	0.8	1.2	1.7
	F7	18.9	29.0	9887	731	282	9.1	13.9	15.9	17.8	22.2
	F5	20.9	7.6	2697	676	353	2.0	3.9	4.4	5.1	5.6
	F3	50.4	14.6	3928	1579	936	7.7	10.5	11.4	12.4	14.0
Open	O8	0.1	7.3	618	713	0	0.0	0.0	1.3	0.4	7.3
	O7	14.1	3.2	36	895	403	0.0	1.3	2.1	3.2	3.2
	O3	16.2	1.2	297	222	0	0.0	0.3	0.8	1.2	1.2
	O6	16.4	2.5	305	345	0	0.0	0.1	1.3	2.5	2.5
	O5	42.2	4.7	301	797	318	0.0	0.7	2.1	4.0	4.7
	O1	72.7	14.3	1317	1660	602	0.0	4.4	10.1	12.8	13.9
	O4	122.9	20.5	2884	2089	739	0.9	7.0	9.8	12.6	20.5
	O2	230.1	16.6	961	1675	721	5.5	13.4	14.9	16.6	16.6

FC: farm contribution to the variation in maize carabid abundances – UAA-S: farm's area within the landscape unit – FI: length of fixed edge (one of the fields along the edge has a fixed land use in all simulations, eg. permanent grassland, wood or hedgerow) – VI: length of variable edges (both fields over the edge can have different covers according the years, eg. crops, temporary grasslands and lucerne) – VI-IN: variable edge length between two fields of the farm – Next 5 variables describes grassland area in the farm in all years simulated (min, average, max and 20/80 percentiles).

Table 3: Correlations of farm contribution to variation of maize carabid beetles with farm characteristics: only correlation with a p-value lower to 0.01 in one of the landscape unit are listed (D: dense landscape unit – O: Open landscape unit)

Variables	D		O	
	Spearman coefficient	P-value	Spearman coefficient	P-Value
VI	0.93	0.002	<i>ns</i>	<i>ns</i>
VI - IN	0.94	0.001	0.78	0.022
Average - p20	0.8	0.017	0.85	0.008
Average - mean	0.78	0.022	0.84	0.001
Average - p80	<i>ns</i>	<i>ns</i>	0.92	0.001

VI: length of variable edges (both fields over the edge can have different covers according to the years, eg.

crops, temporary grasslands and lucerne) – VI-IN: variable edge length between two fields of the farm –

Average p20 / mean / p80: average value of 20th percentile, mean and 80th percentile of grassland area