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Identification and assessment of ecotoxicological hazards attributable to pollutants in
 urban wet weather discharges

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12 <u>Abstract</u>

13 The phenomenon of urbanization leads to considerable pressure on urban/suburban aquatic ecosystems. Urban Wet Weather Discharges (UWWDs) during rainfall events are a major source of pollutants 14 15 leached onto and into urban surfaces and sewers, which in turn affect aquatic ecosystems. The aim of this study is to assess the ecotoxicity of the different compounds identified in UWWDs and identify the 16 17 hazard represented by each of them. To this end, hazard quotients (HQ) were calculated for each compound detected in UWWDs, based on their Predicted No Effect Concentration (PNEC) values and 18 their maximum concentration (MEC_{max}) found in the dissolved part of UWWDs. For the 207 compounds 19 identified in UWWDs, sufficient data existed for 165 of them to calculate their PNEC. The ecotoxicity 20 21 of these compounds is greatly variable. Pesticides represent a high proportion of the wide variety of hazardous compounds whose HQ were calculated (a total of 94 HQ), and they are among the most 22 hazardous pollutants (HQ >1000) transported by stormwater. On the contrary, the hazard of Combined 23 Sewer Overflows (CSO) is mainly linked to heavy metals and pharmaceutical compounds. 24 Consequently, the monitoring of these pollutants is a priority in the future. Lastly, it is to bear in mind 25 that the hazard level of certain pollutants could be underestimated, due to their adsorption on particles, 26 leading their low presence in the dissolved phase of UWWDs. So, an in-depth study of these pollutants 27 28 will be required to clarify their effects on aquatic organisms.

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32 <u>Introduction</u>

33 Urban Wet Weather Discharges (UWWDs) are complex and variable polluted matrices, notably due to the different types of catchments leached by the rain, and the variable flows for different events.¹ 34 35 Many micropollutants have been detected in these effluents as a function of population density and anthropogenic activities. Among them, we can cite here: (i) European priority substances such as 36 Polycyclic Aromatic Hydrocarbons (PAHs),²⁻⁵ diethylexylphtalate,^{4,6}4-nonylphenol,^{4,6} bisphenol A and 37 alkylphenolethoxylate⁷, (ii) other micropollutants like heavy metals (copper, lead, zinc, etc.), 38 polychlorobiphenyls (PCBs), pesticides, phthalates^{3,5,8,9,10}, and (iii) pharmaceutical residues (PC) in the 39 particular case of the Combined Sewer Overflows (CS0)^{11,12}. These micropollutants have also already 40 been detected in receiving river water having received these discharges.^{13,14} Many of them are potentially 41 hazardous for aquatic organisms. That is why UWWDs have toxic effects on aquatic species, ^{15,16,17} and 42 can thus modify biological communities.^{15,18,19,20,21} Treatment to diminish UWWD pollutants through 43 abatement are nonexistent or very limited depending on municipal sewer systems, and raw UWWD are 44 often release into (sub)urban aquatic ecosystems. It is therefore necessary to optimize their management. 45 This optimization can concern: (i) a better monitoring of toxic pollutants and their effects, including 46 continuous monitoring with biosensors;^{22,23} (ii) a specific treatment of pollutants;²⁴ and (iii) the 47 suppression or the substitution at source of the toxic products at the origin of these pollutants (in 48 particular, certain detergents, disinfectants and pharmaceuticals) by less toxic but equally efficient 49 50 products. For that, it is necessary to identify and prioritize the hazardous pollutants present in UWWDs.

51 To establish the list of the compounds most concerned by the ecotoxicological hazard 52 represented by a specific urban effluent, many authors have proposed to formulate and apply a hazard quotient based on the ratio "Concentration of the selected pollutant in filtered polluted 53 matrices"/"Predicted No Effect Concentration (PNEC) of the molecules selected".^{25,26,} This quotient is 54 55 used to distinguish the hazardousness of the different compounds and families of compounds found in UWWDs. Orias and Perrodin²⁵ proposed comparing the maximum Measured Effluent Concentrations 56 (MECmax) measured in hospital effluents to their PNEC value, which is the method used in the current 57 study. This method was also applied by many other authors for contaminated matrices such as rivers 58 polluted by raw or treated municipal wastewaters polluted by various pharmaceuticals,²⁶ pesticides,²⁷ or 59

lakes by antibiotics.²⁸ The choice of using the maximum value of MEC (MECmax) is to be place us in
the worst (extreme conditions) case scenario^{25,27} and in the context of the precautionary principle. This
method was applied here for the first time to the UWWDs, translating its originality.

63 Consequently, the main objectives of our study were to: (i) conduct a review of the ecotoxicity data available for the different compounds commonly detected in UWWDs; (ii) assess the 64 concentrations of these pollutants in UWWDs performing a literature review of UWWDs pollutant 65 compositions; and (iii) calculate the potential hazard of these pollutants for aquatic ecosystems. Several 66 studies have already proposed to identify and prioritize hazardous pollutants in UWWDs.²⁹⁻³² 67 Nevertheless, they generally considered only the stormwater (no studies were reported for the CSO). 68 Moreover, these pollutant hazard assessments were based only on the intrinsic properties of the 69 pollutants (bioaccumulation, biodegradability, physicochemistry, ecotoxicity) and did not take into 70 71 account the concentrations found in situ.

In this article, we present the PNEC values inventoried for all the pollutants identified in UWWDs, and the MECmax found in these discharges. Lastly, we calculated the HQ for each pollutant found in UWWDs. We also chose to focus on combined and separate stormwater systems (named here "CSO" and "stormwater" systems) due to the very different sources and types of pollutants they contain.

76 <u>2. Methodology</u>

77 All the pollutants used and analyzed (on dissolved and particles phases of UWWDs) in this study were classified into families of chemical compounds, as detailed in Table 1. Scientific research 78 papers for UWWD pollutants identification were obtained using databases such as the Web of 79 Knowledge, Science Direct and Google Scholar, using key search terms (e.g. "stormwater", "CSO", 80 "UWWD", "pollutants", "chemicals"). An identified paper's bibliography and citations were also 81 checked to find other suitable studies. Finally, papers used in this study were restricted to peer-reviewed 82 research articles published between 1995 and 2016. Other scientific productions (e.g. conference papers, 83 non-peer reviewed procedures) were not retained for this study. 84

85 2.1 Obtaining and determining the PNEC values of micropollutants in UWWDs

To define the hazard of the different pollutants identified in UWWD physico-chemical studies, 86 87 the first step of our method was to determine their ecotoxicity by seeking or calculating their PNEC. PNEC values were found/calculated for all pollutants sought in physico-chemical studies (on particle 88 89 and dissolved phases). In this study, the PNEC values were obtained for all the pollutants using the following approach: (i) When they were available, PNEC values were obtained from free European 90 databases^{33,34} (INERIS and AGRITOX) containing ecotoxicological data from international scientific 91 92 literature, (ii) When values were not available in these databases, PNEC values were calculated using experimental ecotoxicological data found in the EPA ECOTOX database³⁵, following the method 93 proposed by Orias et Perrodin,³⁶ and in agreement with the guidelines proposed by the EU (European 94 Union) in the Technical Guidance Document (TGD)³⁷ (see Table 2). This method is the following: for 95 each compound, bibliographical research using the EPA ECOTOX database allowed identifying all the 96 97 bioassays performed, the trophic level concerned (decomposers, primary producers, primary consumers and secondary consumers) and the chronicity (or not) of the bioassay (e.g. reproduction). Using all the 98 99 ecotoxicity data, an Extrapolation Factor (EF), calculated following the criteria presented in Table S1, was applied to the lowest toxicity values to establish the PNEC of the compound concerned. No 100 101 differentiation was done between the ecotoxicity data used (e.g. EC_{50} , EC_{10}) for the EF and PNEC 102 calculation. However, several minor modifications in comparison to TGD were made: ecotoxicity data on fungi were considered as data on decomposers while data on periphyton were considered to originate 103 104 from chronic bioassays on decomposers. In the case where only acute ecotoxicity data were available, 105 but not for all trophic levels, PNEC values were calculated with the ECOSAR model, by applying an 106 EF to the lowest ecotoxicological value. In all cases, no PNEC values were calculated if at least one acute ecotoxicity value was available for each trophic level. The chronic algae and macrophyte bioassays 107 did not permit lowering the EF when they were the only ecotoxicological data available. Finally, the EF 108 109 was not reduced when rotifer (Brachionus calyciflorus, noted "Bc") bioassays were the only tests performed on the trophic level of secondary consumers.³⁷ Finally, when no PNEC and ecotoxicity data 110 were available in databases and literature, PNEC values were estimated with modelled ecotoxicological 111 data obtained using the ECOSAR method.³⁸ 112

It is important to point out that the authors of several studies did not indicate the speciation or the isomer of the compounds measured. This is the case of chromium, for which we chose to keep the two speciation forms (chromium III and chromium VI), as it has been reported that the annual concentrations of these two chromium species in rainwater samples are similar.³⁹ In the case of dichloroethane, we chose to keep the two isomers (1-1 and 1-2-dichloroethane), while the hazard quotient was calculated twice for this pollutant, using two PNECs.

119 2.2 Obtaining micropollutant MECmax

120 The second step for UWWD pollutant hazard assessment was the MEC_{max} determination for each of them. As described in Orias and Perrodin²⁵, two different approaches can be used to obtain the 121 122 concentration of micropollutants in UWWDs: (i) a theoretical method based on the calculation of PEC 123 (Predicted Effluent Concentration) with the design of a scenario, and (ii) an experimental approach based on analyses of the pollutants in the dissolved phase of the UWWDs. The second method 124 125 (MECmax determination) is more realistic, because it is based on real measured concentrations in UWWDs, contrary to the calculation of PEC values presenting a high level of uncertainty due to the 126 parameters used (e.g. product consumption). Therefore, this option was selected although not all the 127 compounds were analyzed in the discharges. 128

In this study, we will discuss the different choices made to select concentration data (see section 129 3.5). Firstly, we considered studies providing compound concentration values only in the soluble phase 130 (<1.2 µm). Indeed, ecotoxicity assays performed to calculate PNEC values for the different pollutants 131 have relied only on pure liquid solutions. Moreover, it is impossible to evaluate the bioavailability of 132 133 the contaminants in the particulate phase for aquatic organisms. Additionally, taking into account the concentrations on particles could lead to overestimating the hazard (see section 3.5). Regarding the 134 formulation of the Hazard Quotient (HQ), we also chose to separate concentration values for separate 135 and combined sewer systems, thus we calculated the hazard for both cases. The hazards of these two 136 systems differ, as the source and type of pollutants (i.e. the presence of pharmaceuticals in discharges 137 138 from combined sewer overflows) are also different.

139	For each compound sought in the dissolved part of the stormwater/CSO samples, the MEC of
140	the events observed was reported for each study. Finally, the highest effluent concentration ever
141	measured in UWWDs (MEC _{max}) was determined by seeking the highest MEC reported previously, with
142	the aim to be situated in the worst-case scenario, as explained in the introduction. In further research, it
143	would be interesting to make the same work with the median and the lowest values, but that will be
144	limited by the fact that for some compounds, only one concentration value was available in scientific
145	literature.

146 <u>2.3 Calculation of the hazard quotients</u>

147 Finally, for each compound detected in discharges (section 2.2) with an available PNEC value148 (section 2.1), we calculated an hazard quotient using the equation (1):

149 (1) HQ = MECmax./PNEC

150 With:

151 MECmax: the highest effluent concentration ever measured in UWWDs in μ g.L⁻¹;

152 PNEC: Predicted No Effect Concentration in μ g L⁻¹.

153 It was considered that compounds with an HQ up to 1 are potentially hazardous for the 154 environment, and that compounds with the highest HQ are the most hazardous for aquatic ecosystems.

155 On the contrary, compounds with an HQ lower than 1 were considered slightly hazardous.

In this study, 207 compounds were identified in UWWDs. We found or calculated/modelled a
PNEC value representing the ecotoxicity of 165 of them. By seeking the dissolved proportion analyzed
in UWWDs, we were finally able to determine a MECmax for only 61 compounds in "stormwater" and
49 in "CSO". Consequently, we calculated 94 hazard quotients using equation (1) to prioritize their
hazardousness for freshwater ecosystems.

161 3. Results and discussion

162 3.1 Pollutant classification and available data.

Pollutants analyzed in UWWDs were classified as shown in Table 1. All the PNECs, 163 concentrations and hazard data were analyzed as per the pollutant ranking presented in Table 1.

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Table 1: Families of compounds considered in this study and colour codes associated in the 165 166 various figures.

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Abbreviation	Family of pollutants	Figure colour code
ALK	Alkylphenols	
FR	Flame Retardants	
М	Metals	
0	Others	
ORG	Organotins	
Р	Pesticides	
PAH	Polycyclic Aromatic Hydrocarbons and derivatives	
PBDE	Polybrominated Diphenyl Ethers	
PC	Pharmaceutical Compounds	
PCB	Polychlorobiphenyls	
PFAA	Perfluoroalkyl Acids	
PHT	Phtalates	
VOC	Volatile Organic Compounds	

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3.2 Ecotoxicity of the pollutants sought in UWWDs 170

Of the 207 compounds sought in the UWWD samples (total, dissolved and/or particle phases), 171 172 161 were obtained by calculation using data from the European database and the international literature.

In addition, we formulated ourselves 52 new PNEC values based on experimental data and 4 were 173

modelled using ECOSAR model, as detailed in Table 2. Finally, we listed a total of 165 PNEC values. 174

Figure 1 presents the distribution of available PNEC values for each pollutant family. This figure shows 175

that 5 families present a large number of molecules for which PNEC data are not available: PAH, PC, 176

ALK, O and PCB. This lack of data poses a problem for estimating UWWD compound hazards. 177

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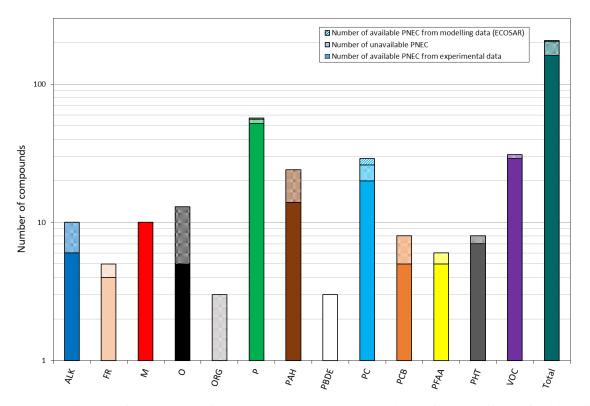


Figure 1: Distribution of the number of molecules and their PNEC available for the different families ofpollutants identified in UWWDs.

184 The 165 PNEC values obtained have been separated into three categories according to the 185 ranking of Orias and Perrodin (2014): a PNEC less than 0.1 μ g L⁻¹ (the most hazardous, the most 186 ecotoxic) (Figure 2), a PNEC between 1 and 0.1 μ g L⁻¹ (intermediate ecotoxicity) (Figure 3) and a PNEC 187 > 1 μ g L⁻¹ (less ecotoxic) (Figure 4). Finally, due to the results obtained, 59 of the compounds were 188 ranked as very ecotoxic, 30 as ecotoxic and 76 as slightly ecotoxic.

189 Among the 59 very ecotoxic compounds, pesticides feature predominantly in the most 190 hazardous category of pollutants sought in UWWDs. Indeed, 33 of the compounds with a PNEC less 191 than 0.1 μ g L⁻¹ are pesticides. Moreover, regarding the PNEC values of a very large number of pesticides sought in UWWDs (54 compounds), they represent a potentially considerable source of hazard for this 192 kind of matrix: 43 compounds have a PNEC value under 1.0 µg L⁻¹ and 33 have a PNEC value under 193 0.1 µg L⁻¹). Two other families represent a significant share of the most ecotoxic pollutants: PCs (7 194 195 compounds) and PAHs (also 7). The share of the other families is lower. The most ecotoxic compound is Endrin, a pesticide with a PNEC of 1.22.10⁻⁸ µg.L⁻¹. However, when examining slightly ecotoxic 196 compounds (PNEC >1 μ g.L⁻¹), VOCs are largely represented, with 24 of the compounds listed in this 197

category. The gap between the share of the other families is narrower: FR (3 compounds), PHT (6),
PFAA (3), PAH (5), P (11), O (5), M (6), PC (10), ALK (1). The least toxic compound sought in
UWWDs is Iohexol, a PC with a PNEC value of 45705 µg.L⁻¹.

201 The large number of very ecotoxic pollutants sought in UWWDs can be explained by the fact that many substances found in UWWDs were developed to eliminate certain families of organisms, thus 202 203 they have a considerable impact on many aquatic species belonging to the same family (the most relevant 204 example is pesticides). Nevertheless, a large number of pollutants of this family present low toxicity 205 regarding PNEC values. This observation may be explained by three possibilities: the pollutant specifically affects the organism targeted without having any effect on the others (the case of similar 206 pesticides), it is only slightly toxic for aquatic organisms, or there are not enough data to lower the 207 208 PNEC value.

We also note that no or insufficient ecotoxicity data are available for 13 of the compounds (3 PC, 6 PAH, 2 ALK, 1 ORG, 1 PFAA) detected in UWWDs. These compounds include a high number of PCs and PAHs. More ecotoxicity data are necessary for these families of compounds to avoid an underestimation of the hazard represented by UWWDs. Moreover, as it can be seen in Table 2, the PNECs of certain pollutants, such as PCBs, propylparaben and cotinine, were calculated using a small number of ecotoxicological data, and consequently their toxicity could be underestimated. However, we should bear in mind that these PNEC values were created with high safety factors (1000).

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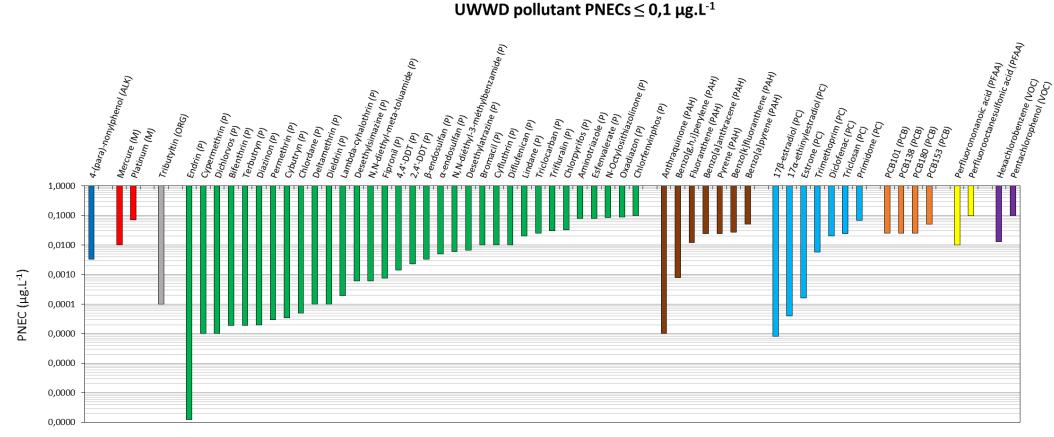
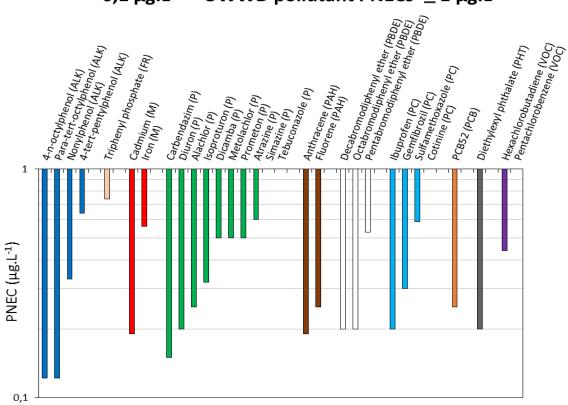


Figure 2: PNEC values lower or equal to 0.1 µg L⁻¹ for the different compounds sought in UWWDs. Letters in brackets correspond to the classification in





0,1 μ g.L⁻¹ < UWWD pollutant PNECs \leq 1 μ g.L⁻¹

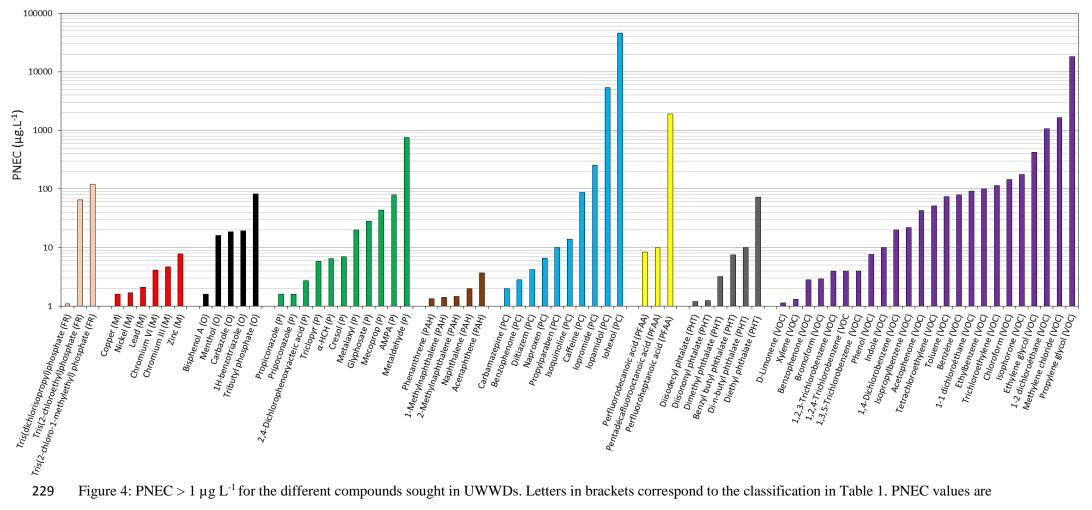
Figure 3: PNEC values between 0.1 µg L⁻¹ and 1 µg L⁻¹ for the different compounds sought in UWWDs. Letters in brackets correspond to the classification in

Table 1. PNEC values are ranked in ascending order into each separated pollutant family.

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UWWD pollutant PNECs > 1 μ g.L⁻¹



230 ranked in ascending order into each separated pollutant family.

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Table 2: Summary of ecotoxicity data and PNEC formulated by calculation or modelling for 60 pollutants sought in UWWDs. (Legend: D: Decomposers; P:

- 233 Primary producers; 1C: Primary consumers; 2C: Secondary consumers; C: ecotoxicity data available for organism community bioassays; Bc: Brachionus
- 234 *calyciflorus*; EF: Extrapolation factor).

					Exp	perim	ental eco	otoxic	ity			E	Data used for PNEC calculation	l				
N°CAS	Compound	Family	Test		Trop	hic le	evel		Chr	onic	test	Specie	Parameter	Endpoint	Value (ug/L)	EF	PNEC (ug/L)	Ref.
			species	D	Р	1C	2C	D	Р	1C	2C	Specie	Farameter	Endpoint	value (ug/L)			
104-40- 5	4-(para)-nonylphenol	ALK	77+C	0	5	19	36+17 (Bc)	0	5	6	11+17 (Bc)	Pimephales promelas	Reproduction	LOEC	3.31E-02	10	3.31E-03	35
80-46-6	4-tert-pentylphenol	ALK	4	0	0	1	3	0	0	1	2	Danio rerio	Development	LOEC	3.20E+01	50	6.40E-01	35
78-51-3	Tris(2- butoxyethyl)phosphate	FR	3	0	0	0	3	0	0	0	0	Danio rerio	Mortality	NOEC	4.00E+03			35
7439- 89-6	Iron	М	6	0	1	2	3	0	1	0	0	Cyprinus carpio	Mortality	LC50	5.60E+02	1000	5.60E-01	35
86-74-8	Carbazole	0	4	0	1	2	1	0	1	1	0	Pimephales promelas	Mortality	LC50	9.30E+02	50	1.86E+01	35
789-02- 6	2,4'-DDT	Р	4	0	0	0	4	0	0	0	2	Oryzias latipes	Reproduction (fertility)	LOEC	2.30E-01	100	2.30E-03	35
50-29-3	4,4'-DDT	Р	40	0	11	21	8	0	11	5	0	Hyallela azteca	Mortality	LC50	7.00E-02	50	1.40E-03	35
309-00- 2	Aldrin	Р	55	0	0	18	37	0	0	0	0	Paratelphusa jacquemontii	Mortality	LC50	9.70E-02			35
218-01- 9	Chrysene	Р	3	0	0	2	1	0	0	0	0	Aedes aegypti	Mortality	LC50	1.70E+03			35
1319- 77-3	Cresol	Р	6	0	1	2	3	0	1	0	0	Gammarus fasciatus	Immobilization	EC50	7.00E+03	1000	7.00E+00	35
52315- 07-8	Cypermethrin	Р	68	0	6	28	33+1 (Bc)	0	6	7	2+1 (Bc)	Palaemonetes argentinus	Mortality	LOEC	1.00E-04	10	1.00E-05	35
6190- 65-4	Desethylatrazine	Р	11	0	8	3	0	0	7	1	0	Stagnicola elodes	Mortality	NOEC	3.30E-01	50	6.60E-03	35
1007- 28-9	Desethylsimazine	Р	10	0	7	2	1	0	7	0	0	Pseudokirchneriella subcapitata	Growth	NOEC	6.00E-01	1000	6.00E-04	35
333-41- 5	Diazinon	Р	131	0	9	53	69	0	9	8	18	Daphnia magna	Reproduction	EC50	2.00E-04	10	2.00E-05	35
60-57-1	Dieldrin	Р	98+C	0	7	43	45+1 (Bc)	0	7	2	1+1 (Bc)	Brachionus calyciflorus	Population growth	LOEC	1.00E-03	10	1.00E-04	35
72-20-8	Endrin	Р	89	0	1	22	66	0	1	0	1	Asellus aquaticus	Mortality	LC50	6.10E-07	50	1.22E-08	35

2682- 20-4	Methylisothiazolinone	Р	3	0	0	1	2	0	0	0	0	Oncorhynchus mykiss	Mortality	LC50	7.00E+01			35
51218- 45-2	Metolachlor	Р	63	0	37	12	14	0	35	6	5	Stuckenia pectinata	Photosynthesis disturbance	LOEC	5.00E+00	10	5.00E-01	35
134-62- 3	N,N-diéthyl-3- méthylbenzamide	Р	6	0	0	3	3	0	0	1	0	Pimephales promelas	Growth	NOEC	6.00E-01	100	6.00E-03	35
134-62- 3	N,N-diethyl-meta- toluamide	Р	6	0	0	3	3	0	0	0	0	Pimephales promelas	Growth (condition index)	NOEC	6.00E-01	1000	6.00E-04	35
26530- 20-1	N-Octylisothiazolinone	Р	6	0	0	1	5	0	0	0	1	Pimephales promelas	Growth	NOEC	8.50E+00	100	8.50E-02	35
52645- 53-1	Permethrine	Р	4	0	1	2	1	0	0	0	0	Daphnid	Mortality	LC50	3.00E-02	1000	3.00E-05	38
1610- 18-0	Prometon	Р	15	0	5	1	9	0	5	1	1	Chlorella fusca var. vacuolata	Population growth (growth rate)	EC01	5.00E+00	10	5.00E-01	35
886-50- 0	Terbutryn	Р	21	0	9	2	10	0	6	0	0	Craticula accomoda	Growth	EC10	1.90E-02	1000	1.90E-05	35
101-20- 2	Triclocarban	Р	11	0	1	5	5	0	1	2	1	Daphnia magna	Reproduction	NOEC	2.50E-01	10	2.50E-02	35
319-84- 6	α-НСН	Р	21	0	1	8	12	0	1	2	4	Lymnaea stagnalis	Reproduction	EC50	6.50E+01	10	6.50E+00	35
33213- 65-9	β-endosulfan	Р	15	1	0	6	8	0	0	0	0	Oncorhynchus mykiss	Mortality	LC50	3.30E+00	1000	3.30E-03	35
319-85- 7	β-НСН	Р	4	0	0	0	4	0	0	0	0	Oryzias latipes	Behavioural changes (general)	NOEC	3.20E+01			35
90-12-0	1-Methylnaphthalene	РАН	6	0	3	1	2	0	1	0	0	Daphnia magna	Immobilization	EC50	1.42E+03	1000	1.42E+00	35
91-57-6	2-Methylnaphthalene	РАН	5	0	2	1	2	0	0	0	0	Oncorhynchus mykiss	Mortality	EC50	1.46E+03	1000	1.46E+00	35
84-65-1	Anthraquinone	PAH	6	0	2	2	2	0	2	0	0	Lemna gibba	Photosynthesis	IC50	1.00E-02	1000	1.00E-05	35
132-65- 0	Dibenzothiophene	PAH	3	0	0	1	2	0	0	0	0	Daphnia magna	Mortality	LC50	4.20E+02			35
119-61- 9	Benzophenone	PC	5	0	0	4	1	0	0	0	1	Daphnia magna	Immobilization	EC50	2.80E+02	100	2.80E+00	35
486-56- 6	Cotinine	PC	1	0	1	0	0	0	1	0	0	Lemna gibba	Biomass	NOEC	1.00E+03	1000	1.00E+00	35
42399- 41-7	Diltiazem	PC	3	0	1	1	1	0	0	0	0	Paspalum vaginatum	Calcium uptake	LOEC	4.16E+03	1000	4.16E+00	35
58-73-1	Diphenhydramine	PC	2	0	0	1	1	0	0	1	0	Ceriodaphnia dubia	Reproduction	LC50	3.94E+03	100	3.94E+01	35
119-65- 3	Isoquinoline	PC	5	0	1	2	2	0	1	0	0	Poecilia reticulata	Mortality	LC50	1.40E+04	1000	1.40E+01	35
94-13-3	Propylparaben	PC	2	0	0	2	0	0	0	1	0	Daphnia magna	Immobilization	NOEC	1.00E+03	100	1.00E+01	35

37680- 73-2	PCB101	PCB	4	0	0	3	1	0	0	0	1	Pimephales promelas	Growth	LOEC	2.50E+00	100	2.50E-02	35
35065- 28-2	PCB138	PCB	2	0	0	0	2	0	0	0	1	Pimephales promelas	Growth	LOEC	2.50E+00	100	2.50E-02	35
35065- 27-1	PCB153	PCB	4	0	0	1	3	0	0	1	1	Pimephales promelas	Growth	LOEC	2.50E+00	50	5.00E-02	35
35065- 29-3	PCB180	PCB	1	0	0	0	1	0	0	0	1	Pimephales promelas	Growth	LOEC	2.50E+00	100	2.50E-02	35
35693- 99-3	PCB52	PCB	2	0	0	1	1	0	0	0	1	Pimephales promelas	Reproduction	NOEC	2.50E+01	100	2.50E-01	35
335-67- 1	Pentadécafluorooctanoic acid	PFAA	12	0	2	3	6 + 1 (Bc)	0	2	2	2	Oryzias latipes	Mortality	LOEC	1.00E+02	10	1.00E+01	35
335-76- 2	Perfluorodecanoic acid	PFAA	3	0	1	1	1	0	1	0	0	Danio rerio	Mortality	LC50	8.40E+03	1000	8.40E+00	35
375-85- 9	Perfluoroheptanoic acid	PFAA	1	0	1	0	0	0	1	0	0	Chlorella vulgaris	Growth	EC50	1.90E+06	1000	1.90E+03	35
375-95- 1	Perfluorononanoic acid	PFAA	4	0	2	1	1	0	2	0	0	Danio rerio	Growth	LOEC	1.00E+01	1000	1.00E-02	35
1763- 23-1	Perfluorooctanesulfonic acid	PFAA	4	0	0	2	2	0	0	1	0	Oryzias latipes	Mortality	LOEC	1.00E+01	100	1.00E-01	35
2058- 94-8	Perfluoroundecanoic acid	PFAA	2	0	0	1	1	0	0	0	0	Daphnia magna	Immobilization	NOEC	5.64E+04			35
117-81- 7	Diethylexyl phthalate	PHT	29	0	3	5	21	0	3	2	3	Danio rerio	Reproduction	NOEC	2.00E+00	10	2.00E-01	35
26761- 40-0	Diisodecyl phtalate	PHT	7	0	1	1	5	0	1	1	0	Daphnia magna	Mortality	LOEC	6.00E+01	50	1.20E+00	35
27554- 26-3	Diisononyl phtalate	PHT	7	0	1	2	4	0	1	1	0	Daphnia magna	Reproduction	NOEC	6.20E+01	50	1.24E+00	35
131-11- 3	Dimethyl phthalate	PHT	9	0	1	2	6	0	1	0	0	Cyprinodon variegatus	Mortality	NOEC	3.20E+03	1000	3.20E+00	35
98-86-2	Acetophenone	VOC	4	0	1	2	1	0	0	0	0	Tetrahymena pyriformis	Population Growth	EC50	4.28E+04	1000	4.28E+01	35
119-61- 9	Benzophenone	VOC	5	0	0	4	1	0	0	1	0	Daphnia magna	Immobilization	EC50	2.80E+02	100	2.80E+00	35
75-25-2	Bromoform	VOC	5	0	1	2	2	0	1	0	0	Cyprinodon variegatus	Mortality	NOEC	2.90E+03	1000	2.90E+00	35
138-86- 3	D-Limonene	VOC	11	0	1	7	3	0	1	0	0	Lymnaea acuminata	Mortality	LC50	1.13E+03	1000	1.13E+00	35
107-21- 1	Ethylene glycol	VOC	20	1	1	7	11	0	1	1	1	Ceriodaphnia dubia	Reproduction	NOEC	4.20E+03	10	4.20E+02	35
120-72- 9	Indole	VOC	3	0	1	2	0	0	1	1	0	Oryzias latipes	Mortality	LC50	1.00E+03	100	1.00E+01	35
124-76- 5	Isoborneol	VOC	1	0	0	0	1	0	0	0	0	Fish	Mortality	LC50	5.90E+04			38

235

236 <u>3.3 Concentration of pollutants in separate and combined systems</u>

Maximum pollutant concentrations found in UWWDs were assessed using available literature and are presented in Tables 3 and 4. MECmax were very variable, with some pollutants (e.g. hexachlorobenzene) found at several ng.L⁻¹, while others (e.g. copper) were found at a few mg.L⁻¹. However, the compound concentrations generally ranged between ng.L⁻¹ and several hundred μ g.L⁻¹.

On one hand, as observed in Table 3, 61 compounds were detected in stormwater samples, and 50 were sought but not detected. We found the highest concentrations for zinc (2.12 mg.L⁻¹). High concentrations of pollutants such as other heavy metals (copper, nickel) and pesticides (such as glyphosate and AMPA) and PAH (e.g. pyrene) were also detected. The main families of pollutants detected at high concentrations were PAH, VOC, ALK, PHT, P and M.

246 On the other hand, out of the 81 chemicals of interest sought in the dissolved part of the CSO 247 samples, 49 were reported (see Table 4). The highest concentration was obtained for zinc (236 μ g.L⁻¹). The main families found at high concentrations were M, P and PC. The occurrence of high 248 concentrations of heavy metals in separate systems and CSO is not surprising due to the source of 249 wastewater and rainfall runoff. More precisely, tyre wear, vehicle brake deposition, conveyance system 250 251 infrastructure (e.g. downspouts and culverts), and atmospheric deposition can lead to the presence of heavy metals.⁴ Moreover, the remobilization of wastewater and in-sewer deposits (due to pollutant 252 water/particle partition) can represent a large share of the pollutant load in CSO, such as Cu and PAH.⁵⁸ 253 Launay et al.⁶⁴ observed the same phenomenon for pharmaceutical compounds and industrial chemicals. 254 255 Contrary to stormwater, we note the presence of PC compounds sought and detected in the dissolved 256 phase (18 compounds), probably due to the contribution of urban wastewater to the pollution of CSO.

If we compare the Maximum concentrations of UWWD pollutants with other urban effluents
(e.g. urban wastewater and hospital wastewater), such as phthalates, heavy metals, pharmaceutical
compounds and pesticides, MEC_{max} were close to those detected in these other matrices.^{60,61,62,63} Some
of them, such as Bisphenol A, concentrations in UWWDs (max. 130 µg L⁻¹) can be even higher than in
wastewater (max. 2.14 µg L⁻¹). This fact was also observed with heavy metals such as copper and zinc.
Consequently, since they undergo hardly any or no treatment, and that most of few current treatment
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263 methods employed do not reduce concentration of many identified pollutants, considerable amounts of264 these substances may enter surface waters.

Nevertheless, it is important to point out that the data collected and discussed above have some limitations. Firstly, few studies have analyzed the dissolved phase of discharges and thus many compounds identified and sought in UWWDs are not taken into account here because they are found in only the particle or total phase (due to their low solubility and high affinity for the particle phase). Consequently, the number of compounds detected here is limited (see Table 3 and Table 4). Secondly, concentrations found in UWWDs need to be considered with caution because, as stated in the study of Huber et al.,⁵² concentrations measured in stormwater samples can be influenced by different parameters such as sample collection, preparation and analysis, which can also affect the concentrations observed in discharges.

286	Table 3: Maximum Measured Effluent Concentrations (MECmax) of micropollutants in the dissolved part sought in stormwater samples (ND: Not detected, *:
287	Unspecified).

Pollutant family	Pollutant	MECmax (µg/L)	Country	Ref.	Pollutant family	Pollutant	MECmax (µg/L)	Country	Ref.
ALK	4-n-octylphenol (ALK)	ND	France	4	PAH	Dibenzothiophene (PAH)	0.0417	USA (Washington D.C.)	2
ALK	4-nonylphenol di- ethoxylate (ALK)	ND	France	4	PAH	Acenaphthylene (PAH)	0.070	Australia; France; Sweden; USA (Alabama); USA (Washington D.C.)	2;4;5 ;47;48
ALK	4-tert-pentylphenol (ALK)	0.7	Sweden	5	PAH	Perylène (PAH)	0.12	USA (Washington D.C.)	2
ALK	Para-tert- octylphenol (ALK) 4-Octylphenol	0.99	France, Sweden	4;5	РАН	Dibenzo[a,h]anthracene (PAH)	0.13	France; Sweden; USA (Alabama); USA (Washington D.C.) Australia; France; Sweden; USA	2;4; 5;47 2;4;
ALK	monoethoxylate (ALK)	4.39	Sweden	5	PAH	Benzo[a]anthracene (PAH)	0.34	(Alabama); USA (South Carolina); USA (Washington D.C.)	2 , 4 , 5 ; 47 ; 48 ; 49
ALK	4-(para)- nonylphenol (ALK)	7.7	France, Sweden	4;5	PAH	Benzo[k]fluoranthene (PAH)	0.40	France; Sweden; USA (Alabama); USA (Washington D.C.)	2;4; 5;47
ALK	4-tert-butylphenol (ALK)	7.82	France, Sweden	4;5	PAH	Benzo[e]pyrene (PAH)	0.42	USA (South Carolina); USA (Washington D.C.)	2;48
ALK	4-nonylphenol mono-ethoxylate (ALK)	8.36	Sweden	5	РАН	Indeno[1,2,3-cd]pyrene (PAH)	505	Australia; France; Sweden; USA (Alabama); USA (South Carolina); USA (Washington D.C.)	2;4; 5;47; 48;49
М	Platinium (M)	ND	France	4	PAH	Benzo[a]pyrene (PAH)	0.54	Australia; France; Sweden; USA (Alabama); USA (South Carolina); USA (Washington D.C.)	2;4; 5;47; 48;49
М	Mercury (M)	0.01	Canada; France	4;41	РАН	Benzo[g,h,i]perylene (PAH)	0.55	Australia; Sweden; USA (Alabama); USA (South Carolina); USA (Washington D.C.)	2;5; 47;48; 49
М	Lead (M)	15.9	Australia; China; France; USA (Alabama); USA (California); *	4;5;47; 50;51;52	РАН	Anthracene (PAH)	0.56	Australia; France; Sweden; USA (Alabama); USA (South Carolina); USA (Washington D.C.)	2;4; 5;47; 48;49
М	Cadmium (M)	31.1	Australia; France; USA (Alabama); USA (California); *	4;47;50; 52;55	РАН	Chrysene (PAH)	0.62	Australia; France; Sweden; USA (Alabama); USA (South Carolina); USA (Washington D.C.)	2;4; 5;47; 48;49
М	Chromium (M)	32	Australia; France; USA (Alabama); USA (California); *	4;47;50; 52;55	РАН	Acenaphthene (PAH)	0.72	Australia; France; Sweden; USA (Alabama) USA (Washington D.C.)	4;5; 47;49
М	Nickel (M)	49	France; USA (Alabama); USA (California); *	4;5;47; 52	РАН	Benzo[b]fluoranthene (PAH)	0.98	Australia; France; USA (Alabama); USA (South Carolina); USA (Washington D.C.)	2;4; 47;48; 49

			Australia; China; France;	4;5;47;				Australia; France; Sweden; USA	2;4;
М	Copper (M)	1520	USA (Alabama); USA	50;51;52	PAH	Phenanthrene (PAH)	1.32	(Alabama); USA (South Carolina);	5;47;
			(California); * Australia; China; France;	,,.				USA (Washington D.C.) Australia; France; Sweden; USA	48;49
М	Zinc (M)	2118	USA (Alabama); USA	4;5;47;	PAH	Fluorene (PAH)	1.37	(Alabama); USA (South Carolina);	2;4; 5;47;
111		2110	(California); *	50;51;52	1 All	Photene (I AII)	1.57	USA (Washington D.C.)	48;49
			(Curronnu),					Australia; France; Sweden; USA	2;4;
0	Bisphenol A (O)	130	Sweden	5	PAH	Fluoranthene (PAH)	14	(Alabama); USA (South Carolina);	5;47;
	•							USA (Washington D.C.)	48;49
ORG	Dibutyltin (ORG)	ND	France	4	PAH	Naphthalene (PAH)	17	Australia; France; Sweden; USA	2;4;5
OKO	Dibutyitiii (OKO)	ND	Talee	4	1 All	Naphulaiene (1 AII)	17	(Alabama); USA (Washington D.C.)	; 47 ; 49
0.0.0					DAI		•	Australia; France; Sweden; USA	2;4;
ORG	Tributyltin (ORG)	ND	France	4	PAH	Pyrene (PAH)	20	(Alabama); USA (South Carolina);	5;47;
	Monobutyltin					Decabromodiphenyl ether		USA (Washington D.C.)	48;49
ORG	(ORG)	0.025	France	4	PBDE	(PBDE)	ND	France	4
	~ /		-			Octabromodiphenyl ether		-	
Р	2,4'-DDT (P)	ND	France	4	PBDE	(PBDE)	ND	France	4
Р	4,4'-DDT (P)	ND	France	4	PBDE	Pentabromodiphenyl ether	ND	France	4
	, , , , ,			4		(PBDE)			4
Р	Alachlor (P)	ND	France	4	PCB	PCB101 (PCB)	ND	France	4
Р	Aldrin (P)	ND	France; USA (Alabama)	4;47	PCB	PCB118 (PCB)	ND	France	4
P P	Atrazine (P) Chlordane (P)	ND ND	France France; USA (Alabama)	4 4;47	PCB PCB	PCB138 (PCB) PCB153 (PCB)	ND ND	France France	4 4
	Chlorfenvinphos		France; USA (Alaballia)	4;47	-			France	4
Р	(P)	ND	France	4	PCB	PCB180 (PCB)	ND	France	4
Р	Chlorpyrifos (P)	ND	France	4	PCB	PCB194 (PCB)	ND	France	4
Р	Desethylsimazine (P)	ND	France	4	PCB	PCB28 (PCB)	ND	France	4
Р	Dieldrin (P)	ND	France	4	PCB	PCB52 (PCB)	ND	France	4
Р	Endrin (P)	ND	France; USA (Alabama)	4;47	PFAA	Perfluoroheptanoic acid	0.011	Japan; Singapour; Switzerland; USA	42; 43;
1		ND	Trance, USA (Alabalila)	+,+/	IIAA	(PFAA)	0.011	(Minnesota)	44;45
Р	Esfenvalerate (P)	ND	France; USA (Alabama)	4;47	PFAA	Perfluoroundecanoic acid	0.045	Japan; Singapour; Switzerland; USA	42;43;
P	Estenvalerate (P)	ND	France; USA (Alaballia)	4;47	ГГАА	(PFAA)	0.045	(Minnesota)	44 ; 45 ; 46
									40 42;43;
Р	Lindane (P)	ND	France	4	PFAA	Perfluorodecanoic acid	0.077	Japan; Singapour; Switzerland; USA	42; 45; 44; 45;
	(_)					(PFAA)		(Minnesota)	46
						Perfluorononanoic acid		Japan; Singapour; Switzerland; USA	42;43;
Р	Simazine (P)	ND	France	4	PFAA	(PFAA)	0.093	(Minnesota)	44;45;
						(11/1/1)		(Winnesota)	46

Р	Trifluralin (P)	ND	France	4	PFAA	Perfluorooctanesulfonic acid (PFAA)	0.16	Japan; Singapour; Switzerland; USA (Minnesota)	42;43; 44;45; 46
Р	α-endosulfan (P)	ND	France	4	PFAA	Pentadecafluorooctanoic acid (PFAA)	0.17	Japan; Singapour; Switzerland; USA (Minnesota)	42;43; 44;45; 46
Р	β-endosulfan (P)	ND	France	4	PHT	Dimethyl phthalate (PHT)	ND	USA (Alabama)	47
Р	Lambda- cyhalothrin (P)	0.007	USA (California)	40	PHT	Di-n-octyl phtalate (PHT)	ND	France	4
Р	Cyfluthrin (P)	0.023	USA (California)	40	PHT	Di-n-butyl phthalate (PHT)	1	Sweden; USA (Alabama)	5;47
Р	Cypermethrin (P)	0.026	USA (California)	40	PHT	Diisodecyl phtalate (PHT)	1.2	Sweden	5
Р	Bifenthrin (P)	0.034	USA (California)	40	PHT	Diisononyl phtalate (PHT)	2	Sweden	5
Р	Metaldehyde (P)	0.05	France	4	PHT	Diethyl phthalate (PHT)	15.9	Sweden; USA (Alabama)	5;47
Р	Permethrin (P)	0.13	USA (California)	40	PHT	Benzyl butyl phthalate (PHT)	16	Sweden; USA (Alabama)	5;47
Р	Deltamethrin (P)	0.25	USA (California)	40	PHT	Diethylhexyl phthalate (PHT)	45	France; Sweden; USA (Alabama)	4;5; 47
Р	Aminotriazole (P)	0.5	France	4;47;48	VOC	1,2,3-Trichlorobenzene (VOC)	ND	France	4
Р	Isoproturon (P)	0.6	France	4	VOC	1,2,4-Trichlorobenzene (VOC)	ND	France	4
Р	Diuron (P)	0.64	France	4	VOC	1,3,5- Trichlorobenzene (VOC)	ND	France	4
Р	AMPA (P)	19	France; Switzerland	4;53;54	VOC	1-2 dichloroéthane (VOC)	ND	France	4
Р	Glyphosate (P)	150	France; Switzerland	4;53;54	VOC	Benzène (VOC)	ND	France	4
	21			<i>, ,</i>	VOC	Chloroform (VOC)	ND	France	4
					VOC	Ethylbenzene (VOC)	ND	France	4
					VOC	Hexachlorobenzene (VOC)	ND	France	4
					VOC	Hexachlorobutadiene (VOC)	ND	France	4
					VOC	Isopropylbenzene (VOC)	ND	France	4
					VOC	Pentachlorobenzene (VOC)	ND	France	4
					VOC	Tetrachloroethylene (VOC)	ND	France	4
					VOC	Toluene (VOC)	ND	France	4
					VOC	Trichloroethylene (VOC)	ND	France	4
					VOC	Xylene (VOC)	ND	France	4
					VOC	Pentachlorophenol (VOC)	0.2	France	4
					VOC	Dichloromethane (VOC)	13	France	4
					VOC	1,4-Dichlorobenzene (VOC)	47	USA (Alabama)	48

290	Table 4: Maximum Measured Effluent Concentrations (MECmax) of the micropollutants sought in the dissolved part of CSO samples (ND: Not detected, *:
291	Unspecified).

Pollutant family	Pollutant	MECmax (µg/L)	Country	Ref.	Pollutant family	Pollutant	MECmax (µg/L)	Country	Ref.
ALK	4-n-octylphenol (ALK)	ND	France	57	PC	Estrone (PC)	ND	South Korea	12
ALK	Para-tert-octylphenol (ALK)	ND	France	57	PC	Diltiazem (PC)	ND	South Korea	12
ALK	4-tert-butylphenol (ALK)	0.080	France	57	PC	Trimethoprim (PC)	ND	South Korea	12
ALK	Nonylphenol (ALK)	0.390	France	57	PC	Meprobamate (PC)	ND	South Korea	12
FR	Tris(2-chloro-1-methylethyl) phosphate (FR)	1.140	South Korea	12	РС	Diclofenac (PC)	ND	South Korea	12
М	Nickel (M)	1.1	France	15;57;58	PC	Gemfibrozil (PC)	0.006	South Korea	12
М	Chromium (M)	1.8	France	15;57;58	PC	Diphenhydramine (PC)	0.023	South Korea	12
М	Cadmium (M)	9.9	France, United Kingdom	15 ; 57 ; 58 ; 59	РС	Primidone (PC)	0.040	South Korea	12
М	Copper (M)	65.4	France, United Kingdom	15 ; 57 ; 58 ; 59	РС	Triclosan (PC)	0.055	South Korea	12
М	Lead (M)	143.4	France, United Kingdom	15;57;58; 59	PC	Carbamazepine (PC)	0.060	South Korea	12
М	Zinc (M)	236.1	France, United Kingdom	15 ; 57 ; 58 ; 59	PC	Propylparaben (PC)	0.069	South Korea	12
О	1H-benzotriazole (O)	0.078	South Korea	12	PC	Sulfamethoxazole (PC)	0.070	South Korea	12
ORG	Dibutyltin (ORG)	ND	France	57	PC	Trimethoprim (PC)	0.073	South Korea	12
ORG	Tributyltin (ORG)	ND	France	57	PC	Benzophenone (PC)	0.136	South Korea	12
ORG	Monobutyltin (ORG)	0.013	France	57	PC	Atenolol (PC)	0.363	South Korea	12
Р	Aldrin (P)	ND	France	57	PC	Acesulfame (PC)	0.727	South Korea	12
Р	Dieldrin (P)	ND	France	57	PC	Naproxen (PC)	0.750	South Korea	12
Р	Desethylatrazine (P)	0.017	France	57	PC	Ibuprofen (PC)	0.773	South Korea	12
Р	Atrazine (P)	0.030	France	57	PC	Sucralose (PC)	0.818	South Korea	12
Р	N,N-diéthyl-3-méthylbenzamide (P)	0.035	South Korea	12	PC	Iopamidol (PC)	1.550	South Korea	12
Р	Isoproturon (P)	0.040	France	57	PC	Iohexol (PC)	1.770	South Korea	12
Р	Triclocarban (P)	0.061	South Korea	12	PC	Caffeine (PC)	2,54	South Korea	12

Р	Aminotriazole (P)	0.460	France	57	PC	Iopromide (PC)	2,66	South Korea	12
Р	Diuron (P)	0.470	France	57	PCB	PCB101 (PCB)	ND	France	57
Р	AMPA (P)	1.600	France; Switzerland	53;57	PCB	PCB118 (PCB)	ND	France	57
Р	Glyphosate (P)	3.4	France; Switzerland	53;57	PCB	PCB138 (PCB)	ND	France	57
PAH	Acenaphthene (PAH)	ND	France	15;57	PCB	PCB153 (PCB)	ND	France	57
PAH	Benzo[a]anthracene (PAH)	ND	France	15;57	PCB	PCB180 (PCB)	ND	France	57
PAH	Benzo[b]fluoranthene (PAH)	ND	France	15;57	PCB	PCB194 (PCB)	ND	France	57
PAH	Benzo[g,h,i]perylene (PAH)	ND	France	15;57	PCB	PCB28 (PCB)	ND	France	57
PAH	Benzo[k]fluoranthene (PAH)	ND	France	15;57	PCB	PCB52 (PCB)	ND	France	57
PAH	Chrysene (PAH)	ND	France	15;57	PHT	Diethylhexyl phthalate (PHT)	ND	France	57
PAH	Dibenzo[a,h]anthracene (PAH)	ND	France	15;57	VOC	Ethylbenzene (VOC)	ND	France	57
PAH	Benzo[a]pyrene (PAH)	0.005	France	15;57	VOC	Tetrachloroethylene (VOC)	ND	France	57
PAH	Benzo[e]pyrene (PAH)	0.006	France	15;57	VOC	Toluene (VOC)	ND	France	57
PAH	Indeno[1,2,3-cd]pyrene (PAH)	0.008	France	15;57	VOC	Trichloroethylene (VOC)	ND	France	57
PAH	Anthracene (PAH)	0.009	France	15;57	VOC	Xylene (VOC)	ND	France	57
РАН	Acenaphthylene (PAH)	0.020	France	15;57	VOC	Hexachlorobenzene (VOC)	0.002	USA (New York)	56
PAH	Pyrene (PAH)	0.021	France	15;57					
PAH	Fluoranthene (PAH)	0.025	France	15;57					
PAH	Fluorene (PAH)	0.027	France	15;57					
PAH	Phenanthrene (PAH)	0.085	France	15;57					
PAH	Naphthalene (PAH)	0.180	France	15;57					

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294 <u>3.4 Hazard of pollutants in separate and combined systems</u>

Using ecotoxicological (PNEC) and chemical (MEC_{max}) results previously enunciated we were able to calculate a very wide range of HQs in discharges (see Figures 5 and 6). The high dispersion of HQ values calculated could be explained by the heterogeneity of PNEC values observed in Figures 2 to 4 and the heterogeneity of pollutant concentrations due to the high variability of anthropic activities, and thus that of the catchment leached by rain, and the variability of rainfall events (length, height, flow rate, etc.). Due to the relatively different nature and composition of stormwater and CSO, pollutant hazards associated to these two matrices are discussed separately in the following sections.

302 <u>3.4.1 Hazard quotients in separate systems</u>

Concerning stormwater samples, only 51 hazard quotients were calculated due to the low 303 304 number of compounds found in the dissolved part (Figure 5). We found 14 compounds with an HQ lower than 1. The lowest hazard quotient was calculated for Metaldehyde (P) with an HQ of 6.7.10⁻⁵. 305 306 On the contrary, the majority of the micropollutants detected in stormwater samples had an HQ as high as 1 and are therefore potentially hazardous for aquatic ecosystems. Among these 37 compounds, 10 307 (27.1%) were pesticides, and 9 PAHs (24.32%). It is also noteworthy that 4 of the 6 most hazardous 308 309 compounds (HQ >1000) were pesticides. Among these very hazardous compounds, we also observed 310 the presence of one ALK (4-p-nonylphenol) and one PAH (fluoranthene). The most hazardous compound found was permethrin, with a hazard quotient of 4166.6. The important hazard from 311 pesticides (pyrethroids) linked to high HQ could be explained by the high ecotoxicity of these pesticides 312 rather than by the concentrations found in discharges (around 100 ng.L⁻¹). As pointed out before, most 313 314 of these pesticides (such as bifenthrin, cypermethrin) have a very low PNEC value, lower than 0.1 ng.L⁻ 1. 315

316 <u>3.4.2 Hazard quotients in combined systems</u>

Regarding CSO, only 43 HQs were calculated, also due to the low number of molecules
measured in dissolved phase in CSO (Figure 6). The HQs obtained were lower than for separate systems.

Indeed, the highest HQ was found for lead, with a value of 68.29. Contrary to separate systems, the 319 320 highest hazards were due to heavy metals, pesticides and pharmaceutical compounds. Only five families 321 were found in this ranking (HQ >1): PC (3 compounds), M (4), PAH (1), P (5) and ALK (1), and were 322 potentially hazardous for aquatic ecosystems. Results highlighted for heavy metals can be explained by the fact that certain heavy metals in these effluents were in high concentrations (e.g. $2.1 \mu g.L^{-1}$ for lead 323 and 1.6 μ g.L⁻¹ for copper), and not because of their low ecotoxicity, as shown by their relatively high 324 325 PNEC values. The presence of hazardous PC in CSO can be easily explained by the discharge of large 326 amounts of these often very ecotoxic molecules coming from urban wastewater in stormwater, without any treatment. Comparing hazard quotients from CSO samples to those of separate systems, we can 327 observe that calculated HQs are often lower (e.g. 65.4 in CSO against 950 in stormwater for copper). 328 329 We can simply explain these results by a dilution of stormwater pollutants by wastewater in CSO.

330 3.4.3 The case of pharmaceutical compounds in CSO studies

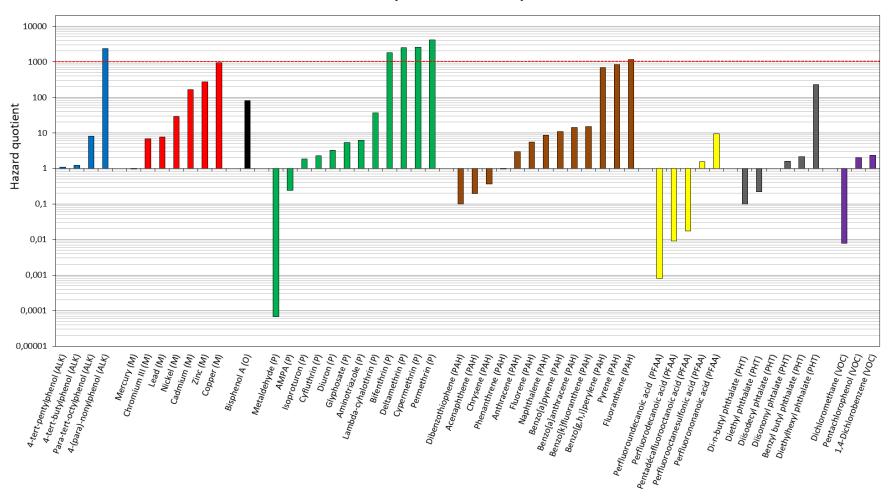
331 With the hazardous pharmaceutical compounds that have been highlighted above, 15 HQs have been calculated only and presented in Figure 6. As urban wastewater is mixed with stormwater in 332 combined sewers, all the pharmaceutical compounds classically consumed in cities could be found in 333 this type of discharge. However, few of them (28) have been taken into account to date in scientific 334 literature. Consequently, our study and all the other ecotoxicological studies dealing with CSO contain 335 336 a bias attributable to the fact that the potentially considerable hazard represented by these types of molecules is not taken into account. Indeed, as shown by Orias et al.³⁶, many of them can have very low 337 338 PNEC values (e.g. 17β-estradiol: 8 pg.L⁻¹, clotrimazole: 0.5 pg.L⁻¹), and/or are widely consumed (e.g. 339 ibuprofen and salicylic acid).

340

3.4.4 The case of molecules (e.g., pesticides) forbidden by European legislation

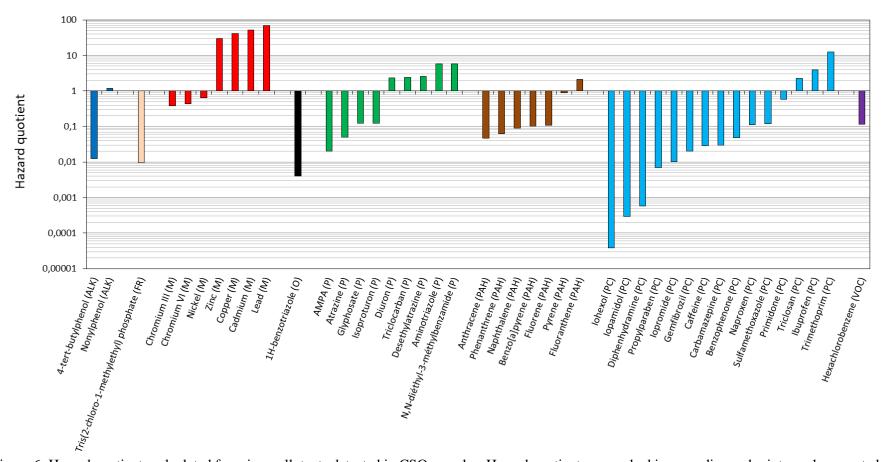
Some of the pesticides involved in the hazard of UWWDs (found in both stormwater and CSO, 341 such as diuron) were forbidden in 2008 by European legislation. However, several European 342 343 studies^{65,66,67} have detected the presence of these pesticides (with concentrations reaching 0.64 μ g L⁻¹ in discharges). This observation can be notably explained by the fact that farmers have stored pesticides 344 How to cite this paper: Gosset, A., Durrieu, C., Orias, F., Bayard, R., Perrodin, Y. (2017) Identification and assessment of ecotoxicological hazards attributable to pollutants in urban wet weather discharges. Environmental Science: Processes & Impacts DOI: 10.1039/C7EM00159B

- 345 purchased in the past and that they continue to use. Moreover, some of them linger on in the environment
- for long periods (due to their high half-life times) and their release can continue due to rainfall events.
- 347 Consequently, it is true to say that the hazard represented by these pesticides continues to exist even if
- they are no longer on the market.



Stormwater pollutant hazard quotients

- 349 Figure 5: Hazard quotients calculated for micropollutants detected in stormwater samples (the red line delimits the compounds with an HQ up to 1000,
- 350 considered very hazardous). Hazard quotients are ranked in ascending order into each separated pollutant family.



CSO pollutant hazard quotients

- 351 Figure 6: Hazard quotients calculated for micropollutants detected in CSO samples. Hazard quotients are ranked in ascending order into each separated
- 352 pollutant family.

353

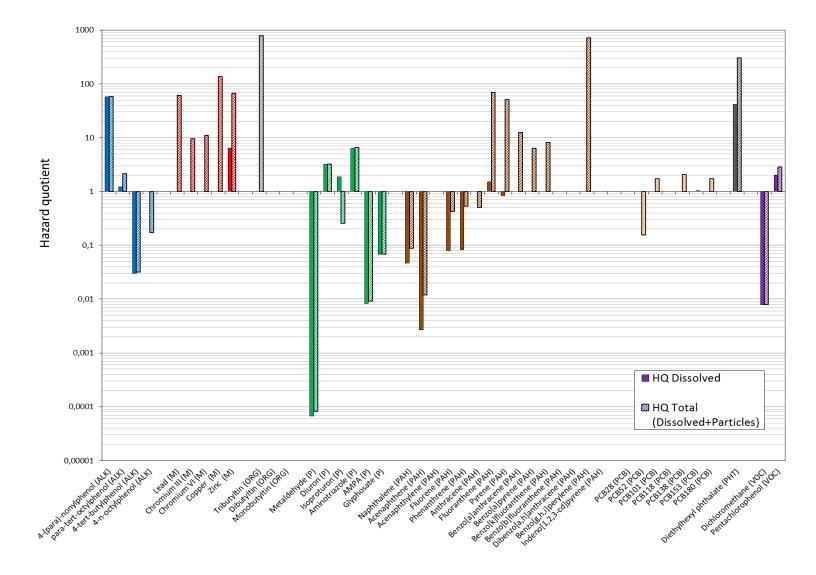
354 <u>3.5 Discussion of the method and data used</u>

The method presented in this study proved to be relevant to detect UWWD pollutant hazards. It 355 could be applied for many other matrices, and included in many risk assessment scenarios or chemical 356 357 prioritization programs. Nevertheless, our method and results are subject to limitations that need to be developed. The first one is the non-consideration of certain parameters, such as the bioconcentration, 358 bioaccumulation³⁰ and biomagnification of pollutants in the aquatic food web, which could amplify 359 considerably the hazard of UWWDs pollutants. Indeed, these phenomena have already been 360 361 demonstrated for many families of pollutants, such as the most hazardous found in UWWDs: heavy metals⁶⁸, pesticides^{69,70} and pharmaceutical compounds.^{71,72,73,74} Other parameters such as pH, 362 temperature, water chemistry (pH, hardness) can also modify pollutant toxicity. The "mixture" effect, 363 364 linked to antagonisms of synergisms effects between sought, but also non-potentially thousand molecules not yet identified in UWWDs, are not considered into this study and can modify their 365 hazardousness. Secondly, the use of NOECs for PNEC values construction can be discussed. The 366 authors are aware that ECx values are more accurate than NOEC. ECx values are calculated from the 367 368 modeling of the dose-response curve, which is itself based on the results of the whole range of concentration, whereas NOEC corresponds only to the identification of the last concentration of the 369 range without effect. However, in the context of this work, it has been chosen to use to the largest set of 370 values available in the literature and in international databases for the two reasons: (1) In the specific 371 372 context of the PNEC calculation (i.e., a no-effect value for all aquatic organisms), considering the scarcity of existing data in the literature, it seems more useful to consider values for one or several 373 374 additional species than using accurate values for a limited number of species. In other words, inter-375 species variability is more important to take into account than intra-species variability, (2) This is the 376 most common approach adopted by international Agencies in charge of the establishment of PNECs for substances, such as INERIS in France, or the European Chemical Bureau in Europe, through the TGD 377 guidelines. Thirdly, our method considers only the maximum concentrations of the different pollutants 378 379 in UWWDs. Consequently, these values represent only the maximum hazard for each compound. This 380 work could also have been done with median values if sufficient data had been available. This problem How to cite this paper: Gosset, A., Durrieu, C., Orias, F., Bayard, R., Perrodin, Y. (2017) Identification and assessment of ecotoxicological hazards attributable to pollutants in urban wet weather discharges. Environmental Science: Processes & Impacts DOI: 10.1039/C7EM00159B

of limited data collection is at the origin of another limit: most of the chemical studies focusing on 381 382 dissolved compound concentrations found in stormwater and CSO have dealt with relatively small areas 383 of the world and many of them were published in Europe. Thus data are lacking for some continents, a 384 factor which limits our studies due to differences in, for example, PC consumption, pesticide use (e.g. glyphosate is widely used in South America), sewer systems, etc. Consequently, we can confidently 385 assert that the HQs calculated here are underestimated and would probably increase if more data were 386 387 available. Fourthly, dissolved part of UWWDs were only considered in this study because the PNEC 388 values were established for dissolved substances and it was impossible to extend them to pollutants whose bioavailability is unknown but present on particles in UWWDs. Indeed, pollutant bioavailability 389 is linked to particle desorption which depends on multiple parameters like pH, molar mass, 390 solubility/lipophily, and the number of carboxyl groups (for organic pollutants).⁷⁵ It is therefore possible 391 that our method underestimated the hazard of these UWWD pollutants for aquatic environments. For 392 example, as in the study of Zgheib et al.⁴ if we had considered the concentration of fluoranthene in the 393 dissolved phase (0.018 μ g L⁻¹) and the total effluent (0.832 μ g L⁻¹), assuming that all the pollutants are 394 available, the hazard index of the pollutants would have increased from 1.5 to 69.33, making it one of 395 396 the most hazardous compounds for aquatic organisms. This point is illustrated in Figure 7. Similar findings were recorded by Baun et al.³⁰ who prioritized 56 organic pollutants in the water phase of 397 398 stormwater, against 121 in the solid phase. Fifthly, many chemical studies have considered only mineral 399 and organic priority pollutants from the EU Water Framework Directive and other international 400 legislations. However, many emerging compounds currently used present proven ecotoxic effects but 401 have not yet been analyzed. For example, pesticides used massively in Europe like metam-sodium and epoxiconazole present relatively low PNEC values (respectively 0.79 and 0.18 μ g L⁻¹). We can therefore 402 expect high hazard quotients linked to these kinds of molecule. Finally, the presence of high 403 404 concentrations of bacteria in CSO has already been detected, for example by Passerat et al. (with concentrations up to 10⁷ cell/L for *E.coli*).⁷⁶ It is now known that bacteria can transform pollutants and 405 create by-products that are more or less toxic than their parent compounds detected in CSO. These 406 407 bacteria can therefore influence the composition of CSO and their toxicity with time.

Despite these limits, our method enabled us to highlight new outlooks in terms of prioritization 408 409 and management of UWWD pollutants. Comparing our results to previous works done with the same 410 purpose, we can observe that considering only intrinsic characteristics (e.g. particle solubility, sorption potential, ecotoxicity) of the pollutants selected, Baun et al.³⁰ classified 56 organic compounds as 411 priority pollutants dissolved in water using their RICH (Ranking and Identification of Chemical 412 Hazards) method against 37 in our case (of which 31 are organic). The difference between the two 413 studies is particularly notable for pesticides (10 against 33 pesticides for the method of Baun et al.³⁰). 414 415 This difference can be explained by non-detected or low environmental concentrations of many pesticides dissolved in water of our studied discharges. On the contrary, we revealed here 9 hazardous 416 PAH and 3 hazardous PHT instead of respectively 5 and 2 in Baun et al.³⁰ An explanation of this 417 discrepancy could be the multiplication of decision filters (sorption/volatilization, technical problems, 418 persistence, bioaccumulation, toxicity) applied by Baun et al. when selecting their compounds, and the 419 420 lack of data for some pollutants that can lead to the rejection or selection of the priority list, as suggested by these authors. In the case of CSO linked hazard, no studies have investigated a similar prioritization 421 method. Consequently, no comparison can be done. 422

423 Looking for the scientific literature, our study enabled us to find new hazardous and priority pollutants. The same observation could be made for the European (Urban Wastewater Directive (UWD 424 91/271/EEC)) and American (USA EPA) priority pollutant lists. In fact, as stated in Table S2, some 425 426 pollutant identified here as hazardous such as diuron or fluoranthene are already under the regulation of 427 international instances. On the contrary, some pollutants are absent from these lists. For example, some 428 heavy metals (e.g. copper and zinc in UWD list), pesticides as glyphosate or pyrethrinoids in stormwater, phthalates (e.g. diisononyl phthalate) and pharmaceuticals such as ibuprofen or trimethoprim in CSO 429 absent in UWD and EPA priority lists, are revealed to be at the origin of an important hazard for aquatic 430 431 receiving ecosystems, and thus should be so the object of reinforced controls and could be proposed to 432 be added to these lists.



433 Figure 7: Example of hazard quotients (HQ) calculated with dissolved and total concentrations of pollutants in stormwater samples from Zgheib et al.⁴

434

435 <u>Conclusion and perspectives</u>

The aim of our study was to develop and apply a methodology to determine and calculate hazard 436 437 quotients associated with the different pollutants sought and found in UWWDs. A wide range of 438 pollutants was sought in the latter, with highly variable ecotoxicity and concentrations. Although many families of pollutants (PC, PAH, VOC, ALK, M, P, O, PFAA and PHT) play a role in the UWWD 439 hazards identified, we emphasize that they are comprise of a large number of pesticides and heavy metals 440 (e.g. pyrethrinoïds, copper and cadmium). The hazards of the pollutants identified in this study have 441 442 different origins: (i) the presence of high concentrations of certain highly ecotoxic compounds (e.g. Pyrene in stormwater), (ii) the presence of high concentrations of slightly ecotoxic compounds (e.g. 443 heavy metals in CSO), (iii) the presence of weak concentrations of certain highly ecotoxic compounds 444 445 (e.g. pesticides like Bifenthrin and Permethrin in stormwater samples), (iv) the presence of intermediate concentrations of compounds with intermediate ecotoxicity (e.g. alkylphenols in UWWDs). We showed 446 that the hazard linked to CSO was lower than that linked to stormwater from separate systems. However, 447 448 these results should be considered with caution because only a few pharmaceutical compounds (just 28 compounds) were sought in CSO, compared to the huge range of drugs consumed by human beings. 449 450 Moreover, no ecotoxicity data were available for 13 of the compounds (3 PC, 6 PAH, 2 ALK, 1 ORG, 1 PFAA) detected in UWWDs, and therefore no PNEC or HQ values could be calculated. Consequently, 451 it will be necessary in further studies to obtain more ecotoxicity data on these compounds. Moreover, 452 current hazard and risk assessment methods must be improved to take into account pollutant 453 454 bioconcentration/accumulation, which can represent a serious risk for higher trophic levels.

To conclude, this work enabled us to identify the potential hazard of UWWDs and the pollutants involved. We now know which pollutants require an in-depth monitoring. We have also identified different points which require in-depth studies. In particular, it is now necessary to better assess the hazard of certain pollutants bound to the particulate phase. Lastly, this work should help managers to identify the largest and most hazardous pollution sources in order to treat them in priority.

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593 Table S1: Extrapolation Factor assessment criteria derivate from Orias and Perrodin³⁶ and TDG³⁷.

	Available data	EF (Extrapolation Factor)
	At least one short-term L(E)C50 from each	1000
	of three trophic levels of the base-set	
	(e.g. algae, daphnids and fishes)	
	One long term chronic data	100
	(e.g. invertebrate or fish)	
	Two long term chronic data from species representing	50
	two trophic levels (e.g.algae and/or daphnids and/or fishes)	10
	Two long term chronic data from species representing three trophic levels (e.g. algae, daphnids and fishes)	10
	Statistics method (e.g. species sensitivity distribution (SSD)),	5-1 (Reviewed case by case)
	field data or model ecosystem	5-1 (Reviewed case by case)
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	Identified as hazardous		Identification as a priority subtance		
Pollutant	In stormwater	In combined sewer overflows	In European Water Framework Directive (WFD, 2000/60/EC) pollutant priority list	In Environmental Protection Agency pollutant priority list	
4-(para)-nonylphenol (ALK)	Yes	No	Yes	No	
4-tert-pentylphenol (ALK)	Yes	No	Yes	No	
4-tert-butylphenol (ALK)	Yes	No	Yes	No	
Nonylphenol (ALK)	No	Yes	Yes	No	
Para-tert-octylphenol (ALK)	Yes	No	Yes	No	
Cadmium (M)	Yes	Yes	Yes	Yes	
Chromium III (M)	Yes	No	No	Yes	
Copper (M)	Yes	Yes	No	Yes	
Lead (M)	Yes	Yes	Yes	Yes	
Nickel (M)	Yes	No	Yes	Yes	
Zinc (M)	Yes	Yes	No	Yes	
Bisphenol A (O)	Yes	No	No	No	
Aminotriazole (P)	Yes	Yes	No	No	
Bifenthrin (P)	Yes	No	No	No	
Cyfluthrin (P)	Yes	No	No	No	
Cypermethrin (P)	Yes	No	Yes	No	
Deltamethrin (P)	Yes	No	No	No	
Desethylatrazine (P)	Yes	Yes	No	No	
Diuron (P)	Yes	Yes	Yes	No	
Glyphosate (P)	Yes	No	No	No	
Isoproturon (P)	Yes	No	Yes	No	
Lambda-cyhalothrin (P)	Yes	No	No	No	
N,N-diéthyl-3-					
méthylbenzamide (P)	No	Yes	No	No	
Permethrin (P)	Yes	No	No	No	
Triclocarban (P)	No	Yes	No	No	
Anthracene (PAH)	Yes	No	Yes	Yes	
Benzo[a]anthracene (PAH)	Yes	No	Yes	Yes	
Benzo[k]fluoranthene (PAH)	Yes	No	Yes	Yes	
Benzo[g,h,i]perylene (PAH)	Yes	No	Yes	Yes	
Benzo[a]pyrene (PAH) Fluoranthene (PAH)	Yes	No	Yes	Yes	
Fluorene (PAH)	Yes	Yes	Yes	Yes	
Naphthalene (PAH)	Yes	No	Yes	Yes	
Pyrene (PAH)	Yes	No	Yes	Yes	
Ibuprofen (PC)	Yes	No	Yes	Yes	
Triclosan (PC)	No No	Yes	No	No	
Trimethoprim (PC)	No No	Yes Yes	No No	No No	
Perfluorooctanesulfonic acid					
(PFAA) Perfluorononanoic acid (PFAA)	Yes	No No	Yes	No	
Benzyl butyl phthalate (PHT)	Yes	No No	Yes	No Ves	
Diethylhexyl phthalate (PHT)	Yes	No	No	Yes	
Dieurymexyr philalate (PHT) Diisononyl phthalate (PHT)	Yes	No	Yes	Yes	
1,4-Dichlorobenzene (VOC)	Yes	No	No	No Vas	
1,4-Dichiotobelizelle (VOC)	Yes	No	No	Yes	

	Pentachlorophenol (VOC)	Yes	No	Yes	Yes
617	Table S2: Comparison of haz	ardous co	mpounds list i	dentified in UWWDs in th	is study with the

618 classification of these compounds as priority pollutants or not by European Water Framework

619 Directive and In Environmental Protection Agency.