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## 5000 years of lacustrine ecosystem changes from Lake Petit (Southern Alps, 2200 m a.s.l.): Regime shift and resilience of algal communities

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#### Abstract

Sediments from Lake Petit (2200 m a.s.l., Southern Alps) are particularly relevant for analysis of coupled landscape palaeoecology and palaeolimnology. Diatom assemblages, organic matter composition of sediments (total nitrogen and organic carbon) and *Pediastrum boryanum* concentrations were obtained from a 144-cm-long core, enabling the reconstruction of the aquatic ecosystem over nearly the last 5000 cal. BP. From 4800 to 4300 cal. BP, Lake Petit was a stable diatom-productive water body dominated by alkaliphilous diatoms (*Staurosirella pinnata*). During this period, nutrients and cations were supplied by the chemical weathering of podzols that developed under conifer woodlands. This overall stability was suddenly interrupted at 4200 cal. BP by a major detrital pulse that was probably climate linked (4200 cal. BP event) and that triggered a drop in diatom productivity and diversity. From 4100 to 2400 cal. BP, diatom productivity progressively decreased, whereas *Pediastrum* developed. Diatom assemblages were more diversified (predominance of *Pseudostaurosira robusta*, *P. brevistriata* and *P. pseudoconstruens*) and reflected a regime of continuous erosion, whereas slopes were colonised by grazed grasslands. Finally, from 2400 cal. BP to the present day, diatom assemblages reveal a slight acidification and nutrient enrichment of waters concomitant with increasing human pressure in the catchment. These results demonstrate the close links between ecosystems and the ready propagation of disturbances throughout watersheds that might lead to abrupt regime shifts in such alpine environments.

#### Keywords

alpine lake, catchment-lake interaction, diatoms, Holocene, reference conditions, 4200 cal. BP event

### Introduction

Determining baseline conditions for freshwater ecosystems is of crucial interest for wildlife conservation. The European Union Water Framework Directive has undertaken actions aimed at restoring a 'good ecological status' in all water bodies, raising the tricky question of the identification of target conditions for restoration (Bennion and Battarbee, 2007; Bennion et al., 2011). These reference conditions may be defined with regard to 'deviation' from 'pre-anthropogenic' conditions of species communities, ecosystem productivity and physico-chemical parameters, but they usually rely on instrumental data and short-term studies covering less than the last two centuries (Battarbee et al., 2011; Bennion and Simpson, 2011; Guilizzoni et al., 2011). However, it is now well known that environments have been deeply impacted by human activities at least at the millennial timescale (Berglund, 2011). Lastly, the definition of a single limit for anthropogenic imprint in the Holocene history of environments remains debatable since human activities and impacts might have occurred asynchronously between sites (Balter, 2013; Steffen et al., 2011).

Beyond the definition of a pre-anthropogenic state, palaeolimnological studies represent a valuable means of obtaining long-term insight into the mechanisms that lead to present day ecosystems (Bennion et al., 2011; Smol, 1992) and into the boundary conditions within which an ecosystem may change (Battarbee et al., 2005). Multi-proxy investigations on lake sediments enable tracking back the dynamics of lacustrine and local vegetation ecosystems over millennia as well as their response to human impact and climate change (Birks and Birks, 2006; Dearing, 2013; Randsalu-Wendrup et al., 2014). Such external forcing may lead to a variety of lake responses, ranging from gradual changes to sudden and drastic switches (Carstensen et al., 2013; Scheffer et al., 1993; Scheffer et al., 2001; Wang et al., 2012), depending on the characteristics of the terrestrial and lacustrine ecosystem. In this context,

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**Figure 1.** Location map of the study site and main geomorphological characteristics of the catchment and bathymetry of Lake Petit. I: glacial cirque; 2: glacial step; 3: moraine; 4: polished bedrock; 5: active debris slope; 6: small dam built in 1947; 7: ancient pastoral enclosures (Suméra et al., 2008).

understanding long-term coupled lake-landscape interactions is decisive in distinguishing factors that trigger lake trajectories (Fritz, 2008; Lotter and Birks, 2003). Regarding such coupled relationships, palaeolimnological studies, based, in particular, on diatom analysis, can provide powerful results on reconstructed hydrochemical variations (Alefs and Müller, 1999; Schmidt et al., 2004). Diatom species are closely related to environmental parameters and respond quickly to ecological perturbations (Bigler et al., 2006; Dam et al., 1994; Smol and Stoermer, 2010).

Alpine lakes are a prime source of insight into long-term lake trajectories because they are relatively simple systems characterised by short aquatic trophic chains and small hydrographical networks. Moreover, these systems are commonly exposed to extreme environmental conditions (Beniston, 2003, 2006), an annually long period of ice cover, persistence of low temperatures and low nutrient availability, all of which render their aquatic communities particularly sensitive to landscape and climatic changes (Druart et al., 1999; Quesada et al., 2006; Rühland et al., 2008). In these lakes, changes in air temperature, precipitation and wind regime can lead to profound hydrological and physicochemical perturbations (Battarbee et al., 2002; Lotter et al., 1997; Sommaruga-Wögrath et al., 1997; Thompson et al., 2005). Mountain-related anthropogenic activities are also a potential source of disturbance of lacustrine ecosystems mainly through changes in nutrient inputs and light availability (Heiri and Lotter, 2003; Lotter and Birks, 2003). In the Southern Alps, human presence is ancient and has been dated by archaeological and charcoal studies from the Bronze Age (ca. 5000-4000 cal. BP; De Lumley and Echassoux, 2009; Walsh et al., 2007) onwards. Forest clearance linked to increasing pastoralism since ca. 3000 cal. BP has been evidenced (De Beaulieu, 1977; Finsinger, 2001; Ortu et al., 2003; Ponel et al., 2011; Walsh et al., 2007). However, few palaeoenvironmental studies have been carried out on the responses to longterm forcing of lakes in the Southern Alps (Gandouin and Franquet, 2002; Larocque and Finsinger, 2008), which differ from those in the Northern Alps in terms, notably because of the Mediterranean climatic influence to which the former are subject.

Sediments from Lake Petit (2200 m a.s.l., Southern French Alps) have turned out to be particularly relevant in a spate of recent studies on the coupling of landscape palaeoecology and palaeolimnology. This relevance stems from (1) exceptional richness in preserved bio-indicators, especially algae (diatoms and *Pediastrum*); (2) a well-constrained archaeological context (Morin and Rosenthal, 2002; Pagès, 2009; Suméra et al., 2008); and (3) an identified tipping point in the terrestrial ecosystem history at 4200 cal. BP from multi-proxy studies of the ambient vegetation and soil dynamics (Brisset et al., 2013). This paper focuses essentially on the lacustrine ecosystem, complementing these previous studies.

By coupling palaeoecological analysis (counting of diatoms and *Pediastrum* algae) and geochemical measurements (carbon, nitrogen and biogenic silica), backed by the palaeoenvironmental studies of Brisset et al. (2012, 2013), this paper aims at (1) reconstructing changes in the lacustrine ecosystem of Lake Petit over the last 4800 years, (2) analysing relationships between the lacustrine and ambient terrestrial ecosystems, particularly through the effect of soil erosion on lacustrine nutrient input, (3) disentangling the influence of human impact and climate variability and (4) evaluating the resilience of the lacustrine ecosystem following perturbations.

## Study site

### Setting

Lake Petit (2200 m a.s.l.; 44°06'789"N; 7°11'342"E) is a small circular body of water 150 m in diameter located in the Southern French Alps about 60 km from the Mediterranean Sea and is the lowest in elevation of a chain of five lakes partly inherited from glacier retreat (Figure 1). The lakes are connected in spring by ice meltwater but remain unconnected the rest of the year. The lake catchment (area: 6 km<sup>2</sup>), culminating at 2600 m a.s.l., is composed of crystalline bedrock (gneiss and migmatites) and is largely covered by alpine meadows; the upper treeline (Larix sp.) is located at about 2100 m a.s.l. Macrophyte vegetation composed of Ranunculus aquatilis colonises the fringes of the lake (Figure 1). The lake surface is usually frozen from October to April. The depth of Lake Petit is up to 7 m in the wake of snow-melt in late spring and decreases to 6.5 m at the end of summer. During summer, continuous water inflow is assured by further melting of residual snow trapped in surrounding moraine and scree. The mean lake-level has been artificially raised by about 1.5 m by the construction of a small dam in 1947 (Beniamino, 2006). This dam is now in ruins, allowing for a nearly permanent water outflow. Measurements of the physico-chemical properties of the lake waters were conducted in spring (17 May 2012). These waters are characterised by 66% of oxygen saturation, a pH of 8 (at a water temperature of 7°C) and are rich in calcium ( $[Ca^{2+}]=14.1\pm0.3$  mg/L), sulphates  $([SO_4^{2-}]=5.2\pm0.1 \text{ mg/L})$  and silica  $([Si^{4+}]=2.7\pm0.1 \text{ mg/L})$ .

Phosphates were absent ([  $PO_4^{3-}$ ]<0.03 mg/L). The measured nitrate concentration (  $NO_3^{-}$ ) was 0.07±0.01 mg/L.

Alpine and Mediterranean influences generate a climate marked by mild winters and dry summers. Mean annual temperature at 1800 m a.s.l. is 5°C, varying from 0.3°C in winter to 9.9°C in summer (Durand et al., 2009), with rainfall occurring mainly in spring and autumn. Mean annual precipitation is 1340 mm at 1800 m a.s.l. Snow depths in winter are relatively important (150–250 cm at 2400 m a.s.l.) due to moisture from the nearby Mediterranean sea. Snow cover duration is about 185 days at 2100 m a.s.l. (Durand et al., 2009).

#### Previous multidisciplinary studies

Sediments of Lake Petit have been previously studied, allowing for a detailed reconstruction of the past terrestrial environment through combined sedimentological, geochemical and palynological approaches (Brisset et al., 2012, 2013). These studies highlighted the prevalence of complex interactions between vegetation and pedogenetic processes in the lake catchment over the last five millennia. The results show, from 4800 to 4200 cal. BP, a progressive development of podzols under coniferous forest conditions and overall slope stability. Degradation of podzols occurred abruptly at 4200 cal. BP, marked by a major detrital pulse in the lake sediment archives. Since 4200 cal. BP, moderately weathered cambisols and alpine meadows associated with ruderal-anthropogenic plants in the catchment have constituted a stable state characterised by uninterrupted terrigenous inputs into the lake.

Archaeological investigations in the catchment have also evidenced the presence of (1) 30 pastoral enclosures, mainly dating to the late Middle Ages (1000–600 cal. BP), but one of which, located upstream of the lake, has been dated from 2140 to 2000 cal. BP (Suméra et al., 2008; Figure 1) and (2) several sites of iron ore extraction and smelting that were in activity from 2200 to 1600 cal. BP (Morin and Rosenthal, 2002; Pagès, 2009). The presence of livestock is abundantly attested between 1300 and 400 cal. BP.

### Materials and methods

Sediment cores were taken in 2009 (PET09P2) and 2012 (PET12P3) in the deepest part of the lake using a UWITEC gravity corer. Correlations between core PET09P2 (144 cm long) and core PET12P3 (75 cm long) are based on lithological markers (colour and texture). Sediments are composed of homogenous yellow to greenish diatomite with millimetre-thick diffuse brownish diatomite-clay laminations. Sedimentological (dry density); geochemical (biogenic silica, carbon and nitrogen); and radiochronological analyses, as well as Pediastrum counts (identified during pollen analysis), were carried out on core PET09P2. The age-depth model is based on short-lived <sup>210</sup>Pb and <sup>137</sup>Cs radionuclides and seven 14C ages obtained from terrestrial macro-remains (see Supplementary Material 1, available online, and Brisset et al. (2013) for further details). This study was conducted on core PET09P2, which covers the last 4800 cal. BP, complemented by some samples from core PET12P3.

#### Diatom and Pediastrum analyses

Diatom analyses were conducted on 45 samples in PET09P2. In order to improve resolution for the uppermost 30 cm, we analysed a further eight samples from core PET12P3. A total of 53 diatom samples (1 cm<sup>3</sup>) were taken in both cores and weighed after drying at 50°C (Supplementary Material 1, available online). Diatom taxonomic and counting analyses were conducted on 0.3g samples and treated using standard procedures (1:1 mixture of  $H_2O_2$ :water, 1:1 mixture of HCl:water and repeatedly rinsed in distilled water; slides were made using Naphrax). For each sample, at least 400 diatom valves were identified and counted using a Nikon Eclipse 80i light microscope (differential interference contrast optics,  $1000 \times$  magnification, numerical aperture (NA)=1.25). Diatom valves were counted in 10 randomly chosen fields to calculate diatom concentrations (number/g dry sediment).

Specimens were identified to their lowest taxonomic level (e.g. variety) following the species concept of Krammer and Lange-Bertalot (1986–1991). Several identified genera and species have been re-assessed using the classification by Round et al. (1990), and synonyms were introduced. Diatom identification was also based on Marciniak (1988 [1986]) and Fallu et al. (2000).

The diatom species percentage diagram was plotted using C2 software (Juggins, 2007). Diatom assemblage zones (DAZs) were defined according to a constrained hierarchical clustering using the CONISS algorithm (Grimm, 1987). The number of significant zones was determined using the 'broken-stick' model (Bennett, 1996) based on the R package 'rioja' (Juggins, 2012). Diatoms were also classified into ecological groups according to their specific preferences in pH and their trophic requirements (Krammer and Lange-Bertalot, 1986–1991; Dam et al., 1994). Species diversity was estimated through species richness and equitability calculated using the Shannon diversity index as follows:

$$Equitability(E) = \frac{Shannon \, diversity \, index(H)}{\ln(species \, richness(S))}$$

In order to compare species richness for samples of different counting sizes, the count sums were standardised using rarefaction analysis (Birks and Line, 1992). Expected species richness E(Sn) was calculated using the R package vegan (Oksanen et al., 2007, retaining a common count sum of 400 ( $E_{s400}$ )).

*Pediastrum coenobia* were extracted from sediments following Bennett (1990) and were identified following Komárek and Jankovská (2001) in 144 samples collected in constant volume (1 cm<sup>3</sup>) at each centimetre of core PET09P2. *Pediastrum* concentrations were expressed in number per millilitre.

## C, N analysis, biogenic silica fluxes and PCA ordination

Analysis of total carbon (TC) and total nitrogen (TN) in the sediment was conducted on 50 samples. The percentages of TC and TN of each 10 mg of dry sediment sample (after crushing and conditioning into tin capsules) were measured by gas chromatography using a FLASH 2000 series Thermo Fisher. Percentages of TC correspond to total organic carbon (TOC) since inorganic carbon is totally absent in the sedimentary record (Brisset et al., 2013). Differences in TOC/TN ratios (abbreviated as C/N ratio), expressed as atomic or weight ratios, can be interpreted as changes in the relative contribution of terrigenous and lacustrine organic matter (Meyers and Teranes, 2001).

Diatoms (D) represent the major compound of biogenic silica in the sedimentary record. Only few cysts of Chrysophyceae (C) were identified during counting (C/D rate=0.01). Biogenic silica (%) was calculated at high resolution (0.2 cm) by Fourier transform infrared spectroscopy (FTIR) and calibrated with measures of wet-alkaline leaching on 28 discrete samples (see details in Brisset et al., 2013). Diatom productivity can be estimated through biogenic silica fluxes (SiO<sub>2biog</sub> flux) calculated as follows:

$$SiO_{2biog} flux (g / cm2 / yr) = dry density (g / cm3) \times sedimentation rate (cm / yr) \times \%SiO_{2biog}$$

Finally, we used principal components analysis (PCA) to evidence, through time, interactions between aquatic communities (diatom species composition, *Pediastrum* concentrations) and organic matter composition (TOC, TN and C/N ratios). For the PCA ordination, 14 dominant diatom species having a relative abundance higher than 5% in the assemblages were selected. The abundance of the detrital fraction (DF) within bulk sediment (% dry mass DF) was integrated in the PCA as a variable representative of soil erosion. The DF represents the addition of major detrital elements (for more details, see Brisset et al., 2013). The 3-cm running mean concentrations (number/cm<sup>3</sup>) of *Pediastrum* were calculated to smooth out high-frequency variability.

### Results

#### Diatom species composition and Pediastrum

A total of 39 genera and 111 taxa were identified. Only species with a relative abundance of >2% are shown in the simplified diagram in Figure 2. Diatom assemblages over the last five millennia are mainly dominated by species of Fragilariaceae. From the base of the core (144 cm) to 105 cm, DAZ 1 is strongly dominated by Staurosirella pinnata, which represents, on average, 74% of total diatoms. From 144 to 116 cm, this species is associated with Pseudostaurosira robusta, which punctually peaks at 25%, Pseudostaurosira pseudoconstruens, Encyonema minutum and Pseudostaurosira brevistriata. From 116 to 105 cm, these species virtually disappear, and S. pinnata becomes very largely dominant, reaching 90% of the total assemblage. From 105 cm to the top of the core, DAZ 2 exhibits an abrupt decline in S. pinnata, replaced by P. robusta. DAZ 2 can be further subdivided into three sub-units according to diatom species composition. DAZ 2a (105-66 cm) is characterised by important concentrations of P. robusta, P. brevistriata, P. pseudoconstruens and Staurosirella lapponica. S. lapponica then progressively decreases in this unit. DAZ 2b (66–22 cm) is defined by a decline in P. robusta, whereas P. pseudoconstruens, E. minutum, Sellaphora pupula, Staurosira construens var.venter, Navicula radiosa, Amphora libyca and Achnanthes oestrupii increase. Finally, DAZ\_2c (22-0 cm) is characterised by S. pinnata, whereas P. robusta decreases to 20%. P. pseudoconstruens, P. microstriata, S. construens var. venter and Opephora mutabilis are also associated with S. pinnata in DAZ 2c.

*Pediastrum* remains have been identified as *P. boryanum* var. *forcipatum*. The mean *Pediastrum* concentration along the core is 1307 nb/mL. The peak concentration is 3700 nb/mL at 100 cm depth. Mean *Pediastrum* concentration is 810 nb/mL in DAZ\_1 and 1450 nb/mL in DAZ\_2 (Figure 2).

### Diatom ecological affinities and diversity

Diatom assemblages are composed of benthic organisms living fixed to different supports (Figure 2). Fragilariaceae are usually considered to be broadly benthic, with the capacity of thriving in the water column during re-suspension of superficial sediment. The dominant species identified in this record (e.g. *Staurosirella pinnata, Pseudostaurosira robusta, Pseudostaurosira pseudoconstruens* and *Pseudostaurosira brevistriata*) are typical of alkaliphilous and oligotrophic environments but are tolerant of trophic variations (Figure 3).

DAZ\_1 is dominated by tychoplanktonic and alkaliphilous species (e.g. *S. pinnata*). Epipelic diatoms (that can grow on mud) represented by *S. lapponica*, *S. pupula* and *A. libyca* increase in DAZ\_2a. Fragilariaceae decreases at the beginning of DAZ\_2b while meso-eutrophic (e.g. *S. pupula*, *N. radiosa* and *S. construens* var. *venter*) and circumneutral species (e.g. *E. minutum*, *A. libyca* and *S. pupula*) increase and peak at about 10% of total

diatoms. Epilithic and epiphytic diatoms (e.g. *Cymbella* spp., *Encyonema* spp. and *Gomphonema* spp.), living fixed on macrophytes and rocks, increase in this unit and in DAZ\_2c (mean value of 6.3%). These results suggest changes in pH, nutrient enrichment and habitat availability for diatom species throughout the record.

Diatom diversity was estimated by expected diatom richness (not shown) and equitability (Figure 3). Expected species richness values range from 17 to 54 and equitability values from 0.16 to 0.75. In DAZ\_1, assemblages are poorly diversified (dominated by *S. pinnata*), and the lowest equitability values (less than 0.2) occur from 115 to 105 cm. DAZ\_2a is characterised by an important development of *P. robusta, P. brevistriata, P. pseudoconstruens* and *Staurosirella lapponica*, which entail an increase in diversity (equitability>0.5). Finally, equitability peaks at 0.8 in DAZ\_2b and DAZ\_2c.

# C, N analysis, biogenic silica fluxes and valve concentrations

TOC values range from 16% to 6% (Figure 3). From 144 to 115 cm. values of TOC attain a maximum of ca. 13%. From 115 to 107 cm, the TOC curve decreases abruptly to the lowest value (6%). From 107 to 0 cm, TOC values are about 8% except between 65 and 25 cm where values sporadically peak to 10%. Mean TN is 0.9%. From 144 to 107 cm, TOC and TN are covariant. In the upper part of the core, TN is maximal between 70 and 25 cm (~0.4%). Mean C/N is 11.5 from 144 to 107 cm and 9.2 from 107 to 0 cm. Maximal biogenic silica values occur from 144 to 120 cm (0.15 g/cm2/yr; Figure 3). Biogenic silica flux decreases to a minimum at 110 cm (0.05 g/cm2/yr) and values remain low in DAZ 2, except from 50 to 40 cm and in the upper 5 cm. Valve concentrations range from 1.27 to  $11.3 \times 10^9$  valves/g dry sediment (Figure 3). Mean concentrations in DAZ 1 are  $7.9 \times 10^9$  valves/g of dry sediment and  $3.5 \times 10^9$  valves/g of dry sediment in DAZ 2. Concentration peaks at 11.3×109 valves/g dry sediment at 37 cm depth.

### PCA ordination and multi-proxy zonation

The total percentage of variation explained by the first two principal components (PC1 and PC2) is 38%, with eigenvalues of  $\lambda_1 = 0.23$  and  $\lambda_2 = 0.15$ , respectively (Figure 4). Variables TOC, C/N and S. pinnata are strongly positively correlated with PC1, whereas P. robusta, A. oestrupii, S. pupula, P. pseudoconstruens and Pediastrum are the most negatively correlated with PC1 (p < 0.01). PC1 separates samples up to and after 4100 cal. BP (corresponding to the limit between DAZ 1 and DAZ 2). Variables positively correlated with PC2 are TN, TOC, A. libyca, N. radiosa, E. minutum, S. pupula, P. brevistriata var. papillosa and P. brevistriata. The DF variable is negatively correlated with PC2 (p < 0.01). The strongest negative weight of DF is between 120 and 105 cm and from 25 to 0 cm. Accordingly, five zones corresponding to the time periods delineating the palaeoenvironmental history of Lake Petit have been defined as follows: 4800-4300, 4300-4100, 4100-2400, 2400-800 cal. BP and from 800 cal. BP to the present.

### Palaeoenvironmental history

Over the last 4800 years, Lake Petit has been dominated by small tychoplanktonic alkaliphilous Fragilariaceae. This group comprises taxa having a wide range of ecological tolerances. They can be found as pioneer assemblages in lagoons progressively isolated from the sea (Stabell, 1985) as well as in new lakes following the retreat of glaciers (Haworth, 1976), in contexts marked by rapid







Figure 3. Detrital and biogenic silica fraction abundance (%; from Brisset et al., 2013); flux of SiO<sub>2biog</sub> (g/cm<sup>2</sup>/yr); diatom valve concentrations (valves/g dry sediment); organic matter (OM) composition (total nitrogen (% TN), total organic carbon (% TOC) and C/N (TOC/TN)); classes of diatom ecological affinities (pH and trophic state); and diversity (equitability). The zoning refers to cluster analysis of diatom assemblages (see Figure 2).



according to core depth and (c) total nitrogen concentration (% TN) plotted against total organic carbon concentration (% TOC). Variables and samples having positive explained variances are shaded in grey (PC1) Figure 4. Principal component analysis of diatom frequency assemblages and environmental proxies: (a) variable loadings on the first and second principal components (PCI and PC2), (b) scores of PCI and PC2 microstriata (Pmi), P. pseudoconstruens (Ppse), P. robusta (Pro), S. pupula (Spu), S. lapponica (Sla), S. pinnata (Spi), E. minutum (Emi), S. construens var. venter (Sco), N. radiosa (Nra), Fragilaria vaucheriae var. continua (Frvc), P. and black (PC2). Variables are as follows: total organic carbon (TOC), total nitrogen (TN), total organic carbon/total nitrogen (C/N), detrital fraction (DF), Pediastrum (P), A. oestrupii (Aoe), P. brevistriata (Pbr), P. brevistriata var. papillosa (Pbrp) and A. lybica (Aly).



**Figure 5.** Lacustrine ((a)% *S. pinnata*, (b) % *P. robusta*, (c) % meso-eutrophic diatoms and (d) biogenic silica flux; this study) and terrestrial ((e) % detrital fraction, (f) % pollen; Brisset et al., 2013) ecosystem proxies compared to hydrological conditions at (g) Lake Saint-Léger (Digerfeldt et al., 1997), (h) Lake Cerin (Magny et al., 2012a) and (i) number of <sup>14</sup>C dates (2σ) of pastorals structures in archaeological stratigraphies in the Southern Alps (Walsh et al., 2007).

changes in physico-chemical conditions. Fragilariaceae can also be abundant in nutrient-rich lakes (Bennion et al., 2001) and in cold, nutrient-poor high mountain lakes with short growing seasons (Catalan et al., 2002; Wögrath and Psenner, 1995). This family is commonly observed in alpine (e.g. Hustedt, 1943; Lotter et al., 1999; Schmidt et al., 2004), arctic and subarctic lakes (e.g. Pienitz et al., 1995; Smol, 1988; Westover et al., 2006). At the species level, differences have also been found in these taxa in response to environmental change, and Fragilariaceae have revealed a good potential for palaeoenvironmental studies (Finkelstein and Gajewski, 2008; Schmidt et al., 2004). These observations are confirmed by the record from Lake Petit where species of Fragilariaceae (and other sub-dominant taxa) highlight a major shift that undoubtedly represents a significant change in the lacustrine ecosystem.

### From 4800 to 4300 cal. BP:A long-lasting lakecatchment equilibrium

Maximal values of biogenic silica flux and concentrations indicate that a major part of the sedimentation was assured by the autochthonous production of diatoms. The high TN and TOC values suggest that organic matter is abundant during this period, and a mean C/N ratio of 11.5 can be interpreted as a mixture of lacustrine algae and terrestrial vascular plants (Meyers and Teranes, 2001). Diatom assemblages are poorly diversified and dominated by *S. pinnata* (Figure 5a) typically observed in alkaline and welloxygenated lakes (Dam et al., 1994; Schmidt et al., 2004). Additionally during this period, (1) minerogenic inputs were low, (2) phosphorus (P) and calcium were preferentially trapped in the lake and (3) the watershed was dominated by chemically weathered podzols under open mixed pine–larch woodlands (Brisset et al., 2013).

The concomitance of the maximal diatom flux with the preferential trapping of organic C, N and P suggests that uptake of nutrients by diatoms was significant and stored via sedimentation processes. These nutrients are determinant in algal growth (Schindler et al., 2008; Vitousek and Howarth, 1991). The results regarding this period, thus, suggest that these nutrients were readily available in their soluble forms. Such nutrient supply might be linked with (1) leaching of dissolved minerals from soils (Filippelli et al., 2006; Likens et al., 1996; McLauchlan et al., 2013; Righi et al., 1999) and (2) terrestrial organic matter input from vegetation. Organic matter enrichment is susceptible to provide nutrients for primary producers through lake degradation processes (Vinebrooke and Leavitt, 1998).

This period is also dominated by alkaliphilous diatom species which might be surprising in crystalline catchments covered by acid soils that are more sensitive to water acidification (Reuss et al., 1987). However, as shown for a large set of Austrian alpine lakes, even in exclusively crystalline catchments, chemical weathering of developed soils might preserve lakes from acidification (Kamenik et al., 2001). Similar lake–catchment processes have been described by Schmidt et al. (2004) in a high-altitude lake of the eastern Alps (Oberer Landschitzsee, Austria), that is, a decrease in water pH throughout the Holocene (but remaining higher than 7) through long-lasting cation losses during progressive development of soil under pine wood cover.

Moreover, chemical weathering rather than particulate runoff in the watershed is likely to have favoured clear waters. Thus, pedogenetic processes might have played a major role in both promoting high diatom productivity and maintaining the alkalinity of water, which suggests a close interdependence of processes occurring on slopes and in the lake.

The period 4800–4300 is also characterised by a relatively low diversity in diatom assemblages. The latter is highly dependent on habitat availability and on the frequency of ecological disturbances (Connell, 1978); therefore, a background of long-lasting favourable conditions may shape stable living aquatic communities comprising few species. In Lake Petit, the low diversity of diatom assemblages might be attributed to poor substrate heterogeneity for benthic diatoms in a context of very low environmental variability (in both frequency and intensity): thus, this relative stability supported the predominance of *S. pinnata* over the other diatom species.

# From 4300 to 4100 cal. BP:A switch to another ecosystem?

During this period, the lacustrine ecosystem switches irreversibly from one of steady low diversity and high diatom production dominated by *S. pinnata* prior to 4300 cal. BP, to a new, low diatom-productive but diversified state dominated by *P. robusta*, *P. brevistriata* and *P. pseudoconstruens* after 4100 cal. BP (Figure 5a, b and d).

This short transitional period occurred over a maximum of 300 years and is characterised by a drop in diatom diversity: *S. pinnata* is almost the only species that persists. This profound change in diatom communities is concomitant with the strongest detrital pulse occurring at 4200 cal. BP (Figure 5e). This pulse generated a dilution of the concentration in biogenic silica, and consequently, the flux of diatoms decreased drastically. This period is also marked by high values of C/N ratio (Figure 3), suggesting high input of both terrestrial minerogenic and organic matter.

This detrital pulse has been interpreted as the erosion of subsurface soil horizons, triggered by more intense and/or more frequent rainfall around 4200 cal. BP (Brisset et al., 2013). This massive increase in allochtonous inputs into Lake Petit necessarily entailed an increase in water turbidity further reinforced by re-suspension of sediments. It can be assumed that high concentrations of suspended mineral particles might, by limiting light penetration in the water column, have reduced photosynthetic activity, diatom development, while enhancing increased conductivity. The persistence of the predominance of *S. pinnata* clearly illustrates that monospecific diatom assemblages might occur in a context of frequent and intense ecological disturbances such as changes in light, nutrient resources and substrate availability (Hayashi, 2011).

# From 4100 to 2400 cal. BP:Towards a new equilibrium

In the wake of the detrital pulse, diatom productivity decreases significantly, while the aquatic ecosystem is still undergoing diversification, as illustrated by the increase in the equitability diversity index (Figure 3). The decrease in diatom production occurs synchronously with an increase in minerogenic detrital input. It may be inferred from this that the persistence of turbidity in the lake waters is likely to have reduced photosynthetic activity. Moreover, during this period, Lake Petit was surrounded by cambisol-type soils under more open vegetation (Brisset et al., 2013). As a result, soluble elements and organic terrestrial matter inputs (and therefore nutrient inputs) decreased significantly from 4100 cal. BP. The decrease in diatom productivity over this period could also be related to this decrease in nutrient inputs.

It must be noted that such interactions between nutrient lowering in lakes and the development of open landscapes have been previously evidenced in the multi-proxy study of Lake Unterer Landschitzsee in the Central Austrian Alps (Schmidt et al., 2002). These authors argued that vegetation changes from forest to grassland and the disappearance of a humus layer might have led to a reduction of organic sourcing of lake fertilisation.

Interestingly, these changes in diatom productivity also correspond to the development of the green alga Pediastrum boryanum, which is typically dominant in subarctic lakes and related to low dissolved organic carbon content (Weckström et al., 2010). Differences in competitive abilities for nutrient resources (e.g. light, Si, N and P) classically control species composition, diversity and succession of algal communities (Tilman, 1986). Thus, the coexistence of diatoms and Pediastrum in Lake Petit might have increased competition for resources, especially for light and nutrients. Lastly, in Lake Sägistalsee (1935 m a.s.l.) in the Swiss Alps, Hofmann (2003) noted an increase in cladoceran diversity related to a large range of shoreline habitats during a period of more open vegetation. If we assume that a similar diversification of primary consumers (such as cladocerans/copepods) occurred in Lake Petit, zooplankton grazing of the 'small-sized' Fragilaria might have been an additional factor in lowering diatom productivity.

The increase in the equitability diversity index (Figure 3) indicates that diatom assemblages were more diversified, leading to the assumption that a larger variety of habitats were available. Diatoms may have taken advantage of possible substrate variability, such as variations in substrate surface roughness, or the existence of gravel beaches or zones of macroalgae surrounding the lake (Downes et al., 1998). Diversity might also have been enhanced by further perturbations (e.g. re-suspension of sediments) due to soil erosion.

# From 2400 to the present: More subtle ecosystem changes

This period is characterised by the expansion of growing epilithic and epiphytic diatoms, suggesting that diversification of habitats is possibly linked to the increase in macrophyte vegetation. In parallel, the development of neutrophilous (e.g. E. minutum and S. pupula) and meso-eutrophic taxa (e.g. S. pupula, N. radiosa and S. construens var. venter) is associated with the weak increase in diatom productivity, which is also concomitant with nutrient enrichment (higher values of TOC and TN). This presupposes a general trend of lake acidification and nutrient enrichment during this period. From 800 cal. BP, higher abundances of S. pinnata and S. construens var. venter and an increase in diatom productivity (remaining lower than between 4800 and 4300 cal. BP) are observed. These recent changes are contemporaneous with several attempts at rural water resource management that have modified Lake Petit and other upstream lakes. These management projects include water supply galleries dug between AD 1820 and 1823 and the now-decommissioned AD 1947 dam (Figure 1). Even if hydraulic engineers did not attain their initial objectives, these activities have certainly modified connectivity between lakes, in terms, for instance, of changes in their water levels and water budgets, but estimating the potential impact of such changes in algal communities remains a difficult task and requires higher resolution analysis.

## Discussion

We have shown that the two major states, separated by a tipping point before and after 4200 cal. BP, that characterised the lacustrine ecosystem trajectories of Lake Petit were primarily linked to local landscape forcing. More subtle changes occur after this date. In this section, we first discuss whether such a response may have been controlled by abrupt or progressive climatic forcing. Following this, we examine the role of long-term human pressures in triggering an aquatic ecosystem-specific trajectory.

### Climate-driven ecosystem changes

Several studies have addressed the potential of diatom ecology for palaeoclimatic studies in circumpolar and alpine lakes rich in Fragilariaceae taxa. Even if diatom distribution is primarily determined by regional water chemistry, climatic gradients also influence algal communities (Laing and Smol, 2000; Lotter et al., 1997; Schmidt et al., 2004). The regime shift in local ecosystems such as that of Lake Petit, might be connected with (1) an abrupt climatic reverse during the 4200 cal. BP event and (2) a low-frequency climate regime shift at the mid–late Holocene transition (ca. 4000–3000 cal. BP).

**Role of abrupt climate change.** Comparison between the Lake Petit record and other palaeoenvironmental proxies strongly suggests a close link between the regime shift in ecosystems around 4200 cal. BP and climatic forcing. This regime shift in Lake Petit might be linked with the rapid climatic oscillation at 4200 cal. BP

identified in the Northern Hemisphere (Booth et al., 2005; Huang et al., 2011; Magny et al., 2009, 2012b; Staubwasser et al., 2003) and in the Mediterranean basin (Bruneton et al., 2002; Drysdale et al., 2006; Miramont et al., 2008; Roberts et al., 2011). This climatic event is contemporaneous with increases in water levels in Lakes Cerin (Figure 5h) and Bourget (Magny et al., 2012a) in Western Europe and Saint-Léger (Digerfeldt et al., 1997; Figure 5g), Ledro and Accesa (Magny et al., 2012b) in the Northern Mediterranean.

In Lake Petit, the important change in diatom assemblages between 4300 and 4100 cal. BP is contemporaneous with a major detrital pulse. Monospecific diatom assemblages (*S. pinnata*) and low diatom productivity could indicate that the lacustrine ecosystem was profoundly constrained by this disturbance. Runoff, by increasing water turbidity, could be a straightforward explanation for the drop in diatom productivity. This argues that more intense and/or more frequent rainfall occurred (Brisset et al., 2013). Regarding these results, it seems that climate has acted mainly as an indirect forcing agent on the lacustrine regime shift by initiating sudden soil erosion processes. These aspects illustrate well abrupt regime shifts due to variations in abiotic parameters in such fragile mountain environments.

Role of trend in climate change. Apart from the abrupt regime shift, a progressive decline in diatom productivity from 4800 to 3600 cal. BP occurs in Lake Petit, which may suggest a long-term effect of changing climate on ecosystem functioning. The highest productivity period (from 4800 to 4300 cal. BP) is related to a distinct dominance of S. pinnata, which is typically associated with high lake water alkalinity and elevated summer water temperatures (Michelutti et al., 2007; Pienitz et al., 1995; Podritske and Gajewski, 2007). In alpine environments, warmer periods have been linked to high lake water pH (Larsen et al., 2006; Michelutti et al., 2007; Psenner and Schmidt, 1992; Wolfe, 2002) and to high diatom productivity through a lengthening of the growing season (Lotter and Bigler, 2000; Smol et al., 2005). Cremer et al. (2001) notably identified an abundance of S. pinnata during the Holocene Climatic Optimum at Lake Basaltsø (East Greenland) and concluded that this diatom likely reflected an increase in lake productivity and higher lake water temperatures. We may, thus, interpret the long-term decrease in diatom productivity highlighted by the drop in S. pinnata frequencies as a progressive decline in the length of the growing season from the end of the Holocene Climatic Optimum until about 3600 cal. BP, probably due to an increase in ice cover duration and/or lower temperatures that reduce diatom growth. A progressive cooling from 4500 to 3300 cal. BP has been shown in the Austrian Alps based on a reconstruction of summer air temperatures inferred from chironomid assemblages at Lake Schwarzsee ob Sölden (Ilyashuk et al., 2011). This may be in agreement with the generally accepted fact that the mid-late Holocene transition (4200 cal. BP according to Walker et al. (2012)) would have been characterised by an increase in humidity and cooler temperatures as evidenced by the glacier advances of the Neoglacial period (Goehring et al., 2012; Holzhauser et al., 2005) and treeline regression (Nicolussi et al., 2005).

Starting from 3600 cal. BP, diatom productivity increases again from 2000 to 1500 cal. BP. In like manner, the rise in diatom productivity added to *S. pinnata* development might be a consequence of a longer growing season. This hypothesis is consistent with the glacier retreat phase in the western Alps during this period (Deline and Orombelli, 2005).

Finally, regarding the 'Little Ice Age', the most recent cold phase, well known in the historical corpus (Pichard and Roucaute, 2014) and by tree-ring growth studies (Corona et al., 2011), there is no evidence of change in diatom assemblages nor of a drop in diatom fluxes in Lake Petit that may allow for an unequivocal recognition of this climate oscillation.

### Human-induced changes in ecosystem functioning

Even if the main ecosystem change dated at 4200 cal. BP certainly appears as climatically driven, ecosystem responses since this date have been more subtle and have not been unequivocally attributed to climatic forcing. The weak increase in diatom productivity and the more meso-eutrophic species suggest that nutrient enrichment occurs from 2400 to 800 cal. BP (Figure 5c and d). This period is not strictly concomitant with a change in vegetation cover in the catchment. Indeed, the frequency of detrital inputs and of ruderal-anthropogenic pollen taxa had progressively increased since 4100 cal. BP. More noteworthy is the continuous presence and the increase in nitrophilous pollen taxa (e.g. *Urtica* and *Mentha*) from 2500 cal. BP, which indicates nitrogen enrichment of soils of the Lake Petit catchment related to intensive local grazing (Figure 5e).

Pollen evidence is supported by archaeological investigations (Suméra et al., 2008) revealing the presence of several pastoral enclosures in the Lake Petit catchment (Figure 1). Transhumant grazing since about the Neolithic period, around 7000 cal. BP, is recognised in pollen diagrams of the Southern Alps (Brisset et al., 2015; Walsh et al., 2014). Pastoral activities may alter alpine ecosystems by producing strong heterogeneities in the distribution of plant species and soil nutrients (Rossignol et al., 2006; Schrama et al., 2013; Singer and Schoenecker, 2003). Specifically, soil nutrient enrichment occurs principally along paths, in domestic livestock resting areas and inside pastoral enclosures. As nutrient transfers from pastures to water bodies occur as particulate nitrogen and organic phosphorus transported by runoff (Kurz et al., 2006), the rise in diatom productivity might have been significantly controlled by anthropogenic-induced nutrient input. Subsequently, diatom productivity decreased at 1300 cal. BP, while more soil erosion occurred in a catchment that had probably become treeless (Brisset et al., 2013).

While pollen data suggest an intensification of human impacts from 1300 cal. BP, diatom assemblages have remained relatively stable. Diatom assemblages indeed tend to show a new equilibrium in the status of the lake, thus suggesting that human pressures were sufficiently buffered by the aquatic ecosystem to preclude another regime shift.

# Can we define 'relevant baseline conditions' for Lake Petit?

Present-day management of aquatic ecosystems is often focused on (1) actions aimed at reversing environmental impacts inferred by the identification of specific disequilibrium (such as eutrophication or loss in diversity) and (2) conservation strategies of both biodiversity and habitats. In this framework, one of the major concerns is the integration of palaeoecology within conservation strategies as a key to understanding biotic and abiotic processes prior to anthropisation (Seddon et al., 2014). For instance, the crucial question of algal biodiversity changes implies acquisition of knowledge beyond ecological monitoring records. Therefore, taking into account long timescales requires enhanced understanding of complex ecosystem changes (especially interactions between terrestrial and lacustrine components), and biogeochemical or nutrient cycling (with the role of surrounding soils and of geology of catchment areas), both of which are rather poorly understood (Geist, 2011).

The palaeoecological study of Lake Petit offers a priori the opportunity to access, retrospectively, this knowledge and to reconstruct pre-anthropogenic lacustrine dynamics. If we assume, regarding diatom assemblages, that Lake Petit has not been significantly influenced by human activities prior to about 2400 cal. BP, we must acknowledge, however, that humans might have exacerbated the effect of climate on the watershed from 4200 cal. BP, via changes in vegetation and sensitivity to erosion, and consequently nutrient sourcing of the lake. Altogether, disentangling the influence of human activities and climate in such regime shift remains difficult since the effects of both are probably superimposed.

Furthermore, the choice of reference conditions implies somehow a 'freeze' of the ecosystem state in its past trajectory regardless of any chaotic event or any unpredictable ecosystem response to external forcing. Yet, as pointed out in a recent review (Vegas-Vilarrúbia et al., 2011), large palaeoecological datasets enable gauging the capacity of 'rare events' in generating ecosystem shifts between 'stable states'. Specifically, data collected at Lake Petit provide a good example of rapid change in the structuring of diatom diversity and habitats triggered by an abrupt event at 4200 cal. BP and by further small incremental changes triggered by human activities beyond a critical threshold.

### Conclusion

Diatom assemblages, organic composition of sediments and *Pediastrum boryanum* concentrations have revealed the lacustrine dynamics of Lake Petit since 5000 cal. BP. Our study has shown, in particular, that diatom assemblages responded rapidly to environmental changes allowing a better understanding of the forcing involved in the tipping point of this alpine environment at 4200 cal. BP. Our results have highlighted the following:

- 1. From 4800 to 4300 cal. BP, Lake Petit was a stable diatom-productive ecosystem dominated by *S. pinnata*, a species encountered in alkaline lakes and related to relatively high summer water temperatures. Nutrient and cation supply were provided by chemical weathering of podzol-type soils under conifer woodlands;
- This overall stability was interrupted following a sudden drop in diatom diversity and productivity between 4300 and 4100 cal. BP in response to a major detrital pulse during the climatic oscillation associated with this event;
- From 4100 to 2400 cal. BP, diatom productivity decreased and *Pediastrum boryanum* developed during the onset of the Neoglacial period. Diatom assemblages, subjected to terrigenous inputs from continuous erosion of pasture soils, were more diversified and mainly dominated by *P. robusta, P. brevistriata* and *P. pseudoconstruens*;
- Finally, from 2400 cal. BP to the present day, diatom assemblages show a slight acidification and enrichment of waters concomitant with increased human pressures in the catchment.

The tipping point in soil dynamics, following soil weakening by erosion processes favoured by deforestation, was reached at 4200 cal. BP probably due to a rapid change in precipitation regime. Soil destabilisation led, through a cascading set of responses, to a new lasting ecosystem state in Lake Petit. These conclusions illustrate the close link between ecosystems and how disturbance can propagate among all compartments of alpine catchments. In comparison, the increase in anthropogenic pressure in the Lake Petit catchment since 3000 cal. BP did not lead to a major regime shift. However, changes in pH and nutrient enrichment are recorded by diatom assemblages and might be an indirect response to pastoral activities in the catchment. Lower temperatures since 4200 cal. BP suggested by diatom assemblages could have participated in landscape changes by limiting soil formation and forest recovery. This climatic influence, in conjunction with increasing human pressure, certainly

contributed to the irreversible changes in both the terrestrial and lacustrine ecosystems. Regarding these results, we can ask the question as to whether recent forest recovery and decline in pastoralism could lead to a Lake Petit ecosystem similar to that of the period 4800–4200 cal. BP. Finally, taking into account long timescales in environmental management requires an enhanced understanding of complex ecosystem interactions. The results obtained at Lake Petit provide a good example of an abrupt regime shift of such fragile mountain systems in response to a rapid climatic event at 4200 cal. BP, further compounded by minor incremental changes triggered by human activities beyond a critical threshold.

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