

## Adjoint-based numerical method using standard engineering software for the optimal placement of chlorine sensors in drinking water networks

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 engineering software for the optimal placement of chlorine
 sensors in drinking water networks

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### 8 Abstract

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7

To obtain representative water quality simulations, unknown model param-9 eters have to be updated by combining information from the water quality 10 model and the sensor outputs. An adjoint-based numerical method has 11 been developed to determine the optimal placement of chlorine sensors in 12 drinking water networks at a low computational cost. From a practical en-13 gineering perspective, the proposed optimal placement corresponds to the 14 set of sensors that minimizes the area in which the unknown model param-15 eters cannot be identified. The numerical strategy is implemented in the 16 hydraulic software EPANET. Using the adjoint framework, we develop and 17 apply an adaptive strategy in a French drinking water network that provides 18 the optimal placement from 1 sensor to 6 sensors. We show that the highest 19 reduction of the non-identifiable area is obtained at the first stages of the 20 adaptive strategy. After 4 sensors, a plateau is reached. 21

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- 22 Keywords: sensor placement, water quality simulations, drinking water
- 23 network, adjoint method

### 24 Software availability

- 25 Name of software: EPANET
- <sup>26</sup> Programming language: C/C++
- 27 Operating system: Windows
- 28 Availability: http://www.epa.gov/water-research/epanet
- 29 Documentation: http://www.epa.gov/water-research/epanet
- <sup>30</sup> User interface: Graphical user interface or Programmer's toolkit
- <sup>31</sup> License: Public domain software that may be freely copied and distributed

### 32 1. Introduction

In drinking water networks, the chlorine concentration field is one of the 33 main indicators of the water quality. Legislation dictates that a minimum 34 level of chlorine at each point in the network has to be ensured. To overcome 35 the lack of measurements in drinking water networks, hydraulic and water 36 quality models are considered. In water network applications, the hydraulic 37 state is generally computed using algebraic equations, *i.e.*, flow continuity 38 at the nodes and headloss in the pipes. Regarding the water quality models, 39 one-dimensional (1D) advection-reaction equations are considered in pipes, 40 and perfect and instantaneous mixing is assumed in pipe junctions. The de-41 crease of the chlorine concentration due to bulk flow reactions and pipe wall 42 reactions, e.g., reaction with the biofilm at the pipe wall, is modeled using 43 a reaction term (Powell et al. (2000)). This term is characterized by the 44

reaction order and the reaction coefficient. The software EPANET (Rossman and Boulos (1996); Rossman (2000)) is commonly used to simulate the
hydraulics and the water quality states.

French water companies can observe a gap higher than 30% between the 48 chlorine sensor outputs and the chlorine concentration obtained from a di-49 rect simulation of the water quality model. This gap may be due to un-50 certainties in the hydraulic state, particularly the water demands, and to 51 the model parameters associated with the chlorine reactions. To represent 52 the variability in water demands, stochastic models are typically considered. 53 The calibration of these models can be achieved using direct measurements 54 (Buchberger and Wells (1996); Bakker et al. (2013); Cominola et al. (2015)), 55 *i.e.*, monitoring of the user water consumption in residences, or indirect mea-56 surements (Kang and Lansey (2009); Alcocer-Yamanaka et al. (2012)), *i.e.*, 57 pressure and/or flow rate outputs into the drinking water network. To lo-58 cate and quantify abnormal water demands due to leaks, inverse techniques 59 based on pressure sensor outputs have been proposed in (Liggett and Chen 60 (1994); Meseguer et al. (2014)). In this inverse problem, the goal is to 61 determine the unknown model parameters by minimizing the gap between 62 the sensor outputs and the simulation. Finally, flow sensor outputs may 63 also be used in inverse techniques. Indeed, an inverse computational fluid 64 dynamics technique has been developed in (Waeytens et al. (2015)) to iden-65 tify the unknown boundary conditions of 2D incompressible Navier-Stokes 66 equations and thus to obtain a high description in 2D of the flow profile in 67 water networks. A detailed description of the flow in 2D or 3D can provide 68 more representative chlorine simulations than using the mean flow velocity, 69 particularly in the distribution mains of the drinking water networks where 70 the flow can be laminar, thus inducing different chlorine propagation veloc-71

<sup>72</sup> ities. Note that the measurement of chlorine or tracer concentrations can
<sup>73</sup> also provide information on the water demands (Jonkergouw et al. (2008);
<sup>74</sup> Al-Omari and Abdulla (2009)).

Regarding the water quality models, first-order reaction kinetics is com-75 monly assumed for the free chlorine decay. Many articles address the iden-76 tification of the reaction coefficient (Sharp et al. (1991); Rodriguez et al. 77 (1997); Munavalli and Kumar (2005); Pasha and Lansev (2012)), but few 78 aim to determine the reaction order and the reaction coefficient (Vasconcelos 79 et al. (1997); Gancel (2006)). Because the reaction coefficient is associated 80 with bulk flow reactions and pipe wall reactions, it is not uniform in the 81 entire network. Nevertheless, to limit the number of unknowns to be de-82 termined, the reaction coefficient is considered to be piecewise constant on 83 subsections of the water network. The choice of the domain decomposition 84 is based on the age, the roughness, the pipe material, the pipe diameter and 85 the flow rate. 86

Because drinking water networks are sparsely instrumented, the use of nu-87 merical tools can indicate to the water companies the coverage area ensured 88 by the existing chlorine sensors and the optimal deployment of new chlorine 89 sensors. A considerable amount of literature addresses the optimal sensor 90 position for detecting a contaminant intrusion in drinking water networks. 91 Three categories can be distinguished: the non-model-based methods using 92 the topology of the water network, the methods based solely on hydraulic 93 simulations (Lee and Deininger (1992); Kessler et al. (1998); Berry et al. 94 (2005); Xu et al. (2008)) and the methods based on hydraulic and water 95 quality simulations (Berry et al. (2006); Preis and Ostfeld (2008); Krause 96 et al. (2008)). The majority of the methods formulate the optimal sensor 97 placement as a multiobjective optimization. The goal is to minimize the 98

<sup>99</sup> non-coverage area, the number of sensors, the time to detection, and so
<sup>100</sup> forth.

In the present article, we propose a numerical strategy based on the adjoint 101 framework to determine and quantify the non-coverage area for a given set 102 of sensor placements. This practical information can be useful for water 103 companies to determine the optimal placement of chlorine sensors for maxi-104 mizing the coverage area for a given number of sensors. The method requires 105 the resolution and the post-treatment of the solution of the adjoint problem, 106 *i.e.*, advection-reaction equations backward in time with virtual chlorine in-107 jection at the position of the sensors and a dynamic back flow. The adjoint 108 framework is used in various applications. First, it provides at a low com-109 putational cost the functional gradient involved in inverse calculations to 110 update the model parameters of the water flow (Liggett and Chen (1994); 111 Waevtens et al. (2015)) and to reconstruct the concentration fields (Elbern 112 et al. (2000); Waevtens et al. (2013, 2017)). Then, it is used in sensitiv-113 ity analyses to study the influence of the physical model parameters on a 114 quantity of interest (Andrews (2013); Kauker et al. (2016)). The adjoint 115 framework is also considered for estimating the modeling or the discretiza-116 tion error on a quantity of interest (Becker and Rannacher (2001); Waeytens 117 et al. (2012); Oden and Prudhomme (2002)). Note that the determination 118 and quantification of the coverage area can also be obtained from (Xu et al. 119 (2008)), which is based on the knowledge of the flow and graph theory. The 120 main advantage of the proposed adjoint approach is that it uses standard 121 hydraulic software such as EPANET. Moreover, the adjoint solution can also 122 be used in an inverse advection-reaction procedure to identify the reaction 123 coefficient. 124

<sup>125</sup> The remainder of this article is organized as follows. Section 2 introduces the

model updating technique for identifying the reaction coefficient of the water
quality model and the definition of the adjoint problem. The adjoint-based
numerical strategy for the optimal chlorine sensor placement is developed
in Section 3. This strategy is applied to a French drinking water network in
Section 4 before drawing concluding remarks and prospects in Section 5.

# 2. Modeling the water quality in drinking water networks and updating the kinetic reaction coefficient

#### 133 2.1. Simulating the water quality in drinking water networks

In drinking water networks, the chlorine concentration is the primary 134 indicator of the water quality. The chlorine propagates in the network 135 according to the flow induced by water demands, and the chlorine con-136 centration decreases due to reactions occurring in the bulk or at the wall. 137 Physical models can be employed to predict the propagation and the re-138 action of chlorine in drinking water networks. Generally, one-dimensional 139 (1D) advection-reaction partial differential equations are considered in the 140 pipes, and the mixing in the junctions is modeled using algebraic equations. 141 The set of equations, detailed in (Rossman (2000)), for modeling the water 142 quality in the drinking water network is called a "direct problem". It can 143 be solved using standard engineering software such as EPANET (Rossman 144 and Boulos (1996)). Let us define the simulated chlorine concentration in 145 the water network as **C**. 146

In practice, water companies may observe a gap higher than 30% between the simulated and measured chlorine concentrations. Hence, to obtain a representative simulation of the water quality, the model parameters, such as the kinetic reaction coefficient, have to be updated. The model updating <sup>151</sup> strategy is described in the following sections.

### 152 2.2. Cost functional used in model updating

To obtain representative water quality simulations, one needs to update 153 the unknown parameters of the model. Herein, we focus on determining the 154 vector  $\mathbf{k} = \{k_1, ..., k_N\}$  of reaction coefficients, where N corresponds to the 155 number of water pipes in the drinking water network. For this purpose, an 156 inverse modeling technique can be employed. Let us choose a quadratic cost 157 functional that quantifies the difference between the sensor outputs  $C_m^{mes}$ 158 and the numerical solution  $\mathbf{C}$  of the water quality model mentioned in the 159 previous section. One seeks the vector  $\mathbf{k}$  of reaction coefficients by solving 160 the following optimization problem: 161

$$\min_{\mathbf{k}} J(\mathbf{C}, \mathbf{k}) = \frac{1}{2} \sum_{m=1}^{N} a_m \int_0^T \left( C_m(\mathbf{k}; x = x_m, t) - C_m^{mes}(t) \right)^2 dt + \beta b \sum_{m=1}^{N} (k_m - k_m^0)^2$$
(1)

where  $\mathbf{C} = \{C_1(x,t), ..., C_N(x,t)\}$  is the vector of chlorine concentrations. The Boolean parameter  $a_m$  is set to 1 (resp. 0) if the  $m^{th}$  water pipe is equipped (resp. is not equipped) with a chlorine sensor recording the concentration level on the observation time interval [0,T]. The position of the  $m^{th}$  sensor is denoted as  $x_m$ . Note that Eq. (1) is a constrained optimization problem. Indeed, the chlorine concentration field  $\mathbf{C}$  has to satisfy the set of water quality equations mentioned in Section 2.1.

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In general, the inverse problem is not well posed. First, in practice, to reduce the number of model parameters to be updated, the reaction coefficient is assumed to be the same on a group of water pipes that have the same characteristics, *e.g.*, age, diameter and material (Fabrie et al. (2010)). Second, a Tikhonov regularization term such as the second term in Eq. (1) can be introduced in the cost functional. This term aims at improving the convexity of the functional. The parameter b ensures the physical homogeneity of both terms, and  $k_m^0$  corresponds to the initial guess of the  $m^{th}$ reaction coefficient.  $\beta$  corresponds to a normalized weighting coefficient.

The minimization of the cost functional (1) can be performed using a gradient-like approach. In the present article, note that the functional gradient is obtained at a low computational cost using the adjoint framework. The methodology to derive the adjoint equation and the gradient formula are presented in the next section.

### 185 2.3. Derivation and physical meaning of adjoint equations

The constrained minimization problem (1) can be rewritten as an unconstrained minimization problem by introducing the Lagrangian  $\mathcal{L}(\mathbf{k}, \mathbf{C}, \mathbf{P})$ and the Lagrange multiplier **P**. The stationarity of the Lagrangian according to the Lagrange multiplier **P** provides the equations of the direct problem mentioned in Section 2.1, whereas the equations of the adjoint problem are obtained from the stationarity of the Lagrangian according to **C**.

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Herein, the adjoint problem is quite similar to the direct problem. It is still an advection problem with a reaction term. In the adjoint problem, chlorine is virtually injected at the sensor location  $x_m$ . The temporal evolution of the chlorine injection is provided by the data misfit. In contrast to the direct problem, the adjoint problem has a final condition in time. Moreover, the flow is reversed in the adjoint problem, *i.e.*, the velocity **v** is replaced by  $(-\mathbf{v})$ .

200 Physically, the adjoint state corresponds to a "sensitivity concentration".

201 Considering a sensor at a given location in the drinking water network, the

<sup>202</sup> "sensitivity concentration" propagates from the sensor location toward the <sup>203</sup> upstream flow direction with increasing reversed time. Hence, it shows that <sup>204</sup> all of the sensitivity of the sensor measurement is located upstream of the <sup>205</sup> sensor location. In other words, the sensor is not sensitive to locations down-<sup>206</sup> stream of the sensor, nor is it sensitive to chlorine concentrations that were <sup>207</sup> present prior to the initial observation time.

208

The adjoint state can provide sensitivity information, but it can also be 209 used to compute the functional gradient  $\nabla J$ , which is involved in model 210 updating techniques, at a low computation cost. Indeed, n + 1 transport 211 reaction problems have to be solved in standard finite-difference techniques, 212 whereas only two transport reaction problems are needed when using the 213 adjoint framework to compute the n components of the functional gradi-214 ent. The functional gradient  $\nabla J$  is obtained from the derivative of the 215 Lagrangian according to the reaction coefficient  $k_m$ . 216

217

To provide a better understanding, the derivation of the adjoint equations and the functional gradient are illustrated on a divergent node of a water network in Appendix B.

# 221 2.4. Practical technique to update the reaction coefficient of the water quality 222 model

Drinking water networks are not massively instrumented with chlorine sensors. As mentioned in Section 2.2, to limit the number of model parameters to be updated, the reaction coefficient is assumed to be uniform on subdomains of the water network that have the same characteristics. To update the vector  $\mathbf{k}$  of reaction coefficients, one can follow the iterative <sup>228</sup> strategy detailed below.

<sup>229</sup> Direct problem:. First, considering an initial guess  $\mathbf{k}^0$  or the kinematic <sup>230</sup> parameters obtained at the end of the previous iteration, the advection-<sup>231</sup> reaction direct problem is solved, thereby providing the concentration field <sup>232</sup> **C** in the water network.

Comparison of simulated chlorine and sensor outputs:. Knowing the simulated concentration field C in the entire water network, we compute the
data misfit at each sensor location.

Adjoint problem:. In a water network, the adjoint problem is defined as an
advection-reaction problem backward in time considering a reversed flow.
It corresponds to the retropropagation of chlorine virtually injected at the
sensor location. The higher is the data misfit, the higher is the chlorine to
be injected.

After changing the time variable t to  $\tau = T - t$ , the final time condition begins as an initial condition. Hence, standard hydraulic and water quality software can be employed to solve the adjoint problem. By solving this problem, we obtain the adjoint state **P**.

Functional gradient:. Let us consider the reaction coefficient  $k_p$  modeling the chlorine reaction in a subdomain  $\Omega_p$  of the water network. This subdomain is composed of  $n_p$  water pipes. Thus, the formula of the derivative of the functional according to  $k_p$  can be expressed as

$$\frac{\partial J}{\partial k_p} = -\sum_{i=1}^{n_p} S_i \int_0^T \int_0^{L_i} C_i P_i dx dt + \beta b(k_p - k_p^0) \tag{2}$$

In Eq. (2), the first term is associated with the sensitivity of the data misfit to the reaction coefficient  $k_p$ , and the second term is dedicated to the <sup>251</sup> functional regularization.

All the components of the functional gradient  $\nabla J$  are computed from Eq. (2).

Updating of the reaction coefficients:. Using the functional gradient  $\nabla J$ as the descent direction, we obtain the updated vector  $\mathbf{k_{new}}$  of reaction coefficients by

$$\mathbf{k}_{new} = \mathbf{k}_{old} - \alpha \nabla \boldsymbol{J} \tag{3}$$

where  $\alpha$  is the descent step. Several solutions are required to determine the descent step minimizing the cost functional J.

If the data misfit functional reaches the measurement error, the model up-dating process is stopped. Otherwise, the iterative process continues.

### 261 3. Optimal chlorine sensor placement using the adjoint framework

- 262 3.1. Theoretical foundations of the optimal chlorine sensor placement method 263 **Proposition 3.1.** Let us denote  $\phi^*(x,t)$  as the modified adjoint solution. 264 If  $\phi^*(x,t) = 0$  in  $\Omega_p \times [0,T]$ , then the reaction coefficient k(x) is not iden-265 tifiable on the subdomain  $\Omega_p$ .
- The modified adjoint solution  $\phi^*(x,t)$  verifies the following transport equations in the pipes of the water networks

$$-\frac{\partial \phi_m^*}{\partial t} - v_m \frac{\partial \phi_m^*}{\partial x} = a_m H(t) \delta(x - x_m), \qquad in \ [0, L_m] \times [0, T], \ m \in \{1, ..., N\}$$

$$\tag{4}$$

The boolean  $a_m$  is equal to 1 (resp. 0) if the  $m^{th}$  water pipe is equipped with a chlorine sensor (resp. is not equipped with a chlorine sensor). As with the adjoint problem, the pipe junctions are governed by the standard equations of convergent or divergent nodes depending on the direction of the flow <sup>272</sup>  $(-v_m)$ . The flow  $(-v_m)$  is considered stationary. Finally, in the modified <sup>273</sup> adjoint problem, the final condition and the Dirichlet boundary conditions <sup>274</sup> vanish.

Note that the modified adjoint problem resembles the adjoint problem. The two differences are as follows. First, no reaction term is involved in the modified adjoint problem. Second, in the adjoint problem, the amplitude of the injected chlorine at the sensor location is given by the data misfit, whereas a constant amplitude in time is considered in the modified adjoint problem.

281

The proof of Proposition 3.1 is given for a reduced water network in Appendix B. Although a reduced water network is considered to facilitate the notations, it includes key elements of a real water network, *i.e.*, divergent and convergent nodes.

### 286 3.2. Numerical method for optimal chlorine sensor placement

For a given number  $n_s$  of chlorine sensors, we seek the placement of chlorine sensors that minimizes the non-identifiable area associated with the reaction coefficient k(x). For this purpose, we propose an adjoint-based numerical method, which is detailed below.

First, the hydraulic v(x, t) has to be simulated in the drinking water network. From the hydraulic, we deduce the reversed velocity field, *i.e.*,  $(-1) \times v(x, t)$ . In agreement with the deployment constrains in the water network, a possible placement of  $n_s$  chlorine sensor is considered. Then, the reversed velocity field is used to obtain the modified adjoint concentration  $\phi^*(x, t)$ . The modified adjoint problem, introduced in Proposition 3.1, corresponds to the retropropagation of chlorine virtually injected at the sensor locations

with a constant unitary amplitude in time. In practice, after making the 298 change of variable  $\tau = T - t$ , this problem can be solved using standard 299 engineering software, e.g., EPANET. In Proposition 3.1, we show that in-300 formation concerning the non-identifiable area associated with the reaction 301 coefficient k(x) can be provided by the modified adjoint concentration field 302  $\phi^*(x,t)$ . The subdomain  $\Omega_p$  having a null modified adjoint concentration on 303 the entire time interval [0,T] corresponds to the non-identifiable area. To 304 quantify and compare different sensor locations in view of minimizing the 305 non-identifiable area, we introduce the dimensionless non-coverage indicator 306  $\eta$ , which is defined as 307

$$\eta = \frac{\text{Total length of water pipes in the non-identifiable area}}{\text{Total length of water pipes in the drinking water network}}$$
(5)

The proposed indicator is rapidly computable and readily usable. Thus, it meets the expectations of hydraulic engineers.

### 310 4. Numerical results using the software EPANET

Let us consider a part of the French drinking water network presented in 311 Figure 1. The water network is composed of two tanks, 298 junctions and 318 312 pipes. The total water pipe length is approximately 15 km. Due to technical 313 and deployment constrains, chlorine sensors cannot be installed anywhere in 314 the drinking water network. Hence, the water company Suez-Environnement 315 has pre-selected 6 potential chlorine sensor positions, as presented in Figure 316 1. The chlorine sensors are useful for comparing the measurements and the 317 water quality simulations. As we previously mentioned, more representative 318 water quality simulations can be achieved by updating the unknown reac-319 tion coefficient k(x). 320

321

Our goal is to determine the best set of chlorine sensors that maximize the coverage area, *i.e.*, minimizing the non-identifiable area associated with the reaction coefficient k(x). In this section, we apply the proposed adjointbased numerical method. It has been implemented in the software EPANET.



Figure 1: Geometry of a French drinking water network with 2 water towers - 6 possible locations of chlorine sensors

Following the methodology presented in the previous section, the first 327 step consists of the hydraulic simulation in the water network using EPANET. 328 As input for the hydraulic simulation, we use estimated varying consumer 329 demands and the initial water level in water towers. The information as-330 sociated with one day in August 2011 was provided by the water company 331 Suez-Environnement. Then, we reverse the simulated flow, *i.e.*, -v(x, t), for 332 simulating the modified adjoint problem  $\phi^*(x,t)$ . We recall that the modi-333 fied adjoint problem is defined as the retropropagation of chlorine virtually 334

injected at the sensor locations with a constant unitary amplitude in time.

336

As a first step, we consider the deployment of only one chlorine sensor in 337 the water network. We test several locations of the sensor, and for each po-338 sition from  $S_1$  to  $S_6$  (see Figure 1), we solve the associated modified adjoint 339 problem. In Figure 2, we show the non-identifiable area when considering a 340 sensor placed at  $S_1$  or at  $S_6$ . The blue color (resp. the red color) denotes 341 the area where the modified adjoint solution vanishes (resp. is not null) on 342 the entire daily time interval. According to the theoretical results of the 343 previous section, the blue color area is associated with the non-identifiable 344 area. To quantify this area, the indicator  $\eta$  defined in Eq. (5) is computed. 345 The results are summarized in Table 1. The non-identifiable area represents 346 87.5% of the water network for a sensor placed at  $S_1$ , whereas it represents 347 73.4% of the water network for  $S_6$ . Moreover, in Figure 2, we observe that 348 a sensor placed at  $S_1$  (resp. at  $S_6$ ) is not able to provide information on 349 chlorine reactions in the lower part of the network (resp. in the upper part 350 of the network). When considering a unique sensor, note that the optimal 351 sensor placement for minimizing the non-identifiable area corresponds to  $S_6$ . 352 To reduce the non-identifiable area, more chlorine sensors should be de-353 ployed in the water network. In the following, combinations of chlorine 354 sensors are studied. In Figure 3, we can observe that when using the com-355 bination of the three chlorine sensors  $S_1 - S_3 - S_6$ , the non-identifiable area 356 represents 54.4% of the drinking water network. Considering these 3 sensors 357 rather than only sensor  $S_1$  provides a 3 km reduction in the non-identifiable 358 area. Nevertheless, increasing the number of sensors does not strongly re-359 duce the non-identifiable area. Indeed, in Figures 3 and 4, we observe that 360 from 3 to 5 chlorine sensors, the non-identifiable area indicator changes from 361



Figure 2: Sensor position  $S_1$ :  $\eta = 87.5\%$  (left), Sensor position  $S_6$ :  $\eta = 73.4\%$  (right)

| Sensor | Non-coverage indicator $\eta$ |  |  |  |
|--------|-------------------------------|--|--|--|
| $S_1$  | 87.5 %                        |  |  |  |
| $S_2$  | 97.5 %                        |  |  |  |
| $S_3$  | 88.5 %                        |  |  |  |
| $S_4$  | 81.9 %                        |  |  |  |
| $S_5$  | 86.3~%                        |  |  |  |
| $S_6$  | 73.4 %                        |  |  |  |

Table 1: Non-coverage area of the drinking water network when considering a unique chlorine sensor

 $_{362}$  54.4% to 48.1%. No improvements are observed when adding a sixth sensor.

An adaptive strategy can be applied to obtain a desired threshold of non-identifiable area. The adaptive process starts by considering a unique chlorine sensor. Using the proposed adjoint-based technique, we retain the sensor placement  $S_I^{opt}$  that has the lowest non-identifiable area indicator  $\eta$ . Then, to continue decreasing the non-identifiable area, an additional chlorine sensor is considered in the drinking water network. The indicator



Figure 3: Combination of three sensors S6 - S1 - S3:  $\eta = 54.4\%$ 



Figure 4: Combination of 5 sensors  $S_6 - S_1 - S_3 - S_4 - S_2$ :  $\eta = 48.1\%$  (right), Combination of 6 sensors  $S_1$  to  $S_6$ :  $\eta = 48.1\%$  (left)

 $\eta$  is computed for all combinations of two sensors, including the sensor  $S_I^{opt}$ determined at the previous stage. Hence, we obtain the optimal combination of two sensors  $(S_I^{opt}, S_{II}^{opt})$ . The adaptive procedure continues until we reach the maximum number of chlorine sensors affordable for the water network. This adaptive strategy has been applied to the investigated water network. The results are summarized in Tables 2 and 3. For a fixed number of sensors, the optimal placement is noted in bold in Tables 2 and 3. Figure 5 presents the evolution of the coverage area indicator  $1 - \eta$  for the optimal placement of chlorine sensors. The highest reduction in the non-identifiable area is obtained at the first stages of the adaptive strategy. After 4 sensors, a plateau is reached. The non-coverage indicator is approximately 50%.

| 1 Sensor         | Non-coverage     | 2 Sensors  | Non-coverage     | 3 Sensors                                  | Non-coverage     |
|------------------|------------------|------------|------------------|--|------------------|
|                  | indicator $\eta$ |            | indicator $\eta$ |  | indicator $\eta$ |
| $S_1$            | 87.5 %           | $S_6, S_1$ | 61.7 %           | $S_6, S_1, S_2$                            | 59.2 %           |
| $S_2$            | 97.5%            | $S_6, S_2$ | 70.9~%           | $\mathbf{S_6}, \mathbf{S_1}, \mathbf{S_3}$ | 54.4~%           |
| $S_3$            | 88.5%            | $S_6, S_3$ | 63.5~%           | $S_6, S_1, S_4$                            | 56.1 %           |
| $S_4$            | 81.9 %           | $S_6, S_4$ | 67.8 %           | $S_6, S_1, S_5$                            | 61.8 %           |
| $S_5$            | 86.3%            | $S_6, S_5$ | 72.1 %           |  |                  |
| $\mathbf{S}_{6}$ | 73.4%            |            |                  |  |                  |

Table 2: Non-coverage area of the drinking water network when considering 1, 2 or 3 chlorine sensors - optimal combinations of sensors are noted in bold

| 4 Sensors  | Non-coverage     | 5 Sensors  | Non-coverage     | 6 Sensors  | Non-coverage     |
|--|------------------|--|------------------|--|------------------|
|  | indicator $\eta$ |  | indicator $\eta$ |  | indicator $\eta$ |
| $S_6, S_1, S_3, S_2$                                     | 53.7 %           | $\mathbf{S_6}, \mathbf{S_1}, \mathbf{S_3}, \mathbf{S_4}, \mathbf{S_2}$ | 48.1%            | $\mathbf{S_6}, \mathbf{S_1}, \mathbf{S_3}, \mathbf{S_4}, \mathbf{S_2}, \mathbf{S_5}$ | 48.1 %           |
| $\mathbf{S_6}, \mathbf{S_1}, \mathbf{S_3}, \mathbf{S_4}$ | 48.8 %           | $S_6, S_1, S_3, S_4, S_5$  | 48.8 %           |  |                  |
| $S_6, S_1, S_3, S_5$                                     | 54.4 %           |  |                  |  |                  |

Table 3: Non-coverage area of the drinking water network when considering 4, 5 or 6 chlorine sensors - optimal combinations of sensors are noted in **bold** 

### 380 5. Conclusions

<sup>381</sup> To obtain representative water quality simulations in drinking water net-

382 works, the unknown model parameters, such as the reaction coefficient,



Figure 5: Synthesis of optimal sensor placement to maximize the coverage area

should be updated using chlorine sensor outputs. In the present article, 383 an adjoint-based numerical method dedicated to drinking water networks 384 has been developed to optimally deploy chlorine sensors in view of minimiz-385 ing the non-identifiable area associated with the reaction coefficient. The 386 computation of the one-dimensional adjoint advection solution using the 387 standard engineering software EPANET allows us to localize and quantify 388 the non-coverage area for a given set of sensors. On a French drinking water 389 network, we applied an adaptive strategy starting from the optimal place-390 ment of 1 sensor to 6 sensors. We showed that the highest reduction in the 391 non-identifiable area is obtained at the first stages of the adaptive strategy. 392 After 4 sensors, a plateau is reached. In the model updating process of the 393 reaction coefficient, the computed adjoint solution can also be used to obtain 394 the functional gradient at a lower computational cost than straightforward 395 difference techniques. Herein, we focus on the optimal sensor placement that 396

minimizes the non-coverage area. The non-coverage area indicator computed from the modified adjoint solution can be used in a more general optimal sensor placement strategy considering the minimization of a multiobjective function. Finally, the proposed method can be extended for the detection of species intrusion in drinking water networks. The modified adjoint solution can highlight the area where species intrusion may not be detected.

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# Appendix A. Illustration of adjoint framework on a divergent node of a water network and computation of the func tional gradient

The advection reaction within a divergent node (see Figure A.6) is represented by the following equations:

$$\begin{aligned} \frac{\partial C_m}{\partial t} + v_m \frac{\partial C_m}{\partial x} + k_m C_m &= 0 & \text{in } [0, L_m] \times [0, T], \ m \in \{1, 2, 3\} \\ C_1(x = 0, t) &= \chi_1(t) & \text{in } [0, T] \\ C_2(x = 0, t) &= C_1(x = L_1, t) & \text{in } [0, T] \\ C_3(x = 0, t) &= C_1(x = L_1, t) & \text{in } [0, T] \\ C_m(x, t = 0) &= c_m^0(x) & \text{in } [0, L_m], \ m \in \{1, 2, 3\} \end{aligned}$$
(A.1)

where *m* denotes the pipe number,  $L_m$  is the length of pipe *m*,  $C_m$  is the chlorine concentration in water pipe *m* as a function of distance *x* and time *t*,  $v_m$  is the flow velocity, and  $k_m$  (resp.  $\alpha_m$ ) represents the reaction coefficient (resp. the reaction order). As usual in drinking water networks, a first-order reaction is considered to model the chlorine reaction. In the following, we take  $\alpha_m = 1$ . The concentration boundary condition is  $\chi_1(t)$ , and  $c_m^0(x)$  denotes the initial chlorine concentration in water pipe *m*.



Figure A.6: Notations for advection reaction through divergent node - the flow velocity  $v_m$  in pipe m is considered positive when it passes from the starting node (x = 0) to the ending node  $(x = L_m)$  - herein,  $v_1 > 0$ ,  $v_2 > 0$  and  $v_3 > 0$ 

To derive the adjoint equations and the gradient formula, we introduce

424 the Lagrangian

$$\mathcal{L}(\mathbf{k}, \mathbf{C}, \mathbf{P}, \boldsymbol{\lambda}, \boldsymbol{\mu}) = J(\mathbf{C}, \mathbf{k}) - \sum_{m=1}^{3} S_m \int_0^T \int_0^{L_m} \left(\frac{\partial C_m}{\partial t} + v_m \frac{\partial C_m}{\partial x} + k_m C_m\right) P_m(x, t) dx dt$$
  
$$-S_1 \int_0^T (C_1(x=0, t) - \chi_1(t)) \lambda_1 dt - \sum_{m=2}^{3} S_m \int_0^T (C_m(x=0, t) - C_1(x=L_1, t)) \lambda_m dt$$
  
$$-\sum_{m=1}^{3} S_m \int_0^{L_m} (C_m(x, t=0) - c_m^0(x)) \mu_m dx$$
  
(A.2)

<sup>425</sup> and the cost functional is defined as

$$J(\mathbf{C}, \mathbf{k}) = \frac{1}{2} \sum_{m=1}^{3} a_m \int_0^T \left( C_m(\mathbf{k}; x_m) - C_m^{mes} \right)^2 dt + \beta b \sum_{m=1}^{3} (k_m - k_m^0)^2$$
(A.3)

426

427 where 
$$\mathbf{k} = \begin{pmatrix} k_1 \\ k_2 \\ k_3 \end{pmatrix}$$
,  $\mathbf{C} = \begin{pmatrix} C_1 \\ C_2 \\ C_3 \end{pmatrix}$ ,  $\mathbf{P} = \begin{pmatrix} p_1 \\ p_2 \\ p_3 \end{pmatrix}$ ,  $\lambda = \begin{pmatrix} \lambda_1 \\ \lambda_2 \\ \lambda_3 \end{pmatrix}$  and  $\mu = \begin{pmatrix} \mu_1 \\ \mu_2 \\ \mu_3 \end{pmatrix}$ 

<sup>428</sup> and  $S_m$  denotes the cross-sectional area of the  $m^{th}$  water pipe.

429

As mentioned in Section 2.3, by writing the stationarity of the Lagrangian according to the Lagrange multipliers **P**,  $\lambda$  and  $\mu$ , we obtain the equations of the direct problem (A.1).

433

434 The equations of the adjoint problem are obtained from the stationarity

 $_{435}$  of the Lagrangian according to C. Let us derive the adjoint equations.

$$\begin{aligned} \frac{\partial \mathcal{L}}{\partial C_{1}} \delta C_{1} &= 0 \Rightarrow \\ & -S_{1} \int_{0}^{T} \int_{0}^{L_{1}} \left( \frac{\partial \delta C_{1}}{\partial t} + v_{1} \frac{\partial \delta C_{1}}{\partial x} + k_{1} \delta C_{1} \right) P_{1}(x, t) dx dt - S_{1} \int_{0}^{T} \delta C_{1}(x = 0, t) \lambda_{1} dt \\ & + \sum_{m=2}^{3} S_{m} \int_{0}^{T} \delta C_{1}(x = L_{1}, t) \lambda_{m} dt - S_{1} \int_{0}^{L_{1}} \delta C_{1}(x, t = 0) \mu_{1} dx = 0, \ \forall \delta C_{1} \\ \frac{\partial \mathcal{L}}{\partial C_{m}} \delta C_{m} &= 0 \Rightarrow \\ & a_{m} \int_{0}^{T} \int_{0}^{L_{m}} \left( C_{m}(x_{m}) - C_{m}^{mes} \right) \delta(x - x_{m}) \delta C_{m} dx dt \\ & -S_{m} \int_{0}^{T} \int_{0}^{L_{m}} \left( \frac{\partial \delta C_{m}}{\partial t} + v_{m} \frac{\partial \delta C_{m}}{\partial x} + k_{m} \delta C_{m} \right) P_{m}(x, t) dx dt \\ & -S_{m} \int_{0}^{T} \delta C_{m}(x = 0, t) \lambda_{m} dt - S_{m} \int_{0}^{L_{m}} \delta C_{m}(x, t = 0) \mu_{m} dx = 0, \ \forall \delta C_{m}, \ m = 2, 3 \\ & (A.4) \end{aligned}$$

436

437 After integrating by parts, one obtains

$$S_{m} \int_{0}^{T} \int_{0}^{L_{m}} \frac{\partial \delta C_{m}}{\partial t} P_{m}(x,t) dx dt = S_{m} \int_{0}^{L_{m}} P_{m}(x,t=T) \delta C_{m}(x,t=T) dx$$
$$-S_{m} \int_{0}^{L_{m}} P_{m}(x,t=0) \delta C_{m}(x,t=0) dx$$
$$-S_{m} \int_{0}^{T} \int_{0}^{L_{m}} \frac{\partial P_{m}}{\partial t} \delta C_{m}(x,t) dx dt$$
$$S_{m} \int_{0}^{T} \int_{0}^{L_{m}} v_{m} \frac{\partial \delta C_{m}}{\partial x} P_{m}(x,t) dx dt = S_{m} v_{m} \int_{0}^{T} P_{m}(x=L_{m},t) \delta C_{m}(x=L_{m},t) dt$$
$$-S_{m} v_{m} \int_{0}^{T} P_{m}(x=0,t) \delta C_{m}(x=0,t) dt$$
$$-S_{m} \int_{0}^{T} \int_{0}^{L_{m}} v_{m} \frac{\partial P_{m}}{\partial x} \delta C_{m}(x,t) dx dt$$
(A.5)

From Eqs. (A.4) and (A.5), we deduce the system of equations associated with the stationarity of the Lagrangian according to the concentration field 440 **C**.

$$\begin{split} &-\frac{\partial P_m}{\partial t} - v_m \frac{\partial P_m}{\partial x} + k_m P_m = \frac{a_m}{S_m} (C(x_m, t) - C_m^{mes}(t)) \delta(x - x_m), \text{ in } [0, L_m] \times [0, T], \ m \in \{1, 2, 3\} \\ &P_m(x, t = T) = 0, \qquad \text{ in } [0, L_m], \ m \in \{1, 2, 3\} \\ &S_1 v_1 P_1(x = L, t) = S_2 \lambda_2 + S_3 \lambda_3, \qquad \text{ in } [0, T] \\ &P_m(x = L_m, t) = 0, \qquad \text{ in } [0, L_m], \ m \in \{2, 3\} \\ &\mu_m = P_m(x, t = 0), \qquad \text{ in } [0, L_m], \ m \in \{1, 2, 3\} \\ &\lambda_m = v_m P_m(x = 0, t), \qquad \text{ in } [0, T], \ m \in \{1, 2, 3\} \end{split}$$
(A.6)

441

In Eq. (A.6), note that the Lagrange multiplier  $P_m$  corresponds to the adjoint state. Thus, the equations of the adjoint are given by

$$\begin{aligned} &-\frac{\partial P_m}{\partial t} - v_m \frac{\partial P_m}{\partial x} + k_m P_m = \frac{a_m}{S_m} (C(x_m, t) - C_m^{mes}(t)) \delta(x - x_m), & \text{in } [0, L_m] \times [0, T], \ m \in \{1, 2, 3\} \\ &P_m(x, t = T) = 0, & \text{in } [0, L_m], \ m \in \{1, 2, 3\} \\ &S_1 v_1 P_1(x = L_1, t) = S_2 v_2 P_2(x = 0, t) + S_3 v_3 P_3(x = 0, t), & \text{in } [0, T] \\ &P_m(x = L, t) = 0, & \text{in } [0, L_m], \ m \in \{2, 3\} \end{aligned}$$
(A.7)

444

From Eq. (A.7), we observe that the adjoint problem of an advection-445 reaction problem through a divergent node corresponds to an advection-446 reaction problem through a convergent node. In contrast to the direct prob-447 lem, the adjoint problem is backward in time, *i.e.*, it has a final condition, 448 and the flow is reversed in the adjoint problem, *i.e.*, the velocity  $v_m$  is re-449 placed by  $(-v_m)$ . The source term at the first line in Eq. (A.7) indicates 450 that chlorine is virtually injected at sensor location  $x_m$ , and its amplitude 451 is given by the data misfit. Hence, the chlorine retropropagates from the 452 sensor locations. 453

<sup>454</sup> In the same way, when considering an advection-reaction direct problem <sup>455</sup> through a convergent node, it can be shown that its adjoint problem begins as an advection-reaction problem backward in time through a divergentnode.

458

The functional gradient can be expressed using the adjoint state. It allows computation of the functional gradient at a lower computational cost than standard finite difference schemes. The analytical formula of the functional gradient can be obtained from the derivative of the Lagrangian according to the reaction coefficient  $k_m$ .

$$\boldsymbol{\nabla} \boldsymbol{J} = \begin{pmatrix} \frac{\partial J}{\partial k_1} \\ \frac{\partial J}{\partial k_2} \\ \frac{\partial J}{\partial k_3} \end{pmatrix} = \begin{pmatrix} \frac{\partial \mathcal{L}}{\partial k_1} \\ \frac{\partial \mathcal{L}}{\partial k_2} \\ \frac{\partial \mathcal{L}}{\partial k_3} \end{pmatrix} = \begin{pmatrix} -S_1 \int_0^T \int_0^{L_m} C_1 P_1 dx dt + \beta b(k_1 - k_1^0) \\ -S_2 \int_0^T \int_0^{L_m} C_2 P_2 dx dt + \beta b(k_2 - k_2^0) \\ -S_3 \int_0^T \int_0^{L_m} C_3 P_3 dx dt + \beta b(k_3 - k_3^0) \end{pmatrix}$$
(A.8)

### <sup>464</sup> Appendix B. Proof of Proposition 4.1 on a reduced water network

In the interests of simplifying notations, let us consider the reduced water network presented in Figure B.7. It is composed of the main elements of a water network, *i.e.*, water pipes, convergent nodes and divergent nodes. Following the methodology presented in Section 2.3 and illustrated in Appendix A, we can show that the adjoint problem associated with the reduced water network corresponds to

$$\begin{aligned} &-\frac{\partial P_m}{\partial t} - v_m \frac{\partial P_m}{\partial x} + k_m P_m = a_m f_m(x,t), & \text{in } [0,L_m] \times [0,T], \ m \in \{1,2,3,4\} \\ &P_m(x,t=T) = 0, & \text{in } [0,L_m], \ m \in \{1,2,3,4\} \\ &P_4(x=L_4,t) = 0, & \text{in } [0,T] \\ &P_m(x=L_m,t) = P_4(x=0,t), & \text{in } [0,T], \ m \in \{2,3\} \\ &S_1 v_1 P_1(x=L_1,t) = S_2 v_2 P_2(x=0,t) + S_3 v_3 P_3(x=0,t), & \text{in } [0,T] \end{aligned}$$
(B.1)

The function  $f_m(x,t)$  corresponds to the virtual injection of chlorine at the sensor location. It is defined by

$$f_m(x,t) = (C(x_m,t) - C_m^{mes}(t))\delta(x - x_m)/S_m.$$
 (B.2)



Figure B.7: Reduced water network: a) Flow direction b) Notations associated with the adjoint problem; in the advection-reaction adjoint problem, the flow is reversed;  $\{x_1, x_2, x_3, x_4\}$  denote the potential locations of chlorine sensors

473 When considering no reaction term in (B.1), we obtain the no-reaction

474 adjoint problem

$$-\frac{\partial \phi_m}{\partial t} - v_m \frac{\partial \phi_m}{\partial x} = a_m f_m(x, t), \quad \text{in } [0, L_m] \times [0, T], \ m \in \{1, 2, 3, 4\}$$
  

$$\phi_m(x, t = T) = 0, \quad \text{in } [0, L_m], \ m \in \{1, 2, 3, 4\}$$
  

$$\phi_4(x = L_4, t) = 0, \quad \text{in } [0, T]$$
  

$$\phi_m(x = L_m, t) = \phi_4(x = 0, t), \quad \text{in } [0, T], \ m \in \{2, 3\}$$
  

$$S_1 v_1 \phi_1(x = L_1, t) = S_2 v_2 \phi_2(x = 0, t) + S_3 v_3 \phi_3(x = 0, t), \quad \text{in } [0, T]$$
  
(B.3)

Finally, if a unitary constant chlorine amplitude in time is injected at the  $i^{th}$  sensor location rather than an amplitude given by the data misfit as in problems (B.1) and (B.3), one obtains the modified adjoint problem associated with sensor i

$$-\frac{\partial \phi_{m,i}^{*}}{\partial t} - v_{m} \frac{\partial \phi_{m,i}^{*}}{\partial x} = H(t)\delta(x - x_{i}), \quad \text{in } [0, L_{m}] \times [0, T], \ m = i$$
  

$$-\frac{\partial \phi_{m,i}^{*}}{\partial t} - v_{m} \frac{\partial \phi_{m,i}^{*}}{\partial x} = 0, \quad \text{in } [0, L_{m}] \times [0, T], \ m \neq i$$
  

$$\phi_{m,i}^{*}(x, t = T) = 0, \quad \text{in } [0, L_{m}], \ m \in \{1, 2, 3, 4\}$$
  

$$\phi_{4,i}^{*}(x = L_{4}, t) = 0, \quad \text{in } [0, L_{4}]$$
  

$$\phi_{m,i}^{*}(x = L_{m}, t) = \phi_{4,i}^{*}(x = 0, t), \quad \text{in } [0, L_{m}], \ m \in \{2, 3\}$$
  

$$S_{1}v_{1}\phi_{1,i}^{*}(x = L_{1}, t) = S_{2}v_{2}\phi_{2,i}^{*}(x = 0, t) + S_{3}v_{3}\phi_{3,i}^{*}(x = 0, t), \quad \text{in } [0, T]$$
  
(B.4)

This problem does not depend on reaction phenomena. Moreover, from the solution of Eq. (B.4), we can deduce the general solution of the modified adjoint problem (4) involved in Proposition 3.1

$$\phi_m^*(x,t) = \sum_{i=1}^N a_i \phi_{m,i}^*(x,t), \qquad \text{in } [0,L_m] \times [0,T], \ m \in \{1,2,3,4\} \quad (B.5)$$

On the one hand, the solution  $\phi_m(x,t)$  of the no-reaction adjoint problem (B.3) can be obtained from the modified adjoint solution  $\phi_{m,i}^*$  (also called <sup>484</sup> "Green function"). Indeed, noting that

$$f_m(t) = [f_m * \delta](t) = [f'_m * H](t)$$
(B.6)

where \* denotes the convolution product in time and H is the Heaviside function, one has

$$\phi_m(x,t) = \sum_{i=1}^N a_i [f'_i * \phi^*_{m,i}](x,t), \ m \in \{1,2,3,4\}$$
(B.7)

487 where  $f'_m$  is the derivative of the function  $f_m$  according to the time variable.

488 We recall that the flow 
$$(-v_m)$$
 is considered stationary in the present proof

$$\phi_m^*(x,t) = 0 \text{ in } [0,L_m] \times [0,T] \Leftrightarrow \phi_{m,i}^*(x,t) = 0 \text{ in } [0,L_m] \times [0,T] \Rightarrow \phi_m(x,t) = 0 \text{ in } [0,L_m] \times [0,T]$$
(B.8)

490

On the other hand, subtracting Eq. (B.1) from (B.3) and noting 
$$\delta P_m(x,t) =$$

492  $P_m(x,t) - \phi_m(x,t)$ , one has

493 From Eq. (B.9), we deduce that

$$\begin{cases} \phi_m(x,t) = 0, & \text{in } [0,L_m] \times [0,T] \\ \delta P_m(x = L_m,t) = 0, & \text{in } [0,T] \\ \delta P_m(x,t = T) = 0, & \text{in } [0,L_m] \end{cases}$$
(B.10)

494 implies

$$\delta P_m(x,t) = 0$$
, in  $[0, L_m] \times [0, T] \Rightarrow P_m(x,t) = 0$ , in  $[0, L_m] \times [0, T]$  (B.11)

495

As the gradient  $\nabla J$  can be expressed using the adjoint solution  $P_m$ 

$$\frac{\partial J}{\partial k_m}\delta k_m = -S_m \int_0^T \int_0^{L_m} C_m P_m \delta k_m dx dt + \beta b(k_m - k_m^0)\delta k_m \qquad (B.12)$$

it can be noted that a null adjoint solution  $P_m(x,t)$  in the pipe m on the entire time interval [0,T] implies that the first term of the gradient component according to the reaction coefficient  $k_m$  vanishes. Hence, the reaction coefficient in pipe m cannot be updated.

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